

# Average molecular weight of surfactants in aerosols

M. T. Latif<sup>1</sup> and P. Brimblecombe<sup>2</sup>

<sup>1</sup>School of Environmental and Natural Resource Sciences, Faculty of Science and Technology, Universiti Kebangsaan Malaysia, 43600 Bangi, Malaysia

<sup>2</sup>School of Environmental Sciences, University of East Anglia, NR4 7TJ Norwich, UK

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Correspondence to: M. T. Latif (talib@pkrisc.cc.ukm.my)

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## Abstract

Surfactants in atmospheric aerosols determined as methylene blue active substances (MBAS) and ethyl violet active substances (EVAS). The MBAS and EVAS concentrations can be correlated with surface tension as determined by pendant drop analysis.  
5 The effect of surface tension was more clearly indicated in fine mode aerosol extracts. The concentration of MBAS and EVAS was determined before and after ultrafiltration analysis using AMICON centrifuge tubes that define a 5000 Da (5 K Da) nominal molecular weight fraction. Overall, MBAS and to a greater extent EVAS predominates in fraction with molecular weight below 5 K Da. In case of aerosols collected in Malaysia the  
10 higher molecular fractions tended to be a more predominant. The MBAS and EVAS are correlated with yellow to brown colours in aerosol extracts. Further experiments showed possible sources of surfactants (e.g. petrol soot, diesel soot) in atmospheric aerosols to yield material having molecular size below 5 K Da except for humic acid. The concentration of surfactants from these sources increased after ozone exposure  
15 and for humic acids it also general included smaller molecular weight surfactants.

## 1 Introduction

The extent and composition of the organic layer which typically covers the surface of atmospheric particles affects surface related aerosol properties such as effect on health, the ability of the particles to activate and form cloud droplets, and the aggregation of particles as well as reactions between the particle and gas phase species  
20 (Peterson and Tyler, 2002, 2003). Since first mentioned by Blanchard (1964), surface active materials in atmospheric aerosols were believed to be derived from the marine environment (Cini and Loglio, 1997; Giovannelli et al., 1988; Loglio et al., 1986, 1985; Nicolotti et al., 2005; Oppo et al., 1999; Rettori et al., 2005). Surface active materials in  
25 the atmosphere are still under study so some points, especially regarding their composition and origin remain under debate. Nevertheless it is clear that they are associated

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with the water soluble oxygenated and macromolecular polar organic substances of fine mode aerosols (Latif and Brimblecombe, 2004).

The exact chemical nature of candidate surface active components in atmospheric aerosols is still did not clear. Several studies showed that the surface active components of both marine and continental aerosols is dominated by aliphatic hydrocarbons and their amphiphilic derivatives (Peterson and Tyler, 2002). This was supported by study of Gilman et al. (2004) which indicated that the carbon length of amphiphilic compound is the most important factor in determining its interfacial stability and longevity. Analysis using TOF-SIMS by Tervahattu et al. showed a preponderance of long chain fatty acids at the surface of atmospheric aerosol particles (Tervahattu et al., 2002a, b, 2005).

Other researchers suggest that the surface active component in atmospheric aerosols correlated with water soluble organic carbon and humic like substances (HULIS) (Bower et al., 2000; Krivacsy et al., 2000; Zappoli et al., 1999). According to Facchini et al. (1999a) surfactants are particularly polyacidic compounds of high molecular mass represent an average 25% of the organic solute. The upper limit for the molecular weight of these compounds, likely to consist of mono and di-carboxylic acid and polycarboxylic acids (Bower et al., 2000), is estimated as 3070 Da (Zappoli et al., 1999). Other studies suggest that the molecular weight of humic like substances in atmospheric aerosols is below 1000 Da (Havers et al., 1998; Kiss et al., 2003; Samburova et al., 2005).

Humic substances are typically seen as soil components, and Aitken et al. (1985) characterized these as yellow to black refractory organic substances of high molecular weight. The components of humic materials are separated on the basis of molecular weight where fulvic acid is typically 500 to 5000 daltons (Da), while humic acid is >5000 Da and can extend to values >1 000 000 Da (Kiss et al., 2003; Yates and Wandruszka, 1999). Likens et al. (1983) showed that organic substances of molecular weight (MW) >1000 Da in precipitation accounted for more than 50% of the total dissolved organic carbon. High molecular weight carbon compounds are believed to be

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involved in cloud condensation (Facchini et al., 1999b; Krivacsy et al., 2000).

The origin of macromolecular humic substances in atmosphere is still unclear. A small amount could derive from wind blown soil. However, (i) secondary production from high molecular weight substances emitted from soil (Mezaros et al., 1997) , (ii) direct biogenic emission and (iii) secondary formation from anthropogenic precursors are thought to be important (Feng and Moller, 2004). They could also derive from photochemical oxidation of volatile organic compounds (Went, 1960). Ozonolysis of organic compounds on aerosols surfaces can alter their chemical properties of (Chughtai et al., 1999) and make the particles more hydrophilic (Weingartner et al., 1997) hinting at a heterogeneous pathway for seconary production of surfactants.

