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Title: Flask measurements of long-lived trace gases at three ground stations in India

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22 Abstract

23 With a rapid growth in population and economic development, emissions of greenhouse gases (GHGs) from the Indian subcontinent have sharply increased during recent decades. 24 However, evaluation of regional fluxes of GHGs and characterization of their spatial and 25 26 temporal variations by atmospheric inversions remain uncertain due to a sparse regional atmospheric observation network. As a result of an Indo-French collaboration, three new 27 atmospheric stations were established in India at Hanle (HLE), Pondicherry (PON) and Port 28 29 Blair (PBL), with the objective of monitoring the atmospheric concentrations of GHGs and other trace gases. Here we present the results of the measurements of CO₂, CH₄, N₂O, SF₆, 30 CO, and H₂ from regular flask sampling at these three stations over the period 2007–2011. 31 For each species, annual means, seasonal cycles and gradients between stations were 32 calculated and related to variations in the natural GHG fluxes, anthropogenic emissions, and 33 34 the monsoon circulations. Covariances between species at the synoptic scale were analyzed to investigate the likely source(s) of emissions. The flask measurements of various trace gases at 35 the three stations show potential to constrain the inversions of fluxes over Southern and 36 Northeastern India. However, this network of ground stations needs further extension to other 37 parts of India to better constrain the GHG budgets at regional and continental scales. 38

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40 1 Introduction

Since the pre-industrial times, anthropogenic greenhouse gas (GHG) emissions have 41 progressively increased the radiative forcing of the atmosphere, leading to impacts on the 42 climate system and human society (IPCC, 2013, 2014a, b). With rapid socio-economic 43 44 development and urbanization during recent decades, a large and growing share of GHG emissions is contributed by emerging economies like China and India. In 2010, India became 45 the world's third largest GHG emitter, next to China and the USA (EDGAR v4.2; Le Quéré 46 47 et al., 2014). Between 1991 and 2010, anthropogenic GHG emissions in India increased by $\sim 100\%$ from 1.4 to 2.8 GtCO₂eq, much faster than rates of most developed countries and 48 economies like the USA (9%) and EU (-14%) over the same period (EDGAR v4.2). Without 49 a systematic effort at mitigation, this trend would continue in the coming decades, given that 50 the per capita emission rate in India is still much below that of the more developed countries. 51 52 For comparison, in 2010, the per capita GHG emission rates were 2.2, 10.9, 17.6, and 21.6 tonCO₂eq/capita for India, the UK, Russia, and the USA, respectively (EDGAR v4.2). In 53 particular, non-CO₂ GHG emissions are substantial in India, most of which are contributed by 54 agricultural activities over populous rural areas (Pathak et al., 2010). In 2010, anthropogenic 55 CH₄ and N₂O emissions in India amounted to 29.6 TgCH₄ (≈ 0.62 GtCO₂eq) and 0.8 TgN₂O 56 (≈0.23 GtCO₂eq), together accounting for 32% of the country's GHG emissions, of which 57 contributions of the agricultural sector were 60 and 73%, respectively (EDGAR v4.2). 58 Reducing emissions of these two non-CO₂ GHGs may offer a more cost-effective way to 59 60 mitigate future climate change than by attempting to directly reduce CO₂ emissions (Montzka et al., 2011). 61

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Effective climate mitigation strategies need accurate reporting of sources and sinks of GHGs.This is also a requirement of the United Nations Framework Convention on Climate Change

65 (UNFCCC). Current estimates of GHG budgets in India, either from the top-down approaches (based on atmospheric inversions) or bottom-up approaches (based on emission 66 inventories or biospheric models), have larger uncertainties than for other continents. For 67 instance, Patra et al. (2013) reported a net biospheric CO_2 sink of -104±150 TgCyr⁻¹ over 68 South Asia during 2007–2008 based on global inversions from 10 TransCom-CO₂ models 69 (Peylin et al., 2013) and a regional inversion (Patra et al., 2011b), while the bottom-up 70 approach gave an estimate of -191±193 TgCyr⁻¹ over the period of 2000–2009 (Patra et al., 71 2013). Notably, these estimates have uncertainties as high as 100–150%, much larger 72 73 compared to those of Europe (~30%, see Luyssaert et al., 2012) and North America (~60%, see King et al., 2015), where observational networks are denser and emission inventories are 74 75 more accurate. Evaluation of N₂O emissions from 5 TransCom-N₂O inversions also exhibited 76 the largest differences over South Asia (Thompson et al., 2014b). A main source of uncertainty is the lack of atmospheric observation datasets with sufficient temporal and 77 spatial coverage (Patra et al., 2013; Thompson et al., 2014b). Networks of atmospheric 78 79 stations that were used to constrain estimates of global GHG fluxes show gaps over South Asia (Patra et al., 2011a; Thompson et al., 2014b, c; Peylin et al., 2013), with Cape Rama 80 (CRI – 15.08°N, 73.83°E, 60m a.s.l.) on the southwest coast of India being the only Indian 81 station (Rayner et al., 2008; Patra et al., 2009; Tiwari et al., 2011; Bhattacharya et al., 2009; 82 Saikawa et al., 2014). Recently a few other ground stations have been established in Western 83 84 India and the Himalayas to monitor GHGs and atmospheric pollutants, which are located in Sinhagad (SNG – 18.35°N, 73.75°E, 1600m a.s.l.; Tiwari and Kumar, 2012; Tiwari et al., 85 2014), Mount Abu (24.60°N, 72.70°E, 1700m a.s.l.; S. Lal, personal communication), 86 Ahmedabad (23.00°N, 72.50°E, 55m a.s.l.; Lal et al., 2015), Nainital (29.37°N, 79.45°E, 87 1958m a.s.l.; Kumar et al., 2010) and Darjeeling (27.03°N, 88.15°E, 2194m a.s.l.; Ganesan 88 et al., 2013). Most of these stations started to measure atmospheric GHG concentrations very 89

90 recently (e.g. Sinhagad – since 2009; Ahmedabad – since 2013; Mount Abu – since 2013; Nainital – since 2006; Darjeeling – since 2011), and datasets are not always available. In 91 addition, aircraft and satellite observations have also been carried out and provided useful 92 93 constraints on estimates of GHG fluxes in this region (Park et al., 2007; Xiong et al., 2009; Schuck et al., 2010; Patra et al., 2011b; Niwa et al., 2012; Zhang et al., 2014). Although 94 inclusion of measurements from South Asia significantly reduces uncertainties in top-down 95 96 estimates of regional GHG emissions (e.g., Huang et al., 2008; Niwa et al., 2012; Zhang et al., 2014), a denser atmospheric observational network with sustained measurements is still 97 98 needed over this vast and fast-growing region for an improved, more detailed, and necessary understanding of GHG budgets. 99

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Besides the lack of a comprehensive observational network, the seasonally reversing Indian 101 monsoon circulations and orographic effects complicate simulation of regional atmospheric 102 103 transport, which contributes to uncertainty of the inverted GHG fluxes (e.g., Thompson et al., 2014b). The Indian monsoon system is a prominent meteorological phenomenon in South 104 Asia, which, at lower altitudes, is characterized by strong southwesterlies from the Arabian 105 106 Sea to the Indian subcontinent during the boreal summer, and northeasterlies during the boreal winter (Goswami, 2005). The summer monsoon is associated with deep convection, 107 which mixes the boundary layer air into the upper troposphere and lower stratosphere 108 (Schuck et al., 2010; Lawrence and Lelieveld, 2010). On the contrary, little deep convection 109 occurs over South Asia during the winter monsoon period, which carries less moisture 110 111 (Lawrence and Lelieveld, 2010). The Indian monsoon also impacts biogenic activities (e.g., vegetation growth, microbial activity) and GHG fluxes through its effects on rainfall 112 variations (Tiwari et al., 2013; Valsala et al., 2013; Gadgil, 2003). Given that accurate 113 atmospheric transport is critical for retrieving reliable inversion of GHG fluxes, an 114

observational network that comprises a range of altitudes including monitoring stations in mountainous regions would be valuable for validating and improving atmospheric transport models.

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119 Since the 2000s, three new atmospheric ground stations have been established in India as part 120 of the Indo-French collaboration, with the objective of monitoring the atmospheric concentrations of major GHGs and other trace gases in flask air samples. Of the three Indian 121 stations, Hanle (HLE) is a high-altitude station situated in the western Indian Himalayas, 122 123 while Pondicherry (PON) and Port Blair (PBL) are tropical surface stations located respectively on the southeastern coast of South India and on an oceanic island in the 124 southeastern Bay of Bengal. In this study, we briefly describe the main features of these 125 stations and present time series of flask air sample measurements of multiple trace gases at 126 HLE, PON, and PBL over the period 2007-2011. Descriptions of the three stations as well as 127 128 methods used to analyze and calibrate the flask measurements are given in Sect. 2. For each station, we measure the atmospheric concentrations of four major GHG species (CO₂, CH₄, 129 N₂O and SF₆) and two additional trace gases (CO and H₂). Among these trace gases, CO₂, 130 CH₄ and N₂O are the three most abundant GHGs in the atmosphere, and the UNFCCC 131 requires each Non-Annex I Party to regularly report anthropogenic emissions of these gases 132 (MoEF, 2012). Sulfur hexafluoride (SF_6) is widely considered as a good tracer for 133 anthropogenic activities with a long atmospheric lifetime and almost purely anthropogenic 134 sources (Maiss et al., 1996), and the Non-Annex I Parties are also encouraged to provide 135 information on its anthropogenic emissions (MoEF, 2012). Although CO and H₂ are not 136 GHGs by themselves, both of them play critical roles in the CH₄ budgets through reaction 137 with the free OH radicals (Ehhalt and Rohrer, 2009). Besides, CO and H₂ are good tracers for 138 139 biomass/biofuel burning (Andreae and Merlet, 2001), an important source of GHG emissions 140 that is quite extensive in India (Streets et al., 2003; Yevich and Logan, 2003). Time series of atmospheric concentrations of all these trace gases are analyzed for each station to 141 characterize the annual means and seasonal cycles, with results and discussions presented in 142 Sect. 3. Gradients between different stations are interpreted in the context of regional flux 143 patterns and monsoon circulations (Sect. 3.1). We examine synoptic variations of CO₂, CH₄ 144 and CO by analyzing the co-variances between species, using deviations from their smoothed 145 fitting curves (Sect. 3.2). Finally, we investigate two abnormal CH₄ and CO events at PBL 146 and propose likely sources and origins (Sect. 3.3). A summary of the paper as well as 147 148 conclusions drawn from these results are given in Sect. 4.

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150 **2** Sampling stations and methods

151 **2.1 Sampling stations**

Figure 1 and S1 in the supplement show the locations of HLE, PON, and PBL. We also 152 present five-day back-trajectories from each station for all sampling dates in April-June 153 (AMJ; Fig. 1a), July-September (JAS; Fig. 1b), October-December (OND; Fig. 1c) and 154 155 January-March (JFM; Fig. 1d), respectively. Note that this four-period classification scheme is slightly different from the climatological seasons defined by the India Meteorological 156 Department (IMD; Attri and Tyagi, 2010), in which months of a year are categorized into the 157 pre-monsoon season (March-May), SW monsoon season (June-September), post-monsoon 158 season (October-December) and the winter season (January and February). We adapted the 159 IMD classification to facilitate better display and further analyses (e.g., Sect. 3.2), making 160 161 sure that samples are fairly evenly distributed across all seasons. The back-trajectories were generated using the Hybrid Single Particle Lagrangian Integrated Trajectory (HYSPLIT4) 162 model (Draxler and Rolph, 2003), driven by wind fields from the Global Data Assimilation 163

System (GDAS) archive data based on National Centers for Environmental Prediction (NCEP)
 model output (<u>https://ready.arl.noaa.gov/gdas1.php</u>).

