

1 **Response to editor comment**

2 By Huan Yu, hyu@nusit.edu.cn

3

4 We revised the paper to address the concerns of editor.

5 1. Please make saturation ratio equations for both SA and ELVOC simultaneously and look how the
6 particle can grow.

7

8 Previously, we did not subtract H_2SO_4 growth rate ($GR_{H_2SO_4}$) from the measured Growth Rate.
9 This is because we didn't measure H_2SO_4 vapor with any instrument.

10 Now we calculate $GR_{H_2SO_4}$ from H_2SO_4 proxy (Mikkonen et al. 2011 equation).

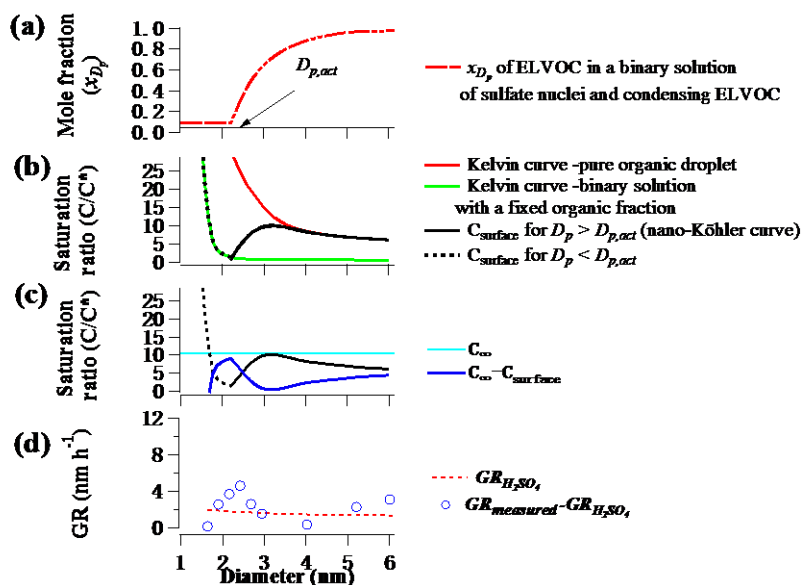
11
$$GR_{H_2SO_4} = \frac{\gamma}{2\rho_{v,H_2SO_4}} \left(1 + \frac{D_{v,H_2SO_4}}{D_p}\right)^2 \left(\frac{8kT}{\pi}\right)^{\frac{1}{2}} \left(\frac{1}{m_p} + \frac{1}{m_{v,H_2SO_4}}\right)^{\frac{1}{2}} m_{v,H_2SO_4} C_{H_2SO_4} \quad \text{Eq.1}$$

12 All parameters in eq.1 are from Nieminen et al. 2010 for H_2SO_4 .

13 Here H_2SO_4 surface concentration over particles is set to zero, i.e., gaseous SA is assumed to
14 condense irreversibly into post-nucleation particles (i.e. >1.7 nm). This is because ammonia, amines,
15 and ELVOC will neutralize or stabilize H_2SO_4 .

16 Then $GR_{elvoc} = GR_{measure} - GR_{H_2SO_4}$. Eq.2

17 The corrected GR_{elvoc} , as well as $GR_{H_2SO_4}$, are again shown in Fig S2.



18

Fig S2.

19 From Fig S2d, We can find: 1. The varying trend of GR_{elvoc} is the same as $GR_{measure}$. Our
 20 arguments in last email still make sense. 2. H_2SO_4 can not explain all of the $GR_{measure}$. $GR_{H_2SO_4}$ is
 21 monotonically decreasing and smaller than $GR_{measure}$.

22 After subtracting H_2SO_4 , ELVOC is fitted using the equation we shown in last email:

$$23 \quad GR_{elvoc} = \frac{\gamma}{2\rho_{v,elvoc}} \left(1 + \frac{D_{v,elvoc}}{D_p}\right)^2 \left(\frac{8kT}{\pi}\right)^{\frac{1}{2}} \left(\frac{1}{m_p} + \frac{1}{m_{v,elvoc}}\right)^{\frac{1}{2}} m_{v,elvoc} (C_{elvoc} - C_{surface}) \quad eq.3$$

$$24 \quad \text{where } C_{surface} = C_{elvoc}^* \exp\left(\frac{4\sigma v}{kTD_p}\right) x_{D_p}$$

$$x_{D_p} = \begin{cases} x_0 & , \quad D_p < D_{p,act} \\ x_0 + \frac{(D_p^3 - D_{p,act}^3)/v_{elvoc}}{(D_p^3 - D_{p,act}^3)/v_{elvoc} + D_{p,act}^3/v_{sulfate}} & , \quad D_p \geq D_{p,act} \end{cases}$$

25 where only x_0 , C_{elvoc} , and C_{elvoc}^* are free parameters and is fitted using measured GR_{elvoc} . x_0 is
 26 fitted to range from 0.07 to 0.42 for all the 8 events in the manuscript. C_{elvoc} , and C_{elvoc}^* are shown
 27 in Table 1 of the manuscript. C_{elvoc} decreases to 63% to 92% of the values without H_2SO_4
 28 subtraction.

29 Figure 9 is revised according, to show the corrected GR_{elvoc} and fitting curves.

30

31 2. Also, please, take into account that both SA and ELVOC concentrations are changing as a
 32 function of time.

33 In Fig S2, the growth rates and ELVOC concentration are for a group of particles with different
 34 size at the same instant in time. We are not tracking the same particle with size evolving as a
 35 function of time.

36

37 This is because GR calculated by our method is not from appearance time method tracking a banana
 38 contour.

39

40 Our method calculates the GR of different sizes at the same instant in time (like Kuang et al. 2012
 41 ACP paper).

42 For an instant in time, if the particle composition is like what we shown in Fig S2a, we get an
 43 instant GR_{elvoc} and C_{elvoc} with Eq. 3. Then in the next time step, the particle composition changed,
 44 so did GR_{elvoc} and C_{elvoc} .

45

46 3. On the other hand, please, revisit your observational data, since I think that it might be too
47 difficult to observe local maximas and minimas in so short time windows.

48

49 We agree with you that, using appearance time method ($GR_{H2SO4} = \frac{Dp2-Dp1}{t2-t1}$), it is not realistic to
50 observe local maxima/minima, for such small D_p intervals and short time windows.

51 So we used $\frac{dN(D_{p1}, D_{p2}, t)}{dt} = J(D_{p1}, t) - J(D_{p2}, t) - CoagSnk(D_{p1}, D_{p2}, t) + CoagSrc(D_{p1}, D_{p2}, t)$

52 For every 4 minutes (PSM scanning cycle), we got a $J(D_{p1}, t)$.

53 Then $GR(D_p, t) = J(D_{p1}, t)/n(D_{p1}, t)$

54 In this way, we did not rely on short $t2 - t1$ and $Dp2 - Dp1$, thus it is possible to obtain GR
55 for different size. Therefore, our GR is for an instant in time, not for a NPF growing period.

56 In the manuscript Section 2.3 and 3.3 we have explained this.

57

58 **Nucleation and growth of sub-3 nm particles in the polluted urban atmosphere of a megacity in**
59 **China**

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67

68 **Abstract**

69 Particle size distribution down to 1.4 nm was measured in the urban atmosphere of Nanjing, China in
70 spring, summer and winter during 2014-2015. Sub-3 nm particle event, which is equivalent to
71 nucleation event, occurred on 42 out of total 90 observation days, but new particles could grow to
72 cloud condensation nuclei (CCN)-active sizes on only 9 days. In summer, infrequent nucleation was
73 limited by both unfavorable meteorological conditions (high temperature and RH) and reduced
74 anthropogenic precursor availability due to strict emission control measures during the 2014 Youth
75 Olympic Games in Nanjing. The limiting factors for nucleation in winter and spring were
76 meteorological conditions (radiation, temperature, and RH) and condensation sink, but for the further

77 growth of sub-3 nm particles to CCN-active sizes, anthropogenic precursors again became limiting
78 factors. Nucleation events were strong in the polluted urban atmosphere. Initial $J_{1.4}$ at the onset and
79 peak $J_{1.4}$ at the noontime could be up to $2.1 \times 10^2 \text{ cm}^{-3} \text{ s}^{-1}$ and $2.5 \times 10^3 \text{ cm}^{-3} \text{ s}^{-1}$, respectively, during the
80 8 nucleation events selected from different seasons. Time-dependent $J_{1.4}$ usually showed good linear
81 correlations with a sulfuric acid proxy for every single event ($R^2 = 0.56 - 0.86$, excluding a day with
82 significant nocturnal nucleation), but the correlation among all the 8 events deteriorated ($R^2 = 0.17$)
83 due to temperature or season change. We observed that new particle growth rate did not increase
84 monotonically with particle size, but had a local maximum up to 25 nm h^{-1} between 1-3 nm. The
85 growth rate behavior was interpreted in this study as the solvation effect of organic activating vapor
86 in newly formed inorganic nuclei.

87

88 **1. Introduction**

89 New particle formation (NPF) is an important source of secondary aerosols in the atmosphere
90 (Kulmala et al., 2004a). Field studies and model simulations have consistently shown that NPF can
91 enhance cloud condensation nuclei (CCN) concentrations and contribute significantly to the global
92 CCN production (Wiedensohler et al., 2009; Yue et al., 2011; Spracklen et al., 2008; Pierce and
93 Adams, 2009; Merikanto, 2009; Yu and Luo, 2009; Matsui et al. 2013). NPF is a two-stage process
94 consisting of formation of clusters and subsequent growth to detectable sizes (Kulmala et al., 2000).
95 Recently, chamber experiments have made substantial progress in revealing the fundamental
96 processes involved in particle nucleation and growth (Kirkby et al., 2011; Almeida et al., 2013;
97 Schobesberger et al., 2013; Riccobono et al., 2014; Ehn et al., 2014; Kürten et al., 2014). However,

98 consistent theories are still under investigation to quantify the processes physically, chemically, and
99 dynamically (Kulmala et al., 2013, 2014). For example, the identity and physico-chemical properties
100 of assisting vapors other than sulfuric acid (H_2SO_4) are uncertain so far. It is also uncertain what
101 mechanisms allow the assisting vapors to overcome strong Kelvin effect over sub-3 nm particles.
102 Existing mechanisms include condensation of extremely low volatility organic compounds (Ehn et
103 al., 2014), nano-Köhler activation (Kulmala et al., 2004b), heterogeneous chemical reactions (Zhang
104 and Wexler, 2002), heterogeneous nucleation (Wang et al., 2013), and adsorption of organics on
105 cluster surface (Wang and Wexler, 2013). However, the relative importance of various mechanisms
106 is unknown.

107 Direct measurements of size- and time dependent nucleation rate and growth rate in sub-3 nm
108 size range are important to constrain the relative contributions from different mechanisms and
109 precursors. Such measurements are also important to evaluate the survival probability of new particle
110 to CCN-active sizes (~ 100 nm for soluble particles at 0.2% super saturation, Pierce and Adams, 2009)
111 and to reveal the limiting factors in the process. Recently, a series of new instruments have been
112 developed to measure sub-3 nm aerosol number concentration and chemical composition, such as
113 condensation particle counters (e.g., PSM, DEG-SMPS, Jiang et al., 2011a; Sipila et al., 2009;
114 Vanhanen et al., 2011), ion spectrometers (e.g., NAIS, Asmi et al., 2009), and mass spectrometers
115 (e.g., Cluster-CIMS, APi-TOF, CI-APi-TOF, Jokinen et al., 2012; Junninen et al., 2010; Zhao et al.,
116 2010). Kuang et al. (2012) developed a de-coupling method to measure size- and time dependent
117 growth rates of sub-5 nm particles. Their results at two urban sites in U.S.A showed that
118 size-resolved growth rates increased approximately linearly with particle size from 1 to 5 nm.

