# 1 Source Attribution of Aerosol Size Distributions and Model

- 2 Evaluation Using Whistler Mountain Measurements and GEOS-
- 3 Chem-TOMAS Simulations
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## 14 Abstract

15 Remote and free tropospheric aerosols represent a large fraction of the climatic influence of 16 aerosols; however, aerosol in these regions is less characterized than those polluted boundary 17 layers. We evaluate aerosol size distributions predicted by the GEOS-Chem-TOMAS global 18 chemical transport model with online aerosol microphysics using measurements from the peak 19 of Whistler Mountain, BC, Canada (2182 m.a.s.l.). We evaluate the model for predictions of 20 aerosol number, size and composition during periods of free tropospheric (FT) and boundary-21 layer (BL) influence at "coarse" 4° x 5° and "nested" 0.5° x 0.667° resolutions by developing 22 simple FT/BL filtering techniques. We find that using temperature as a proxy for upslope flow 23 (BL influence) improved the model measurement comparisons. The best threshold temperature 24 was around 2°C for the coarse simulations and around 6°C for the nested simulations, with 25 temperatures warmer than the threshold indicating boundary-layer air. Additionally, the site was 26 increasingly likely to be in-cloud when the measured RH was above 90%, so we do not 27 compare the modeled and measured size distributions during these periods. With the inclusion 28 of these temperature and RH filtering techniques, the model-measurement comparisons 29 improved significantly. The slope of the regression for N80 (the total number of particles with

1 particle diameter,  $D_p > 80$  nm) in the nested simulations increased from 0.09 to 0.65,  $R^2$ 2 increased from 0.04 to 0.46, and log-mean bias improved from 0.95 to 0.07. We also perform 3 simulations at the nested resolution without Asian anthropogenic (AA) emissions and without 4 biomass-burning (BB) emissions to quantify the contribution of these sources to aerosols at 5 Whistler Peak (through comparison with simulations with these emissions on). The long-range 6 transport of AA aerosol was found to be significant throughout all particle number 7 concentrations, and increased the number of particles larger than 80 nm (N80) by more than 8 50%, while decreasing the number of smaller particles because of suppression of new-particle 9 formation and enhanced coagulation sink. Similarly, BB influenced Whistler Peak during summer months, with an increase in N80 exceeding 5000 cm<sup>-3</sup>. Occasionally, Whistler Peak 10 experienced N80 > 1000 cm<sup>-3</sup> without significant influence from AA or BB aerosol. Air masses 11 12 were advected at low elevations through forested valleys during times when temperature and 13 downwelling insolation were high, ideal conditions for formation of large sources of low-volatility 14 biogenic secondary organic aerosol (SOA). This condensable material increased particle growth and hence N80. The low-cost filtering techniques and source apportionment used in this study 15 16 can be used in other global models to give insight into the sources and processes that shape 17 the aerosol at mountain sites, leading to a better understanding of mountain meteorology and 18 chemistry.

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#### 20 1 Introduction

21 Atmospheric aerosol particles impact human health, climate and visibility. The magnitude of 22 these impacts has a strong dependence on the size, concentration and composition of the 23 particles (Rosenfeld et al., 2008; Clement et al., 2009). These particles can impact climate by 24 acting as seed particles for cloud formation, altering the brightness and/or the lifetime of clouds, 25 or by scattering incoming solar radiation (e.g. Boucher et al., 2013). These impacts of aerosols 26 on clouds and climate are driven by the number concentration of cloud condensation nuclei 27 (CCN), the particles large enough to serve as seeds for condensation of water to form cloud 28 droplets (typically diameters larger than 30-100 nm). Aerosol-cloud interactions are among the 29 most uncertain properties in climate forcing estimations (Boucher et al., 2013). Aerosol size 30 distributions, which are fundamental to aerosol-cloud interactions, evolve in the atmosphere as 31 a direct result of microphysical processes such as condensation, coagulation, nucleation,

primary emissions and deposition. Quantitatively estimating the climatic effect of aerosols
 involves understanding the evolution of aerosol size distributions.

Atmospheric aerosols emitted from or formed near the Earth's surface may remain in the planetary boundary layer (BL) or may be transported to the free troposphere (FT). Aerosols in the FT tend to have longer lifetimes than aerosols in the BL as deposition rates are lower in the FT (Croft et al., 2014). Therefore, aerosols in the FT can be transported over great distances and can affect remote regions where local emissions may be minimal.

8 One important characteristic about measurements at high-elevation mountain surface 9 sites is that there are periods where they can be used to investigate and understand FT 10 aerosols. However, these measurements are frequently influenced by a variety of aerosol 11 sources including advection of BL air with upslope flow. The complexity of air-mass influences 12 at high-elevation sites often makes measurements at these sites difficult to compare to 13 simulations of regional and global models that do not resolve the sub-grid topography. While 14 global models have been used to understand the processes shaping aerosols at mountain-top sites (e.g. Yu and Hallar, 2014), these models have resolution too coarse to explicitly resolve 15 16 topographic meteorology effects of many mountain peaks. Synoptic meteorology, including 17 advection and subsidence, influence the particles observed at mountain sites (Collaud Coen et 18 al., 2011; Gallagher et al., 2011); however, one may expect chemical transport models to 19 resolve these processes if synoptic meteorology is well-represented. A major issue in 20 comparing model simulations to mountain-top measurements is determining the appropriate 21 model layer that accurately represents the high-elevation measurements under the various 22 mountain-top conditions. Therefore, although measurements from these unique sites may be 23 used to evaluate global models, we must first understand how to properly sample the model for 24 comparison to the measurements.

25 The west coast of North America is routinely impacted by trans-Pacific transported 26 aerosol. Long-term measurements have been taken by Environment Canada at Whistler 27 Mountain, Whistler, British Columbia at a site situated approximately 100 km from the west 28 coast of Canada at the peak elevation of 2182 m a.s.l. (50.06°N, 122.96°W) (hereafter referred 29 to as Whistler Peak). Aerosol measurements at Whistler Peak have provided an understanding 30 of the baseline aerosol number concentrations and trace gases, which are characteristic of the 31 lower FT (Macdonald et al., 2011). These measurements often include contributions from Asian 32 anthropogenic aerosol, which has been shown to influence concentrations of carbon monoxide

1 (CO) and particulate matter along the west coast of North America (e.g. Husar et al., 2001; Jaffe 2 et al., 2003; Bailey et al., 2000; Leaitch et al., 2009; Singh et al., 2009; and references therein), 3 and can also lead to enhancements of ground-level ozone and CO concentrations (Macdonald 4 et al., 2011; Jaffe et al., 1999). During the Northern Hemisphere summer, Whistler Peak also 5 experiences influence from North American biomass-burning and biogenic secondary organic aerosol (SOA) that reach the peak through upslope flows or deep boundary layers (e.g. 6 7 Takahama et al., 2011). Global and regional models may help understand how these various 8 sources contribute to aerosol size distributions at Whistler.