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The Hanle (HLE) station (32.780 °N, 78.960 °E, 4517 m a.s.l.) is located in the campus of the 167 Indian Astronomical Observatory (IAO) atop Mt. Saraswati, about 300 m above the 168 169 Nilamkhul Plain in the Hanle Valley of southeastern Ladakh in northwestern Himalayas. The 170 station was established in 2001 as a collaborative project between the Indian Institute of Astrophysics and LSCE, France. The flask sampling inlet is installed on the top of a 3 m mast 171 172 fixed on the roof of a 2m high building, and the ambient air is pumped through a Dekabon tubing with a diameter of 1/4". The area around the station is a cold mountain desert, with 173 sparse vegetation and a small population of ~ 1700 distributed over an area of ~ 20 km². 174 Anthropogenic activities are limited to small-scale crop production (e.g., barley and wheat) 175 and livestock farming (e.g., yaks, cows, goats, and sheep). The nearest populated city of Leh 176 177 (34.25 °N, 78.00 °E, 3480 m a.s.l.) with ~27 000 inhabitants, lies 270 km to the northwest of this station. By virtue of its remoteness, high altitude, and negligible biotic and anthropogenic 178 influences, HLE is representative of the background free tropospheric air masses in the 179 180 northern mid-latitudes. Regular flask air sampling at this station has been operational since February, 2004, and continuous in-situ CO₂ measurements started in September, 2005. Over 181 the period 2007-2011, a total of 188 flask sample pairs were collected at HLE. Back-182 trajectories show that, HLE dominantly samples air masses that pass over northern Africa and 183 the Middle East throughout the year, and those coming from South and Southeast Asia during 184 185 the SW monsoon season (Fig. 1a). More detailed station information of HLE would be found in several earlier publications (Babu et al., 2011; Moorthy et al., 2011). 186

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188 The Pondicherry (PON) station (12.010 °N, 79.860 °E, 20 m a.s.l) is located on the southeast coast of India, about 8 km north of the city of Pondicherry with a population of ~240,000 189 (Census India, 2011). The station was established in collaboration with Pondicherry 190 191 University in 2006. The flask sampling inlet, initially located on a 10 m mast fixed on the roof of the University Guest House, was later moved to a 30 m high tower in June, 2011. The 192 ambient air is pumped from the top of the tower through a Dekabon tubing with a diameter of 193 1/4". The surrounding village Kalapet, has a population of ~9000 (Sivakumar and Anitha, 194 2012). A four-lane highway runs nearly 80 m to the west of the station with a low traffic flow 195 196 especially during the nighttime, while the Indian Ocean stands about 100 m to the east of the station. Moreover, the two nearest megalopolises of Chennai and Bangalore, both with 197 populations of over 6 million (Census India, 2011), are approximately 143 km to the north 198 199 and 330 km to the west of the station. Given its proximity to an urban area and a highway, 200 PON can be influenced by local emissions. Although the highway nearby has a low traffic flow, in-situ measurements at PON (not presented in this paper) do show that this site is 201 202 heavily polluted by local emissions during nighttime. In order to minimize the influences of local GHG sources/sinks, flask air sampling at PON is performed between 12:00 and 18:00 203 local time (LT) (actually 97% of flask samples taken between 12:00 and 14:00 LT), when the 204 sea breeze moves clean air masses towards the land and the boundary layer air is well mixed. 205 Further, we also remove outliers that are likely polluted by local emissions and not 206 207 representative of regional background concentrations (see Sect. 2.3.1 for details). We believe that through these two approaches the local influences at PON should be sufficiently 208 minimized. Flask sampling at PON began in September, 2006 and over the period 2007–2011, 209 210 a total of 185 flask sample pairs were collected at the site. As shown in Fig. 1a, the air masses received at PON are strongly related to the monsoon circulations. During the boreal summer 211 212 when the southwest monsoon prevails, PON is influenced by air masses originating from the

Arabian Sea and South India, whereas during the boreal winter, it receives air masses from the east and northeast parts of the Indian subcontinent, and the Bay of Bengal. During the boreal spring and autumn when the monsoon changes its direction, air masses of both origins are observed.

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218 The Port Blair (PBL) station (11.650 °N, 92.760 °E, 20 m a.s.l.) is located on the small 219 Andaman Islands in the southeastern Bay of Bengal, ~1400 km east of Pondicherry, and roughly 600 km west of Myanmar and Thailand. The station was established in collaboration 220 221 with the National Institute of Ocean Technology (NIOT), India, and flask air sampling was initiated in July, 2009. The flask sampling inlet is located on the top of a 30 m high tower, 222 and the ambient air is pumped through a Dekabon tubing with a diameter of 1/4". The main 223 city on the Andaman Islands, Port Blair, is about 8 km to the north of the station, with a 224 population of ~100,000 (Census India, 2011). Due to its proximity to vegetation and a small 225 226 rural community, the station is not completely free from influences of local GHG fluxes. Therefore, flask samples at PBL are obtained in the afternoon between 13:00 and 15:00 LT, 227 when the sea breeze moves towards the land, to minimize significant local influences. Over 228 229 the period 2009-2011, a total of 63 flask sample pairs were collected at PBL. Backtrajectories show that the air masses sampled at PBL are also controlled by the seasonally 230 reversing monsoon circulations (Fig. 1a), with air masses from the Indian Ocean south of the 231 Equator during the southwest monsoon season, and from the northeast part of the Indian 232 subcontinent, the Bay of Bengal, and Southeast Asia during the northeast monsoon season. 233 234 As for PON, air masses of both origins are detected at PBL during the boreal spring and autumn when the monsoon changes its direction. 235

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237 **2.2 Flask sampling and analysis**

238 2.2.1 Flask sampling

239 In principle, flask samples are taken in pairs on a weekly basis at all three stations. However, in practice air samples are collected less frequently (on average every 10-12 days) due to bad 240 meteorological conditions or technical problems. Whole air samples are filled into pre-241 242 conditioned 1-L cylindrical borosilicate glass flasks (Normag Labor und Prozesstechnik GmbH, Germany) with valves sealed by caps made from KEL-F (PTCFE) fitted at both ends. 243 Besides, a few flasks are equipped with valves sealed by the original Teflon PFA O-ring 244 245 (Glass Expansion, Australia), accounting for ~5.0, 1.2 and 1.1% of air samples respectively for HLE, PON and PBL during the study period. For the air samples stored in flasks sealed 246 with the original Teflon PFA O-ring, corrections are made for the loss of CO₂ (+0.0027 247 ppm/day) and of N_2O (+0.0035 ppb/day) after analyses of the samples. The correction factors 248 are empirically determined based on laboratory storage tests using flasks filled with 249 250 calibrated gases. Drying of the air is performed using 10 g of magnesium perchlorate $(Mg(ClO_4)_2)$ confined at each end with a glass wool plug in a stainless steel cartridge, located 251 upstream of the pump unit. Tests have shown that use of the magnesium perchlorate drier 252 doesn't result in any loss of the target compound. To prevent entrainment of material inside 253 the sampling unit, a 7 µm filter is attached at the end of the cartridge. The flasks are flushed 254 prior to sampling for 10-20 min at a rate of 4–5 L min⁻¹, and the air is compressed in the 255 flasks to about 1 bar over the ambient pressure (pump: KNF Neuberger diaphragm pump 256 powered by a 12V DC motor, Germany, N86KNDC with EPDM membrane). The 257 pressurizing process lasts for less than a minute. 258

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260 2.2.2 Flask analyses

261 On average the flasks arrive at LSCE, France about 150 days after the sampling date. Leakage could occur during shipment, and any flask sample with too low pressure will be 262 flagged in the analyses. Flask samples are analyzed for CO₂, CH₄, N₂O, SF₆, CO, and H₂ with 263 two coupled gas chromatograph (GC) systems. The first gas chromatograph (HP6890, 264 Agilent) is equipped with a flame ionization detector (FID) for CO₂ and CH₄ detection, and a 265 standard electron capture detector (ECD) for N₂O and SF₆ detection. It is coupled with a 266 267 second GC equipped with a reduced gas detector (RGD, Peak Laboratories, Inc., California, USA), for analyzing CO and H₂ via reduction of HgO and subsequent detection of Hg vapor 268 269 through UV absorption. In the following paragraph we summarize the major configurations and parameters of the GC systems (also see Table S1). Further details on the analyzer 270 271 configuration are described in Lopez (2012) and Yver et al. (2009).

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Both GC systems are composed of three complementary parts: the injection device, the 273 274 separation elements and the detection sensors. As flask samples are already dried during sampling, they are only passed through a 5 mL glass trap maintained in an ethanol bath kept 275 at -55°C by a cryocooler (Thermo Neslab CC-65) to remove any remaining water vapor. The 276 277 air samples are flushed with flask overpressure through a 15 mL sample loop for CO₂ and CH₄ analyses, a 15 mL sample loop for N₂O and SF₆ analyses, and a 1 mL sample loop for 278 CO and H₂, at a flow rate of 200 mL min⁻¹. After temperature and pressure equilibration, the 279 air sample is injected into the columns. The CO₂ and CH₄ separation is performed using a 280 Havesep-Q ($12' \times 3/16$ "OD, mesh 80/100) analytical column placed in an oven at 80°C, with 281 a N₂ 5.0 carrier gas at a flow rate of 50 ml min⁻¹. Detection of CH₄ and CO₂ (after conversion 282 to CH₄ using a Ni catalyst and H₂ gas) is performed in the FID kept at 250°C. The flame is 283 fed with H₂ (provided by a NM-H₂ generator from F-DBS) at a flow rate of 100 ml min⁻¹ and 284 zero air (provided by a 75-82 zero air generator from Parker-Balston) at a flow rate of 300 ml 285

min⁻¹. For N₂O and SF₆ separation, a Hayesep-Q (4' \times 3/16" OD, mesh 80/100) pre-column 286 and a Hayesep-Q ($6' \times 3/16''$ OD, mesh 80/100) analytical column, both placed in an oven at 287 80°C, are used together with an Ar/CH₄ carrier gas at a flow rate of 40 ml min⁻¹. Detection of 288 N₂O and SF₆ is performed in the ECD heated at 395°C. For CO and H₂, we use a Unibeads 289 1S pre-column (16.5" \times 1/8" OD; mesh 60/80) to separate the two gases from the air matrix, 290 and use a Molecular Sieve 5Å analytical column ($80^{\circ} \times 1/8^{\circ}$ OD; mesh 60/80) to effectively 291 separate H₂ from CO. Both columns are placed in an oven kept at 105°C. CO and H₂ are 292 analyzed in the RGD detector heated to 265°C. A measurement takes ~5 min and calibration 293 gases are measured at least every 0.45 hour. For CO₂, CH₄, N₂O, and SF₆, we use two 294 calibration gases, one with a high concentration and the other with a low concentration. The 295 calibration and quality control cylinders are filled and spiked in a matrix of synthetic air 296 297 containing N₂, O₂ and Ar prepared by Deuste Steininger (Germany). The concentration of the sample is calculated using a linear regression between the two calibration gases with a time 298 interpolation between the two measurements of the same calibration gas (Messager, 2007; 299 Lopez, 2012). For CO and H₂, we use only one standard and apply a correction for the non-300 linearity of the analyzer (Yver et al., 2009; Yver, 2010). The nonlinearity is verified regularly 301 with 5 calibration cylinders for CO and 8 calibration cylinders for H₂. All the calibration 302 gases themselves are determined against an international primary scale (CO₂: WMOX2007; 303 CH₄: NOAA2004; N₂O: NOAA2005A; SF₆: NOAA2005; CO: WMOX2004; H₂: 304 305 WMOX2009; Hall et al., 2007; Dlugokencky et al., 2005; Jordan and Steinberg, 2011; Zhao and Tans, 2006). Finally, a "target" gas is measured every two hours after the calibration 306 gases as a quality control of the scales and of the analyzers. The repeatability of the GC 307 308 systems estimated from the target cylinder measurements over several days is 0.06 ppm for CO₂, 1 ppb for CH₄, 0.3 ppb for N₂O, 0.1 ppt for SF₆, 1 ppb for CO and 2 ppb for H₂. 309

Additional quality control is made by checking the values of a flask target (a flask filled withcalibrated gases) placed on each measurement sequence.

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For both of the GC systems, data acquisition, valve shunting, and temperature regulation are entirely processed by the Chemstation software from Agilent. Concentrations are calculated with a software developed at LSCE using peak height or area depending on the species.

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317 2.2.3 Uncertainty of flask measurements

318 Uncertainties in the measured concentrations stemmed from both the sampling method and the analysis. Collecting flask samples in pairs and measuring each flask twice allow us to 319 evaluate these uncertainties. A large discrepancy between two analyses of the same flask 320 321 reveals a problem in the analysis system, while a difference between a pair of flasks reflects both analysis and sampling uncertainties. Flask pairs with differences in mole fractions 322 beyond a certain threshold are flagged and rejected (see Table S2 in the supplement for the 323 threshold for each species). The percentages of flask pairs retained for analyses are 65.9-88.3% 324 for CO₂, 88.6-94.1% for CH₄, 74.6-91.5% for N₂O, 92.0-96.8% for SF₆, 68.6-88.3% for CO, 325 326 and 76.2-95.2% for H_2 (Table S3). For each species, we evaluate the uncertainties by averaging differences between the two injections of the same flask (analysis uncertainty) and 327 between the pair of flasks (analysis uncertainty + sampling uncertainty) across all retained 328 flask pairs from the three Indian stations (Table S4). For all species except SF₆, the sampling 329 330 uncertainty turns out to be the major uncertainty, while the analysis uncertainty is equivalent to the reproducibility of the instrument. For SF₆, both uncertainties are extremely low due to 331 332 the small amplitudes and variations of the signals at the three stations.

