

A – POINT BY POINT RESPONSE TO REVIEWS

Answers to the interactive comment on “Characterization of PM10 sources in the central Mediterranean” by Anonymous Referee #1

The manuscript presents interesting results on PM10 chemical characterization in a Central Mediterranean coastal site. These data are useful for improving the understanding of aerosols variability and processes in the Med Basin, and validate modelling outputs in a region with a general lack of data. The work has been carried out correctly, well presented, and data interpretation is in my opinion reasonable within the uncertainty of the methods used.

We thank the reviewer for the positive evaluation of our work and the productive comments and suggestions. These are addressed point-by-point in the following.

The main uncertainty is given by the lack of EC and mostly OC data for the source apportionment. Authors claim on the low mass contribution of EC, but in my opinion its signal is still key in tracing primary combustion emissions. OC would also help improving significantly the share of secondary vs primary PM, mostly for the two combustion sources found. In any case authors have presented the (partial) ECOC data available and compared them with PMF source contributions, which show reasonable results.

We agree with the referee that OC and EC data would add valuable information to this work, as information on the sources and contributions of the carbonaceous aerosol component is also missing in the region of study. Nevertheless, in a recent work (Lucarelli et al., 2015) some of the authors tested the effect of reducing input species in the application of PMF on an experimental data set providing an extended chemical speciation, including elements, EC and OC, and ions. As concerns the test performed removing EC and OC data (and still including elements and ions), PMF results did not significantly vary with respect to the ones obtained including also the carbonaceous species.

This is due to the fact that, although OC is a major aerosol component and EC is an important combustion tracer, OC is a not specific component of PM and EC is not the only combustion tracer. In fact, in this study on Lampedusa, we benefit from other elemental and ionic markers to trace combustion emissions.

Nevertheless, as we agree with the referee about the strength of EC as a marker, lacking a complete EC/OC data-set for the whole campaign, we compared the results for the combustion source as derived by the PMF analysis with the available EC/OC data, although they cover a limited time period. As stated by the referee, the results of this comparison are reasonable. This, together with the evidences from the methodological tests PMF runs with reduced input species, make us confident of the presented results.

We have added some comments on this topic in section 3.3.8.

Still mineral dust and sea salt contributions are much higher than speciation data (why carbonates are not included?), which authors should also ascribe (even if partially) to the 40% of undetermined mass.

We agree that differences between PMF and stoichiometric reconstructions based on speciation data are partially to be ascribed to the chemically undetermined mass (carbon components, including carbonates, and water), as discussed in section 3.4.2.

More in detail, the PMF mineral dust factor is about 30% higher than the mineral dust contribution estimated by the oxide formula. The oxide formula assumes that all the crustal elements are in the oxide form, while some of them, for example Ca or Mg may partially be in the carbonatic form (CaCO₃ instead of CaO, etc.), and this may produce an underestimation of the crustal component. However, we did not measure carbonates and we do not know how much of Ca or Mg are in the carbonatic form. This point is extensively discussed in section 3.4.2 (we have now modified a sentence to better explain it), and we also added a comment on carbonates at the end of section 3.2.

Possible reasons of the difference between the stoichiometric and the PMF sea salt are also discussed in section 3.4.2.

The identified factors are indeed realistic but the source contributions could be affected by large errors due to this fact. I strongly advice to carry out Bootstrapping, Displacement and BS-DISP analysis (EPA PMF5) to explore what are the errors in source contributions.

BS, DISP and BS-DISP are uncertainty methods implemented in the ME-2 and EPA PMF5 tools, but not in the PMF2 model that was used for this work. An explorative analysis performed with EPA PMF5 on the same input data set determined the same sources as presented in this study, although some small differences in the profiles occurred. A comparison of the outputs of different PMF tools is beyond the purposes of this paper, and it was faced in detail in recent papers (Belis 2015a, Belis 2015b and therein cited literature). Nevertheless, we agree with both the referees on the importance of an analysis of the uncertainties: in figure 3 we have now added the uncertainties on the profiles as provided by the PMF2 model. These estimates take into account the uncertainties on the input data and the application of non-negativity criteria, although they do not consider rotational ambiguity (Paatero et al., 2014). However, this aspect was investigated by systematically exploring solutions with FPEAK between -1.0 and 1.0, according to a literature approach (e.g., Vecchi et al., 2008). In the Supplementary Material, profiles of all the identified factors are reported for all the cases with variation in Q value below 5% (all bars are reported with the associated error computed by the PMF2 model). As concerns rotational ambiguity, in any case, it has to be pointed out that the choice of one out of all the rotated solutions is not determined merely by the mathematical outputs of the model: the choice is also driven by physical aspects as the physical sense of the source profiles and temporal evolutions, as well as physical constraints on the reconstruction of the profiles.

Finally, it is useful to recall that in case of large data sets, the importance of both random errors on the input and rotational uncertainty decreases (Paatero et al., 2014), when compared to modelling uncertainty (e.g., constancy of source profiles).

Before doing that I suggest also to explore the solution with the additional Ca-Sr factor. Soil of Lampedusa is rich in Calcite and Dolomite, and we are probably facing a local

resuspension source. Would be interesting to look at the daily contributions of this source (also vs the other crustal source).

We are aware that the soil of Lampedusa is rich in such minerals, and we had already explored the 8th factor, the “Ca-Sr” one, taking in consideration a local dust resuspension source. Nevertheless, we excluded this hypothesis due to the following reasons: this factor had a profile without any contribution from other main crustal elements (e.g., Al and Si); the temporal profile of this factor is very similar to the mineral dust one (in particular, it shows the same peaks during Saharan dust advection episodes); the scatter plots for the crustal elements (e.g., Ca, Si, Al...) do not show any cluster-pattern suggesting that some samples have different inter-elemental ratios and therefore a different mineral composition.

This is in agreement with previous preliminary results from this campaign suggesting that the composition of the Lampedusa soil does not relevantly differ from the average Saharan dust one.

Other minor corrections suggested:

-revise some typos (e.g. Mallorcal)

We are afraid we could not find “Mallorcal” in the text; nevertheless we have gone through the paper to correct as many mistyping as possible.

- some more details of sampling site in section 2.1 would be welcome, although they are discussed throughout the article

A couple of details were added in section 2.1.

-please provide brand of teflon filters

The information was added in section 2.1.

- the comparison of observed values with urban sites from literature might be misleading.

The idea underneath that comparison was to show that, despite the remoteness of the site, PM₁₀ in Lampedusa is characterized by not-negligible aerosol levels, mainly as concerns the secondary ones. This aspect is now explicitly reported in section 3.2.

- Why there are missing data only for Cu??

Actually, in the paper we report that we have missing data for other chemical species than Cu, although in a very low number. In particular, at page 20024, lines 23-24, we report the following: “Missing data were less than 1% of the total number of samples for all chemical species, except soluble elements (5%) and Cu (9%).” The percentage of missing data is higher for Cu due to a problem occurred while measuring a batch of samples with PIXE, resulting in a high Cu background signal that prevented the quantification of Cu in those samples.

- check the Mg oxidation number in page 20028

We have corrected it.

- Saharan dust plume may carry also emissions from Refineries

This information was added in section 3.3.7.

- *I could not visualize Figure 4*

We are sorry for the inconvenience. Figures were checked by the editorial board during the production process, we attach to this document a pdf of Figure 4.

- *The citation Kim et al., 2008 in page 20034 is incorrect. It should be Kim and Hopke, 2008*

We have corrected it.

Additional references not reported in the revised paper:

- Belis, C.A., Pernigotti, D., Karagulian, F., Pirovano, G., Larsen, B.R., Gerboles, M. and Hopke, P.K.: A new methodology to assess the performance and uncertainty of source apportionment models in intercomparison exercises, *Atm. Env.*, 119, 35-44, 2015a.

- Belis, C.A., Karagulian, F., Amato, F., Almeida, M., Argyropoulos, G., Artaxo, P., Beddows, D.C.S., Bernardoni, V., Bove, M.C., Carbone, S., Cesari, D., Contini, D., Cuccia, E., Diapouli, E., Eleftheriadis, K., Favez, O., El Haddad, I., Harrison, R.M., Hellebust, S., Jang, E., Jorquera, H., Kammermeier, T., Karl, M., Lucarelli, F., Mooibroek, D., Nava, S., Nøjgaard, J.K., Pandolfi, M., Perrone, M.G., Petit, J.E., Pietrodangelo, A., Pirovano, G., Pokorná, P., Prati, P., Prevot, A.S.H., Quass, U., Querol, X., Samara, C., Saraga, D., Sciare, J., Sfetsos, A., Valli, G., Vecchi, R., Vestenius, M., Yubero, E., Hopke, P.K.: A New Methodology to Assess the Performance and Uncertainty of Source Apportionment Models II: the Results of Two European Intercomparison Exercises, *Atm. Env.*, accepted for publication, 2015b.

Answers to the interactive comment on “Characterization of PM10 sources in the central Mediterranean” by Anonymous Referee #2

The manuscript reports the result of a quite long PM sampling campaign in the Lampedusa island. This is a peculiar site in a very good position to extract data representative of the Mediterranean basin. In general the article is well written, the results are sound and the methodologies clearly described.

As for reviewer #1, we thank for the positive evaluation of this work and we appreciate the productive comments and suggestions. These are addressed point-by-point in the following, recalling some answers already given to reviewer #1 when both reviewers pointed to the same issues.

I agree with Referee 1 when he/she notes on the possible impact of the missing information on carbonaceous aerosols and I also recommend a more deep analysis of the uncertainties both in the PMF results and in the chemical apportionment directly derived by raw concentration data (for instance: how reliable are the average correction factors used to obtain by PIXE concentration values for light elements? From sample to sample and according to the size distribution of the aerosol they could vary significantly). On the PMF side whatever is the used approach an evaluation on the uncertainty is missing and must be added.

As concerns the uncertainties on PMF results, please refer to the answer given to reviewer #1.

The correction factor for Na concentration as obtained by PIXE was estimated on the basis of PIGE (Particle Induced γ -ray Emission) measurements carried out for the quantification of Na on a subset of samples. PIGE does not suffer from self-absorption problems as it exploits the more energetic γ -ray radiation. Correction factors for the other light elements (Mg, Al, Si) were estimated accordingly on the basis of a simple attenuation model for single spherical particles (Formenti et al., 2010). As a consequence, the uncertainty on the PIXE quantification of light elements also bears a contribution due to these correction factors. These details were added in section 2.2. It has to be highlighted that, as reported in section 3.2, the redundancy of the data set, containing the determination of several elements by different techniques, allowed performing a quality assurance analysis of data (e.g., Na vs Na⁺, see table 1).

As concerns the uncertainties on the chemical apportionment directly derived by raw concentration data, in table 1 are now reported the standard deviations also for sea-salt and mineral dust. In this case, standard deviations mainly reflect the natural variability of the data, and not the analytical uncertainty. Under a strict analytical point of view, roughly a 15% uncertainty may be estimated for the determination of the sea-salt and mineral dust daily contributions, but this does not reflect neither the natural variability of data neither the uncertainty on the main assumption of the method (average composition of seawater and main oxides in the Earth's soil). It is important to note that this latter factor may be the main cause of uncertainty; limits of this method are discussed in the paper, mainly in section 3.4.2.

The comparison between sea salt concentration determined by PMF and by raw data and the discussion on the reason of the quite large discrepancy is questionable. The literature

on the PM composition in the open Mediterranean sea is quite limited but for instance in Schembari et al., Atmospheric Environment 98 (2014) 510-518 the PMF results for sea salt are quite different and some discussion/comment on this issue should be deserved (also because part of the Authors also signed that paper).

We have added a comment on the reported variability of sea salt factors across coastal sites in the Mediterranean sea in section 3.4.2, including the data cited by the reviewer, referring to the open Mediterranean sea.

