

1 **Particle water and pH in the Eastern Mediterranean: Sources variability and**
2 **implications for nutrients availability.**

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16

17 **Abstract**

18 Particle water (LWC) and aerosol pH are important parameters of the aerosol phase, affecting
19 heterogeneous chemistry and bioavailability of nutrients that profoundly impact cloud formation,
20 atmospheric composition and atmospheric fluxes of nutrients to ecosystems. Few measurements
21 of in-situ LWC and pH however exist in the published literature. Using concurrent measurements
22 of aerosol chemical composition, cloud condensation nuclei activity and tandem light scattering
23 coefficients, the particle water mass concentrations associated with the aerosol inorganic (W_{inorg})
24 and organic (W_{org}) components are determined for measurements conducted at the Finokalia
25 atmospheric observation station in the eastern Mediterranean between August and November
26 2012. These data are interpreted using the ISORROPIA-II thermodynamic model to predict pH of
27 aerosols originating from the various sources that influence air quality in the region. On average,
28 closure between predicted aerosol water and that determined by comparison of ambient with dry
29 light scattering coefficients was achieved to within 8% (slope=0.92, $R^2=0.8$, $n=5201$ points).
30 Based on the scattering measurements a parameterization is also derived, capable of reproducing

31 the hygroscopic growth factor ($f(RH)$) within 15% of the measured values. The highest aerosol
32 water concentrations are observed during nighttime, when relative humidity is highest and the
33 collapse of the boundary layer increases the aerosol concentration. A significant diurnal variability
34 is found for W_{org} with morning and afternoon average mass concentrations being 10-15 times lower
35 than nighttime concentrations, thus rendering W_{inorg} the main form of particle water during
36 daytime. The average value of total aerosol water was $2.19 \pm 1.75 \mu\text{g m}^{-3}$, contributing on average
37 up to 33% of the total submicron mass concentration. Average aerosol water associated with
38 organics, W_{org} , was equal to $0.56 \pm 0.37 \mu\text{g m}^{-3}$, thus organics contributed about 27.5% to the total
39 aerosol water, mostly during early morning, late evening and nighttime hours.
40 The aerosol was found to be highly acidic with calculated aerosol pH varying from 0.5 to 2.8
41 throughout the study period. Biomass burning aerosol presented the highest values of pH in the
42 submicron fraction and the lowest values in total water mass concentration. The low pH values
43 observed in the submicron mode and independently of air masses origin could increase nutrient
44 availability and especially P solubility, which is the nutrient limiting sea water productivity of the
45 eastern Mediterranean.

46

47 **1. Introduction**

48

49 Atmospheric particles have the ability to absorb significant amounts of water, which profoundly
50 affects their physical and chemical properties (Khlystov et al., 2005), and impacts on atmospheric
51 processes and health. Ambient concentrations of aerosol liquid water are controlled by the aerosol
52 chemical composition, relative humidity (RH) and temperature, as it is largely in chemical
53 equilibrium with the surrounding water vapor. Liquid water is ubiquitous and exceeds the total
54 aerosol dry mass by 2 to 3 times on a global scale (Liao and Seinfeld, 2005). Therefore, the aerosol
55 liquid water content (LWC) increases the particle size, affecting the particle lifetime and scattering
56 efficiency. LWC and its strong dependence on relative humidity (RH) are the most important
57 contributors to aerosol direct radiative cooling by aerosols (Piliinis et al., 1995). Numerous
58 modeling studies suggest that reactions in aerosol liquid water are an important pathway of
59 secondary organic aerosol (SOA) formation (Carlton and Turpin, 2013; Myriokefalitakis et al.,
60 2011), thus playing an important role in the overall aerosol chemical composition. Aerosol water
61 also has profound impact on the aerosol phase state, being able to transform semi-solid and viscous

62 particles into homogeneous phase that are in equilibrium with their environment (Pöschl and
63 Shiraiwa, 2015). This affects the timescale of heterogeneous reactions and ice nucleation.
64 Nevertheless, despite the abundance and importance of LWC, it is not routinely measured, actual
65 mass concentrations are uncertain, especially in the presence of organic compounds and model
66 predictions of the property are often not evaluated (Nguyen et al., 2014).

67 Apart from the LWC, the pH of aqueous aerosols is another critically important aerosol property
68 that drives many processes related to the aerosol chemical composition and gas-aerosol
69 partitioning (Guo et al., 2015; Surratt et al., 2007; 2010; Meskhidze et al., 2003; Eddingsaas et al.,
70 2010; Myriokefalitakis et al., 2015). Direct measurements of aerosol pH “in situ” are scarce (e.g.,
71 Keene et al., 2002; 2004) and require careful considerations owing to the non-conserved nature of
72 the hydronium ion and partial dissociation of inorganic and organic electrolytes in the aerosol.
73 These challenges have led to the suggestion that indirect alternatives – such as measuring the semi-
74 volatile partitioning of key species sensitive to pH, combined with comprehensive models may
75 provide a reasonably accurate estimate of pH that can be carried out with routine measurements
76 (Hennigan et al., 2015). The most frequently used proxy is the “ion balance”, where the charge
77 balance of measurable anions and cations is calculated, with the exception of the hydronium and
78 hydroxyl ions; a surplus of cations implies an alkaline aerosol and vice versa. Often it is implied
79 that a larger value of the ion balance implies a stronger acidity/alkalinity. As shown by Hennigan
80 et al. (2015) and Guo et al. (2015), the ion balance (and other similar proxies discussed in Hennigan
81 et al., 2015) fail in general to represent the true aerosol pH; only meticulous measurement of semi-
82 volatile species (such as ammonia/ammonium) and other aerosol chemical constituents, combined
83 with appropriate thermodynamic calculations (e.g. with ISORROPIA-II; Fountoukis and Nenes,
84 2007) are able to realistically provide particle pH and LWC (Hennigan et al., 2015). For an
85 accurate and unbiased pH calculation both gas and aerosol phase concentrations are needed,
86 however, when some uncertainty can be tolerated or the level of pH bias is known, aerosol
87 measurements alone can still be quite informative for determining the pH, as demonstrated by Guo
88 et al. (2015).

89 Directly linked to aerosol pH and LWC is the bioavailability of nutrients contained within dust,
90 involving pH-dependent catalyzed redox-reaction pathways; upon deposition, increased
91 availability of these nutrients may promote primary productivity in continental and marine
92 ecosystems (e.g., Meskhidze et al., 2003; Nenes et al., 2011; Mahowald et al., 2008; 2009;

93 Krishnamurthy et al., 2010). Acids (such as sulfuric and nitric) generated in the atmosphere from
94 a variety of anthropogenic and biogenic sources, when mixed with mineral aerosols in sufficient
95 amounts could lower the aerosol pH to values that increase the solubilities of Fe and P-containing
96 minerals by several orders of magnitude (Stumm and Morgan, 1996; Shi et al., 2012). Nenes et al.
97 (2011) have demonstrated that acidification can release considerable amounts of soluble
98 phosphorus from soil-laden minerals (e.g., 81-96% of the total P found in Saharan dust and soil).
99 Apart from P, the transport and deposition of mineral dust is believed to be a major, if not the
100 dominant source of Fe to the remote ocean (Jickells et al., 2005). Considering the mixing of SO₂
101 with advected dust plumes and the subsequent acidification through heterogeneous oxidation of the
102 SO₂ on deliquescent dust particles within the plume, Meskhidze et al. (2003) concluded that for
103 aerosol with pH < 2, 1-2% of the contained Fe would be mobilized within 3-5 days. Meskhidze et
104 al., (2005) also demonstrated that sufficient acidification of Asian dust plumes could drive, upon
105 deposition, a phytoplankton bloom in High Nitrate Low Chlorophyll regions of the North Atlantic.
106 Solomon et al. (2009) also showed that simulated enhancements in particulate soluble iron driven
107 by the chemical dissolution mechanism, can range from 0.5 to 6%, which is consistent with
108 observations over the North Pacific Ocean.

109 The Eastern Mediterranean, being at the nexus of three continents (Europe, Asia and Africa),
110 receives air masses influenced by a spectrum of human (traffic, biomass burning and industry) and
111 natural (dust and marine) sources. It is therefore an ideal location to study atmospheric
112 acidification of aerosols; very few studies however to date have accomplished that.

113 This study uses aerosol chemical composition measurements in conjunction with cloud
114 condensation nuclei (CCN) concentration and light scattering coefficient (σ_{sp}) measurements to
115 model the water mass concentrations (LWC) of aerosols from various sources in the Eastern
116 Mediterranean. These data are then used in combination with the ISORROPIA-II thermodynamic
117 equilibrium model (Fountoukis and Nenes, 2007) to predict the aerosol pH for air masses that
118 influenced air quality in the Eastern Mediterranean during the late summer and fall months of
119 2012.

120

121 **2. Instrumentation and Methods**

122 **2.1 Measurement site**

123 Aerosol measurements were conducted at the Finokalia atmospheric observation station in Crete,
124 Greece (35°20'N, 25°40'E, 250m a.s.l.) between August and November 2012. The site
125 (<http://finokalia.chemistry.uoc.gr/>) is a European supersite for atmospheric aerosol research and is
126 part of the ACTRIS Network (Aerosols, Clouds, and Trace gases Research Infrastructure)
127 (<http://www.actris.eu/>). The station is located at the northeastern part of the island of Crete, facing
128 the Mediterranean Sea covering the whole northern sector. Being away from direct urban
129 influence, the station is representative for background measurements in the Eastern Mediterranean
130 and sampled air masses arriving at the site have most commonly a marine source region or
131 originate from continental Europe and Greek mainland. Moreover, dust events from Northern
132 Africa (Sahara) occur often during spring and autumn. More details for the Finokalia station and
133 the prevailing climatology in the area are given by Mihalopoulos et al. (1997).