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At LSCE, there are regular comparison exercises in which flasks are measured by different 334 laboratories on the same primary scale (e.g., Inter-Comparison Project (ICP) loop, Integrated 335 non-CO2 Greenhouse gas Observing System (InGOS) 'Cucumber' intercomparison project). 336 337 These comparisons allow us to estimate possible biases in our measurements. In Table S4, the bias for each species is calculated over the sampling period using the ICP flask exercise that 338 circulates flasks of low, medium and high concentrations between different laboratories. For 339 340 CO₂, CH₄, SF₆ and CO, the biases are reported against NOAA (NOAA-LSCE) as it is the laboratory responsible for the primary scales for these species. The bias of H₂ is calculated 341 342 against Max Planck Institute for Biogeochemistry (MPI-BGC) in Jena, Germany, which is responsible for the primary scale of H₂. The bias of N₂O is reported against MPI-BGC instead 343 of NOAA. Although NOAA is responsible for the primary scale of N₂O, the instruments they 344 345 use for the N₂O flask analyses and cylinder calibration are not the same as ours. For CH₄, N₂O, SF₆ and H₂, the estimated biases are within the noise level of the instrument and 346 negligible. For CO₂ and CO, we observe a bias of -0.15 ± 0.11 ppm and 3.5 ± 2.2 ppb, 347 respectively (Table S4), which could be due to the nonlinearity of the instrument and/or an 348 improper attribution of the secondary scale values. 349

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351 2.3 Data analyses

352 **2.3.1 Curve-fitting procedures**

For each time series of flask measurements, we calculated annual means and seasonal cycles using a curve-fitting routine (CCGvu) developed by NOAA/CMDL (Thoning et al., 1989). A smoothed function was fitted to the retained data, consisting of a first-order polynomial for the growth rate and two harmonics for the annual cycle (Levin et al., 2002; Ramonet et al., 2002), as well as a low pass filter with 80 and 667 days as short-term and long-term cutoff 358 values, respectively (Bakwin et al., 1998). Residuals were then calculated as the differences between the original data and the smoothed fitting curve. Any data lying outside three 359 standard deviations of the residuals were regarded as outliers and discarded from the time 360 361 series (Harris et al., 2000; Zhang et al., 2007). This procedure was repeated until no outliers remained. These outliers were likely a result of pollution by local emissions and not 362 representative of regional background concentrations. The data discarded through this 363 filtering procedure accounts for less than 4% of the retained flask pairs after flagging (Table 364 S3). Particularly, for PON where observations can be influenced by local emissions, we also 365 366 tried to use CO as a tracer and filtered time series of other species by CO outliers. Results show that this additional filtering does not make significant difference to the trends, seasonal 367 cycles and mean annual gradients (relative to HLE) for all the other species at PON (Table S5, 368 369 Fig. S2). On the other hand, however, the approach may substantially decrease the number of samples used to fit the smooth curve (e.g. ~38% for CH₄) and result in larger data gaps 370 (Table S5, Fig. S2), probably compromising reliability of the analyses. Therefore finally we 371 didn't use CO as a tracer of local emissions for additional filtering. 372

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For each species at each station, the annual means, as well as the amplitude and phases of 374 seasonal cycles, were determined from the smoothed fitting curve and its harmonic 375 component. We bootstrapped the curve-fitting procedures 1000 times by randomly sampling 376 the original data with replacement to further estimate uncertainties of annual means and 377 seasonal cycles. Since the observation records are relatively short, we used all flask 378 379 measurements between 2006 and 2011 to fit the smooth curve when available (Fig. S3). For each species, we also compared results with measurements from stations outside India that 380 belong to networks of NOAA/ESRL (http://www.esrl.noaa.gov/gmd/) and Integrated Carbon 381

382 Observation System (ICOS, https://www.icos-cp.eu/). Locations and the fitting periods of 383 these stations are also given in Table S6, Figs. S1 and S3.

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385 2.3.2 Ratio of species

We analyzed CH₄-CO, CH₄-CO₂, and CO-CO₂ correlations using the residuals from the 386 smoothed fitting curves that represent synoptic-scale variations (Harris et al., 2000; Ramonet 387 388 et al., 2002; Grant et al., 2010). To determine the ratio between each species pair, as in previous studies, we used the slope calculated from the orthogonal distance regression (Press 389 et al., 2007) to equally account for variances of both species (Harris et al., 2000; Ramonet et 390 391 al., 2002; Schuck et al., 2010; Baker et al., 2012). We also bootstrapped the orthogonal distance regression procedure 1000 times and estimated the 1- σ uncertainty for each ratio. 392 The analyses were performed with R3.1.0 (R Core Team, 2014) following the recipes 393 described in Teetor (2011). 394

395

396 3 Results and discussions

397 **3.1 Annual means and seasonal cycles**

398 **3.1.1 CO**₂

Figure 2 shows CO₂ flask measurements and the corresponding smooth curves fitted to the data at HLE, PON and PBL, as well as two additional NOAA/ESRL stations, namely Plateau Assy, Kazakhstan (KZM – 43.25 °N, 77.88 °E, 2519 m a.s.l.) and Waliguan, China (WLG – 36.29 °N, 100.90 °E, 3810 m a.s.l.) (Dlugokencky et al., 2014b). HLE observed an increase in CO₂ mole fractions from 382.3 ± 0.3 to 391.4 ± 0.3 between 2007 and 2011, with annual mean values being lower (by 0.2–1.9 ppm) than KZM and WLG (Fig. 2c and d, Table 1). At PON, the annual mean CO₂ mole fractions were generally higher than at HLE, with 406 differences ranging 1.8–4.3 ppm (Fig. 2a, Table 1). The annual mean CO₂ gradient between PON and HLE reflects the altitudinal difference of the two stations, and a larger influence of 407 CO₂ emissions at PON, mostly from South India (Fig. 1a, EDGAR v4.2). Besides this, as 408 409 shown in Fig. 2a and Table 1, the CO₂ observations at PON are influenced by synoptic scale events, with a large variability of individual measurements relative to the fitting curve (see 410 the relative SDs (RSD) in Table 1). At PBL, the annual mean CO₂ mole fractions were on 411 average 1.2–1.8 ppm lower than that at HLE (Table 1). The negative gradient between PBL 412 and HLE is particularly large during summer, possibly due to clean air masses transported 413 414 from the ocean (Figs. 1a and 2b). Note that caution should be exercised in interpreting the gradient at PBL because of the data gap and short duration of the time series. 415

416

The different CO₂ seasonal cycles observed at the five stations reflect the seasonality of 417 carbon exchange in the northern terrestrial biosphere as well as influences of long-range 418 419 transport and the monsoon circulations. At HLE, the peak-to-peak amplitude of the mean seasonal cycle was 8.2±0.4 ppm, with the maximum early May and the minimum mid-420 September, respectively (Fig. 3, Table 1). The mean seasonal cycle estimated from flask 421 422 measurements at HLE is in good agreement with that derived from vertical profiles of in-situ aircraft measurements over New Delhi (~500 km southwest of HLE) from the 423 Comprehensive Observation Network for Trace gases by Airliner (CONTRAIL, 424 http://www.cger.nies.go.jp/contrail/) project at similar altitudes (R=0.98-0.99, p<0.001, Fig. 425 3a; Machida et al., 2008), and back-trajectories show that they represent air masses with 426 427 similar origins as HLE (Fig. S8), confirming that HLE is representative of the regional free mid-troposphere background concentrations. When comparing with the two other background 428 stations located further north in central and East Asia, a significant delay of the CO₂ phase is 429 seen at HLE compared to KZM and WLG (Fig. 3b, Table 1). We also note that the CO₂ mean 430

431 seasonal cycle at HLE is in phase with the composite zonal marine boundary layer (MBL) reference at 32°N, while for KZM and WLG, an advance in the CO₂ phase by about 1 month 432 is observed compared to the zonal MBL reference (Fig. S4; Dlugokency et al., 2014b). The 433 434 phase shifts in the CO₂ seasonal cycles mainly result from differences in the air mass origins between stations. HLE is influenced by the long-range transport of air masses from mid-435 latitudes around 30 °N, as well as air masses passing over the Indian subcontinent in the 436 boreal summer (Fig. 1a), therefore its CO₂ seasonal cycle is related to the seasonality of 437 vegetation activity over the entire latitude band. KZM and WLG receive air masses passing 438 439 over the Middle East and western Asia as HLE does, but they are also influenced by air masses of more northern origins with signals of strong CO₂ uptake over Siberia during JAS 440 441 (Fig. S5). At WLG, negative CO₂ synoptic events, indicative of large-scale transport of air 442 masses exposed to carbon sinks in Siberia in summer, were also detected by in-situ 443 measurements during 2009-2011 (Fang et al., 2014). Moreover, the back trajectories indicate that WLG and KZM are more influenced than HLE by air masses that have exchanged with 444 the boundary layer air being affected by vegetation CO₂ uptake (Fig. S6a,d,e). This could 445 additionally account for the earlier CO₂ phase observed at KZM and WLG compared to HLE. 446

447

At PON and PBL, the peak-to-peak amplitudes of the CO₂ mean seasonal cycles were 448 7.6 \pm 1.4 and 11.1 \pm 1.3 ppm, with their maxima observed in April. The CO₂ mean seasonal 449 cycle is controlled by changes in the monsoon circulations, in combination with the 450 451 seasonality of CO₂ biotic exchange and anthropogenic emissions in India. During the boreal 452 winter when the NE monsoon prevails, PON and PBL receive air masses enriched in CO₂ from the East and Northeast Indian subcontinent as well as from Southeast Asia, with large 453 anthropogenic CO₂ emissions (EDGAR v4.2; Wang et al., 2013; Kurokawa et al., 2013). 454 455 During April when the SW monsoon begins to develop, the two stations record a decrease in

CO₂ because of the arrival of air masses depleted in CO₂ originating from the Indian Ocean 456 south of the Equator (Fig. 1a, Fig. 3c). Compared to PBL, the CO₂ decrease at PON is less 457 pronounced and longer, probably because of the influence of anthropogenic emissions in 458 South India. The CO₂ mean seasonal cycle at PON is also similar to that observed at CRI 459 (15.08°N, 73.83°E, 60m a.s.l.), another station on the southwest coast of India, yet the 460 seasonal maximum at CRI is reached slightly earlier than at PON in March (Bhattacharya et 461 462 al., 2009; Tiwari et al., 2011, 2014). The SNG station (18.35°N, 73.75°E, 1600m a.s.l.), located over the Western Ghats, observes a larger CO₂ seasonal cycle with a peak-to-peak 463 464 amplitude of ~20 ppm (Tiwari et al., 2014).

465

466 **3.1.2 CH**₄

Figure 4 presents the time series of CH₄ flask measurements at the three Indian stations and 467 the two NOAA/ESRL stations (Dlugokencky et al., 2014a), with their corresponding 468 smoothed curves for 2007–2011. At HLE, the annual mean CH₄ concentration increased from 469 470 1814.8±2.9 to 1849.5±5.2 ppb between 2007 and 2011 (Fig. 4, Table 1). The multiyear mean CH₄ value at HLE was lower than at KZM and WLG by on average 25.7±3.1 and 19.6±7.8 471 ppb (Fig. 4c and d, Table 1), respectively, reflecting the latitudinal and altitudinal CH₄ 472 473 gradients. Indeed, KZM and WLG receive air masses transported from Siberia with large wetland CH₄ emissions in summer, as well as those from regional sources closer to the 474 stations (Fang et al., 2013; Fig. S5), which may further contribute to the positive gradients 475 between these two stations and HLE. At PON and PBL, the annual mean CH₄ mole fractions 476 were higher than those at HLE by as much as 37.4±10.7 and 19.8±24.5 ppb respectively (Fig. 477 478 4a and b, Table 1). The positive gradients indicate significant regional CH_4 emissions, especially during winter when the NE monsoon transports air masses from East and 479 480 Northeast India and Southeast Asia, where emissions from livestock, rice paddies and a variety of waterlogged anaerobic sources and residential biofuel burning are high (EDGAR v4.2; Baker et al., 2012; Kurokawa et al., 2013). The in-situ measurements at Darjeeling, India (27.03°N, 88.25°E, 2194 m a.s.l.), another station located in the eastern Himalayas, also showed large variability and frequent pollution events in CH₄ mole fractions, which largely result from the transport of CH₄-polluted air masses from the densely populated Indo-Gangetic Plains to the station (Ganesan et al., 2013).

487

The CH₄ seasonal cycles exhibit contrasting patterns across stations. As shown in Fig. 5, a 488 489 distinct characteristic of the mean seasonal cycle at HLE is a CH₄ maximum from June to September. Even KZM and WLG do not show a minimum in summer that would be 490 characteristic for the enhanced CH₄ removal rate by reaction with OH. The pronounced HLE 491 feature is consistent with the result from the aircraft flask measurements over India at flight 492 altitudes of 8–12.5 km by the Civil Aircraft for the Regular Investigation of the atmosphere 493 494 Based on an Instrument Container (CARIBIC, http://www.caribic-atmospheric.com/) project (Schuck et al., 2010, 2012; Baker et al., 2012), although a larger seasonal cycle amplitude is 495 found in the CARIBIC composite data due to the rapid vertical mixing over the monsoon 496 497 region and the strong anticyclone that develops in the upper troposphere (Fig. 5a; Schuck et al., 2010). CARIBIC sampled the mid- to upper tropospheric air masses that were earlier and 498 more strongly enriched in CH₄, as a result of the rapid vertical transport of surface air masses 499 by deep convection and subsequent accumulation and confinement of pollutants within the 500 501 strong, closed circulation of the anticyclone (Li et al., 2005; Randel and Park, 2006). Xiong et 502 al. (2009) also reported enhancements of CH₄ during the summer monsoon season over South Asia based on satellite retrievals of CH₄ using the Atmospheric Infrared Sounder (AIRS) on 503 the EOS/Aqua platform as well as model simulations. Moreover, the mean CH₄ seasonal 504 cycle at HLE agrees well with the seasonal variations of CH₄ emissions from wetlands and 505

506 rice paddies and convective precipitation over the Indian subcontinent (Fig. 5b), suggesting that the summer maximum at HLE are likely related to the enhanced biogenic CH₄ emissions 507 from wetlands and rice paddies and deep convection that mixes surface emissions into the 508 509 mid-to-upper troposphere. During the SW monsoon period (June-September), convection over the Indian subcontinent and the Bay of Bengal rapidly mixes surface polluted air with 510 the upper troposphere, therefore concentrations of trace gases would be enhanced at higher 511 altitudes rather than at the surface (Schuck et al., 2010; Lawrence and Lelieveld, 2010). 512 Further analyses of carbon isotopic measurements and/or chemical transport model are 513 514 needed to disentangle and quantify the contributions of meteorology and biogenic emissions to the CH₄ summer maximum at HLE. As stated above, KZM and WLG also record CH₄ 515 increases during summertime, but with smaller magnitudes (Fig. 5a), possibly because they 516 517 are not directly influenced by deep convection from the Indian monsoon system.