We have studied the concentration of surfactants in the atmosphere as anionic methylene blue and ethyl violet active substances (Sukapan and Brimblecombe, 2002; Latif and Brimblecombe, 2004, and Latif et al., 2005). This activity appears to be related to the presence humic like substances (HULIS), so the present work has aimed at determining the range of nominal molecular weight of such surfactants present on ambient aerosols.

## 2 Materials and method

A range of experimental measurements determined various characteristics of our samples.

### 2.1 Surfactant Concentration

Surfactant concentration measured as *Methylene Blue Active Substances* (MBAS) and *Ethyl Violet Active Substances* (EVAS) used in this study was based on the analysis of samples collected at University of East Anglia (UEA), Norwich in United Kingdom, Universiti Kebangsaan Malaysia (UKM), Bangi and Engineering Campus, Universiti Sains Malaysia (USM), Penang and Universiti Malaysia Sabah (UMS), Kota Kinabalu

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as mentioned in previous publications (Latif and Brimblecombe, 2004, 2005).

## 2.2 Surface Tension

Surface tension analysis was carried out with the pendant drop method using a special needle to create a droplet from aerosol extract. A digital camera (Pulnix TM500) was used in conjunction with a Matrox Meteor frame grabber and Image-Pro Plus (Media Cybernetics, USA) software to determine surface tension based on droplet geometry as mentioned in Latif et al. (2005).

## 2.3 Surrogate Acidity

We used the non-seasalt components (nss) of the aerosol as a surrogate of acidity. This was based on the concentration of  $\text{SO}_4^{2-}$ , and  $\text{NO}_3^-$  from the data of Latif et al. (2005) and estimated from the sum:  $2\text{nss}[\text{SO}_4^{2-}] + \text{nss}[\text{NO}_3^-]$

## 2.4 Dissolved Organic Carbon

Total dissolved organic carbon (DOC) in atmospheric aerosols was determined using a Shimadzu TOC-500 carbon analyser. Before analysis, samples were acidified to pH 2 with HCl (5N) (10  $\mu\text{L}$  per 10 mL samples) and bubbled with nitrogen for 10 min to eliminate interference from inorganic carbon. Samples (100  $\mu\text{L}$ ) were directly injected into vertical furnace, filled with platinum-coated silica and oxidised at 700°C. The  $\text{CO}_2$  concentration resulting from combustion was analysed by non-dispersed infrared detector.

## 2.5 Molecular weight of MBAS and EVAS

Centrifugal filter devices, Amicon ultra-15 (PL-5), with a nominal molecular weight limit below 5000 Da (5 K Da), were used to separate the aerosol extracts. The ultrafiltration

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set was pre-rinsed using ultra pure water before being used to avoid the effect of trace surfactants within the apparatus.

The aerosol extracts (15 mL) were spun in a centrifuge (Mistral 1000) at 4000  $\times$  g for thirty minutes. The volume of each aerosol extract after ultrafiltration process was made up to 20 mL in a volumetric flask before being analysed for MBAS and EVAS (as described in Latif and Brimblecombe, 2005). Part of the extract (10 mL) was used for MBAS analysis and another (4 mL) was used for EVAS analysis.

## 2.6 Molecular weight of the Brown Colourant

Centrifugal filter devices, Amicon ultra-15 (PL10 and PL30) with different nominal molecular weight limit, 10 K Da and 30 K Da were used to separate the brown coloured aerosol extracts. The procedures prior to surfactants analysis are as mentioned Sect. 2.1.

## 2.7 Ultra violet/visible spectra of yellow-brown extracts

Ultra violet/visible spectra of the aerosol extracts were determined using a Perkin Elmer UV/Vis spectrometer (Lambda 35) from the range of 200 to 600 nm wavelength using ultra pure water as a blank.

## 2.8 UV Photodegradation

Aerosol extracts (20 mL) in quartz samples tube were photodegraded by exposure to a high intensity UV light from a 1000 W Hanovia mercury lamp for two hours and 24 h. Samples were fan-cooled to maintain a temperature of 50–55°C. The solutions were analysed for MBAS and EVAS after exposure and compared to concentrations in each solution prior to exposure.

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## 2.9 Soot exposure to ozone

Dry diesel and petrol soot and humic acid (50 mg) was exposed to ozone in a Dreschel bottle. The ozone was generated from pure oxygen in a laboratory ozoniser, MX II, in the fume cupboard. The flow rate of ozone from the ozone generator was controlled at 1 l per minute, which gave a concentration in the solution of  $20 \text{ g L}^{-1}$ . The extracts were shaken and filtered using  $0.2 \mu\text{m}$  47 mm cellulos acetate filter papers (Whatman) and a vacuum pump. Concentrations of MBAS and EVAS in the extracts were analysed as described in Latif and Brimblecombe (2004).

## 3 Results and discussion

### 3.1 Surface tension of aerosol extracts

Measurements of surface tension within the pendant drops generated from aerosols both in fine mode and coarse mode aerosol extracts was below  $72 \text{ dyne cm}^{-1}$ , which indicates that there are present surfactants in the aerosols (Fig. 1a and Fig. 1b). These new measurements add to earlier results reported by Facchini et al. (2000); Tucker-mann and Cammenga (2004). We see that the reduction in surface tension is most notable with the extracts from fine mode aerosols (aerosol with diameter size  $< 1.5 \mu\text{m}$ ). The surface tension of the pendant drop generated from fine aerosol extracts significantly reduces with time as surfactant accumulates on the surface.

