1 Seasonality of isoprenoid emissions from a primary rainforest in central Amazonia

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29 Abstract

Tropical rainforests are an important source of isoprenoid and other Volatile Organic 30 Compound (VOC) emissions to the atmosphere. The seasonal variation of these compounds 31 32 is however still poorly understood. In this study, vertical profiles of mixing ratios of isoprene, total monoterpenes and total sesquiterpenes, were measured within and above the 33 canopy, in a primary rainforest in central Amazonia, using a Proton Transfer Reaction -34 Mass Spectrometer (PTR-MS). Fluxes of these compounds from the canopy into the 35 36 atmosphere were estimated from PTR-MS measurements by using an inverse Lagrangian transport model. Measurements were carried out continuously from September 2010 to 37 38 January 2011, encompassing the dry and wet seasons. Mixing ratios were higher during the dry (isoprene -2.68 ± 0.9 ppbv, total monoterpenes -0.67 ± 0.3 ppbv; total sesquiterpenes -39 0.09 ± 0.07 ppbv) than the wet season (isoprene - 1.66 ± 0.9 ppbv, total monoterpenes -40 0.47±0.2 ppbv; total sesquiterpenes - 0.03±0.02 ppbv) for all compounds. Ambient air 41 temperature and photosynthetically active radiation (PAR) behaved similarly. Daytime 42 isoprene and total monoterpene mixing ratios were highest within the canopy, rather than 43 near the ground or above the canopy. By comparison, daytime total sesquiterpene mixing 44 ratios were highest near the ground. Daytime fluxes varied significantly between seasons 45 for all compounds. The maximums for isoprene $(2.53\pm0.5 \text{ umol m}^{-2} \text{ h}^{-1})$ and total 46 monoterpenes $(1.77\pm0.05 \text{ }\mu\text{mol }\text{m}^{-2} \text{ }h^{-1})$ were observed in the late dry season, whereas the 47 maximum for total sesquiterpenes was found during the dry-to-wet transition season 48 $(0.77\pm0.1 \text{ }\mu\text{mol }\text{m}^{-2}\text{ }h^{-1})$. These flux estimates suggest that the canopy is the main source of 49 isoprenoids emitted into the atmosphere for all seasons. However, uncertainties in 50 turbulence parameterization near the ground could affect estimates of fluxes that come from 51 52 the ground. Leaf phenology seemed to be an important driver of seasonal variation of isoprenoid emissions. Although remote sensing observations of changes in leaf area index 53 were used to estimate leaf phenology, MEGAN 2.1 did not fully capture the behavior of 54 seasonal emissions observed in this study. This could be a result of very local effects on the 55 56 observed emissions, but also suggest that other parameters need to be better determined in 57 Biogenic Volatile Organic Compound (BVOC) models. Our results support established findings that seasonality of isoprenoids are driven by seasonal changes in light, temperature 58 and leaf phenology. However, they suggest that leaf phenology and its role on isoprenoid 59

production and emission from tropical plant species needs to be better understood in order to develop mechanistic explanations for seasonal variation in emissions. This also may reduce the uncertainties of model estimates associated with the responses to environmental factors. Therefore, this study strongly encourages long-term measurements of isoprenoid emissions, environmental factors and leaf phenology from leaf to ecosystem scale, with the purpose of improving BVOC model approaches that can characterize seasonality of isoprenoid emissions from tropical rainforests.

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68 Key-words: Isoprene, monoterpenes, sesquiterpenes, leaf phenology, seasonal changes

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70 1. Introduction

Terrestrial vegetation emits high quantities of biogenic volatile organic compounds 71 72 (BVOCs) to the atmosphere (Guenther et al., 2006, 2012), which are removed by oxidation reactions, deposition of reaction products (Lelieveld et al., 2008) and consumption by 73 74 surfaces (Gray et al., 2014). Emissions and subsequent transformations in the atmosphere have been widely explored by the scientific community. However, there is still a need for 75 76 improving our understanding of how BVOC emissions and their reaction products vary seasonally and are involved in atmosphere chemistry, biogeochemical cycling and climate 77 at local, regional and global scales. 78

Despite a large number of BVOC species that have been identified within plants and in emissions from plants, the largest part of the global biogenic emissions and subsequent effect on atmospheric chemistry are thought to be associated with isoprenoids (Laothawornkitkul et al., 2009). The isoprenoids are an important class of organic compounds that include isoprene (containing five carbon atoms - C_5), monoterpenes (10 carbon atoms - C_{10}), sesquiterpenes (15 carbon atoms - C_{15}) and diterpenes (20 carbon atoms - C_{20}) (Guenther, 2002).

Isoprene, as the building block of the higher order isoprenoids, is the dominant compound in emissions from many landscapes and has the single largest contribution to total global vegetation BVOC emission, with an estimated global annual emission of about

400–600 Tg C (see Table 1 of Arneth et al., 2008). Even though there are more than 1000 89 90 monoterpene compounds identified in plants, only a few (less than 12) monoterpenes comprise a large fraction of total monoterpene emissions into the atmosphere (Guenther, 91 2002). Compounds such as α -pinene, *t*- β -ocimene, β -pinene, limonene, sabinene, myrcene, 92 3-carene, camphene, β -phellandrene and terpinolene dominate monoterpene emissions 93 globally (Guenther et al., 2012). However, at regional scales other monoterpene compounds 94 may also be important (Geron et al., 2000; Jardine et al., 2015). Only a few (e.g., β-95 caryophyllene) of about 3000 sesquiterpenes and none of the 2000 diterpenes are known to 96 be emitted into the atmosphere in considerable amounts (Guenther, 2002). However, there 97 are many compounds in the atmosphere that are still unknown or unexplored (Goldstein et 98 99 al., 2007, Park et al., 2013), suggesting that the characterization of sesquiterpene emissions and other trace gases is still an open question. 100

Although models indicate that tropical rainforests are the main source of isoprenoid 101 102 emissions to the global atmosphere (Guenther et al., 2012), estimates of global annual 103 emissions of isoprenoid still have large uncertainties (Guenther et al., 2006). One approach to constraining these estimates, specifically for isoprene, is the use of remotely sensed 104 concentrations of BVOC oxidation products in the atmosphere in order to make top-down 105 model estimates (Barkley et al., 2008, 2009, 2013; Stavrakou et al., 2009, 2015). This 106 approach has also suggested seasonal patterns in the emissions of this organic compound 107 (Barkley et al., 2009). In addition, seasonal variations of isoprene emissions in the 108 Amazonian rainforest are suggested based on comparison of some studies with intensive 109 campaigns in situ (Table 1). This seasonality may be driven by light and temperature 110 seasonal variation and leaf phenology (Barkley et al., 2009), and seasonal changes in 111 112 insolation is probably the main driver of leaf phenology (Jones et al., 2014).

113 Therefore, the objective of this study was to quantify the seasonal variation of 114 mixing ratios and emissions of isoprene, total monoterpenes and total sesquiterpenes in a 115 primary rainforest in central Amazonia and to correlate them to seasonal variations of 116 environmental (temperature and light) and biological (leaf phenology) factors.

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118 **2. Material and methods**

119 **2.1 Site description**

120 Isoprenoid vertical profiles were investigated at the triangular tower (TT34 tower -2°35.37'S, 60°06.92'W) on a plateau of the Cuieiras Biological Reserve, a primary 121 122 rainforest reserve located approximately 60 km northwest of Manaus city, in the central Amazonian Basin, in Amazonas, Brazil (Martin et al., 2010). The vegetation in this area is 123 124 considered mature terra firme rain forest (Pires and Prances, 1985), with a leaf area index of 4.7 (Malhi et al., 2009). The diversity of tree species is above 200 species ha⁻¹ (Oliveira 125 et al., 2008). Annual precipitation is about 2500 mm (Fig. 1a), with December to May 126 being the wetter period. Although severe droughts impacted part of the Amazon basin in 127 2005 and in 2010, those droughts did not affect central Amazonia (Marengo et al., 2008, 128 2011). However, micrometeorological measurements from 1999 to 2012 showed that from 129 August to September the monthly cumulative precipitation can be less than 100 mm per 130 month (Fig. 1a), characterizing this period as dry season. Average air temperature ranges 131 between 24 °C (in April) and 27 °C (in September) (Fig. 1e). Soil moisture near the surface 132 is slightly reduced (10%) during the dry compared to the wet season (Cuartas et al., 2012). 133

The period of this study (from September 2, 2010 to January 27, 2011) represents 134 the second half of the dry season (September 2010 - October 2010), the dry-to-wet 135 transition season (November 2010) and the beginning of the wet season (December 2010 -136 137 January 2011). The whole period of measurements includes the period of low precipitation 138 and when precipitation is increasing (Fig. 1b), and when photosynthetically active radiation (PAR) (Fig. 1d) and air temperature (Fig. 1f) are at their peaks. As October 2010 had more 139 140 precipitation only at the end of the month, for this study October 2010 is also considered as dry season. This is supported by the fact that the length and intensity of the dry season 141 142 varies from year to year (da Rocha et al., 2009).