518

519 In contrast to HLE, the CH₄ mean seasonal cycles at PON and PBL have distinct phases and much larger amplitudes, with minimum CH₄ values during July (Fig. 5c). These not only 520 reflect higher rates of removal by OH, but rather the influence of southern hemispheric air 521 522 transported at low altitudes from the southwest as well as the dilution effect by increased local planetary boundary layer height. In boreal winter, the maxima at PON and PBL are 523 associated with CH₄-enriched air masses transported from East and Northeast India, and 524 Southeast Asia, mostly polluted by agricultural-related sources (e.g., livestock, rice paddies, 525 agricultural waste burning). As PON and PBL, the flask measurements at CRI also showed 526 527 the seasonal maximum CH₄ values during the NE monsoon season, reflecting influences of air masses with elevated CH₄ from the Indian subcontinent (Bhattacharya et al., 2009; Tiwari 528 529 et al., 2013).

530

531 3.1.3 N₂O

Nitrous oxide (N₂O) is a potent greenhouse gas that has the third largest contribution to 532 anthropogenic radiative forcing after CO2 and CH4 (IPCC, 2013). It also becomes the 533 dominant ozone depleting substance (ODS) emitted in the 21st century with the decline of 534 535 chlorofluorocarbons (CFCs) under the Montreal Protocol (Ravishankara et al., 2009). Since the pre-industrial era, the atmospheric N₂O increased rapidly from \sim 270 ppb to \sim 325 ppb in 536 2011 (IPCC, 2013), largely as the result of human activities. Of the several known N₂O 537 538 sources, agricultural activities (mainly through nitrogen fertilizer use) contribute to ~58% of the global anthropogenic N_2O emissions, with a higher share in a predominantly agrarian 539 country like India (~75%; Garg et al., 2012). 540

541

The time series of N₂O flask measurements over the period of 2007–2011 and their smoothed 542 curves are presented in Fig. 6. At HLE, the annual mean N₂O concentration rose from 543 322.2±0.1 to 325.2±0.1 ppb during 2007–2011 (Table 1), with a mean annual growth rate of 544 0.8 ± 0.0 ppb yr⁻¹ (r² = 0.97, p = 0.001), smaller than that at MLO (1.0\pm0.0 ppb yr⁻¹, Table 1). 545 At PON and PBL, the annual mean N_2O mole fractions are higher than at HLE by 3.1 ± 0.3 546 and 3.8±1.7 ppb (Fig. 6, Table 1), respectively. The N₂O gradients between PON, PBL and 547 HLE are larger than typical N₂O gradients observed between stations scattered in Europe or 548 in North America. For example, Haszpra et al. (2008) presented N₂O flask measurements at a 549 continental station – Hegyhátsál, Hungary (HUN – 46.95 °N, 16.65 °W, 248 m a.s.l.) from 550 1997 to 2007. The annual mean N₂O mole fraction at HUN was higher than at Mace Head 551 (MHD) by only 1.3 ppb. We also analyzed N₂O time series of flask measurements during 552 553 2007–2011 at several European coastal stations – BGU in Spain, FIK in Greece, and LPO in France (Table S6), and the N₂O gradients between these stations and MHD were 1.1 ± 0.2 , 554 0.4±0.1, and 2.1±0.6 ppb, respectively (Fig. S10, Table S7). In the United States, N₂O flask 555

measurements from the NOAA/ESRL stations at Park Falls, Wisconsin (LEF - 45.95 °N, 556 90.27 °W, 472 m a.s.l.), Harvard Forest, Massachusetts (HFM – 42.54 °N, 72.17 °W, 340 m 557 a.s.l.) and a continental, high-altitude station at Niwot Ridge, Colorado (NWR - 40.05 °N, 558 105.58 °W, 3523 m a.s.l.) also show that, the annual mean N₂O concentrations at HFM and 559 LEF were higher than that at NWR by only 0.5±0.1 and 0.3±0.1 ppb, respectively (Fig. S10, 560 Table S7). Besides, the N₂O concentrations measured at PON and PBL have a notably higher 561 variability (around the smoothed fitting curve) than that at European and US stations (see 562 relative SDs (RSD) in Table 1 and Table S7). The larger N₂O gradient between PON, PBL 563 564 and HLE, as well as higher variability at PON and PBL, demonstrate the presence of substantial N₂O sources in South Asia and over the Indian Ocean during the observation 565 period. The in-situ measurements at Darjeeling also exhibited N₂O enhancements to be above 566 567 the background level, suggesting significant N₂O sources in this region (Ganesan et al., 2013). These sources may be related to emissions from natural and cultivated soils probably 568 enhanced by extensive use of nitrogen fertilizers, as well as emissions from regions of coastal 569 570 upwelling in the Arabian Sea (Bange et al., 2001; Garg et al., 2012; Saikawa et al., 2014).

571

572 Compared to CO₂ and CH₄, the seasonal cycle of N₂O is very small due to the long lifetime of ~120 years (Minschwaner et al., 1993; Volk et al., 1997), and has a larger uncertainty 573 probably because synoptic events are more likely to mask the seasonal signal. At HLE, PON 574 and PBL, the peak-to-peak amplitudes of the N₂O seasonal cycle are 0.6 ± 0.1 , 1.2 ± 0.5 , and 575 2.2±0.6 ppb, respectively (Table 1). HLE displays a N₂O maximum in mid-August (Student's 576 577 t-test, t=1.78, p=0.06), and a secondary maximum is in January/February but not significant (Student's t-test, t=-0.84, p=0.79) (Table 1, Fig. 7, Table S8 for detailed t-test statistics). The 578 N₂O seasonal cycle at HLE is out of phase with that at other northern background stations 579 such as MHD (Fig. S11, Table S7), where an N₂O summer minimum is always observed, 580

581 likely due to the downward transport of N2O-depleted air from the stratosphere to the troposphere during spring and summer (Liao et al., 2004; Morgan et al., 2004; Jiang et al., 582 2007b). The timing of the summer N₂O maximum at HLE is consistent with that of CH₄ 583 (Table 1; Figs. 5 and 7), giving evidence that the N₂O seasonal cycle may probably be 584 influenced by the convective mixing of surface air, rather than by the influx of stratospheric 585 air into the troposphere. Given that the populous Indo-Gangetic plains have high N₂O 586 emission rates due to the intensive use of nitrogen fertilizers (Garg et al., 2012; Thompson et 587 al., 2014a), during summer, the surface air enriched in N₂O is vertically transported by deep 588 589 convection and enhances N₂O mole fractions in the mid-to-upper troposphere. Like CH₄, the N₂O enhancement during the summer monsoon period (June-September) was also observed 590 by the aircraft flask measurements at flight altitudes 8–12.5 km from the CARIBIC project in 591 592 2008 (Schuck et al., 2010).

593

594 At PON, N₂O also decreases during February–April and reaches a minimum at the end of May. However, the decrease of N₂O does not persist during June-September, which is in 595 contrast with CH₄ (Table 1, Fig. 7a). One reason may be that the air masses arriving at the 596 597 site during the southwest monsoon period is relatively enriched in N₂O compared to CH₄, reflecting differences in their relative emissions along the air mass route. The increase of N₂O 598 at PON during June-August and the maximum during September-October are likely related 599 to N₂O emissions from coastal upwelling along the southern Indian continental shelf, which 600 peak during the SW monsoon season (Patra et al., 1999; Bange et al., 2001). According to 601 602 Bange et al. (2001), the annual N₂O emission for the Arabian Sea is 0.33–0.70 Tg/yr, of which N_2O emissions during the SW monsoon account for about 64–70%. This coastal 603 upwelling N₂O flux is significantly larger than the annual anthropogenic N₂O emissions in 604 South India south of 15 °N, which is estimated to be on average 0.07–0.08 Tg/yr during 605

606 2000–2010 (EDGAR v4.2). At PBL, the maximum and minimum N₂O occur in November and February/March, respectively (Table 1, Fig. 7b). The late N₂O peak at PBL in November 607 may be associated with the N₂O-enriched air masses transported from South and Southeast 608 609 Asia, which could be attributed to natural and agricultural N₂O emissions from this region (Saikawa et al., 2014). It should be noted that, the mean seasonal cycles of N₂O at PON and 610 PBL are subject to high uncertainties because of the short observation periods and data gaps 611 (shaded area in Fig. 7). The N₂O maximum and/or minimum obtained from the mean 612 seasonal cycle are marginally significant for PON and PBL (Table S8 for detailed t-test 613 614 statistics). Therefore, caution should be exercised in interpreting mean seasonal cycles at these stations. Sustained, long-term measurements are needed in order to generate more 615 reliable estimates of the seasonal cycles for the two stations. 616

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618 3.1.4 SF₆

Sulfur hexafluoride (SF_6) is an extremely stable greenhouse gas, with an atmospheric lifetime 619 620 as long as 800–3200 year and a global warming potential (GWP) of ~23,900 over a 100-year time horizon (Ravishankara et al., 1993; Morris et al., 1995; IPCC, 2013). The main sources 621 of atmospheric SF₆ emissions are electricity distribution systems, magnesium production, and 622 semi-conductor manufacturing (Olivier et al., 2005), while its natural sources are negligible 623 (Busenberg and Plummer, 2000). As its sources are almost purely anthropogenic (Maiss et al., 624 1996), SF_6 is widely considered as a good tracer for population density, energy consumption 625 and anthropogenic GHG emissions (Haszpra et al., 2008). 626

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Figure 8 presents the time series of SF_6 flask measurements and corresponding fitting curves at HLE, PON, and PBL. At HLE, the annual mean SF_6 mole fractions increased from 630 6.26 ± 0.03 to 7.38 ± 0.01 ppt between 2007 and 2011, which is in good agreement with the SF₆ trend observed at MLO during the same period (HLE: 0.29 ± 0.05 ppt/yr, $r^2=0.99$, p<0.001; 631 MLO: 0.29 ± 0.03 ppt/yr, r²=0.99, p<0.001; Figs. 8 and S12a, Table 1, Table S9). The annual 632 mean SF₆ gradient between PON and HLE is -0.060 ± 0.030 ppt, whereas the gradient between 633 PBL and HLE is statistically insignificant (-0.002±0.097 ppt). The slight negative gradient 634 between PON and HLE is a reversed signal compared with the SF₆ observations at stations 635 influenced by continental emissions in Europe and United States. For example, the SF₆ mole 636 fractions at HUN over the years of 1997–2007 are higher than those at MHD by on average 637 638 0.19 ppt (Haszpra et al., 2008). We also analyzed the SF₆ gradients between two coastal European stations - BGU (41.97 °N, 3.3 °E, 30 m a.s.l.) and LPO (48.80 °N, 3.57 °W, 30 m 639 a.s.l.) - and MHD, which are 0.10±0.03 and 0.05±0.02 ppt averaged over the period of 2007-640 641 2011, respectively. At HFM, the SF₆ mole fractions are higher than those of the NWR on average by 0.15±0.06 ppt during 2007–2011 (Table S9). Given the long atmospheric lifetime 642 of SF₆, the positive gradients between continental European and US stations and background 643 reference stations suggest significant sources in Europe and the US. On the contrary, the 644 slight negative gradient between PON and HLE implies weak SF₆ emissions over the Indian 645 subcontinent, which is also indicated by recent high-frequency in-situ SF_6 measurements at 646 Darjeeling (Ganesan et al., 2013). It is also worthwhile to note that high SF₆ values occur 647 repeatedly at HLE and PBL in winter, which is likely related to episodic SF₆ pollution events 648 649 from the Middle East, South/Southeast Asia and China (Figs. 8b and S7d).