9 In this study, we use a global chemical transport model with online aerosol microphysics 10 to investigate contributions to Whistler Peak aerosol from BL upslope flow, long-range transport 11 of Asian anthropogenic aerosol, local biomass-burning emissions, and other sources. We 12 compare model simulations to measurements taken from Whistler Peak with the goals of (1) 13 determining how to sample the model for comparison to mountaintop measurements and (2) 14 understanding how various sources influence the aerosol size distributions at Whistler. In the following section, the measurements and model simulations used in this study are described. 15 16 Section 3 describes the results, highlighting the data-filtering techniques, the comparison of 17 measured and simulated particle number concentrations, and the influence of Asian 18 anthropogenic emissions and biomass-burning aerosol on particle size distributions.

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#### 20 2 Methods

#### 21 2.1 Measurement description

22 Continuous high-elevation surface-based aerosol size distribution measurements are taken by 23 Environment Canada at Whistler Peak, located in the Coast Mountain range in southwestern 24 British Columbia (Figure 1). The Whistler Peak site began continuous measurements of particle 25 size distributions, trace gases (e.g.  $O_3$ , CO, SO<sub>2</sub>, NO<sub>x</sub>) and meteorological parameters 26 (temperature, pressure, relative humidity) in March 2002 (Macdonald et al., 2011), and there 27 have been intensive field campaigns in 2006 (e.g. Leaitch et al., 2009; McKendry et al., 2008) 28 and most recently in 2010 with the Whistler Aerosol and Cloud Study (WACS2010, Pierce et al., 29 2012) during the summer of 2010 (June 21 to July 29). We use measurements over a 20-month 30 period from April 2010 to December 2011. Data coverage is shown in Figure 2. Whistler Peak often resides in the lower FT (Gallagher et al., 2011) and therefore is an ideal location to provide 31 32 a baseline of aerosol number concentrations and trace gases for the lower FT, and also to

investigate the influence of long-range trans-Pacific transport of Asian anthropogenic emissions. The high-elevation Whistler Peak site is also ideal to investigate influences of local pollution sources such as biomass-burning aerosols that may appear in the summer during upslope flow or periods when the boundary layer develops fully to include the peak. In this study we focus on particle size distributions and use meteorological measurements to compare to model simulations to measurements taken at Whistler Peak. For detailed descriptions of instrumentation at Whistler Peak, refer to Macdonald et al. (2011).

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## 9 2.2 Model description

10 In this study, the Goddard Earth Observing System chemical transport model, GEOS-Chem 11 (http://geos-chem.org), combined with the TOMAS online aerosol microphysics module (GEOS-12 Chem-TOMAS) as described in D'Andrea et al. (2013), is used to simulate aerosol number 13 concentrations. The sensitivity to Asian anthropogenic aerosols and biomass-burning aerosols 14 is also tested through additional simulations. GEOS-Chem-TOMAS uses GEOS-Chem version 9.02 with 4° latitude by 5° longitude horizontal resolution for coarse simulations, and version 15 16 9.02 with 0.5° latitude by 0.667° longitude horizontal resolution for nested simulations (to be 17 described in more detail in section 2.3). GEOS-Chem-TOMAS uses 47 vertical layers from the 18 surface to 0.01 hPa, and meteorological inputs from the GEOS5 reanalysis 19 (http://gmao.gsfc.nasa.gov). In GEOS-Chem-TOMAS, aerosol size distributions are simulated 20 using 15 size sections from 3 nm to 10 µm. Nucleation rates are predicted using ternary 21 homogeneous nucleation (Napari et al., 2002) tuned globally by a factor of 10<sup>-5</sup> (Westervelt et 22 al., 2013). Ternary nucleation is used when  $NH_3$  mixing ratios are greater than 0.1 pptv, 23 otherwise nucleation rates are predicted by binary homogeneous nucleation (Vehkamäki et al., 24 2002). While these classical nucleation schemes do not get nucleation rates correct for the right 25 reason, the scaled Napari scheme estimated nucleation rates within a factor of 5 and the annual 26 number of nucleation days within 20% at 5 measurement sites in Westervelt et al. (2013). The 27 choice of nucleation scheme impacts aerosol number concentrations, particularly at smaller 28 sizes (Yu et al., 2010), and thus our choice nucleation scheme has some bearing on our results 29 here. Biomass-burning emissions are simulated from the Global Fire Emissions Database 3-30 hourly fire fractions (GFED3) (Mu et al., 2011). The primary black carbon and organic carbon 31 emission size distribution is assumed to be a lognormal distribution with a geometric mean 32 diameter of 100 nm, consistent with field measurements of biomass-burning smoke within the

1 first hour since emission (Janhäll et al., 2010), although larger diameters may be more 2 appropriate for some fires once sub-grid aging has occurred (Sakamoto et al., 2015). The density is assumed to be 2200 kg m<sup>-3</sup> and 1400 kg m<sup>-3</sup> for black carbon and organic carbon 3 respectively (Pierce et al., 2007). All simulations include an additional 100 Tg vr<sup>-1</sup> of 4 5 anthropogenically enhanced secondary organic aerosol (SOA), spatially correlated with anthropogenic carbon monoxide emissions as per Spracklen et al. (2011) and D'Andrea et al. 6 7 (2013). SOA is assumed to be effectively non-volatile, with an average saturated vapor pressure, C<sup>\*</sup>, of less than approximately  $10^{-3} \mu g m^{-3}$  (Pierce et al., 2011). This is consistent with 8 kinetic, gas-phase-diffusion-limited growth with condensation proportional to the Fuchs-9 10 corrected aerosol surface area as per D'Andrea et al. (2013). A complete description of 11 emissions is provided in Stevens and Pierce (2014). Simulations were run from April 2010 12 through December 2011 with one-month spin-up from a pre-spun-up restart file.

13 The surface layer in the simulated 4° x 5° grid box (Figure 1) encompassing Whistler 14 Peak has a mean elevation of approximately 600 m a.s.l., and the surface layer in the simulated 0.5° x 0.667° grid box (Figure 1) encompassing Whistler Peak has a mean elevation of 15 16 approximately 1290 m a.s.l.: however. Whistler Peak resides at an elevation of 2182 m a.s.l. 17 (the simulated grid boxes includes mountainous regions, oceanic regions and urban cities such as Vancouver, B.C.). This implies that the appropriate model layer to represent Whistler Peak 18 19 would be the layer corresponding to approximately 1580 m above the modeled ground level in 20 the coarse simulations and 890 m above the modeled ground level in the nested simulations. 21 While GEOS-Chem does have vertical mixing for the resolved BL and synoptic/convective 22 mixing between the BL and FT, it does not resolve sub-grid vertical transport due to topographic 23 and upslope flows. Therefore, in order to represent the influence of boundary layer air at sub-24 grid scale terrain, conditions were developed to select the model level appropriate for the site 25 conditions and are described in section 3.2.