B – LIST OF RELEVANT CHANGES IN THE MANUSCRIPT

The following relevant changes were made in the manuscript:

- The affiliation of the first author was corrected.
- Further information on the sampling site was added in section 2.1. Text was changed from “As previously mentioned, sampling was performed on Lampedusa (Italy), a small island located in the Central Mediterranean, far from continental pollution sources. More in detail, sampling was carried out at the Station for Climate Observations (35.5° N, 12.6° E, 45 m a.s.l., see Figure 1), maintained by ENEA (Italian Agency for New Technologies, Energy and Sustainable Economic Development)” to “As previously mentioned, sampling was performed on Lampedusa (Italy), a small island located in the Central Mediterranean, far from continental pollution sources (the nearest coast, in Tunisia, is more than 100 km away). More in detail, sampling was carried out on the plateau on the North-Eastern coast, at the Station for Climate Observations (35.5° N, 12.6° E, 45 m a.s.l., see Figure 1), maintained by ENEA (Italian Agency for New Technologies, Energy and Sustainable Economic Development)”.
- Brand of Teflon filter was added in section 2.1. Text was changed from “Particulate was sampled by a low volume sequential sampler (2.3 m³/h) equipped with a PM₁₀ inlet, on 47 mm diameter 2 µm nominal porosity Teflon filters.” to “Particulate was sampled by a low volume sequential sampler (2.3 m³/h) equipped with a PM₁₀ inlet, on 47 mm diameter 2 µm nominal porosity Pall Gelman Teflon filters.”
- Information on the correction of the PIXE concentrations of light elements was added in section 2.2. Text was changed from “The concentrations of lighter elements (Na, Mg, Al and Si) were corrected for self-absorption effects, i.e. lower energy X-rays attenuation inside aerosol particles (Calzolari et al., 2010, Formenti et al., 2010).” to ““The concentrations of lighter elements (Na, Mg, Al and Si) were corrected for self-absorption effects, i.e. lower energy X-rays attenuation inside aerosol particles. Correction factors were estimated on the basis of PIGE (Particle Induced γ-ray Emission) measurements carried out for the quantification of Na (Calzolari et al., 2010), using an attenuation model for single spherical particles (Formenti et al., 2010); the uncertainty on the PIXE quantification of light elements also bears the uncertainty on these correction factors (of the order of few percent).”
- Section 3.2 was modified to make clearer the purpose of the comparison with other data. Text was changed adding the sentences in red: “As concerns mineral dust, the observed concentrations are in the range observed for the crustal component in Mediterranean background sites (Querol et al., 2009), as well as consistent with the mean African dust contributions observed in Mediterranean urban and suburban sites (Pey et al., 2013), due to the character prevalently of long-range transport of this component. Despite the remoteness of the site, not-negligible secondary aerosol levels characterize the PM₁₀ in Lampedusa. Nitrate in Lampedusa accounted for 2.1 µg/m³ as average on the 2-year period: this value appears to be slightly higher than in other Mediterranean Basin background sites (1.7 – 1.9 µg/m³) and in rural and

urban background sites in Italy, and on the lower edge of those measured in Central Europe and UK ($2 - 4 \mu\text{g}/\text{m}^3$) (Querol et al., 2009)."

- The sentence "Nevertheless, some of the unaccounted mass may be partially attributed to carbonates, which were not assessed in this study." was added at the end of section 3.2, according to one of the comments of referee #1.
- Section 3.3 was changed adding more information on the choice of the number of factors and on the uncertainties, as requested by the referees. Added sentences are marked in red: "The eight-factor solution was also discarded, as it was identifying an unlikely source related primarily to Ca and Sr (44% and 21% explained variations, respectively): the hypothesis of facing a local dust resuspension source was discarded as the profile of this source did not contain any contributions from other crustal elements (such as Ca, Si and Al); further, the temporal profile of this factor is very similar to the mineral dust one (in particular, it shows the same peaks during Saharan dust advection episodes).

The seven sources are described in the following sections. Profiles are reported in Figure 3: the left y-axis displays the fraction of the elemental mass concentration, reported in the graph as blue column, whereas the right y-axis refers to the explained variation of each element, depicted as red squares (the contribution of one element in all the profiles sum up to one, when the variation of the element is completely reconstructed by the model). Error bars on the fraction of the elemental mass concentration represent the computed errors as provided by the PMF2 model. These estimates take into account the uncertainties on the input data and the application of non-negativity criteria; nevertheless, they do not consider rotational ambiguity (Paatero et al., 2014), which was in any case investigated by systematically exploring solutions with different FPEAK values (see Supplementary Material), according to a literature approach (Vecchi et al., 2008). The evolution of the daily contribution of each identified source is shown in Figure 4.

- Figure 3 was changed adding the error bars (and restyling it).
- Section 3.3.7 was completed adding the information reported by referee #1. Text was changed from "Several studies report that dust particles plumes from the Saharan area also export biomass-burning particles, mainly during the dry winter season (Tesche et al., 2011; Rodriguez et al., 2011)." to "Several studies report that dust particles plumes from the Saharan area also export biomass-burning particles, mainly during the dry winter season, as well as emissions from refineries (Tesche et al., 2011; Rodriguez et al., 2011)."
- In section 3.3.8, we included the results from a recent study on the effect of reducing input species in the application of PMF on an experimental data set providing an extended chemical speciation. The paragraph was changed from "As previously reported, additional data on the carbonaceous fractions in PM_{10} are available for limited periods in the years 2007-2008. In particular, 24 samplings of PM_{10} on quartz filters for EC/OC analysis were performed from 16 May to 15 August 2007 with integration times ranging from 24 to 72 h. The starting time of the sampling was set in the morning/early-afternoon (between 8 am and 2 pm) due to manual filter change operations. In Figure 7, the EC and OC contents are plotted together with the estimates of the contributions from the combustion

emission source as obtained by PMF. EC and OC are produced in combustion processes of both fossil fuels and biomasses, while OC may have also a biogenic origin. Therefore, EC and, to a smaller extent, OC are good markers of combustion sources, while additional information from other chemical tracers (such as K, levoglucosan, glycolates or radiocarbon (Bernardoni et al., 2013)) is needed for the assessment of the burnt fuel (biomass or fossil). As shown in Figure 7, EC and OC are highly correlated with the combustion emission contribution estimated by the PMF, thus strengthening the attribution of the seventh PMF factor to a combustion source. Small discrepancies, as the one occurring at the beginning of July, may be attributed to episodic important contributions to EC and OC from heavy oil combustion, that is from the primary particulate ship emission source, also reported in Figure 7.”

to “EC and OC are produced in combustion processes of both fossil fuels and biomasses, while OC may have also a biogenic origin. Therefore, EC and, to a smaller extent, OC are good markers of combustion sources, while additional information from other chemical tracers (such as K, levoglucosan, glycolates or radiocarbon (Bernardoni et al., 2013)) may help for the assessment of the burnt fuel (biomass or fossil). Despite OC is a major aerosol component and EC is an important combustion tracer, a recent study has demonstrated that the lack of data for these species does not significantly alter PMF results provided that a complete elemental and ionic speciation is available, so that other markers tracing combustion emissions are available (Lucarelli et al., 2015). Nevertheless, a comparison of the results for the combustion source as derived by the PMF analysis with the available EC/OC data was performed. As previously reported, additional data on the carbonaceous fractions in PM₁₀ are available for limited periods in the years 2007-2008. In particular, 24 samplings of PM₁₀ on quartz filters for EC/OC analysis were performed from 16 May to 15 August 2007 with integration times ranging from 24 to 72 h. The starting time of the sampling was set in the morning/early-afternoon (between 8 am and 2 pm) due to manual filter change operations. In Figure 7, the EC and OC contents are plotted together with the estimates of the contributions from the combustion emission source as obtained by PMF. As may be seen, EC and OC are highly correlated with the combustion emission contribution estimated by the PMF, thus strengthening the attribution of the seventh PMF factor to a combustion source. Small discrepancies, as the one occurring at the beginning of July, may be attributed to episodic important contributions to EC and OC from heavy oil combustion, that is from the primary particulate ship emission source, also reported in Figure 7.

The availability of a complete EC-OC data set for the future campaigns will add valuable information on the sources and contributions of the carbonaceous aerosol component.”

- The citation in section 3.4.1 “Kim et al., 2008” was corrected in “Kim and Hopke, 2008”.
- Section 3.4.2 was changed according to the comments of the referee #1 as follows: text was changed from “The mineral dust estimate from the PMF analysis is about 30% higher than the crustal content calculated as the metal oxides sum. It must be pointed out that no information on carbonates is available, and that the used

analytical techniques are not sensitive to the Oxygen content, thus the estimated metal oxide mass is based on the assumption of an oxygen fraction according to the prevalent oxides of the crustal elements. Nevertheless, PMF may slightly overestimate the source contribution as a consequence of some profile interferences (Mazzei et al., 2008): in this case, the PMF retrieval associated a small amount of SO_4^{2-} and Cl^- to the crustal dust profile, probably due to their co-linearity in case of long-range transport episodes.” to “The mineral dust estimate from the PMF analysis is about 30% higher than the crustal content calculated as the metal oxides sum. It must be pointed out that no information on carbonates is available, and that the used analytical techniques are not sensitive to the Oxygen content, thus the estimated **mineral dust** mass is based on the assumption of an oxygen fraction according to the prevalent oxides of the crustal elements. **This assumption is obviously an approximation as the real mineral dust composition may differ from it; as an example, this approach may underestimate the mineral dust contribution in presence of a significant amount of carbonates.** Nevertheless, PMF may slightly overestimate the source contribution as a consequence of some profile interferences (Mazzei et al., 2008): in this case, the PMF retrieval associated a small amount of SO_4^{2-} and Cl^- to the crustal dust profile, probably due to their co-linearity in case of long-range transport episodes.”.

- Further, according to the comment of referee #2, this paragraph was added at the end of section 3.4.2: “**In any case, it has to be pointed out that some variability in the chemical profiles of the sea-salt factors across the Mediterranean is reported in literature, due to possible reactions occurring in the atmosphere: for Mediterranean coastal sites, some papers report, as for this work, a sea-spray composition in agreement with the marine water one (e.g., Koçak et al., 2007; Amato et al. 2015); other papers report a Cl^- depletion (i.e. Cl^-/Na^+ ratios lower than in marine water) at coastal sites (Koulouri et al., 2008) as well as in open Mediterranean sea (Schembari et al., 2014).**”.
- In table 1, standard deviations for Dust and Sea Salt were added, and their units were changed from ng/m^3 to $\mu\text{g}/\text{m}^3$.
- References were added to the reference list according to the changes made to the text.

C – MARKED-UP MANUSCRIPT VERSION

1 **Characterization of PM₁₀ sources in the central Mediterranean**

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16 **Abstract**

17 The Mediterranean Basin atmosphere is influenced by both strong natural and
18 anthropogenic aerosol emissions, and is also subject to important climatic forcings. Several
19 programs have addressed the study of the Mediterranean basin; nevertheless important pieces
20 of information are still missing. In this framework, PM₁₀ samples were collected on a daily
21 basis on the island of Lampedusa (35.5° N, 12.6° E, 45 m a.s.l.), which is far from continental
22 pollution sources (the nearest coast, in Tunisia, is more than 100 km away). After mass
23 gravimetric measurements, different portions of the samples were analyzed to determine the

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1 ionic content by Ion Chromatography (IC), the soluble metals by Inductively Coupled Plasma
2 Atomic Emission Spectrometry (ICP-AES), and the total (soluble + insoluble) elemental
3 composition by Particle Induced X-ray Emission (PIXE). Data from years 2007 and 2008 are
4 used in this study.

5 The Positive Matrix Factorization (PMF) model was applied to the 2-year long data set of
6 PM₁₀ mass concentration and chemical composition to assess the aerosol sources affecting the
7 Central Mediterranean basin. Seven sources were resolved: sea-salt, mineral dust, biogenic
8 emissions, primary particulate ship emissions, secondary sulphate, secondary nitrate, and
9 combustion emissions. Source contributions to the total PM₁₀ mass were estimated to be about
10 40% for sea-salt, around 25% for mineral dust, 10% each for secondary nitrate and secondary
11 sulphate, and 5% each for primary particulate ship emissions, biogenic emissions, and combustion
12 emissions. Large variations in absolute and relative contributions are found and appear to depend
13 on the season and on transport episodes. In addition, the secondary sulphate due to ship
14 emissions was estimated, and found to contribute by about one third to the total sulphate
15 mass. Results for the sea-salt and mineral dust sources were compared with estimates of the
16 same contributions obtained from independent approaches, leading to an estimate of the water
17 content bound to the sea salt in the marine source.

18 *Keywords:* Mediterranean Basin, aerosol sources, PM₁₀, PMF, Central Mediterranean Sea.

1 **1. Introduction**

2 Atmospheric aerosols are estimated to have negative effects on human health and to play a
3 relevant role in climate, also affecting the hydrological cycle (IPCC). In this context, the
4 Mediterranean Basin is a complex area, due to the strong influences from both natural and
5 anthropogenic emissions, to the peculiar meteorology of the basin, and to conditions (high
6 solar irradiation, summer temperature, humidity) favouring secondary aerosol formation
7 processes (Pey et al., 2009; Im et al., 2012). Information on the aerosol sources and properties
8 in the Mediterranean area is essential for the determination of the impacts on the health of the
9 population living in the countries facing the Basin (among the world's most populated areas)
10 and on the climate of the area. This is particularly important because future climate
11 projections by IPCC consider the Mediterranean among the regions most sensitive to climate
12 change (IPCC).

13 Owing to its importance, Mediterranean atmospheric aerosol and meteorology have been
14 intensively studied (e.g., Salvador et al., 2014; Pey et al., 2013; Pace et al., 2006; Kopanakis
15 et al., 2012; Dayan et al., 1989) and several international programs have focused on the
16 assessment of chemistry and climate and their interactions, nevertheless important pieces of
17 information are still missing. For instance, the source contribution and the forcing of short-
18 lived species (both of them highly space and time variable) remain uncertain, and are not well
19 taken into account in climate simulation.