### 3.2 Correlation with surrogate acidity

Results indicated that the concentration of MBAS and EVAS (Fig. 2a and Fig. 2b) were influenced by the surrogate acidity of aerosol extracts. The estimates of surrogate acidity concentration in aerosols samples collected in Malaysia as can be seen Fig. 3a and Fig. 3b indicating a strong correlation (MBAS:  $n=28$ ,  $r=0.68$ ; EVAS:  $n=28$ ,  $r=0.84$ ) with anionic surfactants as MBAS and EVAS. This suggests that acidic pollutants that

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are likely to come from combustion activities, especially motor vehicles and biomass burning around these sampling stations are related to the surfactants.

### 3.3 Correlation to Dissolved Organic Carbon (DOC)

The results of correlation between anionic surfactants as MBAS and EVAS, and DOC in aerosols are shown in Fig. 4a and Fig. 4b. Overall MBAS in coarse mode aerosols ( $n=39$ ,  $r=0.35$ ,  $p<0.05$ ) and fine mode aerosols ( $n=39$ ,  $r=0.56$ ,  $p<0.05$ ) have significant correlation with DOC. The significant correlation also found between EVAS and DOC ( $n=39$ ,  $r=0.57$ ,  $p<0.05$  for coarse mode and  $n=39$ ,  $r=0.63$ ,  $p<0.05$  for fine mode). The correlation between anionic surfactants as MBAS and EVAS, and DOC found in this study reaffirm the relationship between anionic surfactants and DOC indicated in previous studies e.g. Capel et al. (1990) and Facchini et al. (2000) using other methodologies such as a comparison between DOC and surface tension of cloud and wet aerosols.

These overall relationships expressed by the correlations in our data may conceal some important detail embedded there. The linearity is clearer at higher concentrations than at low concentrations, although difficult to see among the scatter of points. At low concentration the relationship may have almost zero slope. It could be that surfactant only begins to appear in the aerosol system after DOC concentrations exceed  $100 \text{ nmol m}^{-3}$ . Linear regression of the data at higher DOC concentrations suggests a slope of  $1.0 \times 10^{-3}$  for MBAS and  $4.0 \times 10^{-3}$  for EVAS aerosols. These slopes that suggest less than a one molecule of surfactant (as MBAS) is present for every one thousand dissolved organic carbon atoms and one molecule of surfactant (as EVAS) is present for every two hundred and fifty carbon atoms.

### 3.4 Nominal molecular weight of surfactants molecules

The concentration of surfactants after ultrafiltration of fine mode aerosols collected at UEA Norwich are summarised in Fig. 5. Overall, the concentration of MBAS and EVAS

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in aerosols was found higher in the smaller molecular size range (<5 K Da). Several studies have correlated the concentration of atmospheric surface tension with water soluble organic compounds (WSOC) (Bower et al., 2000; Tuckermann and Cammenga, 2004). The distribution of water soluble organic compounds in atmospheric aerosols analysed by Havers et al. (1998) and Zappoli et al. (1999) indicated that the nominal molecular weight of water soluble organic compounds is below 1000 Da and 3070 Da respectively not out of line with the results in this study.

Kiss et al. (2003) in other study estimated the nominal molecular weight of humic like substance in aerosol, which is a possible sources of surfactants in the atmospheric aerosols is around 215–345 Da. The concentration of MBAS after ultrafiltration of PM<sub>10</sub> samples collected in Malaysia (Fig. 4) showed the concentration of surfactants also has potential source in compounds with molecular weight above 5K Da. On the other hand, the distribution of EVAS concentration after ultrafiltration was still dominated by nominal molecular weight below 5 K Da. The difference of sampling area background in Europe e.g. the concentration of surfactants of aerosol collected at Great Hungarian Plain (Kiss et al., 2003) which represent rural site and the sampling areas in Malaysia which are more polluted may result in the different findings; particularly the association of surfactants with molecular weights above 5 K Da.

### 3.5 Yellow/Brown colour of aerosol extracts

The spectrophotometric measurements at 400 nm wavelength correlated with the concentration of surfactants as MBAS ( $r=0.83$ ) and EVAS ( $r=0.76$ ) as shown in Fig. 6a and Fig. 6b. This is much in line with the loose association of perceived colour and amount of dissolved organic material. The intense brown colour of aerosols extracts from samples collected in Malaysia stations also gave a significant correlation ( $p<0.05$ ) between brown colour and MBAS ( $r=0.84$ ) and also EVAS ( $r=0.73$ ). This brown colour was mentioned long ago by Mukai and Ambe (1986) and more recently by Facchini et al. (1999) and was related to a strong ability to act as macromolecular carrying polar group or surfactant in atmospheric aerosols. The relation shown here between brown

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colour and surface activity is a supportive of the importance of HULIS as a surfactant. This is especially true when we note the suggestion from Graber and Rudich (2005) of the association of HULIS with surface tension.

