

Reviewer 1

[We would like to thank for the reviewer for their valuable comments.](#)

General: This article covers measurements of a biomass burning plume sampled in Toronto using a state-of-the-art LS-SP-AMS, ATOFMS and a variety of other instrumentation. There are a few things that can use highlighting, discussion and clarification and I outline those items below. The experimental section is well-written and referenced. References to an unpublished paper (Willis et al 2015) covering details of an AMS factor analysis make it difficult for me to interpret the factor analysis results presented here. I therefore suggest that the authors present an overview of the PMF results in the supplementary section. The authors can do a better job of stating a clear hypothesis – it is stated that the primary goal of this paper is to examine the mixing state of K, rBC, and BBOA – but why exactly? Also, the quantitative capabilities of the AMS measurement can use some more discussion. With a stated mass resolution of 2000 is it really possible to reliably quantify K mixing state? I know that O:C ratios are potentially quantified with AMS, is that possible here? The strengths of this paper include tying the properties of the different particle types to optical properties and volatility. Overall the paper is interesting and is worthy of publication in ACP and I commend the authors for a solid piece of work.

[Response: Please see our response to these scientific comments below.](#)

Specific:

Abstract: it would be nice to have more background about the potential sources of the BBOA given that these appear to be field measurements. There is no mention of the geographic region in the abstract. What sort of biomass burning plume is this? Is it representative of biomass burning plumes worldwide?

[Response: The abstract has been modified accordingly as shown below:](#)

[“Biomass burning organic aerosol \(BBOA\) can be emitted from natural forest fires and human activities such as agricultural burning and domestic energy generation In this work, single particle measurements from a soot-particle aerosol mass spectrometer coupled with a light scattering module \(LS-SP-AMS\) were performed to examine the mixing state of BBOA, refractory black carbon \(rBC\) and potassium \(\$K^+\$, a tracer for biomass burning aerosol\) in an air mass influenced by wildfire emissions transported from Northern Quebec to Toronto, representing aged biomass burning plumes...”](#)

P. 32160, around line 10: There should be a better stated hypothesis here. Why is it so important to determine the mixing state of K and BBOA? What does it mean to optical and hygroscopic properties if they are internally or externally mixed?

[Response: A few sentences have been added to the Introduction of the revised manuscript as shown below:](#)

“...Potassium (K) is abundant in biomass burning plumes and can be used as a tracer of biomass burning aerosol particles. Potassium salts such as KCl, KNO₃ and K₂SO₄ are hygroscopic in nature, and hence the mixing states of potassium salts, BC and BBOA may affect the cloud formation ability and light scattering of individual biomass burning particles. Furthermore, the mixing state of biomass burning aerosol particles can be potentially used as a tracer of different types of burning. Internal mixtures of organic, BC and potassium salts have been observed in biomass burning aerosol to different extents using electron microscopy techniques (Li et al. 2003).”

P. 32160, lines 14-16: Awkward sentence

Response: This sentence has been revised as shown below.

“The primary objective of this study is to investigate the mixing state of potassium, rBC and organic aerosol in aged biomass burning particles, which in this case originated from wildfire emissions transported from Northern Quebec to Toronto, Ontario.”

P. 32160, lines 16-19: Awkward sentence

Response: This sentence has been revised as shown below.

“We use single particle measurements to provide chemical insight into the volatility and optical properties of biomass burning particles observed in this study.”

P. 32163, line 16-21: It is stated that the PMF results will not be discussed here because they are discussed in Willis et al 2015. However, Willis et al 2015 is in review. This leaves me to conclude that the relevance of these results are dependent on how another paper is peer reviewed. I suggest that the authors work with Willis et al to provide a summary of the PMF analysis in the supplementary section of this paper. I have questions about other factors obtained via PMF (see below) and the reference of Willis et al is useless to me at this point.

Response: In addition to the reference to our previous work, some detail of PMF analysis has been added to Section 2.2 of the revised version (see below). Furthermore, time series and mass spectra of four factor solutions are presented in Figures S1 and S2. Note that Willis et al. (2015) has been accepted for publication in Atmospheric Chemistry and Physics, and that the final version will be published in the near future.

“In brief, the bilinear model was solved using the PMF2 algorithm in robust mode (Paatero and Tapper, 1994) and a final solution was selected using the PMF Evaluation Tool (PET) version 2.06 according to the method described by Ulbrich et al. (2009) and Zhang et al. (2011). A four-factor solution was selected by examining the solutions for up to eight PMF factors. Two factors allowed the separation of local/fresh traffic emissions from more aged/transported aerosol. Three

factors allowed the separation of BBOA from a background oxygenated organic aerosol (OOA). Increasing the number of factors from three to four allowed for the separation of local/fresh traffic emissions into HOA-rich and rBC-rich factors. Moving beyond four factors, splitting and/or mixing of BBOA and background OOA factors became evident, so solutions with more than four factors were not considered. Time series and mass spectra of the four-factor solution are presented in Figures S1 and S2.”

P. 32165, line 16: Were the K⁺ and Org_rBC distributions obtained using the same instrument?

Response: Both K⁺ and Org_{rBC} size distributions were measured by LS-SP-AMS. The manuscript has been revised to avoid potential confusion.

P. 32166, lines 1-4: Was this calculated only for rBC containing particles? This is confusing - I suggest the authors give more detail on this calculation and its interpretation in the supplementary section. It is confusing because the authors are invoking spherical morphology while using effective densities for fractal soot.

Response: The calculation is only for rBC-containing particles. The calculation is only for rBC-containing particles. Our instrument detected rBC-containing particles exclusively because the tungsten vaporizer was removed. Even though spherical particle morphology is assumed in the calculation, different effective densities of rBC are used to simulate the effect of the fractal structure of ambient rBC on the coating thickness (i.e. rBC core with different porosity). The above information has been added or clarified in the revised version of manuscript as shown below. In addition, Figure S4 (Figure S1 in the original version) has been updated to show the calculation results by assuming rBC effective densities of 0.3, 0.8 and 1.4 g/cm³. The coating thickness is relatively insensitive to variations in the effective density of rBC core.

“Since a resistively heated tungsten vaporizer was removed from our LS-SP-AMS, only rBC and its coating materials were detected in this study. Assuming a core-shell structure, a spherical rBC core, a uniform thickness of coating, and a BBOA density of 1.3 g/cm³, the single particle measurements can be used to calculate the coating thickness of the rBC core. To simulate the effect of the fractal structure of ambient rBC, the effective density of rBC was varied between 0.3 and 1.4 g/cm³ (i.e., rBC core with different porosity). Figure S4 presents the two-dimensional histograms (mf_{rBC} vs. particle aerodynamic diameter, d_{va}) of BBOA particles identified by cluster analysis, with calculated coating thickness curves overlaid for comparison. For example, a 320 nm BBOA particle (i.e., d_{va} ~ 400 nm, assuming the effective densities of a porous rBC core and BBOA materials are 0.8 and 1.3 g/cm³, respectively) that consists of 5% of rBC by mass would have a BC core of radius ~68 nm covered by ~92 nm of BBOA material. The coating thickness is relatively insensitive to variations in the effective density of rBC core (Figure S4).”

P. 32166, line 9-14: One should not need to solely rely on cluster analysis to tease out trends of HOA type aerosol. For example, the time series for hydrocarbon like OC marker peaks can be plotted. Does such a time series give the typical HOA temporal signature? Also, only the BBOA

factor from the PMF analysis is shown. Was the PMF analysis able to extract an HOA factor? This would be a salient point of discussion.

Response: The PMF results have been added (Section 2.2 and Figures S1 and S2) to illustrate the presence of HOA in this study. It can be found that the PMF analysis for the wider sampling period (i.e., May 31-June 17, 2015) gives two factors (HOA-rich and rBC-rich factors) that are related to local/fresh traffic emissions. Furthermore, cluster analysis that covered a wider range of the sampling period (i.e., May 31-June 17, 2015) in the same field study could identify a HOA-rich particle class as reported in our previous work (Willis et al., 2015). However, most HOA-rich and rBC-rich particles were not detected in the single particle mode possibly due to their small particle size (i.e., smaller than the cut off size of light scattering detection). The above information has been added to Section 3.2 of the revised manuscript.

“A particle class associated with a high mass fraction of hydrocarbon-like organic aerosol materials (HOA-rich particle class reported in Lee et al. (2015)) was not identified. Given that HOA-rich particles were clearly observed during the BB period based on the PMF results (Section 2.2 and Figures S1 and S2), it is possible that the single particle measurement of LS-SP-AMS actually detected a small number of HOA-rich particles during the BB period but cluster analysis could not separate them from the more dominant BBOA particle classes due to similar mass spectral characteristics. It is also important to note that cluster analysis over a wider range of the sampling period (i.e., May 31-June 17, 2015) in the same field study could identify HOA-rich particle class as reported in our previous work (Willis et al., 2015).”

P. 32166, Line 15: what do the percentages correspond to? Number % or mass % of particle type? Mass % of sulfate?

Response: The percentages correspond to Number %. This information has been added to the text and Table 1.

P. 32166, line 27: How was the potassium deconvoluted from the organics here? Was there enough resolution to distinguish between K and organic fragments? I think that some detail on this would be important for the interpretation of K in some of the other BBOA clusters where K may not be as abundant (but still present). If the authors could provide an example of a deconvolution of K peaks from organic peaks this would provide strong evidence of the quantitative capabilities for the technique.

Response: Cluster analysis was performed using unit mass resolution (UMR) mass spectra, and thus the relative contributions of K^+ and organic fragments to m/z 39 cannot be distinguished. Nevertheless, organic fragment (i.e. $C_3H_3^+$) at m/z 39 only gives a small contribution to the mass spectra of PMF-BBOA factor determined by the high resolution ensemble measurements, suggesting that most ion signal of m/z 39 observed in the BBOA-K particle type is due to the presence of K^+ . In addition, if m/z 39 is used as a proxy for K^+ , the results represent the upper limit of K^+ in each particle class. All the above information has been added to Section 3.3 of the revised manuscript.

“Cluster analysis was performed using UMR mass spectra, and thus the relative contributions of K^+ and organic fragments to m/z 39 cannot be distinguished. Nevertheless, the organic fragment (i.e., $C_3H_3^+$) at m/z 39 only gives a small contribution to the mass spectra of PMF-BBOA factor determined by the high resolution ensemble measurements, suggesting that most of the ion signal at m/z 39 observed in the BBOA-K particle type is due to the presence of K^+ .”

“Assuming that the relative ionization efficiency of K^+ (RIE_K) is equal to 2.9 (Drewnick et al., 2006) and using m/z 39 as a proxy for K^+ , the upper limits of average mass fractions of K^+ for BBOA-K and the other BBOA-related particle types are approximately 20% and $< 1\%$, respectively.”

P. 32167, Line 6: Is it possible that the particles having low K are from sources having low K. Another possibility is that the BBOA having low K is SOA. Please discuss.

Response: We agree that, in addition to primary emissions, large quantities of gas-phase organics co-emitted from biomass burning may lead to significant secondary organic aerosol (SOA) production, contributing to the observed BBOA-related particle types with low K^+ content. This information has been added to Section 3.3 of the revised manuscript.

P. 32167, Lines 16-18: One way to compare the two techniques is to compare the NUMBERS of biomass particles classified by ATOFMS and LS-SP-AMS. Was this attempted? If not, I suggest the authors look into this. Inevitably, the ATOFMS is much more sensitive to K and it is not clear how well K can be deconvoluted from organics for the AMS technique and thus the potassium signal may be below the AMS detection limit. Have any studies looked into the quantitative capabilities of the AMS for potassium for biomass burning aerosols? Discuss.

Response: The ATOFMS and the LS-SP-AMS deployed in this study have very different counting efficiencies in terms of particle number (more than 200,000 vs. 7000 particles, respectively), making the direct comparison of particle number difficult. However, the comparison of the particle number fraction of K-rich particles measured by the two instruments suggests that the LS-SP-AMS may underestimate K-rich particle number. This uncertainty has been added to the revised manuscript.

