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Atmos. Chem. Phys. Discuss., 8, 16585–16608, 2008 www.atmos-chem-phys-discuss.net/8/16585/2008/ © Author(s) 2008. This work is distributed under the Creative Commons Attribution 3.0 License.

This discussion paper is/has been under review for the journal *Atmospheric Chemistry and Physics (ACP)*. Please refer to the corresponding final paper in *ACP* if available.

Secondary organic aerosol formation from reaction of tertiary amines with nitrate radical

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Received: 1 July 2008 – Accepted: 6 August 2008 – Published: 1 September 2008

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Published by Copernicus Publications on behalf of the European Geosciences Union.





Abstract

Secondary organic aerosol formation from the reaction of tertiary amines with nitrate radical was investigated in an indoor environmental chamber. Particle chemistry was monitored using a high resolution aerosol mass spectrometer while gas-phase

- species were detected using a proton transfer reaction mass spectrometer. Trimethylamine, triethylamine and tributylamine were studied. Results indicate that tributylamine forms the most aerosol mass followed by trimethylamine and triethylamine respectively. Spectra from the aerosol mass spectrometer indicate the formation of complex non-salt aerosol products. We propose a reaction mechanism that proceeds via
 abstraction of a proton by nitrate radical followed by RO₂ chemistry. Rearrangement of
- the aminyl alkoxy radical through hydrogen shift leads to the formation of hydroxylated amides, which explain most of the higher mass ions in the mass spectra. These experiments show that oxidation of tertiary amines by nitrate radical may be an important night-time source of secondary organic aerosol.

15 **1** Introduction

Aliphatic amines constitute an important class of volatile organic compounds due to their malodorous properties (Chang et al., 2005; Filipy et al., 2006; Rappert et al., 2005). Amines have also shown a propensity to form aerosols (Murphy et al., 2007; Silva et al., 2008) which could have implications on health, visibility and climate change
²⁰ (Dockery et al., 1993; Seinfeld et al., 2003; Watson et al., 2002; IPCC, 2007). Amines are emitted from a variety of anthropogenic sources that include feedlots, car exhaust, biomass burning, sewage treatment plants and industries (Cadle et al., 1980; Westerholm et al., 1993; Mosier et al., 1973; Kataoka et al., 1996; You et al., 2002). In the biosphere, they result from bacterial degradation of nitrogen containing organic
²⁵ matter such as proteins and amino acids (Wright, et al., 1976). There is limited data on emissions of amines from the various sources, but they are generally associated

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with animal husbandry practices. Schade and Crutzen (1995) estimated the global emission of methylamines to be 150 Gg N/year, most of which was trimethylamine.

The lifetime of alkylamines in the atmosphere range from a few hours to tens of hours. Since they do not undergo photolysis in the atmosphere, their main removal
⁵ processes are reactions with atmospheric reactants such as hydroxyl radical, ozone, atmospheric acids and nitrate radical (Tuazon et al., 1994; Pitts et al., 1978; Angelino et al., 2001; Murphy et al., 2007). Some of the products resulting from these reactions may be highly volatile in which case they remain in the gas phase. Others may have sufficiently low vapor pressure to exist in particulate phase. A number of studies have sufficiently low vapor pressure to exist in particulate phase. A number of studies have al., 2002; Murphy et al., 1997; Zhang et al., 2003). Recent field measurements and lab experiments have shown that amine compounds can be a significant portion of aerosol mass especially during the night-time (Silva et al., 2008). However, the chemistry that leads to particulate amines is not well understood.

¹⁵ Few smog chamber experiments have been conducted to probe the chemistry of amines. These experiments have shown that amine reactions result in the formation of both volatile and condensed products. For example, photo-oxidation experiments by Pitts et al. (1978) of ethylamine and triethylamine in an outdoor chamber resulted in aerosol formation. Triethylamine yielded more aerosol than ethylamine. Amine aerosol

chemistry did not get much attention during this time as the focus was on formation of carcinogenic nitrosamines. Years later, studies have shown that the chemistry that leads to the formation of aerosol is more complex and is dependent on the class of amine reacted (Angelino et al., 2001; Murphy et al., 2007). Tertiary amines in particular react to form more aerosol mass than secondary and primary amines (Murphy et al., 2007).