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## 27 2.3 Description of simulations

We test the sensitivity of aerosol size distributions in GEOS-Chem-TOMAS to (a) the removal of Asian anthropogenic emissions, and (b) the removal of biomass-burning emissions. Simulations are summarized in Table 1. In all simulation names, the C indicates coarse simulations (4° x 5° resolution).. The BASE and BASE\_C simulations include all emissions from GEOS-Chem-TOMAS as described in Stevens and Pierce (2014). The noAsia and noAsia\_C simulations 1 remove anthropogenic SO<sub>2</sub>, NH<sub>3</sub>, anthropogenic organic aerosol (including the 2 anthropogenically enhanced SOA from D'Andrea et al. (2013) and elemental carbon from India, 3 China, and southeast Asia following the domain covered in Streets et al. (2003). The noAsia 4 simulations do not mask biogenic SOA and other natural primary emissions. The noBioB and 5 noBioB\_C simulations mask all biomass-burning emissions globally while all other emission 6 sources remain unchanged.

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# 8 3 Results and Discussion

## 9 3.1 Filtering model and measurement data

10 Figures 3a and 3b show 1:1 plots comparing measured and simulated N14 and N80 (the total number of particles with particle diameter,  $D_p > 14$  nm and  $D_p > 80$  nm, respectively) using only 11 12 the model surface layer (averaged elevation of 600 m a.s.l.) at the coarse resolution. All metrics 13 from Figure 3 are summarized in Table 2. In these comparisons, when using the model surface 14 level only (panels a and b), the model consistently over-predicts N14 during times with low measured number concentrations (e.g.  $N14 < 100 \text{ cm}^{-3}$ ), and marginally under-predicts during 15 more polluted conditions (e.g.  $N14 > 1000 \text{ cm}^{-3}$ ), with a slope of 0.11 in the linear regression. 16 17 Also, comparisons with N80 show slightly improved but similar conclusions with over-predictions 18 during cleaner conditions, marginal under-prediction during more polluted conditions, and a 19 slope of 0.14 in the linear regression. The nested simulations yield similar results (Figure 3g and 20 3h), with over-predictions of N14 and N80 in cleaner conditions, and under-predictions of N14 21 and N80 in more polluted conditions. Therefore, it is clear that simulated particle number 22 concentrations in the model surface level alone do not accurately represent the measurements 23 at Whistler Peak under all meteorological conditions. Figure 3c and 3d show 1:1 plots 24 comparing measured and simulated N14 and N80 using only the 1580 m model layer (2200 m 25 a.s.l.) at the coarse resolution, which corresponds to the actual elevation at Whistler Peak. The 26 1580 m level produces marginally improved but similar conclusions to the model surface level. 27 GEOS-Chem-TOMAS consistently over-predicts N14 during clean conditions and under-28 predicts N14 during more polluted conditions, with a slope of 0.21. Similarly, N80 is over-29 predicted during clean conditions and under-predicted during more polluted conditions when 30 assuming the 1580 m level, with a slope of 0.27. The nested simulations once again yield 31 similar results (Figure 3i and 3j), over-predicting N14 and N80 in cleaner conditions, and under-32 predicting N14 and N80 in more polluted conditions. A common trait between the surface and 1 1580 m (860 m) level comparisons is the large model over-predictions during clean 2 measurements. However, with both the surface level and the 1580 m (860 m) level, there may 3 be conditions where the model predicts the correct particle number concentration. Even though 4 the metrics for the coarse simulations are slightly better in some cases than the nested 5 simulations, all of the metrics were overall very poor. Therefore, we must use meteorological 6 conditions to determine how best to compare the model to measurements.

7 A characteristic of Whistler Peak is that the measurement site frequently experiences in-8 cloud conditions (Macdonald et al. 2011). Previous work showed that measurements with RH > 9 90% at the Whistler Peak station corresponded closely to cloudy conditions at the site. When 10 Whistler Peak is in cloud, the measured aerosol size distributions can not be used for model 11 comparison, because some of the particles will be in cloud droplets, either by activation or by 12 diffusive collection. Therefore, the data have been filtered based on the ambient relative 13 humidity i.e. data are not included in the measurement/model comparison when the measured 14 RH is > 90%. Although use of 90% as the threshold value is an estimate, the identification of clouds by this criterion agrees with the visual confirmation of clouds through regular site 15 16 photographs. This RH filter significantly reduces the number of points where the model 17 consistently over-predicts the number of particles during low number concentration conditions. 18 This can be seen in Figure 4, which shows a histogram of the frequency of data points as a 19 function of measured N80 (Figure 4a for coarse simulations, Figure 4b for nested simulations), 20 where the dark gray bars are with the RH filter off and the blue bars are with the 90% RH filter 21 applied. If a particle is activated, it will not be measured and therefore N80 can reach very low number concentrations (N80 < 1 cm<sup>-3</sup>), whereas particles with  $D_p$  between 14 and 80 nm might 22 23 not activate and will be measured. With the RH filter applied, the number of points with measured N80 < 100 cm<sup>-3</sup> is reduced most strongly for all simulations, and nearly all the points 24 with measured N80 < 10 cm<sup>-3</sup> are removed. Therefore, applying the RH filter removes in-cloud 25 26 conditions when we expect measurements to be biased low. It is likely that the RH > 90% filter 27 eliminates some data that was not during cloudy conditions, as some of the data with higher 28 measured N80 are eliminated. We tested other critical RH values as filters, but moving to larger 29 RH values allowed cases with low N80 (e.g.  $< 5 \text{ cm}^{-3}$ ) to be included.