Similar to the above response, cluster analysis was performed using UMR mass spectra, and thus the relative contributions of K^+ and organic fragments to m/z 39 cannot be distinguished. Nevertheless, we believe that most ion signal of m/z 39 observed in BBOA-K particle type is due to the presence of K^+ based on the mass spectra of PMF-BBOA factor determined by the high resolution ensemble measurements. If m/z 39 is used as a proxy for K^+ , the results represent the upper limit of K^+ in each particle class. Furthermore, we agree that the potassium ion signal observed in other BBOA-related particle classes (i.e. BBOA-1 to 4) may be below the AMS detection limit. Figure S3b shows that only 27% of the total particle number used in the clustering gave ion signals above the detection limit of m/z 39 (~ 2 ions). Lastly, although previous work attempted to quantify potassium mass in aerosol particles using AMS (e.g.,

Drewnick et al., 2006), relative ionization efficiency of potassium (RIE_K) is still uncertain for AMS measurements in general. All of the above information has been discussed in Section 3.3 of the revised manuscript.

Reference:

Drewnick, F., Hings, S. S., Curtius, J., Eerdekens, G. and Williams, J.: Measurement of fine particulate and gas-phase species during the New Year's fireworks 2005 in Mainz, Germany, *Atmos. Environ.*, 40, 4316-4327, 2006.

P. 32167, Line 25: is it possible to quantify O:C?

Response: Based on the empirical parameterization of the improved-ambient method reported by Canagaratna et al. (2015) and measured f_{44} values, the oxygen-to-carbon (O:C) ratio (i.e. $O:C = 0.079 + 4.31 \times f_{44}$) of BBOA-4 is estimated to be 0.43, which is higher than O:C ratios of other BBOA-related particle types (0.14-0.34). In addition to organics, CO^+ and CO_2^+ signals can be generated by rBC (Corbin et al., 2014) and thus our estimates likely represent the upper limit of O:C ratios. This information has been added to Section 3.4 of the revised manuscript.

Reference:

Canagaratna, M. R., Jimenez, J. L., Kroll, J. H., Chen, Q., Kessler, S. H., Massoli, P., Hildebrandt Ruiz, L., Fortner, E., Williams, L. R., Wilson, K. R., Surratt, J. D., Donahue, N. M., Jayne, J. T., and Worsnop, D. R.: Elemental ratio measurements of organic compounds using aerosol mass spectrometry: characterization, improved calibration, and implications, *Atmos. Chem. Phys.*, 15, 253-272, 2015.

Corbin, J. C., Sierau, B., Gysel, M., Laborde, M., Keller, A., Kim, J., Petzold, A., Onasch, T. B., Lohmann, U., and Mensah, A. A.: Mass spectrometry of refractory black carbon particles from six sources: carbon-cluster and oxygenated ions, *Atmos. Chem. Phys.*, 14, 2591-2603, 2014.

P. 32169, Lines 13-17: Is there any difference in volatility for the BBOA-K and the rest? Is it possible that the mixing state of the particles can affect the volatility of the particles?

Response: We did not conduct direct single particle measurements of thermally denuded particles so that we cannot directly comment on the volatility of BBOA-K and other particle types. However, BBOA-K and BBOA-1 are mainly composed of low molecular weight organics, suggesting that the two particle classes may be more volatile compared to BBOA-2, BBOA-3 and BBOA-4 as we highlighted in the last paragraph of Section 3.5. The mixing state may affect the measured volatility if aerosol particles composed of less volatile, highly viscous organics but our measurement cannot directly provide information on this topic.

P. 32170, Line 27: It is stated that the properties of the BBOA particles depends on burning conditions, but would also guess that this is also a strong function of source or fuel type as well. I suggest the authors consider that possibility.

Response: We agree that fuel types can also play a role in affecting chemical properties of BBOA. This information has been added to the revised manuscript.

Reviewer 2

We would like to thank the reviewer for their valuable comments

The paper describes measurements of single particle composition from a wildfire plume in Canada. The authors perform cluster analysis to demonstrate that there were 5 different types of particles in the biomass plume, indicating that the emissions are not internally mixed. They demonstrate that BrC appears to be associated with lower/non volatility organics.

Aerosol optical properties and aerosol mixing state are critical issues to resolve regarding the effects of biomass burning on climate. This paper provides new insight into these attributes of wildfire emissions that are consistent with recent laboratory studies and field work. This is an interesting and well written paper. I recommend that it be published after the authors address the following comments.

Specific

Are there issues with sensitivity / uncertainty of single particle analysis? I am not that familiar with the capabilities of the SP-AMS in this regard. The paper mentions issues with the ATOFMS (e.g. high sensitivity to K) but did not really discuss this with the new results. I don't see this changing the conclusions that there are multiple particle classes in the emissions, but it may alter the relative importance of the different classes. More discussion of the uncertainty of the LS-SP-AMS are needed.

Response: In addition to the optical detection limit of light scattering signal (< 0.2 V), multiple scattering signals observed in a trigger event and particles generating weak total ion signals (< 10 ions) were excluded in cluster analysis. The size distribution of total ion signals of the single particle measurement is shown in Figure S3a and the orange dashed line represents the threshold values (10 ions) for real particles.

Similar to the response to Reviewer 1's comments, the ATOFMS and the LS-SP-AMS deployed in this study have very different counting efficiencies in terms of particle number (more than 200,000 vs. 7000 particles), making the direct comparison of particle number difficult. However, the comparison of particle number fraction of K-rich particles measured by the two instruments suggests that the LS-SP-AMS may underestimate the K-rich particle number. This uncertainty has been added to the revised manuscript.

Cluster analysis was performed using UMR mass spectra, and thus the relative contributions of K^+ and organic fragments to m/z 39 cannot be distinguished. Nevertheless, we believe that most ion signal of m/z 39 observed in BBOA-K particle type is due to the presence of K^+ based on the mass spectra of PMF-BBOA factor determined by the high resolution ensemble measurements. If m/z 39 is used as a proxy for K^+ , the results represent the upper limit of K^+ in each particle class. Furthermore, the potassium ion signal observed in other BBOA-related particle classes (i.e. BBOA-1 to 4) may be below the AMS detection limit. Figure S3b shows that only 27% of the total particle number used in the clustering gave ion signals above the detection limit of m/z 39 (~ 2 ions). Lastly, although previous works attempted to quantify potassium mass in aerosol

particles using AMS (e.g., Drewnick et al., 2006), relative ionization efficiency of potassium (RIE_K) is still uncertain for AMS measurements in general. All of the above information has been discussed in Section 3.3 of the revised manuscript.

Reference:

Drewnick, F., Hings, S. S., Curtius, J., Eerdekens, G. and Williams, J.: Measurement of fine particulate and gas-phase species during the New Year's fireworks 2005 in Mainz, Germany, *Atmos. Environ.*, 40, 4316-4327, 2006.

A shortcoming of the experiment design is the very short residence time (1.9 s) in the thermodenuder. The aerosol will certainly not reach equilibrium in this system. Therefore classifying the residual material as low- or non-volatile may be misleading. It may simply be semivolatile material that has not had sufficient time to evaporate. A simple time scale analysis can be done to estimate whether the system has reached equilibrium (e.g. May et al. JGR 2013 doi:10.1002/jgrd.50828). If it has not this should be noted and the caveat added about the residual mass may not actually be low volatility.

Response: We agree that aerosol particles may not reach thermal equilibrium in our system although the thermodenuder was operated at high temperature (250°C), and thus our measurement only represents the upper limit of low-volatility BBOA materials. This information has been added to Section 3.5 of the revised manuscript.

“However, the short residence time (1.9 s) of ambient aerosol particles in the thermodenuder may result in non-equilibrium measurements and thus overestimate low-volatility aerosol mass (An et al., 2007; May et al., 2012).”

References:

An, W. J., Pathak, R. K., Lee, B. and Pandis, S. N.: Aerosol volatility measurement using an improved thermodenuder: Application to secondary organic aerosol, *J. Aerosol Sci.*, 38, 305-314, 2007.

May, A. A., Levin, E. J. T., Hennigan, C. J., Riipinen, I., Lee, T., Collett, J. L., Jr., Jimenez, J. L., Kreidenweis, S. M. and Robinson, A. L.: Gas-particle partitioning of primary organic aerosol emissions: 3. Biomass burning, *J. Geophys. Res.-Atmos.*, 118, 11327-11338, 2013.

Figure 4 – Not sure how much this figure adds. Four of the diagrams are pretty similar (Org dominated) with one showing contribution of K. The contribution of K in one particle type is illustrated in Figure 3. The limited contribution of rBC was made in text.

Response: A ternary diagram is a common approach to present the mixing state of particles measured by ATOFMS in order to visualize the relative contributions of organic, rBC and K in the BBOA-K particles. For that reason we have decided to keep a ternary diagram of BBOA-K and move it to Figure 3. Due to the limited contributions from the other ternary diagrams, they have been moved to the supplementary information.

Figure 6 – The figure shows regression lines and states r^2 values. I was interested in the value of the slopes and their physical interpretation – these slopes should mean something. Basically this way of presenting the data did not really make sense to me. A bit more text describing how to interpret would be helpful. Also, the caption says estimated mass rBC and non- and low-volatility mass – wasn't this measured downstream of the thermodenuder?

Response: The slopes of the linear regression in Figure 5a (blue dashed line) is 1.38 g/cm^3 , which can be interpreted as the average density of thermo-processed particles assuming all particles are spherical and internally mixed, and is similar to the effective particle densities of biomass burning smoke observed in recent laboratory chamber experiments ($1.18\text{-}1.45 \text{ g/cm}^3$) (Li et al., 2015). In addition, the slope of the linear regression in Figure 5b (blue dashed line) is $2.02 \text{ m}^2/\text{g}$, which can be interpreted as the average mass absorption efficiency (MAE) of thermo-processed particles. The above information has been added to Section 3.5 and Figure 5 of the revised manuscript. Note that the MAE of low-volatility BBOA materials can be determined based on the theoretical closure calculation (please see the last response in this document).

The mass of thermo-denuded particles was estimated by the calculation presented in Section 3.5 of the revised manuscript and Section S6 of the Supplementary Information.

Reference:

Li, C., Ma, Z., Chen, J., Wang, X., Ye, X., Wang, L., Yang, X., Kan, H., Donaldson, D. J. and Mellouki, A.: Evolution of biomass burning smoke particles in the dark, *Atmos. Environ.*, 120, 244-252, 2015.

“It is clear that rBC mass alone cannot explain the observed aerosol volatility ($R^2 = 0.31$, black solid circles in Fig. 6a).” This sentence does not make sense. Are you trying to say that rBC mass by itself cannot explain the mass downstream of thermodenuder? Which then implies that there is other low volatility material beyond rBC? What about non-spherical particle issues and SMPS measurements?

Response: Thank you for pointing this out. Yes, that is what we are trying to say. Considering the uncertainties of SMPS measurement due to the non-spherical and fractal structure of rBC particles, the sentence has been revised to highlight the possibility that rBC mass alone may explain the observed aerosol volume of thermo-processed particles during the BB period only if their effective density is very small (less than 0.3 g/cm^3).

“In contrast, significant enhancement of rBC was not observed during the BB period and rBC only has a weak correlation with the aerosol volume of thermo-processed particles for the entire sampling period ($R^2 = 0.31$, slope = 0.43 g/cm^3 , black solid circles and dashed line in Figure 5a). Note that particle mobility diameter measured by SMPS is a function of particle shape and morphology. The non-spherical and fractal structure of rBC can lead to an underestimation of removal efficiency of aerosol mass in the thermodeunder. Although it seems unlikely, rBC mass alone may explain the observed aerosol volume of thermo-processed particles during the BB

period only if their effective density is less than 0.3 g/cm^3 .”

“the strong absorption characteristics observed during the BB period cannot be explained by the presence of rBC alone, because the rBC loading was roughly constant throughout the sampling period ($R^2 = 0.33$, black solid circles in Fig. 6b).” This is based just on r^2 . Or is it based on optical closure using the measured rBC mass and mass absorption cross section? Need to clarify.