#### 3.5.1 Exposure to UV light

Exposure of aerosol extracts from the UEA samples shows almost all MBAS in the solution decomposed after 2 h UV exposure (below 10% left), but there was around 20% EVAS still in the solution after such exposures (Fig. 7). Repetition of UV exposure with longer exposure times (24 h) also gave the same results. Exposure of aerosols extracts collected in Malaysia to UV light also indicated a loss of almost all MBAS while about 10 to 20% EVAS still remained. In parallel the exposure to UV light caused a loss of yellow/brown colour. The small fraction of EVAS that is resistant to destruction by UV light may arise from inorganic materials, perhaps one of the interferents noted by those who use this reagent in analysis (Latif and Brimblecombe, 2004).

#### 3.5.2 Separation of “brown” aerosol extracts using ultrafiltration

The intensity of the brown colour of aerosol extracts declined after ultrafiltration with 30 K Da and 10 K Da, but it not necessarily decreased the concentration of surfactants significantly especially for EVAS as shown in Fig. 8, so not all brown HULIS is surfactant. The UV-Vis spectra of the extracts after ultrafiltration (Fig. 9) with both molecular sized filters show this reduction. The absorbance of the two ultrafiltered extracts of aerosol extracts are similar at about 0.2 for 300 nm, suggesting that both the high and low molecular weight HULIS components are coloured.

### 3.6 Characteristic of possible sources of atmospheric surfactants

Based on our early studies, Latif and Brimblecombe (2004), anionic surfactants determined as MBAS and EVAS were found in soot e.g. diesel and petrol soot and have similar characteristic to humic like substances. The amount of surfactant present on

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these possible sources tended to be enhanced by oxidation e.g. through exposure to ozone and sunlight (Latif and Brimblecombe, 2004). Here we tried to measure the molecular weight of MBAS and EVAS produced by such exposures to ozone.

The molecular weight distribution of surfactants before ozone exposure (white bars) differ among petrol soot, diesel soot and laboratory humic acid (Fig. 10). The distribution of MBAS among the various molecular weight size ranges for laboratory humic acid are found to be quite different from that in petrol and diesel soot (see Fig. 10). Much of this MBAS from humic acid was found in the range of 10 K to 30 K Da range compared to a predominance in the smaller molecular weight range (<5 K Da) for petrol and diesel soot. This results is well related to the finding of Pierson et al. (1983) which indicated that there was little high molecular weight organic fraction in automotive diesel soot compared to combustion of plants materials (Mukai and Ambe, 1986).

After exposure to ozone (hatched bars) the MBAS in petrol soot and diesel soot was detected in the molecular weight ranges below 10 K Da. It should be noticed that in Fig. 10a and Fig. 10b the concentrations have increased after ozone exposure both petrol suggesting the production of MBAS below 10 K Da. The MBAS found in laboratory humic acid was initially found in the higher molecular weight ranges (10 K to 30 K Da), but decreased noticeably on exposure to ozone (Fig. 10c).

The concentration of surfactants as EVAS in soot and humic acid before and after ozone exposure is shown in Fig. 11. These concentrations are somewhat higher than those found for MBAS, which is also typical of aerosol measurements in the atmosphere (Latif and Brimblecombe, 2004). The EVAS is almost entirely found at low molecular weights (<5 K Da) in petrol soot, diesel soot and humic acid solution. These results are quite same to the concentration of EVAS in aerosols extract which found higher in the solution with molecular weight below than 5 K. After ozone exposure the amount of EVAS in the low molecular weight fraction increased, suggestion it was produced on the aerosol, except in the case of diesel soot.

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#### 4 Conclusions

Comparisons between our work and that of earlier studies reinforce the potential links between the presence of HULIS and brown-yellow colour of aerosol extracts and their surfactant nature. As in the case of earlier work we find the presence of much surfactant material, especially in the fine mode aerosol, in the molecular size range <5 K Da, although in very polluted conditions during haze episodes derived from biomass burning in Malaysia substantial amounts were present at molecular weights above 5 K Da.

There was a relationship between sulfate and nitrate and the amount of surfactants, which hinted at combustion sources. There was also a relationship with DOC, although hints that this relationship may have been stronger when the amount of DOC was high, where one molecule of surfactant as MBAS seemed to be present for every 1000 atoms of carbon and one molecule of surfactant as EVAS for every 250 atoms of carbon.

Although our experiments were restricted to relatively high concentrations of ozone (20 ppm) there was evidence that reactions on soot and humic acid could produce surfactants. The surfactants produced were typically of molecular weight below 5 K Da. This is in accord with the observation that much HULIS in the atmosphere is present at low molecular weights. While ozone may produce surfactant compounds it might also destroy them under some conditions. Surfactants can be correlated to the brown and organic HULIS but this study also indicates it is not all brown HULIS that can be considered as surfactants.