20 Without aiming at a thorough review of the published information, it is worthy to recall
21 that PM_{10} concentration is observed to increase from North-Western to South-Eastern
22 Mediterranean due to increasing contributions from both natural and anthropogenic sources
23 (Pey et al., 2013); the Western and Eastern Mediterranean basins have been observed to be
24 characterized by different atmospheric dynamical and chemical processes (e.g., neutralization
25 of secondary aerosol (Querol et al., 2009)). The Mediterranean boundary layer has been

1 observed to generally follow a clear annual behaviour, with minimum vertical extension
2 during summer and maximum altitudes in winter (with an opposite trend with respect to the
3 one commonly observed in continental areas) (Pey et al., 2009, Dayan et al., 1989; Pace et al.,
4 2012); therefore, stagnation of pollutants is favoured in summer over the Mediterranean
5 basin.

6 Most of the studies performed in the Mediterranean area are based on samplings carried
7 out at coastal sites, with strong influences from the continental meteorology and from
8 anthropogenic sources; thus, regional background and remote sites are essential for the study
9 of natural sources and to obtain information on the Mediterranean basin (Querol et al., 2009;
10 Pey et al., 2009; Koulouri et al., 2008; Pace et al., 2006). In this context, the island of
11 Lampedusa represents an ideal site: in fact, Lampedusa's environment is poorly affected by
12 anthropogenic pollution sources from local origin, due to its scarce population and
13 industrialization, and from abroad, as it is more than 100 km far from the nearest coast. As
14 discussed by (Henne et al., 2010), Lampedusa is one of the most remote sites with respect to
15 air quality measurements among 34 measurement stations throughout Europe. In addition,
16 Lampedusa is located in the Central Mediterranean Sea, in an area where few remote or
17 background sampling sites exist, despite the crucial position: in fact, as previously mentioned,
18 Western and Eastern Mediterranean Basin areas are known to be prone to different
19 atmospheric dynamics and pollution conditions.

20 In this work, an investigation aimed at assessing the aerosol sources affecting the Central
21 Mediterranean basin and at quantifying their contributions has been carried out. This study
22 answers to one of the main scientific question of the ChArMEx (Chemistry-Aerosol
23 Mediterranean Experiment) initiative. ChArMEx proposes to set up a coordinated
24 experimental and modelling international effort based on most updated tools, for an
25 assessment of the regional budgets of tropospheric trace species, of their trends, and of their

1 impacts on air quality, marine biogeochemistry, and regional climate. The source
2 apportionment here applied is based on the application of the Positive Matrix Factorization
3 (PMF) model to a 2-year long data set of PM₁₀ mass concentration and chemical composition
4 of samples collected on the Lampedusa Island. It is worth to recall that information on aerosol
5 sources and impact are among the data needed to constrain uncertainties in climate change
6 models, and they are also important for local administrations and policy makers for the
7 development of effective pollution abatement strategies. Studies on the source
8 assessment/apportionment of aerosols, often focused on specific sources such as African dust
9 or ship emissions, are available in literature for sites in the Western Mediterranean (e.g.,
10 Montseny (Cusack et al., 2013) and Palma de Mallorca (Pey et al., 2009) in Spain) and in the
11 Eastern Mediterranean (e.g., the Greek islands of Crete (Koulouri et al., 2008, Hildebrandt et
12 al., 2011) and Rhodes (Argyropoulos et al., 2012)). This work adds an important piece of
13 information still missing in the Central Mediterranean area.

14

15 **2. Materials and methods**

16 *2.1 Sampling site and instrumentation*

17 As previously mentioned, sampling was performed on Lampedusa (Italy), a small island
18 located in the Central Mediterranean, far from continental pollution sources (**the nearest coast,**
19 **in Tunisia, is more than 100 km away**). More in detail, sampling was carried out **on the**
20 **plateau on the North-Eastern coast**, at the Station for Climate Observations (35.5° N, 12.6° E,
21 45 m a.s.l., see Figure 1), maintained by ENEA (Italian Agency for New Technologies,
22 Energy and Sustainable Economic Development), where continuous observations of
23 greenhouse gases concentration (Artuso et al., 2009), total ozone (Gomez Amo et al., 2012),
24 ultraviolet irradiance (Meloni et al., 2005), aerosol properties (Pace et al., 2005, 2006; Di

1 Iorio et al., 2009; Marconi et al., 2014) and other climatic parameters (Di Biagio et al., 2009;
2 di Sarra et al., 2011) are routinely carried out.

3 Daily PM₁₀ samples have been collected at this station since 2004, every third day up to
4 2006 (Becagli et al., 2012; Marconi et al., 2014), and every day since 2007. Particulate was
5 sampled by a low volume sequential sampler (2.3 m³/h) equipped with a PM₁₀ inlet, on 47
6 mm diameter 2 µm nominal porosity **Pall Gelman** Teflon filters.

7 This paper focuses on the 2-years data set relative to the years 2007-2008, when PM₁₀ was
8 collected every day, accounting for a total of 562 samples (due to episodic technical failures,
9 some short sampling interruptions occurred in the period).

10 Additional samplings of PM₁₀ on quartz filters were performed in 2007, from 16 May to 15
11 August, using a low volume Tecora ECHO-PM sampler (2.3 m³/h), with collection times
12 ranging from 24 to 72h, for a total of 24 samples.

13

14 *2.2 Analyses*

15 The mass of the collected PM₁₀ was obtained by gravimetric measurements of the Teflon
16 filters before and after sampling. Before weighting, filters were conditioned for at least 24
17 hours in an environmentally controlled room (35-45% relative humidity, 25°C).

18 Afterwards, every Teflon filter was cut into three portions: a quarter was analyzed for the
19 ionic content by Ion Chromatography (IC), a second quarter for soluble metals by Inductively
20 Coupled Plasma Atomic Emission Spectrometer (ICP-AES); the remaining half filter was
21 analyzed for the total (soluble + insoluble) elemental composition by Particle Induced X-ray
22 Emission (PIXE).

23 More in detail, the quarter of Teflon filter devoted to ion chromatography was extracted in
24 about 10 mL (accurately weighted) of ultra-pure water (MilliQ water, resistivity>18 MΩ) in
25 ultrasonic bath for 15 min and analysed for the ionic content by using 3 Dionex ion

1 chromatographs fed by an auto-sampler. Cations (Na^+ , NH_4^+ , K^+ , Mg^{2+} , Ca^{2+}), inorganic
2 anions (F^- , Cl^- , NO_3^- , SO_4^{2-}), and some organic anions (methanesulphonate – MSA, acetate,
3 formate, glycolate, oxalate) were determined for each sample as reported in (Becagli et al.
4 2011). Detection limits (DLs) were in the range 0.08-0.16 ng/m^3 and reproducibility was
5 better than 5% .

6 Another quarter was extracted in ultrasonic bath for 15 min with MilliQ water acidified at
7 pH 1.5–2 with ultra-pure nitric acid obtained by sub-boiling distillation. This extract was used
8 for the determination of the soluble part of selected metals (Al, As, Ba, Cd, Cr, Cu, Fe, Mn,
9 Mo, Ni, Pb, V, Zn) with an Inductively Coupled Plasma Atomic Emission Spectrometer (ICP-
10 AES, Varian 720-ES) equipped with an ultrasonic nebulizer (U5000 AT+, Cetac
11 Technologies Inc.). The value of pH was chosen because it is the lowest value found in
12 rainwater (Li and Aneja, 1992) and therefore it leads to the determination of the metal
13 fraction more representative of the anthropogenic sources (Becagli et al., 2012; Traversi et al.,
14 2014).

15 PIXE measurements on the remaining half Teflon filters were performed at the INFN-
16 LABEC laboratory, equipped with a 3MV Tandatron accelerator, where research is focused in
17 the fields of cultural heritage (Castelli et al., 2013; Fedi et al., 2013), material science, and
18 atmospheric aerosols (Moreno et al., 2013, Calzolari et al., 2011). The PIXE technique is
19 based on the detection and analysis of the X-rays emitted by the sample after excitation by an
20 accelerated particle beam, and allows the quantification of $Z > 10$ elements. Briefly, samples
21 were bombarded by a 3.2 MeV proton beam ($i \sim 5 \text{ nA}$, $t \sim 300 \text{ s}$) using the set-up widely
22 described elsewhere (Calzolari et al., 2006; Lucarelli et al., 2011); spectra were fitted using the
23 Gupix software package (Campbell et al., 2010) and elemental concentrations were obtained
24 via calibration relative to thin reference standards. Detection limits (DLs) are in the 1-20
25 ng/m^3 range. Uncertainties were determined by a sum of independent uncertainties on

1 certified standard sample thickness (5%), peak areas (from 2 to 20% or higher when
2 concentrations approach DLs) and sampling parameters (of the order of few percent). The
3 concentrations of lighter elements (Na, Mg, Al and Si) were corrected for self-absorption
4 effects, i.e. lower energy X-rays attenuation inside aerosol particles. Correction factors were
5 estimated on the basis of PIGE (Particle Induced γ -ray Emission) measurements carried out
6 for the quantification of Na (Calzolari et al., 2010), using an attenuation model for single
7 spherical particles (Formenti et al., 2010); the uncertainty on the PIXE quantification of light
8 elements also bears the uncertainty on these correction factors (of the order of few percent).

9 As regards the additional samplings carried out in May-August 2007, the 24 samples
10 collected on quartz fibre filters were analyzed for total, elemental and organic carbon
11 fractions (TC, EC and OC) by a Sunset thermo-optical transmittance analyser following a
12 proxy of the IMPROVE protocol (He-580 in (Piazzalunga et al., 2011)).

13

14 *2.3 PMF analysis*

15 The Positive Matrix Factorization (PMF) is a receptor model allowing the determination of
16 the chemical profiles of the aerosol sources affecting the receptor site and the estimation of
17 their contributions to the aerosol load, using as input data the atmospheric concentrations of
18 different chemical species.

19 Briefly, PMF is an advanced factor analysis technique (Paatero, 1997; Paatero and Hopke,
20 2003) that uses realistic error estimates to weigh data values and imposes non-negativity
21 constraints in the computational process, which is based on a weighted least-squares
22 approach. With more detail, PMF model may be written as $X=G \cdot F+E$, where X is the $n \times m$
23 matrix of the m measured chemical species in n samples and G and F are factor matrices to be
24 determined: G is the $n \times p$ matrix of source contributions to the samples; F is the $p \times m$ matrix
25 of factor composition (source profiles); E is the residual matrix.

1 Input data were handled according to (Polissar et al., 1998): measured data were weighted
2 with their overall analytical uncertainty increased by one third of the detection limit (DL);
3 data below the DL were substituted with a value corresponding to DL/2, and a DL*5/6
4 uncertainty was associated to them. Missing data were handled filling the gaps with the
5 geometric mean value calculated over the days with data and associating to them a 400%
6 uncertainty.

7 Mass concentration was put in the PMF analysis as an internal variable, with very high
8 (400%) uncertainty (Norris et al., 2009), as it is suggested when portions of the mass (not
9 directly correlated with the other input variables) are not measured/available: in our case, no
10 data on the carbonaceous fraction is available for the whole considered period. As additional
11 test, PMF was also run with normalization of G and F matrices by multilinear regression
12 analysis on the PM₁₀ gravimetric mass concentrations and gave comparable results for both
13 contributions and profiles of the detected sources with respect to PMF runs performed with
14 internal mass, thus assuring that mass data did not influence the profiles.

15 PMF was run with the algorithm PMF2 (Paatero, 1997) in the robust mode, in order to
16 reduce the influence of outliers. Solutions with different numbers of factors were examined;
17 rotational ambiguity was tested by analyzing multiple FPEAK rotations; several seeds were
18 investigated and values of Q/Q_{exp} were examined (where Q is the weighted least-squares
19 function to be minimized).

20

21 **3. Results and discussion**

22 *3.1 PM₁₀ concentration*

23 The PM₁₀ daily concentration is reported in figure 2: a great variability was observed, with
24 values spanning one order of magnitude, as minima were found to be around 10 $\mu\text{g}/\text{m}^3$ and
25 several episodes reached more than 100 $\mu\text{g}/\text{m}^3$.

1 The average PM₁₀ concentration for the whole period (2007-2008) was $32 \pm 17 \mu\text{g}/\text{m}^3$
2 (median: $29 \mu\text{g}/\text{m}^3$, 25th percentile $21 \mu\text{g}/\text{m}^3$, 75th percentile $39 \mu\text{g}/\text{m}^3$); no relevant difference
3 was observed between the two years 2007 and 2008, with mean PM₁₀ concentrations of 33 ± 15
4 $\mu\text{g}/\text{m}^3$ and $32 \pm 19 \mu\text{g}/\text{m}^3$, respectively. These values are similar to those measured in two
5 different background/rural stations on the island of Crete, Greece (Kopanakis et al., 2012;
6 Gerasopoulos et al., 2006) and in the range observed for urban background sites across
7 Southern Europe (Karanasiou et al., 2014).