30 Whistler Peak may be encompassed by an air mass originating from lower altitudes if 31 the boundary layer is very deep (over 1.5 km) or if there is upslope flow. To separate conditions 32 of upslope flow or deep boundary layer, from free tropospheric conditions, we also define a

1 threshold temperature. When the measured temperature exceeds the threshold temperature, 2 upslope flow is assumed and the model surface layer is used. When the measured temperature 3 is less than the threshold temperature, then FT air is assumed at the peak and the 1580 m (860 4 m) model layer is used. Various temperature thresholds were imposed for determining which 5 model level to use. The temperature-filtered simulated particle size distribution that most 6 accurately represents the measured particle size distribution based on correlation statistics 7 summarized in Table 3 is when a threshold of about 2°C is assumed for the coarse simulations, 8 and about 6°C for the nested simulations. That is, when the measured temperature at Whistler 9 peak is less than 2°C (6°C), the 1580 m (860 m) model layer is assumed and when the 10 measured temperature is greater than 2°C (6°C), the model surface layer is assumed. We note, 11 however, that due to uncertainty in these best threshold temperatures from (1) different metrics 12 giving best values at different temperatures and (2) model errors, there is likely no significance 13 in the different threshold temperatures for the two different resolutions. As temperature is a 14 simple proxy for boundary-layer influence, interannual variability in synoptic conditions at Whistler Peak beyond the measurement period in this study may lead to variability in an ideal 15 16 threshold temperature between years. The observed mean particle number size distributions 17 and various temperature-dependent simulated mean particle number size distributions for the 18 duration of the measurement period are shown in Figure 5a and 5b (coarse and nested, 19 respectively). When using the filtered data and assuming only the model surface level (black 20 dotted line), the predicted total number of particles is much too high for all simulations, and 21 when assuming the 1580 m (860 m) level (black dashed line) the number of small particles is 22 over-predicted and the number of particles larger than roughly 30 nm is under-predicted.

23 Previous studies have used other methods to represent boundary layer influence at 24 Whistler Peak and other high mountaintop sites; however, these methods were used to identify 25 days of BL influence, whereas we seek to sort simulated hourly time points into either BL or FT 26 influence. We therefore synthesized the following methods to test an alternate filter based on 27 N14, rather than attempting to apply each method. Obrist et al. (2008) and Weiss-Penzias et al. 28 (2006) noted diurnal water vapor cycles as indicators of BL influence at Colorado and Oregon 29 mountain peaks; in New Hampshire, Grant et al. (2005) identified days with BL influence using 30 early morning minimum and afternoon maximum temperatures; and daily total particle number (condensation nuclei, CN, in our case, N14) increases indicated BL uplift in Hawaii (Bodhaine, 31 32 1996) and Switzerland (e.g. Baltensperger et al., 1997). At Whistler Mountain, Macdonald et al.

(2011) used temperature data from multiple vertical levels on the mountain to define a stability
index as an indicator of boundary layer influence; however, many of the temperaturemeasurement sites used in Macdonald et al. (2011) were not operational during our time period.
Gallagher et al. (2011) described the frequency of BL influence at Whistler by evaluating how
well the change in CN concentration throughout each day adhered to a typical sinusoidal
pattern, noting that confidence in the influence of vertical transport is higher on days when CN
correlates strongly with water vapor.

8 We tested a filter based on CN changes throughout the day on simulated Whistler 9 measurements informed by these studies. Though some of the studies discussed above used 10 water vapor, we used CN because Gallagher (2010) found that CN was a more robust indicator 11 of BL influence at Whistler. We identified BL influence days using increasing CN concentration 12 from morning to midday (9:00-11:00 average < 11:00-13:00 average < 13:00-15:00 average) 13 and selected 11:00 AM to 5:00 PM local time as BL influence hours within those days when we 14 applied the surface model layer. We assumed all other time points represented the FT and used the 2.5km model layer for those times. However, this filter method was less successful than the 15 temperature filter, with  $R^2 = 0.04$  and slope = 0.15 for regression of simulated versus observed 16 N80 particles (not shown). The correlation did not improve when we relaxed the parameters:  $R^2$ 17 18 was 0.01 and the slope was 0.07 for the N80 regression when we used CN increase from 9:00-19 11:00 to 13:00-15:00 as the BL day criterion (disregarding 11:00-13:00) and no daytime 20 restriction on BL hours.

21 The low performance of this CN-cycle method could be due to overly strict criteria; the 22 BL could influence peak aerosol on days when CN does not increase from morning to early 23 afternoon. In particular, during periods of sustained high pressure systems, for example, when 24 the site was influenced by the boundary layer throughout both day and night, this CN filter would 25 not result in identifying the BL influence. Also, Whistler frequently observes new-particle 26 formation events. Gallagher et al., (2011) estimated that the new-particle formation at Whistler 27 was generally correlated with upslope flow and BL air. However, it is likely that not all new-28 particle formation events are associated with BL air and thus would contribute error to using CN 29 as a classification of BL air. Thus, the methods described above to identify days of BL influence 30 from observed water vapor, timing of temperature extrema, and CN increases may not be 31 robust for sorting simulated hourly time points into either BL or FT influence as we do here.

1 The RH filter combined with the temperature-dependent model level assumption 2 improves comparisons with measurements. Figure 3e and 3f include the RH filter and the 3 temperature filter in the coarse simulated and measured comparison of N14 and N80. With 4 these two filters included in the analysis, the slope of the regression for N14 and N80 both 5 significantly improve (0.44 and 0.54 respectively) as well as the R<sup>2</sup> (0.3 and 0.44 respectively). 6 Similar results are found for the nested simulations (Figure 3k and 3l), with the slope improving 7 even further to 0.65 for both N14 and N80, and the  $R^2$  improves to 0.4 for N14 and 0.46 for N80. 8 A small positive bias still remains in the simulated number concentrations. However, the log-9 mean bias for N14 and N80 improved from 0.61 and 0.67 to 0.08 and -0.03 respectively in the coarse simulations, and from 0.91 and 0.95 to 0.17 and 0.07 respectively in the nested 10 11 simulations (Table 2). For the following sections, only the nested simulations with these best 12 temperature and RH filters will be used.

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## 14 **3.2** Time series and data density

In this section, we address the seasonal cycle of data availability and completeness once the 15 16 filters have been applied. Figure 2 shows a time series of N14 for all measurements (black 17 points), and the temperature and RH filtered points are shown in green when the 860 m 18 simulated layer is selected and red when the surface simulated layer is selected from the nested 19 simulations. Times where black points exist but no red or green points exists show that the 20 model data has been filtered using RH for in-cloud conditions. Periods where there are no 21 points are time periods where the SMPS was not operating at Whistler. The bracketed number 22 in the legend corresponds to the total number of data points for each condition. There are clear 23 seasonal trends in N14 at Whistler Peak, with high particle number concentrations during the 24 summer months, and relatively low particle number concentrations during the winter months. 25 The summer maximum is due in part to the advection of BL air due to upslope flow to the peak 26 as well as influence from biomass burning during the Northern Hemisphere boreal forest fire 27 season, as we will show. For the period of July through September, 77% of the points are 28 identified as BL in 2010 and 65% in 2011. For the period December 2010 through February 29 2011, 100% of the points are identified as FT.