119 Similar results were also observed in the Boreal forest (Kulmala et al., 2013; Lehtipalo et al., 2014).
120 Based on growth rates measured below 2 nm, Kulmala et al. (2013) identified three separate size
121 regimes, which were dominated by different key gas to particle conversion processes.

122 The relative contribution of different precursors and mechanisms to the nucleation and growth of
123 1-3 nm particles may vary greatly with atmospheric conditions (Riipinen et al., 2012). Therefore,
124 sub-3 nm particle measurements in a variety of atmospheric conditions, e.g., remote or urban
125 atmosphere, biogenic- or anthropogenic emission dominated areas, are immensely valuable.
126 Unfortunately, such data are very sparse until now (Jiang et al., 2011b; Kuang et al., 2012; Kulmala
127 et al., 2013; Lehtipalo et al., 2009, 2010, 2011; Yu et al., 2014a, b). China is suffering from severe
128 atmospheric particulate matter pollution in recent years (Chan and Yao, 2008; Yue et al., 2011). To
129 the best of our knowledge, only two studies were conducted in China to measure the occurrence of
130 new particles down to ~1 nm. In these two studies, air ions (Herrmann et al., 2014) or neutral
131 particles (Xiao et al., 2015) were measured by AIS or PSM in two urban locations of Yangtze River
132 Delta region. Both studies were conducted in the winter season.

133 Here we reported the nucleation and growth of sub-3 nm particles in the urban atmosphere of
134 Nanjing, China on arbitrarily selected observation days in spring, summer and winter of 2014-2015.
135 Our aim was to (1) provide new information about the initial steps of NPF based on size- and time
136 resolved nucleation rate and growth rate measurements, and (2) find possible limiting factors behind
137 the seasonal and diurnal variations of nucleation events in the polluted urban atmosphere.

138 **2. Methodology**

139 **2.1 Field measurements**

140 Nanjing is the second largest megacity after Shanghai in the Yangtze River Delta (YRD) region
141 of China (Chan and Yao, 2008). The YRD city cluster, covering 2.1×10^5 km² land with 170 million
142 residents, is one of the most populated and industrialized regions in China. Field measurement was
143 conducted from the third floor (15m above the ground level) of an academic building beside a
144 Chinese national meteorology observatory facility in NUIST campus (32.20° N, 118.71° E, symbol
145 ① in Figure 1). The sampling was carried out during the months of May (May 10-30, 2014), June
146 (June 1-15, 2014), December (December 24-31, 2014), February (February 16-22, 2015), and March
147 (March 1-7, 2015). Total 58 measurement days were arbitrarily selected to represent spring, early
148 summer and winter seasons, but to avoid any rain-event.

149 As part of an intensive summer campaign (12 August-12 September 2014), the summer
150 measurement was conducted at a local governmental meteorology observatory platform (32.06°N,
151 118.70°E) that is 14km south to the NUIST site (② in Figure 1). The instruments were housed in an
152 air conditioned trailer, using exactly the same sampling inlets as the NUIST site. The main aim of the
153 summer campaign was to understand the effects of regional emission control measures during the
154 2014 Young Olympic Games (August 1-September 15) on air quality. Because the two sites locate
155 within the same urban air shed, the measurement provided an opportunity to study seasonal variation
156 of nucleation and its relationship with meteorological variables and gaseous precursors.

157 Sub-3 nm clusters/particles (hereafter referred as particles) were measured with a nano
158 condensation nucleus counter system (nCNC) consisting of a Particle Size Magnifier (PSM model
159 A10, Airmodus Oy, Finland) and a butanol Condensation Particle Counter (model A20, Airmodus
160 Oy, Finland). During the measurement, an ambient air flow of 14 standard liters per minute (slpm)

161 was drawn into building room or trailer via a 72 cm long and 1.0 cm I.D diameter stainless steel (SS)
162 tube, which was extended outside the room/trailer horizontally. PSM then sampled a split flow of 2.5
163 slpm via a SS T-union. The design of the inlet tubing (length and air flow rate) was to minimize the
164 transport loss of nano particles. The size dependent transport survival ratios of sub-3 nm particles in
165 the inlet tubing was estimated (67 %-86 % for 1.4-3.0 nm) and corrected using a particle loss
166 calculator tool (von der Weiden et al., 2009).

167 PSM was operated in a continuous scanning mode with a cycle of 240 steps between saturator
168 flow rates of 0.1 and 1.0 slpm within 240 seconds. The particle cut-off sizes of the nCNC varied with
169 saturation ratios in the saturator (Vanhanen et al., 2011). A step-wise method was used to invert raw
170 scanning data to size spectrum (time resolution: 4 minutes) of sub-3 nm particles, which were
171 classified evenly into 6 size bins, i.e. 1.4-1.6, 1.6-1.9, 1.9-2.2, 2.2-2.4, 2.4-2.7, and 2.7-3.0 nm. The
172 inverted particle number concentrations in the 6 bins were referred as $N_{1.5}$, $N_{1.8}$, $N_{2.0}$, $N_{2.3}$, $N_{2.6}$ and
173 $N_{2.8}$, using mean values of upper and lower size boundaries in each bin. The step-wise method was
174 described in detail by Lehtipalo et al. (2014).

175 Particle size distributions in the range from 3 - 750 nm were obtained by integrating two
176 scanning mobility particle spectrometers (SMPS) with a nano-SMPS (a TSI differential mobility
177 analyzer DMA3085 and a condensation particle counter CPC3776; scanning range: 3 - 64 nm) and a
178 long-SMPS (TSI DMA3081 and CPC3775; scanning range: 64 - 750 nm). During the summer
179 campaign, only the long-SMPS was operated to scan particles from 8 - 350 nm. Scanning cycles of
180 both SMPS systems were 4 minutes, in order to synchronize with the nCNC. The SMPSs sampled
181 ambient air from a separate sampling inlet. The inlet was a 129 cm long and 1.0 cm I.D

182 horizontally-oriented SS tube with an air flow of 14 slpm. The transport loss of particles in the SMPS
183 inlets was corrected using size dependent survival ratios of 85-100% for particles > 3 nm.

184 Sulfur dioxide (SO₂), ozone (O₃), carbon monoxide (CO) and nitrogen oxides (NO and NO₂)
185 concentrations were measured every 1 minute with Thermo Environmental Instruments (model
186 43i-TLE, 49i, 48i, and 42i, respectively). When gaseous SO₂, O₃, NO₂ and CO data were not
187 available, hourly SO₂, O₃, NO₂ and CO were obtained from nearby local Environmental Protection
188 Agency (EPA) monitoring station. PM_{2.5} was monitored with Thermo Scientific TEOM 1405.
189 Meteorological variables including wind speed, wind direction, relative humidity (RH), temperature
190 and solar radiation flux were recorded every 1 hour during the measurement periods. Mean
191 concentrations of PM_{2.5}, SO₂, and O₃ were 79 μg m⁻³, 10 ppbv and 48 ppbv, respectively, during the
192 whole measurement period. Therefore, we regard our measurement environment as a polluted urban
193 atmosphere.

194

195 **2.2 Nucleation event and growth patterns**

196 A criterion was set to determine whether the nCNC detected sub-3nm particles in the atmosphere.
197 The criterion was that total particle concentration reading followed the supersaturation scanning
198 cycle of PSM so that the highest concentrations were measured at lowest cut-off sizes (see also
199 Figure 2 in Lehtipalo et al., 2014). However, it was possible in the step-wise inversion method that
200 the number concentration fluctuation of > 3 nm particles within a 4-min scanning cycle was wrongly
201 inverted to sub-3 nm particles even when sub-3 nm particles actually did not exist according to the
202 above criterion. As a result, the step-wise inversion method always reported a background sub-3 nm

203 particle concentration (N_{sub-3} , i.e. the sum of $N_{1.5}$, $N_{1.8}$, $N_{2.0}$, $N_{2.3}$, $N_{2.6}$ and $N_{2.8}$) of $0.5 \times 10^3 - 2 \times 10^3$
204 cm^{-3} in the nighttime and $3 \times 10^3 - 8 \times 10^3 \text{ cm}^{-3}$ in the daytime. Similar background levels of sub-3 nm
205 particles during non-NPF periods were also reported by other studies that used the nCNC (Kulmala
206 et al., 2013; Lehtipalo et al., 2014; Xiao et al., 2015). Following their procedures, we did not attempt
207 to subtract this background from N_{sub-3} reported in this study.

208 We defined sub-3 nm particle event as sub-3 nm particle occurrence with N_{sub-3} higher than
209 background level persisting for longer than 1 h in the atmosphere. In this study, we used sub-3 nm
210 particle event as an approximate measure of nucleation event. This is because (1) there was an
211 approximately positive linear correlation between N_{sub-3} and nucleation rate ($J_{1.4}$ in this study, see
212 next section) with R^2 of 0.94 (Figure 2), and (2) N_{sub-3} calculation needs only nCNC scanning data
213 and was thus more readily available than $J_{1.4}$ calculation which needs both nCNC and SMPS
214 scanning data. Similar definition has been discussed in our previous studies (Yu et al., 2014a, b).
215 Apparently, a sub-3 nm particle event did not necessarily lead to an NPF event always, but it
216 indicated the intensity and frequency of nucleation in the atmosphere. One focus in this work was to
217 investigate the characteristics of sub-3 nm particle event.

218 Particle growth after nucleation is crucial to determine if nucleated particles could grow to
219 CCN-active sizes. We identified two growth patterns according to size spectrum characteristics in
220 sub-3 nm size range (Figure 3). In a Type A event (Figure 3a or b), size distribution $n(D_p, t)$ was
221 higher at smaller sizes (e.g., 1.4–1.6 nm) than $n(D_p, t)$ at larger sizes (e.g., 2.7–3.0 nm). The size
222 spectrum below 3 nm thus looked like a “volcano”. In a Type B event (Figure 3c or d), $n(D_p, t)$ was
223 lower at smaller sizes than $n(D_p, t)$ at larger sizes (“up-side-down volcano”). For the size range > 3

224 nm, depending on whether a banana-shape growth was seen, we further defined Type A1/A2 and
225 Type B1/B2 events: particles eventually grew to CCN-active sizes in Type A1 and B1 events, while
226 in Type A2 and B2 events banana-shape particle growth to CCN-active sizes was not seen. Therefore,
227 Type A1 and B1 events were equivalent to conventional NPF events based on either DMPS or SMPS
228 measurements.

229 Type B size distribution was more unusual since $n(D_p, t)$ of small particles were less than $n(D_p, t)$
230 of large particles in the sub-3 nm size range. We excluded the possibility of deteriorated nCNC
231 detection efficiencies for small particles due to high particle loading in the sample air. This is
232 because total number concentrations of nCNC during our measurements never approached nCNC
233 upper concentration limit $4 \times 10^5 \text{ cm}^{-3}$, especially in the early stage of nucleation when total particle
234 concentration was rather low. Our nCNC was also calibrated periodically using $\text{H}_2\text{SO}_4\text{-H}_2\text{O}$ particles
235 in a laboratory flow tube to ensure the detection efficiency of the nCNC. The different chemical
236 composition of atmospheric particles could be another factor of lower detection efficiencies. It is
237 well known that organic substances activate less readily in diethylene glycol (e.g. Kangasluoma et al.
238 2014). However, it is accepted in general that larger particles have higher mass fraction of organics
239 than smaller particles in a NPF process. If organic substances activate less readily in DEG, it should
240 be even more difficult to activate larger particles than smaller particles. Therefore, the increasing
241 $n(D_p)$ with D_p (i.e. upside down volcano) could not be simply due to lower detection efficiency of
242 organic substances.