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144 **2.2.** Isoprenoid measurements and data analysis

Ambient mixing ratio measurements of isoprene, total monoterpenes and total sesquiterpenes were carried out using a commercial high sensitivity proton-transfer reaction mass spectrometer (PTR-MS, IONICON, Austria). The PTR-MS was operated in standard

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conditions with a drift tube voltage of 600 V and drift tube pressure of 2.0 mbar (E/N, 136 148 149 Td). During each PTR-MS measurement cycle, the following mass-to-charge ratios (m/z)were monitored: 21 $(H_3^{18}O^+)$, 32 (O_2^+) , 37 $(H_2O-H_3O^+)$ with a dwell time of 20 ms each; 69 150 (isoprene-H⁺), 137 (total monoterpenes-H⁺) and 205 (total sesquiterpenes-H⁺) with a dwell 151 time of 5 s each (Jardine et al., 2011, 2012; Lindinger et al., 1998). The isoprenoid vertical 152 profile was installed with six ambient air inlets at different tower heights (2, 11, 17, 24, 30 153 and 40 m). Air was sequentially sampled during 10 min at each of the six heights, resulting 154 in one complete profile every hour. Average mixing ratios were calculated for the daytime 155 period (10:00 - 16:00, LT) and for the nighttime period (22:00 - 04:00, LT). Calibration 156 slope (m, ppbv/normalized counts per second (PTR-MS signal)) for isoprene, total 157 158 monoterpenes, and total sesquiterpenes were obtained twice in the field using the dynamic solution injection technique (Jardine et al., 2010). Solutions of isoprene, α -pinene, and β -159 caryophyllene standards (> 95% purity, Merk) in 100 mL of cyclohexane were injected into 160 the mixing vial at 0.5, 1.0, 2.0, and 3.0 µL min⁻¹ (30 min each flow rate) with a constant 161 162 dilution flow of 1.0 slpm ultra high purity nitrogen passing through. The linearity of calibrations was significant, being r² of 0.92-0.97 for isoprene, r² of 0.98-0.99 for α -pinene, 163 and r^2 of 0.90-098 for β -caryophyllene. Sample air isoprenoid mixing ratios were 164 calculated by multiplying the calibration slope by normalized counts per second (PTR-MS 165 166 signal) (average of two calibration slopes). Calibration slopes obtained on October 2010 were within 10 % relative to those from the calibration carried out in September 2010 167 (isoprene 7.2 %, α -pinene – 8.2%, and β -caryophyllene – 2.5%). For 4-7 days before each 168 isoprenoid profile measurement period, ultra high purity nitrogen was run into the inlet of 169 the PTR-MS for 2 h in order to obtain the background signals. The limit of detection for 170 isoprene was 0.14 ppby, 0.15 ppby for total monoterpenes and 0.1 ppby for total 171 sesquiterpenes. More details about calibration and experimental design can be obtained in 172 Jardine et al. (2011) and Jardine et al. (2012), in which a subset of these data are already 173 described. While the previous study considered a subset of this data and time period 174 (Jardine et al., 2011, 2012), this study examines the whole dataset and focuses on 175 seasonality of mixing ratios and fluxes. Also, this is the first study in central Amazonia that 176 correlates long-term measurements of isoprenoids, light and temperature, and leaf 177 phenology. 178

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2.3 Isoprenoid gradient flux, and modeled flux estimates - Model of Emissions of Gases and Aerosols from Nature (MEGAN 2.1)

Fluxes of isoprene, total monoterpenes and total sesquiterpenes - for dry, dry-to-wet transition and wet seasons - were estimated using the average daytime (10:00-14:00, LT) concentration vertical profile throughout the canopy and applying an inverse Lagrangian transport model (ILT) (Raupach, 1989; Nemitz et al., 2000; Karl et al., 2004; Karl et al., 2009). The source/sink distributions throughout the canopy were computed according to Eq. (1):

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$$\vec{C} - C_{Ref} = \overleftarrow{D} \cdot \vec{S}$$
 (1)

where \vec{C} is the concentration (g m⁻³) vector for the 6 levels, C_{Ref} is the concentration (g m⁻³) 189 at reference height (40 m), \overrightarrow{D} (m) is a dispersion matrix, and \overrightarrow{S} (mg m⁻² h⁻¹layer⁻¹) is the 190 resulting source/sink vector. \vec{D} is expressed as a function of Lagrangian timescale and 191 profiles of the standard deviation of the vertical wind speed (σ_w), which was normalized to 192 friction velocity (u*). Integration over all source and sink terms (\vec{S}) yielded the canopy 193 scale isoprenoid flux (mg m⁻² h⁻¹). To parameterize \vec{D} , we use the Lagrangian timescale (TI) 194 parameterized according to Raupach (1989) and the vertical profile of the standard 195 deviation of the vertical wind speed scaled to measured friction velocity. The normalized 196 turbulence profile was taken from turbulence measurements inside and above the canopy at 197 this site recorded as part of AMAZE-08 (Amazonian Aerosol Characterization Experiment 198 199 2008) (Karl et al., 2009). The friction velocity was averaged for each season using daytime data (10:00-14:00, LT) measured at a tower (K34 tower - 2° 36' 32.67" S, 60° 12' 33.48" 200 W) that was 2 km away from the tower where isoprenoid profiles were measured (TT34 201 tower). The calculation of \overleftarrow{D} was based on the far- and near-field approach described by 202 Raupach (1989). As some model inputs (i.e., σ_w/u^*) were obtained during the wet season 203 204 at the TT34 tower in 2008 (Karl et al., 2009), changes in canopy structure between the two 205 studies could potentially affect the results of this study. However, previous work carried out at the K34 tower showed that u* along with other averaged turbulence data have quite 206 similar daytime values in both wet and dry seasons (Ahlm et al., 2010; Araujo et al., 2002). 207

208 Once fluxes from the isoprenoid vertical profiles were obtained by the ILT, they 209 were compared with the isoprenoid fluxes estimated by the Model of Emissions of Gases and Aerosols from Nature (MEGAN 2.1). Isoprenoid emissions estimated by MEGAN 2.1 210 are based on a simple mechanistic model that takes into account the main processes driving 211 variations in emissions (Guenther et al., 2012). As described by Guenther et al., (2012), the 212 activity factor for isoprene, monoterpenes and sesquiterpenes (γ_i) considers the emission 213 response to light (γ_P), temperature (γ_T), leaf age (γ_A), soil moisture (γ_{SM}), leaf area index 214 (LAI) and CO₂ inhibition (γ_{CO2}) according to Eq. (2): 215

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$$\gamma_i = C_{CE} LAI \gamma_P \gamma_T \gamma_A \gamma_{SM} \gamma_{CO_2}$$

(2)

where C_{CE} is the canopy environment coefficient. For the present study, the canopy environment model of Guenther et al. (2006) was used. It has a C_{CE} of 0.57. MEGAN 2.1 was run with variation in light and temperature and LAI. Leaf age of the foliage was estimated by the model based on changes in LAI. Soil moisture and CO₂ inhibition activity factors were assigned a value $\gamma_{SM} = 1$ and $\gamma_{CO2} = 1$, respectively, which assumes no variation in these parameters. More details about the model settings can be obtained in Guenther et al. (2012).

Photosynthetic photon flux density (PPFD) and air temperature for all model runs were obtained from the K34 tower measurement time series (Program of Large Scale Biosphere-Atmosphere – LBA). LAI inputs were obtained by satellite observations from NASA MODIS during August 2010 to January 2011. The level-4 LAI product is composited every 8 days at 1-km resolution on a sinusoidal grid (MODIS-NASA, 2015).

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230 2.4 Uncertainties associated with the ILT and BVOC emission modeling

The main source of errors for applying the ILT is related to the parameterization of 231 two combined effects: (1) vertical diffusion coefficient which is based on measured 232 $\sigma(w)/u^*$ profiles, and (2) Langrangian dispersion time scale (Tl). Moreover, some 233 uncertainties may be due to systematic error sources with respect to (3) chemical losses, 234 and (4) the number of source layers. The entire parameterization of combined effect (1) and 235 (2) was tested using data from an earlier study (Karl et al., 2009, Karl et al., 2010), where a 236 237 comparison with eddy covariance measurements was available. Taking the above conservative error assessment, the combined (effect 1 and 2) uncertainty is +/- 30%. 238

To account for chemistry (effect 3) we used a simple modification of the diffusion 239 240 coefficient based on Hamba (1993), relying on the fact that the chemical loss will mainly influence the far field of the parameterization. Based on estimated OH and measured O₃ 241 densities (Karl et al. 2009, Karl et al., 2010) calculated VOC fluxes were corrected 242 accordingly. Due to low OH and O_3 densities in the canopy (<5 x 10⁵ molecules cm⁻³ for 243 OH and <10 ppbv for O₃) the chemical lifetime for isoprene and monoterpenes is 244 considered large compared to the mixing timescale, leading to a chemistry correction on the 245 order of <5% for isoprene and monoterpenes. This systematic error is included, but relies 246 on an estimation of OH for isoprene. The overall uncertainty for isoprene is calculated as 247 0.3 - 4 % by varying in-canopy OH densities between 5 x 10⁵ and 5 x 10⁶ molecules cm⁻³. 248 It is noted that an in-canopy OH density of 5 x 10^6 molecules cm⁻³ is extremely unrealistic 249 in such a dense canopy and only serves as a very conservative upper limit. Those 250 251 assumptions were also considered for sesquiterpene flux estimates. However, a sensitivity test was carried out to show if the increasing ozone concentrations during the dry season 252 253 could effectively affect sesquiterpene lifetime and then sesquiterpene flux estimates. For this test, sesquiterpene lifetime was changed in the ILT model using a range from 2 min to 254 255 8 hours (upper limit used for isoprene and monoterpene flux estimates). The lower limit (2 min) is based on the lifetime calculated for β -caryophyllene when it is exposed to 24-h 256 average of 7 x 10^{11} molecules cm⁻³ of ozone (~30 ppb) (Atkinson and Arey, 2003). If all 257 sesquiterpenes that occur in this site have similar reactivity with ozone as β -caryophyllene, 258 the overall uncertainty for sesquiterpene flux estimates is calculated as up to 20% by 259 varying sesquiterpene lifetime from 8 h to 2 min. It is noted that when considering a 260 261 lifetime range from 8h to 10 min, the uncertainty for sesquiterpene flux estimates is calculated as up to 4%. The 20 % of uncertainty may be important only during the dry 262 season, when ozone mixing ratios can eventually reach 30 ppbv above canopy (40 m) 263 around noontime. 264

We have also investigated the effect of (4) - the number of source layers. If the number of selected source layers is too small, systematic errors of the calculated integrated fluxes arise. We have investigated this effect and found that in the present case, six source layers are sufficient to capture >90% of the flux. In the present setup, the ILT model does not converge for more than nine layers and the numerical solution becomes unstable. If the ILT model was initiated to only calculate two source layers, the integrated flux would be
underestimated significantly (e.g. by up 50%). With 6 source layers we estimate a
systematic error of <10% due to this effect. The combined effect of the systematic errors
(3) and (4) is estimated to be 5-6%.