134

135 **2.2 Instrumentation**

136 PM₁ (Particulate Matter of diameter smaller than 1 μm) aerosol light scattering coefficients (σ_{sp})
137 were measured online with the use of two different nephelometers to infer the LWC. Particle dry
138 scattering was measured with a 3-wavelength Aurora 1000 Integrating Nephelometer, located in
139 an air-conditioned sampling trailer and operated with a diffusional silica dryer upstream, which
140 maintained the RH below 35%. Based on the ISORROPIA model run for metastable ammonium sulfate
141 aerosol at an RH of 30%, the maximum water which can be contained under these conditions is 1.21 μg m⁻³,
142 which is less than 12% of the total submicron aerosol mass. The nephelometer for the particle wet
143 scattering was a Radiance Research M903, located outside the trailer in order to provide a
144 scattering measurement at ambient T and RH. This second nephelometer was equipped with a
145 VAISALA Inc. Humitter 50U Integrated Humidity and Temperature probe in order to record the
146 ambient values of T and RH with an accuracy at 20 °C specified to be better than ±5% RH. Both
147 nephelometers take part in yearly intercomparisons within the ACTRIS network.

148 The real-time, quantitative measurements of the non-refractory components of the submicron
149 aerosol were provided by an Aerosol Research Inc. Aerosol Chemical Speciation Monitor (ACSM;
150 Ng et al., 2011). Ambient air was drawn into the ACSM by a multiple PM₁₀ (Particulate Matter of
151 diameter smaller than 10 μm) aerosol inlet operating at Finokalia, with a temporal resolution of 30

152 minutes. At the inlet of the ACSM and via a critical aperture mounted at the entrance of an
153 aerodynamic lens, the submicron fraction of the aerosol is sampled. The aerodynamic lens of the
154 ACSM allows for the detection of particles up to 700 nm diameter. The focused particle beam is
155 then transmitted through two vacuum chambers and into a detection chamber where the particles
156 impact on a hot plate and flash vaporized. Finally they are detected and characterized with the use
157 of an electron impact quadrupole mass spectrometer. Constituents quantified by the ACSM include
158 organics, sulfate, ammonium, nitrate and chloride. Detection limits for all constituents for 30 min
159 of averaging time are provided by Ng et al. (2011) and are for ammonium, organics, sulfate, nitrate,
160 and chloride are $0.284 \mu\text{g m}^{-3}$, $0.148 \mu\text{g m}^{-3}$, $0.024 \mu\text{g m}^{-3}$, $0.012 \mu\text{g m}^{-3}$, and $0.011 \mu\text{g m}^{-3}$,
161 respectively. Mass concentrations are calculated with the recommended collection efficiency of
162 0.5 for all constituents (Ng et al. 2011) and the main aerosol constituents i.e. organics, sulfate and
163 ammonium are verified by comparison with other concurrent measurements (daily PM_{10} filters;
164 Bougiatioti et al. (2014) supplementary material). PM_{10} filters (Pallflex Tissuquartz, 47 mm
165 diameter) were collected using a built-in system comprised of the four upper stages (stage 8-11)
166 of a low-pressure Berner impactor (PLBI; Berner and Lürzer, 1980). Furthermore, daily PM_{10}
167 filters are also collected on site using a sequential sampler (Leckel, SEQ47/50). A detailed study
168 of the comparison between ACSM and PM_{10} filter values for a large subset of the present data
169 (16/08-30/09/2012) is provided by Bougiatioti et al. (2014).

170 PM_{10} and PM_{10} filters were analyzed by ion chromatography (IC) for anions (Cl^- , Br^- , NO_3^-
171 (nitrate), SO_4^{2-} (sulfate), $\text{C}_2\text{O}_4^{2-}$ (oxalate)) and cations (K^+ , Na^+ , NH_4^+ , Mg^{2+} , Ca^{2+}) using the
172 procedure described by Bardouki et al. (2003). From the calcium, the crustal component can be
173 estimated by subtracting the contribution of sea-salt $\text{Ca}^{2+} = [\text{Na}] \cdot 0.038$ and dust concentration is
174 estimated using the approach of Sciare et al. (2005). Based on the study by Koulouri et al. (2008)
175 for a two-year period at the same sampling site, it has been demonstrated that for fine particles (in
176 that case $D_a < 1.3 \mu\text{m}$) the contribution of the marine source to the aerosol mass was 10%. Similarly,
177 dust contribution in the fine fraction can vary between 6 and 10% for summer and winter,
178 respectively.

179 From 25/6 to 7/8/2014, gas phase ammonia measurements were also conducted at Finokalia, using
180 a wet annular denuder (WAD; Wyers et al., 1993; Spindler et al., 2003) coupled with an ion
181 chromatography system quantifying cations with an hourly resolution in order to check the validity
182 of our calculations in the absence of gaseous phase NH_3 (see section 2.4). Concurrent black carbon

183 (BC) measurements were performed on site using a seven-wavelength aethalometer (Magee
184 Scientific, AE31) with a time resolution of 5 min. Based on previous studies at Finokalia, BC is
185 found mainly on the fine aerosol fraction (Koulouri et al., 2008; Bougiatioti et al., 2014) hence
186 these values are used in addition to the ACSM concentrations to calculate the dry aerosol mass of
187 the PM₁ fraction.

188 Size-selected cloud condensation nuclei (CCN) measurements were obtained using a Droplet
189 Measurement Technologies, Continuous Flow Streamwise Thermal Gradient CCN counter
190 (CFSTGC). Particles of 60, 80, 100 and 120 nm were first size-selected by a differential mobility
191 analyzer (DMA), split into two and one of these flows was introduced in the CCN counter. The
192 total number of condensation nuclei (CN) was measured by a condensation particle counter (CPC;
193 TSI 1772) situated downstream the first DMA. In the CCN counter, the activated droplets are sized
194 and counted by an optical particle counter (OPC) after exiting the growth chamber. The instrument
195 was operated in scanning flow CCN analysis mode (SFCA; Moore and Nenes, 2009), where the
196 flow rate in the growth chamber changes over time, while a constant temperature difference is
197 maintained. In that way the supersaturation changes continuously, providing activation spectra
198 with a high temporal resolution. The flow rate was increased linearly between a minimum and a
199 maximum flow rate and sigmoidal activation curves of CCN vs flow rate were recorded, with the
200 inflection point of the sigmoid representing a critical activation flow, Q_{50} , that corresponds to a
201 critical supersaturation, S^* , above which particles act as CCN. From the critical supersaturation
202 and knowledge of the particle dry diameter, with the application of Köhler theory, the
203 hygroscopicity parameter kappa (Petters and Kreidenweis, 2007) is obtained, which is then used
204 to determine the LWC associated with organic components of the aerosol (Guo et al., 2015).

205 Finally, analysis of 5-day back trajectories of air masses arriving at Finokalia at 1000 and 3000 m
206 above ground level every 6 h was conducted with the help of the HYSPLIT model
207 (www.arl.noaa.gov/hysplit.php; Stein et al., 2015).

208

209 **2.3 Determination of LWC from nephelometers**

210 The particle water was inferred using the approach of Guo et al. (2015), where the ratio of the wet
211 and dry PM₁ scattering coefficients (σ_{sp}) measured by the two nephelometers is used. For this, the
212 aerosol hygroscopic growth factor ($f(RH)$) is calculated according to a well-established method
213 based on the two scattering coefficients: $f(RH) = \sigma_{sp(wet)} / \sigma_{sp(dry)}$ (e.g. Sheridan, 2002; Magi, 2003;

214 Kim et al., 2006). These scattering coefficients in ambient and dry conditions are proportional to
 215 the diameter of average surface $\overline{D_p}$ and the average scattering coefficients $\overline{Q_s}$, so that :

$$216 \quad \overline{D_{p,ambient}} = \overline{D_{p,dry}} \sqrt{f(RH) \overline{Q_{s,dry}} / \overline{Q_{s,ambient}}} \quad (1)$$

217 Assuming that the two average scattering efficiencies are almost equal, LWC is then equal to the
 218 difference between ambient and dry particle volume, and by substitution of the dry diameter of
 219 average surface:

$$220 \quad \overline{D_{p,dry}}^3 = \frac{m_p}{\left(\frac{\pi}{6}\right) \rho_p N_t} \quad (2)$$

221 where N_t is the total number concentration, ρ_p is the density of dry aerosol and m_p is the dry mass
 222 concentration, we arrive at the simplified expression of:

$$223 \quad LWC = [f(RH)^{1.5} - 1] \frac{m}{\rho_p} \quad (3)$$

224 The simplification assumption that the dry and ambient scattering efficiencies are almost equal
 225 introduces an error in the derived LWC, which based on the detailed analysis of Guo et al. (2015)
 226 is of the order of 10% at RH=76.4%, but can reach up to 21% at RH=90%.

227 For this set of measurements the PM₁ dry mass concentration is calculated from the sum of the
 228 ACSM constituents combined with the BC measurements. During the sampling period, the
 229 comparison between the sum of ACSM and BC masses and the submicron mass derived from a scanning
 230 mobility particle sizer (SMPS; TROPOS-type) is very good ($y=0.96x$, $R^2=0.67$ for a period of 2-years;
 231 Figure S3). Therefore the bias introduced by ignoring other refractory constituents than BC is minimal,
 232 which is in agreement with size-segregated aerosol chemical composition measurements already conducted
 233 at Finokalia (e.g. Koulouri et al., 2008). The particle density, ρ_p , was estimated from the particle
 234 composition from the ACSM ammonium, organics and sulfate, using an organic density of 1.35 g
 235 cm⁻³ as determined by Lee et al. (2010) for the same site during the summer of 2008, the density
 236 of ammonium sulfate (1.77 g cm⁻³) and the equation:

$$237 \quad \rho_p = \left[\frac{x_{a/s}}{\rho_{a/s}} + \frac{x_{org}}{\rho_{org}} \right]^{-1} \quad (4)$$

238 where $x_{a/s}$ is the mass fraction of ammonium sulfate and x_{org} is the mass fraction of the organics,
 239 and $\rho_{a/s}$ and ρ_{org} are the densities of ammonium sulfate and the organics, respectively. It must be

240 noted that nitrate is not taken into account as its concentrations are very close to the limit of
241 detection of the ACSM for the PM₁ fraction at the Finokalia site during the studied period. Particle
242 density was calculated to be $1.56 \pm 0.08 \text{ g cm}^{-3}$ (n=6028), with aerosol concentration ranging from
243 1.33 to 16.65 $\mu\text{g m}^{-3}$ and average value of $5.62 \pm 3 \mu\text{g m}^{-3}$. The particle water calculated by this
244 method is hereafter referred to as $f(RH)_{\text{water}}$, with the uncertainty of this calculation being
245 estimated to be 23% (Guo et al., 2015).