25 2007). None of these studies looked at the chemistry of amines with the nitrate radical. In our recent work, we reported results from a series of experiments in which trimethylamine was reacted with ozone, nitrate radical and nitric acid (Silva et al., 2008). These results showed that the reaction of trimethylamine with nitrate radical produced mostly secondary organic aerosol. Formation of aminium salts was minimal. The re-



action also exhibited relatively fast kinetics and high aerosol yields. This paper builds on that work in order to achieve a three-fold objective:

- i) to extend the study of trimethylamine to other tertiary amines,
- ii) to examine the products formed and
- ⁵ iii) to devise a reaction mechanism that is consistent with the observations.

The amines under investigation here are trimethylamine (TMA), triethylamine (TEA) and tributylamine (TBA).

2 Experimental

A series of experiments were conducted in the UC Riverside/CE-CERT environmental
 chamber. The chamber facility has been previously described by Carter et al. (2005). Briefly, the environmental chamber consists of a temperature controlled enclosure that is continually flushed with purified air. Two 90 m³ Teflon (2 mil FEP Teflon) film reactors are hung on a rigid frame that descends during the experiment to maintain a slight positive differential pressure between the bags and the enclosure. All experiments were
 conducted under dry (RH<0.1%) conditions. A Dasibi Environmental Corp. 1003-AH Ozone analyzer monitored ozone concentration. Thermal Environmental Instruments model 42C chemiluminescent NO_x analyzer monitored NO, NO₂ and NO_x, for all ex-

- periments. Particle size and concentration was measured throughout the experiment using an in-house built scanning mobility particle spectrometer (SMPS) located inside
- the chamber enclosure. The SMPS consists of a TSI model 3077 ⁸⁵Kr neutralizer, a TSI model 3081 long column cylindrical differential mobility analyzer, and a TSI model 3760A condensation particle counter. A voltage scan from 40 V to 7000 V provided particle size distribution from 28 nm to 730 nm. Aerosol volumes were wall-loss corrected using a first order decay rate based on number concentrations (Bowman et al., 1997).

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A proton transfer reaction mass spectrometer (PTR-MS, Iconicon Analytik) was used to detect and analyze volatile organic compounds in real time and with high sensitivity. A detailed description of PTR-MS has been given by Lindinger et al. (1997) and only a summary of the key features will be given here. The instrument consists of four main components: an ion source, a drift tube, a mass analyzer (quadrupole), and an ion detector/amplifier. H_3O^+ ions are produced at high concentrations from pure water vapor within a hollow cathode ion source and pass via a Venturi-type inlet into the drift tube. The air sample to be analyzed is introduced into the drift tube and because of their low proton affinities, the major components of air undergo non-reactive collisions with H_3O^+ ions and therefore act as buffer gas. However, any collisions of H_3O^+ ions with the volatile organic compounds (VOCs) present in the gas flow possessing a greater proton affinity than water will result in a proton transfer reaction (Eq. R1).

 $VOC(g)+H_3O^+(g)\rightarrow VOCH^+(g)+H_2O(g)$

(R1)

The resultant ions are mass-selected using a quadrupole mass analyzer and measured as count rates by an electron multiplier detector. The online PTR-MS was used in scan mode to monitor the ion traces from m/z 21 to m/z 200 with a dwell time of 0.2 s per mass.

The aerosol mass spectrometer (AMS, Aerodyne Inc.) has previously been discussed (Jayne et al., 2000; Jimenez et al., 2003; Canagaratna et al., 2007). For the experiments described here, the high resolution time of flight mass spectrometer (HR-ToF-AMS) was used and its details can be found in a paper by DeCarlo et al. (2006). Briefly, it is made up of three differentially pumped chambers; the sampling chamber,

