Atmospheric Speciated Mercury Concentrations on an Island between China and Korea: Sources and Transport Pathways

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Abstract

As a global pollutant, mercury (Hg) is of particular concern in East Asia where anthropogenic emissions are the largest. In this study, speciated Hg concentrations were measured on Yongheung Island, the western most island in Korea, located between China and the Korean mainland to identify the importance of local and regional Hg sources. Various tools including correlations with other pollutants, conditional probability function, and back-trajectory based analysis consistently indicated that Korean sources were important for gaseous oxidized mercury (GOM) whereas, for total gaseous mercury (TGM) and particulate bound mercury (PBM), regional transport were also important. A trajectory cluster based approach considering both Hg concentration and the fraction of time each cluster was impacting the site was developed to quantify the effect of Korean sources and out-of-Korean source. This analysis suggests that contributions from out-of-Korean sources were similar to Korean sources for TGM whereas Korean sources contributed slightly more to the concentration variations of GOM and PBM compared to out-of-Korean sources. The ratio of GOM/PBM decreased when the site was impacted by regional transport, suggesting that this ratio may be a useful tool for identifying the relative significance of local sources vs. regional transport. The secondary formation of PBM through gas-particle partitioning with GOM was found to be important at low temperatures and high relative humidity.

***Keywords** : TGM, GOM, PBM, regional transport, gas-particle partitioning, cluster analysis*

1. Introduction

Mercury (Hg) is the only metal that exists as a liquid at standard conditions (US EPA, 1997) which results in it having a significant vapor pressure and presence in the atmosphere. In the atmosphere, Hg generally does not constitute a direct public health risk at the level of exposure usually found (Driscoll et al, 2007). However, once Hg is deposited into aquatic systems, it can be transformed into methyl-mercury (MeHg) which is very toxic and readily bioaccumulates through the food web (Mason et al.,

1995). Many studies show that one of the major sources of MeHg in aquatic and terrestrial system is atmospheric deposition of inorganic Hg (Landis and Keeler, 2002, Mason et al., 1997). Fish consumption has been considered to be the major exposure pathway of Hg for humans (Mergler et al., 2007; UNEP, 2013). In Korea, You et al. (2012) showed that MeHg concentrations in blood were affected by fish consumption as well as by gender difference. However, rice consumption was also found to be the predominant pathway of MeHg exposure for the inhabitants residing in a highly contaminated area of China (Zhang et al., 2010).

Atmospheric mercury exists in three major inorganic forms, including gaseous elemental mercury (GEM, Hg^0), gaseous oxidized mercury (GOM, Hg^{2+}) and particulate bound mercury (PBM, Hg(p)). The sum of the GEM and GOM is often called as total gaseous mercury (TGM). Due to its high water solubility and deposition velocity, GOM has short atmospheric residence times (~day) and, consequently, its ambient concentration is mainly affected by local sources. Besides the anthropogenic sources, the free troposphere has been identified as an important GOM source (Huang and Gustin, 2012; Weiss-Penzias et al., 2009; Timonen et al., 2013). In contrast, GEM, which comprises more than 95 percent of the total Hg in ambient air, can be transported long distances because it is relatively inert and has low water solubility and deposition velocity (Lin and Pehkonen, 1999). The residence time of PBM is dependent on the size of associated particles, but generally, it has been assumed to be a few days (Fang et al., 2012, Zhang et al., 2001). Measurements of GOM and PBM are challenging and uncertain due to their extremely low concentrations and complex chemical reactivity, and because their chemical forms are not actually known (Pirrone et al., 2013). In most studies, GOM and PBM have operational definitions for the mercury species collected by a KCl coated denuder and by a quartz filter downstream of a KCl denuder, respectively. It is typically assumed that GOM comprises $HgCl_2$, $HgBr_2$, HgO, $Hg(NO_3)_2$, and $HgSO_4$. However, the sampling method including the use of a KCl denuder has been shown to be subject to interferences from ozone, water vapor and possibly other compounds (Lyman et al. 2010; Talbot et al., 2011; Jaffe et al., 2014; Finley et al., 2013; Gustin et al., 2013; Huang et al., 2013; McClure et al., 2014), and it also should be noted that the different Hg(II) compounds have different collection efficiencies by the KCl coated denuder (Gustin et al.,

2015).

In the atmosphere, Hg species can be interconverted through various redox reactions. It is known that GOM can be produced by homogeneous and heterogeneous reactions of GEM with O_3, OH, and Br/BrO (Hedgecock and Pirrone, 2004; Obrist et al., 2011; Subir et al., 2011), but there is no consensus on which oxidants are most important under which environmental conditions. GEM can also be formed through reduction of GOM predominantly in cloud water (Subir et al. 2011, 2012). GOM can also be converted to PBM through gas-particle partitioning, with the partition coefficient, K_p, inversely correlated with temperature and positively correlated with particle surface area (Lyman and Keeler, 2005, Rutter and Schauer, 2007a,b; Liu et al., 2010). Since GEM makes up the bulk of the total Hg in ambient air its formation through reduction processes of divalent Hg may not be important. However, the secondary formations of GOM through the oxidation of Hg^0 followed by gas-particle partitioning formation of PBM can contribute significantly to their ambient concentrations.

The region of largest anthropogenic Hg emissions is East and Southeast Asia, contributing 39.7% of the total global anthropogenic emissions (UNEP, 2013). In Korea, atmospheric Hg emissions have generally decreased since 1990 (Kim et al., 2010). However, Hg levels in Korea are likely to be highly susceptible to Chinese emissions because China alone accounts for about one third of the global total (UNEP, 2013) and Korea is situated just east (and generally downwind) of China. According to the recent studies, Hg concentrations in blood of Koreans are more than 4~8 times higher than those found in US and Germany, and approximately 26% of Koreans have higher blood mercury concentrations than a USA guideline level (http://envhealth.nier.go.kr), indicating that there is an urgent need to identify the Hg sources and pathways controlling Hg concentrations in Korea.

This study was designed to identify the contribution of various Hg sources including direct emissions from anthropogenic and natural sources and indirect secondary formation processes to atmospheric Hg concentrations in Korea. In order to achieve these objectives, Hg concentrations were measured in the western most island in Korea, located in between eastern China and the Korean mainland, so that, depending on wind patterns, the effects of Chinese and Korean Hg emissions could be evaluated. Previously, our group qualitatively evaluated the impact of local Korean sources and

regional Chinese sources on TGM concentrations at the same sampling site (Lee et al. 2014). However, that work was unable to identify the effect of sources on Hg levels in Korea because only TGM was measured whereas all three Hg species are needed since they have very different physical and chemical properties. In this study, the importance of sources and pathways were both qualitatively and quantitatively evaluated using all three Hg species' concentrations measured throughout the extended sampling period.

2. Materials and methods

2.1 Site description

TGM, GOM and PBM were measured on the roof of a three-story building on Yongheung Island (YI), the westernmost island in Korea (Fig. 1). YI is a small island located about 15~20 km west from mainland Korea with a population of 5,815. The Yongheung Coal-fired Power Plant (YCPP), located approximately 4.5 km southwest of the sampling site (Fig 1c), emits about 0.11 ton yr^{-1} of Hg. To the east of the sampling site, industrial (the Incheon industrial complex shown as light violet color in Fig 1b and Fig 1c) and metropolitan (Seoul shown as a pink color in Fig 1b and Fig 1c) areas are located in mainland Korea, and, in the southern direction, there are three large coal fired power plants (Fig 1b). The Hg emission rate of anthropogenic sources in Korea was estimated to be 8.04 ton yr^{-1} in 2010, with cement production being the largest source type (AMAP/UNEP, 2013).

2.2 Sampling and analysis

From January 2013 to August 2014, three atmospheric mercury species: TGM (GEM+GOM), GOM, and PBM (≤2.5 μm) were measured during eight intensive sampling periods (Table 1). TGM concentrations were measured every 5 min using a mercury vapor analyzer (Tekran 2537B). This instrument contains two gold cartridges which collect and thermally desorb Hg alternately. Desorbed Hg is quantifed using cold vapor atomic fluorescence spectrometry (CVAFS). Outdoor air at a flow rate of 1.0 L min^{-1} was transported through a 3-m-long heated sampling line (1/4" OD Teflon) into the analyzer. There is a possibility that some of GOM may be sorbed in the line upstream of the gold traps