243 **2.3 Formation rate and growth rate calculations with a simplified GDE method**

244 Conventional appearance time method determined growth rates (hereafter, *GR*) during the initial
 245 period of NPF by finding the time steps when newly-formed particles appeared at certain size bins
 246 and calculating the *GR* from the time differences between successive size bins (Kulmala et al., 2012;
 247 Lehtipalo et al., 2014). This method was often not applicable to the NPF event with high *GR* below 3
 248 nm, e.g., 0.3 nm/4 min (i.e. 4.5 nm h⁻¹) with size intervals 0.3 nm and scanning time intervals 4
 249 minutes in our measurements. Furthermore, sub-3 nm particles were often generated persistently
 250 throughout the daytime period. Maximum concentrations in the sub-3 nm size bins could appear
 251 around noontime, which is a few hours later than the onset of nucleation. Therefore, we were not
 252 able to pinpoint correctly maximum or 50% maximum concentrations at the onset of nucleation.

253 The rapid growth of small particles in the urban atmosphere was the motivation that we used an
 254 alternative method to calculate growth rate and formation rate. Here, we analyzed 8 events (listed in
 255 Table 1, including both Type A1/A2 and B1/B2 events) in detail, for which complete size spectra
 256 from 1.4 - 750 nm were available without distorted, broken or noisy data. Total 8 size bins were
 257 classified: 6 evenly-divided size bins in sub-3 nm and 2 size bins in 3-30 nm (3-10 and 10-30 nm).
 258 For an aerosol population that is growing through simultaneous condensation and coagulation,
 259 aerosol general dynamic equation (GDE) describes the evolution of number concentration in a size
 260 bin between particle diameters D_{p1} and D_{p2} ($D_{p2} > D_{p1}$) as:

$$261 \frac{dN(D_{p1}, D_{p2}, t)}{dt} = J(D_{p1}, t) - J(D_{p2}, t) - \text{CoagSnk}(D_{p1}, D_{p2}, t) + \text{CoagSrc}(D_{p1}, D_{p2}, t) \quad (1)$$

262

263 where $N(D_{p1}, D_{p2}, t)$ is the number concentration from D_{p1} to D_{p2} , inverted from nCNC or SMPS
 264 scanning data. J is condensational growth flux (i.e. particle formation rate) across the lower ($D_{p,1}$) or
 265 upper ($D_{p,2}$) boundaries of a size bin. In the first size bin of 1.4-1.6 nm, $J(1.4 \text{ nm}, t)$, or simply $J_{1.4}$,
 266 is the unknown formation rate of the smallest particles that we measured.

267 $\text{CoagSnk}(D_{p1}, D_{p2}, t)$ and $\text{CoagSrc}(D_{p1}, D_{p2}, t)$ are the sink and source terms defining the
 268 coagulation out of and into the size bin between D_{p1} and D_{p2} . Assuming bin k has lower boundary
 269 $D_{p,1}$ and upper boundary $D_{p,2}$,

$$270 \quad \text{CoagSnk}(D_{p1}, D_{p2}, t) = N(k, t) \sum_{i=1}^{98} (1 - \theta_{k,i,k}) K_{k,i} N(i, t) \quad (2)$$

$$271 \quad \text{CoagSrc}(D_{p1}, D_{p2}, t) = \frac{1}{2} \sum_{i=1}^{k-1} \sum_{j=1}^{k-1} \theta_{i,j,k} K_{i,j} N(j, t) N(i, t) \quad (3)$$

272 Here $N(i, t)$ is number concentration of bin i . $K_{i,j}$ is coagulation kernel for a collision between
 273 particles from bins i and j . Probability coefficient $\theta_{i,j,k} = 1$, if the volume sum of two coagulating
 274 particles ($v_i + v_j$, here the volume is calculated from the center diameter of a bin) is within the
 275 volume boundaries of bin k . otherwise $\theta_{i,j,k} = 0$. The particle coagulation of total 98 bins was
 276 considered, but the coagulation terms were only needed to be calculated for the smallest 8 bins from
 277 1.4 to 30 nm. According to our calculation, $\text{CoagSrc}(D_{p1}, D_{p2}, t)$ accounted for only 0 - 0.8 % of the
 278 total particle flux into a bin (i.e. $\text{CoagSrc}(D_{p1}, D_{p2}, t) + J(D_{p1}, t)$) in the sub-3 nm size range. This
 279 implied that self coagulation played a negligible role and most of the production flux into a bin is due
 280 to condensational growth from gas molecules.

281 The GDE here is the same as the Eq. 1 by Kuang et al. (2012). In their method, gaseous H_2SO_4
 282 was measured simultaneously and a constant $GR(D_p, t) / GR_{\text{H}_2\text{SO}_4}(D_p, t)$ ratio at a given size over time

283 was assumed. Their $GR(D_p, t)$ was then solved by fitting the GDE to the measured size distributions.
 284 In our study, however, we did not measure gaseous H_2SO_4 . Instead, $J(30 \text{ nm}, t)$ in the largest size
 285 bin, which is the condensational growth flux out of 30 nm, was set to zero. This simplification was
 286 valid in the four Type A2/B2 events when particles never grew to $> 30 \text{ nm}$ (March 4, February 19,
 287 May 20 and May 16). In the rest four Type A1/B1 events (February 18, December 27, May 15, and
 288 August 15), this was also valid during the early NPF period when particles did not grow out of 30 nm
 289 and during the late NPF period when particles grew out of 30 nm completely. During the middle
 290 period of events (usually around 11:00 AM-14:00 PM), $J(30 \text{ nm}, t)$ was underestimated and thus
 291 $J_{1.4}$ could be regarded as a lower estimate. In the four Type A2/B2 events, our calculation showed
 292 that J_{10} was only 0-4% of $J_{1.4}$. Xiao et al. (2015) and Kulmala et al. (2013) measured both $J_{1.5}$ and J_3
 293 using appearance time method. Their J_3 was less than 7% of $J_{1.5}$. Furthermore, $J_{30}/J_{1.4}$ ratio should be
 294 even smaller than $J_{10}/J_{1.4}$ or $J_3/J_{1.5}$ ratios, considering the 8 events were carefully selected to ensure
 295 all sub-30 nm particles were grown from nucleation (not emitted directly from emission sources like
 296 vehicular engine). All these evidences supported that even if J_{30} was set to 0, $J_{1.4}$ would not be
 297 underestimated more than 7% when particles grew cross 30 nm on February 18, December 27, May
 298 15, and August 15.

299 Equation 1 requires the balance of condensational growth (J), coagulation terms (CoagSnk and
 300 CoagSrc) and the changing rate of particle number concentration (dN/dt). Using Eq. (1) we can
 301 therefore calculate the nucleation rate $J(1.4 \text{ nm}, t)$ and formation rates $J(D_p, t)$ across all size bin
 302 boundaries from 1.6 to 10 nm. After the formation rates $J(D_p, t)$ were obtained, $GR(D_p, t)$ was
 303 calculated from $J(D_p, t)/n(D_p, t)$, where $n(D_p, t)$ is size distribution calculated as $n(D_p, t) = \frac{dN(t)}{dD_p}$

304 for each size bin. On the other hand, the appearance time method could still be applied to (1) the size
305 range of > 3 nm where size intervals were large (2-6 nm), and (2) the size range of < 3 nm when GR
306 was small. The results from appearance time method will also be showed in the next section.

307

308 **3. Results and discussion**

309 Section 3.1, sections 3.2-3.4 and section 3.5 were organized, respectively, to address the
310 following 3 issues: (1) seasonal variation, diurnal variation and limiting factors of nucleation event
311 (represented by sub-3 nm particle event) in the polluted urban atmosphere, (2) time- and size
312 dependent nucleation rate and growth rate of sub-3 nm particles, and their implications for nucleation
313 and growth mechanisms, (3) Inhibited particle growth to CCN-active sizes in strong nucleation
314 events of Type B2.

315 **3.1 Seasonal and diurnal variations of nucleation event**

316 As seen from Figure 2, there was an approximate linear correlation between N_{sub-3} and $J_{1.4}$ with
317 the slope of $N_{sub-3}/J_{1.4}$ equal to ~ 160 . This seemed to suggest that the average residence time of
318 new particles in the sub-3 nm size range was 160 seconds before they were scavenged due to
319 coagulation or grew out of 3 nm. The sub-3 nm particles observed at the present work were thus
320 formed in situ in the urban atmosphere and not likely to be carried-over by air transport. In this
321 section we used sub-3 nm particle event as an approximate measure of nucleation.

322 We observed significant seasonal characteristics of nucleation event (Figure 4). Nucleation was
323 rare and weak in summer, while it was commonly observed in all other seasons. During our

324 measurements from 2014 to 2015, nucleation events occurred on 81% of all spring observation days
325 (May 2014), 53% in early summer (June 2014), 10% in summer (August and September 2014), and
326 64% in winter (December 2014, February and March 2015). We compared intensity (N_{sub-3}) and
327 frequency of nucleation events, as well as meteorological variables (temperature, RH, wind speed,
328 and solar radiation flux) and gaseous pollutants (SO_2 , NO_2 , CO and O_3) for spring, summer and
329 winter seasons. June was not shown in Figure 4 for comparison, because it was a transit season from
330 spring (May) to summer (August and September). The data were first averaged over the entire event
331 period for each event; and we then used event-averaged data to create box and whistler plots for the 3
332 seasons. $PM_{2.5}$ was used here as a surrogate of condensational sink (CS), because of the more ready
333 availability of $PM_{2.5}$ data than SMPS data.

334 As shown in Figure 4, nucleation in summer was characterized by lowest frequency, lowest
335 N_{sub-3} ($2.2 \times 10^4 \text{ cm}^{-3}$), and short nucleation period (only 1-2 hours). Strict emission control measures
336 during the 2014 Youth Olympic Games resulted in relatively low $PM_{2.5}$ level ($32 \pm 8 \mu \text{ g m}^{-3}$),
337 which should favor nucleation. However, relatively low SO_2 concentration ($1.4 \pm 0.6 \text{ ppbv}$), high
338 temperature ($26 \pm 2 \text{ }^\circ\text{C}$), and high RH ($74.3 \pm 4.2 \%$) might not be in favor of nucleation. A simple
339 H_2SO_4 proxy ($\text{Radiation} \times SO_2 / PM_{2.5}$) indicated that summer H_2SO_4 concentration was likely to be the
340 lowest among the 3 seasons, which could explain low nucleation intensity/frequency.

341 We further examined diurnal variations of N_{sub-3} and other variables on event and non-event days
342 in winter (Figure 5). Because nucleation in winter was characterized by Type B event
343 (“up-side-down volcano” below 3 nm), event days were further divided to Type B1 and Type B2
344 events depending on whether banana-shape particle growth was seen. The difference between Type

345 B1 and B2 will be discussed later in Section 3.5. During the non-event days, N_{sub-3} ranged from
346 $2.4 \times 10^3 \text{ cm}^{-3}$ in the night to 8.0×10^3 in the day, which was close to background levels. During the
347 event days, N_{sub-3} in the night was close to that of non-event days, but could reach $8 \times 10^4 - 20 \times 10^4$
348 cm^{-3} in the middle of the day. This was more than 10 times higher than those on the non-event days.
349 From Figure 5 we can see that non-event day had higher concentrations of anthropogenic precursors
350 (indicated by SO_2 , NO_2 , and CO), but nucleation seemed to be limited by higher pre-existing particle
351 surface area (indicated by $\text{PM}_{2.5}$), higher temperature and RH, and lower radiation flux.
352 Photochemistry indicators O_3 was also lower during non-event days.

353 Nucleation in spring was characterized by highest frequency (81%) among all seasons. Highest
354 gaseous pollutant concentration of (H_2SO_4 proxy, SO_2 , NO_2 , CO and O_3) and radiation seemed to the
355 favorable factors to explain this. However, N_{sub-3} in spring ($3.3 \times 10^4 \text{ cm}^{-3}$) was much lower than that
356 in winter ($11.2 \times 10^4 \text{ cm}^{-3}$). Unfavorable factors included high pre-existing particle surface area ($\text{PM}_{2.5}$:
357 $112 \pm 68 \mu \text{g m}^{-3}$) and high temperature ($27 \pm 4 \text{ }^\circ\text{C}$) in spring. Integrating the above seasonal and
358 diurnal variation information in Figure 4 and Figure 5, we tentatively identified that the limiting
359 factors for nucleation in our urban atmosphere were (1) radiation, temperature, RH and CS in winter
360 and spring, and (2) temperature, RH and available gaseous precursors in summer.