Response: The original argument is based on a high value of the correlation coefficient between absorption and mass of thermo-processed particles. In the revised version, a theoretical closure calculation is performed to estimate the mass absorption efficiency of low-volatility BBOA materials. Assuming the MAE of rBC (MAE_{rBC}) from biomass burning ranged from 5.3 to $6.4 \text{ m}^2/\text{g}$ (Lack et al., 2012), the average MAE of low-volatility of BBOA (MAE_{BBOA}) observed during the BBOA period is approximately equal to $0.8\text{-}1.1(\pm 0.4) \text{ m}^2/\text{g}$ based on a theoretical closure calculation (i.e., absorption at 405 nm measured by PASS-3 = $m_{rBC} (MAE_{rBC}) + m_{BBOA} (MAE_{BBOA})$, where m_i = mass concentration of species i), which is comparable to MAE of primary organic aerosol emitted from biomass burning (e.g., Lack et al., 2012, 2013). The above estimation indicates that the low-volatility BBOA contributes approximately 33-44% of total light absorption at 405 nm of the thermo-processed particles during the BB period. Furthermore, Healy et al. (2015) showed that absorption of ambient and thermally denuded particles at 405 nm are comparable during the BB period, indicating that almost all of the BBOA absorption was associated with low-volatility materials. This information has been added to Section 3.5 of the revised version. The significance of this finding has been highlighted in the Abstract and Conclusion.

References:

Healy, R. M., Wang, J. M., Jeong, C. -, Lee, A. K. Y., Willis, M. D., Jaroudi, E., Zimmerman, N., Hilker, N., Murphy, M., Eckhardt, S., Stohl, A., Abbatt, J. P. D., Wenger, J. C. and Evans, G. J.: Light-absorbing properties of ambient black carbon and brown carbon from fossil fuel and biomass burning sources, *J. Geophys. Res.-Atmos.*, 120, 6619-6633, 2015.

Lack, D. A., Bahreni, R., Langridge, J. M., Gilman, J. B. and Middlebrook, A. M.: Brown carbon absorption linked to organic mass tracers in biomass burning particles, *Atmos. Chem. Phys.*, 13, 2415-2422, 2013.

Lack, D. A., Langridge, J. M., Bahreini, R., Cappa, C. D., Middlebrook, A. M. and Schwarz, J. P.: Brown carbon and internal mixing in biomass burning particles, *Proc. Natl. Acad. Sci. U. S. A.*, 109, 14802-14807, 2012.

Single particle characterization of biomass burning organic aerosol (BBOA): Evidence for non-uniform mixing of high molecular weight organics and potassium

5

A. K. Y. Lee¹, M. D. Willis¹, R. M. Healy^{2,3*}, J. M. Wang³, C.-H. Jeong³, J. C. Wenger², G. J. Evans³, J. P. D. Abbatt¹

[1] {Department of Chemistry, University of Toronto, Toronto, Canada}

10 [2] {Department of Chemistry and Environmental Research Institute, University College Cork, Ireland}

[3] {Southern Ontario Centre for Atmospheric Aerosol Research, Department of Chemical Engineering & Applied Chemistry, University of Toronto, Toronto, Canada}

*Now at Environmental Monitoring and Reporting Branch, Ontario Ministry of the
15 Environment and Climate Change, Toronto, Canada

Corresponding authors: A. K. Y. Lee (alexky.lee@utoronto.ca)

20 **Abstract**

Biomass burning organic aerosol (BBOA) can be emitted from natural forest fires and human activities such as agricultural burning and domestic energy generation. BBOA is strongly associated with atmospheric brown carbon (BrC) that absorbs near ultraviolet and visible light, resulting in significant impacts on regional visibility degradation and radiative forcing. The mixing state of BBOA can play a critical role in the prediction of aerosol optical properties. In this work, single particle measurements from a soot-particle aerosol mass spectrometer coupled with a light scattering module (LS-SP-AMS) were performed to examine the mixing state of BBOA, refractory black carbon (rBC) and

25

potassium (K, a tracer for biomass burning aerosol) in an air mass influenced by wildfire emissions transported from Northern Quebec to Toronto, representing aged biomass burning plumes. Cluster analysis of single particle measurements identified five BBOA-

related particle types. rBC accounted for 3-14 wt% of these particle types on average.

5 Only one particle type exhibited a strong ion signal for K^+ , with mass spectra characterized by low molecular weight organic species. The remaining four particle types were classified based on the apparent molecular weight of the BBOA constituents. Two particle types were associated with low potassium content and significant amounts of high molecular weight (HMW) organic compounds. Our observations indicate non-

10 uniform mixing of particles within a biomass burning plume in terms of molecular weight and illustrate that HMW BBOA can be a key contributor to low-volatility BrC observed

in BBOA particles. The average mass absorption efficiency of low-volatility BBOA is about 0.8-1.1 m^2/g based on a theoretical closure calculation. Our estimates indicate that low-volatility BBOA contributes ~33-44% of thermo-processed particle absorption at 405

15 nm and almost all of the BBOA absorption was associated with low-volatility organics.

1 Introduction

Biomass burning is one of the major global sources of primary particulate matter, contributing approximately 42% and 74% of global black carbon (BC) and primary organic aerosol, respectively (Bond et al., 2004). Although typical climate forcing calculations only consider BC and dust as light absorbing aerosol components, there is growing evidence that atmospheric brown carbon (BrC, i.e., light-absorbing organic particulate matter with wavelength-dependent absorption in the ultraviolet-visible

20 spectral region) is strongly associated with biomass burning organic aerosol (BBOA), resulting in significant impacts on regional visibility and radiative forcing (Kirchstetter et al., 2004; Chen and Bond, 2010; Lack et al., 2012, 2013; Saleh et al., 2014; Lu et al., 2015; Washenfelder et al., 2015).

30 Recent laboratory studies have investigated the physicochemical and optical properties of BrC from biomass burning emissions in order to improve understanding of its climate

impacts. McMeeking et al. (2014) performed a series of thermodenuder experiments to examine volatility and absorption properties of laboratory-generated BBOA associated with BC. Compared to the evaporative loss of aerosol mass, lower removal efficiencies of aerosol absorption at the wavelengths of 405 and 532 nm were observed, implying the presence of less volatile, light absorbing BrC in the thermo-processed particles. Saleh et al. (2014) also demonstrated that the effective absorptivity of BrC that was generated from various biomass burning experiments is a function of the black carbon-to-organic aerosol (BC-to-OA) ratio and is largely associated with extremely low-volatility organic compounds (ELVOC) that could not be removed efficiently by a thermodenuder operated at 250°C. Mohr et al. (2013) recently identified nitrated phenol compounds as potentially important contributors to BrC in BBOA. However, the chemical characteristics of such extremely low-volatility BrC remain largely uncertain.

High molecular weight (HMW) organic compounds can be one of the major contributors to low-volatility BrC in BBOA. A significant number of HMW organic compounds in BBOA have been observed based on different mass spectrometry analyses (e.g., Elias et al., 1999, Smith et al., 2009). More importantly, Di Lorenzo and Young (2016) recently reported that most BrC absorption of BBOA extracts arises from HMW compounds based on size exclusion chromatography. Dinar et al. (2008) demonstrated that the absorptivity of ambient humic-like substance (HULIS, fulvic acid type) separated from BBOA extracts was enhanced with increasing molecular weight (from 340 to 460 Da), and thus decreasing aerosol volatility. Sun et al. (2007) also reported that the molecular structures of organic compounds that result in absorption spectra similar to that of atmospheric BrC are generally highly oxygenated molecules with more than 18 carbon atoms.

Co-emission of BC and BBOA from biomass burning sources can lead to internal mixing of these components. Potassium (K) is abundant in biomass burning plumes and can be used as a tracer of biomass burning aerosol particles. Potassium salts such as KCl, KNO₃ and K₂SO₄ are hygroscopic in nature, and hence the mixing states of potassium salts, BC and BBOA may affect the cloud formation ability and light scattering of individual

biomass burning particles. Furthermore, the mixing state of biomass burning aerosol particles can be potentially used as a tracer of different types of burning. Internal mixtures of organic, BC and potassium salts have been observed in biomass burning aerosol to different extents using electron microscopy techniques (Li et al. 2003). Due to analytical challenges in single particle characterization, there is limited knowledge of the mixing state of HMW organic compounds in BBOA. Our recent field studies have successfully demonstrated that a soot-particle aerosol mass spectrometer (Aerodyne Research Inc., Billerica, MA) equipped with a light scattering module (LS-SP-AMS) can quantify the mixing state of refractory black carbon (rBC) and organic aerosol in urban environments on a single particle basis (Lee et al., 2015; Willis et al., 2015). Here, we extend the application of this single particle measurement technique to investigate the mixing state of BBOA and other co-emitted species. The primary objective of this study is to investigate the mixing state of potassium, rBC and organic aerosol in aged biomass burning particles, which in this case originated from wildfire emissions transported from Northern Quebec to Toronto, Ontario. We use single particle measurements to provide chemical insight into the volatility and optical properties of biomass burning particles observed in this study.

2 Experiment

2.1 Sampling detail and instrumentation

Measurements were performed at the Southern Ontario Centre for Atmospheric Aerosol Research Facility, located in downtown Toronto (43.66°N, 79.40°W). The sampling period was June 12-17, 2013, and encompassed an air mass heavily influenced by biomass burning (June 14, 14:00 to June 16, 11:00) attributed to wildfire events in Quebec (Healy et al., 2015; Willis et al., 2015). The biomass burning influenced air mass took approximately 2 days transporting from Quebec to Toronto based on back trajectory analysis (Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPPLIT) Model). All instruments were connected to a common sampling line (a 10 cm inner diameter stainless steel tube) with a total flow rate of 170 L min⁻¹. The stainless steel sampling line

was connected to a 2.5 μm cut-off inlet, located 15 m from the roadside and 3 m above ground level.

5 A soot-particle aerosol mass spectrometer (Aerodyne Research Inc., Billerica, MA) equipped with a light scattering module (LS-SP-AMS) was the key instrument for this analysis (see Section 2.2-2.3). Other co-located aerosol instruments included an aerosol time-of-flight mass spectrometer (ATOFMS 3800, TSI Inc., Shoreview, MN), a photoacoustic soot spectrometer (PASS-3, Droplet Measurement Techniques, Boulder, CO), an aerosol chemical speciation monitor (ACSM, Aerodyne Research Inc., Billerica, MA) and a scanning mobility particle sizer (SMPS, Model 3936N76, TSI Inc., Shoreview, MN). The ACSM was used to quantify the ambient mass concentration of non-refractory particulate matter (NR-PM) (Ng et al., 2011) with a time resolution of 30 minute in this study (see details in the Supplementary Information). A thermodenuder (Dekati Inc., Finland) operated at 250 $^{\circ}\text{C}$ with a flow rate of 6 L min^{-1} , was used to remove coating materials associated with rBC particles. The residence time for particles in the heated section of the thermodenuder was 1.9 s and the particle transmission efficiency was approximately 72% (Healy et al., 2015). The PASS-3 and SMPS switched between sampling ambient aerosol and denuded aerosol every 10 minutes using an automated sampling valve. The details of ATOFMS, PASS-3 and thermodenuder for this field campaign have been reported in Healy et al. (2015).

2.2 Soot particle aerosol mass spectrometer

A LS-SP-AMS was used to investigate the mixing state of rBC-containing particles on a single particle basis. The working principle of the LS-SP-AMS has been reported elsewhere (Onasch et al., 2012; Lee et al., 2015). In brief, rBC-containing particles are vaporized at ~ 4000 K by a 1064 nm continuous wave intracavity infrared laser similar to that of the single particle soot photometer (SP2, Droplet Measurement Technologies) (Onasch et al., 2012). The resulting vapour is ionized via 70 eV electron impact and then detected by a high-resolution time-of-flight mass spectrometer operated in V-mode, which provides a mass resolving power of ~ 2000 at m/z 28 (DeCarlo et al., 2006;

Canagaratna et al., 2007). A resistively heated tungsten vaporizer was removed from our instrument so that only rBC and its coating materials were detected (Massoli et al., 2015; Willis et al., 2015). A single slit chopper with ~2% throughput was used. The LS-SP-AMS was operated at one-minute time resolution alternating between bulk mass spectrum
5 (MS), particle time-of-flight (pToF) and single particle modes throughout the study.