Random errors of the ILT parameterization for effects (1) and (2) mostly relate to precision. Systematic errors (3) and (4) mostly relate to accuracy of the parameterization. While there could also be combined effects of random and small systematic errors, that are difficult to assess, we chose an overall conservative error estimate that should reflect precision and accuracy for effects (1) and (2), noting that the 30% should mostly relate to precision. All the uncertainties are one standard error.

280 With respect to uncertainties in model estimates, one of the first quantitative estimates of biogenic VOC emissions (Lamb et al., 1987) included an estimate of 281 uncertainty of 210% based on the propagation of uncertainties in emission factors, emission 282 283 algorithms, amount of biomass, and land use distributions. This "factor of three" 284 uncertainty has continued to be used as a rough assessment of the uncertainty of biogenic VOC emission model estimates applied on regional scales. A more recent study (Hanna et 285 al., 2005) attempted a comprehensive assessment of each model component and concluded 286 that the 95% confidence range on the calculated uncertainty in isoprene emission was about 287 one order of magnitude while the calculated uncertainty for monoterpenes and other VOC 288 was only \pm 20%. Guenther (2013) suggests that the Hanna et al. (2005) study assigns 289 isoprene a higher uncertainty only because more is known about isoprene, and so there are 290 more parameters, and that the lack of observations for quantifying the uncertainties 291 associated with individual model parameters limits the usefulness of this uncertainty 292 293 estimation approach and instead recommends evaluations that consider the results of model comparisons with canopy scale observations. These studies indicate that models tend to 294 agree with observations within $\sim 30\%$ for canopy scale studies with site specific parameters 295 (Lamb et al., 1996) or for regional scale estimates with known land cover (Misztal et al., 296 297 2014) and differ by as much as a factor of two or more for other regional scale studies 298 (Muller et al., 2008; Warneke et al., 2010).

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300 2.5 Canopy light penetration and leaf phenology

The standard canopy environment model of MEGAN 2.1 was used to model light penetration into the canopy (Guenther et al., 2006). Model inputs included the abovecanopy PAR measured (every 30 min) at 50 m on the K34 tower for the whole period of isoprenoid measurements as well as the estimated surface area density of the canopy (m^2 m^{-3}), with measurements carried out in March 2004 using a Light Detection and Ranging sensor (LIDAR) in a transect on the same plateau area of this study (Parker and Fitzjarrald, 2004).

The light penetration was modeled for five canopy layers distributed from the 308 309 canopy top to the ground surface. The thickness of each of the five layers was determined 310 based on the canopy surface area density estimated for every 50 cm from the ground surface to the top canopy (Parker and Fitzjarrald, 2004). The layers were distributed 311 according to a Gaussian curve fit to the canopy surface area densities (from 0.5 m to 48 m). 312 Light absorption was calculated as the difference in the model estimate of downward light 313 at the top and bottom canopy levels. This light absorption corresponded to light that passed 314 through the canopy vertically. Reflectance and scattering were not considered. 315

Leaf phenology was estimated based on the observation of leaf flushing events of 316 the upper crown surfaces of 63 living trees around the K34 tower (~ 2 km far of TT34 317 tower). For this approach, it is assumed that the leaf phenology of the upper crown surfaces 318 of trees around both towers is similar. For the monitoring, a system of data acquisition and 319 storage, based on a Stardot (model Netcam XL 3MP) camera with a 1024 x 768 resolution 320 CMOS sensor, was installed at K34 tower, at 15-20 m above the canopy. The camera 321 viewing angle was south azimuth, perpendicular to the solar transit, centered on 32° of 322 323 depression and pointing out to an area of plateau. Images were logged every 15 s to a passively cooled FitPC2i with heat-tolerant SSD drive. The whole system of data 324 acquisition automatically rebooted after power outages. The images obtained by the camera 325 covered approximately 66° horizontally and 57° vertically, fitting the forest canopy without 326 327 including any area of sky in the image. The most distant trees in the image were located 328 150 m from the camera. The framework was fixed by monitoring the same 63 treetops over four months of observation (October 2010 – January 2011). The analysis of images was 329

based on the number of treetops that showed leaf flushing within one month. For this, one

image was selected at every six days, and then grouped for each month of this study.

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333 2.6 Satellite-derived isoprene emission estimates

334 Top-down isoprene emission estimates over the 0.5 degree region around TT34 tower were obtained by using a grid-based source inversion scheme (Stavrakou et al., 2009) 335 336 constrained by formaldehyde (HCHO) columns. HCHO is an intermediate product of the isoprene degradation process (e.g. Stavrakou et al., 2014). It is measured by UV-visible 337 sensors, such as on the Global Ozone Monitoring Experiment (GOME-2)/MetOp satellite 338 launched in 2006. The source inversion was performed using the global chemistry-transport 339 model IMAGESv2 (Intermediate Model of Annual and Global Evolution of Species) run at 340 a resolution of $2^{\circ} \times 2.5^{\circ}$ and 40 vertical levels from the surface to the lower stratosphere 341 (Stavrakou et al., 2014, 2015). The priori isoprene emission inventory is taken from 342 MEGAN-MOHYCAN-v2 (Stavrakou al., 2014. 343 et http://tropo.aeronomie.be/models/isoprene.htm), and includes updates regarding isoprene 344 emission rates from Asian tropical forests. IMAGESv2 uses HCHO columns retrieved from 345 GOME-2 sensor as top-down constraints and estimates the posterior biogenic isoprene 346 emission on the global scale. Note that given the early morning (9:30) overpass time of the 347 GOME-2 measurement, and the mostly delayed production of formaldehyde from isoprene 348 349 oxidation, the top-down emission estimate is dependent on the ability of MEGAN to simulate the diurnal shape of isoprene emission and on the parameterization of chemical 350 351 and physical processes affecting isoprene and its degradation products in IMAGESv2. For this study, we use daily (24 hours) mean satellite-derived isoprene emissions derived from 352 353 January 2010 to January 2011. More details can be found in Stavrakou et al. (2009, 2014, 2015) and Bauwens et al. (2013). 354

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356 **3. Results and Discussion**

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3.1 Diurnal variation of isoprenoid mixing ratios

Vertical profiles of isoprenoids were analyzed for daytime and nighttime for all the seasons considered in this study. Isoprene (Fig. 2 a, b, c) and total monoterpenes (Fig. 2 d, e, f) had higher mixing ratios during daytime (10:00-16:00, LT) than during nighttime (22:00-04:00, LT) for all seasons, supporting the findings that emissions of isoprene (Alves et al., 2014; Harley et al., 2004) and monoterpenes (Bracho-Nunez et al., 2013; Kuhn et al., 2002, 2004a; Jardine et al., 2015) from Amazonian plant species, at least at this site, are primarily light-dependent and stimulated by increasing temperature.

During daytime, isoprene had a maximum mixing ratio within the canopy. By 365 comparison, at nighttime maximum values occurred above the canopy, and the vertical 366 profiles were similar to those of nighttime air temperature (Fig. 2 j, k, l). As isoprene is not 367 emitted at night, this maximum nighttime abundance of isoprene above the canopy may be 368 due to the daytime residual layer concentrations. In addition, isoprene lifetime increases 369 370 during nighttime owing to the decrease of OH (hydroxyl radical) concentrations in the dark (Goldan et al., 1995) in light of the low concentrations of nitrogen oxides (NO_x) in 371 Amazonia (\leq 3 ppb above the canopy during nighttime in the dry-to-wet transition season) 372 373 (Andreae et al., 2002). Similar results found at another site in central Amazonia suggested that low isoprene concentrations near the ground after sunset could be due to deposition 374 onto and consumption by surfaces (Yáñez-Serrano et al., 2015). Isoprene up-take in the soil 375 has been suggested previously in central Amazonia (Silva, 2010), possibly because of 376 isoprene microbial consumption (Cleveland and Yavitt, 1997; Gray et al., 2014). As with 377 isoprene, higher mixing ratios of total monoterpenes were observed during daytime, 378 379 indicating that they are light-dependent, which agrees with the evidence of recent 380 photosynthetic origin of monoterpenes (Jardine et al., 2015; Loreto et al., 1996).

