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## A transboundary transport episode of nitrogen dioxide as observed from GOME and its impact in the Alpine region

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#### Abstract

High tropospheric NO<sub>2</sub> amounts are occasionally detected by satellite-borne spectrometers even though clouds shield the highly polluted boundary layer. We present a method to investigate such events and apply the model to the high NO<sub>2</sub> vertical tropospheric column densities (VTCs) over middle Europe observed from the Global Ozone Monitoring Experiment (GOME) instrument on 17 February 2001. Our case study shows that pollution originally residing near the ground has been advected to higher tropospheric levels by a passing weather front. With backward trajectories, the NO<sub>2</sub> source region is located in central Germany, the Ruhr area and adjacent parts of the Netherlands and Belgium. The highly polluted air masses are traced by forward tra-10 jectories starting from the GOME columns to move further to the Alpine region. Their impact on the air quality there is modeled by combining the NO<sub>2</sub> VTCs observed by GOME with the trajectory calculations and a given NO<sub>2</sub> lifetime. Considering groundbased in-situ measurements in the Alpine region we conclude that for this episode, 50% to 90% of the NO<sub>2</sub> concentration recorded at the sites can be attributed to trans-15

boundary transport during the frontal passage.

#### 1. Introduction

Nitrogen dioxide (NO<sub>2</sub>) plays a key role in photochemical air pollution. Its amount and distribution is of significance directly for air quality and human health and indirectly as
 an ozone precursor (e.g. Seinfeld and Pandis, 1998). NO<sub>2</sub> is formed according to the photochemical equilibrium on a time scale of minutes from the primarily emitted nitrogen oxide (NO), the bulk of which is of anthropogenic origin (Brasseur et al., 2003). The NO<sub>2</sub> concentration at any time and place is determined by the emission of NO<sub>x</sub> (≡NO+NO<sub>2</sub>), its production and loss reactions as well as meteorological transport and dilution. NO<sub>2</sub> is removed from the troposphere by the reaction with the OH radical to HNO<sub>3</sub> which deposits (Kramm et al., 1995), and the nighttime reaction of N<sub>2</sub>O<sub>5</sub> on sur-

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faces and aerosol (Dentener and Crutzen, 1993). Overall, NO<sub>x</sub> and thus NO<sub>2</sub> exhibit a relatively short but highly variable and condition dependent lifetime in the planetary boundary layer (PBL). Leue et al. (2001) used vertical tropospheric NO<sub>2</sub> column densities observed from GOME over the eastern coast of North America to estimate an NO<sub>2</sub>

- <sup>5</sup> lifetime of 27 h in the PBL. Following the plume leaving the coast of India, Kunhikrishnan et al. (2004) estimated a lifetime of 18 h. However, NO<sub>2</sub> has a longer lifetime in the middle and upper troposphere, being of the order of several days to a week (Jaeglé et al., 1998; Seinfeld and Pandis, 1998). Thus, NO<sub>2</sub> may be transported over larger distances once it is lifted to higher tropospheric levels, for example by frontal systems.
- Numerical model results have shown synoptic-scale vertical advection to be a significant transport mechanism for nitrogen oxides (Hov and Flatoy, 1997). In agreement with this, more than 50% of a passive tracer released in the boundary layer was found to be transported to the free troposphere within 24 h during a frontal passage (Donnell et al., 2001). Aircraft campaigns have verified that significant quantities of trace gases
- are lifted up from the PBL into the free troposphere by a warm conveyor belt associated with the frontal system (Bethan et al., 1998). The importance of synoptic uplift and subsequent long-range transport of trace gases for air pollution monitoring was shown by Stohl et al. (2002). These results illustrated transport phenomena at the time scale of hours to days (the synoptic scale) being relevant for the observed NO<sub>2</sub> distribution and for air guality policy, where accurate source attribution and impact assessment is
- required.

The retrieval of the tropospheric NO<sub>2</sub> column from space-borne instrumentation has become available since the launch of the GOME instrument (Burrows et al., 1999) aboard the European Space Agency (ESA) satellite ERS-2 (Earth Remote Sensing <sup>25</sup> Satellite). During the ESA-DUP project POLPO (Pollution Hot Spot Monitoring from GOME Applied to the Po-basin), high values of tropospheric NO<sub>2</sub> columns detected from space were found to comprise a mixture of local emissions and advected NO<sub>2</sub> (Weiss et al., 2002). Furthermore, GOME measurements were successfully used to observe intercontinental transport events of nitrogen dioxide (Spichtinger et al., 2001; 4, 5103-5134, 2004

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Wenig et al., 2003; Stohl et al., 2003). Therefore, combining satellite and ground-based data with transport models is useful to observe transport phenomena influencing the  $NO_2$  distribution.