- the particle time-of-flight chamber and the analysis chamber. Pressure differential between the smog chamber air and the AMS sampling chamber continuously draws air sample into the instrument at a nominal flow rate of about 1.2 cm s⁻¹ where it passes through an aerodynamic lens that collimates the sample into a tight beam while removing as much gas phase species as possible. After the particles exit the lens, they
- ¹⁵ enter into the particle ToF region then accelerated until they impact the oven. The non-refractory components are flash vaporized at about 600°C and then subjected to a



beam of electrons at 70eV that ionizes them before being sorted by a time of flight mass analyzer and detected using an MCP detector. Ion measurements were alternated between a shorter flight path (V-Mode) that gives higher ion throughput and therefore better sensitivity and the longer flight path (W-mode) that offers better separation but lower sensitivity. The strength of HR-ToF-AMS lies in its ability to offer a high resolving power of up to 4300 times for W-mode and about 2100 times for V-mode (at m/z 200) better than that of the guadrupole AMS (Q-AMS). This ability was useful in definitive

- identification of the product peaks. The AMS was operated in the scanning mode in the mass range between m/z 10 to m/z 500.
- 10 2.1 Experimental protocol

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Each experiment started with thorough cleaning of the bags as outlined in Carter et al. (2005). Background concentrations of NO, and ozone were also determined to be at an acceptable low level (<2 ppb for ozone and NO_{x}). The amines were introduced into the chamber by injecting a known volume of liquid amine into a small glass injection manifold or by introducing a known volume of the gas phase amine into a calibrated bulb based on calculated partial pressures. Pure nitrogen was then passed over the liquid or through the bulb, flushing the amine into the reactors. Injections of amines were verified using PTR-MS. Table 1 lists the initial conditions of all experiments in this paper. After amine injection, ozone was introduced by passing 20 psig of pure air through two UV ozone generators and the reaction was left to proceed until the amount of wall-loss corrected aerosol formed stabilized (usually 1-2 h). Then NO_x was added to the reactors as NO using a calibrated bulb (usually $\sim 1-2h$ after ozone addition). The reaction was left to proceed until the amount of wall-loss corrected aerosol formed leveled off (usually 2–3 h after injection of NO_x). The ozone and NO₂ in the chamber react to form nitrate radical as shown in Eq. (R2). All the experiments were conducted in the dark to simulate night-time NO₃ radical chemistry.

 $O_3 + NO_2(g) \rightarrow NO_3(g) + O_2(g)$



(R2)

3 Results

3.1 TMA reaction

Smog chamber reactions of TMA with nitrate radical have been reported by Silva et al. (2008). The discussion here will only dwell on details pertinent to the objectives of this paper. Figure 1a shows a typical TMA reaction profile. In this experiment, some aerosol formed after injection of ozone. The aerosol concentration leveled off at about $30 \,\mu g \,m^{-3}$ (This was obtained from wall-loss corrected SMPS volume concentration using density of $1 \,g \,cm^{-3}$). Four hours after ozone injection, NO was introduced leading to more aerosol formation leveling off at about $90 \,\mu g \,m^{-3}$. TMA quasi parent ion peak (m/z 60) was monitored by the PTR-MS. This peak started decaying immediately ozone was injected. The decay continued as more TMA was converted into aerosol phase.

Table 2 displays the high intensity peaks that were detected with the PTR-MS and the assigned VOCs. Peaks at m/z 31 (formaldehyde), m/z 45 (acetaldehyde), m/z47 (formic acid), m/z 59 (acetone/propanal) and m/z 61 (acetic acid/propanol) were present in all the three tertiary amines reacted. During the TMA reaction, the PTR-MS also detected m/z 74 (dimethyl formamide), m/z 88 (methyldiformamide), m/z102 (triformamide), m/z 91 (dimethylnitramine) and m/z 75 (dimethylnitrosamine). As expected, dimethylnitramine and dimethylnitrosamine were detected only after injection

of NO_x. The ion peak detected at m/z 44 was specific to TMA reaction and it likely represents methyl-methyleneimine (CH₂=NCH₃). Since imines have double bonds, they can potentially be attacked by ozone to form other products, but at least one study has indicated that this reaction is too slow to be of any significance (Mori et al., 1988).

