

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Long-range transport of Asian dust and air pollutants to Taiwan: observed evidence and model simulation

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Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Abstract

Long-range transport of Asian dust and air pollutants are major environmental concerns of Taiwan during the winter monsoon season when northeasterly winds prevail following passages of cold fronts. Based on hourly measurements of Taiwan Environmental Protection Administration (TEPA) air quality monitoring stations, Lidar and in-situ IC, a significant long-range transport dust and air pollutants event on 18 March 2005 has been identified. During this episode, drastically elevated concentrations of PM₁₀, CO and SO₂ along with the strong northeasterly on 18 March were observed over background Wanli station, with peaks of about 170 μgm⁻³, 1.0 PPM and 14 ppb, respectively. We have found that air masses of air pollutants and Asian dust are transported separately. The major component of the first PM₁₀ peak were air pollutants, evidenced by the consistent peaks of SO₄²⁻ and NO₃⁻ measured by in-situ IC, while no significant depolarization was measured by Lidar. In contrast, the evident non-spherical particles and hourly PM₁₀ concentration consistently varied with Ca²⁺ indicating that mineral dust was the major component of the second peak. Numerical results showed significant agreement of temporal and vertical variation of aerosol concentration with observations. The phenomena of split air parcels between air pollutants and Asian dust transported to Taiwan are strongly associated with the transport paths and stable and dry atmospheric boundary conditions.

1 Introduction

The topic of long-range transport of desert dust during late winter and spring in East Asia is well documented (e.g. Shaw, 1980; Duce et al., 1980; Prospero et al., 1989; Murayama et al., 2001). Many researchers have noted the effect of the high aerosol load caused by frequent Asian dust outbreak, such as reduced visibility, biogeochemical cycles, and the atmospheric radiation budget (e.g. Uematsu et al., 1983; Li et al., 1996; Chun et al., 2001; Sokolik and Toon, 1996; IPCC, 2001; Seinfeld et al., 2004).

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Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

EGU

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Except for the dust storms, anthropogenic emissions over East Asia also received much attention recently. Along with economic growth, East Asia has been a region of rapidly increasing emissions of air pollutants in the past decades. In the winter and spring, Taiwan and vicinity is often under the influence of northeasterly winter monsoon winds originating in central Asia. The winter monsoon not only brings cold air, but can also transport air pollutants and dust over a long distance to Taiwan and even to the northwestern Pacific area (Arimoto et al., 1996; Uno et al., 2001; Uematsu et al., 2003, Takemural et al., 2002). This is evident from satellite observations such as AVHRR, MODIS, and SEAWIFS (Husar et al., 2001; Nakajima, et al., 2003), that show a pronounced enhancement of aerosol optical depth during winter and spring over the western North Pacific as a result of Asian continental outflow. Takemura et al. (2002) pointed out that 50% of the total optical thickness around Japan is due to anthropogenic aerosols during the strong Asian dust events. Based on the qualitative analysis from background stations of Taiwan Environmental Protection Administration (TEPA), Lin et al. (2004) presumed the contribution of long-range transport to PM₁₀ abundance in northern Taiwan to range from 50% to 70% during the northeasterly monsoon. Recently, Lin et al. (2005) estimated that the long-range transport of particulate pollutants in northern and eastern Taiwan contributes to PM₁₀, CO and SO₂ concentrations of about 30 μg/m³, 230 ppb and 0.5 ppb, respectively.

Prospero et al. (2003), in a study of aerosols distributions at Midway Island during 1981–2000, found that pollutants such as sulfate and nitrate ions in the aerosols increased concurrently with mineral dust in spring and winter. Over a long period, the anthropogenic pollutant increases along with mineral dust in the spring. Uno et al. (2001) illustrated a long-range transported case by numerical model and showed that mineral dust and anthropogenic tracer are simultaneously transported to the central Pacific Ocean. However, in our study, the hourly data analysis taken from TEPA air quality monitoring stations showed that most of the dust storm cases usually do not pick up significant air pollutants, or randomly carried pollutants during long-range transport of dust storm episode (Lin et al., 2004). Also, Uematsu et al. (2002) pointed out that

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Asian dusts are occasionally capable of transporting mineral and anthropogenic particles to the western North Pacific. Matsumoto et al. (2003) found out anthropogenic fine particles containing abundant NSS-SO_4^{2-} appeared first and were then followed by large mineral particles that had absorbed NO_3^- . This phenomena is different from that proposed by Prospero et al. (2003) and Uno et al. (2001). The consistent feature found by Prospero et al. (2003) were based on seasonal variations, while the lack of correlation observed at TEPA stations were at time scale of about an hour or less (Lin et al., 2004). Furthermore, Taiwan and Japan lie just off the major emission sources region, so it might take some time to admix the different air masses including various types of aerosols (Uematsu et al., 2002). It is presumed that whether air pollutants are consistently carried by dust storms or not is strongly associated with the paths of the air mass transport and boundary structures. Therefore, the processes of long-range transport are complex and need further discussion.

Matsumoto et al. (2003) showed that the transport patterns of the polluted air masses from the east Asian continent to the northern regions of the northwestern Pacific are often different from those to southern regions. Asian dust transport mainly between 30~40 N (Zhao et al., 2003; Satake et al., 2004) while sulfate and Carbon is between 25~45 N (Satake et al., 2004). More importantly, this is the first time observed and discussed the separately transport phenomenon between Asian dust and air pollutants over low latitude (less than 25 N) location such as Taiwan. This report completes the picture in the region.

In this paper, we investigate an evidently inconsistent transport case of Asian dust and air pollutants air masses to Taiwan. In this study, measurement of TEPA ground stations, In-situ IC and Lidar are used to analyze the episode. The Nested Air Quality Prediction Model System (NAQPMS) is employed to study the process of the long-range transport of dust and air pollutants to Taiwan.

2 Data source and instruments

Measurements of hourly PM_{10} and other pollutants were taken from TEPA air quality monitoring stations. Meteorological parameters, i.e. surface temperature, wind direction and speed, rainfall, and atmospheric concentrations of PM_{10} , NO_x , CO, and SO_2 observed at a coastal background station and a mountain station in northern Taiwan were used in this study. The coastal background station Wanli (Fig. 1) is upwind from the frontal passage and thus ideal for observing the effect of frontal passages on distribution of dust and air pollutants. Measurements at Yangming mountain station (altitude 826 m) at a national park in northern Taiwan turn out to be very valuable also because it is relatively clean and provides information on vertical distribution of trace species.