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The annual mean SF_6 seasonal cycles for HLE, PON, and PBL are presented in Fig. 9. The peak-to-peak amplitudes at the three stations are 0.15 ± 0.03 , 0.24 ± 0.02 , and 0.48 ± 0.07 ppt, respectively (Table 1). At HLE, the SF_6 seasonal cycle is bimodal as for N₂O, with an absolute maximum occurring in November (Student's t-test, t=2.425, p=0.014) and a

secondary maximum in May (Student's t-test, t=2.443, p=0.016) (Table S10 for detailed t-test 655 statistics). Given that SF_6 increases monotonously and that its sources are purely 656 anthropogenic and not subject to seasonally variations (Maiss et al., 1996), the seasonal cycle 657 of SF₆ should be driven by changes in atmospheric circulations, e.g., the SW monsoon 658 convection and stratosphere-atmosphere exchange (Levin et al., 2002). We note that, at HLE, 659 no enhancement of SF₆ during the SW monsoon season is recorded, unlike what is observed 660 for CH₄ and N₂O (Figs. 5 and 7). Although the CARIBIC aircraft flask measurements over 661 the Indian region demonstrated SF_6 enhancements in the upper troposphere at ~30 °N 662 663 (approximately where HLE is located) in August, 2008, they are not related to the deep convection and surface sources that contribute to the summer maxima in CH₄ and N₂O. Back-664 trajectories from the CARIBIC flights showed that the summer enhancements in SF₆ were 665 666 related to air samples collected north of 20°N along the flight routes, where air masses were more influenced by the westerly subtropical jet (and a smaller anticyclone embedded in it 667 over the Arabian Peninsula) rather than the deep convection in the monsoon region 668 (Krishnamurti et al. 2008; Schuck et al., 2010; Fig. S9). Since HLE is not influenced by the 669 westerly subtropical jet in the upper troposphere (also clearly seen by the colors of back-670 trajectories in Fig. S9), the summer enhancements of SF₆ observed by the CARIBIC flights 671 are not detected by the flask measurements at HLE. The absence of SF₆ enhancement in 672 summer at HLE confirms weak SF₆ emissions in India. At PBL, the SF₆ seasonal cycle is 673 674 related to the monsoon circulation and convection (Figs. 9b and S7d). The maximum during November–December (Student's t-test, t=5.138, p<0.001; Table S10) is likely due to frequent 675 episodic SF₆ polluted air masses transported from Southeast Asia and China (Fig. S7d). 676

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678 **3.1.5 CO**

679 Carbon monoxide (CO) plays important roles in atmospheric chemistry, as the dominant sink for the hydroxyl radical (OH, the main tropospheric oxidant) and a precursor of tropospheric 680 ozone under high NO_x (NO+NO₂) concentrations (Logan et al., 1981; Novelli et al., 1998; 681 682 Seinfeld and Pandis, 2006). Although CO does not act as a greenhouse gas, it modulates the atmospheric concentrations of CH₄ (the second anthropogenic greenhouse gas after CO₂) 683 through competition for the OH radicals. At the global scale, it contributes to an indirect 684 positive radiative forcing of 0.23 ± 0.07 Wm⁻² (IPCC, 2013). Besides, CO is an excellent tracer 685 for combustion processes, with emission sources mainly contributed by incomplete 686 687 combustion of fossil fuel and biofuels, and by biomass burning (Granier et al., 2011). In India, biofuel and agricultural waste burning account for 70-80% of the total anthropogenic CO 688 emissions (EDGAR v4.2; Streets et al., 2003b; Yevich and Logan, 2003). 689

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The time series of CO flask measurements and corresponding smoothed curves are shown in 691 692 Fig. 10. Over the period of 2007–2011, HLE recorded a slight decrease in CO mole fractions from 104.7±1.4 to 99.4±2.2 ppb, with an annual rate of -2.2±0.0 ppb yr⁻¹ ($r^2 = 0.65$, p = 0.06). 693 The CO mole fractions at HLE are lower than those at KZM and WLG (Novelli et al., 2014b), 694 by on average 18.8±2.5 and 30.2±7.4 ppb, respectively (Table 1, Fig. 10c and d). The 695 positive gradient between KZM, WLG and HLE does not only reflect decreasing CO with 696 altitude and the N-S global gradient, but also suggests differences in regional emission 697 sources. For example, compared to HLE, the CO signals at WLG are more influenced by 698 transport of polluted air, especially during summer when about 30% air masses pass over 699 700 industrialized and urbanized areas southeast of the station (Zhang et al., 2011). Besides, the positive CO gradient between KZM, WLG and HLE may be further contributed by air 701 masses of northern Siberia origin in summer (Fig. S5), with higher CO emissions from 702 biomass burning and secondary CO from the oxidation of CH₄ and non-CH₄ hydrocarbons 703

704 (Konovalov et al., 2014). At PON and PBL, the annual mean CO mole fractions are higher 705 than that at HLE by on average 82.4±10.7 and 52.5±8.5 ppb, respectively (Table 1, Fig. 10a and b). The PON and PBL stations are influenced by CO regional emissions, mainly due to 706 707 biofuel and agricultural burning over South and Southeast Asia (Lelieveld et al., 2001; Streets et al., 2003a, b; Yevich and Logan, 2003). We also note that, for all the five stations, the CO 708 time series show larger variability with respect to their corresponding smoothed curves than 709 other species do (see the residual SD (RSD) in Table 1, Fig. 10), as a result of the unevenly 710 distributed CO sources and short atmospheric lifetime (Novelli et al., 1992). 711

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As shown in Fig. 11, the CO seasonal cycle at HLE reaches a maximum in mid-March and a 713 minimum by the end of October, with a peak-to-peak amplitude of 28.4±2.3 ppb (Table 1, 714 Fig. 11). The phase of the mean CO seasonal cycle at HLE generally agrees with the ones 715 observed at KZM and WLG, with a lag of up to 1 month in the timing of seasonal minimum 716 717 at the two stations (Table 1, Fig. 11c and d). In contrast with the three stations representative of large-scale free tropospheric air masses, the stations at the maritime boundary layer in the 718 mid-to-high Northern Hemisphere observe the lowest CO values in July or August (Novelli et 719 720 al., 1992, 1998), when the concentration of OH – the major sink of CO – is highest (Logan et al., 1981). The delay in timing of the seasonal CO minimum at the three free troposphere 721 stations in Central and South Asia compared to those boundary layer stations is probably due 722 to the mixing time of regional surface CO emissions and the relatively short lifetime of CO 723 (1-2 months on average). During summer, KZM and WLG sample air masses from Siberia 724 725 impacted by CO fire emissions (Duncan et al., 2003; Kasischke et al., 2005), as well as COpolluted air from urbanized and industrialized area (Zhang et al., 2011), while HLE is 726 influenced by convective mixing of CO emissions from India, either from anthropogenic 727 728 sources or oxidation of VOCs. It is interesting to note that the CO seasonal cycle at HLE does

729 not show an enhancement during JAS as CH₄ and N₂O do (Figs. 5 and 7), possibly as a result of OH oxidation that reduces CO and acts oppositely to vertical transport, and/or differences 730 in seasonal emission patterns between CO and the other two species (Baker et al., 2012). 731 732 However, the CO enhancement during summer was observed in the upper troposphere over South Asia from the CARIBIC aircraft measurements at flight altitudes 8-12.5 km and 733 Microwave Limb Sounder observations at 100–200 hPa (Li et al., 2005; Jiang et al., 2007a; 734 735 Schuck et al., 2010). The differences in the CO seasonal cycles at different altitudes suggest faster transport (and younger air masses) at 10 km than at 5 km due to convection, controlling 736 737 the vertical profile of CO, which makes it difficult to directly compare aircraft measurements in the upper troposphere and column remote sensing observations with surface data. 738

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At PON and PBL, the mean CO seasonal cycles show maxima in the boreal winter and 740 minima in the boreal summer, with peak-to-peak amplitudes of 78.2±11.6 and 144.1±16.0 741 742 ppb, respectively (Fig. 11a and b). A strong and positive correlation is found between detrended CO and CH₄ at PON (r=0.70, p<0.001) and PBL (r=0.84, p<0.001), suggesting that 743 the seasonal cycles of both species are dominated by the seasonally varying atmospheric 744 745 transport. During summer when the southwest monsoon prevails, the surface CO concentrations at PON and PBL are low due to rapid convective uplifting and advection of 746 clean air masses from the ocean. During winter, the two stations are influenced by 747 northeasterly air masses enriched in CO from Northeast India, Southeast Asia and China 748 (back-trajectories in Fig. S7e), probably influenced by biofuel and agricultural waste burning 749 750 in these regions (Yevich and Logan, 2003; Lelieveld et al., 2001).

751

752 **3.1.6 H**₂

Hydrogen (H₂) is the second most abundant reduced trace gas in the troposphere after CH₄, with an average mole fraction of ~530 ppb (Novelli et al., 1999). It plays important roles in tropospheric and stratospheric chemistry and indirectly impacts budgets of CH₄, CO and nonmethane hydrocarbons (NMHCs) through reaction with the OH radicals (Novelli et al., 1999; Ehhalt and Rohrer, 2009). Like CO, H₂ is also a good tracer for incomplete combustion emissions from fossil fuel and biomass/biofuel burning, which is quite extensive in India (Streets et al., 2003b; Yevich and Logan, 2003).

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761 Figure 12 shows the time series of H₂ flask measurements with smoothed curves at HLE, PON, and PBL, respectively. No significant trend was observed at any of the three stations 762 (Table 1, Fig. 12), consistent with the long-term H₂ measurements at other background 763 stations during the last three decades (Novelli et al., 1999; Ehhalt and Rohrer, 2009; Grant et 764 al., 2010). For the year 2008, comparing to KZM and WLG (Novelli et al., 2014a), HLE 765 766 recorded higher H₂ mole fractions by ~40 ppb, reflecting the latitudinal gradient of H₂ with lower concentrations towards northern high latitudes, due to land uptake by soils (Novelli et 767 al., 1999; Price et al., 2007; Hauglustaine and Ehhalt, 2002; Ehhalt and Rohrer, 2009). Note 768 769 that these results based on only one-year comparison need to be confirmed by extended data more up-to-date, which are not available yet. At PON and PBL, the annual mean H₂ mole 770 fractions were higher than at HLE by 29.8±4.1 and 21.8±4.6 ppb, respectively (Table 1; Fig. 771 12). Comparisons with H₂ measurements at Mariana Island, Guam (GMI - 13.39 °N, 772 144.66 °E, 0.00 m a.s.l.) (Novelli et al., 2014a), another maritime station in the western 773 774 Pacific at a similar latitude as PON and PBL, also showed positive gradients of ~40 ppb (Fig. S13c and d; Table S11), suggesting substantial regional H₂ sources over the footprint area of 775 776 PBL and PON. During October-March when the NE monsoon prevails, both PON and PBL receive H₂-enriched air masses from South and Southeast Asia, mainly influenced by fossil 777

fuel combustion and biomass burning (Fig. S7f; GFED v3.1; Hauglustaine and Ehhalt, 2002; Price et al., 2007; Ehhalt and Rohrer, 2009; van der Werf et al., 2010). During April– September, with the northward movement of Intertropical Convergence Zone (ITCZ), the two stations are influenced by advection of air from south of the Equator. For PON, H₂-polluted air masses are occasionally sampled during JAS when the SW monsoon moves over the continent of South India with high population and heavy industry (Fig. S7f; Census India, 2011).

785

786 The mean H₂ seasonal cycles for HLE, PON, and PBL are presented in Fig. 13. At HLE, the peak-to-peak H₂ seasonal amplitude is 15.8±2.2 ppb, less than half of the seasonal amplitudes 787 at BMW (39.6±2.6 ppb) and MID (38.0±2.4 ppb) of similar latitudes (Novelli et al., 2014a), 788 and that at WLG (22.8±3.0 ppb) (Figs. 13d and S14a, Tables 1 and S11). The maximum and 789 minimum of H₂ occur in April and September, respectively. The dampening of the H₂ 790 791 seasonal amplitude with increasing altitude was previously found for another high-altitude continental station at Jungfraujoch, Switzerland (JUN – 46.53 °N, 7.98 °E, 3580.00 m a.s.l.) 792 (Bond et al., 2011), and was also captured by the GEOS-Chem global chemical transport 793 model (Price et al., 2007). Since the soil sink dominates much of the surface H₂ seasonal 794 cycle in the mid-to-high Northern Hemisphere (Hauglustaine and Ehhalt, 2002; Price et al., 795 2007; Bousquet et al., 2011; Yver et al., 2011; Yashiro et al., 2011), the smaller amplitude in 796 the H₂ seasonal cycle at HLE may be attributed to the weakened soil sink with increasing 797 altitude due to vertical mixing (Price et al., 2007; Bond et al., 2011). 798

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At PON and PBL, the mean H_2 seasonal cycles are characterized by the peak-to-peak amplitudes of 21.6±3.4 and 21.3±5.0 ppb respectively, comparable to that at GMI (21.5±1.2

ppb) (Tables 1 and S11, Figs. 13a and b and S14b). At PBL, the H₂ maximum in March– April and a secondary increase during September–October coincide with the double biomass burning peaks in each hemisphere – in March for northern tropics, in August/September for southern tropics (van der Werf et al., 2006; Price et al., 2007; Bousquet et al., 2011; Yver et al., 2011). Given that the seasonal variation of soil H₂ uptake is probably small in the tropics (Price et al., 2007; Bousquet et al., 2011; Yver et al., 2011; Yashiro et al., 2011), this bimodal H₂ seasonal cycle at PBL could be related to biomass burning.