135 although the sampling line and inlet were maintained at 50^{o}C to prevent the GOM sorption. However,
136 it is typically assumed that Tekran 2537 can collect and analyze TGM in most of studies (Temme et al.,
137 2002; Gustin et al., 2013; Han et al., 2014; Zhang et al., 2015). The Tekran 2537B was automatically
138 calibrated daily using an internal permeation source. Manual injections were also used to evaluate
139 these automated calibrations before each sampling campaign using a saturated mercury vapor
140 standard. The relative percent difference between manual injection and automated calibration was
141 <2%. Five-point manual calibration was also performed by injecting Hg vapor into the sampling line
142 two times during the study period. The R^2 ranged from 0.9991 to 0.9997 between mass injected and
143 Tekran reported area, and the average relative percent difference between the mass injected and the
144 mass calculated was 5.5%. The method detection limit (0.04 ng m^{-3}) was calculated as three times the
145 standard deviation obtained after injecting 1 pg of the mercury vapor seven times. The recovery rate
146 (96 ± 3%) was obtained by directly injecting Hg vapor into the sampling line between the sample inlet
147 and the Tekran 2537B in a zero-air stream.
148 GOM and PBM were collected manually using an annular denuder coated with KCl followed by a
149 quartz filter, respectively, at a flow rate of 10 L min^{-1}. To identify any diurnal variations, all samples
150 were separately collected during the daytime (07:00-19:00) and nighttime (19:00-07:00) except during
151 the 7th sampling period when they were measured every 2 hrs. The sampling system including an
152 elutriator, an impactor, a KCl-coated denuder and a filter pack was housed in a custom-made sampling
153 box maintained at 45^{o}C to prevent hydrolysis of KCl. After sampling, the denuder and quartz filter
154 were thermally desorbed using a tube furnace at 525^{o}C and 900^{o}C, respectively, to convert Hg^{2+} to
155 Hg^{0} in a carrier gas of zero air. The heated air was then transported into a Tekran 2537B for
156 quantification. Field blanks for GOM and PBM were collected once for each sampling period, and
157 their average values were 0.23 ± 0.12 pg m^{-3} and 0.25 ± 0.09 pg m^{-3}, respectively.
158 The sampling methods used in this study are currently the most accepted methods for the
159 measurement of atmospheric GOM and PBM, however there are many studies reporting that these
160 methods are subject to interferences from ozone, water vapor and possibly other compounds (Lyman
161 et al. 2010; Talbot et al., 2011; Jeff et al., 2014; Finley et al., 2013; Gustin et al., 2013; Huang et al.,

2013; McClure et al., 2014) although recent side-by-side measurements with two Tekran systems showed good agreement and no impact from added ozone and increasing relative humidity (Edgerton, personal communication). Also, it should be noted that the concentrations of PBM measured during 12-hrs of sampling time (all sampling periods except in the 7^{th}) may have been biased due to Hg loss from filters over the long sampling period; however for model development any loss of PBM is assumed to be the same for each sampling period. Therefore, it should be noted that the GOM and PBM measurements reported in this study may be somewhat biased even though, at present, it is not possible to quantify the magnitude of these uncertainties.

Meteorological data including temperature, wind speed, wind direction, relative humidity and solar radiation were also measured every 5 min at the sampling site using a meteorological tower (DAVIS Inc weather station, Vintage Pro2^{TM}).

Hourly concentrations of SO_2, NO_2, CO, O_3 and PM_{10} were obtained from the national air quality (NAQ) monitoring station (http://www.airkorea.or.kr/) located approximately 8 km east from the sampling site. These concentrations were compared with those measured at another national air quality monitoring station located approximately 24 km west of the Hg sampling site, and there were no statistical differences between sites (p-value< 0.001), indicating that the spatial distribution of these pollutants was relatively uniform across the area

2.3. Backward trajectory and Cluster analysis

Three-day backward trajectories were calculated using the NOAA HYSPLIT 4.7 with GDAS (Global Data Assimilation System) meteorological data which supplies 3-hour, global 1 degree latitude-longitude datasets of the pressure surface. Hourly 3-day back-trajectories were calculated for each hour of sampling, and the arrival heights of both 200 m and 500 m were used to describe the local and the regional transport meteorological pattern. Since GOM and PBM were measured for 12hr for most time periods, hourly trajectories were matched to the 12hr-averaged GOM and PBM concentration; therefore, in total 12 back-trajectories represented one averaged GOM or PBM concentrations. Although there have been many studies using different time scales for measurements

of pollutants and for meteorological data (Amato and Hopke, 2012; Galindo et al., 2011; Kim et al., 2007) this mismatch might increase the uncertainty for the trajectory-based approaches.

The backward trajectories were clustered into groups with similar transport patterns using NOAA HYSPLIT 4.7. This method minimizes the intra-cluster differences among trajectories while maximizing the inter-cluster differences. The clustering of trajectories is based on the total spatial variance (TSV) method. TSV is the sum of all the cluster spatial variances (SPVAR) which is the sum of the squared distances between the endpoints of the cluster's component trajectories and the mean of the trajectories in that cluster. In this study, five clusters were chosen based on a large increase in TSV for larger clusters (Fig. 4S), as described in Draxler et al. (2014) and Kelly et al. (2012). A more detailed description of the clustering process can be obtained in Draxler et al. (2014).

2.5. Conditional Probability Function (CPF)

The conditional probability that a given concentration from given wind direction will exceed a predetermined threshold criterion, was calculated using the following equation.

$$\mathrm{CPF}_{\Delta\theta} = \frac{m_{\Delta\theta}}{n_{\Delta\theta}} \qquad \text{(Eq.1)}$$

where $m_{\Delta\theta}$ is the number of occurrences from wind sector $\Delta\theta$ where the concentration is higher than a criterion value, and $n_{\Delta\theta}$ is the total number of occurrence from this wind sector. In this study, 16 sectors were used ($\Delta\theta = 22.5^{o}$), and calm winds were excluded from the calculation because of isotropic behavior of the wind vane for such conditions. For TGM, two threshold criteria of the upper 10 and 25 percentile were chosen while only the upper 25 percentile was used for GOM and PBM concentrations due to the smaller number of samples. The 1hr-averaged wind direction (WD) data were used for 12hr-averaged GOM and PBM concentrations for most of periods, so that in total 12 WS and WD were used for one averaged GOM and PBM concentrations to create CPF. It should be noted that there are some unquantifiable uncertainties derived from using different time resolutions between measurements of GOM and PBM and WD data although overall trajectories did not diverge significantly at this sampling site for most sampling periods.

2.6. Potential Source Contribution Function (PSCF)

The PSCF model counts each trajectory segment endpoint that terminates within given grid cell. A high PSCF value signifies a potential source location. The PSCF value was calculated as:

$$\text{PSCF value} = \frac{P[B_{ij}]}{P[A_{ij}]} = \frac{m_{ij}}{n_{ij}} \quad \text{(Eq.2)}$$

If N is the total number of trajectory segment endpoints over the study period and if n segment trajectory endpoints fall into the ijth cell, the probability of this event ($P[A_{ij}]$) is calculated by n_{ij}/N. If m_{ij} is the number of segment endpoints in the same ijth cell when the concentrations are higher than a criterion value, the probability of this high concentration event, B_{ij}, is given by $P[B_{ij}] = m_{ij}/N$. The criterion value was the top 25% concentration and the cell size of 0.5° by 0.5° was used for tracing sources. To reduce the uncertainty in a grid cell with a small number of endpoints, an arbitrary weight function W_{ij} was applied when the number of the endpoints in a particular cell was less than three times the average number of endpoints (N_{ave}) for all cells (Fu et al., 2011, Han et al., 2007, Polissar et al., 2001a, 2001b).

$$W_{ij} = \begin{pmatrix} 1.0 & N_{ij} > 3N_{ave} \\ 0.70 & 3N_{ave} > N_{ij} > 1.5N_{ave} \\ 0.40 & 1.5N_{ave} > N_{ij} > N_{ave} \\ 0.20 & N_{ave} > N_{ij} \end{pmatrix} \quad \text{(Eq. 3)}$$

In this study, trajectories were hourly calculated

3. Results and discussion

3.1. General trends of three Hg species

To maintain the consistency of the sampling duration, the 12hr-averaged GOM and PBM concentrations were used for the 7th sampling period to identify the general trends of Hg species. The average TGM, GOM, and PBM concentrations were 2.8 ± 1.1 ng m^{-3}, 8.3 ± 9.7 pg m^{-3}, and 10.9 ± 11.2 pg m^{-3}, respectively (Table 1). Since the GOM concentration was much lower than TGM the reported TGM concentration can be considered a good approximation of the GEM concentration. TGM varied from 0.1 to 18.8 ng m^{-3}; the highest concentration was observed around 2 am on March 18, 2014 (Fig. 2). GOM and PBM concentrations peaked at 50.9 pg m^{-3} during the daytime on March

241 19, 2014 and 43.7 pg m^{-3} during daytime on January 22, 2013, respectively (Fig. 2). The various Hg
242 species did not follow similar concentration patterns although PBM was statistically significantly
243 correlated with TGM (Pearson correlation coefficient, r= 0.235, p-value= 0.03).
244 When the data were grouped into three categories including the first (Apr., 2013, May, 2013, Mar.,
245 2014, May, 2014), the second (Aug, 2013, Aug., 2014.), and the third (Jan., 2013, Feb., 2013) periods,
246 both TGM (ANOVA/Tukey test, p-value<0.001) and PBM (p-value=0.024, Kruska-Wallis test) had
247 the highest concentrations in cold period ((Jan., 2013, Feb., 2013) while there was no statistical
248 difference in GOM concentrations among different categories (p-value= 0.288, Kruskal-Wallis test)
249 (Fig. 3). Observed TGM concentrations were substantially lower than those measured in a suburban
250 and remote site in China and metropolitan areas of Korea (Seoul), but higher than at most North
251 American sites and at a rural site of Korea (Chuncheon) (Table 2). GOM and PBM concentrations
252 were in between those typically found at urban locations and at a rural site in Korea and were much
253 lower than those typically measured in China.
254 The TGM concentration varied diurnally, generally showing morning maximums (07:00-12:00) and
255 minimums during the nighttime. In urban areas, TGM concentrations are typically higher during the
256 nighttime due to a combination of decreased GEM loss by daytime oxidation ,increased use of
257 household heating systems and decreased mixing heights at night (Kim et al., 2012; Han et al., 2014).
258 In contrast daytime peaks have been observed in rural and remote areas, likely due to increased
259 volatilized of Hg^0 from natural sources (Choi et al., 2008; Cheng et al., 2014). Overall these results
260 suggest that TGM concentrations at this site are elevated due to the proximity of regional sources and
261 daily variations are controlled by natural emissions from the ocean and soil surfaces.
262 GOM concentrations were the highest in spring (10.7±10.1 pg m^{-3}) and the lowest in summer
263 (6.2±4.9 pg m^{-3}) with statistically insignificant seasonal variation (Fig. 3). The lack of a GOM
264 seasonal variation for could be an indicator of insignificant secondary formation through
265 photochemical oxidation reactions, but it might be also be due to the small sample numbers and/or
266 relatively long sampling duration (12hr). PBM concentrations did have statistically significant
267 seasonal variations with the highest average concentration in winter (17.8±16.7 pg m^{-3}) and the lowest

268 average concentration in summer (5.8±4.1 pg m^{-3}) (Fig. 3). Higher PBM concentrations in winter
269 were likely caused by increased biomass burning and residential heating, decreased removal from the
270 atmosphere due to the lower precipitation depth, and/or lower temperatures which favor partitioning
271 to the aerosol phase. Previous studies also often observed the highest PBM concentrations in winter
272 (Mao et al., 2012; Amos et al., 2012; Lan et al., 2012).