The aerosol composition was analyzed by an in-situ IC system (Chang et al., 2006), which provides concentrations of water-soluble ions Cl^- , NO_2^- , NO_3^- , SO_4^{2-} , Na^+ , NH_4^+ , K^+ , and Ca^{2+} . The location of the sampling site (25 N, 121.32 E, Fig. 1) is at the weather observatory of the National Taiwan University (NTU), which is located in the southwestern part of the Taipei Basin. The in-situ IC system consisted of gas removal denuders, aerosol collecting device, and ion chromatograph (Model ICS-90, Dionex Corp., Sunnyvale, CA, YSA), and was attached to the bypass flow line of the TEOM monitor to measure the water-soluble inorganic ions $PM_{2.5}$ during spring of 2005 throughout at 15-min. intervals (PM_{10} is not available in this study). At the same location, a Lidar system (hereafter referred to as RCEC/ASNTU Lidar) was set up from 2004. RCEC/ASNTU Lidar is a dual-wavelength Raman and Depolarization Lidar system (manufactured by Zenon SA, Greece). The lidar system employs the second and third harmonics of Nd-YAG laser at 532 nm and 355 nm. This system is operated on a 24 h basis to probe the atmosphere at elevations ranging between 0.3 km and 8 km. More details are shown in Table 1. Meanwhile, Lidars performed by National Institute for Environmental Studies (NIES) located in China and Japan were also employed to compare the model results.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

3 Model description

The NAQPMS is a fully modularized three-dimensional system with various options for representing the physical and chemical processes describing regional and urban atmosphere pollution (Wang et al., 2000; Wang et al., 2002). It consists of two parts, a regional chemical transport model and the Fifth-Generation PSU/NCAR Mesoscale Modeling System, MM5 (Grell et al., 1995), which is used to generate meteorological fields. The initial and boundary conditions for MM5 are obtained from NCAR/NCEP 2.5°×2.5° reanalysis data sets at 6-h intervals. The model can employ flexible horizontal grid resolution with multiple level nested grids for regional applications and urban scale simulations. In this study, 91×73 grid boxes were employed in the east-west and south-north directions with 81 km resolution. Vertically, the model divides the troposphere from surface to 100 hPa into 20 layers using variable-spaced sigma coordinates and the thickness of the surface layer is about 70 m.

A modified dust deflation module is embedded in NAQPMS. The detail methodology for dust emission was described by Wang et al. (2002). In this application, soil dust emission intensity is modified with the vegetation fraction percentages and soil categories based on the approach of Wang et al. (2000):

$$Q_j = C_1 \times u^{*2} \times (1 - u_0^*/u^*) \times W \times R_j \times \sum C_j f_j, \quad (1)$$

where u^* and u_0^* are the friction and threshold friction velocities and C_1 is a constant (2.9×10^{-11}); C_j is the weighting factor of dust loading for j type of vegetation that is obtained from the U.S. Geological Survey (USGS) data with a horizontal resolution of 10 min; f_j is the vegetation fraction percentage of j type of vegetation in a model grid that is derived from the AVHRR data.

The dust particle size is divided into 10 size bins from 0.43 to 42 μm in diameter in terms of the eight stages of aerosol sampler, with three additional size bins above 11 μm for larger particles. Table 2 lists the particle size ranges. The size distribution of dust emission is an important parameter for dust simulation. In this study, the total

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

dust emission flux is apportioned to each size bin based on the observation of the size distribution of soil dust in Chinese source regions within and on the edge of deserts (Zhang et al., 2003, Gao et al., 2002).

The emissions of air pollutants in Asia were obtained from the East Asia inventory by Streets et al. (2003). Meanwhile, in order to evaluate the amount of long-range transport, local emissions are excluded in this study. The model domain in this study is shown in Fig. 8. This model has been successfully used to study the atmospheric trace gases and particles, such as SO_x , dust, O_3 , and acid rain over East Asia (Wang et al., 2000, 2002; Uematsu et al., 2003; Lin et al., 2006¹).

4 Results

4.1 Observation analysis

Figures 2a and b show the time series of the hourly meteorological parameters, PM_{10} and other pollutants from air quality monitoring stations Wanli and Yangming. The data indicates that the cold front arrived around midnight of 17 March 2005 and that the temperature reached the minimum when the wind speed reached the maximum on 18 March 2005. Although wind speed and temperature significantly varied, no rainfall was monitored at Wanli in this case. Before the frontal passage, local pollution dominated on 16 and 17 March and showed clear diurnal variation in air pollutants. During local pollution days, daytime anthropogenic emissions at rush hour lead to the concurrent enhancement of air pollutants such as NO_x , CO, and SO_2 . At night, concentrations of all pollutants drop gradually to their minima. After the cold front passed over northern Taiwan on 18 March, the surface temperature dropped to a minimum, wind speed increased to a maximum and the wind changed to northeasterly. The strong winds swept away most of the local pollutants while NO_x stayed at extremely low values and

¹Lin, C.-Y., Wang, Z., Chou, C. C.-K., Chang, C.-C., and Liu, S.-C.: A numerical study of an autumn high ozone episode over southwestern Taiwan, *Atmos. Environ.*, in revision, 2006.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

showed no diurnal cycle. In this situation, local pollutants become less evident or even disappear all together. Figure 2c shows the ratio of short lived (NO_x) to long lived specie (CO). The values of the ratio (i.e. NO_x/CO) remained low at both Wanli and Yangming stations stably on 18 October and is significantly different from that identified local pollution days. It is important to note that the contributions of local emissions are quite minor on the episode day. Therefore, most of the enhancement of air pollutants can be attributed to the long-range transport.

Figure 2a shows that northeasterly winds with maximum speed of about 12 m/s brought an air mass from the Asian continent with PM_{10} , CO, and SO_2 concentrations that coincidentally peaked at 07:00 LST (i.e. the 55th hour of Figs. 2a and b) on 18 March 2005 at windward Wanli station at about $170 \mu\text{g}/\text{m}^3$, 1.0 PPM, and 14 ppb, respectively. Under the strong northeasterly wind (mean wind speed was 7.86 m/s on 18 March), the average concentrations at Wanli station on 18 March for PM_{10} , CO, and SO_2 were $103 \mu\text{g}/\text{m}^3$, 390 ppb, and 5.3 ppb, respectively. By using two years' dust storm episodes, Lin et al. (2005) estimated that the average inflow concentration of PM_{10} , CO, and SO_2 of dust events were $71 \pm 34 \mu\text{g}/\text{m}^3$, 359 ± 88 ppb, and 1.1 ± 0.5 ppb, respectively. It is apparent that more than 3 times of mean SO_2 concentration arrived via long distance transported over Taiwan for this case. The concentration of PM_{10} , CO, and SO_2 were well correlated at the north tip of coastal station Wanli before 10:00 LST on 18 March. Although there are a few hours' PM_{10} data missing for Yangming station (Fig. 2b), its variation trend is similar to Wanli. Apparently, the consistent hourly variations of PM_{10} and air pollutants in the morning of 18 March are characteristic of the long-range transport event with a significant amount of air pollutants from the Asian continent as propounded by Lin et al. (2004). The concentration of SO_2 (CO) dropped drastically from 14 ppb (1.0 ppm) at 08:00 LST to 3 ppb (0.3 ppm) after 16:00 LST (i.e. the 64th hour of Figs. 2a and b). However, PM_{10} concentration did not consistently drop with air pollutants but peaked at 11:00 LST on 18 March. In other words, there was a second peak at around noon as a result of the variation of hourly PM_{10} at Wanli station.