8 Despite the distance of the sampling site from pollution sources, concentrations exceeding
9 the daily European Union's limit value ($50 \mu\text{g}/\text{m}^3$) were registered in 24 days (out of 278
10 sampling days) in 2007, and in 32 (out of 284 sampling days) in 2008, i.e. for 10% of the
11 sampling period, being the highest concentrations due to Saharan dust events, as shown in
12 (Marconi et al., 2014). The EU legal term of not exceeding the daily limit value of $50 \mu\text{g}/\text{m}^3$ for
13 the percentile 90.4 of the annual values was not respected in 2008 (when percentile 90.4 was
14 $52.5 \mu\text{g}/\text{m}^3$) and nearly reached in 2007 (when percentile 90.4 was $48.9 \mu\text{g}/\text{m}^3$).

15

16 *3.2 PM₁₀ speciation*

17 Average concentrations (and standard deviation) of the main measured elements/species for
18 all the sampling period and in the different seasons are reported table 1. The determination of
19 several elements by different techniques allowed performing a quality assurance analysis of
20 data and studying the solubility of specific elements (Becagli et al., 2012), by comparing results
21 on the total content as obtained by PIXE and on the soluble content as obtained by IC or ICP-
22 AES.

23 Sea-salt and mineral dust contributions were estimated assuming the average composition
24 of seawater and Earth's soil (Mason, 1966). In particular, sea-salt concentration (i.e. [sea-salt])
25 was evaluated, for every day, according to the expression:

1 $[\text{sea-salt}] = [\text{Cl}^-] + [\text{ssNa}^+] + [\text{ssCa}] + [\text{ssMg}] + [\text{ssK}] + [\text{ssSO}_4^{2-}]$

2 where $[\text{Cl}^-]$ is the Cl concentration measured by IC and Na, Mg, K, Ca, and SO_4^{2-}
 3 contributions due to sea-salt ($[\text{ssNa}^+]$, $[\text{ssMg}]$, $[\text{ssK}]$, $[\text{ssCa}]$ and $[\text{ssSO}_4^{2-}]$) are estimated
 4 according to the following equations:

5 $[\text{ssNa}^+] = [\text{Na}^+] - [\text{nssNa}^+] = [\text{Na}^+] - [\text{nssCa}] \cdot (\text{Na/Ca})_{\text{crust}} = [\text{Na}^+] - ([\text{Ca}] - [\text{ssCa}]) \cdot$
 6 $(\text{Na/Ca})_{\text{crust}}$

7 $[\text{ssCa}] = [\text{ssNa}^+] \cdot (\text{Ca/Na})_{\text{seawater}}$

8 $[\text{ssMg}] = [\text{ssNa}^+] \cdot (\text{Mg/Na})_{\text{seawater}}$

9 $[\text{ssK}] = [\text{ssNa}^+] \cdot (\text{K/Na})_{\text{seawater}}$

10 $[\text{ssSO}_4^{2-}] = [\text{ssNa}^+] \cdot (\text{SO}_4^{2-}/\text{Na})_{\text{seawater}}$

11 where $[\text{Na}^+]$ and $[\text{Ca}]$ are the concentrations obtained respectively by IC and PIXE;
 12 $(\text{Na/Ca})_{\text{crust}}$ represents the mean Na/Ca ratio in the Earth crust, assumed to be 0.56 according to
 13 literature (Henderson and Henderson, 2009); $(\text{Ca/Na})_{\text{seawater}}$, $(\text{Mg/Na})_{\text{seawater}}$, $(\text{K/Na})_{\text{seawater}}$ and
 14 $(\text{SO}_4/\text{Na})_{\text{seawater}}$ are the mean elemental ratios in bulk seawater and are set to 0.038, 0.119,
 15 0.037, and 0.253, respectively (Henderson and Henderson, 2009).

16 The contribution of mineral dust ($[\text{mineral dust}]$) was calculated as sum of the metal oxides
 17 for every day based on stoichiometric ratios (e.g., Nava et al., 2012):

18 $[\text{mineral dust}] = 1.35 [\text{nssNa}^+] + 1.66 [\text{nssMg}] + 1.89 [\text{Al}] + 2.14 [\text{Si}] + 1.21 [\text{nssK}] + 1.40$
 19 $[\text{nssCa}] + 1.67 [\text{Ti}] + 1.43 [\text{Fe}]$

20 where “nss” stands for “non-sea-salt” (see previous equations) and $[\text{Al}]$, $[\text{Si}]$, $[\text{Ti}]$, $[\text{Fe}]$ are
 21 the elemental concentrations of Al, Si, Ti, Fe as measured by PIXE; the numerical coefficients
 22 account for the weight of the Oxygen atom in the main oxides of such elements in the average
 23 Earth crust.

24 According to this estimate, sea-salt and mineral dust contribute to PM_{10} in Lampedusa with
 25 $8.3 \mu\text{g}/\text{m}^3$ and $5.8 \mu\text{g}/\text{m}^3$ (corresponding to 26% and 18% of the PM_{10} mass) as average on the

1 2-year period, respectively. None of the components has a clear seasonality, and the temporal
2 pattern of mineral dust shows several episodic peaks. Sea salt levels are of the same order as
3 observed in other coastal Mediterranean sites. As concerns mineral dust, the observed
4 concentrations are in the range observed for the crustal component in Mediterranean
5 background sites (Querol et al., 2009), as well as consistent with the mean African dust
6 contributions observed in Mediterranean urban and suburban sites (Pey et al., 2013), due to the
7 character prevalently of long-range transport of this component.

8 Despite the remoteness of the site, not-negligible secondary aerosol levels characterize the
9 PM₁₀ in Lampedusa. Nitrate in Lampedusa accounted for 2.1 µg/m³ as average on the 2-year
10 period: this value appears to be slightly higher than in other Mediterranean Basin background
11 sites (1.7 – 1.9 µg/m³) and in rural and urban background sites in Italy, and on the lower edge of
12 those measured in Central Europe and UK (2 – 4 µg/m³) (Querol et al., 2009). Sulphate levels
13 (3.6 µg/m³ as average on the 2-year period, out of which 2.9 µg/m³ are nss-sulphate) are also in
14 the range reported for Central and Southern Europe and UK rural background stations (2 – 4
15 µg/m³), although lower than in background sites in the Eastern Mediterranean Basin (5 – 6
16 µg/m³), influenced by long range transport episodes from Eastern Europe (Querol et al., 2009).

17 MSA, a typical marker of biogenic activity, shows the typical seasonality with maxima
18 during summer and minima in winter due to the seasonality in the emission of its precursor
19 dimethylsulfide (DMS; Becagli et al., 2013).

20 As concerns the carbonaceous fraction of aerosol, data are available only for a very limited
21 time (24 samples averaging on 24 to 72 h in the period 16 May-15 August 2007): in such
22 period, on average, TC and EC accounted for 1.9 and 0.4 µg/m³, i.e. for 8.0% and 1.5% of
23 PM₁₀ mass, respectively. Preliminary results on the 2011 campaign on 1-year-long data set
24 estimate TC and EC levels as 9.6% and 0.1% of PM₁₀ mass, thus confirming the order of
25 magnitude of carbonaceous aerosol levels in Lampedusa as measured during the short 2007

1 campaign. In particular, EC appears to be a negligible component of PM₁₀ in Lampedusa, and
2 carbonaceous aerosol appears to be mainly organic; taking into account for non-C atoms in
3 particulate organic matter (POM) by multiplying OC for a factor 2.1 as suggested by (Turpin
4 and Lim, 2001) for nonurban sites, POM mass appears to contribute for about 20% of PM₁₀
5 mass in Lampedusa.

6 The reconstructed mass, i.e. the sum of all the components determined by chemical analysis,
7 corresponds to 61% of the gravimetric mass. When adding to this value the estimated 20% of
8 POM, the unaccounted mass usually attributed to water molecules is about 19%, in agreement
9 with the 10 to 20% range reported in literature for coastal rural and Mediterranean background
10 sites (Koçak et al., 2007; Pey et al., 2009; Querol et al., 2009). **Nevertheless, some of the**
11 **unaccounted mass may be partially attributed to carbonates, which were not assessed in this**
12 **study.**

13

14 *3.3 PM₁₀ sources identification by PMF*

15 The PMF model was applied to the data set of mass concentration and chemical
16 composition relative to the 562 samples collected in the years 2007-2008, considering 21
17 chemical species as measured by either PIXE (Mg, Al, Si, K, Ca, Ti, Mn, Fe, Cu, Br, Sr),
18 ICP-AES (V_{sol}, Ni_{sol}) or IC (Na⁺, NH₄⁻, Cl⁻, NO₃⁻, SO₄²⁻, Oxalates, Glycolate, MSA); all such
19 variables are “good”, according to the signal-to-noise (S/N) ratio criterion by (Paatero and
20 Hopke, 2003). The soluble fraction of V and Ni (V_{sol}, Ni_{sol}) was used instead of their total
21 fraction due to the lower DLs of ICP-AES compared to PIXE, and because it is a more
22 representative marker for anthropogenic sources (Becagli et al., 2012). Missing data were less
23 than 1% of the total number of samples for all chemical species, except soluble elements (5%)
24 and Cu (9%).

1 A seven-factor solution, with FPEAK value 0.1, was selected, on the basis of the Q value,
2 observed residual distributions, factor edges, physical sense of the source profiles and
3 temporal evolutions (F and G matrixes (Paatero, 1997)); observed relative ratios and
4 enrichment factors for the different sources are consistent with literature data, strengthening
5 the results of the analysis. Briefly, the analysis allowed the identification of 7 different
6 sources which are identified as sea-salt, mineral dust, biogenic emission aerosol, secondary
7 nitrate particles, secondary sulphate particles, primary particulate ship emissions, and
8 combustion emission aerosol (see the following sections). All together, these sources
9 reconstruct the measured gravimetric mass within 4%. The six-factor solution was excluded,
10 as it did not separate the secondary nitrate source from the combustion emission one. The
11 eight-factor solution was also discarded, as it was identifying an unlikely source related
12 primarily to Ca and Sr (44% and 21% explained variations, respectively): the hypothesis of
13 facing a local dust resuspension source was discarded as the profile of this source did not
14 contain any contributions from other crustal elements (such as Ca, Si and Al); further, the
15 temporal profile of this factor is very similar to the mineral dust one (in particular, it shows
16 the same peaks during Saharan dust advection episodes).

17 The seven sources are described in the following sections. Profiles are reported in Figure 3:
18 the left y-axis displays the fraction of the elemental mass concentration, reported in the graph
19 as blue column, whereas the right y-axis refers to the explained variation of each element,
20 depicted as red squares (the contribution of one element in all the profiles sum up to one,
21 when the variation of the element is completely reconstructed by the model). Error bars on the
22 fraction of the elemental mass concentration represent the computed errors as provided by the
23 PMF2 model. These estimates take into account the uncertainties on the input data and the
24 application of non-negativity criteria; nevertheless, they do not consider rotational ambiguity
25 (Paatero et al., 2014), which was in any case investigated by systematically exploring

1 solutions with different FPEAK values (see Supplementary Material), according to a literature
2 approach (Vecchi et al., 2008). The evolution of the daily contribution of each identified
3 source is shown in Figure 4.

5 3.3.1 *Sea-salt*

6 Na⁺, Cl⁻, Mg, K and Br are the characterizing elements for this factor (figure 3, panel a);
7 most of their mass concentration is explained by this source (explained variations: 64%, 83%,
8 48%, 30% and 61% respectively). Some SO₄²⁻ is also present. The elemental ratios are in good
9 agreement with those measured in marine water, suggesting a common marine origin for the
10 species characterizing this factor. More in detail, the source profile elemental ratios are reported
11 in Table 2 together with the sea salt literature elemental ratio (Henderson and Henderson,
12 2009). Thus, this factor of the PMF is identified with the primary marine aerosol, sea salt
13 aerosol.

14 The temporal evolution shows no clear seasonality, as it can be expected for this source,
15 since the sampling site is located directly on the coast and sea-salt aerosol is known to be
16 strongly related to wind intensity (Kishcha et al., 2011).

18 3.3.2 *Mineral dust*

19 Several crustal elements such as Al, Si, Ti, Mn, Fe, Sr, Ca, K and Mg characterize this factor
20 (figure 3, panel b). This source reconstructs almost completely the mass concentration of some
21 of these elements, such as Al, Si and Ti (more than 80% of their signal).

22 Enrichment factors (EF) with respect to Al in this profile were calculated using the average
23 Earth's soil composition reported by (Mason, 1966) and are reported in Table 3.

1 The values of these EFs, all close to 1 (with a moderate enrichment only for Ca and Sr which
2 is usually associated with Ca in carbonates), strengthen the identification of this source as
3 “mineral dust”.