An average mass spectrum of the particulate products from the TMA reaction ²⁵ with the ozone/NO_x system is shown in Fig. 1b. The ion peaks at m/z 44.052 (C₂H₆N⁺) and m/z 58.066 (C₃H₈N⁺) represent amine backbone fragments while m/z58.030 (C₂H₄NO⁺), 76.040 (C₂H₆NO₂⁺), m/z 88.037 (C₃H₆NO₂⁺) and m/z 104.040 (C₃H₆NO₃⁺) represent fragments of amines that have been oxidized. Table 3 summa-

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rizes other major ion fragments detected. We examined spectra for evidence of amine-N-oxide, a possible product according to Angelino et al. (2001). Trimethylamine-Noxide (TMAO) parent-ion peak is m/z 75. The ion peak at m/z 75 was detected albeit at low intensity. The HR-ToF-AMS spectrum (Fig. 2a) indicates it consists of two peaks,

⁵ m/z 75.029 (C₂H₅NO₂⁺) and m/z 75.065 (C₃H₉NO⁺). As shown in Fig. 2b, after injection of NO_x only m/z 75.029 was detected. This indicates that TMAO was present after the reaction with ozone, but was reduced after NO_x injection. There was no protonated TMAO (m/z 76.073) detected in the mass spectra during either injection. The plausible explanation for this observation is that ozone indeed reacts with TMA to form TMAO, but it may be vulnerable to radical chemistry itself to form other products.

3.2 TEA reaction

The reaction of TEA with ozone/NO_x followed a pattern similar to that of TMA. After ozone injection, some aerosol was formed leveling off at about $20 \,\mu g \,m^{-3}$. Two hours later, NO was injected resulting in formation of more aerosol that leveled off at about $40 \,\mu g \,m^{-3}$ (Fig. 3a). The amount of aerosol from TEA reaction was less than that of TMA even though TEA has higher molecular weight and should theoretically form more aerosol. Immediately after NO injection, there was sudden buildup of aerosol that quickly dropped before leveling off.

Injection of TEA was confirmed with PTR-MS by monitoring the quasi parent ion peak $(m/z \ 102)$ that gradually decayed during the course of the experiment as more aerosol was being formed. Other volatile products detected specific to the TEA reaction include amides: diethylacetamide $(m/z \ 119)$ and ethylformamide $(m/z \ 74)$; nitro/nitroso species: diethylnitramine $(m/z \ 116)$ and diethylnitrosamine $(m/z \ 103)$ and an imine: ethyl-ethyleneimine $(m/z \ 72)$.

²⁵ The AMS spectrum for TEA reaction is shown in Fig. 3b. The low mass peaks for these reactions include m/z 44.052 ($C_2H_6N^+$), m/z 58.065 ($C_3H_8N^+$), m/z 72.082 ($C_4H_{10}N^+$) and m/z 86.095 ($C_5H_{12}N^+$) that represent fragments from the amine backbone. Other ions detected are m/z 72.046 ($C_3H_6NO^+$), m/z 86.068 ($C_4H_8NO^+$), m/z



102.05 ($C_4H_8NO_2^+$) and m/z 146.07. ($C_6H_{12}NO_3^+$) In each of these cases, one, two, or three oxygen atoms have been added representing oxidized amine fragments. There were fewer high mass peaks in the TEA reaction compared to that of TMA. Shortly after the NO injection, the nitrate fragments (NO^+/NO_2^+) were detected as shown in

- ⁵ Fig. 4a–d. The salt burst was short-lived and did not appear in the other two amines investigated. It is likely that this observation was as a result of sudden build up in nitric acid in the chamber near the injection/sampling port. Apart from this instantaneous period, nitrate salt formation was generally minimal as depicted by the low intensity of NO⁺ and NO⁺₂ peaks at m/z 30 and m/z 46 respectively.
- 10 3.3 TBA reaction

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Figure 5a presents results of the TBA reaction with ozone/NO_x. The amount of aerosol formed after ozone injection (at ~60 μ g m⁻³) was more than the total amount of aerosol formed during the TEA-ozone/NO_x reaction. This was expected because TBA is a twelve-carbon compound and the oxidation products are expected to have lower vapor pressure. The maximum aerosol recorded for the TBA reaction was ≈160 μ g m⁻³ after injection of NO approximately 3 h after ozone injection.