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

In order to identify the sources and to examine how transport paths affect the concentrations of air pollutants in northern Taiwan, a three-day backward trajectory analysis was performed for this case. The analysis was computed using the HYSPLIT (Hybrid Single-Particle Lagrangian -Integrated Trajectory) model (Draxler and Hess, 1998). The backward trajectory analysis was performed for altitudes of 100, 500, and 1000 m at Wanli station in northern Taiwan. Figure 3 shows the 72-h backward trajectory starting from 00:00 UTC (08:00 LST) on 18 March 2005. In this case, the air parcels at 100 m and 500 m mostly move over industrial areas of Mainland China before traveling to Taiwan. The atmospheric condition was considerably stable and was evident that the high pressure center measured nearly 1052 hpa around Mongolia (Fig. 4a) on 16 March 2005. The dry atmospheric conditions and strong pressure gradient led to the occurrence of dust storms over Inner-Mongolia (Fig. 4a). Following the anticyclone outflow, Asian dust was transported to the downstream regions such as Korea, Japan and Taiwan as shown in Fig. 4b. Tracing the air masses trajectory, the sounding of Shanghai (Fig. 5a) at 00:00 UTC on 17 March 2005 shows that the inversion layer was nearly 800 hpa (about 2 km in altitude), caused by the strong subsidence of the cold high. Following the strong prevailing northeasterly flow, the sounding of Taipei (Fig. 5b) also displayed the same inversion height on 18 March 2005, i.e. the subsidence inversion layer at nearly 2 km. During three days of transport, the air parcels of 100 and 500 m are mostly at relatively low altitude (below 1000 m). In other words, air masses are mostly transported in the low boundary during the transport periods. Therefore, the probability for these air parcels of picking up air pollutants over industrial areas should be relatively high.

Figure 6a shows an aerosol depolarization episode monitored by RCEC/ASNTU Lidar on 18 March 2005, during which the significant aerosol depolarization occurred from 09:00 LST to 18:00 LST. The high depolarization occurred below altitude of 1.5 km, ranged from 5% to 10% and lasted for about 8 h. This indicates that the non-spherical particles mainly exist after 09:00 LST on 18 March. It is important to note that mineral dust presumably reached northern Taiwan after 09:00 LST. Furthermore,

the major composition of the secondary peak as a result of the hourly variation of PM₁₀ concentration (Fig. 2a) detected at Wanli station around noon was mineral dust. To the contrary, the significant attenuated backscattering coefficient (Fig. 6b) for 532 nm between 06:00~09:00 LST was about 2~3 times that of the high depolarization period on 18 March. Vertical distribution of enhanced attenuated backscattering coefficient was mainly below 1500 m. It was earlier deduced from hourly data that air pollutants was the major component of the first peak of hourly PM₁₀ concentration at 08:00 LST at Wanli station.

In order to identify the aerosol compositions presumed above, the time series of water-soluble ions Ca²⁺, SO₄²⁻, and NO₃⁻ in PM_{2.5} (PM₁₀ not available in this study) that were monitored by In-situ IC in northern Taiwan are shown in Fig. 7. The water-soluble ion Ca²⁺ is generally considered a good indicator of mineral dust while ions SO₄²⁻ and NO₃⁻ are mainly from anthropogenic sources. As with the earlier analysis on 17 March local pollution dominated before frontal passage, ions SO₄²⁻ and NO₃⁻ peaked during daytime but no significant Ca²⁺ concentration. After frontal passage, the concentration of SO₄²⁻ peaked at 07:00 LST which is about 2–3 h prior to the peak of Ca²⁺ concentration on 18 March. Apparently, air pollutants are the major contributors of PM₁₀ before 08:00 LST, which is consistent with hourly monitored results for ground station (Wanli) and mountain station (Yangming). Concurrently, results of Lidar and In-situ IC show that the transport of air pollutants is prior to the dust aerosol after the frontal passage on 18 March 2005.

The same sequence observed from composition analysis, Lidar and ground stations suggest that northerly flow not only carried dust, but also anthropogenic pollutants to Taiwan for this episode. Air pollutants and dust are not always transported together to Taiwan, for example this case. Apparently, the transport of dust or air pollutants is strongly associated with the boundary structure and transport paths. Unfortunately, currently available data are based on station only without any indications of transport processes. Therefore numerical study obviously plays an irreplaceable role for understanding the transport mechanism of dust or air pollutants to Taiwan.

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

4.2 Simulation results and discussion

Figure 8a shows the simulated spatial distribution of surface dust concentration at 00:00 UTC on 18 March 2005. Concentrations of dust between $100\sim 150\mu\text{g}/\text{m}^3$ are located from south Japan to northern Taiwan. Surface weather report (Fig. 4b) shows that on 18 March 2005 dust was observed from southern Japan to the islands Amami and Okinawa which are located northeast of Taiwan. Tracing back the trajectory of the dust paths, the spatial distribution of a remarkable simulation of dust concentration (Fig. 8b) are in agreement with the surface weather report (Fig. 4a) in the dust source regions around Inner-Mongolia at 00:00 UTC on 16 March 2005. Simulation results of temporal and spatial distribution of Asian dust demonstrated that the air pocket carrying dust via long-range transport was dominated by a high pressure system following the cold front and formed a band-shaped transport to Taiwan (Chen et al., 2004).

Figure 9 shows the time series of vertical profile of the model predicted over northern Taiwan. The long-range transport of dust concentration is mainly below 1.5 km and peaked nearly $150\sim 180\mu\text{g}/\text{m}^3$ at altitude of about 500~1000 m in northern Taiwan from 00:00 UTC (08:00 LST) to 04:00 UTC (12:00 LST) on 18 March. For SO_2 , it peaked about 6~8 ppb along with the continental outflow in the morning and prior to the peak of dust concentration. Figure 10 shows the time series of simulation of dust concentration, sulfate and HNO_3 over northern Taiwan. It is notable that the same peaks of sulfate and HNO_3 pattern were prior to the dust concentration as the observation results on 18 March (Fig. 7). No simulated sulfate and HNO_3 shown by the model on 17 March due to the exclusion of the local emissions. These simulation results agree well with the finding of the high resolution of In-situ IC, i.e. Asian dust and air pollutants are not coincidentally transported by the same air masses. Compared to Lidar results (Figs. 6a and b) over northern Taiwan, simulation dust and SO_2 distributions both are consistent with the characterization of depolarization ratio and attenuated backscattering coefficient in the vertical distribution. However, compared to the background station's monitoring results in northern Taiwan (Fig. 2a), numerical results of air pollutants

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

seem underestimate. It is due to Taiwan's emissions are excluded in this numerical study. In addition, emissions inventory might also one of the important reasons as for the emissions in this study are based on Street et al. (2003), i.e. the emissions are possible underestimating in this study.