In the present case study, an episode of "intracontinental" transport of nitrogen dioxide is investigated. The tropospheric NO<sub>2</sub> retrieved from GOME (Richter and Burrows, 2002) is combined with ground-based in-situ measurements and numerical weather model data. In this manner, the frontal transport of NO<sub>2</sub> within Europe and the resulting impact in the Alpine region is identified. In addition, the relative amounts of local and advected NO<sub>2</sub> are estimated by employing a simple transport model based on trajectories, the GOME NO<sub>2</sub> vertical tropospheric columns (VTCs) and an assumed NO<sub>2</sub> lifetime. The intention is to use the NO<sub>2</sub> information available from GOME and remain independent of emission inventories. Thus, the NO<sub>2</sub> emissions are assumed to be proportional to the population density. The latter is added up along backward trajectories

and serves as a proxy for the NO<sub>2</sub> emissions. A method is developed to convert the

<sup>15</sup> proxies into NO<sub>2</sub> concentration by employing the GOME NO<sub>2</sub> VTCs.

#### 2. Measurement data

#### 2.1. GOME measurements

The GOME instrument aboard ESA's ERS-2 satellite measures upwelling radiance from the atmosphere and extraterrestrial solar irradiance. The amount and distribution

of a variety of significant atmospheric trace constituents may be retrieved from these observations (Burrows et al., 1999). It was designed and optimized primarily to measure the global ozone distribution, but also the amounts of the trace gases BrO, OCIO, H<sub>2</sub>O, HCHO and NO<sub>2</sub> can be obtained.

The GOME instrument comprises a nadir-viewing spectrometer which observes si-<sup>25</sup> multaneously the entire spectral range between 240 and 793 nm at channel dependent spectral resolutions between 0.2 and 0.4 nm. From the spectral features of the trace 4, 5103–5134, 2004

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gases, their vertical columns are retrieved (e.g. Richter and Burrows, 2002). Because of its sun-synchronous polar orbit, GOME measurements in middle latitudes are always taken in the late morning local time. With its swath width of 960 km (single ground pixel size is 320 km across track and 40 km along track) GOME reaches full coverage at the <sup>5</sup> mid-latitudes within 2–3 days.

The NO<sub>2</sub> vertical tropospheric column density (VTC) data used in this study were provided by the University of Bremen (see Richter and Burrows, 2002, for full description). Uncertainties in the VTCs are due to the NO<sub>2</sub> fitting algorithm, instrument characteristics, the method of the stratosphere-troposphere separation and atmospheric parameters such as the NO<sub>2</sub> profile, the surface spectral reflectance, the aerosol loading and the clouds. These atmospheric parameters account for the largest part of the uncertainties. Richter and Burrows (2002) carried out a detailed error analysis for the NO<sub>2</sub> VTCs and found a total uncertainty of up to a factor of 2. In order to reduce the uncertainty, the GOME data used in the present study were reprocessed with an adapted cloud top height (CTH) and an assumed location of the NO<sub>2</sub> layer. Furthermore, different sconarios for the structure of the cloud cover and partly cloudy.

ent scenarios for the structure of the cloud cover (closed cloud cover and partly cloudy taking into account cloud mask data) are used.

2.2. Ground-based in-situ measurements

The Swiss National Air Pollution Monitoring Network (NABEL) provides long-term am-<sup>20</sup> bient air measurements (EMPA, 2002; BUWAL, 2002). To investigate the impact of the mesoscale transport process, rural and remote sites located in the Swiss Plateau or elevated and north-exposed stations are suitable as the air mass inflow during this episode is from northern to north-eastern direction.

Data from the following sites (Fig. 1) are used in this study:

 Rigi (47°04′ N, 8°28′ E, 1030 m a.s.l.): A rural site that is well suited for detection of events advected from northern direction, because it is situated on a northexposed mountain slope. In winter, local influences arising from thermal transport 4, 5103–5134, 2004

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are rare.

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- Chaumont (47°03' N, 6°59' E, 1140 m a.s.l.): This site is also situated at an elevated location at some distance from cities and is surrounded by meadows and grassland. The nearest industrial enterprises are situated 5 to 10 km away.
- Taenikon (47°28' N, 8°54' E, 540 m a.s.l.): A site that samples air from the Swiss Plateau. The surroundings are dominated by agriculture and forests, the highway A1 lies in north-eastern direction at a distance of 4 km.
  - 4. Zugspitze (47°25′ N, 10°59′ E, 2963 m a.s.l.): This site is located at the northern rim of the Alps and, therefore, suitable for the detection air masses advected from the North (Umweltbundesamt, 2002). It is used because no measurements from the high-alpine site Jungfraujoch are available during the time of the case study.