4 The temporal evolution of this factor shows no clear seasonality, being dominated by many
5 narrow peaks. Sixty-hour backward trajectories ending at Lampedusa at 50 m a.s.l.
6 corresponding to the start, middle, and end time of the sampling interval were calculated for the
7 peaks in PM₁₀ concentrations due to mineral dust larger than the 95° percentile (about 32
8 µg/m³), using the NOAA Air Resource Laboratory HYSPLIT transport model (Draxler and
9 Rolph, 2012). The analysis showed that in about 86% of the cases at least one of the three
10 trajectories associated to the sampling overpassed Northern Africa (see Figure 5). This result
11 should be read taking into account the uncertainties in the HYSPLIT trajectories due to the lack
12 of meteorological measurements in the examined region, the low model horizontal resolution,
13 and the possible rapid temporal variation of the air masses during the sampling interval. No
14 relevant difference in the available crustal markers ratios was observed in correspondence of
15 African dust transport episodes, suggesting a similar chemical composition for the Lampedusa
16 soil and the African dust (limitedly to the measured elements), or a negligible impact of local
17 dust. In fact, enrichments in Ca with respect to the average Earth’s soil composition are reported
18 for dust originating from different African areas, e.g. Western Sahara and Moroccan Atlas, due
19 to the high content of minerals such as calcite and dolomite (Avila et al., 1997; Coz et al.,
20 2009). An enhancement of Ca amounts for dust originating from the Algeria-Tunisia sector
21 was shown by (Marconi et al., 2014).

22 The seasonality of African desert dust advection episodes in Lampedusa was discussed in a
23 recent study (Marconi et al., 2014) based on a larger data-set (June 2004 – December 2010), in
24 which ground level aerosol concentrations are compared with aerosol optical depth
25 measurements routinely performed at the Station for Climate Observations of Lampedusa,

1 providing information on the dust content in the entire air column above the sampling site.
2 African desert dust transports often occur above the boundary layer (mainly in summer); in fact,
3 despite the strong seasonality observed for the aerosol optical depth, at the ground level no
4 evident seasonal trend was observed for PM₁₀ or crustal markers, in agreement with the mineral
5 dust temporal pattern obtained by PMF on the 2007-2008 data-set. In particular, as shown by
6 (Marconi et al., 2014), while the aerosol optical depth and the frequency of occurrence of
7 vertically integrated dust events peak in summer, a relative minimum of PM₁₀ concentration and
8 its frequency of occurrence is found in the months of June-July.

9

10 *3.3.3 Biogenic emissions*

11 This factor is mainly characterized by MSA (Methanesulfonic Acid), and explains almost
12 completely (87%) its mass concentration. A small percentage (6%) of the total mass of SO₄²⁻ is
13 also accounted by this source, and constitutes however the prevalent absolute mass contribution
14 to this factor.

15 MSA is used as a tracer for biogenic sulphur emissions (from phytoplankton processes), as it
16 is solely produced by oxidation of dimethylsulfide (DMS). The evaluation of the nssSO₄⁻
17 emitted by the biogenic source, (nss-SO₄²⁻)_{bio}, suffers by the variability of the (nss-SO₄²⁻)_{bio}
18 /MSA emission ratio as this depends on the latitude, the NO_x levels and the temperature
19 (Seinfeld and Pandis, 1998; Bates et al., 1992); this is especially true for the Mediterranean
20 basin, where only a maximum value for such ratio has been estimated in previous studies (in
21 Crete) (Mihalopoulos et al., 1997; Kouvarakis and Mihalopoulos, 2002). The observed SO₄²⁻
22 /MSA ratio in this source, 11.3, is only slightly higher than the aforementioned estimated
23 maximum value (10) and it could be considered as representative for the (nss-SO₄²⁻)_{bio}/MSA
24 ratio for the biogenic emissions in the Central Mediterranean basin. By using the value of 11.3
25 we found that in summer (JJA), when biogenic sulphur emission is at its maximum, 0.5 µg m⁻³

1 of SO_4^{2-} , corresponding to 11% of total SO_4^{2-} , arise from a natural source. This result confirms
2 the order of magnitude of the biogenic sulphur contribution during summer estimated as 17%
3 by (Kouvarakis and Mihalopoulos, 2002) by observations and modelling.

4 The strong seasonality observed for this factor, with minima during winter and maxima in
5 summer, is also supporting its attribution to the biogenic emissions source. A recent study
6 investigated the role of biogenic aerosols in the PM at Lampedusa (Becagli et al., 2013) with
7 respect to the phytoplankton productivity, suggesting that the seasonal evolution of
8 methanesulfonate, with a very evident summer maximum, is mainly driven by phytoplankton
9 physiology and cell stress factors, such as high irradiance and shallow depth of the marine
10 upper mixed layer.

11 12 *3.3.4 Secondary nitrate*

13 This factor is characterized by NO_3^- and Na^+ , with the mass concentration of NO_3^- being
14 explained almost for the 80% by it.

15 The ionic balance for this source is neutral, when taking into account NO_3^- and the two
16 cations Na^+ and Mg^{2+} , as in a marine environment the neutralization of HNO_3 can occur to a
17 large extent on sea-salt components (Bardouki et al., 2003; Metzger et al., 2006).

18 The temporal evolution shows a small seasonality, with somewhat larger values during
19 summer. A similar seasonal behaviour was observed by (Querol et al., 2009) in Eastern
20 Mediterranean Basin (EMB) sites. The same study evidenced that in the EMB nitrate is mainly
21 present in the coarse mode, chemically bound to alkaline ion species, while in the Western
22 Mediterranean Basin (WMB) it is largely in the fine mode, as it is most frequently present as
23 ammonium nitrate. In Lampedusa, as in the EMB, the observed seasonality appears to be
24 mainly regulated by the combination of nitrates with sea-salt cations (Metzger et al., 2006),
25 while in the WMB nitrate is dominated by a strong opposite seasonality, with maxima during

1 winter, due to both the instability of ammonium nitrate in warm environments and intense
2 nitrate episodes in late winter.

3

4 *3.3.5 Secondary sulphate*

5 NH_4^+ , SO_4^{2-} , and Oxalates are the characterizing elements for this source, with most of their
6 signals explained by this factor (explained variations are 87%, 54%, and 45%, respectively).
7 The major absolute contribution to the profile is given by SO_4^{2-} ; the $\text{SO}_4^{2-}/\text{NH}_4^+$ ratio of 3.4
8 indicates that both the compounds $(\text{NH}_4)\text{HSO}_4$ and $(\text{NH}_4)_2\text{SO}_4$ are present.

9 The temporal evolution shows maxima during summer, when the marine boundary layer is
10 more stable and the photochemistry leading to the production of secondary aerosols is
11 enhanced. The same behaviour was observed by (Querol et al., 2009) for sulphate both at
12 Eastern and Western Mediterranean sites. Further, it is worthy to note that sulphate
13 neutralization with NH_4^+ may leave very low amounts of NH_3 available for the neutralization of
14 nitrate, which thus occurs favourably on sea-salt aerosol (see 3.2.4).

15

16 *3.3.6 Primary particulate ship emissions*

17 This factor explains most of the mass of V_{sol} and Ni_{sol} , even if the predominant mass
18 contributions are from Na^+ and SO_4^{2-} . V and Ni have been clearly recognized as markers of
19 heavy oil combustion, and characteristic V/Ni ratios ranging from 2.5 to 3.5 have been reported
20 for ships emissions (Mazzei et al., 2008; Viana et al., 2009; Pandolfi et al., 2011). Further, both
21 V and Ni were found to have roughly the same solubility (80% and 77%, respectively) when
22 originated by heavy oil combustion, in the mild extraction condition (HNO_3 – pH 1.5) used for
23 ICP-AES extraction (Becagli et al., 2012). The $V_{\text{sol}}/\text{Ni}_{\text{sol}}$ ratio for this source is 3.40, in good
24 agreement with the aforementioned literature data. Moreover, the interpretation of this factor as
25 “primary ship emissions” is strengthened by the fact that the observed $\text{SO}_4^{2-}/V_{\text{sol}}$ ratio (6.2 w/w;

1 as V has been found to have a 80% solubility in ships exhausts, such value corresponds to a
2 $\text{SO}_4^{2-}/\text{V}$ ratio of 5.0) is of the same order of magnitude as the ones reported by (Agrawal et al.,
3 2008) for the particulate matter freshly emitted from the main engine of an ocean going
4 container vessel ($\text{SO}_4^{2-}/\text{V}$ ranging from 8.9 to 11.9 depending on the engine operating
5 conditions for the considered vessel). Much higher $\text{SO}_4^{2-}/\text{V}$ values are reported when secondary
6 aerosols (mainly SO_4^{2-}) are formed from the oxidation of the gaseous precursors (e.g., SO_2)
7 (Becagli et al., 2012).

8 A clear seasonal evolution, with maxima during summer, is evident for this source; this is in
9 agreement with the temporal evolution of ship emissions markers (Ni_{sol} , V_{sol} , nssSO_4^{2-})
10 observed by (Becagli et al., 2012) on a larger data set of samples collected at the same
11 Lampedusa site (years 2004-2008). The authors explained the temporal evolution of these
12 markers as due to three possible causes: a) an increased photochemical activity in summer
13 leading to a faster production of secondary aerosols; b) a stronger stability of the marine
14 boundary layer during summer, leading to an increase of their concentrations at the ground
15 level; c) prevalent winds from the Strait of Sicily (one of the main ship paths) during summer.
16 Due to the primary origin of this factor, such seasonal behaviour appears to be more probably
17 linked to the two latter phenomena.

18

19 *3.3.7 Combustion emissions*

20 Several compounds, as K, Cu, NO_3^- , SO_4^{2-} and short carboxylic acids as Oxalates and
21 Glycolates, whose signal is mostly explained by this source, characterize this factor. K and
22 SO_4^{2-} are commonly associated with biomass burning, and short carboxylic acids have been
23 recently associated with coal/waste/biomass burning (Wang et al., 2007). Enhancements of
24 Cu in the aerosol produced by wood combustion have also been reported by several studies

1 (Dall'Osto et al., 2013; Nava et al., 2015). Therefore this factor has been interpreted as a
2 generic combustion source, which includes biomass burning.

3 The temporal evolution (Figure 4) of this factor shows no clear seasonality, even if
4 important contributions during summer may be observed and may be linked to the more
5 frequent occurrence of fires in the Mediterranean region. During spring and autumn the
6 contributions from this source are generally low, while several high contribution days may be
7 observed during winter. Figure 6 shows the combined evolution of the desert dust and
8 combustion emissions contributions to PM₁₀, and some selected cases with air mass
9 trajectories and satellite observations of active fires. Most of the winter cases characterized
10 by elevated contributions of combustion sources display also a high contribution from the
11 mineral dust source (see Figure 6). The analysis of the backtrajectories shows that these days
12 were mostly interested by African dust advection episodes. Several studies report that dust
13 particles plumes from the Saharan area also export biomass-burning particles, mainly during
14 the dry winter season, **as well as emissions from refineries** (Tesche et al., 2011; Rodriguez et
15 al., 2011). Further, small fires, often occurring in biomes as croplands, wooded savannas and
16 tropical forests, are usually below the detection limit of the current generation of surface
17 reflectance/thermal imagery instruments (Randerson et al., 2012). Therefore, winter
18 contributions from this source are likely to be mainly due to small scale biomass burning
19 activity in north Africa; nonetheless, some winter peaks of this source appear also connected
20 with fires and/or biomass burning activity in Central-Eastern Europe (e.g., 13-15/3/2007,
21 Figure 6).

22 Summer peaks have been clearly connected by backtrajectory analysis to fires mainly in
23 South Italy and Sicily, thus supporting the interpretation of this factor as a combustion source,
24 with a main contribution from biomass burning.

25

3.3.8 Combustion emission source identification: additional data

EC and OC are produced in combustion processes of both fossil fuels and biomasses, while OC may have also a biogenic origin. Therefore, EC and, to a smaller extent, OC are good markers of combustion sources, while additional information from other chemical tracers (such as K, levoglucosan, glycolates or radiocarbon (Bernardoni et al., 2013)) may help for the assessment of the burnt fuel (biomass or fossil). Despite OC is a major aerosol component and EC is an important combustion tracer, a recent study has demonstrated that the lack of data for these species does not significantly alter PMF results provided that a complete elemental and ionic speciation is available, so that other markers tracing combustion emissions are available (Lucarelli et al., 2015). Nevertheless, a comparison of the results for the combustion source as derived by the PMF analysis with the available EC/OC data was performed. As previously reported, additional data on the carbonaceous fractions in PM₁₀ are available for limited periods in the years 2007-2008. In particular, 24 samplings of PM₁₀ on quartz filters for EC/OC analysis were performed from 16 May to 15 August 2007 with integration times ranging from 24 to 72 h. The starting time of the sampling was set in the morning/early-afternoon (between 8 am and 2 pm) due to manual filter change operations. In Figure 7, the EC and OC contents are plotted together with the estimates of the contributions from the combustion emission source as obtained by PMF. As may be seen, EC and OC are highly correlated with the combustion emission contribution estimated by the PMF, thus strengthening the attribution of the seventh PMF factor to a combustion source. Small discrepancies, as the one occurring at the beginning of July, may be attributed to episodic important contributions to EC and OC from heavy oil combustion, that is from the primary particulate ship emission source, also reported in Figure 7.