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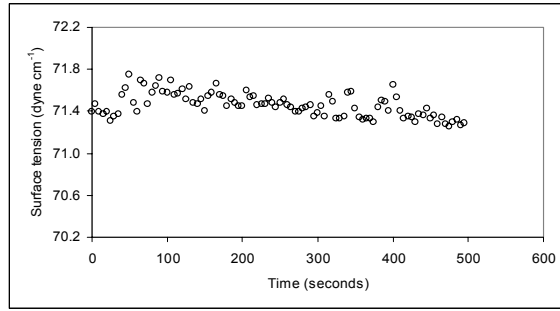
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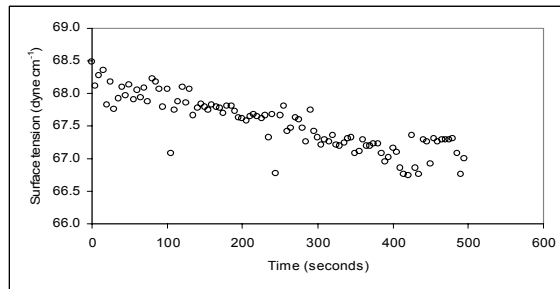
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13820

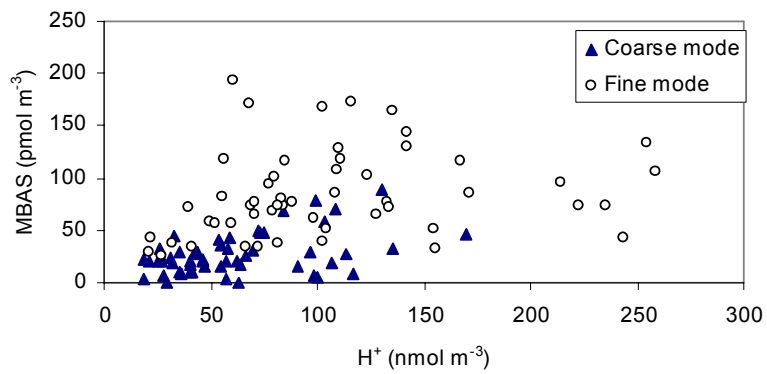


(a) Coarse mode aerosol



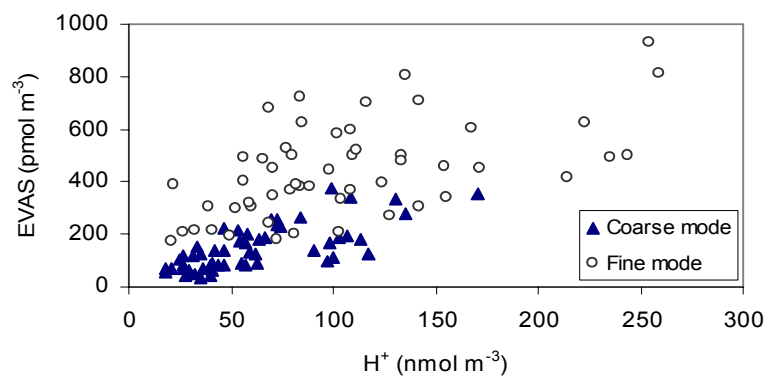
**Fig. 1.** The changing of surface tension of the pendant drop by time.

13821



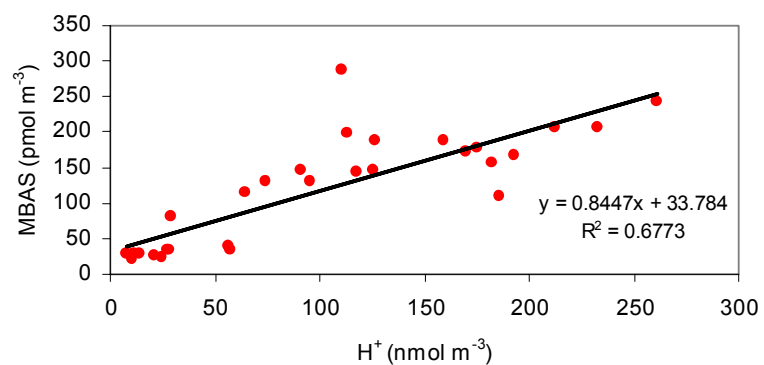
**Fig. 2a.** Concentration of MBAS compared to free  $H^+$  in aerosols collected at UEA.

13822



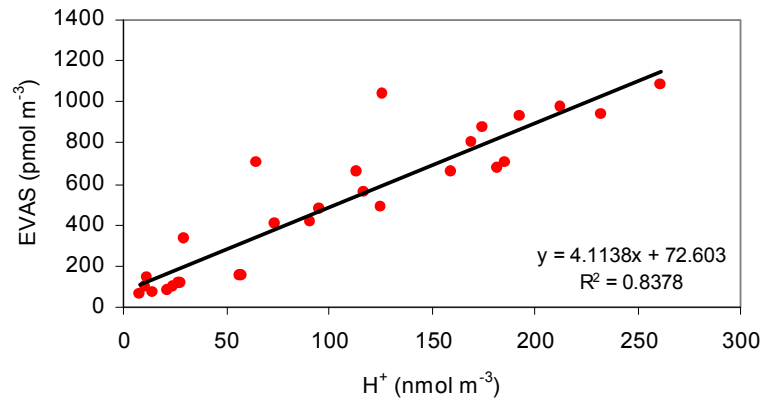
**Fig. 2b.** Concentration of EVAS compared to free H<sup>+</sup> in aerosols collected at UEA.