At these sites, nitrogen oxides are measured with the chemiluminescence detectors Monitor Labs (ML) 8841 (Rigi), ML 9841 A (Chaumont), ECO Physics CLD 700 (Taenikon) and TE 42 Thermo Environmental Systems/TECAN (Zugspitze). All measurement devices are equipped with a molybdenum converter. It should be noted that these catalytic surface converters may reduce not only NO<sub>2</sub>, but other nitrogen species such as PAN, HNO<sub>2</sub> and HNO<sub>3</sub> to NO. Generally, the PAN mixing ratio can be expected to be small during winter conditions. Wunderli and Gehrig (1991) found a mean February PAN mixing ratio at Davos (Switzerland, 1640 m a.s.l.) of 0.1 ppb. Under cold and

- clear weather conditions, however, elevated PAN mixing ratios were measured in winter with an evening maximum at Davos of 0.6 ppb. HNO<sub>2</sub> is rapidly transformed to OH and NO by photolysis and mixing ratios are normally below the detection limit during the day (Winer and Biermann, 1994). Maximum February/March HNO<sub>3</sub> concentrations were found to be in the order of 0.3 ppb at Jungfraujoch (Zellweger et al., 2003) and
- $_{25}$  0.8 ppb in Duebendorf (Zellweger et al., 1999) at 3580 m a.s.l. and 430 m a.s.l., respectively. As the NO<sub>2</sub> mixing ratios are between 5 and 30 ppb in this case study, we consider the interference of the above discussed nitrogen species to be negligible.

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#### 3. Methods

#### 3.1. Trajectory calculations

Backward and forward trajectories are used to investigate the origin of the detected air masses in the studied GOME columns, their further movement to the Alpine region, and their impact there. The trajectories are based on analyzed wind fields with a six hour temporal and 1°×1° geographical resolution provided by the model of the European Centre for Medium-Range Weather Forecast (ECMWF). Three dimensional kinematic backward and forward trajectories are calculated with the software package "Lagranto" (Wernli and Davies, 1997). Their length is chosen to be long enough for source and impact analysis (several days). The NO<sub>2</sub> distribution within the GOME columns is unknown. Therefore, the backward/forward trajectory arrival/starting points cover the GOME columns both horizontally and vertically in order to account for its whole tropospheric volume (320 km×40 km× tropopause height). The arrival/starting time point for this episode is chosen to be on 17 February 2001, 09:00 UTC.

- For a first assessment of the source regions and the further air mass movement (Sects. 4.3 and 4.4), each column is covered with 70 arrival/starting points; 10 levels from 900 hPa up to 400 hPa in 50 hPa steps, each level containing 7 points. For a more detailed transport assessment (Sect. 3.2 and 4.6), a much higher density of backward and forward trajectories is calculated for each GOME column with 128 (4 along track ×32 across track) arrival/starting points in the horizontal for 23 height levels between 950 hPa and 400 hPa in 25 hPa steps (resulting in a total of 2944 trajectories per column).
  - 3.2. Transport model based on trajectories and GOME observations

A simple transport model based on backward and forward trajectories, the GOME NO2

<sup>25</sup> VTCs, population distribution data, and an assumed NO<sub>2</sub> lifetime is employed to estimate the amount of advected NO<sub>2</sub> at the ground measurement sites in the Alpine re4, 5103–5134, 2004

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gion. The method comprising three steps is discussed in detail in the following sections and is illustrated in Fig. 2. Firstly, the population number is added up and transported along the backward trajectories (as a proxy for NO<sub>2</sub> emissions) and is corrected for the NO<sub>2</sub> lifetime (Sect. 3.2.1). Secondly, above an effective cloud top height, the NO<sub>2</sub> proxies (population number) are summed vertically and converted to number of NO<sub>2</sub> molecules employing the GOME NO<sub>2</sub> VTCs (Sect. 3.2.2). Thirdly, the further movement of the NO<sub>2</sub> polluted air masses is traced with forward trajectories, corrected for NO<sub>2</sub> lifetime and converted into NO<sub>2</sub> mixing ratio when trajectories reach ground near levels (Sect. 3.2.3).

<sup>10</sup> 3.2.1. Pollution uptake and NO<sub>2</sub> lifetime correction

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The NO<sub>2</sub> uptake along the backward trajectories is simulated by passing the trajectories over a 1°×1° resolved population grid (UNEP, 1997) assuming the NO<sub>2</sub> emission to be proportional to the population in the respective grid cell. Following the trajectory forward in time, pollution is taken up when the trajectory reaches a ground near level (50 hPa above model ground). In this case, the number of the population in the respective grid cell is added to the trajectory as a proxy for NO<sub>2</sub> emissions. Further, the NO<sub>2</sub> proxy transported along the trajectory is decreasing every time step according to an assumed NO<sub>2</sub> lifetime of 24 h.

- 3.2.2. Conversion of NO<sub>2</sub> proxies to number of NO<sub>2</sub> molecules
- Each backward trajectory arrives in the appropriate GOME column containing a population number as a proxy for NO<sub>2</sub>. This quantity is converted to a number of NO<sub>2</sub> molecules by applying a conversion factor deduced from relating the GOME measurement with the appropriate proxy values. However, two main issues complicate this approach:
- The sensitivity of the GOME instrument for NO<sub>2</sub> within the column is not constant. For a cloudy pixel, the sensitivity is maximal at the cloud top height (CTH) and

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decreases for NO<sub>2</sub> situated above or within the cloud (Hild et al., 2002).

