

1 **Responses to comments of anonymous referees**

2 We would like to thank the anonymous referees for reviewing our manuscript and their
3 useful comments and suggestions. Please find below the reviewer's comments (black)
4 and our detailed responses (blue).

5 6 **Anonymous Referee #1**

7
8 This study presents 14C level in carbonaceous aerosols, and inorganic ions and
9 anhydrosugars in PM_{2.5} during the spring period of 2013 in Beijing and Guangzhou.
10 The study concludes that both primary and secondary matter from fossil sources played
11 a key role in the blooming phase of the pollution episode. In my opinion, several critical
12 issues need to be addressed in this study as detailed below.

13 1. The chemical components of PM_{2.5} in both Guangzhou and Beijing have evidently
14 seasonal variations. The sources of major aerosol chemical components are expected to
15 be highly variable with seasons, and the causes for haze formation could also be
16 different from season to season. For example, secondary inorganic aerosols were the
17 dominant sources in summer while biomass burning sources were in autumn/winter in
18 Beijing (Zhang et al., 2013; Cheng et al., 2013; Du et al., 2014). Heavy haze mostly
19 occurred in winter instead of in spring in Beijing. Why this study picked a spring period?
20 Will the results be applicable to other seasons? Moreover, only a few data samples are
21 presented in this study to illustrate one aerosol pollution episode during the spring
22 period in Guangzhou and Beijing. How is the representative of the episode case in
23 spring for these two cities?

24 We appreciate the reviewer's comment. We agree with the reviewer's opinion that PM_{2.5}
25 concentrations and its emission sources would vary seasonally with the highest PM_{2.5}
26 mass concentration probably appears in winter season in some years. Given this,
27 numerous studies have been preferentially performed to investigate the chemical,
28 physical, and microbial characteristics of PM_{2.5} particles during the winter season both
29 in Beijing and Guangzhou (Cao et al. 2003; Sun et al. 2006; Li et al. 2013; Zhao et al.
30 2013; Sun et al. 2013; Cao et al. 2014; Liu et al. 2014; Quan et al. 2014; Wang et al.
31 2014; Zhang et al. 2014; Zhang et al. 2015; Zheng et al. 2015).

32 However, the season of winter is not the only season during when severe air pollution
33 occur. Zhang et al. (2013) reported that the average PM_{2.5} concentrations in Beijing
34 were 126 (39-355) µg/m³, 138 (41-226) µg/m³, 135 (45-251) µg/m³, and 139 (48-355)
35 µg/m³, for spring, summer, autumn, and winter, respectively. Based on the results from
36 this study, it is easily to find that PM_{2.5} concentration in spring is practically the same
37 with winter during the year of 2009-2010. In our study, the average PM_{2.5} concentration
38 in Beijing during the 2013 spring was 218 µg/m³, which is 1.4-fold higher the winter

39 of 2013 (*Huang et al. 2014*) and is 8.7-fold higher health threshold ($25 \mu\text{g}/\text{m}^3$)
40 suggested by World Health Organization (WHO). In Guangzhou, the average $\text{PM}_{2.5}$
41 concentration during 2013 spring is $90.6 \mu\text{g}/\text{m}^3$, which is 1.3-fold higher the winter of
42 2013 (*Huang et al. 2014*) and 3.6-fold higher the WHO health threshold. Thus, in
43 addition to winter season, more detailed studies are urgently needed to explore the
44 origins and formations of $\text{PM}_{2.5}$ in spring.