246

247 **2.4 LWC and pH prediction from chemical composition**

248 The water vapor uptake by aerosol establishing equilibrium for ambient temperature and relative
249 humidity conditions is influenced by both inorganic and organic components. LWC, therefore, is
250 directly dependent on aerosol chemical composition and meteorological conditions, as well. As
251 proposed by Guo et al. (2015) and explained below, we calculated the particle water associated
252 with inorganics (W_{inorg}) and the particle water associated with organics (W_{org}). The sum of these
253 two ($W_{\text{inorg}} + W_{\text{org}}$), equal to the total particle water, is then compared to the LWC determined by
254 the two nephelometers.

255 First the particle pH was calculated by the thermodynamic model ISORROPIA-II for the PM₁
256 aerosol fraction. W_{inorg} was calculated based on a thermodynamic equilibrium between an
257 inorganic aerosol (NH₄-SO₄-NO₃-Cl-Na-Ca-K-Mg-water) and its gas phase precursors. For the
258 current analysis, the inputs to ISORROPIA-II are the inorganic ions measured by the ACSM with
259 a 30 min time resolution (except for Cl which is at the detection limit of the instrument at the
260 specific site), the average daily values for Na⁺, Cl⁻, Ca²⁺, K⁺ and Mg²⁺ determined by ion
261 chromatography analysis of PM₁ filters, and RH and T measured by the ambient nephelometer
262 probe. The use of daily values for the macromineral elements obtained from the filter analysis
263 could introduce potential artifacts in the calculation, but is believed to be limited as almost 90%
264 of the fine fraction mass in the area can be attributed to ammonium sulfate and organics (Koulouri
265 et al., 2008; Bougiatioti et al., 2009;2013).

266 The contribution of the organic components to particle water (W_{org}) was determined from the
267 organic hygroscopicity parameter (κ_{org}) from the observed CCN activities of the organic fraction
268 (Cerully et al., 2014):

269

$$W_{org} = \frac{m_s}{\rho_s} \frac{K_{org}}{\left(\frac{1}{RH} - 1\right)} \quad (5)$$

270 where m_s and ρ_s are the organic mass concentration from the ACSM and an organic density,
 271 respectively determined as described in section 2.3.

272 Finally, the particle pH is calculated by the thermodynamic model ISORROPIA-II (Fountoukis
 273 and Nenes, 2007) based on the calculated equilibrium particle hydronium ion concentration in the
 274 aerosol. As ISORROPIA-II does not take into account the LWC associated with the organic
 275 aerosol, a recalculation of pH is made by considering the predicted particle hydronium ion
 276 concentration per volume of air (H^+_{air}) and the total predicted water ($W_{inorg} + W_{org}$) (Guo et al.,
 277 2015):

278

$$pH = -\log_{10} H^+_{aq} = -\log \frac{1000H^+_{air}}{W_{inorg} + W_{org}} \quad (6)$$

279 where the modeled concentrations for LWC and H^+_{air} are $\mu\text{g m}^{-3}$, and H^+_{aq} (mol L^{-1}) is the
 280 hydronium concentration in an aqueous solution and W_{inorg} and W_{org} are in $\mu\text{g m}^{-3}$. ISORROPIA-
 281 II has been previously tested and was able to predict the equilibrium partitioning of ammonia and
 282 nitric acid to within measurement uncertainty (Nowak et al., 2006; Fountoukis et al., 2009;
 283 Hennigan et al., 2015; Guo et al., 2015). Here, ISORROPIA-II was run in the “forward mode”
 284 assuming a metastable aerosol state. It should be noted that gas phase measurements of ammonia
 285 ($\text{NH}_{3(g)}$) were generally not available for the whole measurement period and the sum of NH_3 and
 286 NH_4^+ is assumed to be equal to NH_4^+ . Therefore it is expected that the pH could be underpredicted
 287 by at most one unit (Guo et al., 2015; Weber et al., 2016).

288 To assess the uncertainty in our calculations by not including gas phase NH_3 , a sensitivity analysis
 289 is performed by adding different amounts of $\text{NH}_{3(g)}$ to the system and quantifying the response in
 290 pH. Initial results of ISORROPIA, were compared to results obtained after adding 0.5, 1.2, 3.2 and
 291 $5 \mu\text{g m}^{-3}$ of ammonia. The values of 1.2 and $3.2 \mu\text{g m}^{-3}$ were the median and maximum values of
 292 the gas phase measurements conducted with the WAD during summer 2014, respectively. These
 293 values are also within the observed values reported by Guo et al. (2015). A 3-year study conducted
 294 at the Finokalia station (Kouvarakis et al., 2001) reported that $\text{NH}_{3(g)}$ concentrations during
 295 summertime have an average of $0.27 \mu\text{g m}^{-3}$ (range from 0.07 to $0.68 \mu\text{g m}^{-3}$). Thus, a lower value
 296 ($0.5 \mu\text{g m}^{-3}$) to represent these measurements was also applied and finally $5 \mu\text{g m}^{-3}$ was selected

297 as an extreme value which is very close to the European critical level for NH₃, established to 8 µg
298 m⁻³ as an annual mean (Air Quality Guidelines for Europe, 2000). The results of the sensitivity study
299 are provided in detail in the supplementary material and clearly show that neglecting the gas phase NH₃ in
300 the calculations leads to an underestimate of around 0.5 units in the pH (from 1.38 to 1.85 median values)
301 for the NH₃ range reported for Finokalia.

302 The above results are further supported by the partitioning of nitric acid between the condensed
303 (NO₃⁻) and gas phase (HNO₃). Assuming ideal solutions for the average conditions during the
304 study period derived from ISORROPIA-II, it is calculated that little nitrate should be present in
305 the aerosol phase when pH is lower than 3 (Figure S5 Supplementary). Indeed, with an average
306 concentration of nitrates of 0.12±0.06 µg m⁻³ in the fine mode and the resulting partitioning
307 coefficient of less than 0.2, it is derived that the maximum value of pH that can be observed for
308 the current conditions is 2. This value is in total accordance with the upper limit of pH of 1.85 derived
309 from the formerly presented sensitivity analysis.

310 For days where NH_{3(g)} measurements exist (summer 2014), following the same methodology as
311 before, the measured concentrations of NH_{3(g)} and predictions by the model are in fairly good
312 agreement. Observed gas phase concentrations have an average of 0.79±0.27 µg m⁻³ while predicted
313 concentrations have an average of 0.65±0.32 µg m⁻³. Additionally, we directly compared the pH
314 and LWC values derived from ISORROPIA in the forward mode when calculations were initiated
315 with total (NH₃+NH₄⁺) versus paired results (n=328) for the respective ones initiated with only
316 particulate phase NH₄⁺. The results show that for the specific periods, the addition of NH₃ in the
317 calculations has a minor effect to both pH (y=0.965x, R²=0.584) and LWC (y=1.055x, R²=0.993).

318

319

320 **3. Results and Discussion**

321 **3.1 PM₁ chemical composition and f(RH)_water**

322 For the measurement period, the average values for the main aerosol constituents were 1.85±0.94,
323 2.31±1.61, 0.81±0.58 and 0.52±0.22 µg m⁻³ for organics, sulfate, ammonium and BC, respectively.

324 In terms of contribution to the PM₁ mass concentration, the two most abundant components of the
325 submicron range were sulfate and organics, with mass fractions of 39.6 and 33.8% respectively,
326 followed by ammonium (14.8%), BC (9.3%) and nitrate (2.1%). Chloride has a negligible
327 contribution to the total submicron mass concentration. The time series of the main aerosol

328 constituents and their overall contribution, as measured by the ACSM, are portrayed in Figure 1.
329 Nevertheless, a fluctuation of 10-20% in sulfate and/or ammonium concentrations is not expected
330 to be reflected in a pH change, given the logarithmic scale of the property, which is consistent with
331 the findings of Weber et al. (2016) that pH has a weak sensitivity to a wide range of SO_4^{2-} and
332 NH_3 . This is further elaborated by the sensitivity analysis of gas phase ammonia, where the 5-fold
333 increase of added amount of NH_3 causes a unit change in pH, on average. More details about the
334 ACSM measurements and performance can be found in Stavroulas et al. (in preparation, 2016).
335 The chemical composition is expected to influence the water content of aerosol, as well. Mean RH
336 and T during the study period were $57\pm 11\%$ and 27.4 ± 3.7 °C. As described in the section 2.3,
337 $f(RH)_{\text{water}}$ was calculated from the data from the two nephelometers. With the use of equations
338 1 and 2, we calculated LWC from the nephelometers for the whole measurement period. The
339 average value of $f(RH)_{\text{water}}$ was 2.19 ± 1.75 $\mu\text{g m}^{-3}$, which according to the dry mass
340 measurements, can contribute on average, up to 33% of the total submicron mass concentration.
341 We also sought to establish a link between $f(RH)$ and RH and taking into account all available
342 scattering data (n=7044) the following parameterization has been established:

$$343 \quad f(RH) = 1.067(\pm 0.004) + 1.99(\pm 1.05) \cdot 10^{-7} RH^{3.547(\pm 0.035)} \quad (7)$$

344 Based on this, we reconstructed the time series of the hygroscopic growth factors, with a very good
345 correlation between calculated and measured values ($y=0.99x$, $R^2=0.85$) (also see Supplementary
346 Material). This parameterization does not appear to be influenced by changes in the chemical
347 composition. Uncertainties in the RH measurement by the probe are in the order of 5% while dry
348 and wet aerosol light scattering coefficients measurement uncertainty is in the order of 20%.

349

350 **3.2 Inorganic and organic water predictions**

351 **3.2.1 Inorganic aerosol water**

352 The thermodynamic model ISORROPIA-II (<http://isorrophia.eas.gatech.edu>; Fountoukis and
353 Nenes, 2007) was used to predict the contribution of inorganic species to LWC. The water
354 attributed to the inorganic component of the aerosol has an average value of 1.77 ± 1.45 $\mu\text{g m}^{-3}$. The
355 lowest values are observed during August, probably because of the higher temperatures that
356 enhance evaporation of water from the aerosol. The timeseries for aerosol water associated both
357 with inorganic and organic aerosol components are shown in Figure 2, where it can be seen that
358 most of the time, the variability of the two water components follow each other closely, with water

359 concentrations associated with organics being about 1:10 (12±9%) of those associated with the
360 inorganic aerosol components.