The vertical profile of total sesquiterpene mixing ratios differed from that of isoprene and total monoterpenes for all seasons. Total sesquiterpenes had higher mixing ratios near the ground and at the sub-canopy level (17 m) than above the canopy (Fig. 2 g, h, i) (P < 0.05). Daytime and nighttime vertical profiles had similar shape, but total sesquiterpene mixing ratios were higher during the nighttime. Even though sesquiterpene emissions for some plant species are both light- and temperature-dependent (Duhl et al.,

2008), results reported here indicate that sesquiterpene emissions are not strongly light-387 dependent in this site, suggesting that their daily variation is driven primarily by 388 temperature. Since some studies have shown that sesquiterpenes are found in the essential 389 oil stored in Amazonian forest trees (e.g. Lima et al., 2005), emissions from these storage 390 structures would not be expected to be light-dependent. In contrast, the monoterpenes, 391 while also present in Amazonian tree essential oil (e.g. Fidelis et al., 2012; Lima et al., 392 2005), appear to be dominated by emissions that occur with no storage (e.g. Loreto et al., 393 1996; Jardine et al., 2015), similar to isoprene emission processes. Another reason for the 394 higher total sesquiterpene mixing ratios at nighttime might be because of the reduction of 395 oxidative reactions owing to the decrease of OH concentrations in the dark (Goldan et al., 396 397 1995) and low concentrations of nitrogen oxides (NO_x) (Andreae et al., 2002), ozone, and nitrate (NO₃) in Amazonia (Martin et al., 2010). In addition, ozonolysis of sesquiterpenes 398 399 during daytime can reduce ambient sesquiterpene concentrations, as previously reported for a subset of these data (Jardine et al., 2011). With daytime ozone mixing ratios up to 40 400 401 ppbv (40 m) during the dry season, sesquiterpene lifetime with respect to ozonolysis above the canopy (40 m) can be 2 min during the daytime and 5 min during the nighttime (Jardine 402 403 et al., 2011). Additionally, sesquiterpene concentrations can build up near the surface, because during nighttime the storage in the forest dominates (80-90%) and is significantly 404 405 larger than the turbulent flux (Karl et al., 2004).

406

407 **3.2** Seasonal variation on isoprenoid mixing ratios and emissions

Vertical profiles of isoprene had higher mean mixing ratios in the dry season, 408 followed by the dry-to-wet transition season and wet season (top panel of Fig. 3 a). The 409 reduction of isoprene mixing ratios from the dry season to dry-to-wet transition season was 410 411 up to 20% and from dry season to wet season was up to 65%. During the dry season, the higher mixing ratios and emissions of isoprene have been attributed to the higher insolation 412 and higher temperatures compared to the wet season and, for this reason, higher isoprene 413 concentrations at the top of the canopy are expected. Nevertheless, in contrast to the 414 415 observations of Yañez-Serrano et al. (2015), who reported maximum daytime mixing ratios of isoprene at the top of the canopy for both dry and wet seasons, this study showed the 416

highest isoprene mixing ratios inside the canopy (11 m) during the dry season, with this
maximum moving to the upper canopy during the dry-to-wet transition season (24 m).

Isoprene emissions inferred from concentration vertical profiles were estimated to 419 be highest in the sub-canopy (16 m) during the dry season and in the upper canopy (28 m) 420 during the dry-to-wet transition season and the wet season (Fig. 4 a). Even though there 421 were differences in which layer was the highest emitter of isoprene within the canopy, 422 mean isoprene emissions into the atmosphere were about the same for the dry season and 423 the dry-to-wet transition season $(1.37 \pm 0.7 \text{ mg m}^{-2} \text{ h}^{-1} \text{and } 1.41 \pm 0.1 \text{ mg m}^{-2} \text{ h}^{-1}$. 424 respectively). Both of these seasons had higher isoprene emissions than during the wet 425 season $(0.52 \pm 0.1 \text{ mg m}^{-2} \text{ h}^{-1})$ (Fig. 4 b). 426

427 The maximum absorption of PPFD by canopy, calculated based on PPFD penetration profile modeled by the standard MEGAN 2.1 canopy environment model, 428 occurred right above the maximum of estimated surface area density of the canopy, with 429 430 the absorption of PPFD being higher during the dry season, followed by the wet season and 431 the dry-to-wet transition season (Fig. 3 b). This maximum PPFD absorption at the upper canopy agreed with the maximum of isoprene mixing ratios (top panel of Fig. 3 a) and 432 emissions (Fig. 4 a) during the dry-to-wet transition season. It differed, however, when 433 434 compared to peaks of isoprene mixing ratios and emissions during the dry season and the wet season. 435

436 One reason for this difference could be the isoprene oxidation in the atmosphere and within plant, especially at the top of the canopy. During the dry season the ratio of methyl 437 vinyl ketone + methacrolein + hydroperoxides (MVK+MAC+ISOPOOH) (Liu et al., 2013) 438 to isoprene was higher compared to the dry-to-wet transition and the wet season (data not 439 440 shown). This higher ratio may indicate an increased oxidative capacity of the atmosphere during the dry season. Moreover, a small source of MVK+MAC+ISOPOOH was observed 441 at the top of the canopy (Jardine et al., 2012). Under conditions of high abiotic stress, as 442 can occur in the dry season, elevated isoprene oxidation rates in plants can be observed and 443 444 isoprene oxidation products might be directly emitted by plants (Jardine et al., 2012).

445 Another important factor might be leaf phenology and/or leaf demography. 446 Different tree species have different isoprene emissions rates, and these rates depend upon

the leaf ontogenetic stage. Isoprene emitters can flush at different canopy levels seasonally, 447 and changes in within-canopy isoprene vertical profiles would be expected as a result. 448 Moreover, as more leaf flushing was observed at the upper canopy during the wet-to-dry 449 transition and early dry season, this caused leaves in the age group of 3-8 months to reach 450 the highest abundance in late dry season and early wet season (Nelson et al., 2014). The 451 period with the high abundance of leaves in this age group is coincident with the period 452 453 when gross ecosystem productivity and landscape-scale photosynthetic capacity is most efficient (Restrepo-Coupe et al., 2013). Here, results show maximum isoprene emission at 454 the upper canopy during the dry-to-wet transition season (Fig. 4 a), which is coincident 455 with the period of high abundance of healthy efficient leaves at the canopy top (Nelson et 456 457 al., 2014) and also coincident with the maximum isoprene emission shown in young mature leaves in the dry-to-wet transition season (Alves et al., 2014). Similarly, higher isoprene 458 459 emissions during the late dry season have also been related to the increase of active biomass in southern Amazonia (Kesselmeier et al., 2002; Kuhn et al., 2004a, 2004b). 460

461 Although the isoprene mixing ratios reported here are within the range of previously reported values in central Amazonia for the dry season and the dry-to-wet transition season 462 (Greenberg and Zimmerman, 1984; Rasmussen and Khalil, 1988; Zimmerman et al., 1988) 463 464 and for the wet season (Yáñez-Serrano et al., 2015), these results are the lowest observed fluxes of isoprene to atmosphere reported for the Amazonia. However, this could be due to 465 features associated with the site of this study, such as the relatively open canopy caused by 466 the proximity to a dirt road and perhaps a relatively low fraction of isoprene emitting 467 species. Isoprene fluxes measured previously at the same tower site during the wet season 468 were similar (Karl et al., 2009). 469

470 Total monoterpenes also showed a strong seasonal variation with maximum mixing 471 ratios during the dry-to-wet season, followed by the dry season and the wet season (middle 472 panel of Fig. 3 a). Taking mixing ratios of the dry-to-wet transition season as a reference, total monoterpene mixing ratios showed an increase of up to 20% from the dry season to 473 474 the dry-to-wet transition season, and a decrease of up to 50% from the dry-to-wet transition season to the wet season. Although total monoterpene mixing ratios were somewhat higher 475 476 in the dry-to-wet transition season than during the dry season, total monoterpene fluxes 477 inferred by the vertical profiles were slightly higher during the dry season $(1.47 \pm 0.06 \text{ mg})$

 $m^{-2} h^{-1}$) compared to the dry-to-wet season $(1.29 \pm 0.2 \text{ mg m}^{-2} h^{-1})$ (Fig. 4 e), indicating that 478 479 the production is higher in the dry season and losses are also higher, leading to lower mixing ratios. In comparison, emissions from these two seasons were considerably higher 480 than during the wet season $(0.36 \pm 0.05 \text{ mg m}^{-2} \text{ h}^{-1})$ (Fig. 4 e). This again indicates that 481 higher insolation and air temperature during the dry season and dry-to-wet transition season 482 compared to the wet season increased the atmospheric concentrations of monoterpenes and, 483 considering the enhanced ozone mixing ratios during the dry season, this may influence the 484 seasonal pattern in monoterpene ozonolysis loss rates (Jardine et al., 2015). These results 485 486 agree with branch level measurements that showed higher monoterpene emissions during the dry-to-wet transition season compared to the wet-to-dry transition season (Kuhn et al., 487 488 2004a). However, results reported here differ from those presented for the southern Amazonia, where monoterpene mixing ratios were higher during the wet season than 489 490 during the dry season (Kesselmeier et al., 2002). Although only a few studies have been carried out with the objective of investigating monoterpene seasonal variations, factors 491 492 other than light and temperature might influence monoterpene emissions from vegetation, including the oxidative capacity of the atmosphere and leaf phenology (Kesselmeier et al., 493 2002; Kuhn et al., 2004a). 494

495 Total monoterpene mixing ratios and fluxes, during the dry season and the dry-towet transition season, were similar to values reported for other sites in central Amazonia 496 (Karl et al., 2007; Yáñez-Serrano et al., 2015). However, the monoterpene comparison of 497 reported studies is a difficult endeavor given that some techniques measured total 498 monoterpenes and others measured some specific monoterpene compounds, and also 499 because monoterpene fragmentation during measurements (PTR-MS) could affect the 500 absolute values of these compounds. Therefore, further efforts are needed in order to 501 characterize the seasonal abundance and the seasonal species-specific composition of 502 monoterpenes in the Amazonia. 503