Direct calibration of the ionization efficiency for nitrate (IE_{NO_3}) is not possible without the tungsten vaporizer. Before removal of the tungsten vaporizer from the LS-SP-AMS, ammonium nitrate (NH_4NO_3) particles generated by a constant output atomizer (TSI Inc.,
10 Model 3076) were dried using a diffusion dryer, and subsequently size-selected at 300 nm using a differential mobility analyzer (DMA, TSI Inc., Model 3081) for determining the mass-based ionization efficiency of nitrate (mIE_{NO_3}). Similarly, the LS-SP-AMS was calibrated for rBC quantification using dried 300 nm Regal Black particles (Regal 400R Pigment Black, Cabot Corp.), a calibration standard recommended by Onasch et al.
15 (2012) to determine the mass-based ionization efficiency of rBC (mIE_{rBC}). The mass of individual dried 300 nm Regal Black particles was ~11.2 fg based on aerosol particle mass analyzer measurements (APM model 3600, Kanomax Inc.) (Willis et al., 2014). The relative ionization efficiency of rBC ($RIE_{rBC} = mIE_{rBC}/mIE_{NO_3}$) was 0.2 ± 0.05 (1σ uncertainty). Assuming that RIE_{rBC} remains unchanged, mIE_{NO_3} and IE_{NO_3} were
20 calculated based on measured values of mIE_{rBC} (after removing the tungsten vaporizer) and the known RIE_{rBC} . The average of the calculated mIE_{rBC} values was 189 ± 20 ions/pg based on four independent calibrations performed throughout the study. The calculated IE_{NO_3} was then used with recommended RIE values (Jimenez et al., 2003, i.e., nitrate = 1.1, sulfate = 1.2, organics = 1.4 and ammonium = 4) to quantify non-refractory aerosol
25 species associated with rBC (referred to as NR- PM_{rBC}). Note that our previous studies have shown that the uncertainties of the recommended RIE for NR- PM_{rBC} for the SP-AMS can be up to 50% (Willis et al., 2014; Lee et al., 2015).

The ensemble MS and pToF data were processed using the AMS data analysis software
30 (Squirrel, version 1.51H for unit mass resolution (UMR) data and Pika, version 1.10H for high resolution peak fitting, <http://cires.colorado.edu/jimenez-group/ToFAMSResources/>

ToFSoftware/index.html) with the corrected air fragment column of the standard fragmentation table (Allan et al., 2004, DeCarlo et al., 2006). Signals for rBC were quantified by the sum of carbon ion clusters (C_x^+ , i.e. C_1^+ - C_9^+) using high-resolution mass spectral data. The average C_1^+ to C_3^+ ratio (0.6) obtained from Regal Black calibration was used to correct the interference in C_1^+ from the non-refractory organics in ambient aerosol. Collection efficiency (CE) for rBC particles was determined in this study using beam width probe measurements described previously (Willis et al 2014). Ambient rBC-containing particles had an average beam width (σ) = 0.46 ± 0.03 mm, and hence a CE of 0.6 was applied for absolute quantification of rBC and NR-PM_{rBC} (Willis et al., 2014). The CE applied does not impact calculations of mass fraction of rBC (mf_{rBC}) in the rBC-containing particles.

Positive matrix factorization (PMF) analysis that covered a wider range of the sampling period (i.e., May 31-June 17, 2015) was performed in our previous work to investigate the potential sources and characteristics of rBC and organic aerosol (Willis et al., 2015). In brief, the bilinear model was solved using the PMF2 algorithm in robust mode (Paatero and Tapper, 1994) and a final solution was selected using the PMF Evaluation Tool (PET) version 2.06 according to the method described by Ulbrich et al. (2009) and Zhang et al. (2011). A four-factor solution was selected by examining the solutions for up to eight PMF factors. Two factors allowed the separation of local/fresh traffic emissions from more aged/transported aerosol. Three factors allowed the separation of BBOA from a background oxygenated organic aerosol (OOA). Increasing the number of factors from three to four allowed for the separation of local/fresh traffic emissions into HOA-rich and rBC-rich factors. Moving beyond four factors, splitting and/or mixing of BBOA and background OOA factors became evident, so solutions with more than four factors were not considered. Time series and mass spectra of the four-factor solution are presented in Figures S1 and S2.

2.3 Single particle measurements and data analysis

The design and working principles of the LS module installed in different versions of the AMS have been previously described in detail (Cross et al., 2009; Freutel et al., 2013; Liu et al., 2013; Lee et al., 2015). The LS module consists of a 405 nm continuous wave laser (CrystaLaser, LC BCL-050-405) that overlaps perpendicularly and completely with the particle beam ~0.265 m downstream of the chopper (Cross et al., 2007). Scattered light from sampled particles is collected using an ellipsoidal mirror and detected with a photomultiplier tube (PMT). A LS signal can be obtained if a sampled particle is larger than the optical detection limit. The 50% cut-off mobility diameter (d_m) (i.e., a diameter that represents 50% counting efficiency of particle number when compared to a condensation particle counter) of the LS module determined using dried NH_4NO_3 particles is ~250 nm, which is approximately equal to a vacuum aerodynamic diameter (d_{va}) of ~340 nm (the density and shape factor of NH_4NO_3 particles are 1.72 g/cm^3 and 0.8, respectively). Once a light scattering signal is detected, the computer is triggered to save the whole pToF trace, from which the single particle mass spectrum is subsequently obtained.

Single particle measurements were analyzed based on a particle categorization procedure, fragmentation table modification and k-means clustering algorithm described previously (Lee et al., 2015), except that m/z 39 (K^+) was also included in the cluster analysis. In addition to the optical detection limit of the LS signal ($< 0.2 \text{ V}$), multiple scattering signals observed in a trigger event and particles generating weak total ion signals (< 10 ions) were excluded in cluster analysis. The size distribution of total ion signals of the single particle measurement is shown in Figure S3a and the orange dashed line represents the threshold values for real particles. The tungsten vaporizer was removed from the LS-SP-AMS, and hence the instrumental background for K^+ predominantly due to surface ionization on the tungsten vaporizer was minimal. LS signals and single particle mass spectra with unit mass resolution (UMR) were processed using the AMS LS data processing software (Sparrow, version 1.04E, <http://cires.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/index.html>). The built-in k-means clustering algorithm in IGOR Pro (WaveMetrics Inc., version 6.2.2.2) was used to analyze single

particle data for up to 25 clusters (Lee et al., 2015). Multiple clusters were manually merged into final particle classes if they exhibited similar mass spectral features and size distributions. Note that k-means clustering has been extensively applied to analyze single particle data measured from the ATOFMS (Rebotier and Prather, 2007; Friedman et al.; 5 2009; Giorio et al., 2012; Gross et al., 2010; Healy et al., 2010; Healy et al., 2013; Pagels et al., 2013) and the results are highly consistent with other clustering algorithms such as ART-2a and hierarchical clustering (Rebotier and Prather, 2007; Giorio et al., 2012).

3 Results and Discussion

10 3.1 Observations of biomass burning influenced air mass

A high loading of rBC-associated organics (referred to as Org_{rBC}) was observed during a period of strong biomass burning influence due to transported wildfire emissions originating from Quebec (Figure 1a, June 14, 14:00 to June 16, 11:00, referred to as “BB period”) (Healy et al., 2015). Potassium (K^+ , a tracer of BBOA) measured by the LS-SP-AMS is in good agreement with the K-containing particle number counts measured by the 15 ATOFMS, and is strongly correlated with increased levels of Org_{rBC} (Figure 1a-c), indicating a significant contribution of BBOA to the total Org_{rBC} during the BB period. This is in good agreement with results of a PMF analysis, which indicate that a PMF-BBOA factor correlates well with K^+ (Figure 1b) (Willis et al., 2015). The average mass 20 fractions of m/z 60 (a marker ion for BBOA) and 44 to total organic (f_{60} and f_{44}) during the BB period were 0.013 and 0.09, respectively, which is consistent with previous field observations for BBOA (Cubison et al., 2011). Even though K^+ and Org_{rBC} observed within the BB period likely have the same origin, different peak positions in K^+ and 25 Org_{rBC} size distributions measured by the LS-SP-AMS suggest non-uniform mixing of these two aerosol components in biomass burning emissions (Figure 1d).

3.2 Single particle observations – BBOA particle types

Cluster analysis of LS-SP-AMS single particle mass spectra was performed to investigate the mixing states of K^+ , Org_{rBC} and rBC, yielding distinct particle classes (Table 1).

Approximately 90% of the particles were classified into five BBOA-related particle types, namely BBOA-K, BBOA-1, BBOA-2, BBOA-3, and BBOA-4 (see discussion below). rBC was only a minor constituent of these particle classes ($mf_{rBC} = 0.03-0.14$), which is in agreement with the PMF-BBOA factor. Since a resistively heated tungsten vaporizer was removed from our LS-SP-AMS, only rBC and its coating materials were detected in this study. Assuming a core-shell structure, a spherical rBC core, a uniform thickness of coating, and a BBOA density of 1.3 g/cm^3 , the single particle measurements can be used to calculate the coating thickness of the rBC core. To simulate the effect of the fractal structure of ambient rBC, the effective density of rBC was varied between 0.3 and 1.4 g/cm^3 (i.e., rBC core with different porosity). Figure S4 presents the two-dimensional histograms (mf_{rBC} vs. particle aerodynamic diameter, d_{va}) of BBOA particles identified by cluster analysis, with calculated coating thickness curves overlaid for comparison. For example, a 320 nm BBOA particle (i.e., $d_{va} \sim 400 \text{ nm}$, assuming the effective densities of a porous rBC core and BBOA materials are 0.8 and 1.3 g/cm^3 , respectively) that consists of 5% of rBC by mass would have a BC core of radius $\sim 68 \text{ nm}$ covered by $\sim 92 \text{ nm}$ of BBOA material. The coating thickness is relatively insensitive to variations in the effective density of rBC core (Figure S4).

An rBC-rich particle type with mf_{rBC} of $\sim 0.8-0.9$ was identified in our previous observations in downtown Toronto likely arising from fresh traffic exhaust (Lee et al., 2015; Willis et al., 2015). rBC-rich particles accounted for only about 1% of the total particles detected during this biomass burning episode, in part due to the lower cut-off size limitation of the LS module (Lee et al., 2015). A particle class associated with a high mass fraction of hydrocarbon-like organic aerosol materials (HOA-rich particle class reported in Lee et al. (2015)) was not identified. Given that HOA-rich particles were clearly observed during the BB period based on the PMF results (Section 2.2 and Figures S1 and S2), it is possible that the single particle measurement of LS-SP-AMS actually detected a small number of HOA-rich particles during the BB period but cluster analysis could not separate them from the more dominant BBOA particle classes due to similar mass spectral characteristics. It is also important to note that cluster analysis over a wider range of the sampling period (i.e., May 31-June 17, 2015) in the same field study could

identify HOA-rich particle class as reported in our previous work (Willis et al., 2015). Accumulation mode SO₄-rich and NO₃-rich (3 and 5 % of particle number) particle classes with low rBC content ($m_{f_{rBC}} < 0.02$) were also identified, similar to our previous observations in downtown Toronto (Lee et al., 2015).

5

3.3 Mixing of K⁺ in BBOA

The five BBOA-related particle classes give relatively strong total ion signals within the BB period as illustrated in Figure 1c, which is in good agreement with the PMF-BBOA factor and K⁺ (Figure 1b). Figure 2 demonstrates the average mass spectra (grey and green trace) of each BBOA-related particle class and the mass spectral histograms (normalized to the number of particles in each particle class, red trace) that show the relative frequency of a signal at each m/z. Only one particle type, BBOA-K, exhibits a strong ion signal at m/z 39 with mass spectra characterized by low molecular weight organic fragment ions (Figure 2a). The ion fraction of m/z 39 for BBOA-K and other BBOA-related particle classes are 0.33 and less than 0.02, respectively (Figure 3a and Table 1). The inset in Figure 3a presents the ion fraction distribution of m/z 39 for BBOA-K (ranging from 0.1 to 0.8). Cluster analysis was performed using UMR mass spectra, and thus the relative contributions of K⁺ and organic fragments to m/z 39 cannot be distinguished. Nevertheless, the organic fragment (i.e., C₃H₃⁺) at m/z 39 only gives a small contribution to the mass spectra of PMF-BBOA factor determined by the high resolution ensemble measurements, suggesting that most of the ion signal at m/z 39 observed in the BBOA-K particle type is due to the presence of K⁺.