361

362 **3.2.2 Organics: hygroscopicity and aerosol water**

363 The contribution of the organic submicron fraction of the aerosol to the particle water was
364 calculated from the combination of the CCN and chemical composition measurements, as
365 described in section 2.4. Assuming that the cumulative aerosol hygroscopicity can be represented
366 as the sum of contribution of the inorganic (expressed mostly by ammonium sulfate) and organic
367 fraction of the aerosol, the measured hygroscopicity can be calculated by the sum:

$$368 \quad \kappa = \varepsilon_{inorg} \kappa_{inorg} + \varepsilon_{org} \kappa_{org} \quad (8)$$

369 where ε_j and κ_j are the volume fraction and hygroscopicity of the inorganic and organic species.

370 Once the aerosol species concentrations are determined, the corresponding volume fractions for
371 ammonium sulfate and organics are calculated, and a set of hygroscopicity parameter equations is
372 produced. As mentioned in the section 2.2, hygroscopicity parameters for 60, 80, 100 and 120 nm
373 particles are measured. In a former CCN study at the same site (Bougiatioti et al., 2011) it was
374 shown that from many different particle sizes, the characteristic hygroscopicity parameter κ^* of
375 the 100 nm particles was the closest one to the κ determined by PM₁ filter measurements. Therefore
376 we created two different sets of kappa equations along with the volume fractions to calculate the
377 total κ^* , one for the 100 and another one for the 120 nm. With the subsequent application of
378 multivariate regression analysis to the set of n=2429 and 1801 equations (for 100 and 120 nm
379 respectively), κ_{org} is determined to be 0.28±0.01 based on the κ_{100nm} and 0.24±0.01 based on the
380 κ_{120nm} . The average value of 0.26 is used in equation 5 to calculate the time series of W_{org} . The
381 average value W_{org} was thus found to be 0.56±0.37 $\mu\text{g m}^{-3}$, which constitutes on average ~27.5%
382 of the calculated $f(RH)$ _water. The overall uncertainty of this calculation is estimated by Guo et
383 al. (2015) to be around 30%. That study for south eastern United States of America, found a higher
384 contribution of organic species (on average 35%) to the total water that can be explained by the
385 dominance of organics in the submicron range aerosol, with an average mass fraction of 67%.

386

387 **3.2.3. Aerosol water**

388 Aerosol water is the sum of $W_{org}+W_{inorg}$, determined as previously explained. This can be further
389 compared with the time series of the $f(RH)$ _water deduced from concurrent observations by the

390 two nephelometers (section 2.3), independent from the aerosol water calculated as here explained.
391 The particle water predicted from the sum of the organic and inorganic contributions to the water
392 agrees very well with the water measured by the nephelometers (Figure 3). More specifically, the
393 total predicted water is highly correlated and on average within 10% of the measured water, with
394 slope=0.92 and $R^2=0.8$ for the whole measurement period (n=5201 points).

395 The diurnal variability of the calculated water components, along with the total measured water,
396 ambient T and RH is shown in Figure 4. It must be noted at this point that the presented diurnal
397 variability corresponds to the entire period of the study and this pattern seems to be independent
398 of the geographical sector and/or source region, based on the backtrajectory analysis. The diurnal
399 variability of aerosols from different sources, further presented at Section 3.4, have been explicitly
400 studied and found not to differ significantly, therefore the total diurnal variability was selected as
401 being representative of the whole measurement period. Predicted and measured aerosol LWC
402 diurnal variabilities are in very good agreement. As expected, the highest LWC values are observed
403 during nighttime, when RH is also at its maximum, resulting in significant water uptake. W_{org}
404 shows a significant diurnal variability with morning and afternoon average mass concentrations
405 being 10-15 times lower than nighttime ones. Thus, during daytime W_{inorg} is the main component
406 of particle water as the average values for the predicted water ($W_{org}+W_{inorg}$) are very close to the
407 ones of the inorganic water alone. On the other hand, during nighttime, the two averages start to
408 diverge and this could be attributed to the higher contribution of organic water during nighttime,
409 when the photochemical activity and temperature are minimum and RH high.

410

411 **3.3 Aerosol pH**

412 The predicted pH for the aerosols collected at Finokalia during the studied period was highly acidic
413 with an average value of 1.25 ± 1.14 (median 1.51) and varying between -0.97 and 3.75. pH varied
414 by almost 1 unit throughout the day. This can be translated as an almost 10-fold increase of the
415 H^+_{air}/LWC ratio from early morning to mid-day and to a lesser extent during early night. This
416 significant variation in pH can be partially explained by the diurnal variation of H^+ and its increase
417 which coincides temporally with the decrease in pH (Figure 5) and by the reduction of aerosol
418 water (LWC) during daytime compared to the higher LWC during nighttime (Figure 4). This
419 implies the diurnal variability of pH is mostly driven by the reduction of aerosol water during
420 daytime compared to the higher aerosol water during nighttime.

421 When studying the diurnal variability of pH for the different sources/geographical sectors
422 (supplementary material) it occurs that pH differs depending on the source and origin of the air
423 masses. Air masses from the northwestern, northeastern Europe and Turkey, as well as from Greek
424 mainland, do not exhibit significant diurnal variability. When no distinction is made between
425 source regions, the pattern of the overall pH variability with a drop in pH values during midday as
426 presented in Fig. 5, is similar to the diurnal variability of pH for air masses influenced by mineral
427 dust (coming from the SW) and by biomass burning, but with pH values from biomass burning
428 being 1-1.5 units higher. Nevertheless, the mean pH values for each one of the source regions are
429 very close to the overall pH mean value when no distinction of origin is made.

430 Furthermore, accounting only for the water associated with the inorganic aerosol component,
431 particle water is underestimated by around 9%, thus resulting in a slightly lower pH (more acidic)
432 by 0.07-0.38 units of pH (Figure 6a). As seen in section 3.2, W_{org} is on average 27.5% of the total
433 water, as a result the pH increases by 0.14 units when the organic water is included (Figure 6a).
434 When the contribution of organic water is taken into account, the recalculation of pH gives an
435 average value of 1.38 ± 1.11 (median 1.65). The pH calculated by ISORROPIA-II correlates very
436 well with the pH corrected using equation 6 for inclusion of the organic water ($W_{org} + W_{inorg}$)
437 ($R^2=0.98$, Figure 6b). Subsequently, if organic mass and organic hygroscopicity data are not
438 available, ISORROPIA-II solely based on inorganic composition data, will provide an adequate
439 estimate of both H^+_{air} and W_{inorg} and thus of pH.

440

441 **3.4 Aerosol water and pH based on aerosol fraction and source region**

442 Apart from the time-resolved aerosol water and pH calculations for the submicron aerosol fraction,
443 aerosol water and pH were also calculated based on the daily PM_{10} filter chemical composition
444 analysis. It should be noted at this point that there is a strong link between aerosol pH and aerosol
445 size distribution due to the chemical composition changes in the aerosol size. For PM_1 which is in
446 equilibrium and chemical composition is determined by high time resolution measurements leads
447 to higher accuracy in the pH calculations. On the other hand, as PM_{10} particles are not in
448 equilibrium, the respective acidity results are subject to biases and are less quantitative than the
449 submicron results. For comparison purposes the corresponding PM_1 measurements, based on
450 ACSM and the respective PM_1 ionic concentrations for cations such as Na^+ , Ca^{2+} , Mg^{2+} , corresponding to
451 the fine fraction, have also been averaged to daily values. Characteristic source regions/sources

452 were subsequently selected based on back-trajectory analysis and chemical tracers. Furthermore,
453 biomass burning events have been identified based on source apportionment of BC and Positive
454 Matrix Factorization (PMF) analysis of the organic mass spectra obtained from the ACSM as
455 detailed in Bougiatioti et al. (2014). Dust events were identified by backtrajectory analysis of the
456 air masses, originating from Northwestern Africa and confirmed as by containing large amounts
457 of particulate matter ($45\text{-}55 \mu\text{g m}^{-3}$ in PM_{10} mass) and high concentrations of crustal ions, such as
458 Ca^{2+} (median concentration during dust events of $1.8 \mu\text{g m}^{-3}$ versus $0.6 \mu\text{g m}^{-3}$ for the rest of the
459 days) . The main measured and calculated parameters are presented in Table 1.

460 It can be seen that, in general, for anthropogenic-laden air masses (e.g. from Istanbul, Black Sea
461 and Continental Europe) aerosol mass resides mainly (60-70%) in the submicron aerosol and thus
462 the submicron fraction seems also to drive the pH of PM_{10} , while most of the aerosol water (~70%)
463 is present in the coarse fraction. Biomass burning exhibits the highest values of pH in the
464 submicron fraction and the lowest values in total water mass concentration. Interestingly enough,
465 the value of 2.77 for biomass burning pH is well above the value of 2, which favors the partitioning
466 of nitrate to the aerosol phase (e.g., Meskhidze et al., 2003) The correlation observed during fire
467 events between biomass burning organic aerosol and particulate nitrates (e.g. Bougiatioti et al.,
468 2014) may be explained by these higher pH levels, rather than just high levels of nitric acid and/or
469 ammonia. As found in the literature, higher pH levels of biomass burning aerosol can also be
470 attributed to higher concentrations of non-sea salt potassium, which as an ionic species influences
471 pH (Zhang et al. 2015), rendering aerosol more alkaline. Based on the calculated ratios for pH and
472 particle water, it can be derived that the ions that set the bulk pH is mostly driven by the fine mode;
473 the LWC however is associated mostly with the supermicron aerosol mass. Hygroscopic sea salt
474 components are also found in the coarse mode, which can justify the larger contribution of PM_{10}
475 total water, while the submicron fraction may be composed mostly of less hygroscopic, organic
476 components (Bougiatioti et al., 2015). Even though strong dust events were lacking during the
477 study period, numerous weaker events were captured. From these events it occurs that aerosol
478 mass, total water and pH are mostly driven by the coarse aerosol fraction. This is consistent with
479 the findings of Koulouri et al. (2008) that most (86%) of the mass of crustal components were
480 found in coarse particles. In terms of mass concentration, the two natural sources, namely air
481 masses of marine and dust origin, contain the largest amounts of total submicron water. Even
482 though total water exhibits the highest mass concentrations for the specific sources, the relative

483 contribution of organic water to the submicron fraction is the lowest (16% for dust and 17% for
484 marine, based on the median values). For the two natural sources, the ratio between submicron pH
485 and calculated pH for the PM₁₀ fraction is the lowest observed (Table 1), namely 30±17% (median
486 25%) and 6±19% (median 5%).