45 Furthermore, few studies regarding radiocarbon measurements nowadays were
46 performed in Chinese cities, and most of them mainly focused on winter season. In an
47 early study, for example, *Huang et al. (2014)* and *Zhang et al. (2015)* have already
48 presented the radiocarbon levels of organic carbon (OC) and elemental carbon (EC) in
49 the winter season of Beijing and Guangzhou. According to the results of *Zhang et al.*
50 *(2015)*, ~50-60% of OC and 70-77% of EC were derived from the fossil-fuel related
51 sources. These results are impacted severely by the coal combustion in Beijing due to
52 the central heating in winter season. In spring, our results showed that 41% of OC and
53 67% of EC on average came from the fossil sources in Beijing, which is lower than that
54 in winter (*Zhang et al. 2015*), evidently implying the more impacts from biomass-
55 burning smokes on Beijing during the spring season. In addition, we further explore the
56 radiocarbon levels of water-soluble OC and water-insoluble OC, respectively, and we
57 found that the radiocarbon signals of water-soluble OC is significantly different from
58 water-insoluble OC in both cities. In other words, some critical information would be
59 lost if atmospheric scientists only analyze the radiocarbon level of OC. We also
60 combined the measurements of biomass-burning organic tracers with an aim of
61 studying the contributions of fossil and non-fossil sources to the secondary organic
62 carbon (SOC).

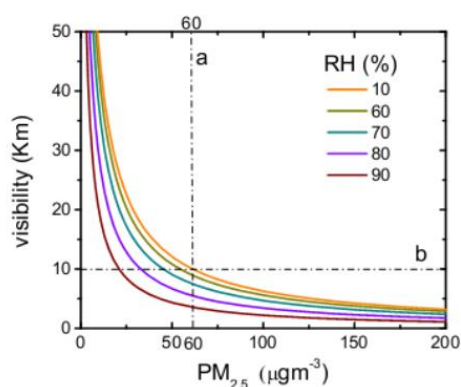
63 In this study, we collected 35 $\text{PM}_{2.5}$ samples in total (21 in Beijing and 14 in
64 Guangzhou, respectively). The sum of selected samples for radiocarbon measurements
65 were 9 and the total radiocarbon data were 45 including OC, water-soluble OC, water-
66 insoluble OC, EC, and total carbon. We didn't measure the radiocarbon levels of all
67 samples we collected as the cost for radiocarbon measurements is really expensive. In
68 fact, the radiocarbon data we presented in this study are comparable to previous and
69 recent studies, such as *Szidat et al. (2006)* (24 data, 1 site for 1 season), *Liu et al. (2013a)*
70 *(14 data, 1 site for 1-year)*, *Zhang et al. (2015)* (48 data, 4 cities for 1 season), *Zong et*
71 *al (2015)* (8 data, 1 site for 1 season), and *Andersson et al. (2015)* (20 data, 3 cities for
72 1 season). Given that radiocarbon is the most powerful tracer at present for determining
73 the fossil and non-fossil sources and we have analyzed the radiocarbon levels of
74 different carbon fraction of $\text{PM}_{2.5}$ samples from low to high concentrations in both cities,
75 we think that our dataset presented in the manuscript could represent the season we
76 studied. In addition, although this study is performed in the spring season only, the
77 results presented in this manuscript can provide an important implication and

78 inspiration to the future studies regarding $PM_{2.5}$ carbonaceous aerosols and the method
79 used in this study probably can be employed in other regions and seasons in the world
80 in the future in order to get a better understanding of the air pollution.

81

82 2. While haze is mainly caused by aerosol pollution, they are two different definitions
83 and should not be used interchangeably. Besides PM level, relative humidity is another
84 key factor for haze formation. This work does not provide any information on aerosol
85 optical properties and meteorological parameters. It only presents aerosol pollution
86 episodes in Beijing and Guangzhou, rather than haze processes. Moreover, inorganic
87 and organic particulate matters have different mass scattering efficiencies, it is
88 inappropriate to conclude haze is predominantly driven by organic matter and nitrate
89 only based on mass concentrations. The focus of the study may need to be modified
90 based on available data and analysis conducted in the study.