10

15

20

Ternary diagrams further illustrate the relative abundance of K⁺ (representing the upper limit), rBC and organic ion signals in BBOA-K (Figure 3b) and other BBOA-related particle types (Figure S5). These observations highlight that (1) K⁺ is not uniformly mixed in BBOA although it has been extensively used as a tracer for BBOA, and (2) a large number fraction of BBOA particles are characterized by low K⁺ and rBC content. In addition to primary emissions, large quantities of gas-phase organics co-emitted from biomass burning may lead to significant secondary organic aerosol (SOA) production,

30

contributing to the observed BBOA-related particle types with low K^+ content. Assuming that the relative ionization efficiency of K^+ (RIE_K) is equal to 2.9 (Drewnick et al., 2006) and using m/z 39 as a proxy for K^+ , the upper limits of average mass fractions of K^+ for BBOA-K and the other BBOA-related particle types are approximately 20% and < 1%, respectively.

While BBOA-K accounted for ~15% of the detected particle number (Table 1), our complementary ATOFMS measurements revealed that particles with a strong K^+ signal accounted for over 60% of the detected particle number for the same period. The comparison suggests that the LS-SP-AMS measurement may underestimate potassium-rich particle number. However, ATOFMS is highly sensitive towards K^+ (Gross et al., 2000), thus even particles with low potassium mass content can exhibit high positive ion intensities for K^+ but chemical matrix effects make the mass quantification of K^+ difficult. On the other hand, for the LS-SP-AMS measurement, only 27% of the total particle number used in the clustering gave ion signals above the detection limit of m/z 39, which is approximately 2 ions as shown in Figure S3b. Furthermore, RIE_K is still uncertain for AMS measurements in general. Note that ATOFMS and LS-SP-AMS cannot detect BBOA particles that are without any components absorbing at a wavelength of 266 nm and in the absence of rBC, respectively. Overall, a direct comparison between the two measurements is not straightforward as the mass quantifications of K^+ at the single particle level for both instruments remain highly uncertain.

3.4 Mixing of high molecular weight organic compounds in BBOA

The remaining four BBOA particle types were classified according to the apparent molecular weight of BBOA constituents (Figure 2b-2e). The average value of f_{60} for BBOA-4 (0.004) is smaller than those of other BBOA particle types (ranging from 0.008 to 0.015). Note that m/z 44 (a tracer ion for organic acids) is the main contributor to BBOA-4, whereas other BBOA particle types are dominated by ions at m/z 29 and 43, suggesting that BBOA-4 is more oxidized relative to other BBOA particles. Based on the empirical parameterization of the improved-ambient method reported by Canagaratna et

al. (2015) and measured f_{44} values, the oxygen-to-carbon (O:C) ratio (i.e., $O:C = 0.079 + 4.31 \times f_{44}$) of BBOA-4 is estimated to be 0.43, which is higher than O:C ratios of other BBOA-related particle types (0.14-0.34). In addition to organics, CO^+ and CO_2^+ signals can be generated by rBC (Corbin et al., 2014) and thus our estimates likely represent the upper limit of O:C ratios of each BBOA-related particle types.

Figure 4a illustrates the normalized cumulative histograms of mass-to-charge ratios (from m/z 12 to 628, excluding m/z 39 and fragments with strong interferences from air and ammonium, see Lee et al. (2015)) for each BBOA particle class, all BBOA particle types and ensemble Org_{rBC} within the BB period. Two particle types, BBOA-3 and BBOA-4, are associated with significant amounts of HMW organic fragments (i.e., more than 50% of the ion signal arose from organic fragments at m/z 100 or larger). For BBOA-2, ~23% of the ion signal resides in HMW organic fragments. BBOA-K and BBOA-1 are mainly composed of low molecular weight organics with over 96% of their ion signal arising from organic fragments smaller than m/z 100. Further, a significant number of observed BBOA particles are composed of HMW organic fragments, up to 80% by mass (Figure 4b). These observations demonstrate that (1) the mixing state of BBOA is not uniform in terms of organic molecular weight, and (2) a large number fraction of BBOA single particle mass spectra (~40%) has major contributions from HMW organic fragment ions, which are detected even though a hard ionization technique (electron impact) was used in the LS-SP-AMS.

HMW mass spectral signatures, some of which are associated with polycyclic aromatic hydrocarbons, were also observed in the ATOFMS data for the same period (see Figure S8 for the average ATOFMS mass spectrum). However, those spectra only accounted for ~0.5% of the total number of spectra detected by the ATOFMS during the BB period. Even though the ATOFMS has been previously demonstrated to detect HMW oligomeric species in both ambient air and chamber experiments (Gross et al., 2006; Denkenberger et al., 2007; Healy et al., 2010), its sensitivity towards HMW organics may be influenced by chemical matrix effects, making the detection and quantification of HMW organics in ambient particles challenging (Reilly et al., 2000).

Although the ensemble size distributions of K^+ and Org_{rBC} are observably different (Figure 1d), the size distributions of BBOA-K and other BBOA-related particle classes are similar to each other except that BBOA-4 peaks at a slightly larger particle diameter (the insets in Figure 2a-e), likely due to measurement bias of the LS module towards larger particle sizes (Lee et al., 2015). Nevertheless, it is worth noting that there is no significant difference between the ensemble size distributions of low ($< m/z$ 100) and high ($> m/z$ 100) molecular weight organics (Figure 1d). This suggests that ensemble size distributions are insufficient to visualize external mixing of the BBOA particle types with different average molecular weight, and also highlights the relevance and unique capabilities of single particle mass spectrometry observations.

3.5 Volatility and optical properties of BBOA

To assess the volatility of BBOA, the mass fraction of total BBOA, regardless of rBC content, retained in thermo-processed particles was estimated based on the SMPS and ACSM measurements (see details in Supplementary Information). The removal efficiency of aerosol volume, based on SMPS measurements, was 60-80% for the majority of the sampling period and decreased to ~40% on June 15, when the site was most heavily impacted by the wildfire emissions (Figure S6a). The estimation based on ACSM measurements also illustrates that approximately 60% of the total BBOA was retained in the thermo-denuded particles on June 15 to match the SMPS observations, given the fact that BBOA dominated the total aerosol mass on June 15 (Figure S6b, see details in Supplementary Information). However, the short residence time (1.9 s) of ambient aerosol particles in the thermodenuder may result in non-equilibrium measurements and thus overestimate low-volatility aerosol mass (An et al., 2007; May et al., 2012).

A strong correlation between the thermo-denuded SMPS aerosol volume and the estimated low-volatility aerosol mass is observed for the entire sampling period ($R^2 = 0.93$, blue open circles in Figure 5a) when all rBC and ~60% of the total BBOA is

assumed to remain after the thermodenuder. The slope of the linear regression (blue dashed line) is 1.38 g/cm³, which can be interpreted as the average density of thermo-processed particles assuming all particles are spherical and internally mixed, and is similar to the effective particle densities of biomass burning smoke observed in recent laboratory chamber experiments (1.18-1.45 g/cm³) (Li et al., 2015). In contrast, significant enhancement of rBC was not observed during the BB period and rBC only has a weak correlation with the aerosol volume of thermo-processed particles for the entire sampling period ($R^2 = 0.31$, slope = 0.43 g/cm³, black solid circles and dashed line in Figure 5a). Note that particle mobility diameter measured by SMPS is a function of particle shape and morphology. The non-spherical and fractal structure of rBC can lead to an underestimation of removal efficiency of aerosol mass in the thermodeunder. Although it seems unlikely, rBC mass alone may explain the observed aerosol volume of thermo-processed particles during the BB period only if their effective density is less than 0.3 g/cm³.

The absorption of thermo-denuded aerosol at a wavelength of 405 nm (B_{abs}) was measured by the PASS-3 (Figure S6d and 5b). Similar to the case of aerosol volatility, the strong absorption characteristics observed during the BB period cannot be explained by the presence of rBC alone, because the rBC loading was roughly constant throughout the sampling period ($R^2 = 0.33$, black solid circles in Figure 5b). The correlation between thermo-denuded particles and aerosol absorption at 405 nm is improved significantly if low-volatility BBOA (60% of the total BBOA) is included in the calculation ($R^2 = 0.84$, blue open circles in Figure 5b). The slope of the linear regression to these data (blue dashed line in Figure 5b) is 2.02 m²/g, which can be interpreted as the average mass absorption efficiency (MAE) of thermo-processed particles at 405 nm.

Our previous work reported that the absorption enhancement attributable to lensing at 781 nm was not observed for rBC during the BB period based on the PASS-3 measurement (Healy et al., 2015). Furthermore, a recent study has reported that the interference on absorption measurement due to pyrolyzed ambient organic aerosol in thermodenuder operated at 250 °C is minimal (Liu et al., 2015). Therefore, these

observations suggest that low-volatility BBOA components could be a key contributor to BrC in the thermo-denuded particles. Assuming the MAE of rBC (MAE_{rBC}) from biomass burning ranged from 5.3 to 6.4 m^2/g (Lack et al., 2012), the average MAE of low-volatility of BBOA (MAE_{BBOA}) observed during the BBOA period is approximately equal to 0.8-1.1(± 0.4) m^2/g based on a theoretical closure calculation (i.e., absorption at 405 nm measured by PASS-3 = $m_{rBC} (MAE_{rBC}) + m_{BBOA} (MAE_{BBOA})$, where m_i = mass concentration of species i), which is comparable to MAE of primary organic aerosol emitted from biomass burning (e.g., Lack et al., 2012, 2013). The above estimation indicates that the low-volatility BBOA contributes approximately 33-44% of thermo-processed particle absorption at 405 nm during the BB period. Furthermore, Healy et al. (2015) showed that absorption of ambient and thermally denuded particles at 405 nm are comparable during the BB period, suggesting that almost all of the BBOA absorption was associated with low-volatility materials (Saleh et al., 2014; Di Lorenzo and Young, 2016).

Lastly, it is particularly important to note that the ion fractions of BBOA-2, BBOA-3 and BBOA-4 from the single particle measurements accounted for ~61% of the total ion signal, which is equivalent to aerosol mass as most of the observed BBOA particles are dominated by organics, from all BBOA-related particle classes during the BB period. This implies that medium to high molecular weight organic compounds may contribute significantly to low-volatility BrC (i.e. 60% of the total BBOA) in the biomass burning affected atmosphere.

4 Conclusions and atmospheric implications

This is the first study to quantify the mixing state of BBOA, K^+ and rBC in ambient biomass burning aerosol using the LS-SP-AMS. BBOA-related particle types are composed of low rBC mass ($mf_{rBC} = 0.03-0.14$). Even though K^+ has been considered as a tracer of BBOA, our results demonstrate that K^+ is not uniformly mixed in BBOA and that over 80% of BBOA particles observed here are characterised by low K^+ content (i.e., < 1 wt% by assuming RIE_K of 2.9). High molecular weight organic compounds were clearly observed in some BBOA particles. A large fraction of BBOA particles (BBOA-3

and BBOA-4) detected in this study is composed of HMW organic compounds with low rBC and K⁺ content.

Thermodenuder experiments indicate the presence of low-volatility BrC in BBOA. Based on SMPS and aerosol mass spectrometry measurements, we estimated that about 60% of the total BBOA mass was retained in the thermo-processed particles. This is consistent with recent observations, which demonstrated that laboratory-generated BBOA particles could be composed of extremely low-volatility BrC depending on the burning conditions and fuel types (Saleh et al., 2014). Liu et al. (2015) also observed low-volatility secondary BrC that was associated with large rBC particles in wintertime UK. Notably, BBOA-2, BBOA-3 and BBOA-4 accounted for ~61% of ion signals from BBOA-related particle classes during the BB period, highlighting that medium to high molecular weight organic compounds may contribute significantly to low-volatility BrC. The average mass absorption efficiency of low-volatility BBOA is about 0.8-1.1 m²/g based on a theoretical closure calculation. The estimation indicates that the low-volatility BBOA contributes ~33-44% of the thermo-processed particles absorption at 405 nm during the BB period and almost all of the BBOA absorption was associated with low-volatility organics. The observed non-uniform mixing state of BBOA in terms of molecular weight also suggests that it may be inappropriate to assume uniform distributions of aerosol volatility and absorption for BBOA particles.