5 The substantial consistent between observation and model results are also shown at other Lidar sites over Beijing (116.3E, 39.9N), and Miyakojima (125.3E, 24.7N), (Fig. 1). Beijing is located upstream of Taiwan under the northeasterly monsoon while Miyakojima is located on an island off north-eastern Taiwan. Figure 11 shows the time series of Lidar observation at Miyakojima. The significant depolarization ratio is below 1.5 km on 18 March at Miyakojima has a similar temporal and vertical distribution but stronger than the result of RCEC/ASNTU Lidar (Fig. 6) as for Miyakojima is located on the upstream of Taiwan. The time series of simulation long-range transport dust over Miyakojima (not shown) also resemble to northern Taiwan (Fig. 9) due to their close location. Our model also simulated (Fig. 13a) the enhancements of the depolarization ratio (Fig. 12a) below 3 km between 00:00 UTC and 12:00 UTC over Beijing on 16 March. Meanwhile, the vertical and temporal variation of attenuated backscattering coefficient was also consistent to our simulation of high SO₂ concentration at Beijing (Figs. 12b and 13b). It is apparent that our model results provide important messages for this issue. Figures 14a–d shows the vertical cross section along the path of the altitude of the 1 km trajectory that is displayed in Fig. 3. Taiwan is located at position of “g” in Figs. 14a–d. It is important to note that not only the significant concentrations of dust aerosols, but also air pollutants are transported below the altitude of 1.5 km before reaching Taiwan (Figs. 14a–d). As mentioned earlier, the atmospheric condition is stable and dry caused by a strong subsidence of the cold high for this episode. A strong subsidence such as this can inhibit the air masses by admixing those transported in the lower boundary. When the air masses passed over East China sea, the boundary layer subsequently developed (Figs. 14b, c and d). This is probably a result of a boundary layer over the East China Sea upwind of Taiwan, which is known for turbulence mixing because of the cold winter monsoon winds over the relatively warm Kuroshio Current.

Long-range transport of Asian dust and air pollutants to TaiwanC.-Y. Lin et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[⏪](#)[⏩](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

Revisiting the trajectory path (Fig. 3) at 00:00 UTC on 17 March, nearly all air parcels are located over costal industrial areas. The dust aerosols have nearly reached the mainland costal area while a high SO₂ concentration is between the mainland coast and Taiwan (Fig. 14a). In the other words, air pollutants transported to Taiwan are prior to dust air masses before they admix properly. A significant concentration of SO₂ is continually emitted over the industrial areas of China. Air pollutants are pushed by the dust air masses in the stable atmospheric conditions of the boundary layer as they pass by the major pollutant emission areas of China (Figs. 14a and b). Simulation results also demonstrated that the long-range transport of air pollutants to Taiwan is prior to dust for this case.

Since major paths came from the continent and no evident rainfall occurred at stations Yangming and Wanli (Figs. 2a and b), the effect of the wet deposition might have weakened before reaching northern Taiwan on 18 March. Furthermore, vertical mixing between dust aerosol and major coastal air pollutant masses are indistinctive due to stable weather conditions and air masses mostly traveled over continental. Therefore, the high SO₂ concentration (peaked nearly 14 ppb) via long-range transport can be observed over Taiwan after wet deposition of the front and the turbulence mixing of cold winter monsoon winds over the relatively warm Kuroshio current. The above processes might have been the major reasons that significant concentrations of air pollutants could be transported for long distances prior to dust transport to Taiwan for this episode.

The observed different composition could be significantly influenced by the transport paths and different vertical mixing processes (Zhou et al., 1990; Matsumoto et al., 2003; Arimoto et al., 2004; Lin et al., 2004). This paper provides the fundamental information to understand what controls the composition of Asian outflow to Taiwan, especially in the low latitude region. Except for this case that shows disagreement in transport between mineral dust and air pollutants, we noted that dust aerosol (mineral particles) arriving over northern Taiwan together with air pollutants also frequently occurred. Therefore, more cases need to be studied to make precise deductions re-

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

garding the vertical mixing and transport processes of various types of aerosols to Taiwan.

5 Summary

During late winter and spring, high mineral dust concentrations have frequently been observed over East Asia. Along with the long-range transport of Asian mineral dust, anthropogenic air pollutants are occasionally carried by the prevailing winter monsoon. Data from a background coastal station and mountain station by TEPA showed an episode in which a frontal passage was accompanied by a substantial mass of air pollutants on 18 March 2005. Drastically elevated concentrations of PM₁₀, CO and SO₂ along with the strong northeasterly were observed over background Wanli station, with peaks of about 170 μgm^{-3} , 1.0 PPM and 14 ppb, respectively. Monitoring by RCEC/ASNTU Lidar showed that significant aerosol depolarization ratio occurred below 1.5 km between 09:00 LST and 18:00 LST on 18 March 2005. Meanwhile, the significant attenuated backscattering coefficient for 532 nm was mainly below 500 m and about 2~3 h earlier than significant depolarization. Continuous measurement of hourly soluble ions demonstrated that air pollutants were the major contributors of PM₁₀ before 0800LST while Ca²⁺ peaked nearly two hours later. Results from TEPA ground stations, RCEC/ASNTU Lidar, and continually In-situ IC consistently showed that the transport of air pollutants was prior to the dust aerosol after the frontal passage for this case.

Results from NAQPMS air quality model suggests that the predicted time series of Asian dust distribution agree excellent with the observation of Lidars that were measured in northern Taiwan, Beijing and Miyakojima. Overall concentration of air pollutants is underestimated by simulation compared to that of ground stations over northern Taiwan. This is probably due to the emissions input in this study are based on the early inventory (Street et al., 2003). In addition, local emissions were excluded in this study; it might also be one of the reasons for this concern. Under the stable and dry atmo-

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

spheric conditions, air pollutants transported to Taiwan are prior to dust also predicted by the model. As dust and air pollutants are from different sources and are transported in dry and stable weather conditions to Taiwan the phenomena of split air parcels for air pollutants and Asian dust transported to Taiwan are strongly associated with the transport paths, and stable and dry atmospheric boundary conditions.

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Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

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Long-range transport of Asian dust and air pollutants to TaiwanC.-Y. Lin et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

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Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

Table 1. RCEC/ASNTU Lidar Characteristics.