91 Thanks for the comment. We agree that both PM level and relative humidity are key
92 factors influencing the formation of haze pollution (atmospheric visibility < 10 km).
93 Please see Fig. r1 below. $PM_{2.5}$ level is the most key indicator of the haze for a given
94 range of relative humidity. In this study, the relative humidity of those selected samples
95 for radiocarbon measurements were in the range of 20-60% in Beijing, under which the
96 curve line in Fig. r1 is very close, implying that relative humidity played a limited role
97 on the haze and the haze formation is practically controlled by the $PM_{2.5}$ level. As for
98 Guangzhou samples, the relative humidity ranged from 70% to 90%, under which both
99 humidity and $PM_{2.5}$ exert impacts on haze formation (Fig. r1). However, all these
100 samples we selected in Guangzhou have a $PM_{2.5}$ concentration higher than $60 \mu\text{g}/\text{m}^3$
101 when the humidity play a limited role on the haze formation. Thus, we think that it is
102 reasonable to use $PM_{2.5}$ as an indicator to tracer the haze formation in this study.



103

104 Fig. r1. A plot of atmospheric visibility versus $PM_{2.5}$ concentration at different relative
105 humidity (Liu et al. 2013b)

106

107 In addition, we have presented aerosol optical depth (AOD) and meteorological

108 parameters in Supporting Information. Please see Fig. S1 and Fig. S2. In the 3.3 section
109 of manuscript, we have discussed the role of relative humidity in the formation of
110 secondary organic carbon. In this sampling campaign, we fortunately observed a
111 dynamic process of haze formation in Guangzhou. Please see Fig. S4 and the 3.4 section
112 of manuscript. Thus, in the Introduction section of manuscript, we said “the source
113 dynamics of individual primary and secondary aerosols during the haze bloom-decay
114 process in Guangzhou basing day-to-day time serials and Beijing basing low-to-high
115 PM_{2.5} concentrations were investigated as well”. We agree that inorganic and organic
116 particulate matters have different mass scattering efficiencies, while the dataset
117 presented in this study mainly focused on the analysis of radiocarbon, organic tracers,
118 and ions. For the haze formation process in Guangzhou, we did find that fossil-derived
119 secondary organic matter and nitrate made a largest impact on the PM_{2.5} concentrations.
120 We seriously considered this reviewer’s comments and we still think that our statements
121 on the roles of organic matter and nitrate is reasonable in this study.

122
123 3. Page 34959, line 20. The OM is the biggest contributor to PM_{2.5} in Beijing and
124 Guangzhou. However, 41~49% fraction of PM_{2.5} were unidentified. Such a large
125 percentage of unknown mass fractions of PM_{2.5} make people wonder if the conclusions
126 are valid. The authors are suggested to reconstruct the PM_{2.5} mass based on daily data.
127 The OM conversion factors are 2.1 and 1.3 to WSOC and WSIC, respectively. However,
128 these conversion factors are obtained from a rural site in Hungary from January to
129 September 2000 (Kiss et al., 2002). These factors may not be suitable for application in
130 the urban area of Beijing and Guangzhou.

131 Thanks for the comment. In this study, 49% and 41% of PM_{2.5} in Beijing and
132 Guangzhou were not unidentified basing on the measurement of organic carbon (OC),
133 elemental carbon (EC), and 8 water-soluble ions, respectively. The percentage of
134 unidentified fraction in PM_{2.5} largely depends on the numbers of measured chemicals
135 and the errors of analytical methods (*Andrews et al. 2000*). In Taiwan, ~35-40% of
136 PM_{2.5} can’t be identified basing on the measurement of 10 elements, 8 water-soluble
137 ions, OC and EC (*Chen et al. 2001*). In a recent study, after measuring the elementals,
138 OC, EC, and water-soluble ions, *Huang et al. (2014)* found that the percentages of
139 unidentified fraction of PM_{2.5} in China were in the range of 10-36%. Relatively higher
140 proportion of unidentified fraction in PM_{2.5} probably because we didn’t analyze the
141 elementals in this study. We have made corresponding explanation in the revised
142 manuscript.