The availability of a complete EC-OC data set for the future campaigns will add valuable information on the sources and contributions of the carbonaceous aerosol component.

1

2 *3.4 PM₁₀ source apportionment*

3 Figures 8 and 9 report the relative and absolute seasonal contributions of the seven
4 identified sources (winter – December, January, February; spring – March, April, May;
5 summer – June, July, August; fall – September, October, November). During all the seasons,
6 natural sources (seas-salt, mineral dust, and biogenic emissions) give the largest contribution
7 to PM₁₀, accounting for a minimum of 62% in spring to a maximum of 76% during fall. The
8 maximum relative and absolute contribution of sea-salt is observed in winter, while relative
9 contributions during the other seasons are comparable. This is consistent with a higher wind
10 velocity during the winter months (7.4 m/s, versus 5.6 m/s in summer, 6.5 m/s in spring, and
11 6.1 m/s in fall during 2007 and 2008).

12 The mineral dust maximum relative contribution is found during fall (37%), and is around
13 20% in the other seasons. As expected, the maximum relative contribution of biogenic
14 emissions is during summer, and is negligible in winter.

15 Among the anthropogenic sources, primary particulate ship emissions and secondary
16 sulphate display a clear and similar seasonality, with maxima in summer. No significant
17 dependence on the season is observed for the relative contributions of secondary nitrate and
18 combustion emissions. It is worth to note that the not explained mass is around 10% across
19 the seasons.

20 The annual average relative contributions are reported with their uncertainties in Figure 10:
21 the main contribution to aerosol comes from sea-salt, accounting for 40% of the mass; the
22 second contribution is from mineral dust (25%). Taking into account also the biogenic
23 emissions, accounting for 4% of the mass, the contribution of natural sources to the PM₁₀ in
24 Lampedusa is estimated to be around 68% on average. As concerns secondary nitrates and
25 sulphates, they are estimated to contribute by 9% and 11% to the total PM₁₀, respectively.

1 Finally, primary particulate ship emissions have been observed to give an average 4%
2 contribution, while 5% of PM₁₀ was found to be due to combustion emissions. Uncertainties
3 on the contributions were estimated propagating the uncertainties obtained by the PMF model
4 on the G matrix.

5 In absolute terms, the following contributions were estimated: sea-salt, $12.9 \pm 1.1 \mu\text{g}/\text{m}^3$;
6 mineral dust, $7.8 \pm 0.5 \mu\text{g}/\text{m}^3$; biogenic emissions, $1.1 \pm 0.2 \mu\text{g}/\text{m}^3$; secondary nitrate, $2.8 \pm$
7 $0.3 \mu\text{g}/\text{m}^3$; secondary sulphate, $3.4 \pm 0.4 \mu\text{g}/\text{m}^3$; primary particulate ship emissions, 1.2 ± 0.2
8 $\mu\text{g}/\text{m}^3$; combustion emissions, $1.5 \pm 0.3 \mu\text{g}/\text{m}^3$.

9

10 *3.4.1 Evaluation of the total contribution from ship emissions to PM₁₀ in Lampedusa*

11 An estimate of the contribution of the shipping emissions to the secondary sulphate
12 formation was performed following previous studies (Kim and Hopke, 2008; Viana et al.,
13 2009), based on the observed correlation between the daily contributions from the primary
14 particulate ship emission source and those from the secondary sulphate source as assessed by
15 PMF analysis. The contributions from these two sources are plotted one against the other in
16 Figure 11: the solid line represents the minimum amount of secondary sulphate that is
17 expected to be associated to the observed primary particulate from ship emissions. Such line
18 was graphically drawn aiming at excluding outliers and its slope indicates that, at the receptor
19 site, at least $0.9 \mu\text{g}/\text{m}^3$ of secondary sulphate are expected per unit of detected primary
20 particulate from ship emissions. Therefore, primary particles contribute by 50% or less to the
21 total PM₁₀ originating from ship emissions. (Kim et al., 2008) reported about $0.8 \mu\text{g}/\text{m}^3$ of
22 secondary sulphate per $1 \mu\text{g}/\text{m}^3$ of primary oil combustion particles in PM_{2.5} in 3 sites in
23 Seattle (WA, USA). A larger ratio (about 2.1) was found by (Viana et al., 2009) for PM₁₀ in
24 Melilla (Spain). Both oil combustion particles and secondary aerosols are found mainly in the
25 fine aerosol fraction, therefore the ratio between the two components (secondary sulphate and

1 primary particulate) of the ship emissions source is likely to be roughly the same for PM_{2.5}
2 and PM₁₀. It must be pointed out however that we retrieved a minimum value, and the actual
3 contribution of secondary sulphate depends on many factors and, in many cases, may be
4 much higher than the limit.

5 Taking into account this minimum secondary sulphate contribution, we derive a minimum
6 SO₄²⁻/V_{sol} ratio of about 190 for ship emissions (primary particulate + secondary sulphate). It
7 is worthy to note that a SO₄²⁻/V_{sol} ratio = 200 was proposed as characteristic lower ratio for
8 ship emissions from a previous study performed at Lampedusa (Becagli et al., 2012) on a more
9 extended dataset (June 2004 – December 2008).

10 Further, ship emissions on the whole (primary particulate + secondary sulphate) are
11 estimated to contribute to PM₁₀ in Lampedusa by approximately 7% (about 4% primary
12 particles and 3% secondary sulphate). These values confirm the order of magnitude of the
13 estimated impact of ship emissions in European coastal areas, which is reported to be around
14 1–7% (EEA techn. report, 2013).

15

16 *3.4.2 Comparison of contribution estimates by PMF and speciation*

17 The PMF estimates of the contributions from sea-salt and mineral dust were compared
18 with those obtained according to the average composition of seawater and Earth's soil
19 (paragraph 3.2).

20 The mineral dust estimate from the PMF analysis is about 30% higher than the crustal
21 content calculated as the metal oxides sum. It must be pointed out that no information on
22 carbonates is available, and that the used analytical techniques are not sensitive to the Oxygen
23 content, thus the estimated **mineral dust** mass is based on the assumption of an oxygen fraction
24 according to the prevalent oxides of the crustal elements. **This assumption is obviously an**
25 **approximation as the real mineral dust composition may differ from it; as an example, this**

1 approach may underestimate the mineral dust contribution in presence of a significant amount
2 of carbonates. Nevertheless, PMF may slightly overestimate the source contribution as a
3 consequence of some profile interferences (Mazzei et al., 2008): in this case, the PMF retrieval
4 associated a small amount of SO_4^{2-} and Cl^- to the crustal dust profile, probably due to their co-
5 linearity in case of long-range transport episodes.

6 Coming to sea-salt, the stoichiometric approach gives an estimate of the average contribution
7 of $8.3 \mu\text{g}/\text{m}^3$, to be compared with the PMF estimated contribution for the sea-salt source of
8 $12.9 \mu\text{g}/\text{m}^3$. Since all the sea-salt elements are measured, this discrepancy is likely to be due to
9 other compounds not visible with the applied analytical techniques, such as water or, to a lesser
10 extent, organic compounds. In fact, although gravimetric measurements are performed in a
11 controlled atmosphere (50% relative humidity, 25°C), some water may endure and it is likely
12 to be bound to hygroscopic aerosol components such as sea-salt; this is in agreement with
13 recent studies that have shown water exists in ambient aerosol also at low relative humidity
14 (Plinis et al., 2014; Engelhart et al., 2011). Therefore, taking into account for water, the PMF
15 contribution from the sea-salt source may be possibly split in two contributions, namely “dry
16 sea-salt” and “ H_2O ”, accounting for about 26% and 14% of the PM_{10} mass in Lampedusa,
17 respectively. This estimate lays in the 10 to 20% range reported in literature for unaccounted
18 mass in Mediterranean background/rural sites (Koçak et al., 2007; Pey et al., 2009; Querol et al.,
19 2009), which is generally attributed to water molecules of moisture.

20 In any case, it has to be pointed out that some variability in the chemical profiles of the sea-
21 salt factors across the Mediterranean is reported in literature, due to possible reactions occurring
22 in the atmosphere: for Mediterranean coastal sites, some papers report, as for this work, a sea-
23 spray composition in agreement with the marine water one (e.g., Koçak et al., 2007; Amato et
24 al. 2015); other papers report a Cl^- depletion (i.e. Cl^-/Na^+ ratios lower than in marine water) at

1 coastal sites (Koulouri et al., 2008) as well as in open Mediterranean sea (Schembari et al.,
2 2014).

4 **4. Conclusions**

5 On average, each of the sources “primary particulate ship emissions”, “biogenic emissions”
6 and “combustion emissions” contributes by about 5% to the total PM₁₀ mass; “secondary nitrate”
7 and “secondary sulphate” account for about 10% each; the “mineral dust” contribution is around
8 25%, while “sea-salt” constitutes about 40% of the PM₁₀. According to a stoichiometric approach,
9 the latter may possibly split in two contributions, namely “dry sea-salt” and “H₂O”, accounting
10 for roughly 25% and 15% of the PM₁₀ mass in Lampedusa, respectively. Further, about one third
11 of the total contribution of secondary sulphate appears to originate from ship emissions. Large
12 variations in absolute and relative contributions are found and appear to depend on the season and
13 on transport episodes.

14 Improvements to this source apportionment will be possible taking into account for the
15 carbonaceous fraction (no extensive dataset is available for it for the considered years, 2007 and
16 2008).

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23 multidisciplinary program MISTRAL.

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1 TABLES

	Year		Winter (DJF)		Spring (MAM)		Summer (JJA)		Autumn (SON)	
n	562		136		156		142		128	
	ng/m ³		ng/m ³		ng/m ³		ng/m ³		ng/m ³	
	Av	Sd								
Na	3137	1996	3584	1972	3293	1937	2953	2108	2686	1860
Mg	374	187	377	171	405	201	345	190	367	177
Al	360	549	234	432	347	570	339	446	530	682
Si	896	1389	581	1106	889	1435	830	1097	1308	1756
S	1534	806	1051	387	1644	762	2098	904	1281	634
Cl	4708	3725	5727	3945	5263	3783	3592	3663	4212	3052
K	314	223	267	190	319	238	324	198	346	256
Ca	862	1120	694	916	922	1277	761	964	1077	1237
Ti	30	46	20	35	30	49	29	41	41	56
V	5.2	5.4	2.8	3.0	6.1	5.6	8.4	6.5	3.2	3.3
Mn	5.3	7.1	3.5	5.5	5.7	7.7	5.5	6.1	6.5	8.4
Fe	304	445	200	333	316	471	290	379	415	547
Ni	2.0	1.7	1.2	1.0	2.4	1.9	3.1	1.9	1.3	1.0
Cu	1.8	1.3	1.3	1.4	1.7	1.3	2.5	1.2	1.5	1.1
Br	15	11	20	13	15	10	13	11	12	9
Sr	7	8	6	5	8	10	6	7	7	8
Na ⁺	3191	2191	3481	2230	3629	2352	2912	2211	2670	1751
NH ₄ ⁺	673	452	477	313	757	481	954	441	463	340
K ⁺	161	88	153	84	172	82	183	103	131	69
Ca ²⁺	615	717	475	433	670	859	532	508	784	905
Cl ⁻	4304	3743	5111	3807	4924	4019	3278	3646	3829	3072
NO ₃ ⁻	2128	1148	1866	958	2434	1232	2283	1218	1861	1028
SO ₄ ²⁻	3652	2074	2478	1110	4011	2059	4961	2276	3010	1674
Ossal.	144	85	101	58	151	83	194	80	126	88
Glyc.	6	5	5	4	8	6	7	6	6	4
MSA	22	23	4	4	22	19	46	25	14	14
V _{sol}	4.0	4.3	2.1	1.7	4.1	4.2	7.4	5.6	2.3	1.7
Mn _{sol}	3.1	3.9	1.9	3.9	3.3	4.2	3.4	3.0	3.9	4.2
Ni _{sol}	1.4	1.5	0.7	0.5	1.6	1.6	2.5	1.8	0.8	0.6
Cu _{sol}	1.2	1.1	0.7	1.1	1.3	1.3	1.8	0.9	0.9	0.8
AS _{sol}	0.1	0.1	0.1	0.1	0.2	0.2	0.2	0.1	0.1	0.1
Pb _{sol}	2.9	2.9	2.3	2.9	3.5	2.4	3.4	2.5	2.5	3.4
	µg/m ³		µg/m ³		µg/m ³		µg/m ³		µg/m ³	
Dust	5.8	8.2	4.0	6.5	5.8	8.8	5.3	6.8	8.1	10.1
SeaS.	8.3	6.8	9.6	7.1	9.6	7.2	7.0	6.8	6.9	5.4

Table 1 – Average concentrations (Av) and standard deviation (Sd) of the main measured elements/species for all the sampling period and in the different seasons.