273 In Korea emissions of PBM from anthropogenic sources are much smaller then gaseous emissions
274 (the proportion of GEM, GOM, and PBM released are 64.4%, 28.8%, and 6.8%)(Kim et al., 2010).
275 The fact that PBM concentrations are similar to GOM even though significantly less PBM is released
276 suggests that a significant portion of atmospheric PBM may be due to secondary formation through
277 gas-particle partitioning. This process is characterized by a partition coefficient, $K_{p,}$ which is inversely
278 correlated with temperature (Rutter and Schauer, 2007), possibly causing the distinct seasonal
279 variation in PBM concentrations at the sampling site.

280 The relationship between K_p, defined as:

281 $$K_p = \frac{PBM/PM}{Hg_{gas}} \quad \text{(Eq. 4)}$$

282 where PM represents the particle mass, and Hg_{gas} is the concentration of gaseous Hg and relative
283 humidity (RH) was examined (PM10 concentration was used for PM in Eq. 4 in this study). RH was
284 included since in recent studies K_p was found to increase at high relative humidity in colder seasons
285 (Lyman and Keeler, 2005; Liu et al., 2010). Note that the sampling site was located in a coastal area
286 with generally high RH.

287 Some previous studies suggested that all gaseous mercury species including Hg^0 may deposit on
288 particles (Xiu et al., 2005; 2009); however, others suggested that the gas-particle partitioning of GOM
289 occurred but assumed that the adsorption of Hg^0 on particles was negligible due to its high vapor
290 pressure (Amos et al., 2012; Rutter and Schauer, 2007). Consistent with this hypothesis, we found a
291 statistically significant multiple linear relationship between K_p with temperature and relative humidity
292 (Fig. 4):

293 $$\mathrm{Log}(K_p) = -2.518 - 0.036(\mathrm{T}) + 0.017(\mathrm{RH}) \quad \text{(Eq. 5)}$$

294 where T and RH indicate atmospheric temperature (^{o}C) and relative humidity (%), respectively. The

multiple linear equation fit the data well (R^2=0.29, R=0.54, p-value<0.001), and both variables of temperature and relative humidity were statistically significant (p-value<0.001). The partial correlation coefficients were -0.389 for temperature (p-value<0.001) and 0.375 for RH (p-value=0.001). When each of the temperature and the RH was used as a single independent variable the log(K_p) regression equation was still significant with a Pearson correlation coefficient of -0.42 (R^2=0.18, p-value<0.001) and 0.39 (R^2= 0.15, p-value< 0.001), somewhat lower than that from the multiple regression (Fig. 1S). Amos et al. (2012) found an empirical gas-particle partitioning relationship between K_p and T using the Hg data obtained from five monitoring sites in United States and Canada and two laboratory experiments, and average equation is :

$$\mathrm{Log}\left(1/K_p\right) = (10 \pm 1) - (2500 \pm 300)/T \qquad \text{(Eq. 6)}$$

where T is in Kelvin unit. When a partial linear regression was performed with T in this study similar equation was derived: $Log\ (1/K_p) = 13.5 - 3362.7/T$. In Amos et al. (2012), the coefficients, ß and y_0, ranged from -1600 to -3300 and 6 to 13, respectively, and R^2 ranged from 0.16 to 0.57 in different monitoring sites. Rutter and Schauer (2007a) also determined the relationship between $Log(1/K_p)$ for urban aerosols with inverse temperature and found the slope and intercept were -4250±480 and 10±2, respectively, with R^2 of 0.77. Another study using datasets from 10 AMNet sites located in North America found R^2 ranged between 0.04 and 0.53. In summary the values derived in this study fall between those reported by other studies. Somewhat lower R^2 in this study is probably caused by smaller number of samples, different composition of aerosol, different GOM species, and/or longer sampling duration.

Han et al. (2014) also found a significant multiple linear relationship between the ratio of PBM/GOM with temperature and relative humidity at a rural site (R^2= 0.613, ß for T= -0.774, ß for RH= 0.33) but not at an urban site. The lower correlation coefficient and the beta values found in this study compared to those from Han et al. (2014) is probably due to the greater impact of primary anthropogenic sources around the sampling site. A large anthropogenic source near the sampling site is likely to weaken the relative contribution of gas-particle partitioning (secondary formation) to the variations of ambient PBM and GOM concentration because both Hg species can be also strongly

affected by the primary source emission. In the study of Amos et al. (2012), the highest R^2 was shown in Experimental Lakes Area of Canada and Reno, NV where no large anthropogenic sources of Hg is located.

In previous researches, GOM concentration measured using KCl denuder is subject to interferences under the conditions of high ozone and relative humidity (Gustin et al., 2012; Huang et al., 2013; Gustin et al., 2015) although it is currently the most accepted method. To evaluate the possible error on interpreting result, K_p was re-calculated with re-calculated GOM concentration using the empirical equation suggested by McClure et al. (2014) (RH=0.63 GOM loss % + 18.1). Since McClure et al. (2014) suggested this equation at RH of 21 to 62%, the GOM concentrations collected only when RH was from 20% to 65% were re-calculated. The multi-linear relationship was compared between re-calculated K_p and original K_p, and the results were very similar to each other (Fig. 2S).

3.2. Tracing sources of Hg species

Correlations between Hg and other pollutant concentrations are often used to identify sources. For example good correlations with SO_2 and CO typically indicate the impact of coal combustion (Pirrone et al., 1996; Han et al., 2014), and a strong correlation between Hg and CO has often been used as an indicator for regional transport because both pollutants have similar sources and do not easily decompose by reaction and undergo deposition during transport (Weiss-Penzias et al., 2003, 2006; Kim et al., 2009) although a few recent studies showed the significant bromine-induced oxidation of GEM in the mid-latitude marine boundary layer as well as in the polar atmosphere (Ariya, 201; Obrist et al., 2011) . A good correlation between Hg and NO_2 suggests the site is being impacted by local sources because the lifetime of NO_2 is relatively short compared with that of CO (Seinfeld and Pandis, 2006). In this study TGM concentrations were well correlated with SO_2, CO, and PM_{10} concentrations but not with NO_2 concentrations (Table 3), indicating that regional transport of TGM emitted from coal combustion was impacting the site throughout much of the sampling period. PBM concentrations also had a statistically significant relationship with TGM and CO suggesting regional transport is also important for PBM, but GOM was not correlated with any other pollutant suggesting it is impacted to

a greater extend by local sources (see additional discussion below).

CPF plot shows that the top 25% TGM concentrations were associated with winds from the NNW and eastern direction, pointing towards northeastern China and inland Korean sources; however, when the criterion was set to the top 10% the winds from NNW became less important and the sources located in southern and eastern areas from the sampling site were identified as an important source direction (Fig. 5). The CPF plot for GOM is significantly different from the one for PBM. High PBM concentrations were associated with northern winds while GOM concentrations were enhanced during southeastern winds.

These results suggest that for PBM regional transport from Chinese and North Korea sources were more important than Korean sources; in contrast coal fired power plants located in the southern direction rather than regional transport impacted GOM concentration. It should be noted that this result is in apparent conflict with the finding that there was no relationship between GOM and SO_2 concentrations. Total SO_2 emissions from power plants in China (18.6 Tg yr^{-1}) are much larger than in Korea (0.09 Tg yr^{-1}), and SO_2 emission rates per capita and per area in China also greatly surpass those in Korea (Lu et al., 2010). Much larger SO_2 emissions in China raise the background SO_2 concentration in the region and may mask any correlation between GOM and SO_2 even if coal fired power plants located in the southerly direction from the sampling site impacted GOM concentrations. In support of this hypothesis the pollution rose indicates that high SO_2 concentrations are associated with westerly winds while high GOM concentrations are associated with southerly winds. It should be noted that only the top 25% of GOM and PBM concentrations were used as the criteria for the CPF plot because the number of samples for both species were significantly less than for TGM due to their longer sampling duration (12hr).