143 Few studies associated with the conversion factors of WSOC and WIOC are performed
144 simultaneously in China. While, the conversion factors used in this study are
145 comparable to those previous studies conducted in both urban and rural sites. For
146 example, 1.3 was employed to estimate WIOM in 3 urban and 1 rural sites in United

147 States (*Sun et al. 2011*). *Favez et al. (2009)* used 1.4 and 2.1 as the conversion factors
148 for WIOC and WSOC in Paris, respectively. In Beijing, 2.1 was also used as a
149 conversion factor to calculate the WSOM (*Chen et al. 2014*). Thus, it should be
150 reasonable to employ 1.3 and 2.1 to calculate the WIOM and WSOM in this study,
151 respectively. We have added these references in the revised manuscript.

152

153 4. Page 34959, the authors are suggested to provide direct evidences that the higher
154 calcium concentration in Beijing is related with dust storm during the sampling period,
155 as it could also come from road or/and construction fugitive dust.

156 Thanks for the comment. This suggestion has been employed in the revised manuscript.

157

158 5. Page 34964-34965. The growth of carbonaceous aerosols and nitrate also related to
159 the combined influence of boundary layer height, humidity, chemical reaction and their
160 thermodynamic properties. It would be better to provide more convincing evidence to
161 show the growth is more related to the sources rather than the meteorological factors.

162 Thanks for the comment. We agree with the reviewer's point that many factors would
163 exert impacts on the growth of carbonaceous aerosols and water-soluble ions. While,
164 for a given city during one season, the metrological parameters would not change
165 significantly with the exception of some special events. In the manuscript, we gave an
166 implication that different humidities between Beijing and Guangzhou would result in
167 their different SOC composition. In the section 3.4 of manuscript, we firstly discussed
168 an integrated haze process in Guangzhou and a trend of low-to-high PM_{2.5} in Beijing
169 using a combined measurement of radiocarbon, organic tracers, and water-soluble ions.
170 We also explore the potential impacts of air masses on the growths of chemical species
171 in this section. For example, C/C₀ values would change significantly when the air
172 masses reaching Guangzhou from south region. A more detailed discussion regarding
173 all metrological parameters, thermodynamic properties, and chemical reactions needs a
174 modeling work which is beyond the scope of this study.

175

176 **Anonymous Referee #3**

177 This study showed the measurements of radiocarbon, anhydrosugars, and water-soluble
178 ions in PM_{2.5} collected in Guangzhou and Beijing, China. The authors found that non-
179 fossil fuel sources make a large contribution to the total carbonaceous aerosols in
180 Chinese megacities. The authors believed that both primary and secondary species are
181 important to the haze formation in Chinese cities. As for me, the results presented in
182 this paper are interesting and will expand our understanding of bad air pollution. This
183 study has a clear logic writing and completely within the scope of ACP. Therefore, I
184 recommend its publication after the following issues are addressed.

185 Comments:

186 1. My main concern is the limited period of sampling: only about 12 samples in each
187 site without considering the different seasons or period of years that could influence the
188 CAs emissions and sources. More information about the experiment should be
189 presented in the text, such as reproducibility of the experiment result.

190 Thanks for the comment. As for the sample numbers and the sampling season, please
191 see our responses above (Line 63-80). In the revised manuscript, we have added the
192 reproducibility of the measurements of WIOC, EC, WSOC, and the water-soluble ions.

193

194 2. Please add the corresponding literature to support Lev/Gal/Man is the biomass
195 burning-specific organic tracers. Line 213: 24-hour or annual standard?

196 Thanks for the comment. The literature and standard has been shown in manuscript.

197

198 3. Line 245: “the EC” → “EC”. In addition, I agree with you that EC suspending over
199 urban areas are dominated by fossil-fuel combustion. What about rural region or remote
200 areas? Does any study focused on this? I think it would be better for readers to
201 understand this paper if the authors can cite same papers conducted in rural/remote.
202 Readers can get a direct comparison results for radiocarbon levels in mind.