2
3
4

1

Ratio	Mg/Na ⁺	SO ₄ /Na ⁺	Cl/Na ⁺	K/Na ⁺	Ca/Na ⁺	Br/Na ⁺
PMF profile	0.13	0.23	1.68	0.04	0.05	0.005
Henderson and Henderson 2009	0.119	0.253	1.796	0.037	0.038	0.006

2

Table 2 – Elemental ratios for the sea-salt source.

3

1

EF(Mg)	EF(Si)	EF(K)	EF(Ca)	EF(Ti)	EF(Mn)	EF(Fe)	EF(Sr)
0.90	0.83	0.92	2.81	1.21	0.79	1.05	1.72

2

Table 3 – Enrichment factors for the mineral dust source.

3

1 **FIGURE CAPTION**

2

3 Figure 1 – Sampling site location.

4

5 Figure 2 – Daily and monthly mean PM₁₀ concentrations in the years 2007-2008.

6

7 Figure 3 – Profiles (fraction of the elemental mass concentration, blue columns) and
8 explained variations (red squares, referring to the right y-axis) of the identified PM₁₀ sources.

9

10 Figure 4 – Temporal evolution of the PM₁₀ contributions from the identified sources.

11

12 Figure 5- HYSPLIT backward trajectories ending at Lampedusa at 50 m a.s.l. corresponding
13 to the peaks in the PM₁₀ contributions due to mineral dust source (see text for the details).

14

15 Figure 6 – Temporal evolution of the “combustion emissions” and “mineral dust” sources;
16 examples of trajectories from fires are evidenced in red, together with an example of
17 advection from the Saharan region (green).

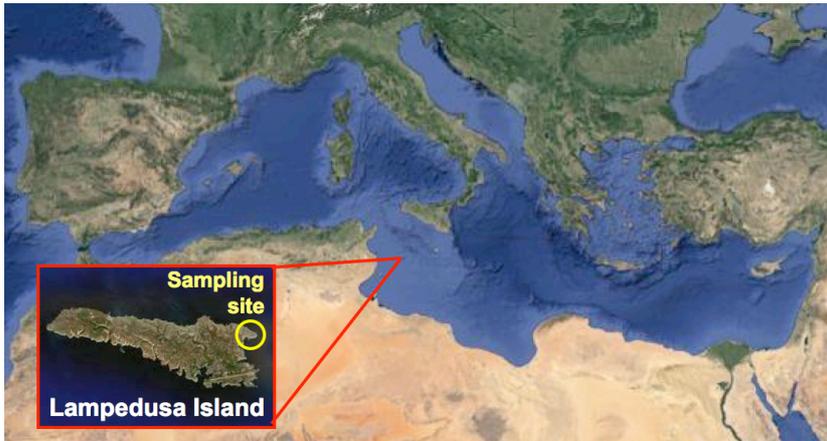
18

19 Figure 7 – Comparison between EC and OC data and PMF estimates of the contributions of
20 combustion emissions and primary particulate ship emissions in the period of EC/OC data
21 availability. Results for EC and OC are reported with solid lines, black and red, respectively,
22 referring to the secondary axis (empty rounds report the single sample result attributing it to
23 the central sampling day).

24

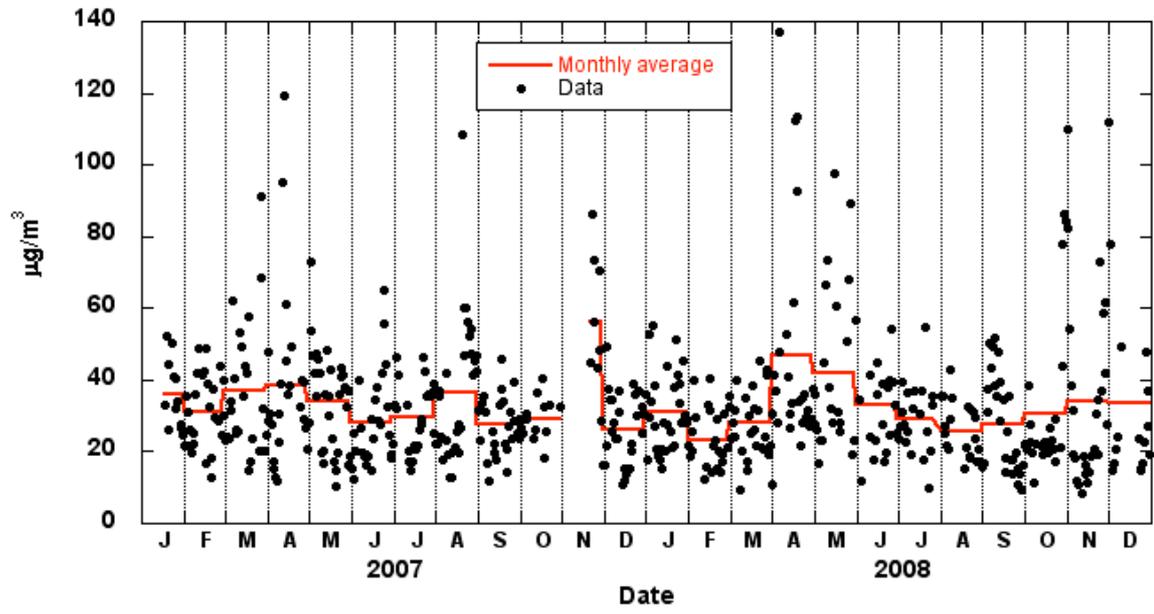
1 Figure 8 – Seasonal relative contributions to PM₁₀ concentration (average over the years
2 2007-2008). Factors may sum to more than 100% due to number rounding.
3
4 Figure 9 – Seasonal absolute contributions to PM₁₀ concentration (average over the years
5 2007-2008).
6
7 Figure 10 – Relative annual contributions to the PM₁₀ concentration (average over the years
8 2007-2008).
9
10 Figure 11 – Correlation plot between the contributions from secondary sulphate and from
11 primary particulate emitted by ships as obtained by the PMF analysis.
12

1 Figure 1



2

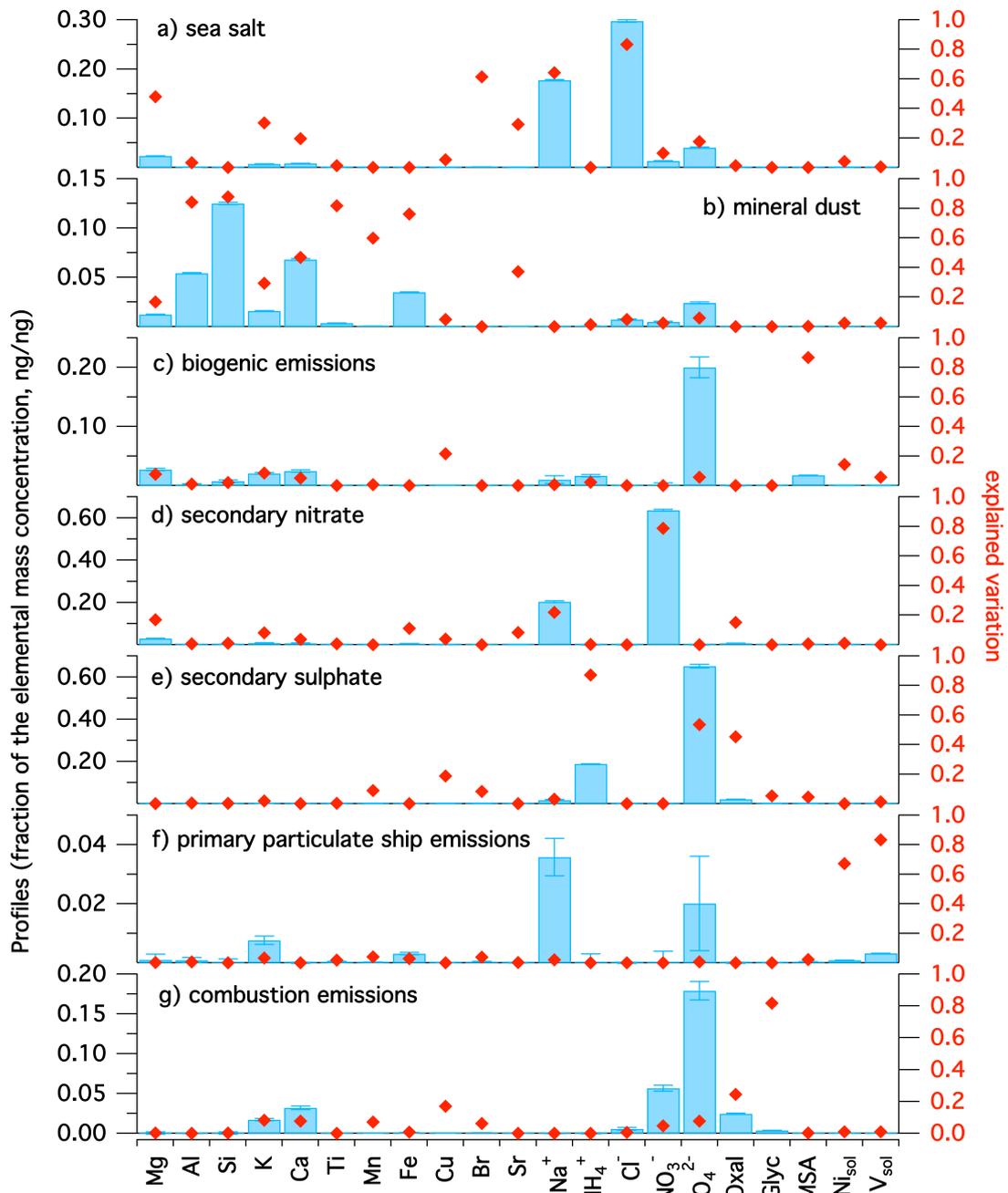
1 Figure 2



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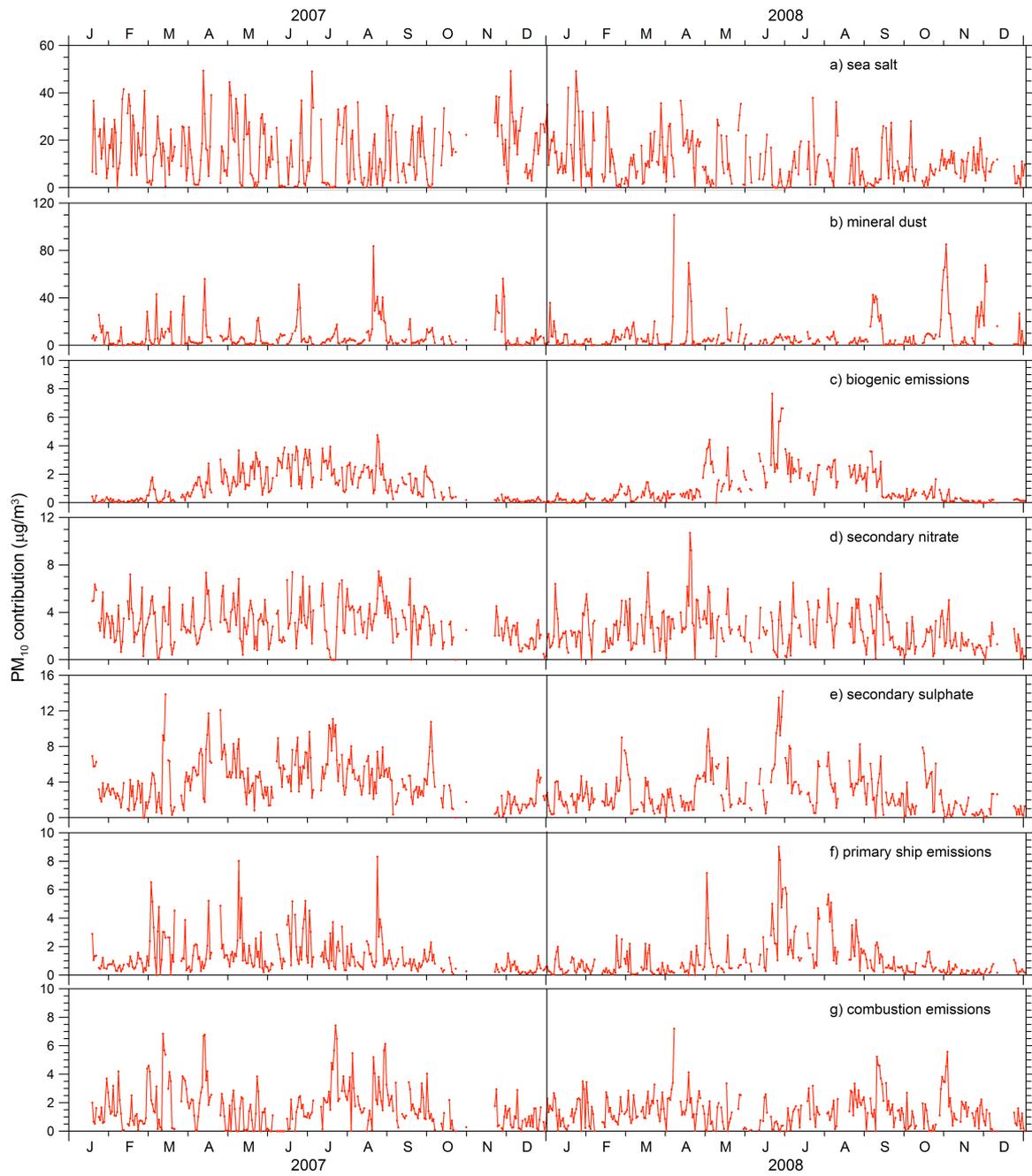
1 **Figure 3**