809

810 **3.2 Synoptic variations**

811 In this section we analyze synoptic variations of CO₂, CH₄, and CO by examining correlations between species, after subtracting the smoothed curve from the original data. 812 Ratios of trace gas mole fractions or their enhancements have been widely used in previous 813 studies to partition contributions from different source types and origins (Langenfelds et al., 814 2002; Paris et al., 2008, Lopez et al., 2012), to estimate emissions of one species given 815 816 emissions of another one that is better-known (Gamnitzer et al., 2006; Rivier et al., 2006; Turnbull et al., 2006; Schuck et al., 2010), and to provide valuable constraints on inversion of 817 sources and sinks of trace gases (Xiao et al., 2004; Pison et al., 2009). 818

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820 **3.2.1** ΔCH₄/ΔCO

Figure 14 shows scatterplots of CH₄ and CO residuals with the orthogonal distance regression lines at HLE, PON, and PBL for different seasons. A significant and positive correlation between CH₄ and CO residuals (hereafter Δ CH₄/ Δ CO, unit ppb ppb⁻¹) is found for all three stations throughout the year. Furthermore, the Δ CH₄/ Δ CO ratio also shows seasonal variation at each of the three stations. The most prominent feature is the occurrence of maximum 826 slopes in July-September (also October-December at PON), especially at HLE and the generally higher ratios at this station. Wada et al. (2011) and Niwa et al. (2014) also reported 827 increased summer $\Delta CH_4/\Delta CO$ over the western North Pacific, according to the in-situ 828 829 measurements at several surface stations and aircraft flask measurements in the midtroposphere. The main process for this seasonal variation of $\Delta CH_4/\Delta CO$ might be the 830 enhanced emissions of biogenic CH₄ in summer (e.g., wetland and rice paddy emissions; 831 Streets et al., 2003a; Yan et al., 2003) combined with concurrent lower anthropogenic CO 832 emissions in summer than in winter (due to less residential fuel use for heating, see Streets et 833 834 al., 2003a). The faster photochemical destruction of CO by increased OH during summer cannot explain such large changes (less than 15% according to Wada et al. (2011)). 835

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At HLE, the $\Delta CH_4/\Delta CO$ ratio varies from 1.2±0.3 to 4.0±1.2 ppb ppb⁻¹ throughout the year, 837 with a maximum in JAS, corresponding to the summer monsoon season (Fig. 14a-d). Based 838 839 on the CARIBIC flights between 10 and 12 km from Frankfurt, Germany to Chennai, India, Baker et al. (2012) derived a $\Delta CH_4/\Delta CO$ ratio in the range 1.88(±0.22) to 4.43(±0.56) in JAS 840 over South Asia. The maximum $\Delta CH_4/\Delta CO$ observed during summer in the mid-to-upper 841 842 troposphere may be the result of higher biogenic CH₄ emission over the Indian subcontinent, lower CO emissions, combined with frequent widespread convective uplift of surface air 843 during the SW monsoon (Schuck et al., 2010; Baker et al., 2012). The CARIBIC flights 844 recorded similar $\Delta CH_4/\Delta CO$ values to HLE, confirming that convection plays a dominant 845 role compared to advection during the SW monsoon season. Outside the SW monsoon season, 846 847 both the CARIBIC flights and HLE do generally not record strong effects of surface emissions due to the weakened vertical transport. With respect to the $\Delta CH_4/\Delta CO$ ratios for 848 January-March, April-June and October-December, our estimates are 1.5 to 4 times that of 849 850 the ratios determined for air masses with signatures of fossil fuel combustion, according to

851 several aircraft and ground observations in East and Southeast Asia (Table S12; Sawa et al., 2004; Lai et al., 2010; Wada et al., 2011; Niwa et al., 2014), which rules out fossil fuel 852 combustion as an explanation for the higher ratios. Our ratios are comparable to the 853 854 $\Delta CH_4/\Delta CO$ values inferred for air masses of Siberian origin during winter (Table S12; Harris et al., 2000; Chi et al., 2013), and we also obtain similar estimates of $\Delta CH_4/\Delta CO$ from the 855 flask measurements at KZM over the study period (The $\Delta CH_4/\Delta CO$ ratios for KZM are 856 0.8±0.2, 1.7±0.2 and 1.5±0.3 ppb ppb⁻¹ for AMJ, OND and JFM, respectively), which are 857 influenced by air masses originating from North Africa, the Middle East, and Central Asia as 858 859 seen at HLE (see back-trajectories in Fig. S5). Given that oil and gas production accounts for 50-70% of CH₄ emissions in these regions (EDGAR v4.2) and that over dry areas the 860 daytime boundary layer is higher which favors injection of surface emissions into the 861 862 troposphere, the preferential enrichment in CH₄ relative to CO at HLE may tentatively be attributed to fossil CH₄ emissions over gas extraction regions and transported eastwards by 863 westerlies (Harris et al., 2000; Tohjima et al., 1996). 864

865

At PON and PBL, the $\Delta CH_4/\Delta CO$ ratios are in general considerably higher than 0.3 for all 866 867 seasons, putting them in the range of ratios indicative of urban/industrial sources (Table S12; Harriss et al., 1994; Sawa et al., 2004; Xiao et al., 2004; Bakwin et al., 1995; Lai et al., 2010; 868 Wada et al., 2011; Niwa et al., 2014). However, this does not rule out contributions from 869 biomass/biofuel burning with emissions having a typical $\Delta CH_4/\Delta CO$ ratio less than 0.3 870 (Mauzerall et al., 1998; Andreae and Merlet, 2001; Mühle et al., 2002). Considering that 871 872 biofuel and agriculture waste burning are the primary energy sources in rural India (Streets et al., 2003a; Yevich and Logan, 2003; Venkataraman et al., 2005), CO emissions from biofuel 873 burning must be substantial (Lelieveld et al., 2001). This is the case for NE India located 874 upwind of PON and PBL when the NE monsoon prevails during December-March. 875
Nevertheless, the relatively low $\Delta CH_4/\Delta CO$ derived from biomass/biofuel burning could be increased by CH₄ emissions from livestock with similarly distributed sources (EDGAR v4.2). Emissions of both trace gases from livestock and biomass/biofuel burning in the Indian subcontinent compiled by EDGAR v4.2 also indicate a CH₄ to CO ratio of 0.64–0.69 over the period of 2000–2008, close to the atmospheric measurements of $\Delta CH_4/\Delta CO$ at PON and PBL during JFM (Fig. 14h and l).

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883 **3.2.2** ΔCH₄/ΔCO₂

The $\Delta CH_4/\Delta CO_2$ ratios are strongly influenced by the high variability of CO₂ and the 884 885 interpretation is complex. Unlike the positive correlation between CH₄ and CO consistently observed at all three stations, the relationships between CH4 and CO2 residuals exhibit 886 scattered and differences in the residual slopes for different stations and seasons (Fig. 15). At 887 HLE, no significant correlations are found during AMJ, JAS, and OND (Fig. 15a-c), because 888 CH₄ and CO₂ have distinct biogenic and/or photochemical sources and sinks over the 889 northern mid-latitudes. During JFM when biogenic CO₂ fluxes and anthropogenic emissions 890 are positive to the atmosphere, there is a significant and positive relationship between CH₄ 891 and CO₂, with a Δ CH₄/ Δ CO₂ ratio of 45.6±1846.8 ppb ppm⁻¹ (r=0.37, p=0.03; Fig. 15d). This 892 value is close to the ratio of CH₄ and CO₂ anthropogenic emissions over North Africa (39.1– 893 46.2 mmol mol⁻¹), Central Asia (44.4–49.5 mmol mol⁻¹) and to a lesser degree the Middle 894 East (25.8–28.4 mmol mol⁻¹) during the period of 2000–2010 (EDGAR v4.2), corresponding 895 to the back-trajectories reaching HLE (Fig. 1a). It should be noted that this estimate of 896 $\Delta CH_4/\Delta CO_2$ is subject to large uncertainty according to the standard deviation calculated 897 with 1000 bootstrap replications (Fig. 15d), implying that CH₄ and CO₂ sources of various 898 types and origins influence the HLE records. 899

At PON, in contrast to HLE, positive correlations occur between CH₄ and CO₂ residuals for 901 all seasons except OND, with a $\Delta CH_4/\Delta CO_2$ ratio of 6.7±2.4 ppb ppm⁻¹ (r=0.72, p<0.001) in 902 AMJ and 8.5 ± 0.9 ppb ppm⁻¹ in JAS (r=0.74, p<0.001), respectively (Fig. 15e and f). The 903 904 relatively narrow ranges of slopes compared to that for HLE and PBL likely suggest colocated urban and industrial sources in South India upwind of PON during April-September 905 (see back-trajectories in Fig. 1a). Emissions from biofuel burning could be a common source 906 907 for both CH₄ and CO₂, given the substantial biofuel use in South India (Yevich and Logan, 2003) and the biofuel burning emission ratio of CH₄ and CO₂ derived from previous studies 908 (5–10 mmol mol⁻¹; Andreae and Merlet, 2001). Note that the CARIBIC flask measurements 909 over India south of 20°N indicate a negative correlation between CH₄ and CO₂ at the altitudes 910 of 10-12 km during July-September, 2008 (r=-0.80, p=0.002; Fig. S15a), interpreted as the 911 912 concurrent strong uptake of CO₂ with enhanced emissions of CH₄ during the SW monsoon. During JFM when the NE monsoon predominates, CH₄ is positively correlated with CO₂ with 913 a $\Delta CH_4/\Delta CO_2$ ratio of 31.9±1635.7 ppb ppm⁻¹ (r=0.45, p=0.02; Fig. 15h). Like at HLE, this 914 ratio is subject to large uncertainty due to variability in CH₄ and CO₂ sources. The ratio based 915 on the CARIBIC observations in the upper troposphere (10-12 km) is 23.5±41.4 ppb ppm⁻¹ 916 (r=0.67, p=0.004; Fig. S15b). The inconsistency of the $\Delta CH_4/\Delta CO_2$ ratios estimated from the 917 two datasets suggest that the flask measurements at the surface station PON do provide 918 information more specific for constraining estimates of regional CH₄ and CO₂ fluxes. 919

920

Finally, at PBL, the prominent feature of the CH_4 – CO_2 relationship is the significant and negative correlation observed during JAS, with a $\Delta CH_4/\Delta CO_2$ ratio of -14.6±16.4 ppb ppm⁻¹ (r=-0.73, p=0.007; Fig. 15j). Since the time series of flask measurements at PBL is relatively short and has large data gaps (Fig. S3), correlations between trace gases could be influenced by abnormal pollution events. For example, excluding the event with CH_4 residuals > +20 ppb (corresponding to the observation at PBL on 16 September 2009, the point marked with black circle in Fig. 15j) would substantially decrease the strength of negative correlation between CH_4 and CO_2 (r=-0.54, p=0.09). We will investigate the CH_4 enriched event further in Sect. 3.3.

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931 **3.2.3** ΔCO/ΔCO₂

As shown in Fig. 16, at HLE, CO is positively correlated with CO₂ during AMJ, with a 932 $\Delta CO/\Delta CO_2$ ratio of 35.8±12.1 ppb ppm⁻¹ (r=0.53, p=0.001; Fig. 16a). During JFM, there is 933 934 no significant relationship between CO and CO₂ (r=0.15, p=0.39; Fig. 16d). However, excluding an abnormal event with $\Delta CO_2 = -1.8$ ppm on 8 January 2007 (the point marked 935 with black circle in Fig. 16d) would give a significant and positive correlation between CO 936 and CO₂, with a Δ CO/ Δ CO₂ ratio of 55.7±259.1 ppb ppm⁻¹ (r=0.40, p=0.02; the red solid line 937 in Fig. 16d). This ratio is less than half the emission ratio of CO to CO₂ from forest/grassland 938 939 biomass burning (Mauzerall et al., 1998; Andreae and Merlet, 2001), but higher than ratios of anthropogenic combustion sources in developed countries that are typically in the range of 940 10–15 ppb ppm⁻¹ (e.g., Suntharalingam et al., 2004; Wada et al., 2011; Takegawa et al., 2004). 941 942 This could be attributed not only to the lower combustion efficiency of fuels in North Africa, the Middle East, and Central Asia where air masses at HLE originate from, but also to 943 additional contribution from biofuel burning with relatively high CO to CO₂ emission ratios 944 (e.g., fuelwood, charcoal, agricultural residuals; Andreae and Merlet, 2001). Besides, the 945 relatively high $\Delta CO/\Delta CO_2$ in JFM compared to AMJ may further indicate a contribution of 946 947 CO emissions from residential biofuel burning in winter (Wada et al., 2011), especially in developing countries within the footprint area. 948