203 Thanks for the comment. In the manuscript, “the EC” has been replaced by “EC” and
204 the new study regarding remote areas has been added.

205

206 4. Line 281: POC during atmospheric transportation may experience aging processes. I
207 suggest the authors adding “gas” before “reactions”.

208 Thanks for the comment. This suggestion has been accepted.

209

210 5. Page 34955, Line 23: Both ambient OC and EC can be produced by the combustion
211 activities such as coal and biofuel. I am curious why their conversion factors are slightly
212 different. Specifically, the conversion factors mentioned in this study are 1.10 and 1.06
213 for EC and OC, respectively. The authors should clarify this.

214 We would like to thank the reviewer for this comment. Atmospheric non-fossil OC is
215 derived from biomass burning and primary biogenic emissions (solid and gas phases)
216 while non-fossil EC only come from biomass burning. Conversion factor for biogenic
217 emissions was obtained from the $^{14}\text{CO}_2$ level in the background locations during
218 sampling time. However, tree age need to be considered for the conversion factor of
219 biomass burning derived OC and EC because atmospheric $^{14}\text{CO}_2$ level are changing
220 every year due to the nuclear bomb effect in the early of 1960s. That is why small
221 difference exists between the conversion factors of EC and OC. We have simply
222 mentioned this in the section of 2.3.

223

224 6. Page 34958, Line 5 – 9: To my knowledge, the ratio of biomass burning OC to

225 levoglucosan is highly unstable in different emission sources. One of the most possibly
226 most important reasons, in my opinion, is the various biomass types. In particular,
227 Guangzhou is in south China while Beijing is in north China. How the author get the
228 corresponding ratios to calculate the primary biomass-burning OC?

229 Thanks for the comment. Levoglucosan (Lev) and its isomers, i.e., galactosan (Gal),
230 and mannosan (Mann), are excellent organic tracers of biomass-burning smoke. We
231 agree that the ratio of OC/Lev in biomass-burning smoke is largely depending on
232 biomass types. The burnings of different biomass types would result in different ratios
233 of Lev/Mann, Mann/Gal, and OC/Lev (*Liu et al. 2014*). Thus, in this study, we
234 simultaneously analyzed Lev, Gal, and Mann in Guangzhou and Beijing to constrain
235 the value of OC/Lev. Then, we found that the biomass types for burning in Beijing is
236 different from that in Guangzhou and we got their different OC/Lev values.

237

238 References:

239 Andersson, A., Deng, J., Du, K., Zheng, M., Yan, C., Sköld, M. and Gustafsson, O.,
240 2015. Regionally-varying combustion sources of the January 2013 severe haze
241 events over eastern China. *Environmental Science & Technology*, 49(4), pp.2038-
242 2043.

243 Andrews, E., Saxena, P., Musarra, S., Hildemann, L.M., Koutrakis, P., McMurry, P.,
244 Olmez, I. and White, W.H., 2000. Concentration and composition of atmospheric
245 aerosols from the 1995 SEAVS experiment and a review of the closure between
246 chemical and gravimetric measurements. *Journal of the Air & Waste Management*
247 *Association*, 50(5), pp.648-664.

248 Cao, C., Jiang, W., Wang, B., Fang, J., Lang, J., Tian, G., Jiang, J. and Zhu, T.F., 2014.
249 Inhalable microorganisms in Beijing's PM_{2.5} and PM₁₀ pollutants during a severe
250 smog event. *Environmental Science & Technology*, 48(3), pp.1499-1507.

251 Cao, J.J., Lee, S.C., Ho, K.F., Zhang, X.Y., Zou, S.C., Fung, K., Chow, J.C. and Watson,
252 J.G., 2003. Characteristics of carbonaceous aerosol in Pearl River Delta Region,
253 China during 2001 winter period. *Atmospheric Environment*, 37(11), pp.1451-1460.