487

488 **3.5 Atmospheric implications**

489 The pH values for submicron and PM₁₀ fractions is also expected to affect many processes, such
490 as the solubility and therefore the bioavailability of nutrients and especially phosphorus (P) and
491 iron (Fe). This is of crucial importance for seas such as the eastern Mediterranean for which
492 primary productivity is P limited (Krom et al., 1991). One of the striking results of the aerosol pH
493 calculations (presented in Table 1) is that with the exception of biomass burning, pH of all
494 submicronic aerosols including those originating from desert dust is highly acidic with values
495 typically below 2. Although the nutrient flux from the fine mode is certainly smaller than in the
496 coarse mode, it can be transported much further away from source regions before deposition, and
497 be considerably more acidified (hence bioavailable). Based on published literature (Nenes et al.,
498 2011; Meskhidze et al., 2003) under acidic conditions similar to the derived pH aerosol values, the
499 acid-mobilized dissolution of such nutrients is expected to be of great importance.

500 Markaki et al. (2003), by analyzing PM₁₀ aerosols collected in Greece (Finokalia) and Turkey
501 (Erdemli) for dissolved inorganic (DIP) and Total (TP) phosphorus levels have shown that
502 although the SW sector, corresponding to the Sahara, presented TIP levels by a factor of two higher
503 than the European (NE or NW) sectors, the lowest values for the DIP/TIP ratio are associated with
504 the SE and SW sectors, and the highest with the N sectors (difference of almost a factor of 2). The
505 explanation for that behavior is given by the ratio of pH in PM₁ and PM₁₀ aerosols that is presented
506 in Table 1. In aerosols from the NW sector only a very small difference in pH between the two
507 fractions is observed, while in desert aerosols the large amount of dust present in the coarse mode
508 highly influences pH of PM₁₀ leading to a significant decrease of acidity and thus P solubility.
509 However as dust events can transport large amounts of P (up to a factor of three higher than in air
510 masses coming from Europe), any future change in PM₁/PM₁₀ ratio caused by dust events will
511 significantly influence P availability and deposition to the eastern Mediterranean sea and thus to
512 its productivity.

513

514 **Conclusion**

515 Using the aerosol chemical composition measurements by ACSM, thermodynamic models, such
516 as ISORROPIA-II, and CCN measurements we calculated the particle water associated with both
517 inorganics (W_{inorg}) and organics (W_{org}) aerosol components. The sum of these two ($W_{inorg}+W_{org}$)
518 equal to the total particle water is then compared to the LWC determined by the two nephelometers
519 operating in dry and wet mode, respectively. Predicted aerosol water was compared to LWC
520 determined from ambient versus dry light scattering coefficients. At Finokalia the sum of W_{inorg}
521 and W_{org} was highly correlated and in close agreement with the measured LWC (on average within
522 10%), with slope=0.92 and $R^2=0.8$ for the whole measurement period (n=5201 points). As
523 expected, the highest fine aerosol water values are observed during nighttime, when RH is also at
524 its maximum, resulting in significant water uptake. W_{org} shows a significant diurnal variability
525 with morning and afternoon average mass concentrations being 10-15 times lower than nighttime
526 concentrations. Thus, during daytime W_{inorg} is the main form of particle water as the average values
527 for the predicted water ($W_{org}+W_{inorg}$) are very close to the ones of the inorganic water alone. On
528 the other hand, during nighttime, the contribution of organic water to the total aerosol water
529 increases becoming significant during nighttime, when the photochemical activity and temperature
530 are minimum. The average concentration of total aerosol water was found to be $2.19\pm 1.75 \mu\text{g m}^{-3}$,
531 which according to the dry mass measurements, can contribute on average up to 33% of the total
532 submicron mass concentration. The average aerosol water associated with organics, W_{org} , was
533 found to be $0.56\pm 0.37 \mu\text{g m}^{-3}$, which constitutes about 27.5% of the total calculated water. The
534 overall uncertainty of this calculation is estimated to be around 30%.

535 Based on aerosol water, particle pH is also calculated, a parameter which is very important for the
536 atmospheric implications of aerosols but is difficult to measure directly. In the eastern
537 Mediterranean during the studied period, aerosol pH varied from 0.5 to 2.8 indicating that the
538 aerosol was highly acidic throughout the period. Biomass burning aerosol presented the highest
539 values of pH in the submicron fraction and the lowest values in total water mass concentration.
540 Interestingly enough, the value of 2.8 for biomass burning pH is well above the value of 2, which
541 is a lower limit for the occurrence of nitrate's condensation. The correlation, which has been
542 observed during fire events between biomass burning organic aerosol and particulate nitrates, may
543 be explained by these pH levels, which favor nitrate condensation. It can be seen that, in general,
544 for anthropogenic-laden air masses (e.g. from Istanbul, Black Sea and Continental Europe) aerosol

545 mass resides mainly in the submicron aerosol (60-70%) and thus the submicron fraction seems
546 also to drive the pH of PM₁₀, while most of the aerosol water (~70%) is present in the coarse
547 fraction. pH is driven by the fine mode while on the contrary, the grand majority of the aerosol
548 water is present in the coarse mode. This may be due to the important contributions of the
549 hygroscopic sea salt components in the coarse mode and of the less hygroscopic organics in the
550 fine mode.

551 In terms of mass concentration, the air masses affected by the two natural sources, namely of
552 marine and dust origin, contain the largest amounts of total submicron water. Even though total
553 water exhibits the highest mass concentrations for the specific sources, the relative contribution of
554 organic water to the submicron fraction is the lowest (~19% for dust and ~15.5% for marine). For
555 these two natural sources, the ratio between submicron pH and calculated pH for the PM₁₀ fraction
556 is the lowest observed, 30±17% (median 25%) and 6±19% (median 5%).

557 The low pH values observed during the studied period in the submicron mode and independently
558 of air masses origin could have significant implications for nutrient availability and especially for
559 P solubility which is the nutrient limiting sea water productivity of the eastern Mediterranean.

560

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572 **References**

- 573 APHA, AWWA, WEF, (2000) Standard Methods for the Examination of Water and Wastewater,
574 20th ed. American Public Health Association, Washington DC.
- 575 Bougiatioti, A., Zampas, P., Koulouri, E., Antoniou, M., Theodosi, C., Kouvarakis G., Saarikoski,
576 S., Mäkelä, T., Hillamo, R., and Mihalopoulos, N.: Organic, elemental and water-soluble organic
577 carbon in size segregated aerosols, in the marine boundary layer of the Eastern Mediterranean,
578 *Atmos. Environ.*, 64, 251-262, 2013.
- 579 Bougiatioti A., Stavroulas, I., Kostenidou, E., Zampas, P., Theodosi, C., Kouvarakis, G.,
580 Canonaco, F., Prévôt, A.S.H., Nenes, A., Pandis, S.N., and Mihalopoulos, N.: Processing of
581 biomass-burning aerosol in the eastern Mediterranean during summertime, *Atmos. Chem. Phys.*,
582 14, 4793-4807, doi:10.5194/acp-14-4793-2014, 2014.
- 583 Bougiatioti, A., Bezantakos, S., Stavroulas, I., Kalivitis, N., Kokkalis, P., Biskos, G.,
584 Mihalopoulos, N., Papayannis, A., and Nenes, A.: Influence of biomass burning on CCN number
585 and hygroscopicity during summertime in the eastern Mediterranean, *Atmos. Chem. Phys.*
586 *Discuss.*, 15, 21539-21582, doi:10.5194/acpd-15-21539-2015, 2015.
- 587 Carlton, A. G. and Turpin, B. J.: Particle partitioning potential of organic compounds is highest in
588 the Eastern US and driven by anthropogenic water, *Atmos. Chem. Phys.*, 13, 10203–10214,
589 doi:10.5194/acp-13-10203-2013, 2013.
- 590 Cerully, K. M., Bougiatioti, A., Hite Jr., J. R., Guo, H., Xu, L., Ng, N. L., Weber, R., and Nenes,
591 A.: On the link between hygroscopicity, volatility, and oxidation state of ambient and water-
592 soluble aerosol in the Southeastern United States, *Atmos. Chem. Phys. Discuss.*, 14, 30835-
593 30877, doi:10.5194/acpd-14-30835-2014, 2014.
- 594 Eddingsaas, N. C., VanderVelde, D. G., and Wennberg, P. O.: Kinetics and Products of the Acid-
595 Catalyzed Ring-Opening of Atmospherically Relevant Butyl Epoxy Alcohols, *J. Phys. Chem.*
596 *A*, 114, 8106–8113, doi:10.1021/Jp103907c, 2010.
- 597 Fountoukis, C., and Nenes, A.: ISORROPIA II: a computationally efficient thermodynamic
598 equilibrium model for K^+ - Ca^{2+} - Mg^{2+} - NH_4^+ - Na^+ - SO_4^{2-} - NO_3^- - Cl^- - H_2O aerosols, *Atmospheric*
599 *Chemistry and Physics*, 7, 4639-4659, 2007.
- 600 Fountoukis, C., Nenes, A., Sullivan, A., Weber, R., Van Reken, T., Fischer, M., Matias, E., Moya,
601 M., Farmer, D., and Cohen, R. C.: Thermodynamic characterization of Mexico City aerosol
602 during MILAGRO 2006, *Atmospheric Chemistry and Physics*, 9, 2141-2156, 2009.

603 Guo, H., Xu, L., Bougiatioti, A., Cerully, K.M., Capps, S.L., Hite, J.R., Carlton, A.G., Lee, S.-H.,
604 Bergin, M.H., Ng, N.L., Nenes, A., and Weber, R.J.: Particle water and pH in the southeastern
605 United States, *Atmospheric Chemistry and Physics*, 15, 5211-5228, doi:10.5194/acp-15-5211-
606 2015, 2015.