Laser	Nd:YAG (Big-Sky CFR-400)
Wavelength	532/355 nm
Pulse energy	65/60 mJ
Repetition rate	20 Hz
Transient Recorder	12 bits A/D converter at 20 MHz and 250 MHz photon counting (Licel TR20-40)
Height Resolution	7.5 m
Telescope	diameter 40 cm, focal length 160 cm
Channels	532 nm(depolization), 355 nm, and 387 nm (nighttime only)

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

Table 2. Particle size ranges and mass size distributions of vertical dust flux for three major desert regions of China used in this study.

Bin	Size Range in Diameter, μm	Middiameter μm	MSD %
1	0.43–0.65	0.53	0.03
2	0.65–1.1	0.84	0.04
3	1.1–2.1	1.52	0.07
4	2.1–3.3	2.63	0.1
5	3.3–4.7	3.95	0.12
6	4.7–7.0	5.76	0.15
7	7.0–11.0	8.85	0.18
8	11.0–17.5	14	0.11
9	17.5–27	21.5	0.09
10	27–42	34	0.11

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

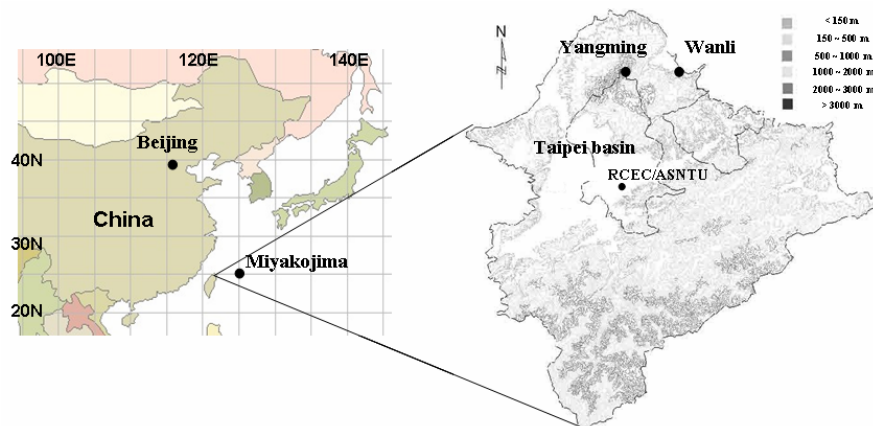


Fig. 1. Locations of Lidar sites for RCEC/ASNTU, Beijing and Miyakojima and the TEPA background stations Yangming and Wanli.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

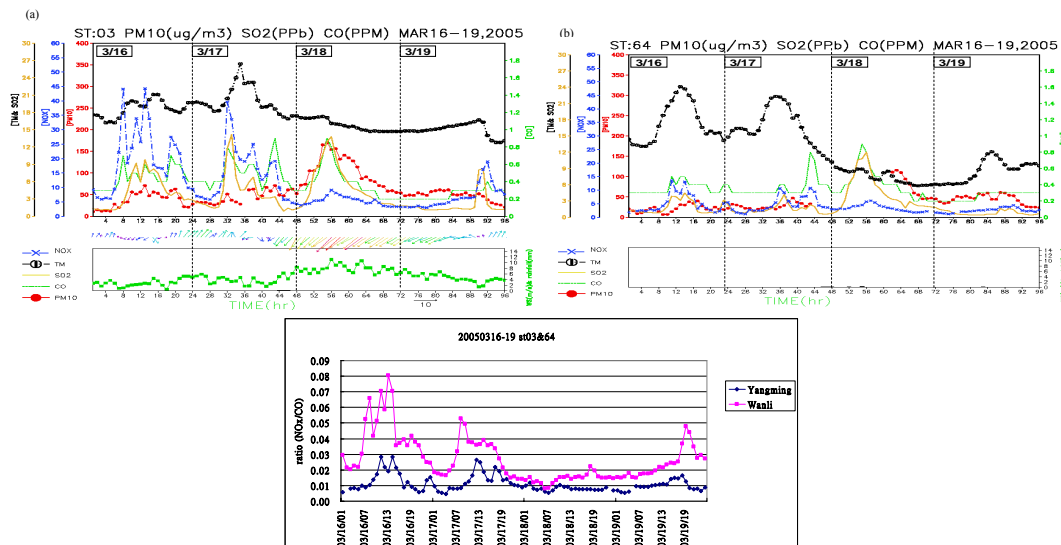


Fig. 2. The time series of hourly PM₁₀ (red closed circles), CO (dashed line), SO₂ (solid), NO_x + O₃ (triangle), NO_x (cross), surface temperature (open circle with vertical line), wind speed (closed square, bottom panel) and rainfall (black bar, bottom panel) are shown at **(a)** Wanli station **(b)** Yangming station (from 01:00 LST 16 March to 24:00 LST 19 March 2005). Surface wind at Wanli station is represented by wind vector while wind field is not available at Yangming station. Data missing is between 45th and 60th for PM₁₀ at Yangming station. **(c)** The time series of hourly ratio of (NO_x/CO) at stations Wanli and Yangming from 01:00 LST March to 24:00 LST 19 March 2005.

[Title Page](#)
[Abstract](#)
[Introduction](#)
[Conclusions](#)
[References](#)
[Tables](#)
[Figures](#)
[◀](#)
[▶](#)
[◀](#)
[▶](#)
[Back](#)
[Close](#)
[Full Screen / Esc](#)
[Printer-friendly Version](#)
[Interactive Discussion](#)

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

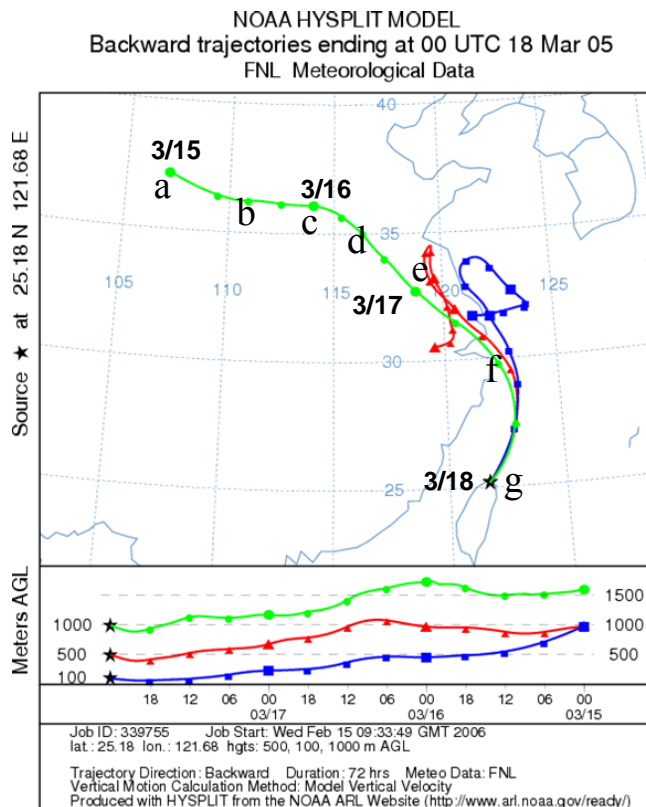


Fig. 3. Result of the HYSPLIT model 3-day backward trajectory analysis started at 00:00 UTC (08:00 LST), 18 March 2001 at altitudes of 100, 500 and 1000 m at Wan-Li station in northern Taiwan. The top and bottom panels display horizontal and vertical motion. Symbols denote the location of the air parcel every 6 h. The letters a to g are represented the geographic locations in Fig. 14 in x-axis.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