Acknowledgement

This work was financially supported by Natural Sciences and Engineering Research Council (NSERC) of Canada, the Canada Foundation for Innovation and the Marie Curie Action FP7-PEOPLE-IOF-2011 (Project: CHEMBC, No. 299755).

References:

Allan, J., Delia, A., Coe, H., Bower, K., Alfarra, M., Jimenez, J., Middlebrook, A., Drewnick, F., Onasch, T., Canagaratna, M., Jayne, J., and Worsnop, D.: A generalised

- method for the extraction of chemically resolved mass spectra from Aerodyne aerosol mass spectrometer data, *J. Aerosol Sci.*, 35, 909-922, 2004.
- 5 An, W. J., Pathak, R. K., Lee, B. and Pandis, S. N.: Aerosol volatility measurement using an improved thermodenuder: Application to secondary organic aerosol, *J. Aerosol Sci.*, 38, 305-314, 2007.
- Bond, T. C., Streets, D. G., Yarber, K. F., Nelson, S. M., Woo, J. H., and Klimont, Z.: A technology-based global inventory of black and organic carbon emissions from combustion, *J. Geophys. Res.-Atmos.*, 109, D14203, DOI: 10.1029/2003JD003697, 2004.
- 10 Canagaratna, M. R., Jayne, J. T., Jimenez, J. L., Allan, J. D., Alfarra, M. R., Zhang, Q., Onasch, T. B., Drewnick, F., Coe, H., Middlebrook, A., Delia, A., Williams, L. R., Trimborn, A. M., Northway, M. J., DeCarlo, P. F., Kolb, C. E., Davidovits, P., and Worsnop, D. R.: Chemical and microphysical characterization of ambient aerosols with the aerodyne aerosol mass spectrometer, *Mass Spectrom. Rev.*, 26, 185-222, 2007.
- 15 Canagaratna, M. R., Jimenez, J. L., Kroll, J. H., Chen, Q., Kessler, S. H., Massoli, P., Hildebrandt Ruiz, L., Fortner, E., Williams, L. R., Wilson, K. R., Surratt, J. D., Donahue, N. M., Jayne, J. T., and Worsnop, D. R.: Elemental ratio measurements of organic compounds using aerosol mass spectrometry: characterization, improved calibration, and implications, *Atmos. Chem. Phys.*, 15, 253-272, 2015.
- 20 Chen, Y., and Bond, T. C.: Light absorption by organic carbon from wood combustion, *Atmos. Chem. Phys.*, 10, 1773-1787, 2010.
- Cross, E. S., Onasch, T. B., Canagaratna, M., Jayne, J. T., Kimmel, J., Yu, X. -Y., Alexander, M. L., Worsnop, D. R., and Davidovits, P.: Single particle characterization using a light scattering module coupled to a time-of-flight aerosol mass spectrometer, *Atmos. Chem. Phys.*, 9, 7769-7793, 2009.
- 25 Corbin, J. C., Sierau, B., Gysel, M., Laborde, M., Keller, A., Kim, J., Petzold, A., Onasch, T. B., Lohmann, U., and Mensah, A. A.: Mass spectrometry of refractory black carbon particles from six sources: carbon-cluster and oxygenated ions, *Atmos. Chem. Phys.*, 14, 2591-2603, 2014.
- 30 Cross, E. S., Slowik, J. G., Davidovits, P., Allan, J. D., Worsnop, D. R., Jayne, J. T., Lewis, D. K., Canagaratna, M., and Onasch, T. B.: Laboratory and ambient particle density determinations using light scattering in conjunction with aerosol mass spectrometry, *Aerosol Sci. Technol.*, 41, 343-359, 2007.
- 35 Cubison, M. J., Ortega, A. M., Hayes, P. L., Farmer, D. K., Day, D., Lechner, M. J., Brune, W. H., Apel, E., Diskin, G. S., Fisher, J. A., Fuelberg, H. E., Hecobian, A., Knapp, D. J., Mikoviny, T., Riemer, D., Sachse, G. W., Sessions, W., Weber, R. J., Weinheimer, A. J., Wisthaler, A., and Jimenez, J. L.: Effects of aging on organic aerosol

- from open biomass burning smoke in aircraft and laboratory studies, *Atmos. Chem. Phys.*, 11, 12049-12064, 2011.
- 5 DeCarlo, P. F., Kimmel, J. R., Trimborn, A., Northway, M. J., Jayne, J. T., Aiken, A. C., Gonin, M., Fuhrer, K., Horvath, T., Docherty, K. S., Worsnop, D. R., and Jimenez, J. L.: Field-deployable, high-resolution, time-of-flight aerosol mass spectrometer, *Anal. Chem.*, 78, 8281-8289, 2006.
- Denkenberger, K. A., Moffet, R. C., Holecek, J. C., Rebotier, T. P., and Prather, K. A.: Real-time, single-particle measurements of oligomers in aged ambient aerosol particles, *Environ. Sci. Technol.*, 41,5439-5446, 2007.
- 10 Di Lorenzo, R. A., and Young, C. J.: Size separation method for absorption characterization in brown carbon: Application to an aged biomass burning sample, *Geophys. Res. Lett.*, 43, 458-465, 2016.
- 15 Dinar, E., Riziq, A. A., Spindler, C., Erlick, C., Kiss, G., and Rudich, Y.: The complex refractive index of atmospheric and model humic-like substances (HULIS) retrieved by a cavity ring down aerosol spectrometer (CRD-AS), *Faraday Discuss.*, 137, 279-295, 2008.
- Drewnick, F., Hings, S. S., Curtius, J., Eerdekens, G., and Williams, J.: Measurement of fine particulate and gas-phase species during the New Year's fireworks 2005 in Mainz, Germany, *Atmos. Environ.*, 40, 4316-4327, 2006.
- 20 Elias, V. O., Simoneit, B. R. T., Pereira, A. S., Cabral, J. A., and Cardoso, J. N.: Detection of high molecular weight organic tracers in vegetation smoke samples by high-temperature gas chromatography-mass spectrometry, *Environ. Sci. Technol.*, 33, 2369-2376, 1999.
- 25 Freutel, F., Drewnick, F., Schneider, J., Klimach, T., and Borrmann, S.: Quantitative single-particle analysis with the Aerodyne aerosol mass spectrometer: development of a new classification algorithm and its application to field data, *Atmos. Meas. Tech.*, 6, 3131-3145, 2013.
- 30 Friedman, B., Herich, H., Kammermann, L., Gross, D. S., Arneth, A., Holst, T., and Cziczo, D. J.: Subarctic atmospheric aerosol composition: 1. Ambient aerosol characterization, *J. Geophys. Res.-Atmos.*, 114, D13203, DOI: 10.1029/2009JD011772, 2009.
- Giorio, C., Tapparo, A., Dall'Osto, M., Harrison, R. M., Beddows, D. C. S., Di Marco, C., and Nemitz, E.: Comparison of three techniques for analysis of data from an Aerosol Time-of-Flight Mass Spectrometer, *Atmos. Environ.*, 61, 316-326, 2012.
- 35 Gross, D. S., Galli, M. E., Silva, P. J., and Prather, K. A.: Relative sensitivity factors for alkali metal and ammonium cations in single particle aerosol time-of-flight mass spectra, *Anal. Chem.*, 72, 416-422, 2000.

- Gross, D. S., Atlas, R., Rzeszutarski, J., Turetsky, E., Christensen, J., Benzaid, S., Olson, J., Smith, T., Steinberg, L., Sulman, J., Ritz, A., Anderson, B., Nelson, C., Musicant, D. R., Chen, L., Snyder, D. C., and Schauer, J. J.: Environmental chemistry through intelligent atmospheric data analysis, *Environ. Modell. Softw.*, 25, 760-769, 2010.
- 5 Gross, D. S., Galli, M. E., Kalberer, M., Prevot, A. S. H., Dommen, J., Alfarra, M. R., Duplissy, J., Gaeggeler, K., Gascho, A., Metzger, A., and Baltensperger, U.: Real-time measurement of oligomeric species in secondary organic aerosol with the aerosol time-of-flight mass spectrometer, *Anal. Chem.*, 78, 2130-2137, 2006.
- 10 Healy, R. M., Hellebust, S., Kourtchev, I., Allanic, A., O'Connor, I. P., Bell, J. M., Healy, D. A., Sodeau, J. R., and Wenger, J. C.: Source apportionment of PM_{2.5} in Cork Harbour, Ireland using a combination of single particle mass spectrometry and quantitative semi-continuous measurements, *Atmos. Chem. Phys.*, 10, 9593-9613, 2010.
- 15 Healy, R. M., Sciare, J., Poulain, L., Crippa, M., Wiedensohler, A., Prevot, A. S. H., Baltensperger, U., Sarda-Estevé, R., McGuire, M. L., Jeong, C. -H., McGillicuddy, E., O'Connor, I. P., Sodeau, J. R., Evans, G. J., and Wenger, J. C.: Quantitative determination of carbonaceous particle mixing state in Paris using single-particle mass spectrometer and aerosol mass spectrometer measurements, *Atmos. Chem. Phys.*, 13, 9479-9496, 2013.
- 20 Healy, R. M., Wang, J. M., Jeong, C. -H., Lee, A. K. Y., Willis, M. D., Jaroudi, E., Zimmerman, N., Hilker, N., Murphy, M., Eckhardt, S., Stohl, A., Abbatt, J. P. D., Wenger, J. C., and Evans, G. J.: Light-absorbing properties of ambient black carbon and brown carbon from fossil fuel and biomass burning sources, *J. Geophys. Res.-Atmos.*, 120, 6619-6633, 2015.
- 25 Jimenez, J. L., Jayne, J. T., Shi, Q., Kolb, C. E., Worsnop, D. R., Yourshaw, I., Seinfeld, J. H., Flagan, R. C., Zhang, X. F., Smith, K. A., Morris, J. W., and Davidovits, P.: Ambient aerosol sampling using the Aerodyne Aerosol Mass Spectrometer, *J. Geophys. Res.-Atmos.*, 108, 8425, DOI: 10.1029/2001JD001213, 2003.
- 30 Kirchstetter, T. W., Novakov, T., and Hobbs, P. V.: Evidence that the spectral dependence of light absorption by aerosols is affected by organic carbon, *J. Geophys. Res.-Atmos.*, 109, D21208, DOI: 10.1029/2004JD004999, 2004.
- Lack, D. A., Bahreni, R., Langridge, J. M., Gilman, J. B., and Middlebrook, A. M.: Brown carbon absorption linked to organic mass tracers in biomass burning particles, *Atmos. Chem. Phys.*, 13, 2415-2422, 2013.
- 35 Lack, D. A., Langridge, J. M., Bahreini, R., Cappa, C. D., Middlebrook, A. M., and Schwarz, J. P.: Brown carbon and internal mixing in biomass burning particles, *Proc. Natl. Acad. Sci. U. S. A.*, 109, 14802-14807, 2012.