607 Hennigan, C. J., Izumi, J., Sullivan, A. P., Weber, R. J., and Nenes, A.: A critical evaluation of
608 proxy methods used to estimate the acidity of atmospheric particles, *Atmos. Chem. Phys.*, 15,
609 2775– 2790, doi:10.5194/acp-15-2775-2015, 2015.

610 Jickells, T.D., Z. S. An, K. K. Andersen, A. R. Baker, G. Bergametti, N. Brooks, J. J. Cao, P. W.
611 Boyd, R. A. Duce, K. A. Hunter, et al., Global iron connections between desert dust, ocean
612 biogeochemistry, and climate. *Science* 308:67–71, 2005.

613 Keene, W. C., A. A. P. Pszenny, J. R. Maben, and R. Sander: Variation of marine aerosol acidity
614 with particle size, *Geophys. Res. Lett.*, 29(7), doi:10.1029/2001GL013881, 2002.

615 Keene, W.C., Pszenny, A.A.P., Maben, J.R., Stevenson, E., Wall, A.: Closure evaluation of size-
616 resolved aerosol pH in the New England coastal atmosphere during summer, *J. Geophys. Res.*,
617 109, D23307, doi:10.1029/2004JD004801, 2004.

618 Khlystov, A., Stanier, C.O., Takahama, S. and Pandis, S.N.: Water content of ambient aerosol
619 during the Pittsburgh Air Quality Study, *J. Geophys. Res. Atmospheres*, Vol 110, Issue D7,
620 doi:10.1029/2004JD004651, 2005.

621 Kim, J., Yoon, S.-C., Jefferson, A., and Kim, S.-W.: Aerosol hygroscopic properties during Asian
622 dust, pollution, and biomass burning episodes at Gosan, Korea in April 2001, *Atmospheric* 635
623 *Environment*, 40, 1550-1560, 10.1016/j.atmosenv.2005.10.044, 2006.

624 Koulouri, E., Saarikoski, S., Theodosi, C., Markaki, Z., Gerasopoulos, E., Kouvarakis, G., Mäkelä,
625 T., Hillamo, R., and Mihalopoulos, N.: Chemical composition and sources of fine and coarse
626 aerosol particles in the Eastern Mediterranean, *Atmos. Environ.*, 42, 6542-9550, 2008.

627 Kouvarakis, G. Mihalopoulos, N., Tselepidis, T., Stavrakakis, S.: On the importance of
628 atmospheric nitrogen inputs on the productivity of Easter Mediterranean, *Global Biochem.*
629 *Cycles*, 15, 805-818, 2001.

630 Kreidenweis, S. M., Petters, M. D., and DeMott, P. J.: Singleparameter estimates of aerosol water
631 content, *Environ. Res. Lett.*, 3, 035002, doi:10.1088/1748-9326/3/3/035002, 2008.

632 Krishnamurthy, A., Moore, J.K., Mahowald, N., Luo, C., and Zender, C.S.: Impacts of atmospheric
633 nutrient inputs on marine biogeochemistry, *J. Geophys. Res.*, 115, G01006,
634 doi:10.1029/2009JG001115, 2010.

635 Krom, M.D., Kress, N., Brenner, S., and Gordon, L.I.: Phosphorus limitation of primary
636 productivity in the eastern Mediterranean Sea, *Limnol. Oceanogr.*, 36 (3), 424-432, 1991.

637 Lee, T., Sullivan, A. P., Mack, L., Jimenez, J. L., Kreidenweis, S. M., Onasch, T. B., Worsnop, D.
638 R., Malm, W., Wold, C. E., Hao, W. M., and Collett, J. L.: Variation of chemical smoke marker
639 emissions during flaming vs. smoldering phases of laboratory open burning of wildland fuels,
640 *Aerosol Sci. Technol.*, **44**, 1–5, doi:10.1080/02786826.2010.499884, 2010.

641 Liao, H. and Seinfeld, J. H.: Global impacts of gas-phase chemistry aerosol interactions on direct
642 radiative forcing by anthropogenic aerosols and ozone, *J. Geophys. Res.-Atmos.*, 110, D18208,
643 doi:10.1029/2005jd005907, 2005.

644 Magi, B. I.: Effects of humidity on aerosols in southern Africa during the biomass burning season,
645 *Journal of Geophysical Research*, 108, doi:10.1029/2002jd002144, 2003.

646 Mahowald, N., Jickells, T.D., Baker, A.R., Artaxo, P., Benitez-Nelson, C.R., Bergametti, J., Bond,
647 T.C., Chen, Y., Cohen, D.D., Herut, B., Kubilay, N., Losno, R., Luo, C., Maenhaut, W.,
648 McGee, K.A., Okin, G.S., Siefert, R.L., and Tsukuda, S.: Global distribution of atmospheric
649 phosphorus sources, concentrations and deposition rates, and anthropogenic impacts, *Global*
650 *Biogeochem. Cycles*, 22, GB4026, doi:10.1029/2008GB003240, 2008.

651 Mahowald, N. M., Engelstaedter, S., Luo, C., Sealy, A., Artaxo, P., Benitez-Nelson, C., Bonnet,
652 S., Chen, Y., Chuang, P.Y., Cohen, D.D., Dulac, F., Herut, B., Johansen, A.M., Kubilay, N.,
653 Losno, R., Maenhaut, W., Paytan, A., Prospero, J.M., Shank, L.M., and Siefert, R.L.:
654 Atmospheric iron deposition: Global distribution, variability, and human perturbations, *Annu.*
655 *Rev. Mater. Sci.*, 1, 245–278, doi:10.1146/annurev.marine.010908.163727, 2009.

656 Meskhidze, N., Chameides, W.L., Nenes, A., and Chen, G.: Iron mobilization in mineral dust: Can
657 anthropogenic SO₂ emissions affect ocean productivity?, *Geophys. Res. Lett.*, 30 (21),
658 doi:10.1029/2003GL018035, 2003.

659 Meskhidze, N., Chameides, W. and Nenes, A. (2005) Dust and Pollution: A Recipe for Enhanced
660 Ocean Fertilization? *J. Geoph. Res.*, **110**, D03301, doi:10.1029/2004JD005082

661 Mihalopoulos, N., Stephanou, E., Kanakidou, M., Pilitsidis, S., and Bousquet, P.: Tropospheric
662 aerosol ionic composition above the eastern Mediterranean area, *Tellus B*, **49**, 314–326,
663 1997. Moore, R.H., and Nenes, A.: Scanning Flow CCN Analysis-A method for fast
664 measurements of CCN spectra, *Aerosol Sci. Technol.*, 43:12, 1992-1207, 2009.

665 Myriokefalitakis, S., Tsigaridis, K., Mihalopoulos, N., Sciare, J., Nenes, A., Kawamura, K.,
666 Segers, A., and Kanakidou, M.: In-cloud oxalate formation in the global troposphere: a 3-D
667 modeling study, *Atmos. Chem. Phys.*, 11, 5761–5782, doi:10.5194/acp-11-5761-2011, 2011.

668 Myriokefalitakis, S., Daskalakis, N., Mihalopoulos, N., Baker, A. R., Nenes, A., and
669 Kanakidou, M.: Changes in dissolved iron deposition to the oceans driven by human activity: a
670 3-D global modelling study, *Biogeosciences*, 12, 3973-3992, doi:10.5194/bg-12-3973-2015,
671 2015. Nenes, A., Krom, M. D., Mihalopoulos, N., Van Cappellen, P., Shi, Z., Bougiatioti, A.,
672 Zampas, P., and Herut, B.: Atmospheric acidification of mineral aerosols: a source of
673 bioavailable phosphorus for the oceans, *Atmos. Chem. Phys.*, 11, 6265–6272, doi:10.5194/acp-
674 11-6265-2011, 2011.

675 Ng, N.L., Herndon, S.C., Trimborn, A., Canagaratna, M.R., Croteau, P.L., Onasch, T.B., Sueper,
676 D., Worsnop, D.R., Zhang, Q., Sun, Y.L., and Jayne, J.T.: An Aerosol Chemical Speciation
677 Monitor (ACSM) for routine monitoring of the composition and mass concentration of ambient
678 aerosol., *Aerosol Sci. Technol.*, **45** (7), 780-794, 2011.

679 Nguyen, T.K.V., Petters, M.D., Suda, S.R., Guo, H., Weber, R.J., and Carlton, A.G.: Trends in
680 particle-phase liquid water during the Southern Oxidant and Aerosol Study, *Atmos. Chem.*
681 *Phys.*, 14, 10911-10930, doi:10.5194/acp-14-10911-2014, 2014.

682 Nowak, J. B., Huey, L. G., Russell, A. G., Tian, D., Neuman, J. A., Orsini, D., Sjostedt, S. J.,
683 Sullivan, A.P., Tanner, D. J., Weber, R. J., Nenes, A., Edgerton, E., and Fehsenfeld, F. C.:
684 Analysis of urban gas phase ammonia measurements from the 2002 Atlanta Aerosol Nucleation
685 and Real-Time Characterization Experiment (ANARChE), *Journal of Geophysical Research*,
686 111, 10.1029/2006jd007113, 2006.

687 Petters, M.D., and Kreidenweis, S.M.: A single parameter representation of hygroscopic growth
688 and cloud condensation nucleus activity, *Atmos. Chem. Phys.*, 7, 1961-1971, doi: 10.5194/acp-
689 8-6273-2008, 2007.

690 Pilinis, C., Pandis, S. N., and Seinfeld, J. H.: Sensitivity of Direct Climate Forcing by Atmospheric
691 Aerosols to Aerosol-Size and Composition, *J. Geophys. Res.-Atmos.*, 100, 18739–18754,
692 doi:10.1029/95jd02119, 1995.

693 Pöschl, U., and Shiraiwa, M.: Multiphase chemistry at the atmosphere-biosphere interface
694 influencing climate and public health in the Anthropocene, *Chem. Rev.*, 115, 4440-4475,
695 doi:10.2021/cr500487s, 2015

696 Roberts, G.C., and Nenes, A.: A continuous-flow streamwise thermal-gradient CCN chamber for
697 atmospheric measurements, *Aerosol Sci. Technol.*, 39, 206-221,
698 dpo:10.1080/027868290913988, 2005.