361 Out of total 90 measurement days, March 4, 2015 in winter was the only day that we observed
362 significant nocturnal nucleation. Sunrise and sunset were at 6:29 AM and 18:00 PM local time on
363 March 4, but nucleation were observed persistently from 4:00 AM - 20:00 PM. N_{sub-3} increased from
364 $3.5 \times 10^3 \text{ cm}^{-3}$ at 4:00 AM to $6.3 \times 10^4 \text{ cm}^{-3}$ before sunrise. During 10:00-11:00 AM, peak N_{sub-3}
365 reached $3 \times 10^4 \text{ cm}^{-3}$, 3 times higher than the average of all other event days in winter. Apparently,

366 nocturnal nucleation on March 4 could not be explained as carry-over of daytime particles nor being
367 associated with photochemistry. This implied the existence of certain dark nucleation source. There
368 are a number of observations that have also shown nighttime particle formation events in various
369 atmospheric conditions (Junninen et al., 2008; Lehtipalo et al., 2011; Lee et al., 2008; Ortega et al.,
370 2009, 2012; Russell et al., 2007; Suni et al., 2008; Svenningsson et al., 2008; Yu et al., 2014), but the
371 mechanisms behind the nocturnal nucleation are yet still highly speculative. With our instrument
372 capability in this work, we could not deduce any valuable information on the nocturnal nucleation
373 mechanism, except that we found the air mass on 04 March was relatively clean (both *CS* and gases,
374 mean *CS*: 0.15 s^{-1}), and temperature and RH (mean: 4.4°C and 33%) were favorable for nucleation.

375

376 **3.2 Size- and time dependent formation rates of sub-3 nm particles**

377 We observed 23 Type A events and 9 Type B events during the measurements. The different
378 size distribution patterns (Figure 3) were probably linked to the mechanism or intensity of nucleation
379 and growth. To address this issue, we first compared the formation rates and growth rates in two
380 types of events. Formation rates J of 1.4, 1.6, 1.9, 2.2, 2.4, 2.7 and 3.0 nm particles were shown in
381 Figure 6 (upper panels) for typical Type A and Type B events. It is obvious that $J_{1.4}$ was much higher
382 on February 18 (Type B) than that on May 15 (Type A). A clear time dependence of J was observed.
383 For example, $J_{1.4}$ was $60 \text{ cm}^{-3} \text{ s}^{-1}$ at the onset of the nucleation event on May 15 and increased to 300
384 $\text{cm}^{-3} \text{ s}^{-1}$ in the middle of the day. In the type B event on February 19, the initial and peak $J_{1.4}$ were
385 2.1×10^2 and $1.2 \times 10^3 \text{ cm}^{-3} \text{ s}^{-1}$ respectively. Therefore, our method provided more information of
386 nucleation than conventional calculation methods that usually showed only an averaged J at the onset

387 of a nucleation event. Our method was also different from Kulmala et al. (2013). Their
388 time-dependent formation rate on an event day was equal to size distribution $n(D_p, t)$ times a constant
389 growth rate at the onset of the event obtained with the appearance time method.

390 The diurnal variation of J implied that nucleation was probably linked to sunlight induced
391 photochemistry. We calculated the correlations between $J_{1.4}$ and an H_2SO_4 proxy for the 8 events of
392 our interest. The H_2SO_4 proxy was calculated following $[H_2SO_4] = 8.21 \times 10^{-3} k \cdot Radiation \cdot$
393 $[SO_2]^{0.62} \cdot (CS \cdot RH)^{-0.13}$ (Eq. 8 of Mikkonen et al. 2011), where k is the temperature-dependent
394 reaction-rate constant. Figure 7a show that good linear correlation was usually seen for every single
395 event with R^2 ranging from 0.72 to 0.86 for 6 out of 8 events. A moderate R^2 of 0.56 was obtained
396 for August 15. R^2 was lowest (0.34) on March 4, 2015. This is not surprising because we know
397 March 4 was the only day with nocturnal nucleation during the measurement period. The H_2SO_4
398 proxy was also calculated using the derivation of Petäjä et al. (2009), which resulted in lower R^2 of
399 $\log J_{1.4}$ vs. $\log [H_2SO_4]$ for all 8 events. Therefore, in this study we used Mikkonen H_2SO_4 proxy, as
400 it was derived with more comprehensive datasets than Petäjä et al. (2009). The slopes of $\log J_{1.4}$ vs.
401 $\log [H_2SO_4]$ were close to 1 in all events (0.82-1.17, excluding March 4), indicating activation theory
402 can explain the nucleation mechanism in our urban atmosphere.

403 If data points of all the 8 events were put together, the linear correlation between H_2SO_4 proxy
404 and $J_{1.4}$ deteriorated (slope=1.1, $R^2=0.17$, Figure 7b). In spite of considerable scattering, most of data
405 points located between $J_{1.4}=10^{-4.1} \times [H_2SO_4]$ and $J_{1.4}=10^{-6.3} \times [H_2SO_4]$. An interesting finding was
406 that the scattering of $J_{1.4}$ vs. $[H_2SO_4]$ proxy among all 8 events was probably due to temperature or
407 season change (Figure 7b). More specifically, with the same level of H_2SO_4 proxy, $J_{1.4}$ was higher in

408 winter with lower temperature than in spring/summer with higher temperature. There were two
409 possibilities behind the deteriorated linear correlation between H₂SO₄ proxy and $J_{1,4}$: (1) inaccurate
410 H₂SO₄ proxy and (2) other varying factors in nucleation mechanism. First, it was very likely that
411 H₂SO₄ concentrations in our polluted urban atmosphere were overestimated by the H₂SO₄ proxy of
412 Mikkonen et al. (2011), which was based on statistic regression of historical datasets from relatively
413 clean Europe/USA atmosphere. The extent of overestimation may vary with the levels of predictor
414 variables (e.g., SO₂, temperature, CS). Mean SO₂ mixing ratios were 10.5 and 7.3 ppbv in
415 spring/summer and winter during our measurements, respectively. These were 1 order of magnitude
416 higher than SO₂ mixing ratios at the 6 European and USA sites (mean values: 0.23-3.4 ppbv,
417 Mikkonen et al., 2011). Our CS in the 8 events was on the order of magnitude of 10⁻² s⁻¹, again higher
418 than 10⁻³ s⁻¹ in Mikkonen et al. (2011). Mikkonen et al. (2011) had already pointed out that the
419 predictive ability was lower for long term data due to atmospheric condition changes in different
420 seasons.

421 Second, organic condensing vapor concentrations in particle growth events were higher in winter
422 than those in spring/summer (Table 1, see Section 3.4). If the organics were also involved in
423 nucleation, $J_{1,4}$ should be enhanced in winter. The enhancement of nucleation by organics (most
424 likely anthropogenic organics in our urban atmosphere) could be supported by the comparison of $J_{1,4}$
425 dependences on H₂SO₄ between our study and the measurements in the Boreal forest: besides
426 possible H₂SO₄ overestimation, $J_{1,4}=10^{-4.1} \times [\text{H}_2\text{SO}_4] - 10^{-6.3} \times [\text{H}_2\text{SO}_4]$ in our sites was much higher
427 than $J_{1,5}=1.06 \times 10^{-7} [\text{H}_2\text{SO}_4]^{1.1}$ in Hyytiälä during active aerosol formation periods (Kulmala et al.,

428 2013). At last, low temperature itself might enhance nucleation in winter (Brus et al., 2011) via
429 increasing the saturation ratios of all nucleation precursors (e.g., water, H₂SO₄, organics).

430

431 **3.3 Size- and time dependent growth rates of sub-3 nm particles**

432 Particle size distribution $n(D_p)$ and corresponding $GR(D_p)$ at an instant in time during the events
433 were shown in Figure 6 middle and lower panels. A local minimum of $n(D_p)$ at 2.4 nm, followed by a
434 local maximum somewhere between 2.5 and 10 nm, was seen on May 15, 2014. Such size
435 distribution characteristics on May 15, 2014, as well as on all other Type A event days, was also
436 observed by Kulmala et al. (2013) in the Boreal forest (Figure 1A and S9A in their paper) and by
437 Jiang et al. (2011b) in the urban area of Atlanta, USA (Figure 1 in their paper). We further examined
438 the growth rates in the size range of 1-3 nm on May 15, 2014. It was shown that there was a local
439 maximum of $GR(D_p)$ at 2.4 nm. This could explain why $n(D_p)$ was increasing in 2.4-3 nm size range:
440 when particle condensational flow out of a size bin was slowed down, it was possible that particles
441 flowing into the size bin accumulated, leading to particle number increase in the bin.

442 We saw more unusual behaviors of $n(D_p)$ and $GR(D_p)$ in the Type B event on February 18
443 (Figure 6 right panels): $GR(D_p)$ decreased monotonically in the size range of 1.4 - 3 nm, and
444 accordingly $n(D_p)$ increased monotonically at the same time. A high $GR(D_p)$ of 25 nm h⁻¹ was
445 observed at 1.6 nm and $GR(D_p)$ decreased rapidly to 1.7 nm h⁻¹ at ~3 nm. If we consider that $GR(D_p)$
446 below 1.6 nm would eventually decrease due to strong Kelvin effect of all possible precursors
447 (H₂SO₄ or organics), the overall trend of $GR(D_p)$ in the Type B event was in fact the same as Type A:

448 for the smallest clusters, growth rate was small (possibly below 1 nm h^{-1}) and increased with D_p . It
449 reached a local maximum somewhere between 1-3 nm, after which $GR(D_p)$ decreased with D_p . For a
450 typical NPF event, $GR(D_p)$ would eventually increase again after a local minimum between 3-10 nm.
451 The difference between the Type A event (February 18) and Type B event (May 15) was the D_p of
452 local maximum $GR(D_p)$ (2.4 nm vs. 1.6 nm).

453 The interesting behaviors of $n(D_p)$ and $GR(D_p)$ in our urban atmosphere were different from the
454 stereotyped understanding that steady-state cluster size distribution $n(D_p)$ decreases with D_p in
455 nucleation and GR increases monotonically with D_p in an NPF event. It should be pointed out that if
456 we calculated the overall GR in 1.4-3 nm, $GR_{1.4-3}$ was 3.6 nm h^{-1} on May 15 and 4.4 nm h^{-1} on
457 February 18, which were still smaller than GR_{3-20} during the initial period of the events (7.7 and 6.0
458 nm h^{-1} , calculated using appearance time method). Table 1 showed that a faster GR_{3-20} than $GR_{1.4-3}$
459 were quite common, except in two events on May 16 and 20 when particles did not grow beyond 3
460 nm. Overall, GR was still increasing with increasing D_p .

461 Kuang et al. (2012) had also reported a local maximum of GR at $\sim 2.6 \text{ nm}$ in an NPF event
462 measured in Atlanta, USA (Figure 1b in their paper). In this study we further point out that GR could
463 decrease monotonically with D_p in 1-3 nm range in strong nucleation events. Our GR was calculated
464 from a simplified GDE method, however, the decrease of GR in 1-3 nm size range could be easily
465 inferred from the size spectra shown in Figure 3 or Figure 6 middle panels: for a D_{p2} that was larger
466 than D_{p1} , particle formation rate $J(D_{p2})$ must be smaller than $J(D_{p1})$. If we observed a higher $n(D_{p2})$
467 than $n(D_{p1})$, $GR(D_p)$ that was equal to $J(D_p)/n(D_p)$ must be smaller at D_{p2} than D_{p1} .