- Lee, A. K. Y., Willis, M. D., Healy, R. M., Onasch, T. B., and Abbatt, J. P. D.: Mixing state of carbonaceous aerosol in an urban environment: single particle characterization using the soot particle aerosol mass spectrometer (SP-AMS), *Atmos. Chem. Phys.*, 15, 1823-1841, 2015.
- 5 Li, C., Ma, Z., Chen, J., Wang, X., Ye, X., Wang, L., Yang, X., Kan, H., Donaldson, D. J., and Mellouki, A.: Evolution of biomass burning smoke particles in the dark, *Atmos. Environ.*, 120, 244-252, 2015.
- Li, J., Posfai, M., Hobbs, P. V., and Buseck, P. R.: Individual aerosol particles from biomass burning in southern Africa: 2, Compositions and aging of inorganic particles, *J. Geophys. Res.-Atmos.*, 108, 8484, DOI: 10.1029/2002JD002310, 2003.
- 10 Liu, S., Russell, L. M., Sueper, D. T., and Onasch, T. B.: Organic particle types by single-particle measurements using a time-of-flight aerosol mass spectrometer coupled with a light scattering module, *Atmos. Meas. Tech.*, 6, 187-197, 2013.
- Liu, S., Aiken, A. C., Gorkowski, K., Dubey, M. K., Cappa, C. D., Williams, L. R., Herndon, S. C., Massoli, P., Fortner, E. C., Chhabra, P. S., Brooks, W. A., Onasch, T. B., Jayne, J. T., Worsnop, D. R., China, S., Sharma, N., Mazzoleni, C., Xu, L., Ng, N. L., Liu, D., Allan, J. D., Lee, J. D., Fleming, Z. L., Mohr, C., Zotter, P., Szidat, S., and Prevot, A. S. H.: Enhanced light absorption by mixed source black and brown carbon particles in UK winter., *Nat. Commun.*, 6, 8435-8435, 2015.
- 15 Lu, Z., Streets, D. G., Winijkul, E., Yan, F., Chen, Y., Bond, T. C., Feng, Y., Dubey, M. K., Liu, S., Pinto, J. P., and Carmichael, G. R.: Light Absorption Properties and Radiative Effects of Primary Organic Aerosol Emissions, *Environ. Sci. Technol.*, 49, 4868-4877, 2015.
- 20 Massoli, P., Onasch, T. B., Cappa, C. D., Nuumaan, I., Hakala, J., Hayden, K., Li, S., Sueper, D. T., Bates, T. S., Quinn, P. K., Jayne, J. T., and Worsnop, D. R.: Characterization of black carbon-containing particles from soot particle aerosol mass spectrometer measurements on the R/V Atlantis during CalNex 2010, *J. Geophys. Res.-Atmos.*, 120, 2575-2593, 2015.
- 25 May, A. A., Levin, E. J. T., Hennigan, C. J., Riipinen, I., Lee, T., Collett, J. L., Jr., Jimenez, J. L., Kreidenweis, S. M., and Robinson, A. L.: Gas-particle partitioning of primary organic aerosol emissions: 3. Biomass burning, *J. Geophys. Res.-Atmos.*, 118, 11327-11338, 2013.
- 30 McMeeking, G. R., Fortner, E., Onasch, T. B., Taylor, J. W., Flynn, M., Coe, H., and Kreidenweis, S. M.: Impacts of nonrefractory material on light absorption by aerosols emitted from biomass burning, *J. Geophys. Res.-Atmos.*, 119, 12272-12286, 2014.
- 35 Mohr, C., Lopez-Hilfiker, F. D., Zotter, P., Prevot, A. S. H., Xu, L., Ng, N. L., Herndon, S. C., Williams, L. R., Franklin, J. P., Zahniser, M. S., Worsnop, D. R., Knighton, W. B.,

- Aiken, A. C., Gorkowski, K. J., Dubey, M. K., Allan, J. D., and Thornton, J. A.: Contribution of Nitrated Phenols to Wood Burning Brown Carbon Light Absorption in Detling, United Kingdom during Winter Time, *Environ. Sci. Technol.*, 47, 6316-6324, 2013.
- 5 Ng, N. L., Herndon, S. C., Trimborn, A., Canagaratna, M. R., Croteau, P. L., Onasch, T. B., Sueper, D., Worsnop, D. R., Zhang, Q., Sun, Y. L., and Jayne, J. T.: An Aerosol Chemical Speciation Monitor (ACSM) for Routine Monitoring of the Composition and Mass Concentrations of Ambient Aerosol, *Aerosol Sci. Technol.*, 45, 780-794, 2011.
- 10 Onasch, T. B., Trimborn, A., Fortner, E. C., Jayne, J. T., Kok, G. L., Williams, L. R., Davidovits, P. and Worsnop, D. R.: Soot Particle Aerosol Mass Spectrometer: Development, Validation, and Initial Application, *Aerosol Sci. Technol.*, 46, 804-817, 2012.
- 15 Pagels, J., Dutcher, D. D., Stolzenburg, M. R., McMurry, P. H., Gaelli, M. E., and Gross, D. S.: Fine-particle emissions from solid biofuel combustion studied with single-particle mass spectrometry: Identification of markers for organics, soot, and ash components, *J. Geophys. Res.-Atmos.*, 118, 859-870, 2013.
- Rebotier, T. P., and Prather, K. A.: Aerosol time-of-flight mass spectrometry data analysis: A benchmark of clustering algorithms, *Anal. Chim. Acta*, 585, 38-54, 2007.
- 20 Reilly, P. T. A., Lazar, A. C., Gieray, R. A., Whitten, W. B., and Ramsey, J. M.: The elucidation of charge-transfer-induced matrix effects in environmental aerosols via real-time aerosol mass spectral analysis of individual airborne particles, *Aerosol Sci. Technol.*, 33, 135-152, 2000.
- 25 Saleh, R., Robinson, E. S., Tkacik, D. S., Ahern, A. T., Liu, S., Aiken, A. C., Sullivan, R. C., Presto, A. A., Dubey, M. K., Yokelson, R. J., Donahue, N. M., and Robinson, A. L.: Brownness of organics in aerosols from biomass burning linked to their black carbon content, *Nat. Geosci.*, 7, 647-650, 2014.
- Smith, J. S., Laskin, A., and Laskin, J.: Molecular Characterization of Biomass Burning Aerosols Using High-Resolution Mass Spectrometry, *Anal. Chem.*, 81, 1512-1521, 2009.
- 30 Sun, H., Biedermann, L., and Bond, T. C.: Color of brown carbon: A model for ultraviolet and visible light absorption by organic carbon aerosol, *Geophys. Res. Lett.*, 34, L17813, DOI: 10.1029/2007GL029797, 2007.
- Ulbrich, I. M., Canagaratna, M. R., Zhang, Q., Worsnop, D. R., and Jimenez, J. L.: Interpretation of organic components from Positive Matrix Factorization of aerosol mass spectrometric data, *Atmos. Chem. Phys.*, 9, 2891-2918, 2009.
- 35 Washenfelder, R. A., Attwood, A. R., Brock, C. A., Guo, H., Xu, L., Weber, R. J., Ng, N. L., Allen, H. M., Ayres, B. R., Baumann, K., Cohen, R. C., Draper, D. C., Duffey, K. C.,

Edgerton, E., Fry, J. L., Hu, W. W., Jimenez, J. L., Palm, B. B., Romer, P., Stone, E. A., Wooldridge, P. J., and Brown, S. S.: Biomass burning dominates brown carbon absorption in the rural southeastern United States, *Geophys. Res. Lett.*, 42, 653-664, 2015.

5 Willis, M. D., Lee, A. K. Y., Onasch, T. B., Fortner, E. C., Williams, L. R., Lambe, A. T., Worsnop, D. R., and Abbatt, J. P. D.: Collection efficiency of the soot-particle aerosol mass spectrometer (SP-AMS) for internally mixed particulate black carbon, *Atmos. Meas. Tech.*, 7, 4507-4516, 2014.

10 Willis, M. D., Healy, R. M., Riemer, N., West, M., Wang, J. M., Jeong, C.-H., Wenger, J., Evans, G. J., Abbatt, J. P. D., and Lee, A. K. Y.: Quantification of black carbon mixing state from traffic: implications for aerosol optical properties, , *Atmos. Chem. Phys.*, accepted.

Zhang, Q., Jimenez, J. L., Canagaratna, M. R., Ulbrich, I. M., Ng, N. L., Worsnop, D. R., and Sun, Y.: Understanding atmospheric organic aerosols via factor analysis of aerosol mass spectrometry: a review, *Analyt. Bioanalyt. Chem.*, 401, 3045-3067, 2011.

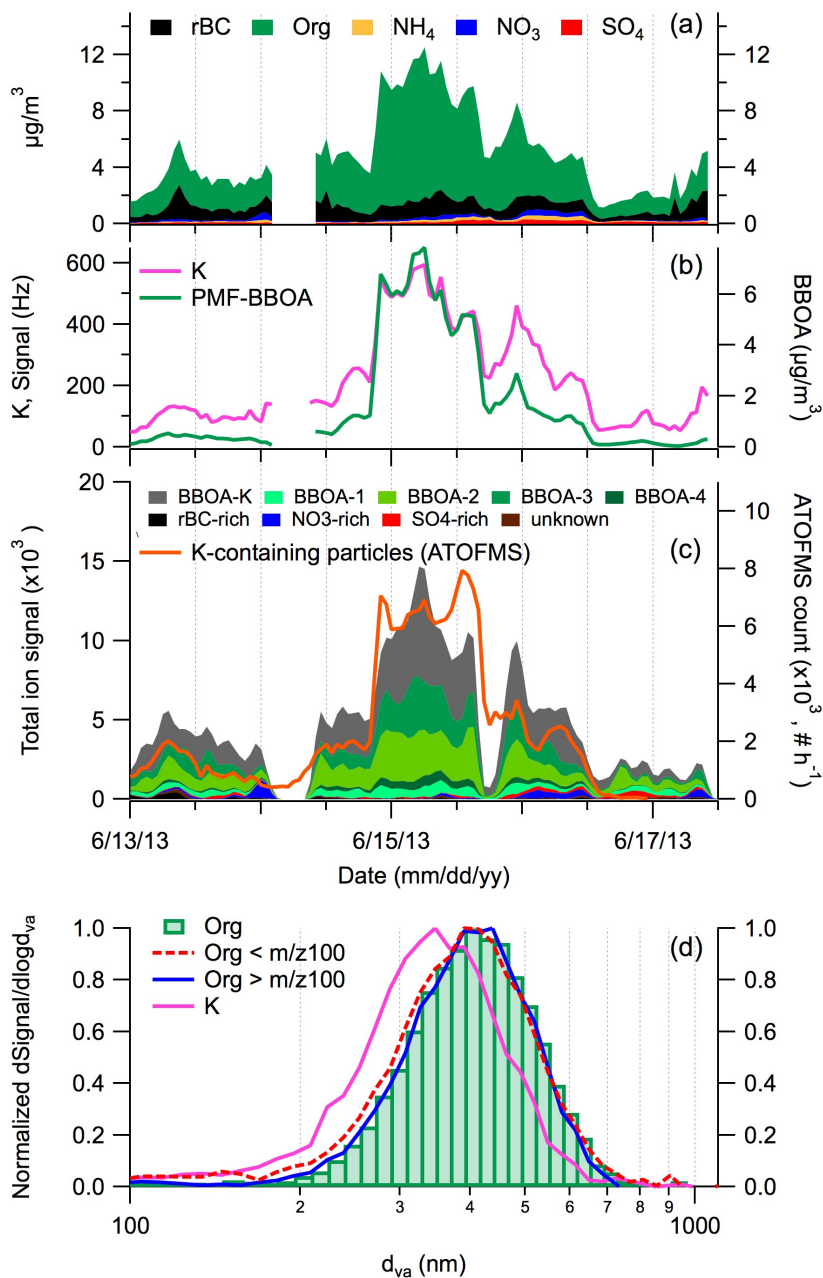
Tables:

Table 1: Summary of LS-SP-AMS single particle cluster analysis

Particle class	Number of particles	<u>Number</u> Percentage (%)	Average mass fraction of rBC (m_{rBC})	Average ion fraction of m/z 39	Average ion fraction of m/z 100-650
SO ₄ -rich	200	3	0.04	0.01	0.27
NO ₃ -rich	344	5	0.05	0.02	0.05
rBC-rich	78	1	0.86	0.04	0.07
BBOA-K	1020	15	0.14	0.33	0.02
BBOA-1	872	12	0.03	0.02	0.04
BBOA-2	1645	24	0.03	0.02	0.23
BBOA-3	1956	28	0.03	0.01	0.51
BBOA-4	745	11	0.05	0.01	0.56
Unknown	139	2	--	--	--

5

Figures:



5

Figure 1: Time series of (a) rBC and NR-PM_{rBC} (Org, SO₄, NO₃, NH₄), (b) K⁺ and PMF-BBOA factor measured by the LS-SP-AMS, (c) number of K-containing particles measured by the ATOFMS (right axis) and total ion signal of all particle classes identified by cluster analysis of LS-SP-AMS data (left axis). (d) Size distributions of K⁺ and Org associated with rBC measured by the LS-SP-AMS.