2

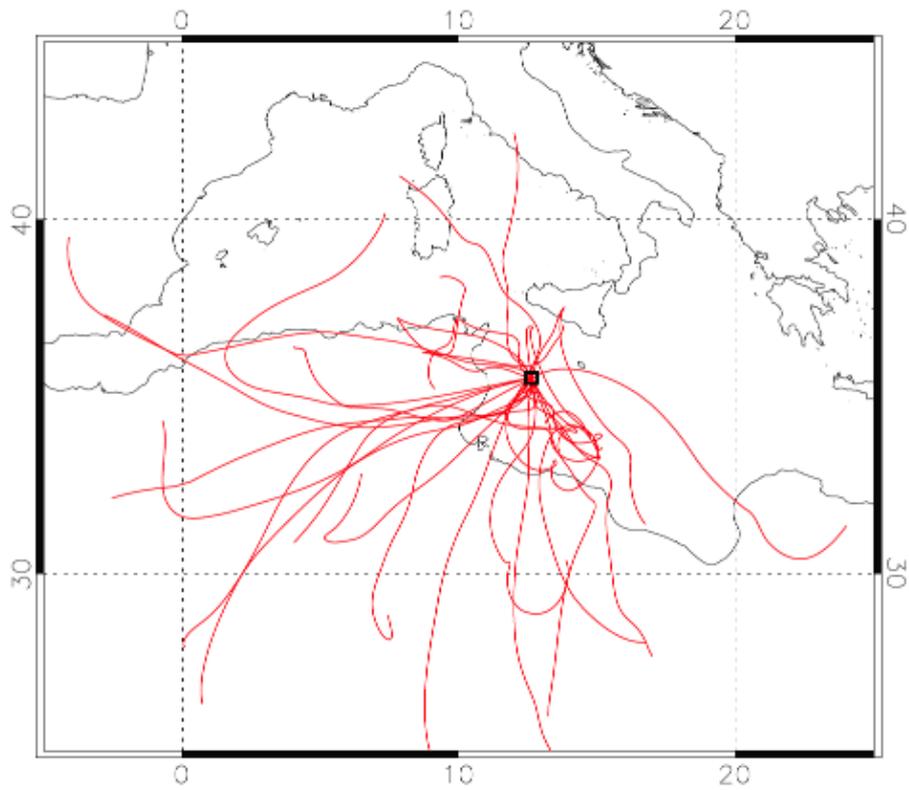
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1 Figure 4



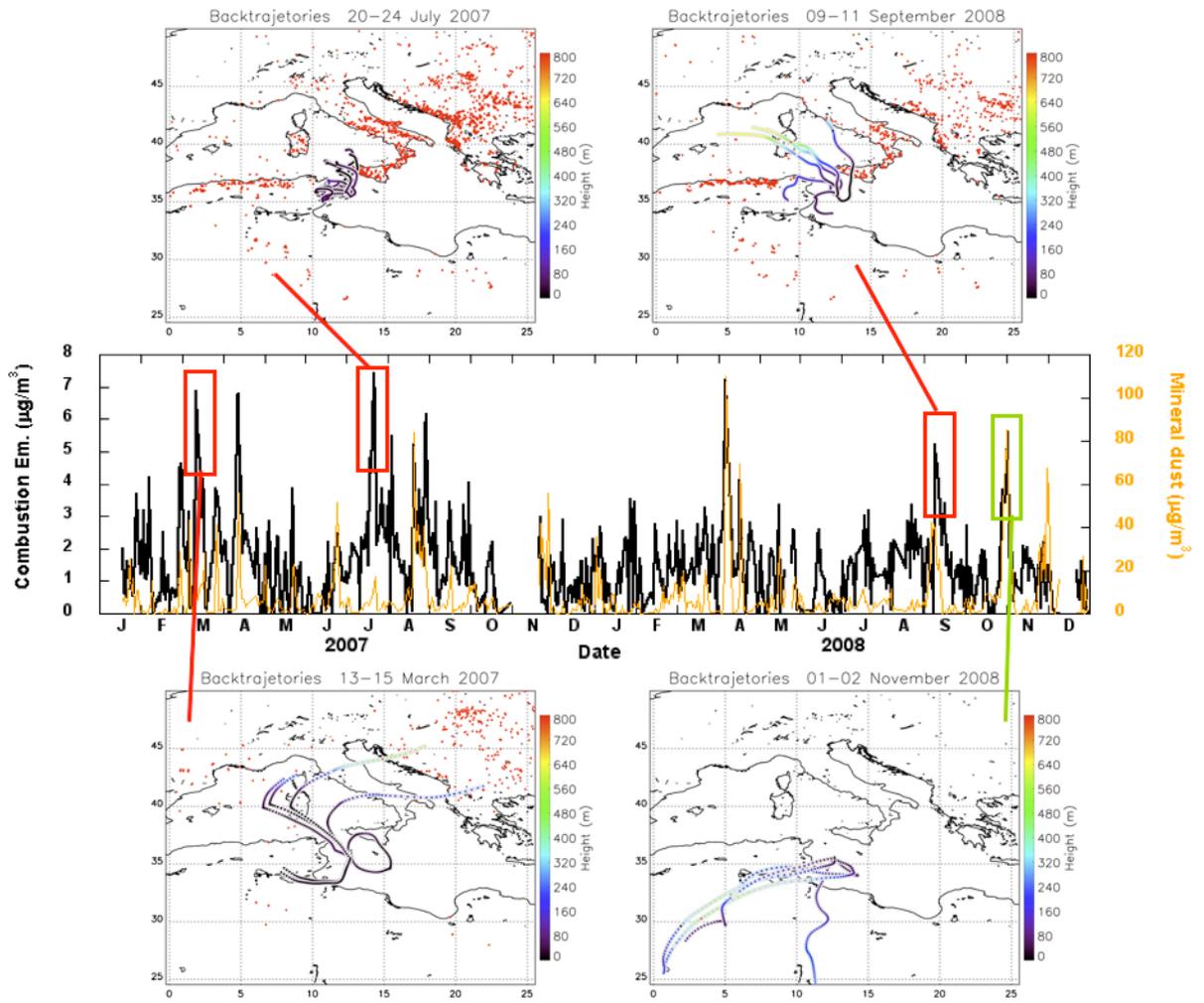
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1 Figure 5



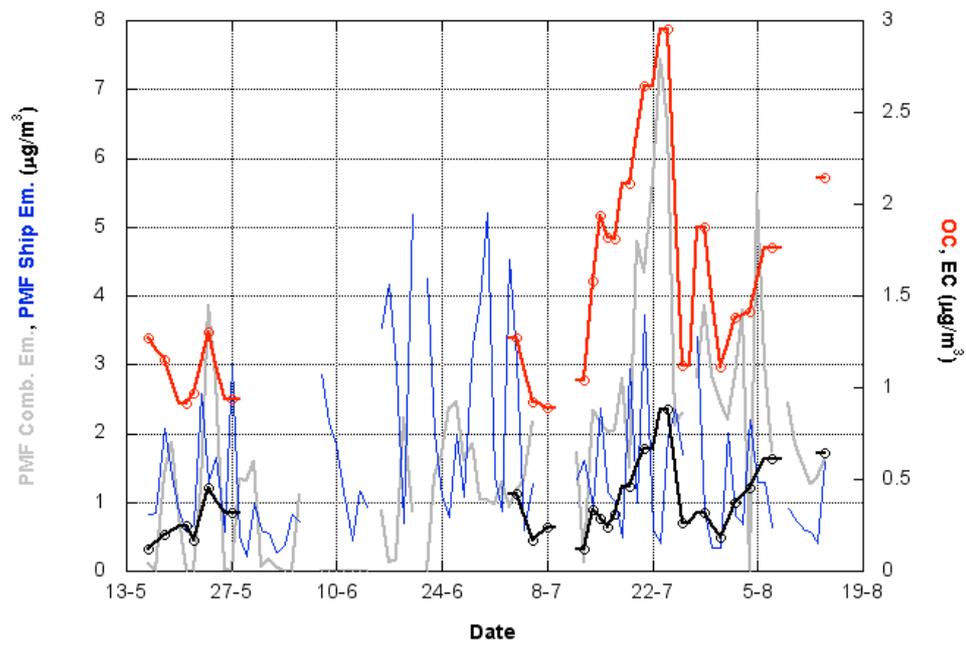
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1 Figure 6



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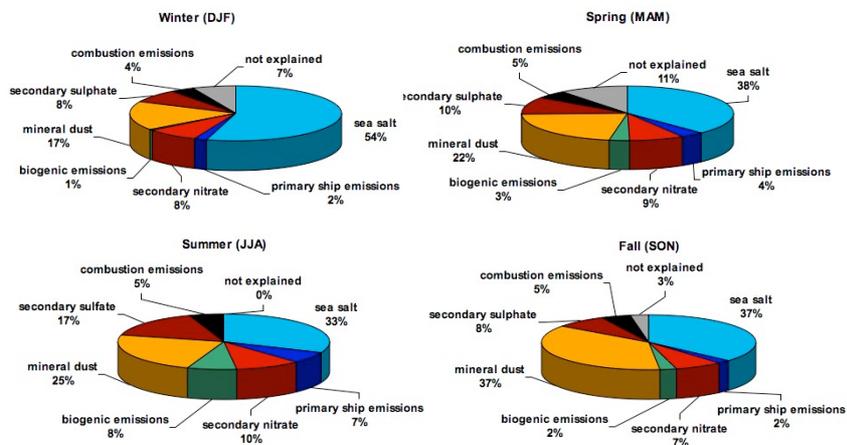
1 Figure 7



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3

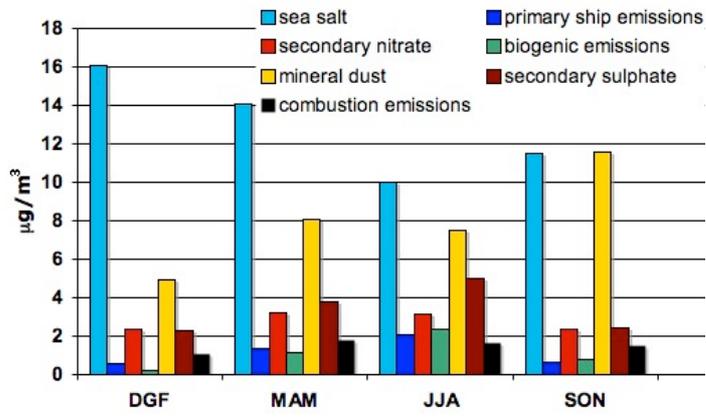
1 Figure 8



2

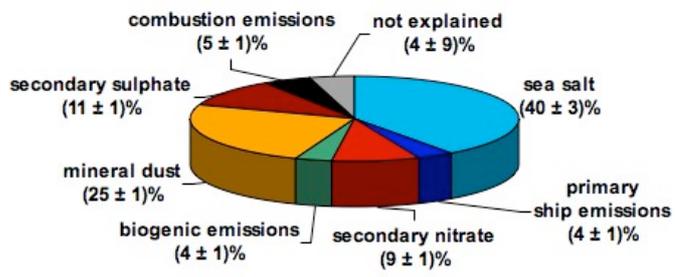
3

1 Figure 9



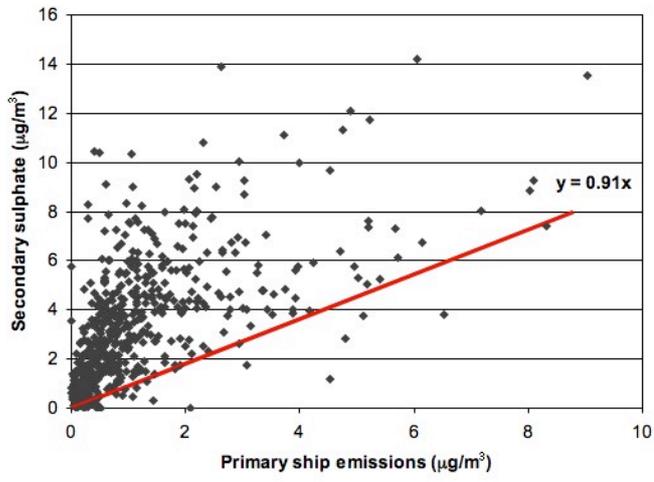
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1 Figure 10



2

1 Figure 11



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3