2. Ground near air mass trajectories based on a 1°×1° resolved wind field might not properly reproduce the frontal movement.

To account for both, we define an effective cloud top height  $(CTH_{eff})$  to be the lowest height in the troposphere above which the NO<sub>2</sub> contributes significantly to the GOME signal. Because GOME sees to some extent into the clouds, the  $CTH_{eff}$  is lower than the real CTH. The  $CTH_{eff}$  is assumed to be constant across the considered GOME pixels (Fig. 3a) and the population numbers above it are added up to column values per GOME pixel. Linearly fitting the proxy VTCs with the GOME NO<sub>2</sub> VTCs yields the adequate  $CTH_{eff}$  where the explained variance (R<sup>2</sup>) maximizes. The population numbers of each trajectory are subsequently converted to number of NO<sub>2</sub> molecules with the conversion factor. The latter corresponds to the ratio between the the sum of the considered proxy VTCs and the sum of the appropriate GOME NO<sub>2</sub> VTCs.

# 3.2.3. Pollution propagation, $NO_2$ lifetime and conversion to $NO_2$ mixing ratios at ground stations

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The NO<sub>2</sub> molecules are propagating along the forward trajectories to the Alpine region. Again, the NO<sub>2</sub> loading is decreasing every time step according to an assumed NO<sub>2</sub> lifetime of 24 h. When a trajectory reaches near ground levels over a 0.5°×0.5° grid superimposed over the Alpine region, the value is assigned to this grid cell. For the
measurement sites in Taenikon and at Rigi and Chaumont, the near ground level is defined to be 50 hPa deep. For the more elevated Zugspitze station, a layer from 50 to 100 hPa from model ground is used. In a last step, the hourly resolved numbers of molecules in the grid cells corresponding to the measurement sites are recalculated to mixing ratios using temperature and pressure data and the vertical box extension of 50 hPa.

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#### 3.2.4. Errors of the method

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Main error sources of the described approach are (i) the assumption of a constant chemical lifetime of  $NO_2$  of 24 h, (ii) the fact that a single air mass trajectory does not implicitly account for dilution, (iii) the uncertainties in the GOME  $NO_2$  VTCs, and (iv) the error associated with the assumption of the population distribution being a representative proxy for  $NO_x$  emissions.

(i) Assuming a constant lifetime for  $NO_2$  is a strong simplification given its complex atmospheric chemistry. However, a relatively long lifetime of 24 h is chosen because the photolysis is relatively weak in winter due to the increased solar zenith angle (Dickerson et al., 1982; Parrish et al., 1983; Wiegand and Bofinger, 2000), and the  $NO_x$  equilibrium is heavily biased towards  $NO_2$ . Furthermore, the loss of  $NO_x$  by reaction with OH is slow because the amount of OH is low in winter (Seinfeld and Pandis, 1998).

(ii) The dilution is to some extent taken into account by the large number of trajectories and their divergence. Furthermore, the effect of dilution can be expected to be rather weak for the studied episode, because the frontal transport lifts up an extended air mass having enhanced  $NO_2$  concentrations and, thus, relatively small concentration gradients.

(iii) For the present study the GOME error is reduced compared to the standard processing by reprocessing the data (Sect. 4.1).

 $_{20}$  (iv) A basic assumption is that the population distribution constitutes a proxy for NO<sub>x</sub> emissions. This seems plausible taking into account the fact that NO<sub>x</sub> is mainly of anthropogenic origin (Brasseur et al., 2003).

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#### 4. Case study – results and discussion

#### 4.1. GOME observation on 17 February 2001

GOME recorded high NO<sub>2</sub> VTC values over central Europe on 17 February 2001 (Fig. 3a). As can be seen in the Meteosat image (Fig. 3b), clouds shielded the ground. Therefore, a significant amount of nitrogen dioxide has to be located within and/or above the cloud cover. However, as pointed out by Richter and Burrows (2002), Leue et al. (2001) and Hild et al. (2002), the albedo effect of underlying clouds can increase the NO<sub>2</sub> signal. Therefore, a reprocessing of the GOME NO<sub>2</sub> VTCs is carried out using an assumed NO<sub>2</sub> layer, the CTH, and the structure of the cloud cover. A closed cloud cover is assumed in the reprocessing version A. In version B, cloud mask information is used and the total tropospheric column (where cloud free) and above cloud columns are combined. The CTH is estimated from live cams in the Alps to be located at an altitude of 3000 m a.s.l. The NO<sub>2</sub> is assumed to occur in a layer between 4000 and 6000 m a.s.l., whereas for standard processing an a priori NO<sub>2</sub> profile with a constant mixing ratio in the lowermost 1.5 km of the troposphere is used (Richter and Burrows, 15 2002). In fact, the real vertical NO<sub>2</sub> distribution is not known, but an inaccurate estimation of this layer turned out to be not very critical. The two reprocessing scenarios A and B yielded quite similar column values as shown in Table 1. Furthermore, the reprocessing results in distinctly lower values than the standard processing, which is due to the above mentioned albedo effect of the clouds (Richter and Burrows, 2002; 20 Leue et al., 2001; Hild et al., 2002). For the further analysis, the overall cloudy case is used.