10

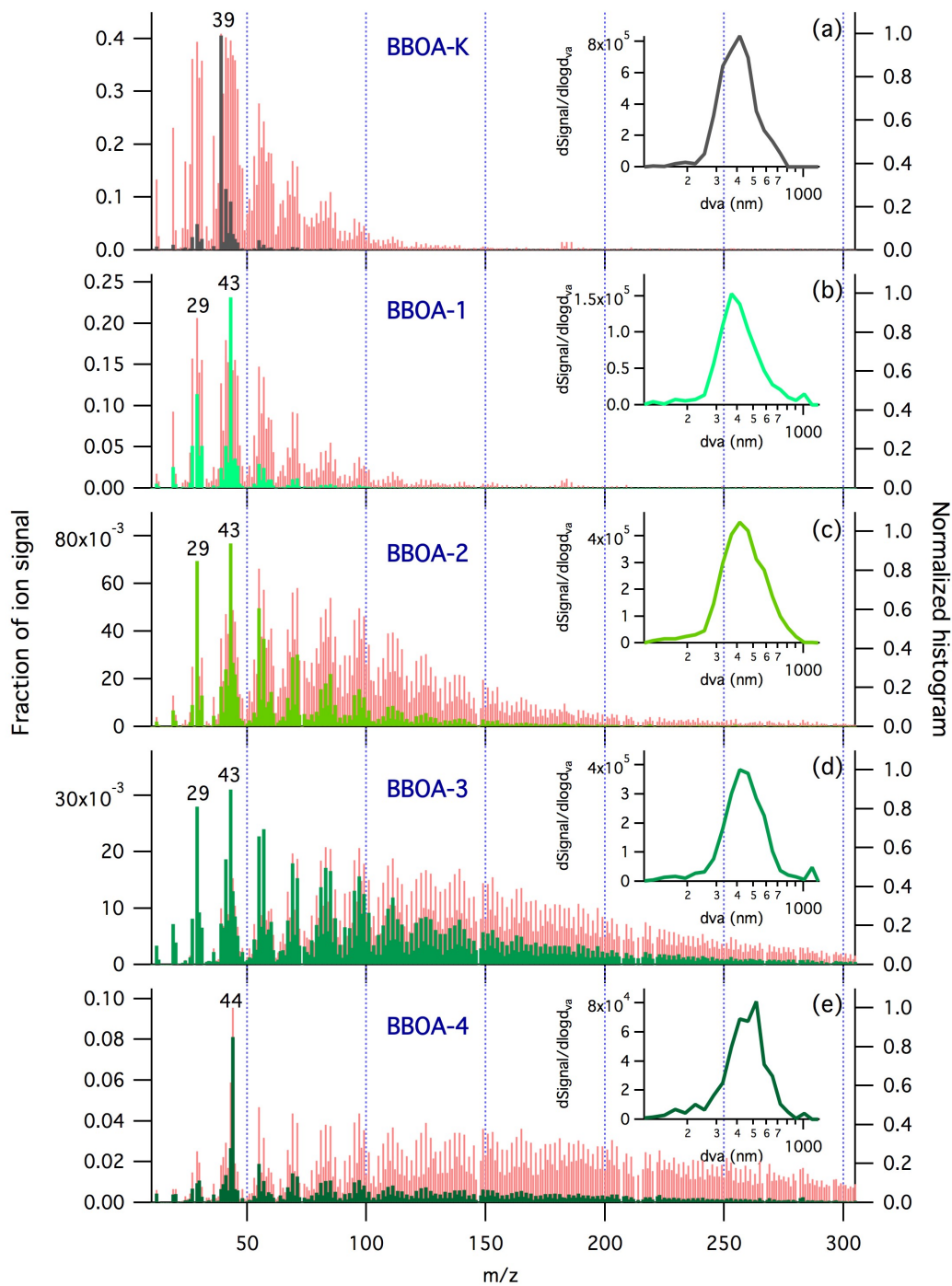


Figure 2: Average mass spectra (left axis), normalized mass spectral histograms (showing the relative frequency of a signal at each m/z , right axis), and size distributions (insets) of the five BBOA-related particle classes identified by cluster analysis of LS-SP-AMS data.

5

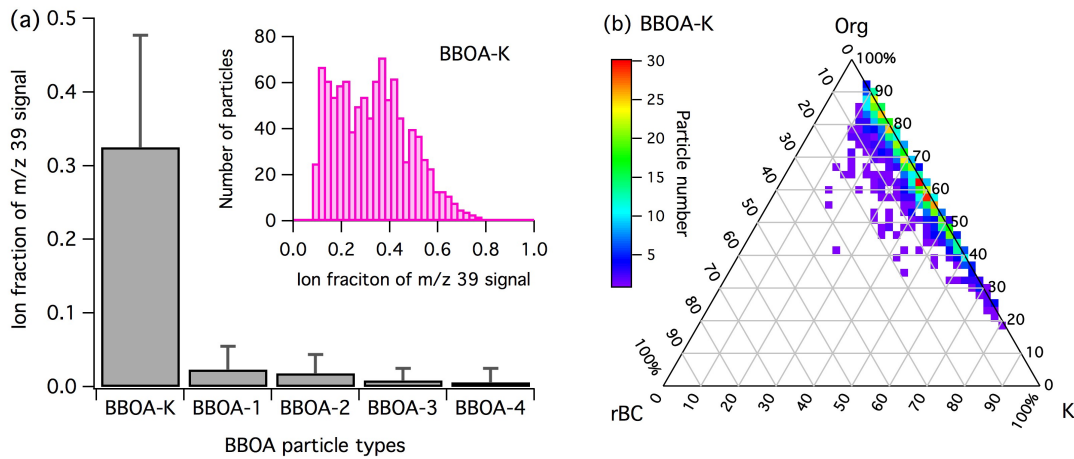
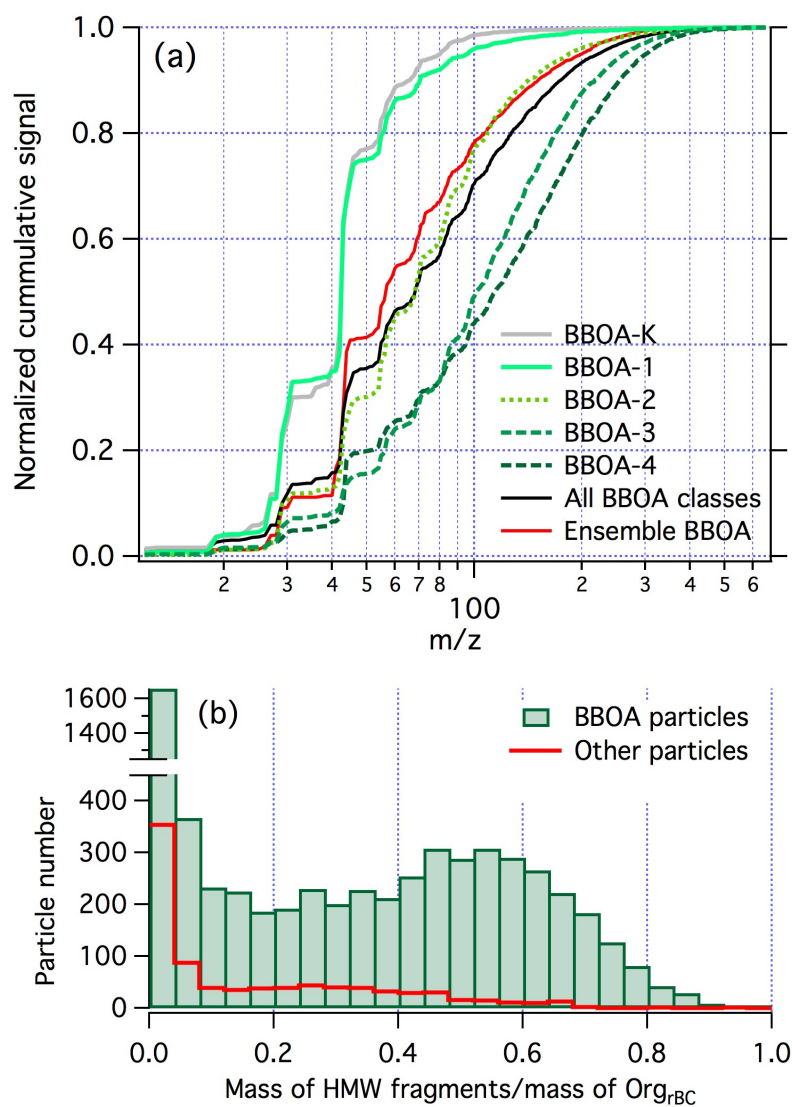


Figure 3: (a) Ion fraction of m/z 39 signal (a tracer of K^+) for each BBOA-related particle class. Histogram of the ion fraction of m/z 39 for the BBOA-K particle class (inset). (b) Ternary plot of rBC , Org_{BC} and K ion signal for the five BBOA particle types



5 Figure 4: (a) Normalized cumulative histogram of mass-to-charge ratios for each BBOA-related particle class. (b) Histograms of the mass fraction of HMW organic fragments ($> m/z 100$) in different particle types.

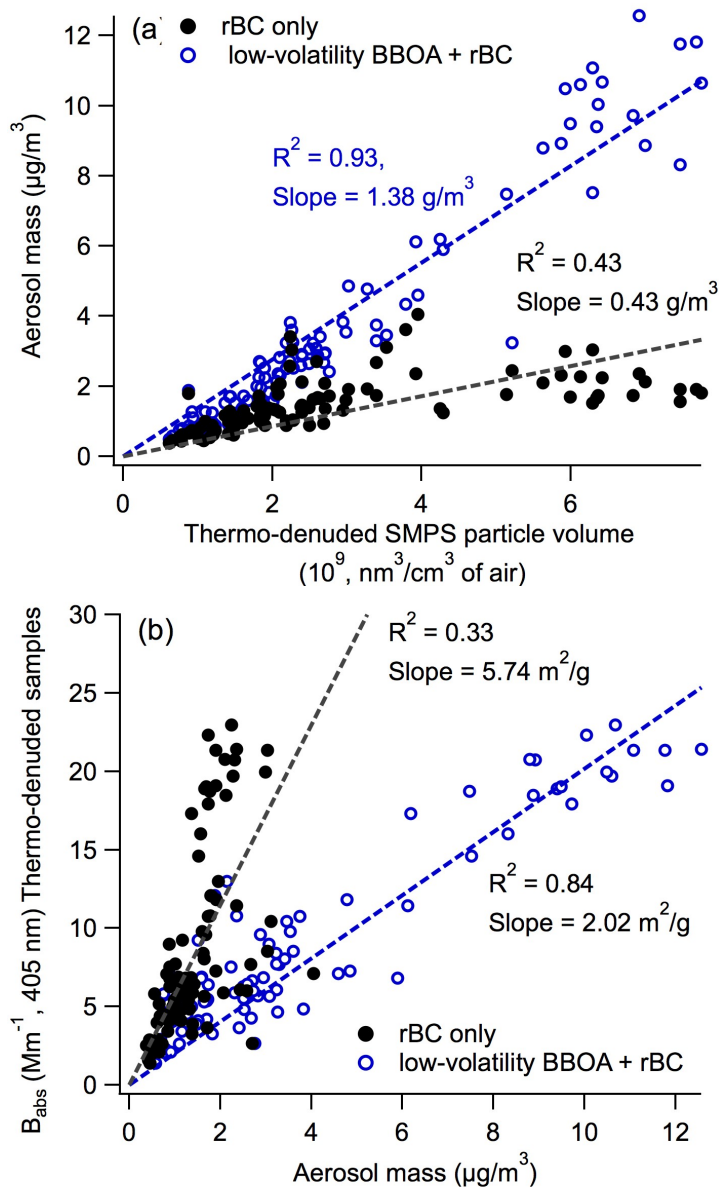


Figure 5: (a) Correlation of the estimated total of non- and low-volatility aerosol mass (rBC + 60% of the total BBOA) and thermo-denuded particle volume measured by the SMPS. (b) Correlations of aerosol absorption at 405 nm of thermo-denuded particles (B_{abs}) and the estimated total of non- and low-volatility aerosol mass