254 Chen, J., Qiu, S., Shang, J., Wilfrid, O.M., Liu, X., Tian, H. and Boman, J., 2014. Impact
255 of relative humidity and water soluble constituents of PM_{2.5} on visibility impairment
256 in Beijing, China. *Aerosol and Air Quality Research*, 14, pp.260-268.

257 Chen, K.S., Lin, C.F. and Chou, Y.M., 2001. Determination of source contributions to
258 ambient PM_{2.5} in Kaohsiung, Taiwan, using a receptor model. *Journal of the Air &*
259 *Waste Management Association*, 51(4), pp.489-498.

260 Favez, O., Cachier, H., Sciare, J., Sarda-Estève, R. and Martinon, L., 2009. Evidence
261 for a significant contribution of wood burning aerosols to PM_{2.5} during the winter
262 season in Paris, France. *Atmospheric Environment*, 43(22), pp.3640-3644.

263 Huang, R.J., Zhang, Y., Bozzetti, C., Ho, K.F., Cao, J.J., Han, Y., Daellenbach, K.R.,

264 Slowik, J.G., Platt, S.M., Canonaco, F., Zotter, P., Wolf, R., Pieber, S.M., Bruns, E.A.,
265 Crippa, M., Ciarelli, G., Piazzalunga, A., Schwikowski, M., Abbaszade, G.,
266 Schnelle-Kreis, J., Zimmermann, R., An, Z., Szidat, S., Baltensperger, U., and
267 Haddad, I.E., 2014. High secondary aerosol contribution to particulate pollution
268 during haze events in China. *Nature*, 514(7521), pp.218-222.

269 Li, Z., Gu, X., Wang, L., Li, D., Xie, Y., Li, K., Dubovik, O., Schuster, G., Goloub, P.,
270 Zhang, Y. and Li, L., 2013. Aerosol physical and chemical properties retrieved from
271 ground-based remote sensing measurements during heavy haze days in Beijing
272 winter. *Atmospheric Chemistry and Physics*, 13(20), pp.10171-10183.

273 Liu, D., Li, J., Zhang, Y., Xu, Y., Liu, X., Ding, P., Shen, C., Chen, Y., Tian, C. and
274 Zhang, G., 2013a. The use of levoglucosan and radiocarbon for source apportionment
275 of PM_{2.5} carbonaceous aerosols at a background site in East China. *Environmental
276 Science & Technology*, 47(18), pp.10454-10461.

277 Liu, X.G., Li, J., Qu, Y., Han, T., Hou, L., Gu, J., Chen, C., Yang, Y., Liu, X., Yang, T.
278 and Zhang, Y., 2013b. Formation and evolution mechanism of regional haze: a case
279 study in the megacity Beijing, China. *Atmospheric Chemistry Physics*, 13(9),
280 pp.4501-4514.

281 Liu, J., Li, J., Zhang, Y., Liu, D., Ding, P., Shen, C., Shen, K., He, Q., Ding, X., Wang,
282 X. and Chen, D., 2014. Source apportionment using radiocarbon and organic tracers
283 for PM_{2.5} carbonaceous aerosols in Guangzhou, South China: contrasting local-and
284 regional-scale haze events. *Environmental Science & Technology*, 48(20), pp.12002-
285 12011.

286 Quan, J., Tie, X., Zhang, Q., Liu, Q., Li, X., Gao, Y. and Zhao, D., 2014. Characteristics
287 of heavy aerosol pollution during the 2012–2013 winter in Beijing, China.
288 *Atmospheric Environment*, 88, pp.83-89.

289 Wang, Y., Yao, L., Wang, L., Liu, Z., Ji, D., Tang, G., Zhang, J., Sun, Y., Hu, B. and
290 Xin, J., 2014. Mechanism for the formation of the January 2013 heavy haze pollution
291 episode over central and eastern China. *Science China Earth Sciences*, 57(1), pp.14-
292 25.