#### 4.2. Meteorological situation

After a frontal passage on 13 February, a high pressure system installed over southern England and moved slowly to the east, causing the inflow of dry continental polar air to Switzerland during 14–16 February (from Alpine Weather Statistics; MeteoSwiss,

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1985). The flat pressure distribution associated with the high pressure system led to stagnating air over middle Europe. On the 16th, a cold-front formed over north-eastern Europe and moved southward, developing to an occlusion during 17 February (Fig. 4). Cold-fronts are often associated with lightning activity, the latter being known to be
<sup>5</sup> an in-situ source of NO<sub>x</sub> in the free troposphere (e.g. Huntrieser et al., 2002). During the time of our case study, however, no lightning activity was recorded (Wetterzentrale Karlsruhe), and NO<sub>2</sub> production by lightning can, therefore, be excluded.

4.3. Backward trajectories - NO<sub>2</sub> source

The source region of the air pollution detected by the GOME instrument is analyzed with backward trajectories (Fig. 5), focusing on the two NO<sub>2</sub> pixel groups (each composed of two GOME pixels as defined in Fig. 3a) over the border triangle of France, Germany and Switzerland (pixel group 1) and over south-eastern Germany (pixel group 2).

The trajectories indicate that the air masses arriving in the GOME columns are advected from western directions above the Northern Atlantic. They descend over Southern Scandinavia and the North Sea and subsequently turn clockwise over the region of the Netherlands, Belgium and Germany. This is a result of the presence of the high pressure system which installed over Europe before the frontal passage (Fig. 4). Near the ground, the vertical distribution of the trajectories is investigated in detail, because the residence time of an air parcel within the PBL potentiates the uptake of anthropogenic NO<sub>2</sub>. To find the potential source regions, only trajectories residing at a ground distance of less than 100 hPa for at least one (hourly) time step are considered (Fig. 5). Further, in the horizontal projections, time steps along these trajectories are omitted when their distance from ground exceeds 150 hPa. The narrow spatial distonace of the hourly time steps (red date in Fig. 5) show that the sit was rather stagnant

tances of the hourly time steps (red dots in Fig. 5) show that the air was rather stagnant during 15 and 16 February. As a result, the air parcels were exposed to NO<sub>x</sub> emissions for several hours in the highly populated region of the Netherlands, Belgium, Central Germany and the Ruhr area, the latter known to be one of the European hot spot reACPD

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gions concerning  $NO_2$  pollution (e.g. Weiss et al., 2002). The air mass trajectories then exhibit a sudden movement to southern direction associated with frontal uplifting (Fig. 5). These air masses are highly likely to be responsible for the enhanced  $NO_2$  in the pixel groups 1 and 2. The ground near pollution reaches heights of up to 500 hPa above surface pressure (Fig. 5). This corresponds to lifting up to up to 5000 m above model ground. Even higher levels were reached by air formerly traveling at a ground distance of more than 100 hPa.

Comparing the backward calculations for GOME pixel groups 1 and 2 indicates that the air from pixel group 2 over south-eastern Germany is more affected by polluted PBL air, because the density of the trajectories at less than 50 hPa is higher, and their uplifting takes place later. This explains the higher NO<sub>2</sub> VTCs in these GOME columns.

4.4. Forward trajectories – impact in the Alpine region

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The forward trajectories show that the polluted air masses detected by GOME move to the Alpine region including Switzerland (Fig. 6). For both GOME pixel groups, the
forward trajectories' distance to the ground decreases again. This is explained by both an effective subsidence and the increasing altitude of the ground in southern direction. The air masses reach heights below 100 hPa over the Alpine region (Fig. 6). The forward trajectories from GOME pixel group 1 stay relatively bundled and reach Switzerland in its north-western part. The trajectories from GOME pixel group 2, however, show a much larger spread and the polluted air reaching Switzerland can therefore be expected to be more diluted.

4.5. In-situ measurements in the Alpine region

The February 2001 time series of NO<sub>2</sub> measurements from the NABEL and Zugspitze stations indicate several events with high NO<sub>2</sub>, the two most outstanding ones around <sup>25</sup> 7–8 and 17–18 February (Fig. 7). The highest NO<sub>2</sub> mixing ratio of this month was

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measured in the evening on 17 February at the Rigi station with a value of 30 ppb.

The NO<sub>2</sub> measurements at Chaumont and Zugspitze reached their maximum values at the same time with 27 ppb and 6 ppb, respectively. At the Taenikon station, the highest value was measured on 8 February (29 ppb), but the value on the evening of the 17th is, nevertheless, a distinct event with 26 ppb NO<sub>2</sub>. What makes the event on

<sup>5</sup> 17–18 February special is the fact that all four stations show an almost simultaneous and significant increase in their nitrogen dioxide mixing ratios. This is in contrast to the event on 7–8 February, when very high NO<sub>2</sub> values are limited to Taenikon. There, local pollution events are frequent due to the relatively near highway and the inversion layer above the Swiss Plateau which increases pollution concentrations especially in winter.