Among the eight sampling periods, the second period (April, 2013) had the highest TGM, PBM and the second highest GOM average concentration, and SO_2, NO_2, CO, and PM10 were also quite high (Table 1). During this period, TGM was statistically well correlated with SO_2 (r= 0.55), NO_2 (r=0.56), and CO (r=0.36), with the highest Pearson correlation coefficient with NO_2, the characteristic local pollutant. In addition, the CPF plot (for TGM) and the back-trajectories were also associated with

easterly winds transporting air masses from major Korean urban areas, supporting the previous suggestion that inland sources enhanced all three Hg concentrations during the second sampling period.

In contrast the fifth sampling period had the lowest GOM, PBM and the second lowest TGM concentrations (Table 1). Note however that the TGM concentrations for the first couple of days reached approximately 5 ng m^{-3} and gradually decreased to about 1 ng m^{-3} during the last days of sampling (Fig. 2), indicating that there was likely two different sources affecting Hg concentrations during this period. Back-trajectories associated with high TGM concentrations passed through northeastern China, North Korea and the industrial/metropolitan areas of Korea before arriving at the sampling site whereas trajectories during the low TGM concentrations spent long periods within the ocean boundary layer. Although Hg can be emitted from ocean surface (Han et al., 2007; UNEP, 2013) heavy rain and low solar radiation occurring during the last two days of this period probably inhibited emissions of Hg from the ocean surface.

3.2.1 GOM/PBM ratio

According to the CPF results, the winds from NW and NE of the sampling site were responsible for the elevated PBM concentrations while easterly winds pointing towards inland Korea were associated with increased GOM concentrations (Fig. 5). The finding that regional transport of TGM and PBM to the site is important is supported by their significant correlation with CO (Table 3). In order to identify the relative importance of local sources relative to regional transport, the ratio of GOM/PBM was used as an indicator because the atmospheric residence time of GOM is widely regarded to be shorter than that of PBM (even though there is no consensus on what specific chemical forms are collected by KCl-coated denuders). The GOM/PBM ratio should be higher if local sources are more important, and the GOM/PBM ratio is likely to decrease as regional transport becomes more important. In this study the GOM/PBM ratios were categorized into three groups: low (0~2), middle (2~8), and high (>8) and the frequency of wind direction was compared (Fig. 6). The result clearly indicates that the southerly and southeasterly winds were associated with high GOM/PBM ratios and

that the westerly and northerly winds indicative of regional transport from China prevailed at lower GOM/PBM ratios. There was a weak negative correlation between the ratio of GOM/PBM and CO concentration at a significance level of 0.1 (p-value= 0.089) (Fig. S3), supporting the assertion that the GOM/PBM ratio decreased with the increased effect of regional transport. Lynam and Keeler (2005) also found that high GOM/PBM ratio was observed with influences from local sources and low GOM/PBM ratios appeared with influence from regional sources in Detroit. In Korea, Kim et al. (2009) found a significant increase in the PBM/GOM ratio during high PM2.5 concentration events caused by regional transport from China.

The reciprocal of this ratio (i.e. GOM/PBM) was used to calculate K_p which indicates that the secondary formation of PBM through gas-particle partitioning was became more important as the significance of regional transport increased, suggesting that secondary production was favored when air underwent regional transport, as has been shown previously (Lynam and Keeler, 2005). It should be noted that there are uncertainties related to different aerosol composition, particle surface area variations, and/or temperature change which can affect both K_p and the GOM/PBM ratio.

3.2.2 PSCF results

In order to locate potential source areas in more detail, PSCF was used. For TGM, potential sources were located in Liaoning, Shandong, and Henan provinces of China along with the southern area of Korea (Fig. 7). Liaoning province, where large non-ferrous smelters are situated, is the province with the largest Hg emissions in China; Shandong and Henan provinces are also large Hg emission areas, emitting about 30-40 ton yr^{-1} (Fu et al., 2012) in part due to a large lead smelter (Wang et al, 2014), and biomass burning (Huang et al., 2011).

The probable source areas of PBM identified by PSCF were similar to those for TGM, indicating that both Chinese and inland Korean sources enhanced PBM concentrations, with the exception of metropolitan (Seoul) and industrial (Incheon) areas located in northwestern South Korea which emerged as more prominent source areas for PBM than for TGM (Fig. 7). Only Korean sources including metropolitan (Seoul) and the industrial areas in southern Korea were identified as probable

source areas for GOM (Fig. 7); regional transport of GOM from China was not important. The Yellow Sea between China and Korea was also associated with high PSCF values, possibly indicating the shipping ports located on the western coast of Korea as important source areas. However, it should be noted that it might be a trailing effect derived by relatively short sampling duration. A trailing effect is often observed, especially with a limited number of measurements or short sampling period, since PSCF evenly distributes weight along the path of trajectories so that PSCF results often identify areas upwind and downwind of real sources as a source area (Han et al., 2007). However, it should be also noted that the marine boundary layer provides good conditions for active Hg oxidation reactions due to an abundance of oxidants (Auzmendi-Murua et al., 2014); therefore, the possibility of areas over the ocean being a GOM source should not be excluded.

It should be noted that different temporal resolutions for trajectories (hourly) and concentrations (every 12hr) were used for GOM and PBM. Since trajectory directions can significantly change over the course of 12hrs there is a possibility that some source areas could be misidentified, especially for more distant regional sources. However, upon investigation it was determined that over the 12 hr sampling periods the trajectories did not diverge significantly at this sampling site for most sampling periods.

3.2.3 Source attribution based on cluster analysis

In an effort to quantify the contribution of national and foreign sources to the measured Hg concentrations the back trajectories were grouped into five clusters using the trajectory cluster analysis feature of HYSPLIT. Among the five clusters, clusters 1 and 5 represent trajectories originating from outside (South) Korea whereas the trajectories grouped in the cluster 4 originated and passed through the (South) Korean peninsula (Fig. 8). Clusters 2 and 3 contain trajectories from China and the Korean peninsula, but cluster 2 was more associated with Liaoning province and North Korea while cluster 3 originated more from Shandong and Henan provinces. Clusters 1 through 5 contributed 12%, 31%, 26%, 20%, and 11% of the total time, respectively, and the associated concentrations with each cluster are shown in Table 4. Concentration ranges of three Hg species for

each cluster were shown as the box-and-whisker plots in the supplementary file (Fig. 5S). When considering that cluster 4 is associated with the local transport from inland Korea and the cluster 1 and 5 are associated with the regional transport from outside of (South) Korea, the maximum and 75th percentile values as well as the arithmetic average are higher in cluster 4 for GOM and PBM than those in the clusters 1 and 5 (Fig. 5S).

The TGM concentration was the highest for cluster 5; however, GOM and PBM concentrations had the lowest averages for this cluster. Cluster 5 contains the back-trajectories originating from Mongolia and Russia and passing through northeastern China before arriving at the sampling site, which suggests regional transport was important for this cluster (Fig. 8). Average CO concentrations were pretty similar for all clusters, but it was the second highest for cluster 5 (cluster 2 was highest). The highest total average GOM and PBM concentrations were associated with cluster 4 which includes trajectories distributed over the Korean peninsula, suggesting that Korean sources were responsible for the enhanced GOM and PBM concentrations. For cluster 4, the highest Pearson correlation coefficient between GOM and PBM concentrations (r=0.721) was observed, indicating that both Hg species were emitted from similar sources. For other clusters, there were no statistically significant correlations between GOM and PBM except for cluster 2 (r=0.209, p-value<0.001). In addition, both average NO_2 concentration (19.4 ±14.9 ppb) and the correlation coefficient between NO_2 and TGM (r= 0.688) were the highest for cluster 4, supporting the finding of impact from Korean sources.

In order to consider both Hg concentration and the fraction of time for each cluster, the following equation was used to quantify the effect of Korean and out-of- Korean sources to the Hg concentration at the receptor site.

$$\text{source contribution of cluster, i} = \frac{(\frac{N_i}{N_{total}})\times C_j}{\sum_{i=1}^{n}\left\{(\frac{N_i}{N_{total}})\times C_j\right\}} \quad \text{(Eq.6)}$$

where (N_i/N_{total}) indicates the percentage of time associated with the cluster, *i*, *i* is the cluster number, *n* is the number of clusters (equal to five in this study), and C_j indicates the average Hg concentration associated with the cluster, *i*. Compared to the other clusters, the source contributions of clusters 1 and 5, which represent regional transport, were relatively low for all Hg species (Table 4). Cluster 4

contributed more significantly, especially for GOM and PBM, indicating the importance of Korean sources. The source contribution of cluster 2 was the highest for PBM compared to other Hg species, suggesting that North Korean sources were an important contributor to the high PBM concentrations measured, likely due to coal and biomass burning in North Korea (Kim et al., 2013; NI, 2001; NI, 2003).

In order to quantify the contribution of Korean vs. out-of-Korean sources (note that "Korean" means "South Korean" throughout the manuscript), the source contributions of the clusters were used. Clusters 1 and 5 were used to represent the effect of sources outside of Korea and the cluster 4 was used to indicate the effect of sources in Korea. Since clusters 2 and 3 contain mixed trajectories from Korea and out-of-Korea their contribution was divided evenly between in and out of Korea. The results indicate that the sources in Korea and outside Korea contributed about 50% each to the concentration variation of TGM measured at the sample site during the sampling period while the Korean sources affected GOM and PBM more significantly, accounting for approximately 52.3% and 53.4%, respectively (Table 4). These results augment the CPF and PSCF results which only use concentrations that are in the top 25^{th} percentile. While CPF and PSCF found that for high concentration events Korean sources were most important for GOM while for TGM and PBM regional transport from China and North Korea were also important, the cluster based approach suggests that for all 3 species Korean and out-of-Korean sources contributed approximately 50% each to the concentration variations seen by the site. When the geometric mean concentrations were used for each cluster a similar result was obtained for relative contributions as the results using arithmetic mean concentrations.