13823



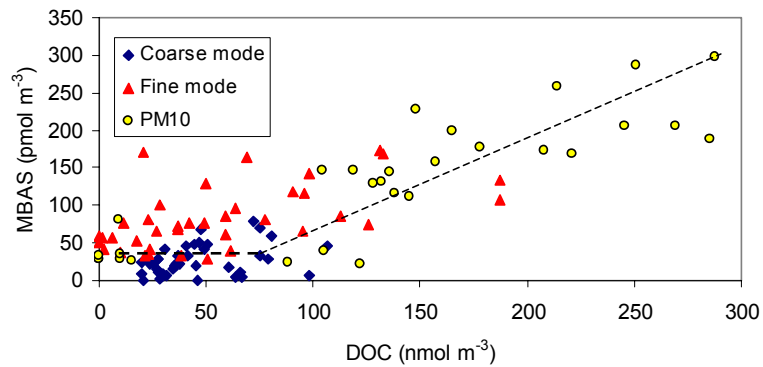
**Fig. 3a.** Concentration of MBAS compared to free H<sup>+</sup> in aerosols collected in all sampling stations in Malaysia.

13824



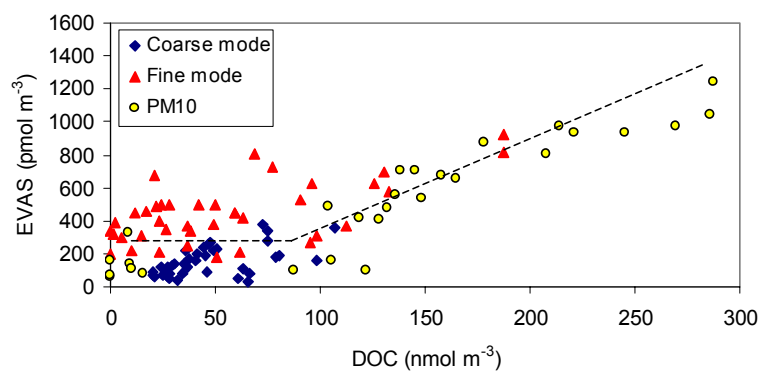
**Fig. 3b.** Concentration of EVAS compared to free H<sup>+</sup> in aerosols collected in all sampling stations in Malaysia.

13825



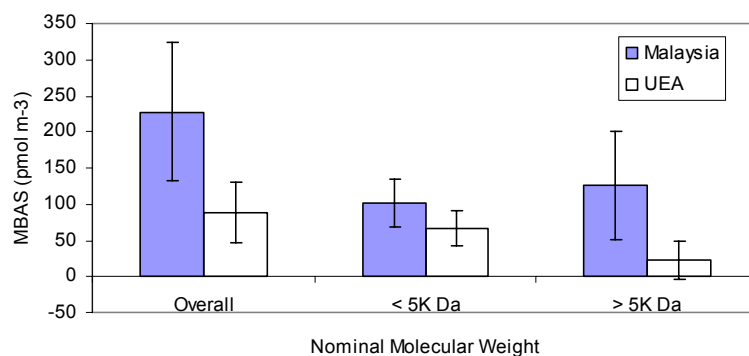
**Fig. 4a.** Concentration of MBAS compared to DOC in aerosols collected at UEA (as fine and coarse) and Malaysia (Bangi, Penang and Kota Kinabalu).

13826



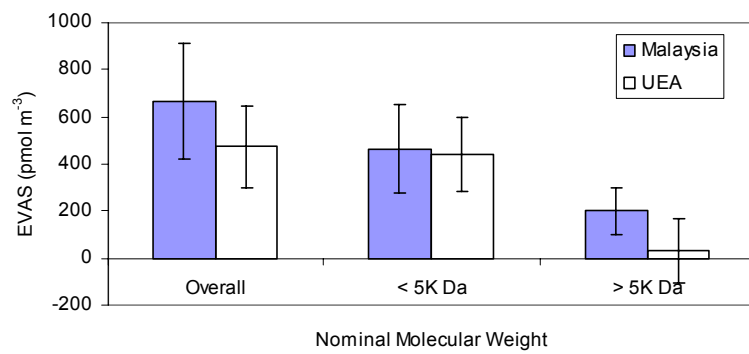
**Fig. 4b.** Concentration of EVAS compared to DOC in aerosols collected at UEA (as fine and coarse) and Malaysia (Bangi, Penang and Kota Kinabalu).

13827



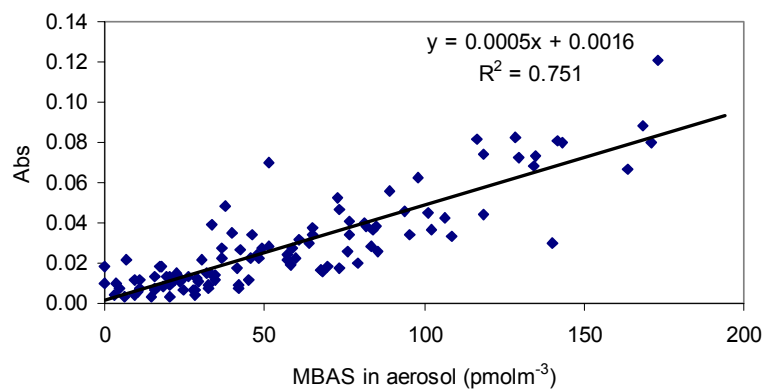
**Fig. 5a.** Distribution of MBAS in fine mode aerosols samples collected at UEA and  $\text{PM}_{10}$  in Malaysia (Bangi, Penang and Kota Kinabalu) based on nominal molecular weight.