At PON, a positive and significant correlation between CO and CO₂ is found during AMJ, 950 with a $\Delta CO/\Delta CO_2$ ratio of 13.4±76.8 ppb ppm⁻¹ (r=0.46, p=0.03; Fig. 16e). This ratio is 951 similar to the ratios determined for air masses influenced by both fossil fuel emissions and 952 953 biomass/biofuel burning during the same seasons. For example, based on the in-situ measurements in the upper troposphere during the CARIBIC flights between South China 954 and Philippines in April 2007, Lai et al. (2010) reported the $\Delta CO/\Delta CO_2$ ratios of 15.6–29.3 955 ppb ppm⁻¹ during pollution events influenced by both biomass/biofuel burning and fossil fuel 956 combustion in Indochinese Peninsula. At PBL, CO is significantly and negatively correlated 957 with CO₂ during JAS (r=-0.68, p=0.01; Fig. 16j). However, we note that the CH₄ abnormal 958 event discussed in Sect. 3.2.2 is enriched in CO as well, and the negative relationship 959 between CO and CO₂ would no longer exist if we removed the event (r=-0.45, p=0.16). The 960 simultaneous enhancement of CO and CH₄ may suggest possible influences of biomass 961 burning episodes, which we will explore in detail in Sect. 3.3. During JFM, no significant 962 relationship is found between CO and CO₂ for PON or PBL (Fig. 16h and l). 963

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965 **3.3 Elevated CH₄ and CO events at PBL**

In this section, we discuss two elevated CH_4 and CO events at PBL during the SW monsoon season. Significant enhancements of CH_4 and CO were observed on September 16, 2009 (July 29, 2011), with residuals from smoothed curves as high as 34.2 (29.2) ppb and 36.2 (17.9) ppb for CH_4 and CO, respectively. We further analyzed CH_4 and CO measurements at Bukit Kototabang (BKT – 0.20 °S, 100.32 °E, 845.00 m a.s.l.), Indonesia, located upwind of PBL when the southwest monsoon prevails. The flask measurements at BKT detected enhanced CH_4 and CO with a magnitude of 38.0 and 66.1 ppb on September 8, 2009, about 973 one week before the occurrence of the first CH₄ and CO event at PBL (Fig. 17a). The in-situ measurements at BKT also showed CH₄ and CO enhancements about one week before the 974 second event at PBL, lasting over the period of 17 July-21 July 2011 (Fig. 17b). The 975 976 coincidence of the two abnormal CH₄ and CO events at PBL and BKT possibly suggests influences of polluted air masses with common sources and origins. Moreover, the fire 977 radiative power (FRP, mWm^{-2}) during the sampling dates implies that the two abnormal CH₄ 978 and CO events could be related to fire emissions in Indonesia (GFAS product version 1.0; 979 Kaiser et al. 2012; Fig. S16). Note that the mechanisms we propose for the abnormal CH_4 and 980 981 CO events and the possible linkage between PBL and BKT during the SW monsoon season are still speculative. Model experiments are needed to further confirm these hypotheses. 982

983

984 **4 Conclusions**

In this paper we present the results of flask measurements of CO₂, CH₄, N₂O, SF₆, CO, and 985 H₂ at three stations in India: Hanle (HLE), Pondicherry (PON) and Port-Blair (PBL), over the 986 987 period of 2007–2011. Of these three stations, HLE is located at a high altitude and regarded as a continental background station in the mid-latitude of the Northern Hemisphere; PON is a 988 tropical surface station located on the southwest coast of India, while PBL is an oceanic 989 990 station located on the Andaman Islands, of similar latitude to PON. With a total of 188, 185, and 63 flask pair samples collected respectively from HLE, PON and PBL between 2007 and 991 2011 (for PBL between 2009 and 2011), and analyzed at LSCE, the program represents an 992 important logistical and analytical effort to produce a unique dataset of atmospheric trace gas 993 observations over the Indian subcontinent. The observed records will serve as an important 994 995 source of information to infer regional patterns of trace gas fluxes and atmospheric transport in this under-documented region. Several conclusions and implications are drawn from the 996 997 first analyses of the datasets.

999 The annual gradients of the atmospheric mole fractions observed at PON and PBL, with respect to HLE as a reference, suggest significant emission sources of CO₂, CH₄, N₂O, CO, 1000 and H₂ over the footprints of those stations. In particular, the annual mean N₂O mole 1001 1002 fractions at PON and PBL are higher than at HLE by 3.1±0.3 and 3.8±1.7 ppb, notably larger than the typical N₂O gradients observed between stations in Europe or North America, 1003 1004 indicating substantial N₂O emissions. The analyses of the atmospheric mole fractions with 1005 back-trajectories at the three stations further confirmed emission sources from South and NE India, and SE Asia, all of which are populous with high demand for food and energy, and 1006 thus high emissions from industrial, residential, and/or agricultural sectors. On the other hand, 1007 despite of substantial anthropogenic GHG emissions in India (whether based on national 1008 inventories or atmospheric observations), unlike the USA and EU countries, its SF₆ emissions 1009 1010 are rather weak.

1011

1012 The seasonal cycles for each trace gas reflect not only the seasonal variations of natural sources/sinks and anthropogenic emissions over the Indian subcontinent, but also the 1013 seasonally varying atmospheric transport, especially the monsoon circulations (including 1014 1015 convection). Strong influences of the monsoon circulations are well depicted by the contrasting phases of CH₄ seasonal cycles between HLE and PON/PBL. At HLE, the distinct 1016 CH₄ maximum during June-September is likely related to the enhanced biogenic CH₄ 1017 emissions from wetlands and rice paddies in summer, combined with deep convection 1018 associated with the SW monsoon that mixes surface emissions into the mid-to-upper 1019 1020 troposphere. By contrast, the CH₄ seasonal cycles at PON and PBL have seasonal minima during the SW monsoon season, reflecting influences of southern hemispheric air depleted in 1021 1022 CH₄ transported at low altitudes, as well as high rates of OH oxidation. Covariance between 1023 species variations at the synoptic scale further helps identification and attribution of different sources and sinks, like fossil fuel combustion, biofuel burning and biogenic emissions. 1024 Besides, measurements of δ^{13} C-CO₂ have been recently started for HLE, and the 4-D 1025 distributions of CO₂ and CH₄ have been realistically simulated using a chemical transport 1026 model (LMDz-OR-INCA, Hauglustaine et al., 2004; Folberth et al., 2006) with zoom over 1027 South and East Asia (manuscript in preparation). Both of them may serve as valuable tools to 1028 disentangle and quantify contributions of different sources and meteorology to trace gas 1029 1030 signals.

1031

1032 Apart from the flask measurements of trace gases presented in this study for the three stations, in-situ continuous measurements of CO₂ and CH₄ have also been deployed at HLE, PON and 1033 PBL in parallel, which would considerably contribute to the value of the stations through 1034 high-frequency air sampling. While the three stations have the potential to provide useful 1035 1036 constraints on estimates of trace gas fluxes over South and NE India (for example, Swathi et al. (2013) reported considerable reduction in the uncertainty of inverted CO₂ fluxes over 1037 temperate Eurasia by the inclusion of measurements at HLE), the monitoring network 1038 1039 requires further expansion to sample air masses from other parts of the Indian subcontinent. Recently a few other atmospheric ground stations have been established in western India 1040 (Bhattacharya et al., 2009; Tiwari et al., 2011; Tiwari et al., 2014; Tiwari and Kumar, 2012) 1041 and the Himalayas (Kumar et al., 2010; Ganesan et al., 2013), with their concentration 1042 footprints covering Central India (e.g., the Sinhagad station; Tiwari et al., 2014; Tiwari and 1043 1044 Kumar, 2012), the Indo-Gangetic Plains and a large extent of the Himalayas (e.g., the Dajeeling station; Ganesan et al., 2013). More efforts are needed to develop a comprehensive 1045 observation network with adequate spatial and temporal coverage in this region. 1046

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1567 **Table**

Table 1 Annual mean values, trend, and average peak-to-peak amplitudes of trace gases at HLE, PON, PBL and the two additional NOAA/ESRL stations – KZM and WLG. For each species at each station, the annual mean values and average peak-to-peak amplitude are calculated from the smoothed curve and mean seasonal cycle, respectively. The residual standard deviation (RSD) around the smoothed curve and the Julian days corresponding to the maximum (D_{max}) and minimum (D_{min}) of the mean seasonal cycle are given as well. Uncertainty of each estimate is calculated from 1 s.d. of 1000 bootstrap replicates.

| | HLE | PON | PBL | KZM | WLG | | | |
|---------------------------|------------|-------------|-------------|------------|------------|--|--|--|
| CO ₂ (ppm) | | | | | | | | |
| Annual mean 2007 | 382.3±0.3 | 386.6±0.9 | _ | 382.7±0.2 | 384.2±0.2 | | | |
| Annual mean 2008 | 384.6±0.5 | 388.1±0.9 | _ | 385.7±0.2 | 386.0±0.2 | | | |
| Annual mean 2009 | 387.2±0.2 | 389.0±0.6 | _ | - | 387.4±0.2 | | | |
| Annual mean 2010 | 389.4±0.1 | 391.3±1.5 | 387.6±0.7 | _ | 390.1±0.2 | | | |
| Annual mean 2011 | 391.4±0.3 | _ | 390.2±0.6 | _ | 392.2±0.2 | | | |
| Trend (yr ⁻¹) | 2.1±0.0 | 1.7±0.1 | _ | _ | 2.0±0.0 | | | |
| (Trend at MLO: 2.0±0.0) | | | | | | | | |
| RSD | 0.7 | 4.0 | 1.5 | 1.5 | 1.4 | | | |
| Amplitude | 8.2±0.4 | 7.6±1.4 | 11.1±1.3 | 13.8±0.5 | 11.1±0.4 | | | |
| D _{max} | 122.0±2.9 | 111.0±13.4 | 97.0±26.0 | 75.0±2.6 | 100.0±1.5 | | | |
| D_{min} | 261.0±3.0 | 327.0±54.3 | 242.0±7.7 | 205.0±2.1 | 222.0±1.6 | | | |
| CH ₄ (ppb) | | | | | | | | |
| Annual mean 2007 | 1814.8±2.9 | 1859.2±6.7 | - | 1842.6±2.4 | 1841.0±1.8 | | | |
| Annual mean 2008 | 1833.1±5.4 | 1856.1±10.4 | _ | 1856.6±2.3 | 1845.6±1.5 | | | |
| Annual mean 2009 | 1830.2±1.7 | 1865.7±5.1 | _ | _ | 1851.8±1.9 | | | |
| Annual mean 2010 | 1830.5±2.1 | 1876.9±9.1 | 1867.5±15.4 | _ | 1857.6±1.4 | | | |
| Annual mean 2011 | 1849.5±5.2 | _ | 1852.0±7.6 | _ | 1859.9±1.2 | | | |
| Trend (yr ⁻¹) | 4.9±0.0 | 9.4±0.1 | - | _ | 5.3±0.0 | | | |
| (Trend at MLO: 6.2±0.0) |) | | | | | | | |
| RSD | 9.1 | 34.4 | 22.4 | 14.6 | 12.3 | | | |
| Amplitude | 28.9±4.2 | 124.1±10.2 | 143.9±12.4 | 22.7±4.7 | 17.5±2.2 | | | |
| D _{max} | 219.0±4.6 | 337.0±6.1 | 345.0±87.6 | 236.0±43.2 | 222.0±6.2 | | | |
| D _{min} | 97.0±58.9 | 189.0±10.7 | 193.0±13.5 | 338.0±39.0 | 340.0±96.6 | | | |
| N ₂ O (ppb) | | | | | | | | |
| Annual mean 2007 | 322.2±0.1 | 324.8±0.3 | - | | | | | |
| Annual mean 2008 | 322.9±0.1 | 326.3±0.3 | _ | | | | | |
| Annual mean 2009 | 323.5±0.1 | 326.7±0.3 | _ | | | | | |
| Annual mean 2010 | 324.0±0.1 | 327.1±0.5 | 329.0±0.5 | | | | | |
| Annual mean 2011 | 325.2±0.1 | - | 327.9±0.3 | | | | | |