468

469 3.4 Growth rate due to condensing organic vapor on newly formed nuclei in sub-3 nm sizes

470 Apparently, the complicated growth rate behaviours in our polluted urban atmosphere can not be
471 explained by H₂SO₄ condensation alone, not only because H₂SO₄ condensational growth rate
472 ($GR_{H_2SO_4}$, calculated from the H₂SO₄ proxy and shown as black dashed lines in Figure 6) was smaller
473 than the measured growth rate (GR_{meas}), but also because $GR_{H_2SO_4}$ curve should follow a
474 monotonically decreasing trend in > 1 nm sizes assuming a collision-only condensational growth
475 without vaporization (Nieminen et al., 2010).

476 Nano-Köhler theory (Anttila et al., 2004; Kulmala et al., 2004b, c) suggested that when a soluble
477 organic vapor is dissolved in newly formed nuclei of aqueous-phase sulfate at certain size between
478 1-3 nm, the surface organic vapor pressure is lowered and thus assists the growth of the nuclei. Here,
479 we continued our discussion based on the nano-Köhler theory to provide an explanation of GR
480 behaviours observed in our urban atmosphere. We first subtract $GR_{H_2SO_4}$ from GR_{meas} to obtain the
481 growth rate due to a condensing organic vapor (hereafter, denoted as ELVOC, extremely low
482 volatility organic compound):

$$GR_{meas,elvoc} = GR_{meas} - GR_{H_2SO_4}$$

483 where $GR_{H_2SO_4}$ is calculated from the H₂SO₄ proxy concentration [H₂SO₄] following Nieminen et al.
484 (2010),:

$$485 \quad GR_{H_2SO_4} = \frac{\gamma}{2\rho_{v,H_2SO_4}} \left(1 + \frac{D_{v,H_2SO_4}}{D_p}\right)^2 \left(\frac{8kT}{\pi}\right)^{\frac{1}{2}} \left(\frac{1}{m_p} + \frac{1}{m_{v,H_2SO_4}}\right)^{\frac{1}{2}} m_{v,H_2SO_4} [H_2SO_4] \quad (4)$$

486 and all parameters in Eq.4 are taken from Nieminen et al. (2010) for H₂SO₄.

487 The size-dependent growth rate due to the uptake of ELVOC was expressed as

$$GR_{elvoc} = \frac{\gamma}{2\rho_{v,elvoc}} \left(1 + \frac{D_{v,elvoc}}{D_p}\right)^2 \left(\frac{8kT}{\pi}\right)^{\frac{1}{2}} \left(\frac{1}{m_p} + \frac{1}{m_{v,elvoc}}\right)^{\frac{1}{2}} m_{v,elvoc} (C_{elvoc} - C_{surface}) \quad (5)$$

where C_{elvoc} is gas-phase ELVOC concentration far from the particle. The net uptake of ELVOC is driven by the difference of C_{elvoc} and equilibrium surface concentration over the particle $C_{surface}$. $C_{surface}$ is determined by the pure component saturation vapor pressure C_{elvoc}^* , particle curvature $\exp\left(\frac{4\sigma v}{kTD_p}\right)$ and particle composition:

$$C_{surface} = C_{elvoc}^* \exp\left(\frac{4\sigma v}{kTD_p}\right) x_{D_p}$$

x_{D_p} is the mole fraction of water-soluble ELVOC in the pseudobinary solution consisting of ELVOC and the aqueous sulfate nuclei. The pseudobinary solution was treated ideal here. An example of x_{D_p} as a function of D_p was shown in Figure 8a. Nuclei activation diameter $D_{p,act}$ is the size that ELVOC fraction begins to increase significantly. For $D_p < D_{p,act}$, x_{D_p} is approximated with a fixed value (x_0). For $D_p > D_{p,act}$, x_{D_p} increases significantly with the organics being added to the sulfate core of $D_{p,act}$ size. The size dependent x_{D_p} is approximated as

$$x_{D_p} = \begin{cases} x_0 & , \quad D_p < D_{p,act} \\ x_0 + \frac{(D_p^3 - D_{p,act}^3)/v_{elvoc}}{(D_p^3 - D_{p,act}^3)/v_{elvoc} + D_{p,act}^3/v_{sulfate}} & , \quad D_p \geq D_{p,act} \end{cases}$$

Considering strong Kelvin effect, $C_{surface}$ decreases with increasing D_p for $D_p < D_{p,act}$ (Figure 8b dashed black line). For $D_p > D_{p,act}$, the rapidly increasing organic fraction in the small size regime of 2-3 nm raises the equilibrium $C_{surface}$ of ELVOC first. Then for 3-6 nm particles that are dominated by organics, $C_{surface}$ decreases to merge with the Kelvin curve of a pure organic droplet (red line, Figure 8b). The complete equilibrium curve of $C_{surface}$ in 1-6 nm (dashed + solid black

505 lines) was shown in Figure 8c. The blue line represented the calculated $C_{\text{elvoc}} - C_{\text{surface}}$. The trend
506 of $C_{\text{elvoc}} - C_{\text{surface}}$ coincided with the size dependence of the measured GR corrected by H_2SO_4
507 ($GR_{\text{meas,elvoc}}$, Figure 8d blue circle). $D_{p,act}$ corresponded to the size with local maximum
508 $GR_{\text{meas,elvoc}}$.

509 We fitted GR_{elvoc} with $GR_{\text{meas,elvoc}}$ in sub-3 nm sizes at an instant in time by adjusting 3 free
510 parameters in Eq. (5): x_0 , C_{elvoc} , and C_{elvoc}^* . Other parameters like surface tension (0.02 N m^{-1}) and
511 molar volume ($135.5 \text{ cm}^3 \text{ mol}^{-1}$) of ELVOC were taken from Kulmala et al. (2004b). Molecule
512 diameter d_v (0.8 nm) and condensed-phase density ρ_v (1.5 g cm^{-3}) of ELVOC were taken from Ehn
513 et al. (2014). Uptake coefficient γ was calculated following Nieminen et al. (2010). The fitting
514 results in Figure 9 showed that the dependence of $GR_{\text{meas,elvoc}}$ on D_p below 3 nm could be well
515 reproduced by Eq. (5) for both Type A and Type B events. Free parameter x_0 determined the
516 magnitude of the dashed black line in Figure 8b. x_0 was fitted to be 0.07-0.42 for the 8 events. C_∞
517 was sensitive to the local maximum $GR_{\text{meas,elvoc}}$ at the $D_{p,act}$. C_{elvoc} , and C_{elvoc}^* determined the
518 local minimum $GR_{\text{meas,elvoc}}$ at the right side of $D_{p,act}$. Therefore, C^* and C_∞ were basically
519 determined by the measured GR (local maximum and local minimum) and not sensitive to x_0 . As
520 shown in Table 1, the condensing organic vapor concentrations C_{elvoc} were $1.7 \times 10^7 - 1.7 \times 10^8 \text{ cm}^{-3}$.
521 The saturation vapor concentration C_{elvoc}^* were $3.3 \times 10^6 - 5.2 \times 10^7 \text{ cm}^{-3}$. They were within the orders
522 of magnitude of $10^7 - 10^8 \text{ cm}^{-3}$ and $10^6 - 10^7 \text{ cm}^{-3}$ suggested by Kulmala et al. (2004b), respectively.

523 For comparison, the GR calculated from appearance time method was also shown in Figure 9
524 (blue cross) for $> 3 \text{ nm}$ particles on May 15, February 18 and February 19, as well as for sub-3 nm
525 particles on May 20 when particle growth was relatively slow. It can be seen that the GR on May 20

526 calculated from the two methods agreed well with each other, lending credit to our GDE method.
527 The GR in >10 nm sizes was usually underestimated by GR_{elvoc} . This could be interpreted as other
528 condensing vapors with higher volatility may contribute to particle growth in the larger particles. It
529 should be noted that the appearance time method followed the time steps when newly-formed
530 particles appeared in successive size bins and thus the GR calculated from appearance time method
531 as not the growth rates at the same instant in time.

532 For all the 8 nucleation events, Table 1 summarizes the measured values of overall growth rate
533 in 1.4-3 nm ($GR_{1.4-3}$), maximum growth rate in 1.4-3 nm ($GR_{max,1.4-3}$), overall growth rate in 3-20 nm
534 (GR_{3-20}), nucleation rate ($J_{1.4}$), activation diameter ($D_{p,act}$), CS , and temperature (T) during the event
535 periods with maximum nucleation rates. Corresponding estimates of Mikkonen H_2SO_4 proxy, C_{elvoc} ,
536 and C_{elvoc}^* were shown in the right 3 columns. It can be seen that in comparison with more
537 conventional Type A events, Type B events usually occurred with (1) higher $J_{1.4}$, $GR_{max,1.4-3}$, $GR_{1.4-3}$,
538 C_{elvoc} , and CS ; (2) smaller $D_{p,act}$; and (3) lower T . However, the H_2SO_4 proxy and GR_{3-20} were
539 similar in Type A and Type B events. Based on these estimations, we concluded that higher ELVOC
540 concentration C_{elvoc} was the key factor leading to the higher $J_{1.4}$ and $GR_{1.4-3}$, which in turn resulted
541 in the different size spectrum pattern in Type B events (“up-side-down volcano”) from in Type A
542 events (“volcano”).

543 It should be noted that the organic vapor concentrations C_{elvoc} in this study were not directly
544 measured, but estimated based on Eq. (4-5). C_{elvoc} , $[H_2SO_4]$, mole fraction x_{D_p} and growth rates
545 calculated using Eq. (4-5) were for an instant in time. Aerosol dynamic processes, such as nucleation,
546 coagulation, and the condensation growth of H_2SO_4 and water vapor, were not considered explicitly

547 in Eq. (4-5). In addition, bulk thermodynamics was applied in Eq. (5) for extremely small
548 clusters/particles of sub-3 nm sizes. Therefore, although our calculation provided a possibility to
549 explain the size dependence of growth rate observed in the polluted urban atmosphere, C_{elvoc} in this
550 study was subject to uncertainties in (1) the growth rate derived from the GDE method, (2) the theory
551 by which the growth rate was related to the organic vapor concentration, and (3) H_2SO_4 level which
552 was calculated using Mikkonen et al. (2011) proxy.

553

554 **3.5 Inhibited particle growth to CCN-active sizes in strong nucleation events of Type B2**

555 Type B2 was strong nucleation event that produced rather high concentrations of new particles
556 in sub-20 nm size range (Figure 3d). High concentrations of activating vapor in these events (e.g.,
557 C_{∞} : $1.4\text{-}2.0 \times 10^8 \text{ cm}^{-3}$ on February 18 and March 4) should favor a banana-shape NPF event with fast
558 growth of particles >20 nm, due to weakened Kelvin effect. However, it was puzzling to us why new
559 particles accumulated in 2-20 nm and did not grow further on Type B2 event days (see Figure 3d).
560 We first examined the air mass trajectory characteristics of Type B2 events. Compared with Type B1,
561 Type B2 was characterized by long range transport air masses from far north of China and Mongolia.
562 The lumped trajectories with insignificant wind direction change imply that the air mass in Type B2
563 event was quite uniform. In addition, meteorological and chemical variables (high solar radiation
564 flux and wind speed, low temperature, $\text{PM}_{2.5}$, SO_2 , NO_2 , CO and O_3 , green lines in Figure 5)
565 collectively suggested that Type B2 was typical regional event in homogeneous cold air masses.
566 Therefore, the interrupted growth of new particles was not likely to be a result of wind direction
567 change.