It should be noted that errors always exist in calculating trajectories, causing uncertainties in all trajectory-based approaches. Trajectory errors vary considerably from case to case; Stohl (1998) suggested uncertainties might be 20% of the distance travelled by trajectories while Draxler (1996) found that the final error was about 10% of the travel distance.

4. Conclusion and Implications

This study was initiated to identify the sources affecting speciated mercury concentrations measured on an island located between mainland Korea and Eastern China. Various tools were used to locate and quantify the sources, including correlations with other pollutants, CPF, and the back-trajectory based analysis (PSCF and cluster analysis). The results consistently show that Korean sources are most important for GOM while for other Hg species (TGM and PBM) regional transport from China and North Korea were also important. Existing methods including PSCF and CPF are able to locate the source direction and areas, but do not consider the frequency of the wind directions which can affect the long-term concentrations at the receptor site. For example, if the Hg concentration is high with easterly winds both CPF and PSCF identify the eastern areas as important source areas even if, in fact, winds are rarely blowing from east. In this work, it is true that sources located in the eastern direction from the sampling site are likely to be important for enhancing Hg concentrations, but based only on CPF and PSCF results it cannot be said that their contribution to the concentration variations at the site is also high.

Do address this problem a new approach that considers both the cluster frequency and the Hg concentration associated with each cluster was used to quantify the source contribution at the sampling site. On average, contributions from out-of-Korean sources were similar to Korean sources for TGM whereas Korean sources contributed slightly more to the concentration variations of GOM and PBM compared to the out-of-Korean sources. However, in general, conclusions using this approach are more uncertain when the concentration ranges are similar between clusters. Additional work is needed with this approach to determine if a different statistic (other than mean) would provide better results when there are not distinct concentration differences between clusters. In addition, uncertainties exist in the source attribution approach based on cluster analysis because the trajectories inevitably overlap between different clusters since the cluster analysis accounts for both variations in transport speed and direction simultaneously. Nevertheless, this new approach can augment existing methods including CPF and PSCF to help identify source contributions to the concentration variations at the sampling site.

The ratio of GOM/PBM proved to be a useful tool for identifying the relative significance of local

sources vs. regional transport. The GOM/PBM ratio decreased as the effect of regional transport increased and vice versa since GOM has a shorter atmospheric residence time than PBM. The K_p calculated using the reciprocal of the GOM/PBM ratio was negatively correlated with atmospheric temperature and positively correlated with relative humidity, suggesting that the secondary formation of PBM was an important source of atmospheric PBM concentration at low temperature and high relative humidity. This result also suggests that the secondary formation of PBM becomes more favored when the air undergoes regional transport rather than local transport.

Acknowledgements

This work was funded by the National Research Foundation of Korea (NRF) grant funded by the Korea government (MSIP) (No. 2015R1A2A2A03008301) and the Korea Ministry of Environment (MOE) as "the Environmental Health Action Program". This research was also supported by 2014 Research Grant from Kangwon National University (No. C1011758-01-01).

Author Contributions

The work presented here was carried out in collaboration between all authors. Gang S. Lee analyzed data and wrote the paper. Pyung R. Kim performed the experiments and interpreted the results. Young J. Han defined the research theme, interpreted the results, and wrote the paper. Yong S. Seo, Seung. M. Yi, and Thomas M. Holsen also interpreted the results and approved the final paper.

Conflicts of Interest

The authors declare no conflict of interest.

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855 **Table list**

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Table 1 Summarized concentrations of speciated Hg and other typical pollutants for each sampling period

Sampling periods	TGM (ng m^{-3})	GOM (pg m^{-3})	PBM (pg m^{-3})	SO_2 (ppb)	NO_2 (ppb)	CO (ppm)	O_3 (ppb)	PM_{10} (μg m^{-3})
1st : 2013.01.17 ~ 01.23	3.5 ± 0.8	5.8 ± 8.8	17.0 ± 16.5	5.8 ± 4.2	26.7 ± 12.4	0.5 ± 0.2	21.7 ± 13.5	55.9 ± 38.5
2nd : 2013.02.25 ~ 03.01	3.7 ± 0.9	13.2 ± 14.8	19.5 ± 19.6	7.3 ± 3.0	28.8 ± 19.1	0.7 ± 0.2	28.3 ± 14.7	83.6 ± 28.6
3rd : 2013.04.08 ~ 04.13	2.1 ± 0.4	4.3 ± 5.6	15.6 ± 13.8	4.9 ± 1.7	8.6 ± 3.3	0.4 ± 0.1	46.8 ± 6.0	45.8 ± 21.4
4th : 2013.05.20 ~ 05.25	2.8 ± 1.0	4.2 ± 5.9	6.7 ± 7.3	5.7 ± 2.5	17.2 ± 5.9	0.5 ± 0.1	36.2 ± 19.4	51.9 ± 21.6
5th : 2013.08.19 ~ 08.24	2.3 ± 0.9	3.2 ± 3.0	5.4 ± 4.6	4.7 ± 1.6	10.2 ± 3.8	0.5 ± 0.1	40.9 ± 29.6	34.2 ± 13.7
6th : 2014.03.17 ~ 03.21	2.6 ± 1.2	12.8 ± 15.7	7.3 ± 3.5	7.5 ± 0.1	32.9 ± 75.1	0.4 ± 0.2	39.0 ± 12.5	66.7 ± 44.4
7th : 2014.05.26 ~ 05.31	2.8 ± 0.7	13.5 ± 9.0	9.8 ± 12.2	5.8 ± 2.3	12.4 ± 5.6	0.5 ± 0.1	85.5 ± 23.3	124.5 ± 44.9
8th : 2014.08.19 ~ 08.23	2.4 ± 1.1	10.7 ± 3.7	6.3 ± 3.6	3.4 ± 1.2	8.7 ± 6.0	0.4 ± 0.1	38.7 ± 16.5	32.0 ± 22.3
Average	2.8 ± 1.1	8.3 ± 9.7	10.9 ± 11.2	5.6 ± 5.0	18.2 ± 28.7	0.5 ± 0.2	42.1 ± 25.8	61.9 ± 42.8

Table 2. Comparisons of measured Hg concentrations with those reported in other studies.

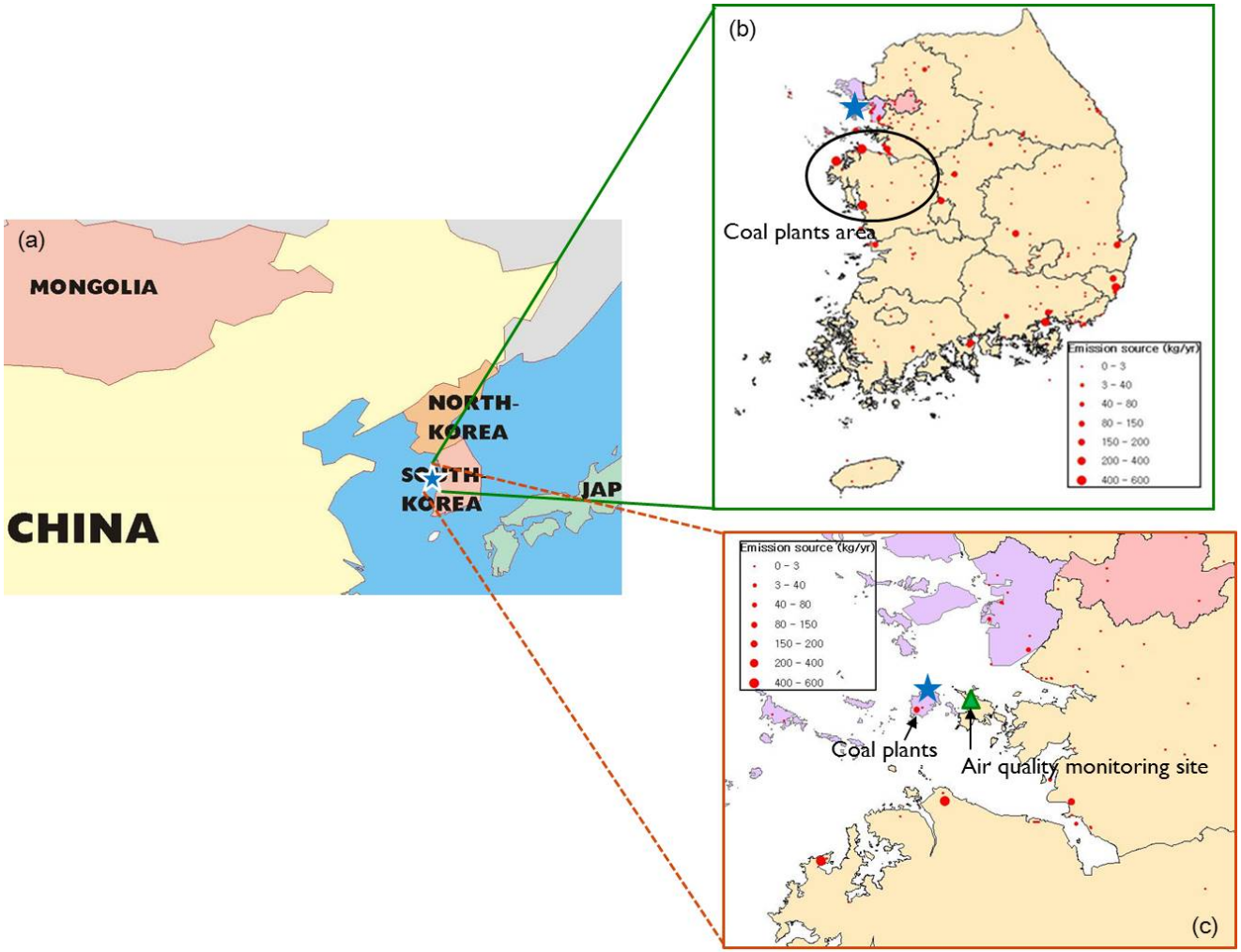
Country	Site	Remarks	Year	TGM (ng m^{-3})	GOM (pg m^{-3})	PBM (pg m^{-3})	Reference
Korea	Seoul	Urban	2005-2006	3.22 ± 2.10	27.2 ± 19.3	23.9 ± 19.6	Kim et al., 2009
	Seoul	Urban	2006-2009	3.72 ± 2.96	11.3 ± 9.5	13.4 ± 12.0	Han et al., 2014
	Chuncheon	Rural	2006-2009	2.12 ± 1.47	2.7 ± 2.7	3.7 ± 5.7	Han et al., 2014
	Yongheung	**Island**	**2013-2014**	**2.8 ± 1.1**	**9.8 ± 9.9**	**10.6 ± 12.0**	**This study**
China	Guiyang	Urban	2009	9.7 ± 10.2	35.7 ± 43.9	368.0 ± 676.0	Fu et al., 2011
	Xiamen	Suburban	2012-2013	3.5	61.05	174.41	Xu et al., 2015
	Mt. Changbai	Remote	2005-2006	3.58 ± 1.78	65 ± 84	77 ± 136	Wan et al., 2009a,b
USA	Chicago	Urban	2007	2.5 ± 1.5	17 ± 87	9 ± 20	Gratz et al., 2013
	Reno, Nevada	Suburban	2007-2009	2.0 ± 0.7	18 ± 22	7 ± 7	Lyman and Gustin, 2009
	Rochester, NY	Urban	2008-2009	1.49	4.08	6.57	Huang et al., 2010
Canada	Nova Scotia	Urban	2010-2011	1.67 ± 1.01	2.07 ± 3.35	2.32 ± 3.09	Cheng et al., 2014
	Ontario	Rural	2006-2007	1.17	15.10	16.40	Baya and Van Heyst, 2010