The simultaneous pollution event on 17th and 18th and the similar mixing ratios at the Swiss sites cannot be explained by local sources, because these are strongly different in the north of each of the considered measurement stations. Furthermore, thermal uplifting, which is typical for elevated sites, is not likely to occur during the night and early morning hours, when the maximum  $NO_2$  mixing ratios were observed. Thus,

the higher NO<sub>2</sub> amounts measured at the elevated stations are expected to be due to larger scale meteorological transport of pollution. It is noteworthy that the maximum NO<sub>2</sub> mixing ratios of 26 ppb measured in Taenikon, 27 ppb at Chaumont and 30 ppb at Rigi are quite similar, even though the annual mean NO<sub>2</sub> amounts show a pronounced

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- decrease with increasing height (Taenikon: 7.0 ppb, Rigi: 4.8 ppb, Chaumont: 3.7 ppb; BUWAL, 2002). These similar mixing ratios indicate a rather well mixed and extended air mass, the latter additionally confirmed by the simultaneous increasing mixing ratio at Zugspitze which is located about 290 km away from Rigi. These findings agree well with the above described air masses that have taken up pollution over an extended region
- <sup>25</sup> and having been transported over mesoscale distances of several hundred kilometers.

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#### 4.6. Transport model

#### 4.6.1. Effective cloud top height and conversion with GOME observations

Our method proposed in Sect. 3.2 and illustrated in the flow chart of Fig. 2 is used to model the transport of NO<sub>2</sub> for this case study. The core element of it is the conversion of population numbers assigned to the trajectories as a proxy for NO<sub>2</sub> to number of NO<sub>2</sub> molecules considering the CTH<sub>eff</sub>. Figure 8 illustrates an intermediate step where the population numbers assigned to the trajectories are added up above CTH<sub>eff</sub> to a 2-dimensional proxy for the NO<sub>2</sub> VTCs. This is done for different CTH<sub>eff</sub>s ranging between 750 and 875 hPa. Especially when CTH<sub>eff</sub> is in the range of 800 to 875 hPa
(which is lower than the real CTH), the modeled NO<sub>2</sub> distribution resembles the GOME observation. The band structure of high NO<sub>2</sub> present in Fig. 8 agrees with the shape of the occlusion in Fig. 4 and indicates the uplift of previously ground-near pollution.

The conversion factor is the ratio between the sum of the 12 GOME NO<sub>2</sub> VTCs (Fig. 3a) and the sum of the 12 vertical columns of the proxies (calculated above the CTH<sub>eff</sub> for each of the 12 GOME pixel locations shown in Fig. 3a). The explained variance ( $R^2$ ) of the linear fit between the two quantities for different CTH<sub>eff</sub> is shown in Fig. 9. As the  $R^2$  culminates at about 840 hPa, we use this CTH<sub>eff</sub> as the relevant level for our further calculation and use the 825 and 850 hPa levels to determine the sensitivity of our model with respect to CTH<sub>eff</sub>. The conversion function (Fig. 9) lets us expect that small changes at this level do not have a large impact on the results. With the conversion factor associated with 840 hPa, the population number of every

trajectory is converted to an NO<sub>2</sub>amount, which further propagates along the forward trajectory.

4.6.2. Comparing in-situ measurements and modeled impact

The model results are shown, together with the measured time series of  $NO_2$ , in Fig. 10. The modeled  $NO_2$  mixing ratios for the Zugspitze station match the measured series

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surprisingly well. The measured variations in the  $NO_2$  mixing ratio are reproduced by the model and the modeled maximum value of about 4 ppb is in the order of the measured value of 6 ppb. Assuming that no local pollution occurred at this high-alpine site, the bulk of the 6 ppb  $NO_2$  can be considered as being due to frontal transport.

<sup>5</sup> This implies, that the model approach somewhat underestimates the transported NO<sub>2</sub>. The time displacement of the modeled relative to the measured time series of about 1 to 2 h is most likely explained by inaccuracies in the ground-near trajectories.

The model output for Taenikon is less correlated with the measurement. At the beginning of the episode, the increasing mixing ratios are well shown by the model results, but the nearly constant values during the night are overestimated. However, the time range of the overall event is reproduced well and the order of magnitude of the NO<sub>2</sub> mixing ratio is matched.