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#### 31 3.3 Influence of Asian anthropogenic emissions

1 Whistler Peak experiences conditions where trans-Pacific FT air transports anthropogenic 2 aerosol from Asia and influences aerosol size distributions. Figure 6a and 6b show 1:1 plots for 3 measured and simulated N14 and N80, respectively, from the BASE and noAsia (Asian 4 anthropogenic emissions turned off) simulations, where the gray crosses represent all points 5 (implementing the temperature and RH filter as discussed earlier) and the green crosses 6 represent all points where N14<sub>BASE</sub> – N14<sub>noAsia</sub> > 50 cm<sup>-3</sup> and N80<sub>BASE</sub> – N80<sub>noAsia</sub> > 50 cm<sup>-3</sup>, where 50 cm<sup>-3</sup> is our criteria for "high Asian anthropogenic influence". Thus, we are using the 7 8 simulations to characterize periods of high Asian aerosol influence. Comparing the BASE 9 simulation with the noAsia simulation indicates that Asian anthropogenic aerosol influences N14 10 and N80 during both clean and polluted periods.

11 The overall impact that transport from Asia has on the number size distribution at 12 Whistler is shown in Figure 6c, which shows the simulated contribution to particle number 13 concentration due to Asian anthropogenic aerosol (BASE - noAsia) as a function of particle diameter,  $D_p$ , averaged over the measurement period when N80<sub>BASE</sub> – N80<sub>noAsia</sub> > 50 cm<sup>-3</sup>. 14 There is a negative contribution to particle number concentrations  $(dN/dlog_{10}D_{0})$  exceeding 200 15 cm<sup>-3</sup> for particle sizes with diameters smaller than approximately 20 nm, and a positive 16 contribution to particle sizes with diameters larger than around 70 nm, not exceeding 50 cm<sup>3</sup>. 17 18 These results imply that the Asian emissions are increasing the concentration of particles larger 19 than about 100 nm during transport; however, significantly decreasing the concentration of 20 particles less than 20 nm. These larger particles are suppressing nucleation and acting as a 21 coagulation sink for smaller particles; both of these effects cause the presence of Asian aerosol 22 emissions in the model to reduce the number of Aitken-mode particles. Figure 6d shows a time 23 series of the percent (green) and absolute (black) contribution to N80 due to Asian anthropogenic aerosol. Periods where the contribution to N80 exceeds 30% correspond 24 generally to colder months and at particle number concentrations less than 800 cm<sup>-3</sup>, as these 25 are the periods where the BL height is the lowest and Whistler Peak is influenced predominantly 26 27 by FT air masses carrying Asian anthropogenic aerosol.

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#### 3.4 Influence of North American boreal forest fires

30 Whistler Peak also experiences periods of increased concentrations of biomass-burning 31 aerosol. Similar to Figs. 6a and 6b, Figure 7a and 7b show 1:1 plots for measured and 32 simulated N14 and N80 respectively, from the BASE and noBioB (biomass-burning emissions

1 turned off) simulations. The gray crosses represent all points (implementing the temperature 2 and RH filter as discussed earlier) and the red crosses represent all points where N14<sub>BASE</sub> - $N14_{noBioB} > 100 \text{ cm}^{-3}$  and  $N80_{BASE} - N80_{noBioB} > 100 \text{ cm}^{-3}$ , where 100 cm<sup>-3</sup> is our criteria for "high 3 4 biomass-burning influence". We are using the simulations to characterize periods of high 5 biomass-burning aerosol influence. Comparing the BASE simulation with the noBioB simulation 6 indicates that biomass-burning aerosol influences all particle sizes (N14 and N80), and the 7 biomass-burning emissions contribute to many of the periods with the highest particle number 8 concentrations in both the model and the measurements. This result is consistent with biomass-9 burning plumes, which contain high number concentrations of Aitken- and/or accumulation-10 mode particles (Janhäll et al., 2010; Sakamoto et al., 2015).

11 The impact of biomass-burning aerosol on the particle size distribution is quantified in 12 Figure 7c, which shows the simulated contribution to particle number concentration due to 13 biomass-burning aerosol (BASE - noBioB) as a function of particle diameter, D<sub>n</sub>, averaged over the year when  $N80_{BASE} - N80_{noBioB} > 100 \text{ cm}^{-3}$ . For particle sizes with  $D_p > 20 \text{ nm}$ , there is a 14 positive contribution to particle number concentrations  $(dN/dlog_{10}D_p)$  with enhancements in the 15 accumulation mode exceeding 150 cm<sup>-3</sup> between 100-200 nm. The size at which this peak 16 17 occurs is likely sensitive to the emissions size assumed in GEOS-Chem-TOMAS, which is 18 currently set to a mode centered at 100 nm. Biomass-burning emissions size in aerosol models 19 is uncertain because it is difficult to capture sub-grid aging due to coagulation and chemistry 20 (Sakamoto et al., 2015). There is also a marginal reduction of nucleation-mode particles, which 21 is likely due to an increase in the coagulation sink due to the enhancement of accumulation 22 mode particles. A clear seasonality of Whistler Peak experiencing biomass-burning events is 23 shown in Figure 7d, which shows a time series of the percent (red) and absolute (black) 24 contribution to N80 due to biomass-burning aerosol. Expectedly, N80 contributions exceeding 25% and upwards of 5000 cm<sup>-3</sup> at Whistler Peak only occur during the summer months (North 25 American forest fire season). 26

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## 28 3.5 High aerosol loading but low biomass-burning/Asian anthropogenic influence

In addition to the periods with biomass burning influence, Whistler Peak has periods with high
particle number concentrations without influence from biomass-burning emissions. Figure 8
shows 2-day back trajectories (HYSPLIT version 4.9 (Draxler and Hess, 1997, 1998; Draxler,
1999)) for July 2010 including only times with low biomass burning or Asian anthropogenic

influence (N80<sub>BASE</sub> - N80<sub>noAsia</sub> < 50 cm<sup>-3</sup> and N80<sub>BASE</sub> - N80<sub>noBioB</sub> < 100 cm<sup>-3</sup>) and N80<sub>BASE</sub> > 1 2 1000 cm<sup>-3</sup>. Figure 8 is colored by (a) altitude above ground level, (b) surface air temperature, 3 and (c) downwelling insolation. During July 2010, the trajectories of these air parcels with high 4 particle number concentrations advect to Whistler Peak through the heavily wooded (montane 5 and coastal forests) interior of British Columbia at elevations below 1000 m (Figure 8a). This 6 region is largely characterized by series of forested valleys with high emissions of biogenic 7 volatile organic compounds known to be precursors for SOA formation. The emissions of 8 biogenic vapors are highly dependent on surface air temperature and downwelling insolation. 9 Increasing surface air temperature and solar insolation yield increases in biogenic emissions (Guenther et al. 2006). Paasonen et al. (2013) and Leaitch et al. (1999) showed that regions 10 11 influenced by biogenic VOC emissions have increased concentrations of CCN-sized particles 12 (e.g. N80) with increasing temperature due to high SOA concentrations and increased growth of 13 Aitken-mode particles to these larger sizes. Figures 8b and 8c show that most of the trajectories experienced warm temperatures (>292 K) and high downwelling insolation (>700 W m<sup>-2</sup>) prior to 14 arriving at Whistler. The advection of relatively warm BL air that has passed over BVOC-15 16 generating forests to Whistler Peak is a likely reason for high particle concentrations at Whistler 17 Peak in the absence of biomass-burning influence. However, metropolitan areas such as 18 Vancouver and Seattle could also contribute aerosol to some of the trajectories. The importance 19 of biogenic aerosol during this period was identified as part of the Whistler Aerosol and Cloud 20 Study (e.g. Pierce et al., 2012; Ahlm et al., 2013).