| Trend (yr ⁻¹) | 0.8±0.0 | 0.8±0.1 | _ | | | | | |
|---------------------------|-----------------|------------------|------------------|------------|-----------|--|--|--|
| (Trend at MLO: 1.0±0.0) | | | | | | | | |
| RSD | 0.3 | 1.4 | 1.1 | | | | | |
| Amplitude | 0.6±0.1 | 1.2±0.5 | 2.2±0.6 | | | | | |
| D _{max} | 227.0±11.8 | 262.0±83.2 | 313.0±42.6 | | | | | |
| D _{min} | 115.0±16.4 | 141.0 ± 48.2 | 65.0±33.4 | | | | | |
| SF ₆ (ppt) | | | | | | | | |
| Annual mean 2007 | 6.26±0.03 | 6.19±0.01 | - | | | | | |
| Annual mean 2008 | 6.54 ± 0.03 | 6.49±0.02 | _ | | | | | |
| Annual mean 2009 | 6.79±0.01 | 6.77±0.01 | _ | | | | | |
| Annual mean 2010 | 7.17 ± 0.01 | 7.08 ± 0.02 | $7.10{\pm}0.07$ | | | | | |
| Annual mean 2011 | 7.38±0.01 | _ | 7.45 ± 0.03 | | | | | |
| Trend (yr ⁻¹) | 0.29 ± 0.05 | 0.31±0.05 | _ | | | | | |
| (Trend at MLO: 0.29±0.03) | | | | | | | | |
| RSD | 0.07 | 0.05 | 0.12 | | | | | |
| Amplitude | 0.15±0.03 | 0.24 ± 0.02 | 0.48 ± 0.07 | | | | | |
| D _{max} | 320.0±8.3 | 327.0±12.1 | 342.0±59.9 | | | | | |
| D _{min} | 211.0±65.1 | 204.0±3.3 | $210.0{\pm}18.1$ | | | | | |
| CO (ppb) | | | | | | | | |
| Annual mean 2007 | 104.7±1.4 | 200.5±7.8 | - | 121.7±1.7 | 141.0±4.3 | | | |
| Annual mean 2008 | 103.1±2.1 | 175.3±13.1 | _ | 123.7±1.7 | 129.0±2.9 | | | |
| Annual mean 2009 | 98.9±1.9 | 174.3±4.8 | _ | _ | 131.9±3.7 | | | |
| Annual mean 2010 | 99.0±1.2 | 185.1±8.7 | 157.6±20.4 | _ | 130.2±3.9 | | | |
| Annual mean 2011 | 99.4±2.2 | _ | 145.9±9.9 | _ | 124.0±2.3 | | | |
| Trend (yr ⁻¹) | -2.2±0.0 | 0.4±0.1 | _ | _ | -1.9±0.0 | | | |
| (Trend at MLO: -1.6±0.0) | | | | | | | | |
| RSD | 6.5 | 32.0 | 30.8 | 11.8 | 22.5 | | | |
| Amplitude | 28.4±2.3 | 78.2±11.6 | 144.1 ± 16.0 | 37.1±4.4 | 38.6±5.1 | | | |
| D _{max} | 79.0±11.4 | 4.0±160.2 | 12.0±117.9 | 72.0±5.0 | 94.0±38.2 | | | |
| D _{min} | 297.0±5.3 | 238.0±46.1 | 213.0±23.0 | 318.0±6.1 | 331.0±6.2 | | | |
| H ₂ (ppb) | | | | | | | | |
| Annual mean 2007 | 539.6±2.1 | 574.5±2.4 | - | 502.4±2.0 | 500.9±1.5 | | | |
| Annual mean 2008 | 533.2±3.2 | 558.2±5.3 | - | - | - | | | |
| Annual mean 2009 | 533.3±1.6 | 562.4±1.6 | _ | _ | _ | | | |
| Annual mean 2010 | 533.5±1.8 | 563.9±2.3 | 558.6±2.4 | _ | _ | | | |
| Annual mean 2011 | 536.9±1.5 | _ | 555.4±1.6 | _ | _ | | | |
| Trend (yr ⁻¹) | -0.5±0.0 | -1.3±0.1 | _ | _ | _ | | | |
| RSD | 6.6 | 8.4 | 7.0 | 13.3 | 9.5 | | | |
| Amplitude | 15.8±2.2 | 21.6±3.4 | 21.3±5.0 | 16.7±4.0 | 22.8±3.0 | | | |
| D _{max} | 120.0±8.7 | 96.0±9.6 | 99.0±8.8 | 120.0±34.2 | 51.0±13.4 | | | |
| D _{min} | 266.0±39.6 | 219.0±10.3 | 353.0±87.8 | 341.0±78.3 | 298.0±6.5 | | | |

1576 Figures

Figure 1 (a) Five-day back-trajectories calculated for all sampling dates over the period
2007–2011 at Hanle (HLE), Pondicherry (PON), and Port Blair (PBL) during April–June
(AMJ), July–September (JAS), October–December (OND) and January–March (JFM),
respectively. Back-trajectories are colored by the elevation of air masses at hourly time step.
(b) Map of terrain over the zoomed box in (a), showing locations of HLE, PON and PBL.
The digital elevation data are obtained from NASA Shuttle Radar Topographic Mission
(SRTM) product at 1km resolution (http://srtm.csi.cgiar.org)



Figure 2 Time series of CO_2 flask measurements at (a) HLE and PON, (b) HLE and PBL, (c) HLE and KZM, and (d) HLE and WLG. The open circles denote flask data used to fit the smoothed curves, while the crosses denote discarded flask data lying outside 3 times the residual standard deviations from the smoothed curve fits. For each station, the smoothed curve is fitted using Thoning's method (Thoning et al., 1989) after removing outliers.



1590

1591 Figure 3 (a) The mean CO₂ seasonal cycle at HLE, in comparison with the mean seasonal cycles derived from the in-situ CO₂ measurements over New Delhi at different altitude bands 1592 (3-4 km, 4-5 km, and 5-6 km) by the CONTRAIL project (2006–2010). (b) The mean CO₂ 1593 1594 seasonal cycles at HLE, KZM and WLG. (c) The mean CO₂ seasonal cycles at HLE, PON 1595 and PBL. For each station, the mean seasonal cycle is derived from the harmonics of the smoothed fitting curve in Fig. 2. Shaded area indicates the uncertainty of the mean seasonal 1596 1597 cycle calculated from 1 s.d. of 1000 bootstrap replicates. For the CONTRAIL datasets, CO₂ measurements over New Delhi were first averaged by altitude bands. A fitting procedure was 1598 1599 then applied to the aggregated CO₂ measurements to generate the mean season cycle for different altitude bands. 1600



Figure 4 Time series of CH_4 flask measurements at (a) HLE and PON, (b) HLE and PBL, (c) HLE and KZM, and (d) HLE and WLG. The open circles denote flask data used to fit the smoothed curves, while the crosses denote discarded flask data lying outside 3 times the residual standard deviations from the smoothed curve fits. For each station, the smoothed curve is fitted using Thoning's method (Thoning et al., 1989) after removing outliers.



1606 1607

1608 Figure 5 (a) The mean CH_4 seasonal cycles observed at HLE, KZM and WLG. The mean CH₄ seasonal cycle derived from aircraft flask measurements by the CARIBIC project is also 1609 presented. The CARIBIC flask measurements in the upper troposphere (200-300 hPa) during 1610 2005–2012 are averaged over the Indian subcontinent (10°N-35°N, 60°E-100°E) by month to 1611 generate the mean seasonal cycle. The error bars indicate 1 standard deviation of CH₄ flask 1612 measurements within the month. (b) The seasonal variations of CH₄ emissions from rice 1613 paddies and wetlands over the Indian subcontinent. The CH₄ emissions from rice paddies are 1614 extracted from a global emission map for the year 2010 (EDGAR v4.2), imposed by the 1615 1616 seasonal variation on the basis of Matthews et al. (1991). The CH₄ emissions from wetlands are extracted from outputs of a global vegetation model (BIOME4-TG, Kaplan et al., 2006). 1617 The seasonal variation of deep convection over the Indian subcontinent is also presented, 1618 indicated by convective precipitation obtained from an LMDz simulation nudged with 1619 ECMWF reanalysis (Hauglustaine et al., 2004). The CH₄ emissions and convective 1620 precipitation are averaged over the domain 10-35 °N, 70°-90°E to give a regional mean 1621 estimate. (c) The mean CH₄ seasonal cycles observed at HLE, PON and PBL. For each 1622 station, the mean seasonal cycle is derived from the harmonics of the smoothed fitting curve 1623 1624 in Fig. 4. Shaded area indicates the uncertainty of the mean seasonal cycle calculated from 1 1625 s.d. of 1000 bootstrap replicates.





Figure 6 Time series of N₂O flask measurements at (a) HLE and PON, (b) HLE and PBL. 1628 The open circles denote flask data used to fit the smoothed curves, while crosses denote 1629 discarded flask data lying outside 3 times the residual standard deviations from the smoothed 1630 curve fits. For each station, the smoothed curve is fitted using Thoning's method (Thoning et 1631 al., 1989) after removing outliers. 1632



1634

Figure 7 The mean N₂O seasonal cycles observed at (a) HLE and PON, (b) HLE and PBL.
For each station, the mean seasonal cycle is derived from the harmonics of the smoothed
fitting curve in Fig. 6. Shaded area indicates the uncertainty of the mean seasonal cycle
calculated from 1 s.d. of 1000 bootstrap replicates.



Figure 8 Time series of SF_6 flask measurements at (a) HLE and PON, (b) HLE and PBL, (c) HLE and KZM, and (d) HLE and WLG. The open circles denote flask data used to fit the smoothed curves. For each station, the smoothed curve is fitted using Thoning's method (Thoning et al., 1989) after removing outliers.



1645

Figure 9 The mean SF_6 seasonal cycles observed at (a) HLE and PON, (b) HLE and PBL. For each station, the mean seasonal cycle is derived from the harmonics of the smoothed fitting curve in Fig. 8. Shaded area indicates the uncertainty of the mean seasonal cycle calculated from 1 s.d. of 1000 bootstrap replicates.



Figure 10 Time series of CO flask measurements at (**a**) HLE and PON, (**b**) HLE and PBL, (**c**) HLE and KZM, and (**d**) HLE and WLG. The open circles denote flask data used to fit the smoothed curves, while the crosses denote discarded flask data lying outside 3 times the residual standard deviations from the smoothed curve fits. For each station, the smoothed curve is fitted using Thoning's method (Thoning et al., 1989) after removing outliers.



1658

Figure 11 The mean CO seasonal cycles observed at (a) HLE and PON, (b) HLE and PBL,
(c) HLE and KZM, and (d) HLE and WLG. For each station, the mean seasonal cycle is
derived from the harmonics of the smoothed fitting curve in Fig. 10. Shaded area indicates
the uncertainty of the mean seasonal cycle calculated from 1 s.d. of 1000 bootstrap replicates.



Figure 12 Time series of H_2 flask measurements at (a) HLE and PON, (b) HLE and PBL, (c) HLE and KZM, and (d) HLE and WLG. The open circles denote flask data used to fit the smoothed curves, while the crosses denote discarded flask data lying outside 3 times the residual standard deviations from the smoothed curve fits. For each station, the smoothed curve is fitted using Thoning's method (Thoning et al., 1989) after removing outliers.



1671

Figure 13 The mean H₂ seasonal cycles observed at (**a**) HLE and PON, (**b**) HLE and PBL, (**c**) HLE and KZM, and (**d**) HLE and WLG. For each station, the mean seasonal cycle is derived from the harmonics of the smoothed fitting curve in Fig. 12. Shaded area indicates the uncertainty of the mean seasonal cycle calculated from 1 s.d. of 1000 bootstrap replicates.



1677

Figure 14 The relationships between ΔCH_4 and ΔCO at HLE (**a**–**d**), PON (**e**–**h**), and PBL (**i**– 1679 **I**) for April–June (AMJ), July–September (JAS), October–December (OND), and January– 1680 March (JFM). For each panel, ΔCH_4 and ΔCO are estimated as residuals from smoothed 1681 curves. The $\Delta CH_4/\Delta CO$ ratio is the slope of the fitting line from the orthogonal distance 1682 regression, with the SD calculated from 1000 bootstrap replications.



1683
Figure 15 The relationships between ΔCH_4 and ΔCO_2 at HLE (**a**–**d**), PON (**e**–**h**), and PBL (**i**–**l**) for April–June (AMJ), July–September (JAS), October–December (OND), and January– March (JFM). For each panel, ΔCH_4 and ΔCO_2 are estimated as residuals from smoothed curves. The $\Delta CH_4/\Delta CO_2$ ratio is the slope of the fitting line from the orthogonal distance regression, with the SD calculated from 1000 bootstrap replications. For ΔCH_4 and ΔCO_2 that is not significantly correlated, the fitting line is not plotted.



1692

Figure 16 The relationships between ΔCO and ΔCO_2 at HLE (**a**–**d**), PON (**e**–**h**), and PBL (**i**– 1694 **I**) for April–June (AMJ), July–September (JAS), October–December (OND), and January– 1695 March (JFM). For each panel, ΔCO and ΔCO_2 are estimated as residuals from smoothed 1696 curves. The $\Delta CO/\Delta CO_2$ ratio is the slope of the fitting line from the orthogonal distance 1697 regression, with the SD calculated from 1000 bootstrap replications. For ΔCO and ΔCO_2 that 1698 is not significantly correlated, the fitting line is usually not plotted.



1700

1701 **Figure 17** (a) The relationship between ΔCH_4 and ΔCO at PBL (colored by red) and BKT (colored by grey) during July–September (JAS) over the period of 2007–2011. ΔCH_4 and 1702 1703 ΔCO are estimated as residuals from smoothed curves. The $\Delta CH_4/\Delta CO$ ratio is the slope of the fitting line from orthogonal distance regression (ODR), with the SD calculated from 1000 1704 1705 bootstrap replications. Two abnormal events at PBL are labeled, with enhancements of CH₄ and CO on September 16, 2009 and July 29, 2011, respectively. Enhancements of CH₄ and 1706 1707 CO are also observed at BKT on Sep. 8, 2009. (b) Hourly CH₄ and CO measurements at BKT in July, 2011 (BMKG & EMPA). Enhancements of CH4 and CO are observed during July 17-1708 1709 21, 2011.