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For the two elevated sites at Rigi and Chaumont, the time series modeled for the grid cells corresponding to the sites did not explain the measured NO<sub>2</sub> time series ad-

- equately. Taking into account adjacent grid cells, the time series found there match the measured mixing ratios much better, which might be due to uncertainties in the ground near trajectories. The latter explanation is additionally supported by the fact that for both stations, the corresponding adjacent grid cells are found to be shifted towards the Swiss Plateau. This favors the following explanation: the air mass flow is approximately
- <sup>20</sup> parallel to the mountain ridges on which the two stations are located (the Jura Mountains for Chaumont and the foothills of the Alps for Rigi, Fig. 6). However, the complex Swiss topography strongly affecting the flow is not properly resolved in the ECMWF model with a resolution of 1°×1°, displacing the ground near trajectories somewhat. The modeled NO<sub>2</sub> in the appropriate adjacent grid cells (Fig. 10) explain the shapes of
- the measured NO<sub>2</sub> time series at Rigi and Chaumont quite reasonable although for the Chaumont station, the modeled mixing ratios are distinctly lower than the measured values. A reason for that could be that the Chaumont station is most affected by air masses from pixel group 1 (Figs. 1, 3a, and 6). Figure 6 indicates that the air from pixel group 2 has to cover a larger distance to reach this most western station and can,

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therefore, be expected to have a minor influence. The colors of pixel group 1 in Fig. 8 further show that the  $NO_2$  proxies for these two pixels might be somewhat underestimated by our method, because for this pixel group with the second highest column values (Fig. 3a), we would expect higher population numbers. This underestimation is probably due to the assumption of a cloud cover with a uniform height.

However, during the investigated episode, roughly 50 to 90% of the measured NO<sub>2</sub> are found to be caused by transboundary transport (Fig. 10). This seems plausible taking into account that the mean NO<sub>2</sub> mixing ratios are much lower. The latter are determined by averaging the NO<sub>2</sub> mixing ratios over the February months of the years 1999 to 2001 for the four stations (with the investigated period from 17 February, 0:00 MET, to 18 February, 12:00 MET, excluded). Typical nighttime NO<sub>2</sub> mixing ratios from 0.5 ppb at Zugspitze to about 5 ppb for the elevated sites at Rigi and Chaumont are found. Only the Swiss Plateau station Taenikon shows evening mixing ratios exceeding 10 ppb. Altogether we conclude from the comparison of the measured with the modeled NO<sub>2</sub>
mixing ratios that the GOME NO<sub>2</sub> VTCs combined with our simple model can explain

<sup>15</sup> mixing ratios that the GOME NO<sub>2</sub> VTCs combined with our simple model can expl the exceptional events with high mixing ratios measured at the four sites.

#### 5. Summary and conclusions

5

The origin and impact of air pollution observed by GOME above a cloud cover has been investigated in a case study. The suggested method to model the NO<sub>2</sub> transport is based on trajectory analysis. The NO<sub>2</sub> uptake is assumed to be proportional to the population density and is further propagated along trajectories (and corrected for lifetime) to constitute a proxy for the NO<sub>2</sub> amount. The GOME measurements are used to convert these proxies to NO<sub>2</sub> concentrations using an effective cloud height. The impact of the pollution transport to the remote alpine area has been modeled and compared with the ground-based measurements.

In the case study of 17 February 2001, it is shown that polluted boundary layer air was lifted up by a frontal system. This led to nitrogen dioxide being located within

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and/or above an extended cloud cover and therefore being detectable by the GOME instrument. Employing backward trajectories, the source region of the NO<sub>2</sub> pollution is identified as being in the central part of Germany, the Ruhr area, and adjacent regions in the Netherlands and Belgium. Forward trajectory calculations indicate the s subsequent movement of the polluted air masses to the Alpine region and predict that measurement stations in Switzerland and southern Germany must be affected by the

- frontal transported NO<sub>2</sub>. The rural in-situ measurements used in this study observed enhanced NO<sub>2</sub> mixing ratios during the night, thus confirming the prediction. Employing a simple model based on air mass trajectories and the GOME observations, we
- estimate that at least 50% of the NO<sub>2</sub> peak mixing ratios measured at the stations during the frontal passage can be attributed to transboundary transport. Thus, although NO<sub>2</sub> is usually considered to be a short lived trace gas mainly being influenced by local sources and sinks, this study shows that significant amounts can be transported over mesoscale distances of several 100 km under favorable meteorological conditions.
- Although the time resolution of the GOME measurements is limited, this case study 15 demonstrates the gain of a synergistic use of satellite observations and meteorological data, as atmospheric transport of pollutants can be followed over time and in 3dimensional space. When the GOME instrument observes a cloudy scene, the lowest part of the troposphere is shielded from the measurement device observing in the visi-
- ble spectral range. Pollution can directly be detected only within and above the clouds. 20 Nevertheless, together with meteorological data such as trajectories, the polluted air masses are traceable backward and forward in time giving evidence of transboundary pollution transport. Clearly, in the future, spatially and temporally higher resolved measurements from space (possibly geostationary) with instrumentation similar to that of
- GOME would facilitate and enhance the assessment of air quality. 25