21

#### 22 4 Conclusions

23 Continuous high-elevation surface-based aerosol size distribution measurements have been 24 taken by Environment Canada at Whistler Peak (50.06°N, 122.96°W, 2182 m a.s.l.), located in 25 the Coast Mountain range in southwestern British Columbia. Whistler Peak is influenced by long-range transport of trans-Pacific air masses in the free troposphere (FT) or local boundary 26 27 layer (BL) air being lifted thermodynamically or orographically to the mountain peak. In this 28 study, we use measurements from Whistler Peak and simulations from the global chemical 29 transport model GEOS-Chem-TOMAS at 4° x 5° (coarse) and 0.5° x 0.667° (nested) 30 resolutions, to investigate the source attribution and processes that shape the aerosol at 31 Whistler Peak.

1 To compare simulations to measurements at Whistler Peak, it was necessary to develop 2 filtering techniques to determine whether there was BL or FT influence at Whistler Peak. We 3 found that using the measured temperature at Whistler Peak as a proxy for upslope flow, we 4 could improve our agreement with measurements, and that temperature was a better proxy than 5 a CN proxy that had been previously used as a proxy for BL air at Whistler (Gallagher et al., 2011), although it is possible that better proxies exist. The best threshold temperature we found 6 7 was 2°C for the coarse simulations and 6°C for the nested simulations, with warmer 8 temperatures indicating upslope flow. If the temperature was colder than 2°C (6°C) in the coarse 9 (nested) simulations, the simulated layer corresponding to the actual elevation of Whistler Peak 10 was used (1580 m (860 m) level above the model surface, 600 m (1290 m)). Whistler Peak also 11 often experiences in-cloud conditions when the RH measurements are larger than 90%. When 12 in cloud, activated particles are not measured, so aerosol size distribution measurements 13 should not be used. Therefore, we filter out measurements when the measured RH exceeded 14 90%. With the inclusion of these two filtering techniques, the model measurement comparisons of N14 and N80 improved significantly in the coarse (nested) simulations with the slope of the 15 regression improving to 0.44 (0.65) and 0.67 (0.65) respectively, R<sup>2</sup> improving to 0.30 (0.40) and 16 0.44 (0.46) respectively, and log-mean bias improving to 0.08 (0.17) and -0.03 (0.07). 17 18 respectively.

19 Due to the high elevation of Whistler Peak, the measurement site is often influenced by 20 long-range transport of Asian anthropogenic aerosol. To investigate this, a base simulation 21 (BASE) was compared to a simulation with Asian anthropogenic emissions turned off (noAsia) 22 at the nested resolution. High Asian influence periods were determined when the difference in 23 particle number concentrations between the BASE simulation and the noAsia simulation exceeded 50 cm<sup>-3</sup>. The long-range transport of Asian anthropogenic aerosol was found to occur 24 25 during periods with low total particle number concentrations and increase the number of 26 particles larger than 80 nm (N80), but decrease the number of smaller particles due to 27 suppression of new-particle formation and increases in coagulation. The influence of Asian 28 anthropogenic aerosol was found to be most prevalent during the winter months, when the BL 29 height is the lowest and long-range transport dominates the aerosol at Whistler Peak.

30 Whistler Peak experiences BL air influence particularly during summer months, and 31 during fire seasons upslope flow or deep boundary layers can transport biomass-burning 32 aerosol to the peak. To investigate this, the BASE simulation was compared to a simulation with biomass-burning emissions turned off (noBioB) at the nested resolution. Similar to the noAsia comparison, high influence periods were determined when the difference between the BASE simulation and the noBioB simulation exceeded 100 cm<sup>-3</sup>. Biomass-burning aerosol was found to increase particle numbers of sizes larger than 20 nm, particularly at sizes near the biomassburning source diameters in GEOS-Chem-TOMAS (100 nm), at Whistler Peak.

6 Occasionally, Whistler Peak measured N80 in excess of 1000 cm<sup>-3</sup> without significant 7 influence from Asian anthropogenic or biomass-burning aerosol. We used back trajectories to 8 investigate these high particle number concentration periods. The air masses during these 9 periods were found to flow at low elevations through forested valleys, when both the 10 temperature and downwelling insolation were high. These conditions are ideal for biogenic 11 emissions and low-volatility biogenic secondary organic aerosol (SOA) formation. Based on this 12 back-trajectory analysis, it is hypothesized that this possible source of SOA could be a large 13 source of condensable material, which could increase particle growth and hence increase N80.

14 Mountain measurement sites are difficult to simulate in global chemical transport models. By using simple filters on simulated data, we were able to improve model-measurement 15 16 comparisons. We were also able to test the sensitivity of the simulations to Asian anthropogenic 17 emissions and local biomass burning to determine source apportionment at a high elevation 18 mountain site. These low-cost techniques could be used in other global models to more 19 accurately represent mountain measurement sites, leading to a better understanding of 20 mountain meteorology and chemistry; however, the details of the filtering likely need to be tuned 21 for different models and mountains.