568 As seen from Figure 5, meteorological variables on Type B2 days were generally more favorable
569 in aiding particle growth than on Type B1 days: lower PM_{2.5}, lower temperature, and higher solar
570 radiation flux. The unfavorable factors in Type B2 events, however, included lower concentrations of
571 SO₂, NO₂, and CO (anthropogenic emission indicators), lower secondary photochemical product
572 indicators O₃ and lower particle phase sulfate in 100-1000 nm (X. Ge, private communication, 2015,
573 X. Ge conducted simultaneous AMS measurement during our measurement periods). All these
574 evidences suggested that further particle growth in Type B2 events was limited by certain
575 condensing vapor other than ELVOC. Consequently, although there was a pool of sub-20 nm
576 particles, they were not further “activated” due to the low availability of this condensing vapor.
577 Following the terminology of Donahue et al. (2011, 2012), we called this condensing vapor LVOC
578 (low volatility organic compounds)

579 The above hypothesis was sound if we considered that the identity of LVOC for the growth of
580 particles > 20 nm could be different from ELVOC for sub-3 nm particle growth. Hirsikko et al.
581 (2005) observed that GR_{3-20} demonstrated an opposite seasonal cycle to $GR_{1.3-3}$: GR_{3-20} was higher in
582 summer, whereas $GR_{1.3-3}$ was higher in winter. This suggested that the condensing vapors were
583 different in identity for particles of different sizes. Hirsikko et al. (2005) attributed the condensing
584 vapors for GR_{3-20} to biogenetic VOCs. In our urban atmosphere, according to Figure 5, LVOC was
585 more likely to be from anthropogenic sources associated with SO₂, NO_x and CO emissions. A picture
586 of the growth process was thus like this: ELVOC of lower volatility, lower concentration and higher
587 water solubility activated inorganic nuclei and accelerated particle growth in smaller sizes. This in
588 turn assisted in the condensation of LVOC of high volatility, low solubility, but with larger amount

589 of mass. The further growth of particles >20 nm, which means significant increment of particle mass,
590 needed continuous supply of LVOC from the anthropogenic sources. On the Type B2 days, LVOC
591 supply was not adequate (low SO_2 , CO and NO_x). As a result, continuous banana-shape particle
592 growth did not take place.

593 **4. Conclusion**

594 NPF can contribute to CCN only after going through nucleation, initial growth steps and
595 subsequent growth to CCN-active sizes. This study provided the evidences of limiting factors in
596 these processes in a polluted urban atmosphere in China. We observed atmospheric nucleation events
597 on 42 out of total 90 observation days, but particles could grow to CCN-active sizes on only 9 days,
598 which was equivalent to 9 conventional NPF events. In summer, strict emission control measures
599 during the 2014 Youth Olympic Games resulted in relatively low $\text{PM}_{2.5}$ and anthropogenic trace
600 gases (SO_2 , NO_2 , CO and O_3) levels. Infrequent nucleation was thus limited by both low
601 concentrations of gaseous precursors and high temperature and RH in summer. In more polluted
602 winter and spring atmosphere, precursor supply was not limiting anymore; nucleation occurred once
603 meteorological conditions were favorable (i.e. low CS and temperature/RH, higher solar radiation).
604 However, for the further growth of sub-3 nm particles to CCN-active sizes, anthropogenic gaseous
605 precursors again became limiting factors.

606 A simplified GDE method was used in this study to calculate particle formation rates first and
607 then growth rates. Nucleation events were strong in the polluted urban atmosphere of Nanjing. Initial
608 $J_{1.4}$ at the onset and peak $J_{1.4}$ at the noontime could be up to $2.1 \times 10^2 \text{ cm}^{-3} \text{ s}^{-1}$ and $2.5 \times 10^3 \text{ cm}^{-3} \text{ s}^{-1}$,
609 respectively, during the 8 nucleation events selected from different seasons. The diurnal variation of

610 $J_{1.4}$ implied that nucleation was usually linked to sunlight induced photochemistry. Time-dependent
611 $J_{1.4}$ showed good linear correlations with the H₂SO₄ proxy for every single event, except a day with
612 significant nocturnal nucleation. However, the correlation between $J_{1.4}$ and the H₂SO₄ proxy for all 8
613 events deteriorated, which might reflect the effect of temperature or assisting vapor concentration in
614 the nucleation. The deteriorated correlation could also be due to the lower predictive ability of the
615 H₂SO₄ proxy in the polluted urban atmosphere for different seasons.

616 In all nucleation events, a local maximum growth rate was observed between 1-3 nm with GR up
617 to 25 nm h⁻¹. This means GR was not monotonically increasing with particle size. The overall $GR_{1.4-3}$,
618 however, was still smaller than GR_{3-20} , if particles could grow beyond 3 nm. The local maximum
619 growth rate was interpreted as the solvation effect of organic activating vapor in newly formed
620 inorganic nuclei. Based on our estimation, high ELVOC concentration of $2.3 \times 10^7 - 2.0 \times 10^8$ cm⁻³ was
621 expected to be the key factor leading to high $GR_{1.4-3}$. The varying GR of new particle in turn resulted
622 in the different particle growth patterns that we observed in Nanjing.

623 Our results call for a more robust proxy of gaseous H₂SO₄ to be developed for polluted urban
624 conditions. The study also highlighted the importance of estimating or measuring activating organic
625 vapor levels (using CI-APi-TOF, for example) in the initial growth steps of atmospheric NPF. Our
626 year-round measurement data provided valuable size evolution data of sub-3 nm clusters/particles to
627 evaluate previous aerosol dynamic models of new particle formation. A robust dynamic model was
628 needed to appropriately treat all aerosol and gas-phase processes in the initial growth steps.

629

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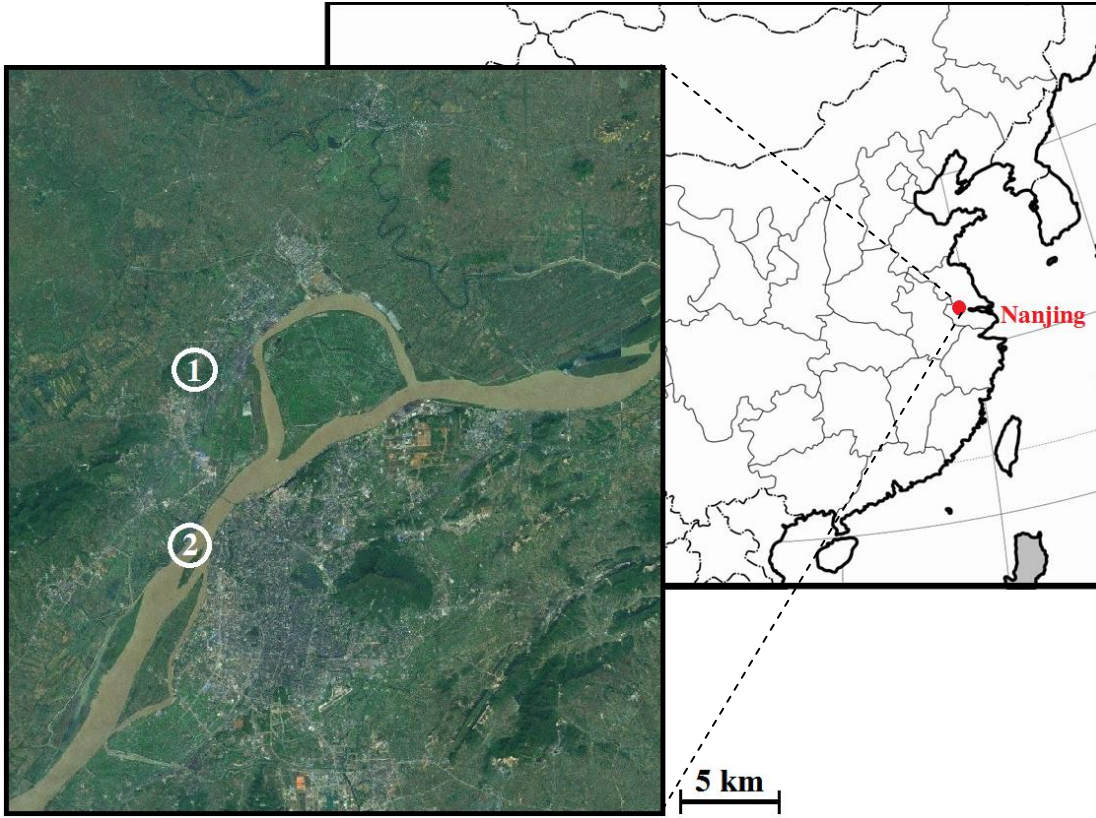
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888 *Res.*, 115, D08205, doi: 10.1029/2009JD012606, 2010.

889 Table 1. Activation diameter ($D_{p,act}$), maximum growth rate in 1.4-3 nm ($GR_{max, 1.4-3}$), overall growth rate in 1.4-3 nm ($GR_{1.4-3}$), overall growth
890 rate in 3-20 nm (GR_{3-20}), nucleation rate ($J_{1.4}$), condensation sink (CS), and temperature (T) of selected nucleation events. Estimated gas-phase
891 condensing vapor concentrations C_{elvoc} , pure saturation concentration of condensing vapor over flat surface C_{elvoc}^* , and Mikkonen H₂SO₄ proxy
892 were shown in right 3 columns. All data were for the time periods with maximum nucleation rates.

Type	Date	$D_{p,act}$ (nm)	$GR_{max,1.4-3}$ (nm h ⁻¹)	$GR_{1.4-3}$ (nm h ⁻¹)	GR_{3-20} (nm h ⁻¹)	$J_{1.4}$ (cm ⁻³ s ⁻¹)	T (°C)	CS (10 ⁻² s ⁻¹)	Mikkonen H ₂ SO ₄ proxy (cm ⁻³)	C_{elvoc} (cm ⁻³)	C_{elvoc}^* (cm ⁻³)
A1	May 15, 2014	2.4	6.4	3.6	7.7	3.0×10 ²	20.8	1.6	2.9×10 ⁷	3.5×10 ⁷	6.3×10 ⁶
A1	Aug 15, 2014	2.4	14.5	7.1	7.7	2.0×10 ²	26.1	1.8	3.1×10 ⁷	8.5×10 ⁷	2.1×10 ⁷
A2	May 16, 2014	2.4	3.8	1.9	0	95	25.3	1.9	1.4×10 ⁷	2.5×10 ⁷	4.6×10 ⁶
A2	May 20, 2014	2.2	2.9	1.6	0	92	24.1	1.9	1.3×10 ⁷	1.7×10 ⁷	3.3×10 ⁶
B1	Feb 18, 2015	1.6	25.9	4.4	6.0	1.1×10 ³	8.2	3.3	3.9×10 ⁷	1.4×10 ⁸	3.0×10 ⁷
B1	Dec 27, 2014	1.6	17.7	4.2	5.5	1.9×10 ²	7.6	2.8	3.5×10 ⁷	1.1×10 ⁸	2.2×10 ⁷
B2	Feb 19, 2015	1.9	25.0	8.9	10.1	8.0×10 ²	7.4	3.2	3.7×10 ⁷	1.7×10 ⁸	5.2×10 ⁷
B2	Mar 4, 2015	1.9	18.0	5.8	8.7	2.5×10 ³	3.9	2.2	4.8×10 ⁷	1.3×10 ⁸	1.1×10 ⁷