13828



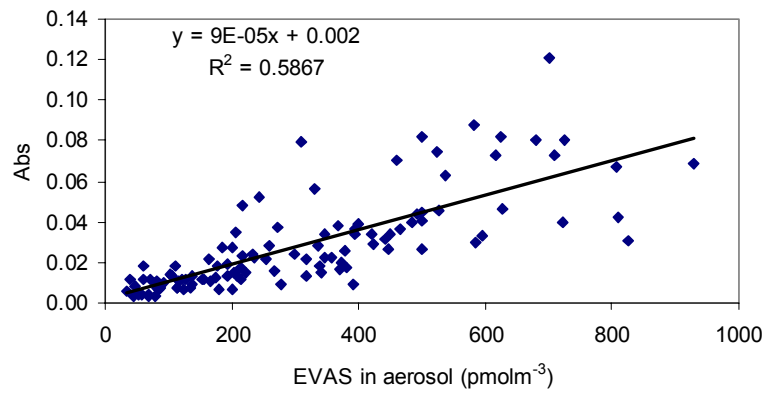
**Fig. 5b.** Distribution EVAS in fine mode aerosols samples collected at UEA and PM<sub>10</sub> in Malaysia (Bangi, Penang and Kota Kinabalu) based on nominal molecular weight.

13829



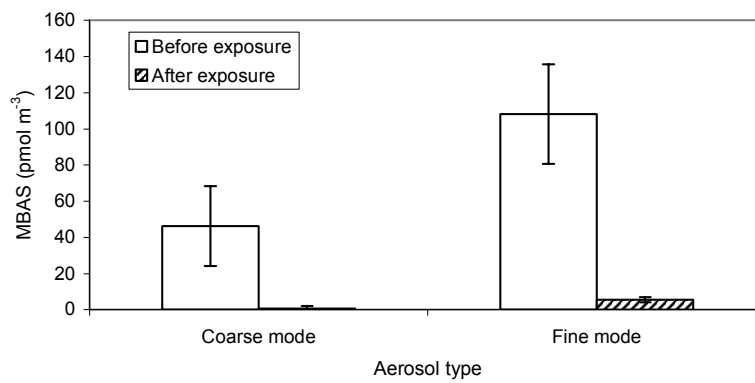
**Fig. 6a.** Absorbance of aerosol extracts (UEA, Norwich) at 400nm as a function of MBAS concentration.

13830



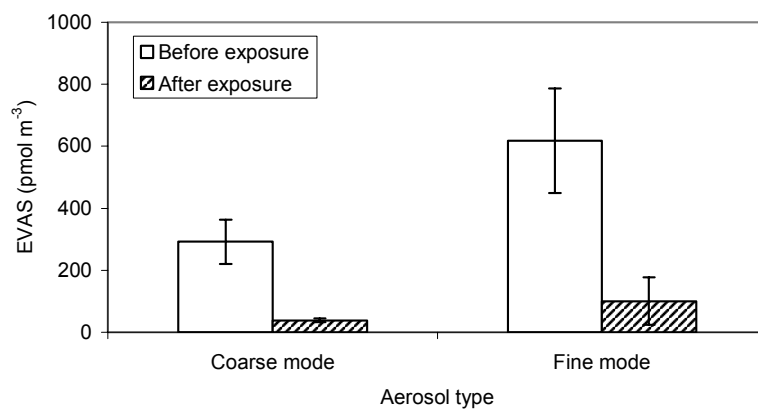
**Fig. 6b.** Absorbance of aerosol extracts (UEA, Norwich) at 400 nm as a function of EVAS concentration.

13831



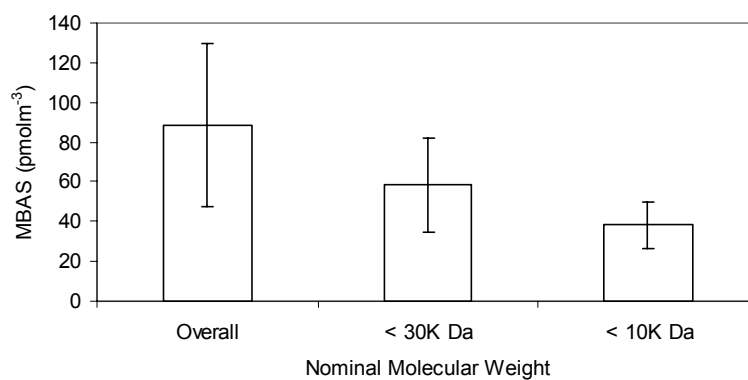
**Fig. 7a.** Concentration of MBAS in coarse and fine mode aerosols before and after UV light exposure.