Average vertical profiles of total sesquiterpene mixing ratios were higher in the dryto-wet transition season, followed by the dry season and the wet season (bottom panel of Fig. 3a). Taking mixing ratios of the dry-to-wet transition season as a reference, total sesquiterpene mixing ratios increased up to 30% from the dry season to the dry-to-wet transition season and decreased by up to 55% from the dry-to-wet transition season to the

wet season. During the dry season and the dry-to-wet transition season, the maximum total 509 510 sesquiterpene mixing ratios were observed near the ground. During the wet season, the maximum mixing ratio was at 17 m (sub-canopy). According to Jardine et al. (2011), 511 during the daytime many sesquiterpenes (46%–61% by mass) are rapidly oxidized by ozone 512 as they undergo within-canopy ozonolysis and contribute to the scarcity of total 513 sesquiterpenes above and near the top of the canopy. Considering that higher insolation and 514 515 also higher ozone concentrations were observed during the dry season (ozone daily average of ~ 23 ppbv and ~ 10 ppbv at 40 m in the dry and wet seasons, respectively), an important 516 fraction of the sesquiterpenes emitted by vegetation could be rapidly oxidized by ozone, 517 518 leading to significantly lower mixing ratios of total sesquiterpene during the dry season 519 (Jardine et al., 2011), which creates a need to account for sesquiterpene oxidation within the canopy when calculating emission rates. 520

Another potential reason for higher mixing ratios of total sesquiterpenes near the 521 ground is that emission could come from surface sources including litter, roots and soil 522 microbes and fungi. Silva (2010) presented surface BVOC emissions at this site, and the 523 524 results suggested that the litter decomposition could be an important source of sesquiterpenes to the atmosphere. Litter production is higher during the dry than during the 525 wet season (Luizão et al., 1989), which could lead to higher amounts of litter at the end of 526 the dry season. Rain starting to increase in the dry-to-wet transition could contribute to 527 528 more decomposition of the litter storage, which can potentially increase sesquiterpene emissions during the processes of decomposition of dead organic matter. Although the 529 ecological functional role of these sesquiterpenes is not known, abiotic emissions from the 530 litter have a specific signature that can be similar to the concentration profile in the green 531 leaf content (Austin et al., 2014) and in sufficient concentration BVOCs can have the 532 533 capacity of attracting and repelling soil organisms to a specific location (Austin et al., 2014). Therefore, higher sesquiterpene emissions from the litter could be a signal to the 534 fauna related to the decomposition process and represent an important step of the 535 biogeochemical cycling. 536

537 In contrast to the mixing ratios, the source-sink distribution analysis made from the 538 vertical profiles of total sesquiterpenes indicated that the main source of these compounds 539 is the canopy (24 m) (Fig. 4 g), and the integration of sources and sinks showed that the

highest total sesquiterpene emission rates going into the atmosphere was during the dry-to-540 wet transition season $(0.77 \pm 0.1 \text{ mg m}^{-2} \text{ h}^{-1})$, followed by the dry season $(0.38 \pm 0.2 \text{ mg m}^{-2})$ 541 h^{-1}), and the wet season (0.34 ± 0.2 mg m⁻² h⁻¹) (Fig. 4 h). However, although Nemitz et al. 542 (2000) have suggested that limitations on the σ_w/u^* parameterization close to the ground do 543 not affect the net flux above the canopy, here we strongly suggest future studies focus on 544 better characterizing the turbulence and oxidation processes at this site, in order to verify 545 the source-sink distribution of sesquiterpenes within the canopy and the emissions from the 546 canopy to atmosphere. This should include speciated sesquiterpene measurements in order 547 to account for their specific reactivity with ozone and other oxidants. 548

Relative emissions can be calculated as emissions normalized to standard conditions 549 of above-canopy PAR of 1500 µmol m⁻² s⁻¹ and temperature of 30 °C. Based only on light, 550 temperature and LAI variation, relative emissions estimated by MEGAN 2.1 were 551 maximum during the dry season for isoprene, α -pinene, and β -caryophyllene (Fig. 4 c, f, i), 552 when the highest light and temperature were observed. This prediction differs from the ILT 553 554 flux estimates (Fig. 4 b, e, h), which showed similar emissions between the dry and the dryto-wet season for isoprene and total monoterpenes and maximum emission during the dry-555 to-wet season for total sesquiterpenes. The overall uncertainties related to ILT flux was 556 557 calculated as \pm 36% and MEGAN estimates are considered to be in agreement with observations when they are within $\sim 30\%$. However, more observation studies are needed in 558 order to evaluate the degree of observation-modeling agreement and to improve model 559 approaches, especially for total monoterpenes and total sesquiterpenes, which could present 560 larger uncertainties due to the lack of information about atmospheric concentrations and 561 reactivity of monoterpene and sesquiterpene chemical species in Amazonia. 562

563 To compare the seasonal variation of isoprenoid emissions with changes in environmental (light and temperature) and biological (LAI) factors in more detail, monthly 564 565 fluxes of isoprenoids were compared to PAR at 51 m, air temperature at 51 m, and LAI (satellite observations - MODIS) (Fig. 5). The highest fluxes of isoprene and total 566 567 monoterpene were observed when PAR was at its highest (October 2010) (Fig. 5 b, c), and when there is high abundance of healthy efficient leaves (Nelson et al., 2014). The 568 569 similarity in the behavior of isoprene and monoterpene emissions is supported by the evidence of the photosynthetic origin of monoterpenes (Jardine et al., 2015; Loreto et al., 570

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571 1996). Interestingly, in September 2010 total monoterpene emissions were higher than 572 isoprene emissions. This could be related to the higher source of monoterpenes in the upper 573 canopy compared to isoprene during this month. When there are more young leaves at the 574 upper canopy during the first half of the dry season (Nelson et al., 2014), high emissions of 575 monoterpenes can be expected. Total sesquiterpene fluxes tracked neither PAR nor air 576 temperature, having the highest emission when PAR and air temperature were decreasing 577 (November 2010) (Fig. 5 d).

578 Predictions from MEGAN 2.1 again differed from measured emissions (Fig. 5 b, c, d), showing a reduction in emissions from September 2010 to January 2011. Major 579 580 quantitative differences between ILT and MEGAN estimates can be shown for isoprene in 581 September, when ILT estimates represented only 4 % of the MEGAN estimates; for total monoterpenes in December, when ILT estimates accounted for 14 % of the MEGAN 582 estimates; and for total sesquiterpenes in November, when ILT estimates were 232% of the 583 584 MEGAN estimates". These differences may be related to local effects, especially leaf 585 phenology and changes in the atmospheric oxidative capacity over the seasons. In order to evaluate the potential effect of leaf phenology on emissions, leaf flushing, PAR, isoprene 586 and total monoterpenes at canopy scale were compared in Fig. 6. They closely tracked each 587 other during the 4 months of measurements. For the period of this study, the analysis of 588 canopy images for every six days from October 2010 to January 2011 showed a decrease in 589 leaf flushing from the end of the dry season to the wet season, which was similar to the 590 decrease of isoprene and total monoterpene emissions and PAR. Results from 28 months 591 (October 2010-January 2013) of canopy imaging have shown that the highest number of 592 treetops with leaf flushing occurred during the wet-to-dry transition season (June-July), 593 594 accounting for 35-50 % of treetops with leaf flushing, followed by a subsequent decrease until the end of the wet season (Tavares, 2013) (Fig. 6). Correspondingly, the results of the 595 present study suggest that lowest emissions might be expected in the June-July time period. 596 These results agree with those presented by Barkley et al. (2009) using remote sensing, 597 598 suggesting that seasonal changes in isoprene emissions may be strongly affected by leaf 599 phenology in the Amazonia.

600 In order to verify if the seasonal trend of the isoprene emissions observed in this 601 study can also be observed in a 0.5° grid cell around TT34 tower, isoprene emissions

estimated based on tower vertical profile concentrations are compared with estimates 602 603 constrained by satellite measurements of HCHO in Fig. 7. The top-down estimates have a seasonal cycle that is similar to the bottom-up approach. Compared to the dry season, 604 fluxes decrease by 40% during the wet and the wet-to-dry transition season from April to 605 July (Stavrakou et al., 2015), in qualitative agreement with the conclusions drawn in 606 Barkley et al. (2009). The inferred dry season isoprene flux is about twice that of the wet-607 to-dry season. It peaks in September and gradually drops from October to January (Fig. 7), 608 as a result of decreasing temperature and solar radiation, affecting the oxidation of isoprene 609 leading to HCHO formation. The ground-based estimates exhibit a much stronger month-610 to-month variation, with flux estimates of 5 times higher in October compared to 611 612 September and December. The small increase of the flux between December and January is not observed by the satellite observations. Despite these differences, partly due to reduced 613 representativeness when comparing local measurements with flux estimates from a 0.5° 614 grid cell, this comparison shows that both large (satellite) and small (ground-based) scales 615 616 agree that there are enhanced isoprene emissions during the dry season followed by a reduction towards the wet season. 617

618 The results reported here are associated with a small footprint area. This together with the huge biodiversity of tropical rainforests makes it impossible to generalize these 619 620 results to the regional scale. Moreover, although some previous reports have suggested significant seasonal variations of BVOCs based on in situ measurements in different sub-621 regions of Amazonia, when those investigations (summarized in Table 1) and this study 622 were compared, high variability is apparent among values of mixing ratios and fluxes. This 623 variability could be due to: (1) different methodologies, (2) sampling in different seasons, 624 (3) sampling in different regions (e.g., south, north, west, eastern Amazonia), (4) sampling 625 in different ecotones of the same region, (5) different statistical analyses, and (6) perhaps 626 due to small data sets that are not statistically significant to characterize emissions of a 627 specific site. 628

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630 **3.3** Comparison with model predictions of seasonal isoprenoid emissions in Amazonia