699 Sheridan, P. J.: Spatial variability of submicrometer aerosol radiative properties over the Indian
700 Ocean during INDOEX, *Journal of Geophysical Research*, 107, 10.1029/2000jd000166, 2002.

701 Shi, Z., Krom, M.D., Jickells T.D., Bonneville, S., Carslaw, K.S., Mihalopoulos, N., Baker, A.R.,
702 Benning, L.G.: Impacts on iron solubility in the mineral dust by processes in the source region
703 and the atmosphere: A review, *Aeol. Res.* 5, 21–42, doi:/10.1016/j.aeolia.2012.03.001, 2012.

704 Solmon, F., Chuang, P.Y., Meskhidze, N., and Chen, Y.: Acidic processing of mineral dust iron
705 by anthropogenic compounds over the north Pacific Ocean, *J. Geophys. Res., Atmospheres*,
706 114 (D2), doi:10.1029/2008JD010417, 2009.

707 Spindler, G., Hesper, J., Brüggemann, E., Dubois R., Müller, Th., and Herrmann, H.: Wet annular
708 denuder measurements of nitrous acid: laboratory study of the artefact reaction of NO₂ with
709 S(IV) in aqueous solution and comparison with field measurements, *Atmos. Environ.*, 37, 2643-
710 2662, 2003.

711 A. F. Stein, R. R. Draxler, G. D. Rolph, B. J. B. Stunder, M. D. Cohen, and F. Ngan, 2015: NOAA's
712 HYSPLIT Atmospheric Transport and Dispersion Modeling System. *Bull. Amer. Meteor.*
713 *Soc.*, **96**, 2059–2077, doi: <http://dx.doi.org/10.1175/BAMS-D-14-00110.1>, 2015.

714 Stookey, L. C.: Ferrozine - a new spectrophotometric reagent for iron. *Analytical Chemistry* 42,
715 779–781, 1970.

716 Stumm, W. and Morgan, J. J.: *Aquatic Chemistry*, Wiley- Interscience, New York, 1996.

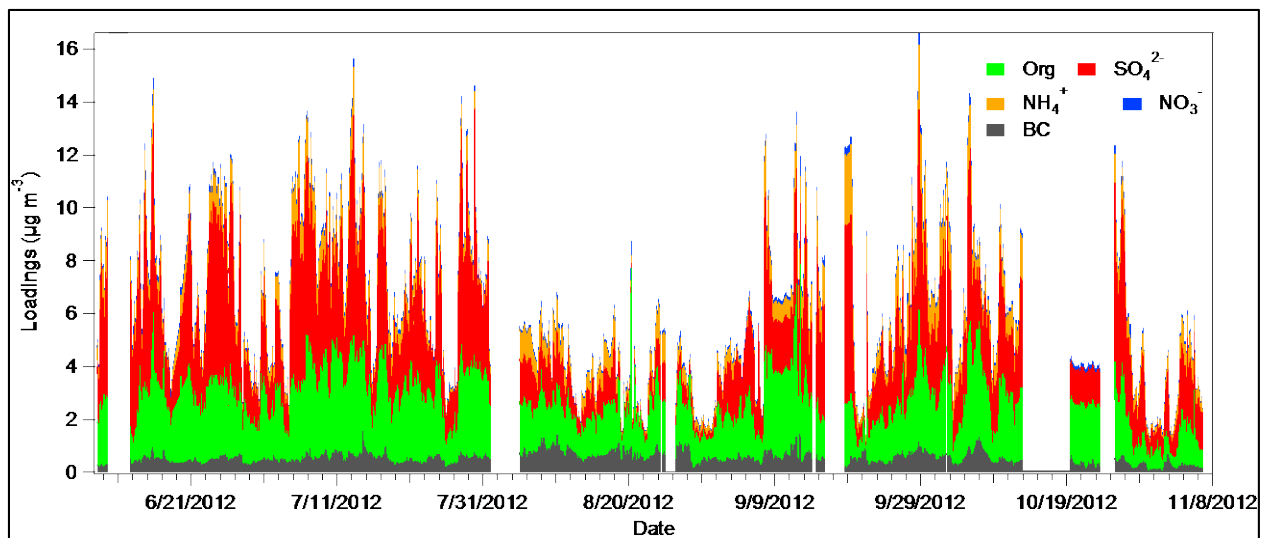
717 Surratt, J. D., Lewandowski, M., Offenberg, J. H., Jaoui, M., Kleindienst, T. E., Edney, E. O., and
718 Seinfeld, J. H.: Effect of acidity on secondary organic aerosol formation from isoprene, *Environ.*
719 *Sci. Technol.*, 41, 5363–5369, doi:10.1021/es0704176, 2007.

720 Surratt, J. D., Chan, A. W., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S. P.,
721 Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in secondary
722 organic aerosol formation from isoprene, *P. Natl. Acad. Sci. USA*, 107, 6640–6645,
723 doi:10.1073/pnas.0911114107, 2010.

724 Weber, R.J., Guo, H., Russel, A.G., and Nenes, A.: High aerosol acidity despite declining
725 atmospheric sulfate concentrations over the past 15 years, *Nat. Geosci.*, *in press*, 2016.

726 Wyers, G.P., Otjes, R.P., Slanina, J.: A continuous-flow denuder for the measurement of ambient
727 concentrations and surface-exchange fluxes of ammonia, *Atmos. Environ.*, 27, 2085–2090,
728 1993.

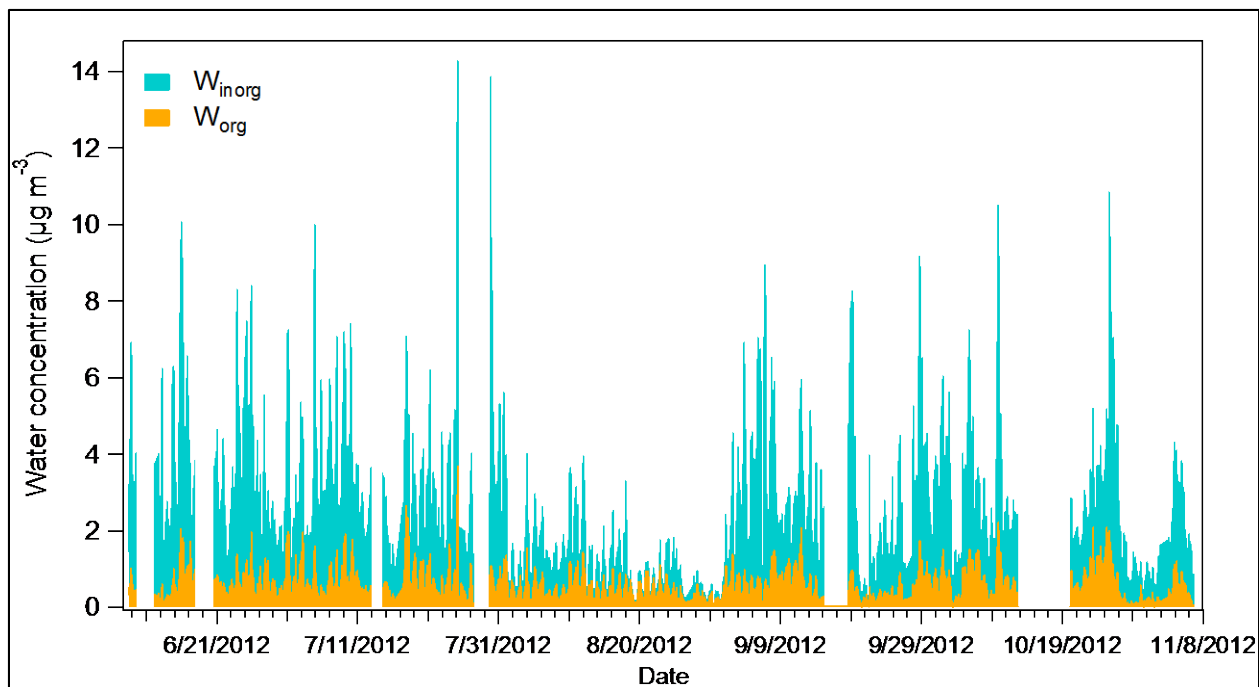
729 Zhang, Z., Gao, J., Engling, G., Tao, J., Chai, F., Zhang, L., Zhang, R., Sang, X., Chan, C., Lin,
730 Z., and Cao, J.: Characteristics and applications of size-segregated biomass burning tracers in
731 China's Pearl River Delta region, *Atmos. Environ.*, 102, 290-301, 2015.



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733 **Figure 1:** Time series of the main submicron aerosol components at Finokalia, Crete, during the
734 measurement period in 2012.

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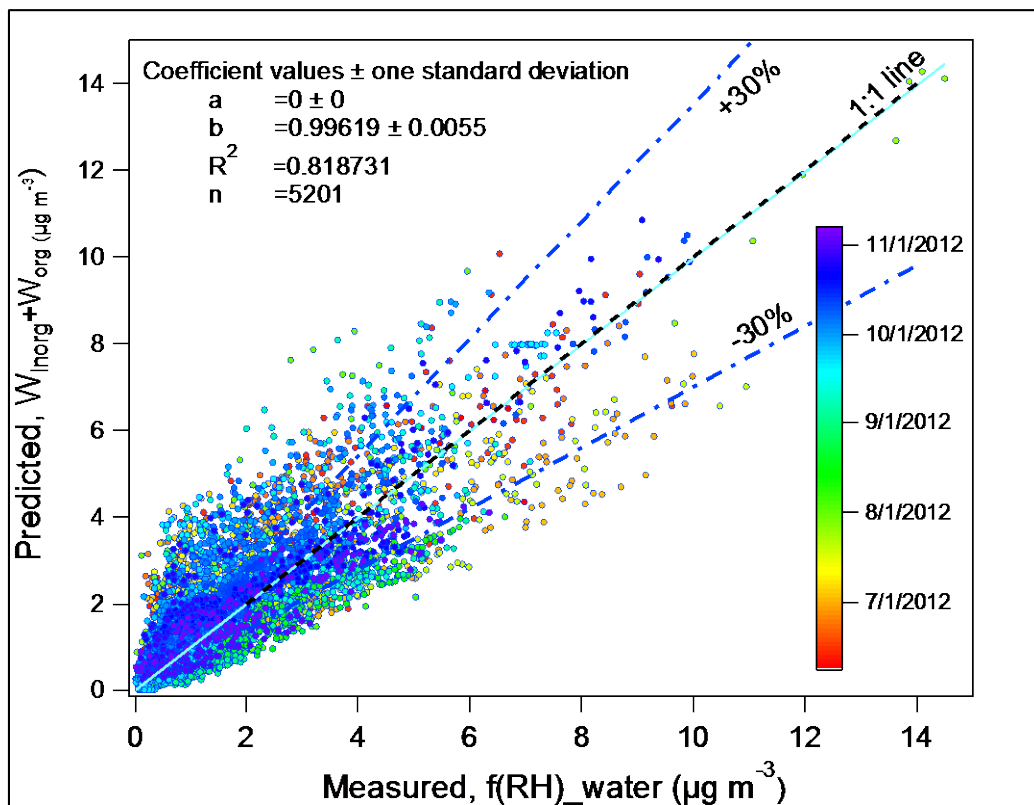


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737 **Figure 2:** Time series of the two calculated aerosol water components, the water associated with
738 the inorganic (W_{inorg}) and the organic (W_{org}) aerosol components.