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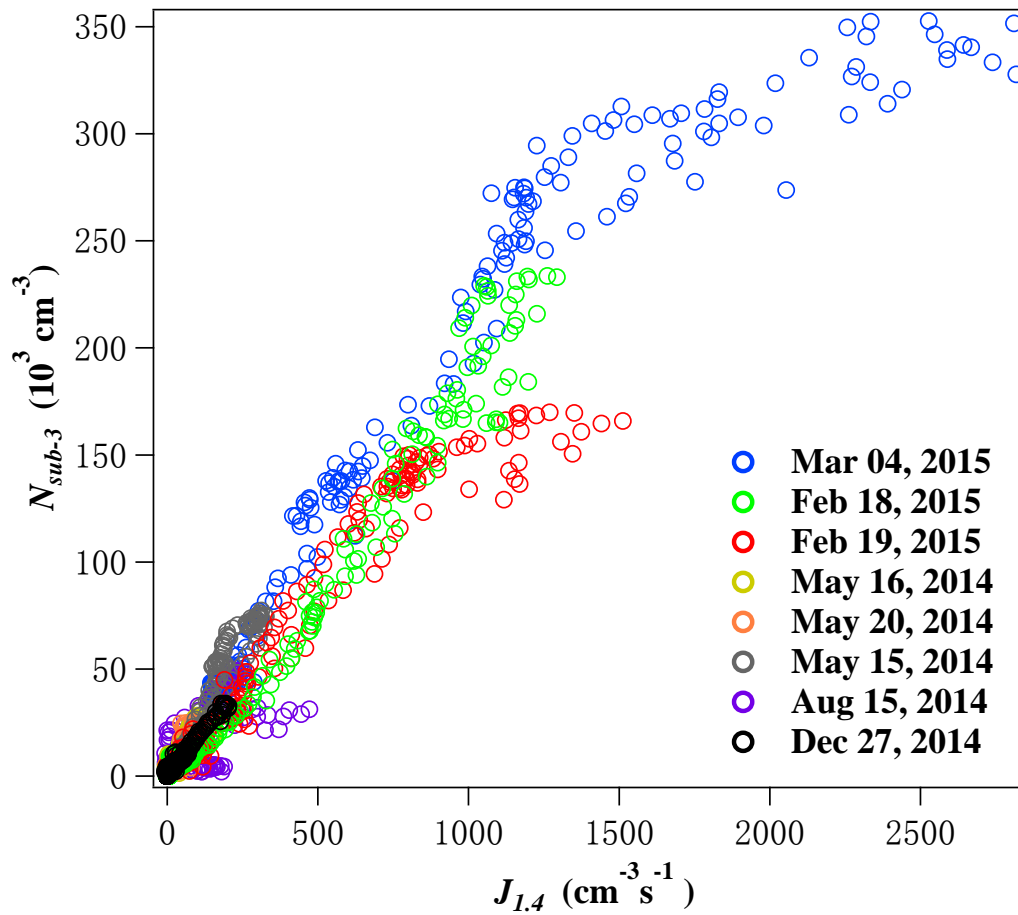
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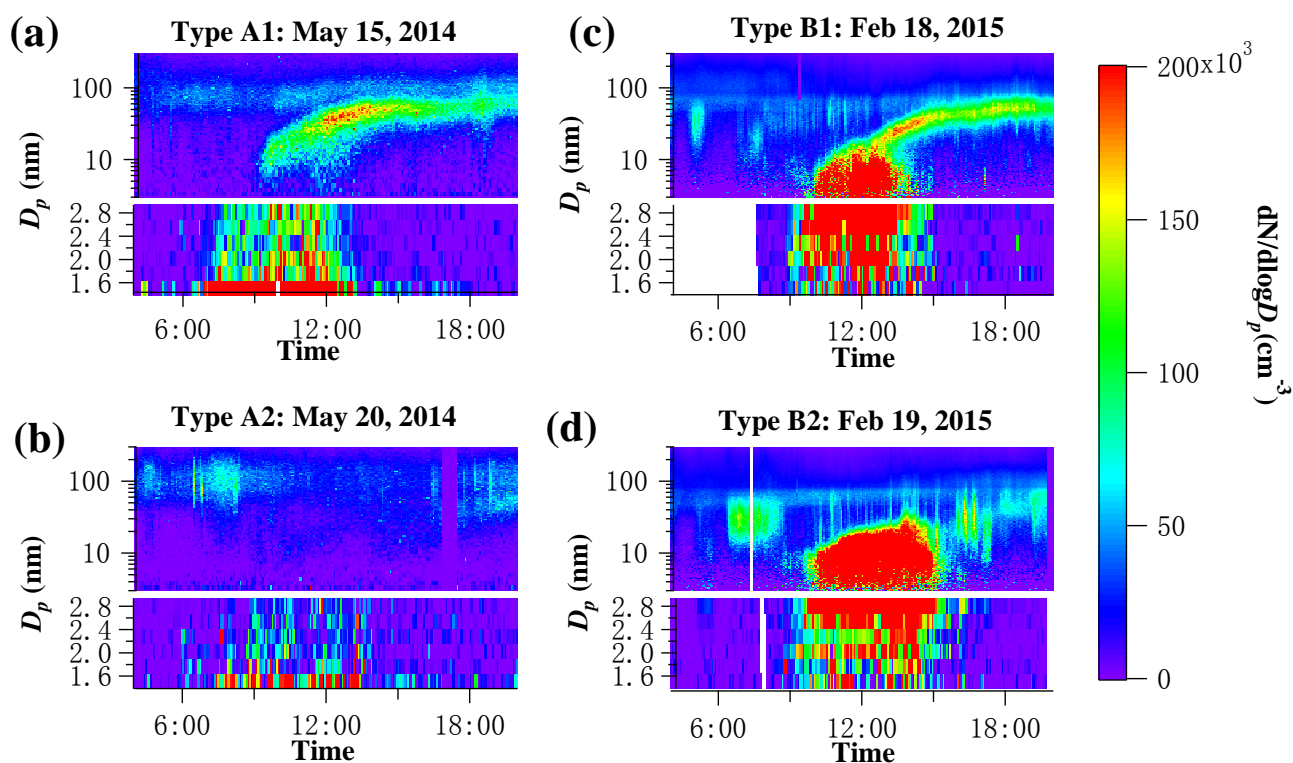
Figure 1. Locations of two urban measurement sites in Nanjing, the second largest megacity in the Yangtze River Delta region, China. ① is the NUIST site and ② is the summer measurement site.



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900 Figure 2. N_{sub-3} vs. $J_{1.4}$ in the 8 nucleation events in February, May, December and August during
 901 2014-2015. The events were indicated by different colors (blue: March 1, 2015; green: February
 902 18, 2015; red: February 19, 2015; purple: August 15, 2014; black: December 27, 2014; grey:
 903 May 15, 2014; orange: May 20, 2014; yellow: May 16, 2014)

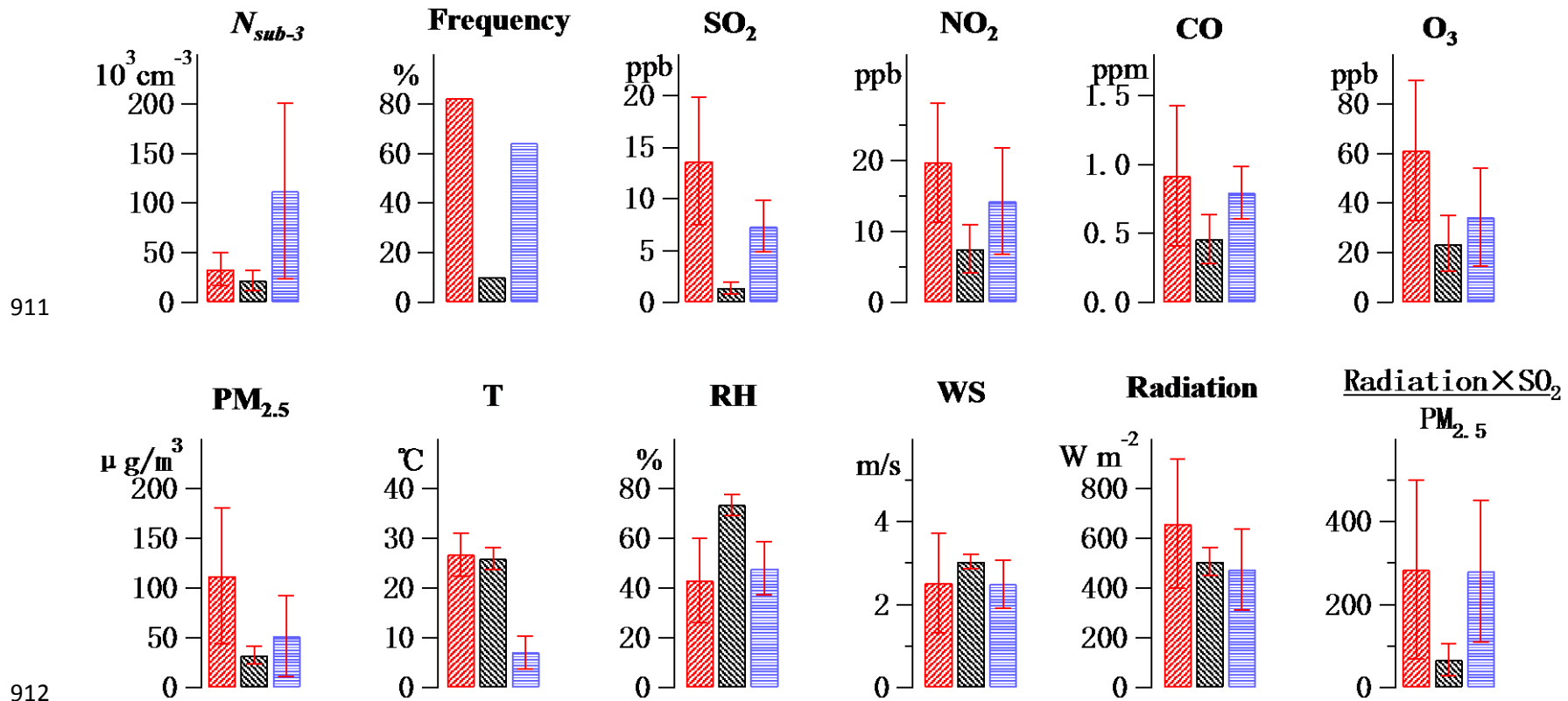
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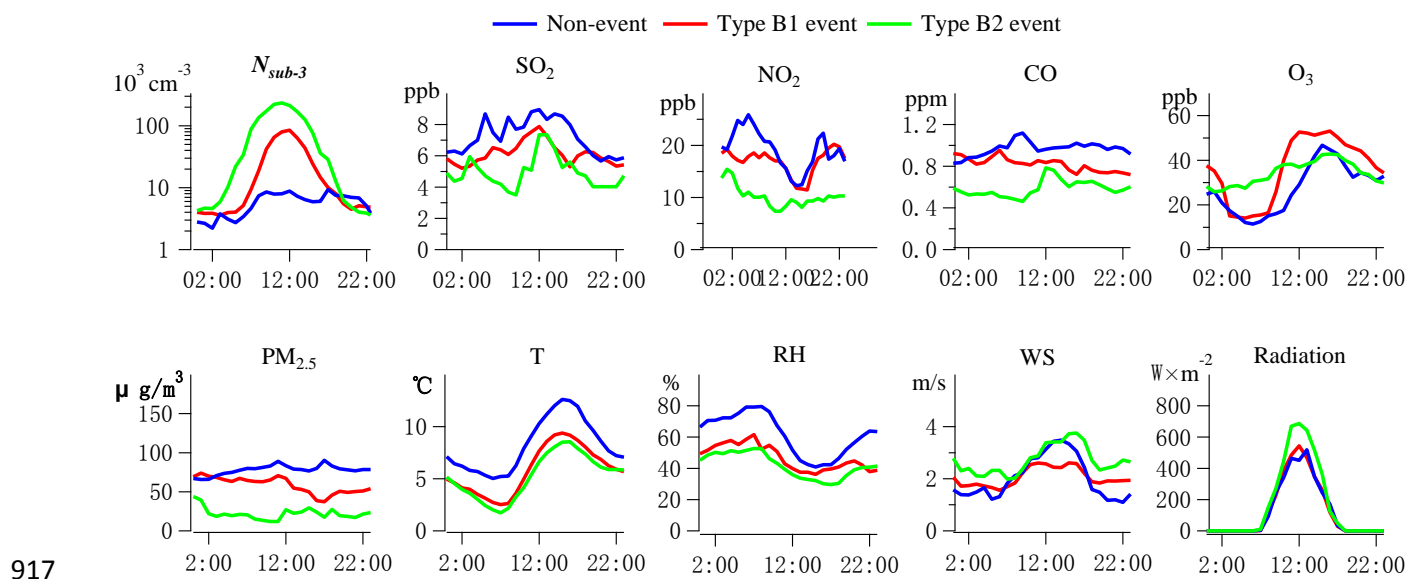
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907 Figure 3. Size spectra of typical (a) Type A1 event on May 15, 2014; (b) Type A2 event on May 20,
 908 2014; (3) Type B1 event on February 18, 2015 and (d) Type B2 event on February 19, 2015 during
 909 our measurement period. Size spectra from 3-300 nm (logarithmic scale) and 1.4-3 nm (linear scale)
 910 were obtained using SMPS and nCNC, respectively.