Although the canopy scale isoprenoid emission measurements presented herediffered from those modeled by MEGAN 2.1 (Figs. 4 and 5), which assume that variations

are driven primarily by light, temperature and leaf area, in terms of seasonal variation, 633 634 MEGAN 2.1 estimates of isoprene emission agreed fairly well with the satellite-derived isoprene emission, which suggests that other factors at this site could influence isoprene 635 emissions locally. As already mentioned, leaf phenology may cause important effects on 636 local emissions. As MEGAN 2.1 was driven with local variations in PAR and air 637 temperature, and with regional variations of LAI (satellite observations at 1-kilometer 638 resolution), this regional variation in LAI may not represent the local effect of LAI 639 variation on local emissions, since vegetation in Amazonia is phenologically distinct due to 640 the huge biodiversity of this ecosystem (Silva et al., 2013). Furthermore, as the canopy 641 structure might vary seasonally due to leaf phenology/demography, the pattern of light 642 643 penetration/absorption and then leaf temperature may change as well; thus, this, together with the differences in emissions among species and among leaf ontogenetic stages, could 644 have an important impact on seasonal changes of local emissions. 645

646 Besides the effects of light, temperature and leaf phenology/demography, some 647 efforts have been made to include effects of CO₂ variation (Arneth et al., 2007; Guenther et al., 2012) as well as the link between photosynthesis and emission (Grote et al., 2014; 648 Morfopoulos et al., 2013, 2014; Unger et al., 2013) into isoprene emission models at 649 regional and global scales. However, the current regional and global BVOC emission 650 models predict much smaller seasonal variations (Guenther et al., 2006, 2012; Muller et al., 651 2008; Unger et al., 2013) compared to the measurements in Amazonia (Table 1). 652 Furthermore, satellite observations indicate that the current understanding of the processes 653 controlling seasonal variations is insufficient, and models do not simulate the unexpected 654 shutdown of isoprene emission in the Amazonia during the wet-to-dry transition season 655 (Barkley et al., 2009). 656

Many recently published studies have used the MEGAN model and the majority have focused on improving our understanding of isoprene emissions. Although other models have been developed on the basis of known biochemical processes (Grote et al., 2014; Morfopoulos et al., 2014; Unger et al., 2013), the general framework and processes simulated are similar. The biochemical basis of isoprene production and release must be further understood to develop mechanistic explanations for variation in isoprene emission 663 (Monson et al., 2012), which may reduce uncertainties associated with the responses to 664 environmental factors.

Seasonal variation of isoprene emissions might be explained by the change in 665 energy supply from photosynthesis throughout the seasons (e.g. Grote et al., 2014). This is 666 supported by the generally strong correlation between isoprene emission and gross 667 photosynthetic capacity reported for Amazonian tree species (Kuhn et al., 2004b), and by 668 the fact that higher demography of healthy efficient leaves (Nelson et al., 2014) coincides 669 with the period of most efficient landscape-scale photosynthesis and photosynthetic 670 capacity (Restrepo-Coupe et al., 2013). However, more measurements are needed to 671 examine this relationship which should follow PAR variation. Additionally, since canopy 672 673 structure may explain some variation in biomass growth over tropical landscapes due to differences in the pattern of light penetration and absorption by the canopies (Stark et al., 674 675 2012), measurements of canopy structure may also help to explain some of the differences in isoprenoid emissions among the Amazonian sub-regions. 676

677 Therefore, at least for the Amazonian rainforest, models currently do not fully capture seasonal variations in isoprenoid emissions, especially for monoterpenes and 678 679 sesquiterpenes, which are less investigated compared to isoprene. The scarcity of measurements in Amazonia prevents the development and evaluation of accurate model 680 681 approaches. Thus, this study strongly encourages future *in situ* measurements in Amazonia, including at leaf level, in order to verify changes driven by seasonal variations in leaf area, 682 leaf age, phenology and emission response to soil moisture, and the short-term and long-683 term temperature and light environment. 684

685

686 4. Summary and conclusions

In this study, we present the first in situ measurements that show a seasonal trend in isoprenoid emissions for a primary rainforest of central Amazonia. Isoprenoid emissions peak at the end of the dry season and at the dry-to-wet transition season. Under conditions of high insolation and high temperatures joined together with the high demography of photosynthetically efficient leaves (Caldararu et al., 2012; Myneni et al., 2007; Nelson et al., 2014; Samanta et al., 2012), isoprenoid metabolic pathways may experience more favorable conditions for synthesizing these compounds in the dry season and the dry-to-wet transition season. This is especially for the case of isoprene and monoterpenes, which are
light- and temperature-dependent and are affected by the recent production of
photosynthetic substrates.

697 Although some studies have suggested that there are no seasonal variations in canopy structure and greenness in Amazonia (e.g. Morton et al., 2014), results reported 698 here present a seasonal variation of leaf flushing and suggest maximum leaf demography in 699 700 the late dry season, which generally agrees with the assumption that a greenup during the 701 dry season in Amazonia may drive increasing isoprene emissions as suggested by satellite retrievals (Barkley et al., 2009). Moreover, this study also suggests that seasonal changes in 702 703 the atmospheric oxidative capacity could have an important impact on the seasonality of at 704 least some isoprenoid concentrations and above canopy emissions, especially for sesquiterpenes. Their quantification is challenged by rapid atmospheric chemical reactions 705 catalyzed by high insolation and higher ozone concentrations in the dry season. 706

707 MEGAN 2.1 estimates did not fully capture the behavior observed with the 708 isoprenoid emissions based on in-situ PTR-MS measurements (inverse Lagrangian transport model). Model emissions of isoprene and total monoterpenes were overestimated, 709 especially during September 2010 (dry season) and December 2010 (wet season), 710 711 respectively. Total sesquiterpenes were underestimated during November 2010 (dry-to-wet transition season). This difference between MEGAN 2.1 flux estimates and fluxes 712 estimated by the PTR-MS vertical mixing ratio profiles could be due to experimental errors 713 or the influence of very local effects on the seasonal emissions measured in this site, 714 because satellite-derived isoprene emissions agree fairly well with MEGAN 2.1 emission 715 estimates and the ground observations do not agree with the satellite data or the model, 716 717 principally in September. Perhaps the isoprene pattern observed at the site is due to a very 718 local effect of leaf flushing by isoprene emitting species around this tower, but this is not seen on the regional scale where there are different species distributions. 719

Generally, current models assume that seasonal variation of BVOC emissions in the Amazonian rainforest are primarily based on light and temperature variations. These model simulations capture only a part of the actual variation and have uncertainties associated with the insufficient understanding of mechanistic processes involved in the seasonality of these compounds. Nevertheless, because the number of measurements and sites is limited in

Amazonia, there is a scarcity of information, which hinders further model improvements. In 725 726 summary, our results demonstrate strong seasonality and suggest that important processes are taking place during the transition seasons. Also, results reveal the need for long-term 727 and continuous BVOC observations from leaf level to ecosystem level, and also suggest 728 that standardized measurement procedures are required in order to compare the different 729 Amazonian sub-regions. This may advance understanding of the seasonality of BVOC 730 exchanges between forest and atmosphere, providing the information needed to improve 731 BVOC emission estimates for climate and air quality modelling studies. 732

733

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Table1: Isoprene and monoterpenes from different regions in the Amazonian rainforest: comparison of estimates and direct measurements of mixing ratios and fluxes.

študy	Site	Technical approach	isoprene (ppbv)	isoprene (mg m ⁻² h ⁻¹)	sum of su Mt [†] (ppbv) (m	m of Mt [†] 1g m ⁻² h ⁻¹)	season	comments
Central Amazonia				. 8 /		0 /		
Greenberg and Zimmerman, 1984	Manaus/Humaitá- Amazonas, Brazil	GC-FID, canister samples (near ground to 30m)	2.40 (1-5.24) ^a		2.86		Dry (Aug-Sep 1980)	mean - daytime range i not reported
Greenberg and Zimmerman, 1984	Manaus/Humaitá- Amazonas, Brazil	GC-FID, canister samples (flights from treetop to 2 km)	2.27 (0.38-4.08) ^a		5.47		Dry (Aug-Sep 1980)	mean - daytime range i not reported
Greenberg and Zimmerman, 1984	Manaus/Humaitá- Amazonas, Brazil	GC-FID canister samples (flights from 2km to Tropopause)	0.19 (0.14-0.22) ^a		1.91		Dry (Aug-Sep 1980)	mean - daytime range i not reported
acob and Wofsy, 1988*	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas, Brazil	Inverse modeling approach using Zimmerman et al. 1988 data		1.58			Dry (July-Aug 1985)	mean average of 24 hour
Zimmerman et al., 1988*	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas Brazil	GC-FID, teflon bag on tethered balloon (30m)	2.65 [1.39-3.38] ^b		0.27 [0.15-0.54] ^b		Dry (July-Aug 1985)	median and interquartil range (24h)
Zimmerman et al., 1988*	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas Brazil	GC-FID, teflon bag on tethered balloon (305m)	1.73 [1.03-2.15] ^b		0.15 [0.04-0.33] ^b		Dry (July-Aug 1985)	median and interquartil range (24h)
Zimmerman et al., 1988*	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas, Brazil	GC-FID, teflon bag on tethered balloon (up to 305m)		3.1		0.23	Dry (July-Aug 1985)	mean daytime (08:00 16:00, LT)
Rasmussen and Khalil, 1988	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas, Brazil	GC-FID, canister samples (near ground level)	2.77 (±0.4)				Dry (July-Aug 1985)	mean daytime (11:00 15:00, LT)
Rasmussen and Khalil, 988	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas, Brazil	GC-FID, canister samples (aircraft flights from 150m to 5000m)	1.5 (±0.75)				Dry (July-Aug 1985)	daytime
Davis et al., 1994*	ABLE - Adolfo Ducke Forest Reserve - Manaus- Amazonas, Brazil	Mixed Layer Gradient approach using Zimmerman et al. 1988 data		3.63 (±1.4)			Dry (July-Aug 1985)	mean daytime (08:00 18:00, LT)
Cesselmeier et al., 2000	Balbina - ~100 km north of Manaus-Amazonas,	GC-MS, cartridge samples (outside forest)	6.55 (±1.26)		0.63 (±0.19)		Wet (Apr 1988)	mean daytime (09:30 15:00, LT)
	Brazil	GC-MS, cartridge samples (inside Canopy)	3.55 (±0.07)		0.24 (±0.04)		Wet (Apr 1988)	mean daytime (09:30 15:00, LT)
		GC-MS, cartridge on tethered balloon (200- 500m)	~3		~0.2		Wet (Apr 1988)	mean of 24h

Cont. **Table1:** Isoprene and monoterpenes from different regions in the Amazonian rainforest: comparison of estimates and direct measurements of mixing ratios and fluxes.