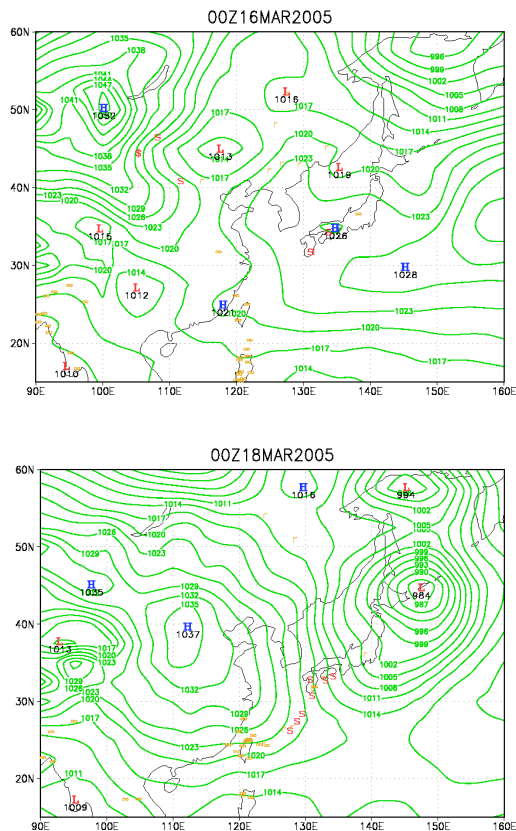


Fig. 4. Surface weather chart and dust storm report (denoted by dollar signs) at 00:00 UTC on **(a)** 16 March 2005 **(b)** 18 March 2005. Symbols of H and L show the location of major high and low pressure systems.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

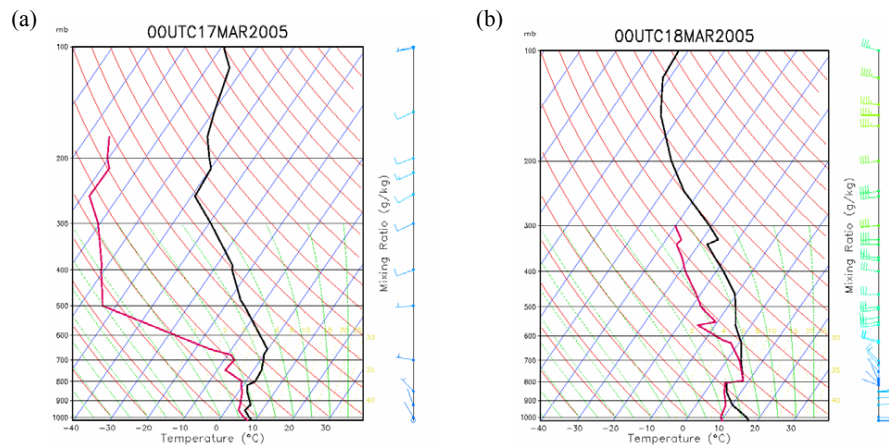


Fig. 5. (a) Sounding from Shanghai (station 58362) at 00:00 UTC 17 March 2005 (b) Sounding from Taipei (station 46692) at 00:00 UTC 18 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

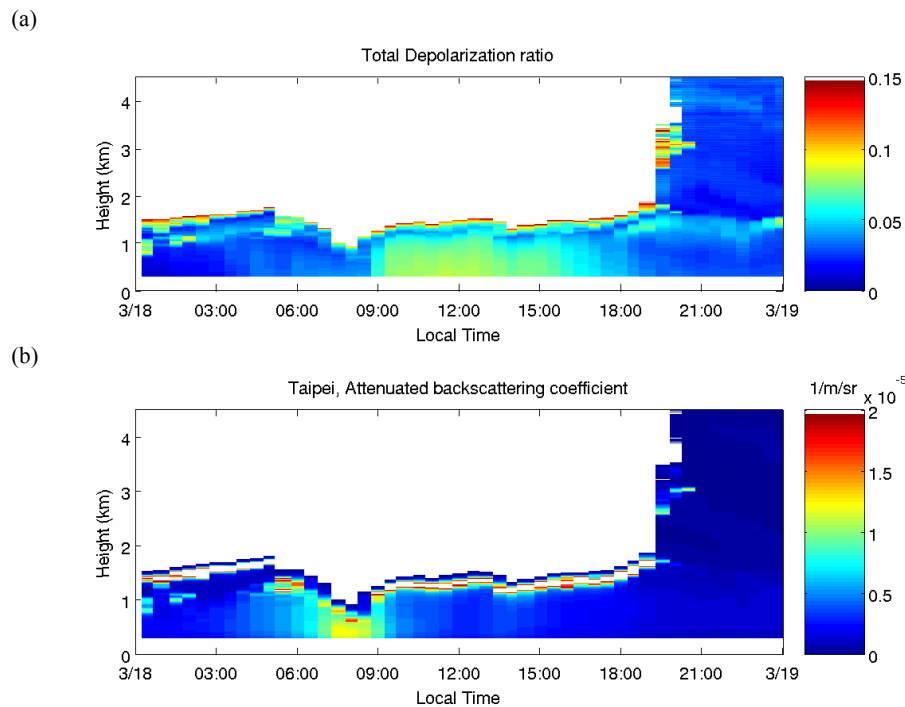


Fig. 6. Time series of Lidar observed in northern Taiwan on 18 March 2005 (a) depolarization ratio (b) attenuated backscattering coefficient.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

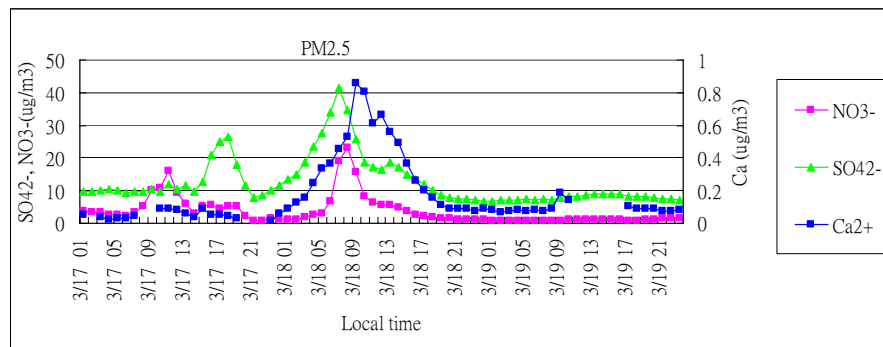


Fig. 7. Time series of hourly soluble ions for Ca^{2+} , SO_4^{2-} and NO_3^- in $\text{PM}_{2.5}$ during 17 March 2005~19 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

Long-range transport
of Asian dust and air
pollutants to Taiwan

C.-Y. Lin et al.