13832



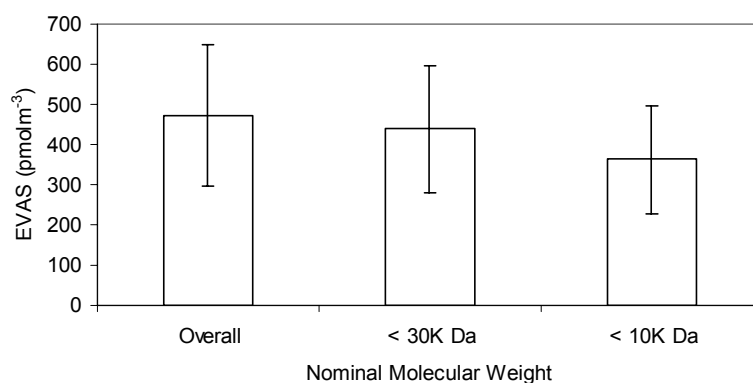
**Fig. 7b.** Concentration of EVAS in coarse and fine mode aerosols before and after UV light exposure.

13833



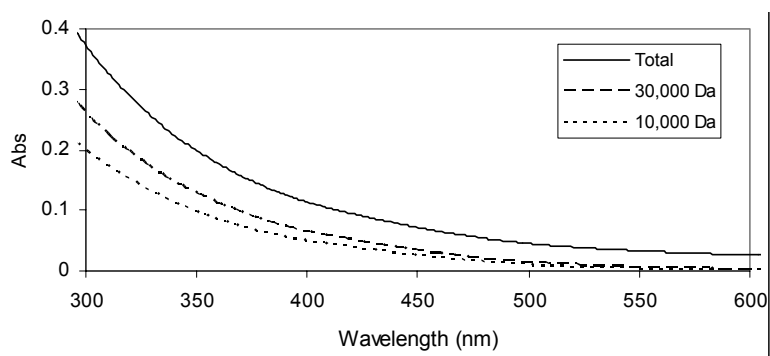
**Fig. 8a.** Separation of brown colour of aerosol extracts (Norwich station) and its MBAS surfactants concentration.

13834



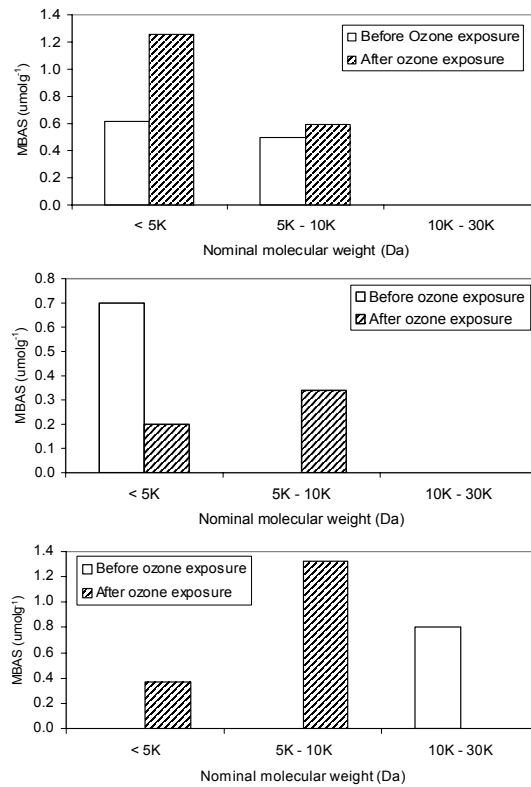
**Fig. 8b.** Separation of brown colour of aerosol extracts (Norwich station) and its EVAS surfactants concentration.

13835



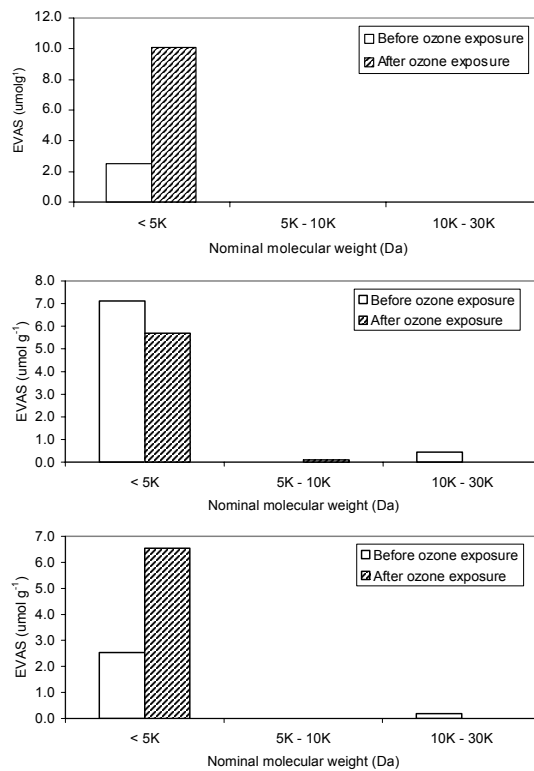
**Fig. 9.** UV-Vis spectra of overall aerosol extracts and after ultrafiltration with 30K and 10K Da.

13836



**Fig. 10.** MBAS concentration of petrol soot, diesel soot and humic acid before and after expose to ozone.

13837



**Fig. 11.** MBAS concentration of petrol soot, diesel soot and humic acid before and after expose to ozone.

13838