študy	Site	Technical approach	isoprene (ppbv)	isoprene (mg m ⁻² h ⁻¹)	sum of Mt [†] (ppbv)	sum of Mt [†] (mg m ⁻² h ⁻¹)	season	comments
Central Amazonia								
Cesselmeier et al., 2000	Cuieiras Biological Reserve (C14-ZF2) - Manaus-Amazonas, Brazil	GC-MS, cartridge samples (inside and above canopy)	6.7 ±1.07		0.73 ±0.24		Wet (Apr 1988)	daytime
Stefani et al. 2000	Cuieiras Biological Reserve (K34-ZF2) - Manaus-Amazonas, Brazil	GC-MS, cartridge on Relaxed Eddy Accumulation (~53m)		3.6 -5.4		0.72 - 0.9	Aug 1999 and Jan 2000	range of daytime averag normalized fluxes for th whole period of measurement
Andreae et al., 2002	Cuieiras Biological Reserve (K34-ZF2) - Manaus Amazonas Brazil	GC-MS, cartridge on Relaxed Eddy		2.88		0.36	Dry-Wet (Nov 1999-Jan 2000)	midday values
Ciccioli et al., 2003	Cuieiras Biological Reserve (K34-ZF2) - Manaus-Amazonas, Brazil	GC-MS, cartridge on Relaxed Eddy Accumulation (~51m)		5.11 max.		1.36 max.	Dry (July 2001)	midday values
Greenberg et al., 2004	Balbina - ~100 km north of Manaus-Amazonas, Brazil	GC-MS, cartridge on tethered balloon (200- 1000m)	2.86 [2.25-3.64] ^b		0.21 [0.17-0.31] ^b		Wet (March 1998)	median and interquartiles daytime (12:00-15:00, LT)
Greenberg et al., 2004	Balbina - ~100 km north of Manaus-Amazonas, Brazil	Box model		5.3		0.23	Wet (March 1998)	maximum midday emission fluxes estimated for the ecoregion
Carl et al., 2007 ^{††}	Cuieiras Biological Reserve (C14-ZF2) - Manaus-Amazonas, Brazil	PTR-MS, Disjunct Eddy Covariance (~ 54 m)	7.8 ±3.7	8.3 ±3.1	0.87 ±0.3	1.7 ±1.3	Dry (Sep 2004)	mean daytime (12:00-14:00 LT)
⟨arl et al., 2007 ^{††}	Cuieiras Biological Reserve (C14-ZF2) - Manaus-Amazonas Brazil	PTR-MS, Mixed Layer Gradient (up to ~1200 m)	5.5 ±2.6	12.1 ±4.0	0.52 ±0.2	3.5 ±1.2	Dry (Sep 2004)	mean daytime (10:00-11:30 LT)
Cuhn et al., 2007**	Cuieiras Biological Reserve (K34-ZF2)- Manaus-Amazonas, Brazil	GC-FID, cartridge on Relaxed Eddy Accumulation (~51m)		2.4 ±1.8 (max. 6.1)		0.44 ±0.49 (max. 1.9)	Dry (July 2001)	mean daytime (10:00-15:00 LT)
Suhn et al., 2007**	Cuieiras Biological Reserve (K34-ZF2) - Manaus-Amazonas, Brazil	GC-FID, cartridge on Surface Layer Gradient (28, 35,5, 42,5, 51m)		3.9 ±4.1 (max. 12.8)		0.43 ±0.65 (max. 2.1)	Dry (July 2001)	mean daytime (10:00-15:00 LT)
Suhn et al., 2007**	Cuieiras Biological Reserve (K34-ZF2) - Manaus-Amazonas, Brazil	GC-FID, cartridge samples, Mixed Layer Gradient (50-3000m)		4.2 ±5.9 (max. 15.7)			Dry (July 2001)	mean daytime (10:00-18:00 LT)
Carl et al., 2009	Cuieiras Biological Reserve (TT34-ZF2) - Manaus-Amazonas Brazil	PTR-MS, Gradient flux (2, 10.9, 16.7, 23.9, 30.3 and 39.8 m)		0.7 ±0.2			Wet (Feb 2008)	mean daytime (11:00-17:00 LT); flux at 35 m
Rizzo et al., 2010 ^{††}	Cuieiras Biological Reserve (C14-ZF2)- Manaus-Amazonas,	PTR-MS, Disjunct Eddy Covariance (54 m)	7.8	8.4	0.29	0.93	Dry (Sep 2004) Dry (Sep 2004)	max. at early afternoon max. at noon

Cont. **Table1:** Isoprene and monoterpenes from different regions in the Amazonian rainforest: comparison of estimates and direct measurements of mixing ratios and fluxes.

študy	Site	Technical approach	isoprene (ppbv)	isoprene (mg m ⁻² h ⁻¹	sum of) Mt [†] (ppbv)	sum of Mt (mg m ⁻² h ⁻¹	season	comments
Central Amazonia								
Silva, 2010	Cuieiras Biological Reserve (K34-ZF2)-	GC-MSFID cartridge samples at 1m	3.2 ±0.9		0.28 ±0.13		Wet (May 2009)	mean daytime (07:00-17:00 LT)
	Manaus-Amazonas, Brazil	GC-MSFID cartridge samples at 10m	4.6 ± 0.94		1.09 ± 0.35		Wet (May 2009)	mean daytime (07:00-17:00 LT)
		GC-MSFID cartridge samples at 20m	6.17 ± 1.03		0.75 ± 0.17		Wet (May 2009)	mean daytime (07:00-17:00 LT)
ardine et al., 2011 [‡]	Cuieiras Biological Reserve (TT34-ZF2)- Manaus-Amazonas, Brazil	PTR-MS, Gradient profile (2, 11, 17, 24, 30 and 40 m)			~ 0.78		Dry-Wet (Sep- Dec 2010)	mean daytime 10:00-16:00 LT) at 40 m
ardine et al., 2011 [‡]	Cuieiras Biological Reserve (TT34-ZF2)- Manaus-Amazonas, Brazil	PTR-MS, Gradient profile (2, 11, 17, 24, 30 and 40 m)				~ 1.47	Dry-Wet (Sep- Dec 2010)	mean daytime 10:00-16:00 LT) at 35 m
ardine et al., 2012 [‡]	Cuieiras Biological Reserve (TT34-ZF2)- Manaus-Amazonas, Brazil	PTR-MS, Gradient profile (2, 11, 17, 24, 30 and 40 m) and Gradient flux		~1.43			Dry-Wet (Sep- Dec 2010)	mean daytime (10:00-16:00 LT); flux at 40 m
l'áñez-Serrano et al., 2015	ATTO site - Manaus Manaus-Amazonas, Brazil	PTR-MS, Gradient profile (0.05, 0.5, 4, 24, 38, 53 and 79 m)	5.22 ±1.5		0.75 ±0.18		Dry (Sep 2013)	Isoprene, daytime median (12 15:00, LT). Mt, daytime median (15-18:00 LT)
l'áñez-Serrano et al., 2015	ATTO site - Manaus Manaus-Amazonas, Brazil		1.5 ±0.78		< 0.23		Wet (Feb-Mar 2013)	Isoprene, daytime median (12 15:00, LT). Mt, daytime median (15-18:00, LT)
Γhis study [‡]	Cuieiras Biological Reserve (TT34-ZF2) - Manaus-Amazonas Brazil	PTR-MS, Gradient profile (2, 11, 17, 24, 30 and 40 m) and Gradient flux	2.68 ±0.9	1.37 ± 0.7	0.67 ±0.3	1.47 ±0.06	Dry (Sep-Oct 2010)	mean daytime (10:00-14:00, LT) at 40m
Γhis study [‡]	Cuieiras Biological Reserve (TT34-ZF2) -	PTR-MS, Gradient profile (2, 11, 17, 24, 30 and 40 m) and Gradient flux	2.65 ±1.33	1.41 ±0.1	0.85 ±0.4	1.29 ±0.2	DWT [¥] (Nov 2010)	mean daytime (10:00-14:00, LT) at 40m
Гhis study [‡]	Cuieiras Biological Reserve (TT34-ZF2) - Manaus-Amazonas, Brazil	PTR-MS, Gradient flux (2, 11, 17, 24, 30 and 40 m) and Gradient flux	1.66 ±0.9	0.52 ±0.1	0.47 ±0.2	0.36 ± 0.05	Wet (Dec 2010 -Jan 2011)	mean daytime (10:00-14:00, LT) at 40m
Eastern central Amazon	nia							
Rinne et al., 2002	Tapajós National Forest - Santarém-Pará, Brazil	GC-MS cartridge on Disjunct Eddy Accumulation (~ 45m)	5 max.	2.4			Dry (July 2000)	Afternoon values 30 °C and 1000 µmol m ⁻² s ⁻¹
Greenberg et al., 2004	Tapajós National Forest - Santarém-Pará, Brazil	GC-MS, cartridge on tethered balloon (200- 1000m)	0.74 [0.6-1] ^b	(0.08 [0.03-0.06] ^b		Wet (Jan-Feb 2000)	median and interquartiles daytime (12:00-15:00, LT)
Greenberg et al., 2004	Tapajós National Forest - Santarém-Pará, Brazil	Box model		2.2		0.18	Wet (Jan-Feb 2000)	maximum midday emissio fluxes estimated for th ecoregion

Cont. **Table1:** Isoprene and monoterpenes from different regions in the Amazonian rainforest: comparison of estimates and direct measurements of mixing ratios and fluxes.