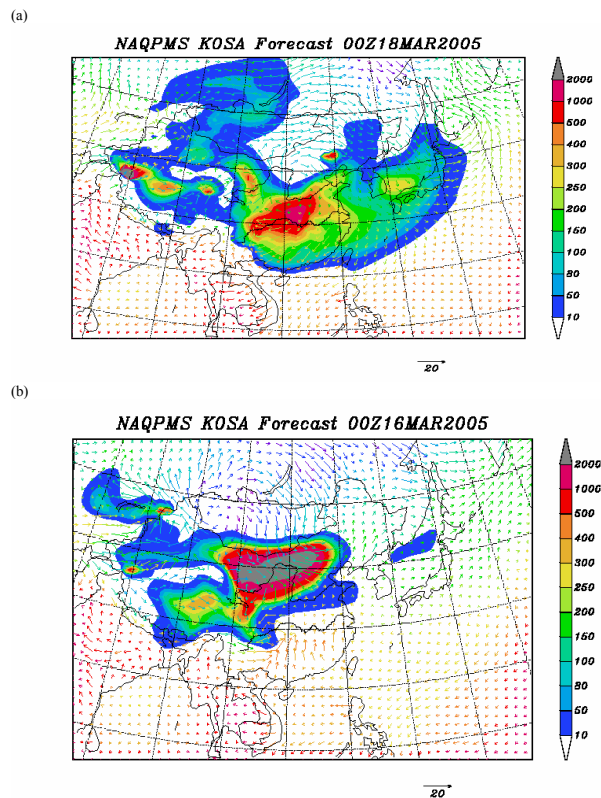


Fig. 8. Simulated distribution of ground level dust concentration and wind field at **(a)** 00:00 UTC (08:00 LST) 18 March 2005, **(b)** 00:00 UTC (08:00 LST) 16 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

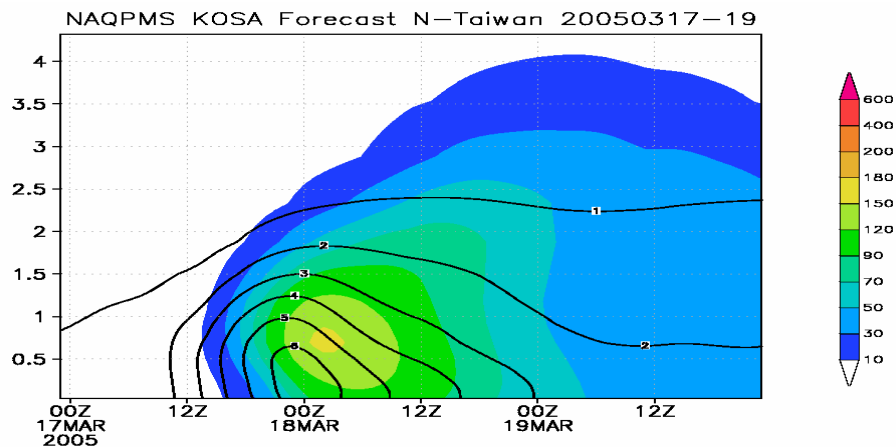


Fig. 9. Simulated vertical profiles of dust concentration (shaded, unit: $\mu\text{g}/\text{m}^3$) and SO_2 (contour, unit: ppb) over northern Taiwan from 00:00 UTC 17 March to 24:00 UTC 19 March 2005.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

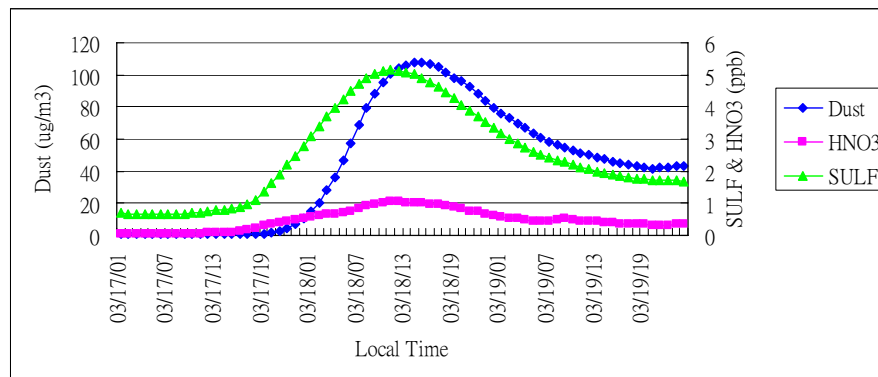


Fig. 10. Simulated time series of Dust, HNO₃ and sulfate concentration over northern Taiwan from 01:00 LST 17 March to 24:00 LST 19 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

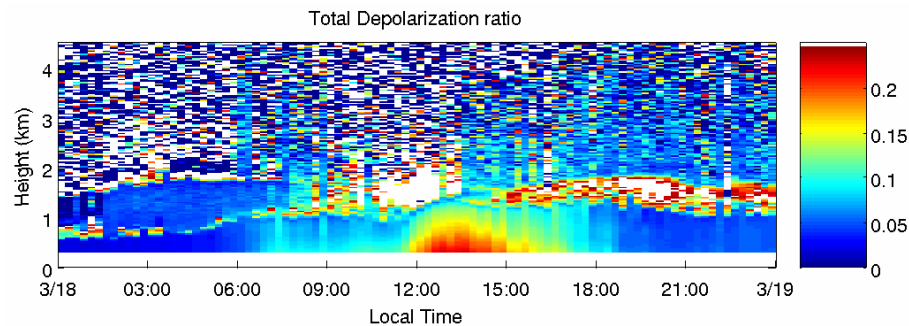
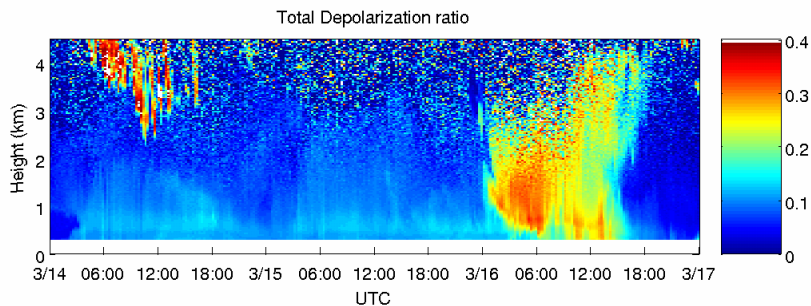