293 Sun, Y., Zhuang, G., Tang, A., Wang, Y. and An, Z., 2006. Chemical characteristics of
294 PM_{2.5} and PM₁₀ in haze-fog episodes in Beijing. *Environmental Science &
295 Technology*, 40(10), pp.3148-3155.

296 Sun, Y.L., Wang, Z.F., Fu, P.Q., Yang, T., Jiang, Q., Dong, H.B., Li, J. and Jia, J.J., 2013.
297 Aerosol composition, sources and processes during wintertime in Beijing, China.
298 *Atmospheric Chemistry and Physics*, 13(9), pp.4577-4592.

299 Sun, Y., Zhang, Q., Zheng, M., Ding, X., Edgerton, E.S. and Wang, X., 2011.
300 Characterization and source apportionment of water-soluble organic matter in
301 atmospheric fine particles (PM_{2.5}) with high-resolution aerosol mass spectrometry
302 and GC–MS. *Environmental Science & Technology*, 45(11), pp.4854-4861.

303 Szidat, S., Prévôt, A.S., Sandradewi, J., Alfarra, M.R., Synal, H.A., Wacker, L. and
304 Baltensperger, U., 2007. Dominant impact of residential wood burning on particulate
305 matter in Alpine valleys during winter. *Geophysical Research Letters*, 34(5),
306 pp.L05820

307 Zhao, X.J., Zhao, P.S., Xu, J., Meng, W., Pu, W.W., Dong, F., He, D. and Shi, Q.F., 2013.
308 Analysis of a winter regional haze event and its formation mechanism in the North
309 China Plain. *Atmospheric Chemistry and Physics*, 13(11), pp.5685-5696.

310 Zhang, J.K., Sun, Y., Liu, Z.R., Ji, D.S., Hu, B., Liu, Q. and Wang, Y.S., 2014.
311 Characterization of submicron aerosols during a month of serious pollution in
312 Beijing, 2013. *Atmospheric Chemistry and Physics*, 14(6), pp.2887-2903.

313 Zhang, Y.L., Huang, R.J., El Haddad, I., Ho, K.F., Cao, J.J., Han, Y., Zotter, P., Bozzetti,
314 C., Daellenbach, K.R., Canonaco, F., Slowik, J.G., Salazar, G., Schwikowski, M.,
315 Schnelle-Kreis, J., Abbaszade, G. Zimmermann, R., Baltensperger, U. Prévôt, A.S.H.,
316 and Szidat, S., 2015. Fossil vs. non-fossil sources of fine carbonaceous aerosols in
317 four Chinese cities during the extreme winter haze episode of 2013. *Atmospheric*
318 *Chemistry and Physics*, 15(3), pp.1299-1312.

319 Zhang, R., Jing, J., Tao, J., Hsu, S.-C., Wang, G., Cao, J., Lee, C.S.L., Zhu, L., Chen,
320 Z., Zhao, Y., and Shen, Z., 2013. Chemical characterization and source
321 apportionment of PM_{2.5} in Beijing: Seasonal perspective. *Atmospheric Chemistry*
322 *and Physics*, 13(14), pp.7053-7074.

323 Zheng, G.J., Duan, F.K., Su, H., Ma, Y.L., Cheng, Y., Zheng, B., Zhang, Q., Huang, T.,
324 Kimoto, T., Chang, D. and Pöschl, U., 2015. Exploring the severe winter haze in
325 Beijing: the impact of synoptic weather, regional transport and heterogeneous
326 reactions. *Atmospheric Chemistry and Physics*, 15(6), pp.2969-2983.

327 Zong, Z., Chen, Y., Tian, C., Fang, Y., Wang, X., Huang, G., Zhang, F., Li, J. and Zhang,
328 G., 2015. Radiocarbon-based impact assessment of open biomass burning on
329 regional carbonaceous aerosols in North China. *Science of The Total Environment*,
330 518, pp.1-7.