PTR-MS quasi parent peaks of m/z 60 (TMA) and m/z 102 (TEA) showed clear decay in the course of the experiment. This was not the case with TBA quasi parent peak $(m/z \ 186)$ which was not detected even though TBA has the highest proton affinity of the three tertiary amines (Green-Church et al., 2000). A possible explanation could be that the ion transmission efficiency was lower for TBA parent ion than the other two amines. Volatile products detected by the PTR-MS for the TBA reaction were similar to those formed by TMA and TEA (Table 2) and include an amide (dibutylbutanamide; m/z 200), a nitro-compound (dibutylnitramine; m/z 175), a nitroso-compound (dibutylnitrosamine; m/z 159) and an imine (butyl-butyleneimine; m/z 128).

Figure 5b shows the AMS mass spectrum for the TBA reaction where high intensity ion peaks were detected at m/z 44.052 (C₂H₆N⁺), m/z 58.065 (C₃H₈N⁺), m/z

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72.082 (C₄H₁₀N⁺), m/z 86.095 (C₅H₁₂N⁺) and m/z126 (C₈H₁₆N⁺). The oxygenated fragments from TBA reaction were m/z 100 (C₅H₁₀NO⁺), m/z 142 (C₈H₁₆NO⁺), m/z 174 (C₈H₁₆NO₃⁺) and m/z 230 (C₁₂H₂₄NO₃⁺). Others ions are listed in Table 3.

3.4 Reaction mechanism

- ⁵ Patterns in both gas phase and aerosol phase products emerged from all three tertiary amines reacted. Homologous volatile products include amides, nitro/nitroso compounds and imines. Formation of aminium salt with nitric acid was minimal, based on the low intensities of nitrate ions (m/z 30 and 46). The high mass fragments $C_3H_6NO_3^+$ (m/z 104), $C_6H_{12}NO_3^+$ (m/z 146) and $C_{12}H_{24}NO_3^+$ (m/z 230) from TMA,
- TEA and TBA reactions are similar. These high mass fragments indicate that each tertiary amine gains three oxygen atoms during the reaction while losing three hydrogen atoms. Because molecules containing one nitrogen atom in the structure must have an odd molecular mass, the ion peak at *m*/*z* 104 for TMA (C₃H₆NO₃⁺) is a fragment ion of a larger molecule. Further, the degree of unsaturation suggests the structure contains
 only one double bond. This ion is replicated for TEA at *m*/*z* 146 (C₆H₁₂NO₃⁺) and for TBA at *m*/*z* 230 (C₁₂H₂₄NO₃⁺). The structure for all three tertiary amines is consistent with the formation of a di-hydroxylated amides. This structure can account for all high mass fragments present in the spectra for all three compounds with the exception of the TMA reaction where it appeared to be forming dimers, possibly oligomers (Silva et an et al. 2007).

²⁰ al., 2008).

With the high resolution aerosol mass spectra and gas phase species detected by PTR-MS, we propose a reaction mechanism for the ozone/NO_x interaction with tertiary amines as shown in Fig. 6. Some of the products detected are shown with bold numbers (1-11) in Fig. 6 and the text below.

The reaction with the nitrate radical starts with an initial abstraction step of the hydrogen atom from one of the alkyl groups in the amine. The resulting amine alkyl radical then reacts solely with O_2 to form an aminyl alkylperoxy radical, which reacts with NO to form an aminyl alkoxy radical. The aminyl alkoxy radical then either reacts

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with O₂ to form dimethylformamide (1) or decomposes to form formaldehyde (2) and an aminyl alkyl radical. The resulting aminyl alkyl radical can react with NO or NO₂ to form dimethylnitrosamine (3) and dimethylnitramine (4) respectively or it can lose a proton to form methyl-methyleneimine (5). This procedure, starting with the initial abstraction of proton from the alkyl group can repeat itself until methyldiformamide (6) and triformamide (7) are formed.