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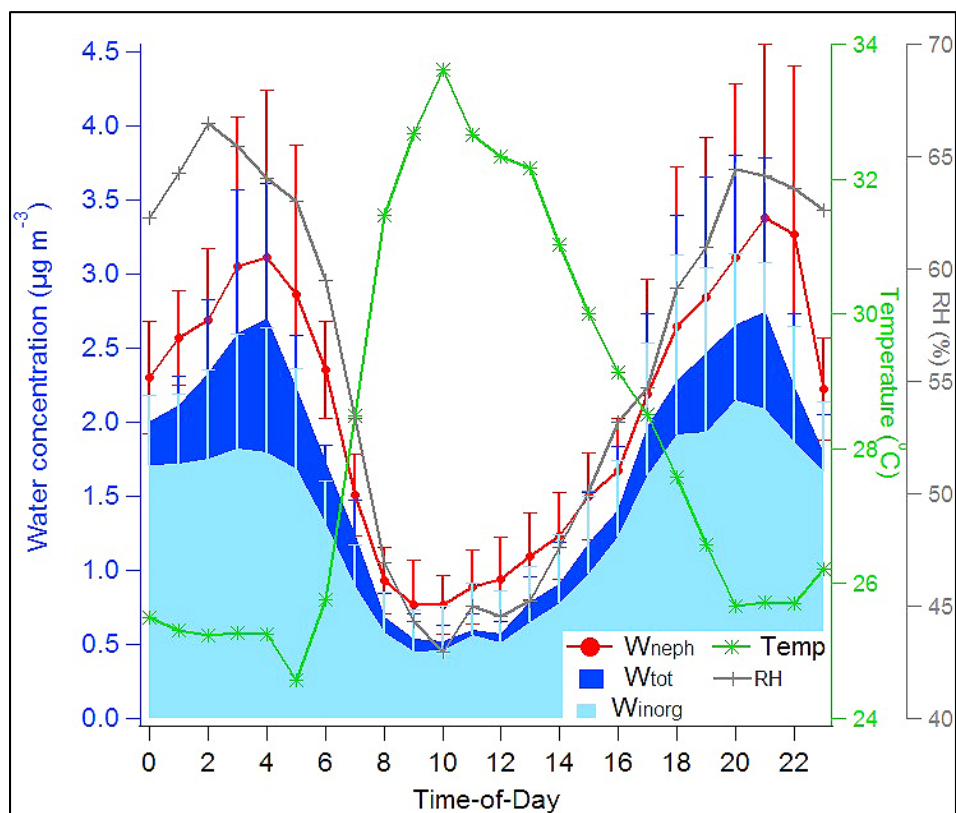
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742 **Figure 3:** Correlation between calculated and measured LWC of aerosol.

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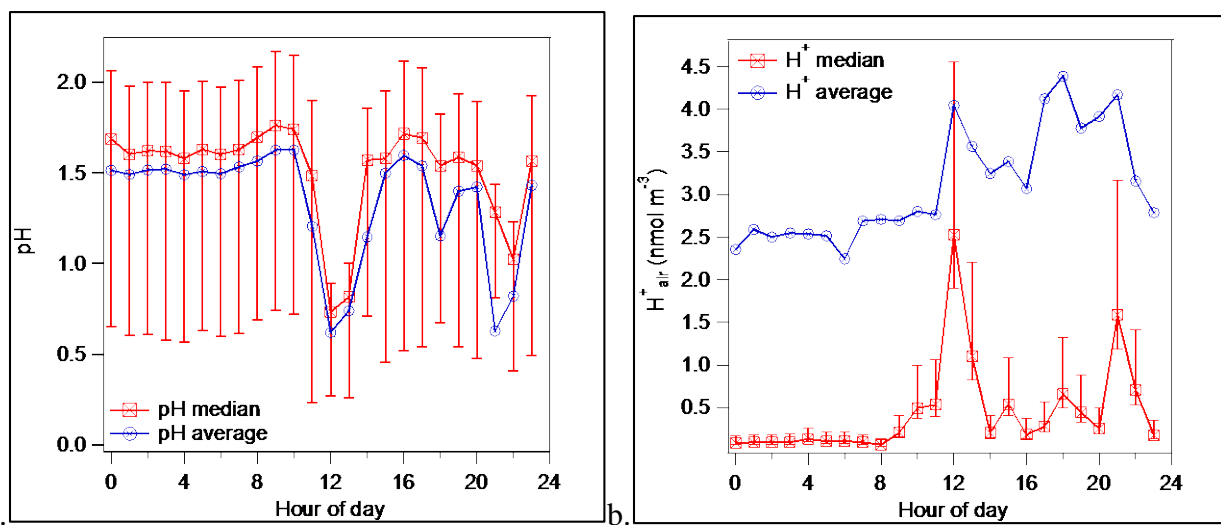


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746 **Figure 4:** Diurnal profiles of predicted and measured water, along with measured RH and T.

747 Average hourly averages and standard deviations plotted as error bars in local hour are shown.

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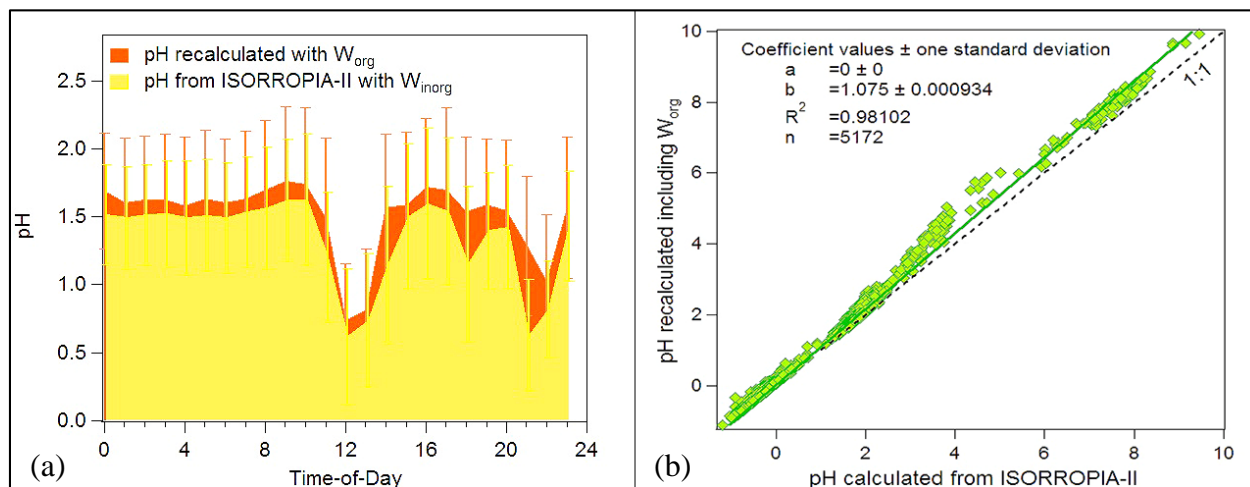


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751 **Figure 5:** Diurnal profile of pH (a) and H^+ (b) calculated based on total predicted water and
752 hydronium concentrations predicted by ISORROPIA-II. Average and median (with 25%, 75%
753 quantiles) values are provided.

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758 **Figure 6:** a) Diurnal variability of pH based on total predicted water (orange using both W_{org} and
759 W_{inorg}) and only the inorganic water (yellow using only W_{inorg}), respectively. Average hourly
760 values and standard deviation in the form of error bars are provided, b) Comparison between the
761 pH predicted by ISORROPIA-II based on inorganic aerosol composition and that recalculated
762 including also the organic component.

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776 **Table 1:** Average values of water content (total and organic) and pH with standard deviation and
 777 median values in parenthesis.

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Source / region	PM ₁			PM ₁ /PM ₁₀		
	water _{tot} (μg m ⁻³)	water _{org} (μg m ⁻³)	pH	mass	water _{tot}	pH
Biomass Burning (n=7)	1.33±1.22 (1.56)	0.37±0.33 (0.39)	2.77±0.88 (2.67)	0.48±0.16 (0.47)	0.08±0.09 (0.05)	0.89±0.2 (0.93)
Istanbul & Black Sea (n=7)	1.72±0.54 (1.21)	0.81±0.34 (0.75)	1.92±0.24 (2.00)	0.72±0.09 (0.69)	0.33±0.21 (0.26)	0.94±0.67 (0.88)
Continental Europe (n=9)	1.52±0.64 (1.32)	0.48±0.34 (0.47)	1.30±0.19 (1.26)	0.60±0.18 (0.64)	0.31±0.17 (0.28)	0.58±0.47 (0.50)
Dust (n=7)	2.48±0.54 (2.44)	0.48±0.22 (0.39)	1.48±0.44 (1.59)	0.34±0.08 (0.38)	0.39±0.08 (0.29)	0.30±0.17 (0.25)
Marine (n=7)	2.38±1.24 (1.91)	0.37±0.17 (0.34)	0.44±0.67 (0.43)	0.58±0.27 (0.57)	0.83±0.6 (0.70)	0.06±0.19 (0.05)

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