Table 3. Correlation coefficients and p-values (in parenthesis) for speciated Hg with other pollutants for the whole sampling period. Correlation coefficients with an asterisk indicate a statistically significant relationship at $\alpha = 0.05$.

	TGM	**GOM**	**SO_2**	**NO_2**	**CO**	**O_3**	**PM_{10}**
TGM	-	-0.132 (0.233)	**0.115*** (0.001)	0.063 (0.074)	**0.571*** (<0.001)	-0.055 (0.115)	**0.401****(<0.001)
GOM	-0.132 (0.233)	-	0.025 (0.822)	0.022 (0.846)	-0.149 (0.180)	0.143 (0.197)	0.128 (0.248)
PBM	**0.235*** (0.030)	0.021 (0.855)	-0.006 (0.954)	0.008 (0.941)	**0.215*** (0.048)	0.029 (0.794)	0.139 (0.206)

Table 4. Estimated contribution of Korean and out-of-Korean sources on variations of speciated Hg concentration.
*: TGM is shown in ng m^{-3} while for both GOM and PBM the units are pg m^{-3}.

Cluster	Cluster frequency (%)	Average concentration*			Source contribution (%)			Korean (%)			Out-of-Korean (%)		
		TGM	GOM	PBM	TGM	GOM	PBM	TGM	GOM	PBM	TGM	GOM	PBM
1	12	2.2	6.9	7.7	9.6	10.5	8.6				9.6	10.5	8.6
2	31	2.8	8.3	12.6	31.5	32.5	36.5	15.7	16.3	18.2	15.7	16.3	18.2
3	26	2.9	8.1	10.0	27.3	26.6	24.3	13.7	13.3	12.1	13.7	13.3	12.1
4	20	2.6	9.0	12.3	18.9	22.8	23.0	18.9	22.8	23.0			
5	11	3.2	5.5	7.4	12.8	7.6	7.6				12.8	7.6	7.6
Korean								**48.3**	**52.3**	**53.4**			
Out-of-Korean											**51.7**	**47.7**	**46.6**



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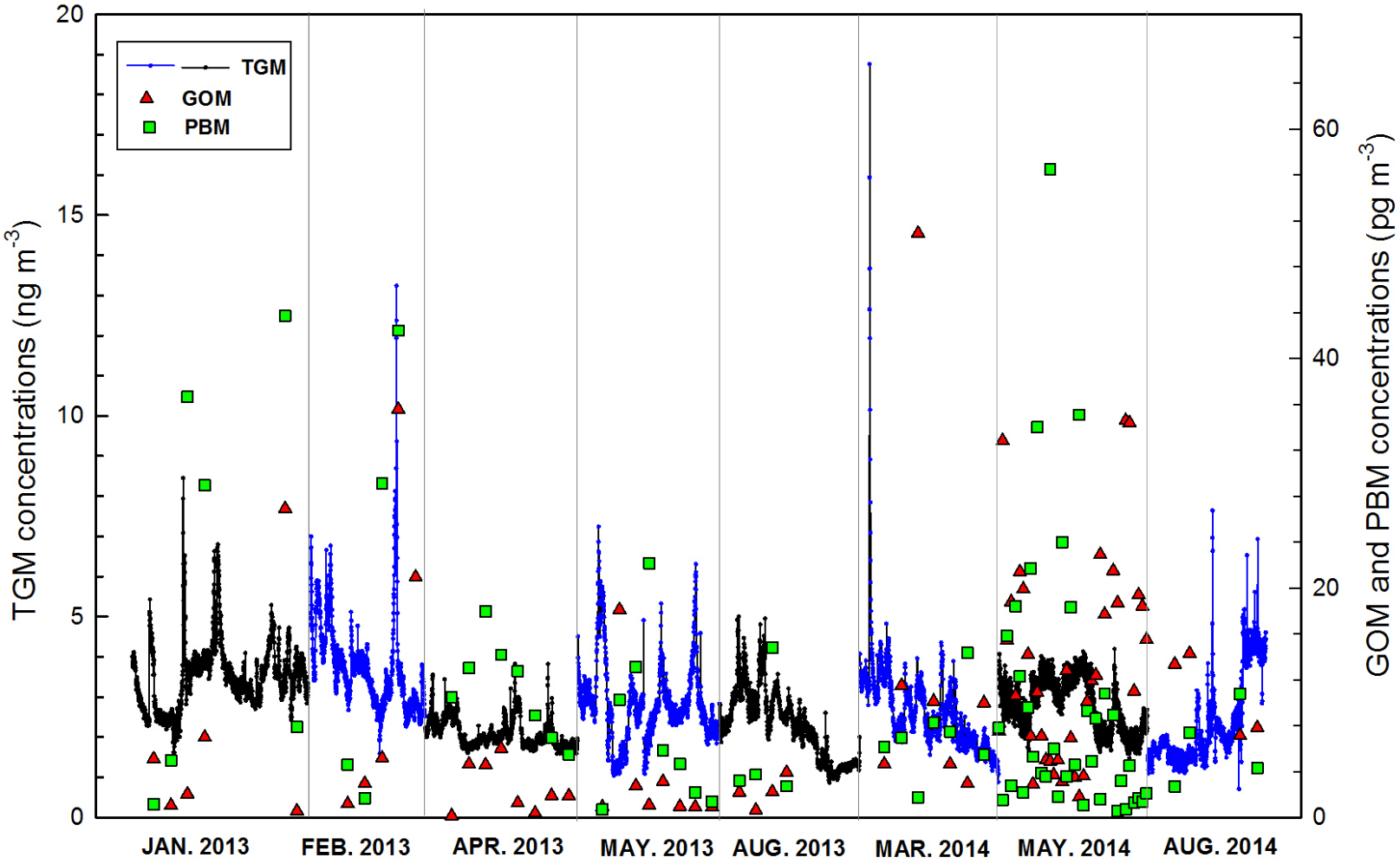