For aerosol formation, we propose that the reaction proceeds via rearrangement of the aminyl alkoxy radical through 1,4 H-shift to form a hydroxyaminyl alkyl radical that then reacts with oxygen, NO and oxygen again to form hydroxymethyl-¹⁰ methylformamide (8). This procedure can be repeated until formyl-hydroxymethylformamide (9) and N, N-bis(hydroxymethyl) formamide (10) are formed. The vapor pressures of the hydroxylated products are calculated to be much lower than the corresponding aldehydes. For example, the theoretical estimation of the vapor pressure of dihydroxymethyl-formamide is ~ 6.38×10^{-8} Torr, at 25°C, and would therefore be ¹⁵ consistent with a condensed phase species (Scifinder Scholar, 2008). The major high mass AMS peaks for TMA reaction such as $C_3H_6NO_3^+$ (m/z 104) and $C_3H_6NO_2^+$ (m/z 88) can well be explained from loss of proton and OH respectively from N, Nbis(hydroxymethyl) formamide. Formation of aerosol via the ozone only route has been previously reported by Tuazon et al. (1994). In this mechanism for example, TMA is

²⁰ oxidized to TMAO (11) as hypothesized by Angelino et al. (2001).

Since TMA products exhibit oligomeric properties, this is possible through fusing together of two or more $C_3H_6NO_3^+$ at the OH end with loss of one proton resulting in a repeating unit of m/z 103. This explains AMS peaks such as m/z 207 (two $C_3H_6NO_3^+$ with a loss of proton) and m/z 191 (two $C_3H_6NO_3^+$ with a loss of OH). Since this repeating behavior was not seen in TEA and TBA (perhaps due to steric hindrance), it also explains why TMA, a three-carbon compound formed more aerosol

than TEA which is a six-carbon compound.

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4 Implication

This study shows that tertiary amines, major constituents of animal husbandry practices do react with nitrate radical, an important night-time oxidant. The products include both gas phase species and secondary organic aerosol formation (SOA).

Acknowledgements. This research was supported by the National Science Foundation under grants ATM-0735396 and ATM-0449778 and the W. M. Keck Foundation.

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Table 1. Initial conditions and results obtained from reaction of tertiary amines with ozone/NO_x.

Amine	[Amine]	[NO _x]	[Ozone]	$[Aerosol]_{max}$
	(ppb)	(ppb)	(ppb)	$(\mu m^3 cm^{-3})$
Trimethylamine	100	50	300	90
Triethylamine	100	40	245	40
Tributylamine	100	35	255	110

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Table 2. Major PTR-MS	S peaks and	assigned	VOCs.
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TMA			TEA			TBA		
m/z	Formula	VOC	m/z	Formula	VOC	m/z	Formula	VOC
31	НСНО	Formaldehyde	31	HCHO	Formaldehyde	31	НСНО	Formaldehyde
33	CH ₃ OH	Methanol	45	CH ₃ CHO	Acetaldehyde	45	CH ₃ CHO	Acetaldehyde
44	CH ₃ NCH ₂	Methylmethyleneimine	47	HCOOH/CH3CH2OH	Formic Acid/Ethanol	47	HCOOH CH ₃ CH ₂ OH	Formic Acid Ethanol
45	CH ₃ CHO	Acetaldehyde	59	(CH ₃) ₂ CO CH ₃ CH ₂ CHO	Acetone Propanal	59	(CH ₃) ₂ CHO CH ₃ CH ₂ CHO	Acetone Propanal
47	HCOOH CH ₃ OH	Formic Acid Ethanol	60	CH ₃ CH ₂ NO	Nitrosoethane	61	CH ₃ COOH CH ₃ (CH ₂) ₂ OH	Acetic acid Propanol
59	(CH ₃) ₂ CO/CH ₃ CH ₂ CHO	Acetone/Propanal	61	(CH ₃) ₂ CO/CH ₃ CH ₂ CHO	Acetic Acid/Propanol	73	CH ₃ (CH ₂) ₂ CHO	Butanal
60	(CH ₃) ₃ N	Trimthylamine	72	CH ₃ CH ₂ NCHCH ₃	Ethylidine-Ethanamine	75	CH ₃ (CH ₂) ₃ OH	Butanol
61	CH ₃ COOH CH ₃ (CH ₂) ₂ OH	Acetic Acid Propanol	74	CH ₃ CH ₂ NCHO	EthylFormamide	100	CH ₃ (CH ₂) ₃ NO ₂	Nitrobutane
62	CH ₃ NO ₂	Nitromethane	76	CH ₃ CH ₂ NO ₂	Nitroethane	128	C4H9NCH(CH2)2CH3	Butylimine
74	(CH ₃) ₂ NCHO	Dimethylformamide	102	(CH ₃ CH ₂) ₃ N	Triethylamine	159	[CH ₃ (CH ₂) ₃] ₂ NNO	DibutyInitrosamine
75	(CH ₃) ₂ NNO	Dimethylnitrosamine	103	(CH ₃ CH ₂) ₂ NNO	Diethylnitrosamine	175	[CH ₃ (CH ₂) ₃] ₂ NNO ₂	DibutyInitramine
88	CH ₃ N(CHO) ₂	Methyldiformamide	116	(CH ₃ CH ₂) ₂ NCH ₂ CHO	Diethylacetamide	186	[CH ₃ (CH ₂) ₃] ₃ N	Tributylamine
90	CH ₂ OHCH ₃ NCHO	Hydoxymethylmethylformamide	119	(CH ₃ CH ₂) ₂ NNO ₂	Diethylnitramine	200	[C4H9]2N(CH2)3CHO	Dibutylbutanamide
91	(CH ₃) ₂ NNO ₂	Dimethylnitramine						
102	(CHO) ₃ N	Triformamide						