študy	Site	Technical approach	isoprene (ppbv)	isoprene (mg m ⁻² h ⁻¹)	sum of Mt [†] (ppbv)	sum of Mt [†] (mg m ⁻² h ⁻¹)	season	comments
Eastern central Amazon	ia			· • • /		· • • · ·		
Frostdorf et al., 2004	Tapajós National Forest - Santarém-Pará, Brazil	GC-FID, canister samples (54, 64 m)	1.9 ±1.2; 1.3 ±0.8				Wet (Jan- May 2002)	mean daytime (11:00-14:00 LT)
Frostdorf et al., 2004	Tapajós National Forest - Santarém-Pará, Brazil	GC-FID, canister samples (54, 64 m)	1.4 ±0.5; 1.0 ±0.4				WDT [₩] (June- July 2002)	mean daytime (11:00-14:00 LT)
Frostdorf et al., 2004	Tapajós National Forest - Santarém-Pará, Brazil	GC-FID, canister samples (54, 64 m)	2.8 ±0.9; 2.5 ±0.8				Dry (Aug-Nov 2002)	mean daytime (11:00-14:00 LT)
Western Amazonia		· · ·						
Helmig et al., 1998	Peru - 500 km west of Iquitos	GC-MS, cartridge on tethered balloon (up to 1600 m)	3.31, 1.39, 0.16		0.21, 0.06, 0.015		July 1996	Median daytime (ground mixed layer and above mixed layer)
Helmig et al., 1998	Peru - 500 km west of Iquitos	GC-MS, cartridge samples, Mixed Layer Gradient		7.4		0.42		mean daytime
Helmig et al., 1998	Peru - 500 km west of Iquitos	GC-MS, cartridge samples, Mixed Layer Budget		8.1		0.41		mean daytime
Southern Amazonia								
Kesselmeier et al., 2002	Jaru Biological Reserve, Rondônia, Brazil	GC-FID, cartridge samples (8-52m)	~4		~0.8		WDT [₩] (May 1999)	mean daytime (11:00-18:00 LT)
Kesselmeier et al., 2002	Jaru Biological Reserve, Rondônia, Brazil	GC-FID, cartridge samples (8-52m)	~12		~0.8		DWT [¥] (Sep- Out 1999)	mean daytime (11:00-18:00 LT)
Greenberg et al., 2004	Jaru Biological Reserve, Jaru-Rondônia, Brazil	GC-MS, cartridge on tethered balloon (200- 1000m)	6.89 [2.78- 7.73] ^b		0.83 [0.56-2.65] ^b		Wet (Feb 1999)	median and interquartiles daytime (12:00-15:00, LT)
Greenberg et al., 2004	Jaru Biological Reserve, Jaru-Rondônia, Brazil	Box model		9.8		6.1	Wet (Feb 1999)	maximum midday emission fluxes estimated for the ecoregion
Simon et al., 2005	Jaru Biological Reserve, Rondônia, Brazil	Lagrangian transport sub- model.		~5.9			WDT ^{¥¥} (May 1999)	midday values
Simon et al., 2005	Jaru Biological Reserve, Rondônia, Brazil	Modeling using data of Kesselmeier et al., 2002		~8.2			DWT [¥] (Sep- Out 1999)	midday values

Cont. Table1: Isoprene and monoterpenes from different regions in the Amazonian rainforest: comparison of estimates and direct measurements of mixing ratios and fluxes.

Study	Site	Technical approach	isoprene (ppbv)	isoprene	sum of	sum of Mt^{\dagger}	season	comments	
Southorn Amazonia				(mg m n)	Mt [*] (ppbv)	(mg m n)			
southern Amazonia									
Aquino, 2006	Jaru Biological Reserve,	GC-FID, canister samples	4.5 ±0.9; 4.0 ±1.2				Wet (Feb-May	mean daytime	(11:00-16:00
	Rondônia, Brazil	(50, 60 m)					2002)	LT)	
Aquino, 2006	Jaru Biological Reserve,	GC-FID, canister samples	2.1 ±2.0; 1.8 ±1.8				WDT ^{¥¥} (Jun	mean daytime	(11:00-16:00
1	Rondônia, Brazil	(50, 60 m)					2002)	LT)	
Aquino, 2006	Jaru Biological Reserve,	GC-FID, canister samples	4.6 ±2.7; 4.0 ±2.5				Dry (Jul-Sep	mean daytime	(11:00-16:00
•	Rondônia, Brazil	(50, 60 m)					2002)	LT)	
Aquino, 2006	Jaru Biological Reserve,	GC-FID, canister samples	3.4 ± 1.2 ; 3.0 ± 0.5				DWT [¥] (Out-	mean daytime	(11:00-16:00
-	Rondônia, Brazil	(50, 60 m)					Nov 2002)	LT)	

Note: Seasons follow determination of each study. For some studies the exact times of sample collection are not available and then not reported. Statistics differed among studies. The most of studies showed mean values but others presented median values and/or just a range of all values measured.

†Mt - monoterpenes; ^a- range of variation;

^b - interquartile ranges based on median "[]";
 *, ** studies derived from the same observational data base;
 ^t ^{††} studies derived from part of the same observational data base;

 $^{\text{4}}$ DWT - dry-to-wet transition season;

^{¥¥}WDT - wet-to-dry transition season.



Figure 1: Precipitation, PAR and air temperature measured at K34 tower (~2 km far of TT34 tower); a) relative frequency (%) of monthly cumulative precipitation from 1999 to 2012, b) monthly cumulative precipitation from July 2010 to June 2011 (measured in 30 minute intervals for 24 hours) (bars), and average of monthly cumulative precipitation from 1999 to 2012 (red line); c) relative frequency of monthly PAR from 1999 to 2012 (measured every 30 min during 06:00-18:00, LT), d) monthly average PAR from July 2010 to June 2011 (measured every 30 min during 06:00-18:00, LT); e) relative frequency of monthly air temperature from 1999 to 2012, f) monthly average air temperature from July 2010 to June 2011 (measured in 30 minute intervals for 24 hours). Figures on the right side cover the period of this study; grey areas represent the period of dry season; and blue line at (b) represents 100 mm month⁻¹. Error bars represent one standard deviation.



Figure 2: Daytime (10:00-16:00, LT) and nighttime (22:00-04:00, LT) average vertical profiles of isoprene (a, b, c), total monoterpenes (d, e, f), total sesquiterpenes (g, h, i), and air temperature (j, k, l) of the dry season (DS), the dry-to-wet transition season (DWT) and the wet season (WS). Error bars represent one standard deviation.



Figure 3: Daytime (10:00-16:00, LT) vertical profiles of mixing ratios of isoprene, total monoterpenes and total sesquiterpenes from the dry season to the wet season; and estimated surface area density of the canopy at this study site (ground-based measurements carried out in March/2004 using LIDAR - Light Detection And Ranging) (Parker and Fitzjarrald, 2004) (a). Vertical profile of photosynthetic photon flux density (PPFD) penetration and absorption by the canopy from the dry season to the wet season modeled by MEGAN 2.1 (b). Daytime (10:00-16:00, LT) air temperature profiles from dry season to wet season measured at K34 tower (c). In fig. 1a the top and the bottom x axis represent isoprenoid mixing ratios and estimated surface area density of the canopy, respectively. Error bars represent one standard deviation.



Figure 4: Daytime (10:00-14:00, LT) source-sink distribution inside and above the canopy, cumulative flux estimation, and relative emission modeled by MEGAN 2.1 of isoprene (a, b, c), total monoterpenes (TMt) (d, e, f) and total sesquiterpenes (TSt) (g, h, i) from the dry season to the wet season. Error bars represent one standard deviation.



Figure 5: Monthly averages of air temperature and PAR (measured at K34 tower during 10:00-14:00, LT), and LAI (MODIS, 8-day observations) (a). Monthly averages of fluxes of isoprene (b), total monoterpenes (TMt) (c) and total sesquiterpenes (TSt) (d). Flux based on in-situ PTR-MS measurements (inverse Lagrangian transport model - estimates for 10:00-14:00, LT, at TT34 tower) are represented by solid squares and one standard deviation; fluxes modeled by MEGAN 2.1 (estimates for 10:00-14:00, LT) are shown by solid lines and filled areas that represent one standard deviation. Isoprene flux modeled by MEGAN 2.1 in (b) were divided by five. Error bars represent one standard deviation.



Figure 6: Estimated monthly leaf flushing (light green line) (Tavares, 2013), and monthly average of PAR measured from October 2010 to January 2013 at K34 tower (06:00-18:00, LT) (black line). For the period of this study, leaf flushing is also represented by the analysis of canopy images for every six days from October 2010 to January 2011 (red circles). Monthly averages of fluxes of isoprene (dark green line) and total monoterpenes (blue line) (estimated for 10:00-14:00, LT, at TT34 tower). Grey areas represent the period of the dry season.



Figure 7: Comparison of monthly isoprene emissions based on in-situ PTR-MS measurements (inverse Lagrangian transport model) and satellite-derived estimates and MEGAN 2.1 estimates. Satellite-derived estimates are from January 2010 to January 2011, and ground-based estimates are from September 2010 to January 2011. Satellite-derived and MEGAN 2.1 estimates were divided by 2.5 and 5, respectively. Grey area represents the period of the dry season. Error bars represent one standard deviation.