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	Simulation	Resolution	Asian Anthropogenic Emissions	Biomass Burning Emissions
_	BASE_C	4° x 5°	yes	yes
	noAsia_C	4° x 5°	no	yes
	noBioB_C	4° x 5°	yes	no
	BASE	0.5° x 0.667°	yes	yes
	noAsia	0.5° x 0.667°	no	yes
	noBioB	0.5° x 0.667°	yes	no

2 Table 1 – Summary of the GEOS-Chem-TOMAS simulations used in this study. Note that C

3 indicates coarse simulations (4° x 5°).

	m		R <sup>2</sup>		LMB	
coarse	N14	N80	N14	N80	N14	N80
surface layer	0.11	0.14	0.05	0.17	0.61	0.67
1580 m layer	0.21	0.27	0.19	0.36	0.06	-0.01
2°C T and 90% RH filter	0.44	0.54	0.30	0.44	0.08	-0.03
nested	N14	N80	N14	N80	N14	N80
surface layer	0.09	0.09	0.02	0.04	0.91	0.95
860 m layer	0.27	0.26	0.19	0.24	0.23	0.26
6°C T and 90% RH filter	0.65	0.65	0.4	0.46	0.17	0.07

2 Table 2 – Summary of the slope of the linear regression (m), correlation (R<sup>2</sup>), and log-mean bias

3 (LMB) for coarse and nested simulations. These statistics are found by comparing the average

4 values of the aerosol number concentrations during the measurement period to measurements

5 at Whistler Peak. Bolded numbers represent the best statistical result between all simulations.

coarse	-3°C	0°C	2°C	3°C	4°C	6°C
R <sup>2</sup>	0.35	0.40	0.44	0.43	0.43	0.43
m	0.48	0.51	0.54	0.53	0.53	0.51
LMB	0.08	0.02	-0.03	-0.06	-0.11	-0.18
nested	3°C	5°C	6°C	7°C	9°C	11°C
R <sup>2</sup>	0.43	0.45	0.46	0.44	0.42	0.41
m	0.67	0.66	0.65	0.62	0.58	0.53
LMB	0.18	0.11	0.07	0.05	0.00	-0.03

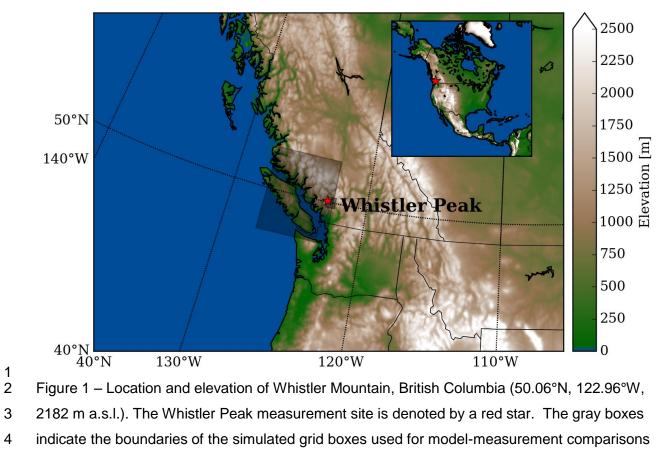
2 Table 3 – Summary of the slope of the linear regression (m), correlation (R<sup>2</sup>), and log-mean bias

3 (LMB) for different temperature cutoffs for coarse and nested simulations. These statistics are

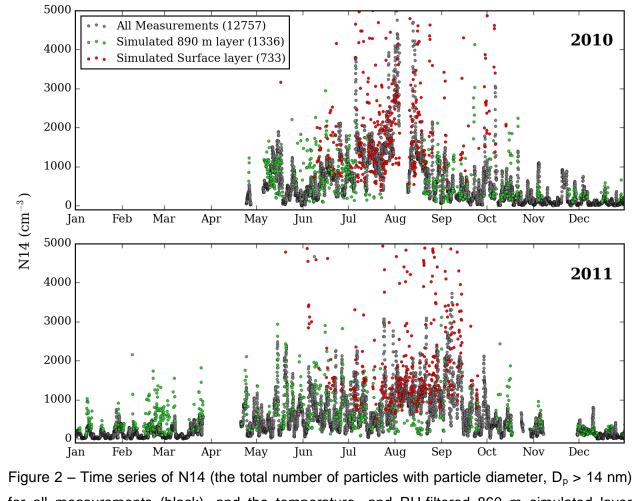
4 found by comparing the simulated aerosol number concentrations during the measurement

5 period to measurements at Whistler Peak for various temperature cutoffs with the 90% RH

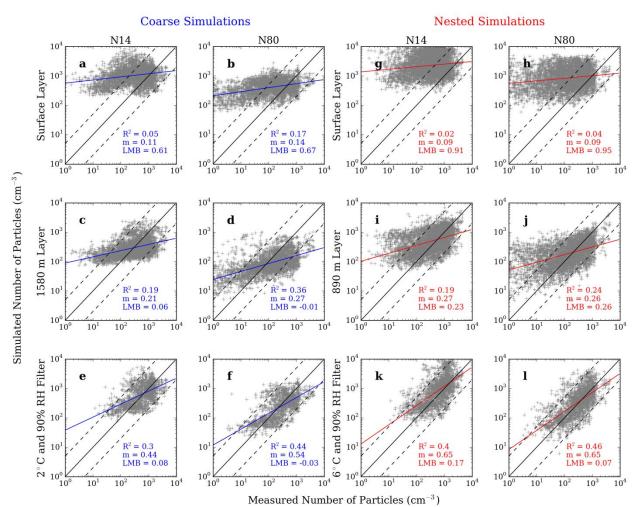
6 filtering included. Bolded numbers represent the best statistical result between all simulations.



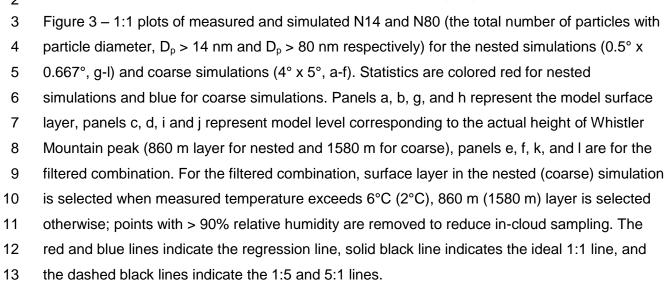
(0.5° x 0.667° resolution for nested and 4° x 5° resolution for coarse). 

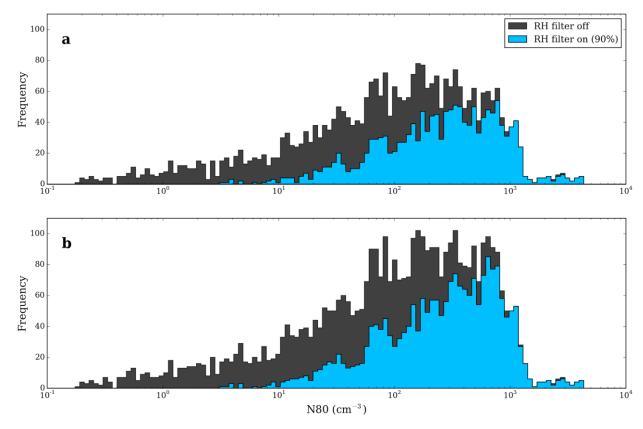


for all measurements (black), and the temperature- and RH-filtered 860 m simulated layer (green) and surface layer (red). The bracketed number in the legend corresponds to the total number of data points for each condition.