Acknowledgements. This work was partly funded by the European Space Agency (ESA), the Swiss Federal Agency for Environment, Forest and Landscape (SAEFL) and the Swiss Federal Institute for Materials Testing and Research (EMPA). We gratefully acknowledge H. Wernli (ETH Zurich) for the software package "Lagranto" and for his support concerning trajectory cal-

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culation. We further acknowledge the Swiss National Air Pollution Monitoring Network (NABEL) and the Umweltbundesamt (Germany) for providing ground-based measurements, MeteoSwiss for the weather charts, the European Centre for Medium-Range Weather Forecast (ECMWF) for providing meteorological wind fields and the British Atmospheric Data Centre (BADC) for the Meteosat images

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**Table 1.** GOME  $NO_2$  vertical tropospheric column densities (VTCs) for the standard processing and the two reprocessing scenarios A and B. The  $NO_2$  VTCs are the average values for the two pixel groups shown in Fig. 3a.

	standard processing	reprocessing A	reprocessing B
NO <sub>2</sub> profile	standard a priori profile, constant mixing ratio in the lowermost 1.5 km (Richter and Burrows, 2002)	NO <sub>2</sub> layer between 4000 and 6000 m a.s.l.	NO <sub>2</sub> layer between 4000 and 6000 m a.s.l.
cloud cover	_	closed cloud cover	partly cloudy
cloud top height	_	3000 m a.s.l.	3000 m a.s.l.
$\frac{\text{NO}_2 \text{ VTCs}}{[\times 10^{15} \text{ molec cm}^{-2}]}$	38 (pixel group 1) 64 (pixel group 2)	13 (pixel group 1) 17.5 (pixel group 2)	12 (pixel group 1) 16.5 (pixel group 2)

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**Fig. 1.** The location of the three remote ground-based measurement stations operated by the Swiss National Air Pollution Monitoring Network (NABEL) and the Zugspitze station (Umwelt-bundesamt, 2002) which are used in this case study.





**Fig. 2.** Flow chart illustrating the proposed method (Sect. 3.2) to model NO<sub>2</sub> mixing ratios at ground, employing population numbers assigned to the trajectories and subsequent conversion with GOME NO<sub>2</sub> vertical tropospheric column densities (VTCs). (CTH<sub>*eff*</sub> = effective cloud top height, VMR = volume mixing ratio, VC = vertical column, pixel = GOME pixel).





**Fig. 3.** (a) GOME track on 17 February 2001, showing high  $NO_2$  VTCs above clouds. For the highest values (pixel groups 1 and 2, each consisting of 2 GOME pixels) trajectories are discussed qualitatively. The transport model calculation is based on all of the 12 framed pixels found relevant for the investigated air mass. (b) Meteosat visible wavelength image detected at 9:30 a.m. UTC on 17 February 2001, indicating the extended cloud cover over middle Europe.

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**Fig. 4.** Weather maps showing the southward moving cold front on 16 February 2001 (left) developing into an occlusion on 17 February (right) approaching the Alpine region.

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**Fig. 5.** Backward trajectories arriving on 17 February 2001, 9:00 UTC, in the two GOME pixel groups 1 (a) and 2 (b) defined in Fig. 3a. The colored dots mark hourly trajectory time steps with a trajectory height above model ground of less than 150 hPa (blue), less than 100 hPa (green) and less than 50 hPa (red), respectively. For clarity, in the horizontal projections, only trajectories with a distance from ground of less than 100 hPa for at least one hourly time step are shown and trajectory points exceeding 150 hPa ground distances are omitted. The source regions of the detected NO<sub>2</sub> are likely to be located where the trajectories reside near the ground (red dots).

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**Fig. 6.** The forward trajectories of the two GOME pixel groups 1 and 2 (defined in Fig. 3a) starting on 17 February 2001, 9:00 UTC, indicate the polluted air masses proceed to Switzerland and the Alpine region and further to south-west.

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**Fig. 7.**  $NO_2$  mixing ratios measured at the NABEL ground stations Chaumont, Rigi and Taenikon and at the high-alpine site at Zugspitze in February 2001.



**Fig. 8.** Population numbers (assigned to the backward trajectories as a proxy for NO<sub>x</sub> emissions) added up above different  $CTH_{eff}$  between 750 hPa and 875 hPa to a 2-dimensional proxy for the NO<sub>2</sub> VTCs.

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**Fig. 9.** Explained variance  $R^2$  of the linear fit between the vertical column proxies and the GOME NO<sub>2</sub> VTCs (solid line) and the respective conversion factors (dashed line) for different CTH<sub>eff</sub>.



**Fig. 10.** NO<sub>2</sub> mixing ratios measured during the investigated episode (black) and results of the model assessment of the advected additional NO<sub>2</sub> (red). The thick red curve corresponds to the optimal  $CTH_{eff}$  (840 hPa), the thin curves result from the adjacent (trajectory) levels in 825 (upper curve) and 850 hPa (lower curve). For the Zugspitze and Taenikon stations, the time series from the grid cells containing the measurement sites are used, whereas for Rigi and Chaumont, adjacent grid cells show more reliable results (see text).

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