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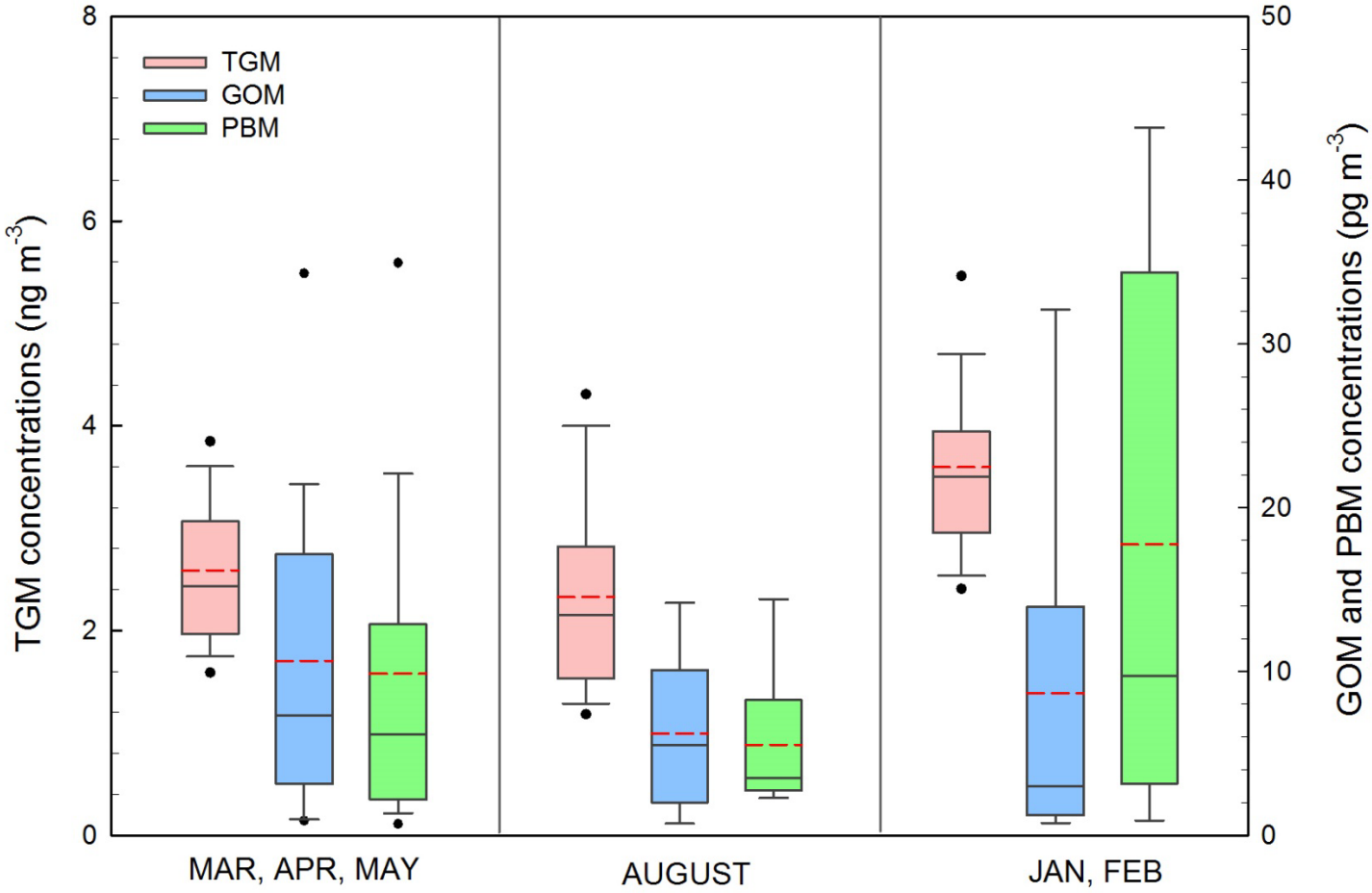

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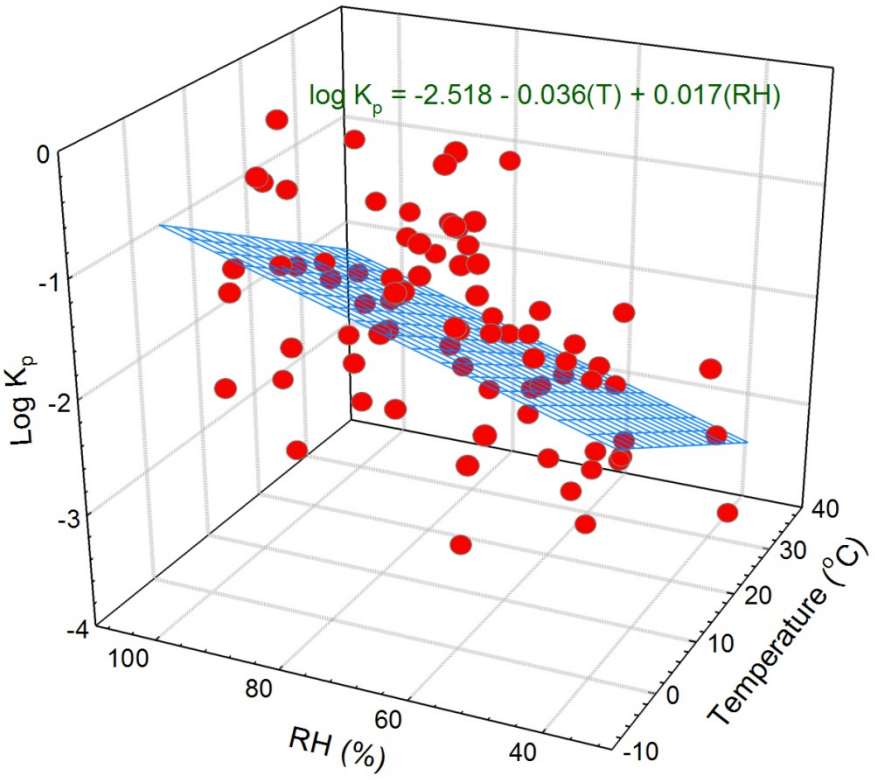

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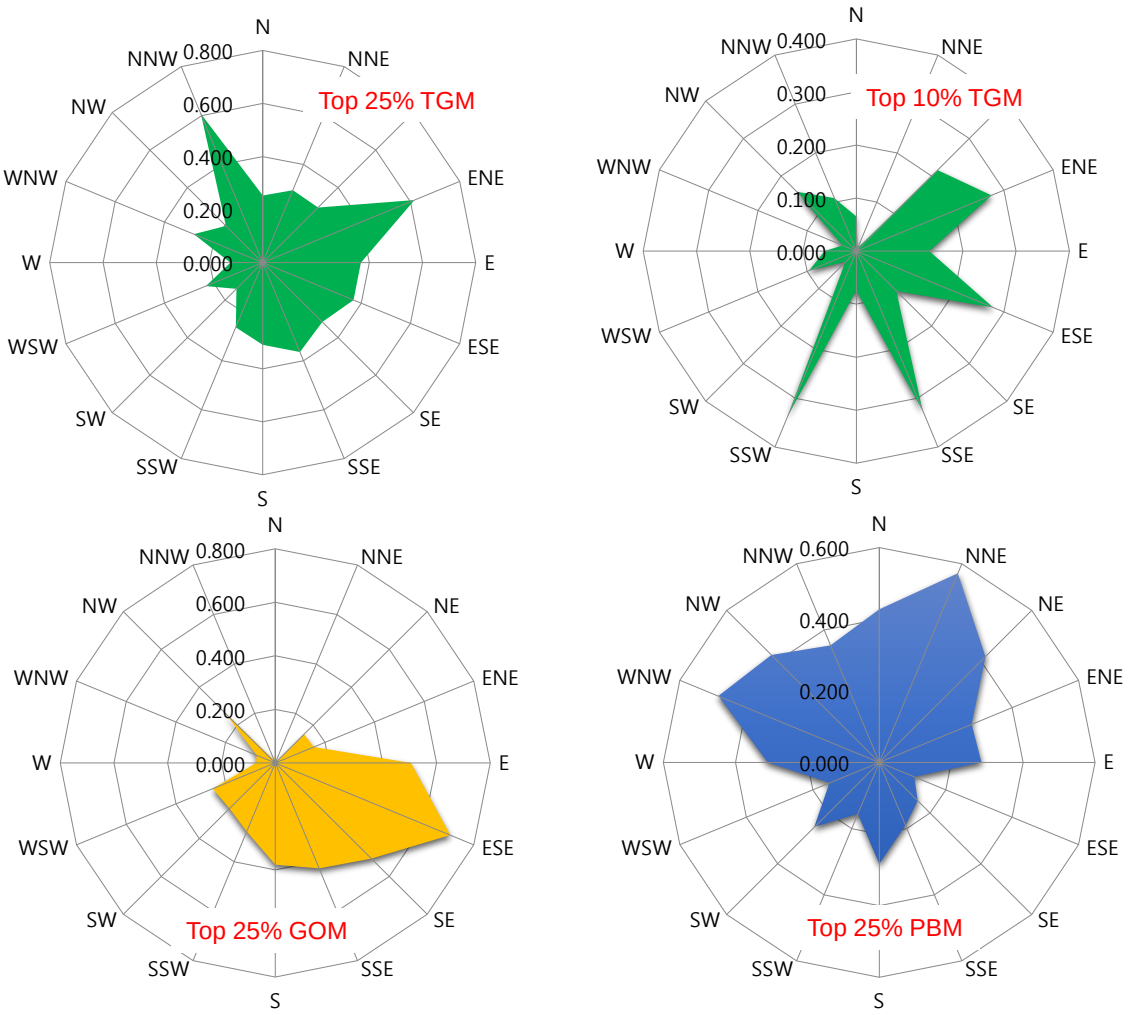