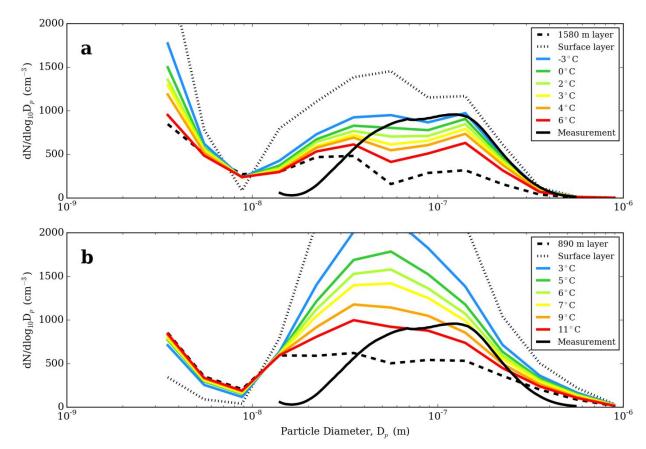




2 Figure 4 – Histogram of the frequency of data points as a function of measured N80 (the total

- 3 number of particles with particle diameter, Dp > 80 nm), for (a) coarse simulations (4° x 5°), and
- 4 (b) nested simulations (0.5° x 0.667°). The dark gray bars are with the relative humidity filter off
- 5 and the blue bars are with the 90% relative humidity filter applied.





2 3

Figure 5 – Observed and temperature-dependent simulated mean particle number size distributions for the measurement period for the (a) coarse simulations (4° x 5°), and (b) nested resolution (0.5° x 0.667°). The solid black line indicates the measured data, the black dashed line indicates the 860 m simulated layer, the black dotted line indicates the simulated surface layer and the colored lines indicate various temperature thresholds, where cool colors indicate colder threshold values, and warmer colors indicate warmer threshold values.

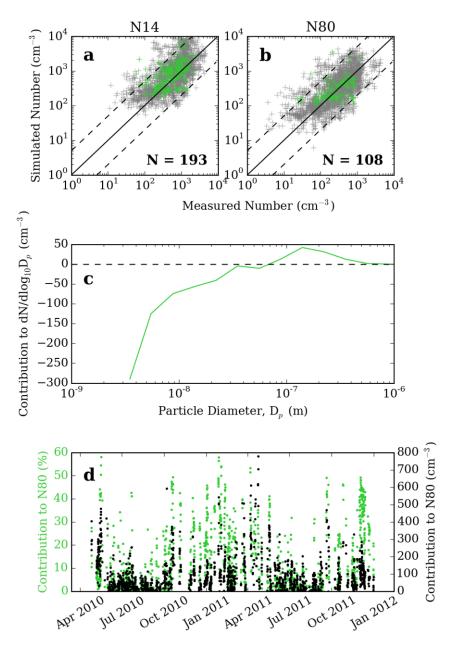


Figure 6 – 1:1 plots for measured and simulated (nested resolution) (a) N14 and (b) N80 where the gray crosses represent all points (implementing the temperature and RH filter), the green crosses represent all points where BASE – noAsia > 50 cm<sup>-3</sup>, and the number of points is given by N. (c) The simulated contribution to particle number concentration due to Asian anthropogenic aerosol as a function of particle diameter, D<sub>p</sub>, averaged over the year when N80<sub>BASE</sub> – N80<sub>noAsia</sub> > 50 cm<sup>-3</sup>. (d) Time series of the percent (green) and absolute (black) contribution to N80 due to Asian anthropogenic aerosol.

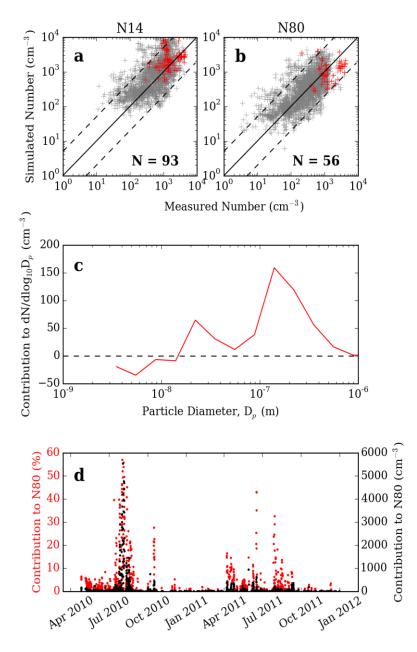


Figure 7 – 1:1 plots for measured and simulated (nested resolution) (a) N14 and (b) N80 where the gray crosses represent all points (implementing the temperature and RH filter), the red crosses represent all points where BASE – noBioB > 100 cm<sup>-3</sup>, and the number of points is given by N. (c) The simulated contribution to particle number concentration due to biomass burning aerosol as a function of particle diameter,  $D_p$ , averaged over the year when N80<sub>BASE</sub> – N80<sub>noBioB</sub> > 100 cm<sup>-3</sup>. (d) Time series of the percent (red) and absolute (black) contribution to N80 due to biomass burning aerosol.

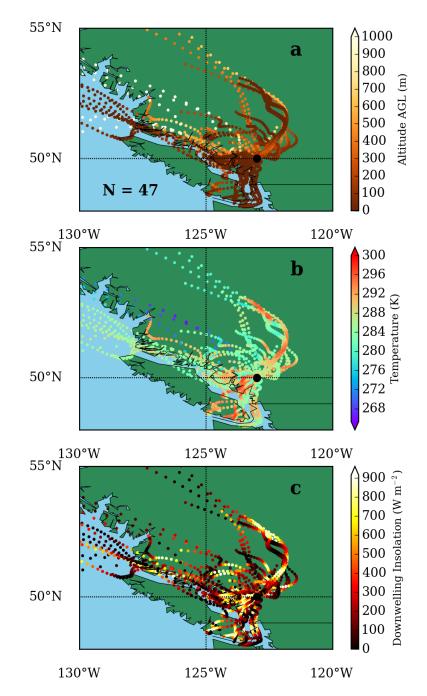


Figure 8 – 2-day back trajectories for July, 2010 including only times with low biomass burning
or Asian anthropogenic influence (N80<sub>BASE</sub> – N80<sub>noAsia</sub> < 50 cm<sup>-3</sup> and N80<sub>BASE</sub> – N80<sub>noBioB</sub> < 100</li>
cm<sup>-3</sup>) and N80<sub>BASE</sub> > 1000 cm<sup>-3</sup>. The trajectories are colored by (a) altitude above ground level,
(b) surface air temperature, and (c) downwelling insolation. The end point of the trajectory is set
to 100 m above ground level. The black circle represents Whistler Peak and the number of back
trajectories is given by N.