Fig. 11. Results of time series of depolarization ratio at Miyakojima Lidar on 18 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

(a)



(b)

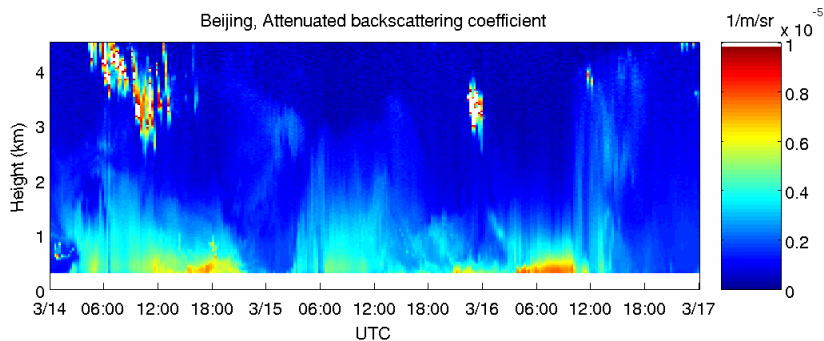


Fig. 12. Results of time series of Lidar at Beijing **(a)** depolarization ratio **(b)** attenuated backscattering coefficient from 00:00 UTC 14 March to 24:00 UTC 16 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Long-range transport
of Asian dust and air
pollutants to Taiwan**

C.-Y. Lin et al.

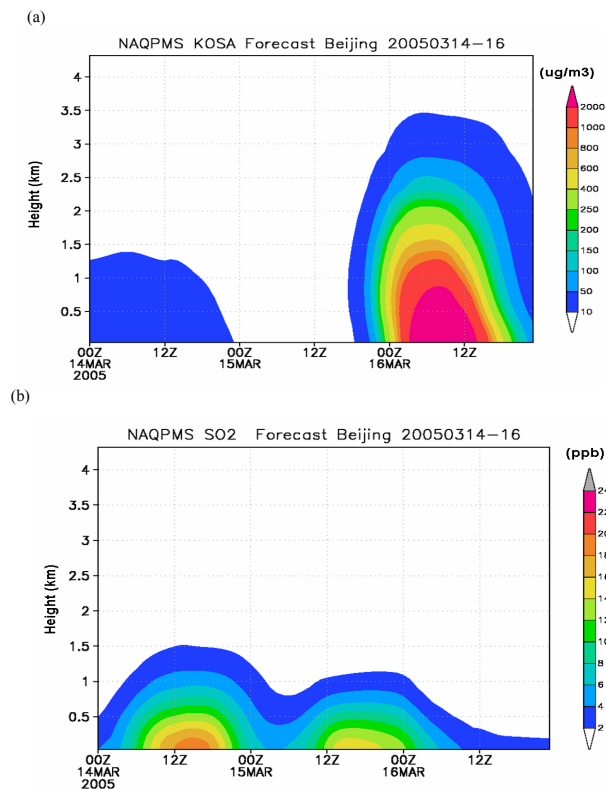


Fig. 13. Simulated vertical profiles of **(a)** dust concentration ($\mu\text{g}/\text{m}^3$) **(b)** SO_2 (ppb) over Beijing from 00:00 UTC 14 March to 24:00 UTC 16 March 2005.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

Long-range transport of Asian dust and air pollutants to Taiwan

C.-Y. Lin et al.

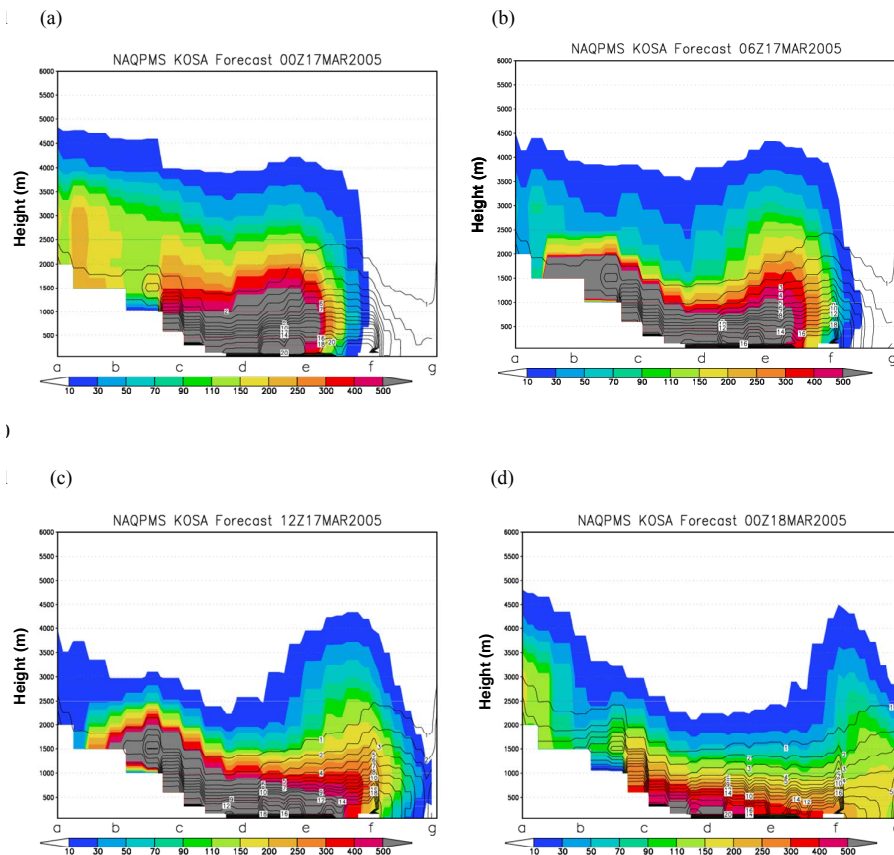


Fig. 14. Simulation results of temporal variations of dust (shaded) and SO₂ concentrations (contour lines) of the vertical profile along the path of the altitude of 1000 m in Fig. 3 at (a) 00:00 UTC 17 March 2005 (b) 06:00 UTC 17 March 2005 (c) 12:00 UCT 17 March 2005 (d) 00:00 UTC 18 March 2005. The letters a to g in x-axis are represented the geographic locations in Fig. 3.

[Title Page](#)
[Abstract](#)
[Introduction](#)
[Conclusions](#)
[References](#)
[Tables](#)
[Figures](#)
[◀](#)
[▶](#)
[◀](#)
[▶](#)
[Back](#)
[Close](#)
[Full Screen / Esc](#)
[Printer-friendly Version](#)
[Interactive Discussion](#)