Table 3. Major AMS	ion peaks and	assignments.
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m/z	TMA Formula	m/z	TEA Formula	m/z	TBA Formula
				,-	
30.010	CH ₂ O	30.000	NO	44.027	C ₂ H ₄ O
30.036	CH_4N^+	30.036	CH_4N^+	44.051	C₂H ₆ N ⁺
43.982	CO_2^+	44.024	$C_2H_4O^+$	58.065	C₃H ₈ N⁺
44.052	$C_2 \overline{H}_6 N^+$	44.052	$C_2 H_6 N^+$	72.080	$C_4H_{10}N^+$
45.993	NO ₂ ⁺	58.066	C ₃ H ₈ N ⁺	86.092	$C_5H_{12}N^+$
46.03	$CH_4^-NO^+$	72.046	C ₃ H ₆ NO ⁺	100.080	$C_5H_{10}NO^+$
58.066	C₃H ₈ N⁺	72.082	$C_4H_{10}N^+$	114.090	$C_6H_{12}NO^+$
76.040	$C_2H_6NO_2^+$	86.068	C ₄ H ₈ NO ⁺	126.130	$C_8H_{16}N^+$
88.040	$C_3H_6NO_2^+$	86.095	$C_5H_{12}N^+$	142.120	C ₈ H ₁₆ NO⁺
104.040	$C_3H_6NO_3^+$	102.050	$C_4H_8NO_2^+$	158.120	$C_8H_{16}NO_2^+$
191.070	$C_{6}H_{11}N_{2}O_{5}^{+}$	130.080	$C_6H_{12}NO_2^+$	174.110	$C_8H_{16}NO_3^+$
207.060	$C_{6}H_{11}N_{2}O_{6}^{+}$	146.070	$C_6H_{12}NO_3^{\overline{+}}$	214.180	$C_{12}H_{24}NO_{2}^{+}$
	C C		Ũ	230.170	C ₁₂ H ₂₄ NO ₃ [∓]

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Fig. 4. Expanded product peaks for the TEA reaction (from HR-ToF-AMS): (a) m/z 30 after ozone injection, (b) m/z 30 shortly after NO_x injection, (c) m/z 46 after ozone injection, (d) m/z 46 shortly after NO_x injection.

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Fig. 5. Panel **(a)** Temporal profiles of m/z 186 (from PTR-MS) and aerosol volume concentration (from SMPS) for the reaction of TBA with ozone/NO_x. Panel **(b)** Aerosol mass spectrum taken at the end of the experiment after the SMPS volume concentration leveled off.







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Fig. 6. Proposed mechanism for the reaction of TMA with nitrate radical. TMA oxidation by ozone (Tuazon et al., 1994) is also shown.

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