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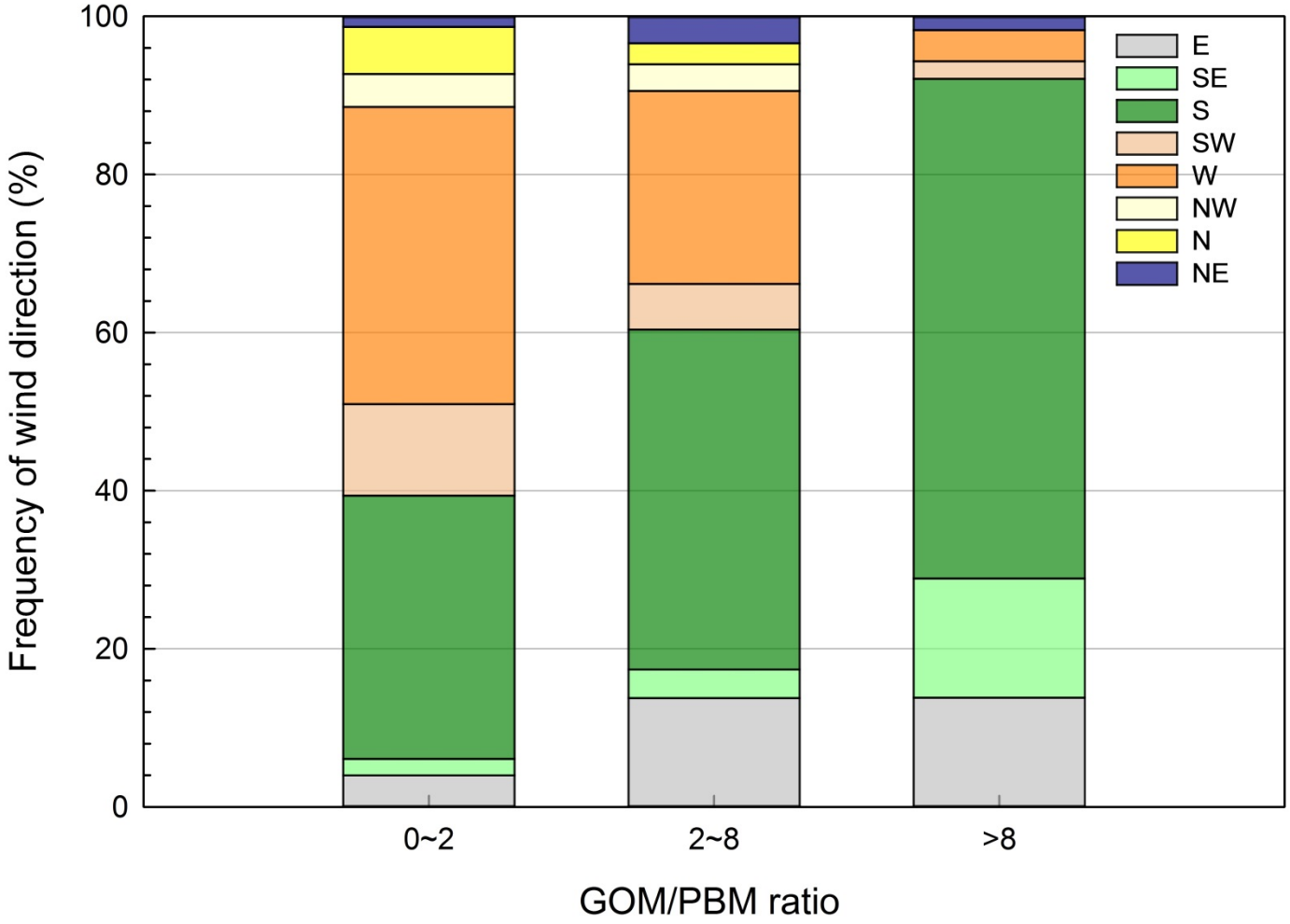

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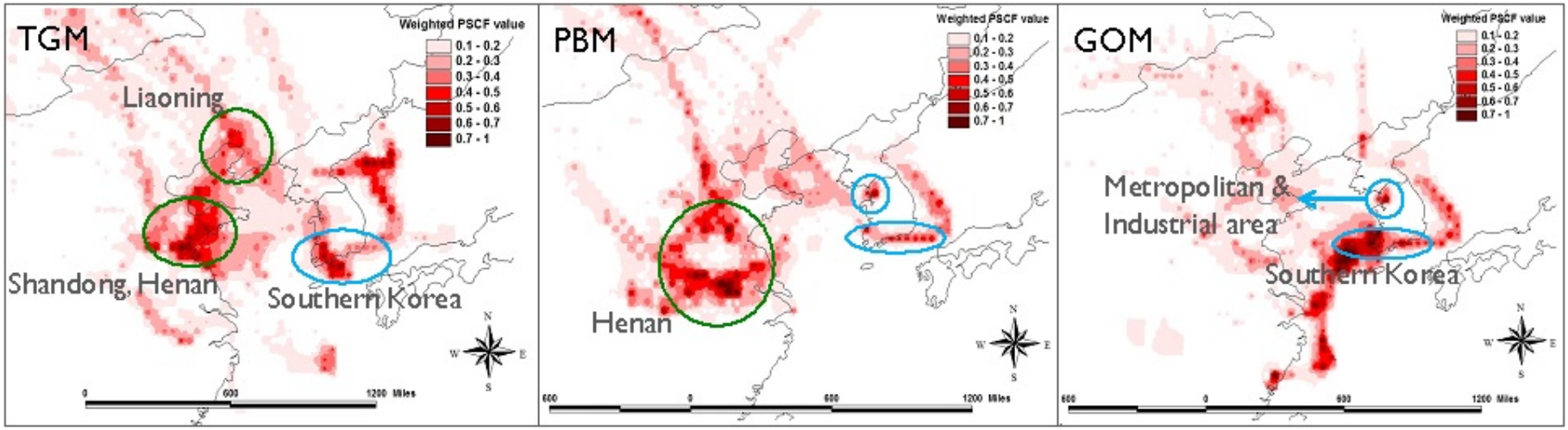

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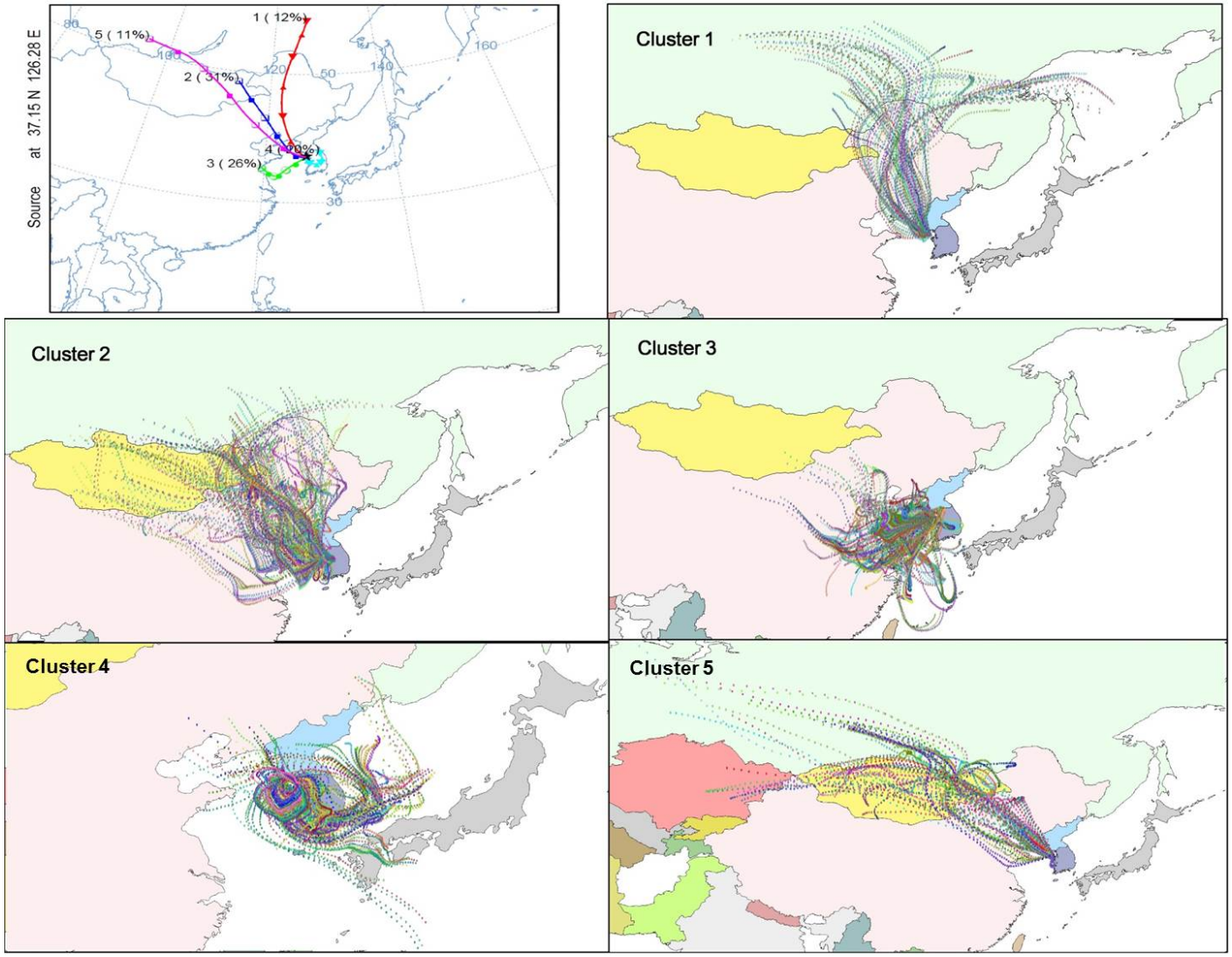

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