Spring Summer Winter



913 Figure 4. Mean and standard deviation of event-averaged N_{sub-3} , anthropogenic trace gases (SO_2 , NO_2 , CO and O_3), $\text{PM}_{2.5}$, and meteorological
 914 variables (temperature, RH, wind speed (WS), solar radiation, and radiation $\times \text{SO}_2 / \text{PM}_{2.5}$) for nucleation events in spring (n=17), summer (n=3)
 915 and winter (n=14). Nucleation frequency (the percentage of event days out of total measurement days) was also shown.



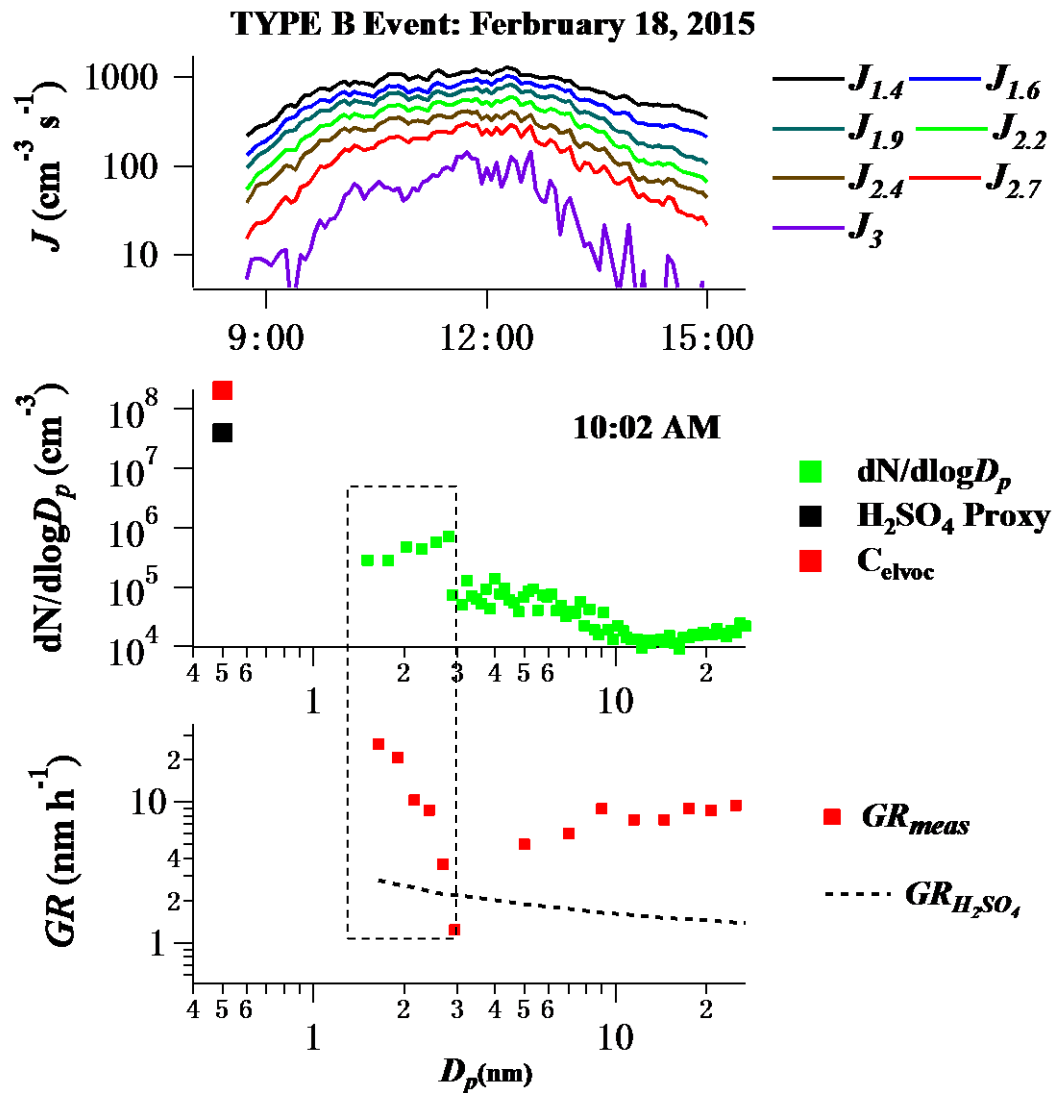
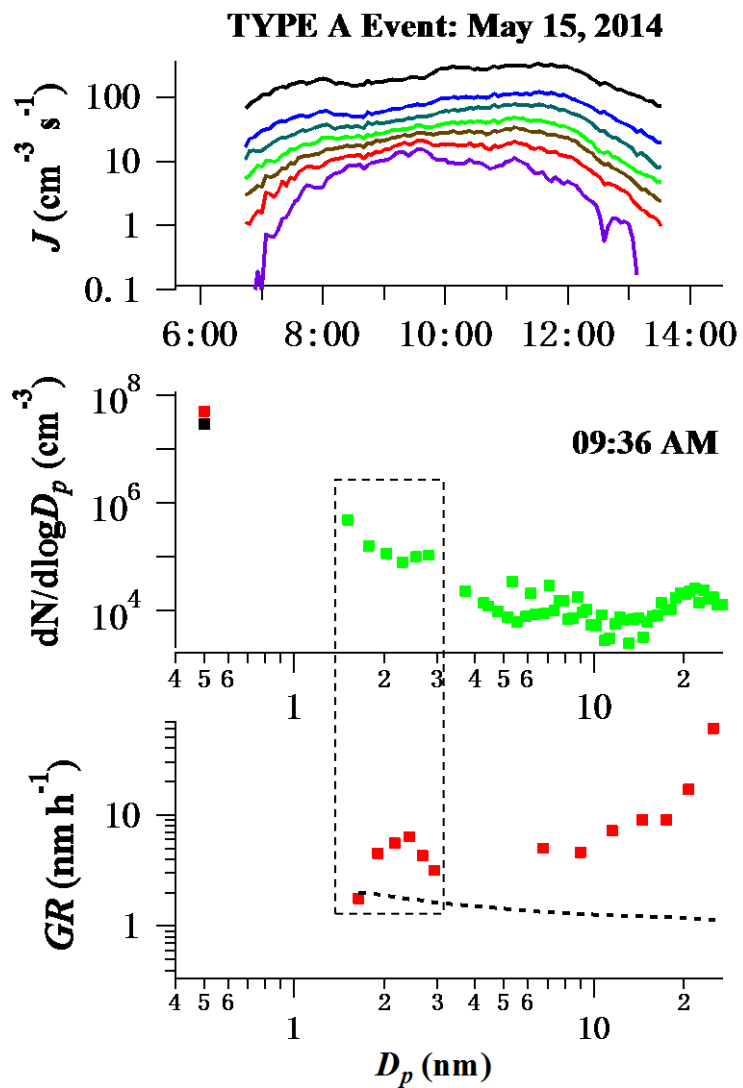
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918 Figure 5. Diurnal variations of mean N_{sub-3} , anthropogenic trace gases (SO_2 , NO_2 , CO and O_3), $\text{PM}_{2.5}$,

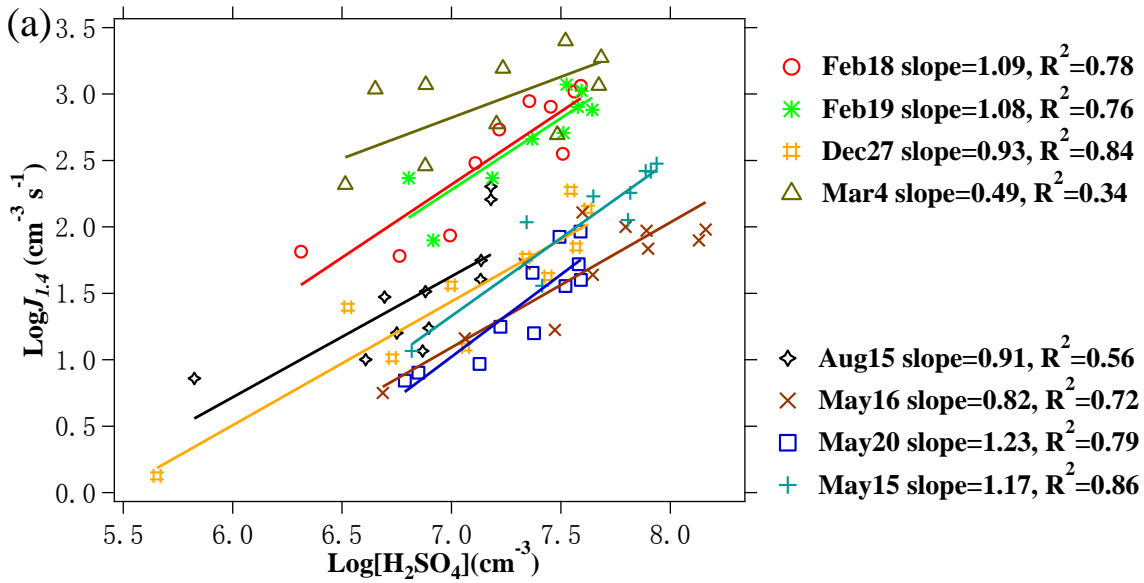
919 and meteorological variables (temperature, RH, wind speed, and solar radiation flux) on non-event

920 days (n=8, blue line) and event days (n=3 for Type B1 event, red line and n=6 for Type B2 event,

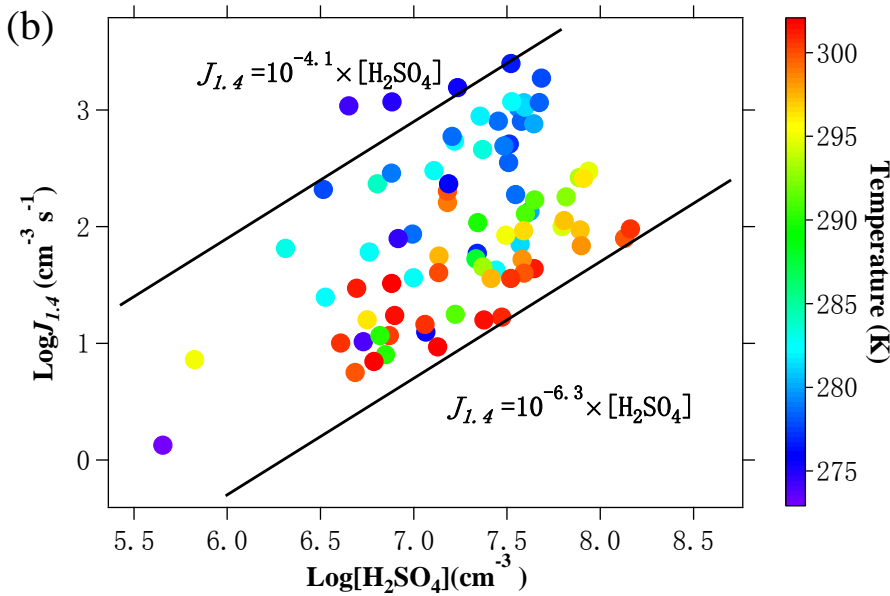
921 green line) during winter measurement period.



923 Figure 6. Upper: formation rates (or equivalently, particle growth fluxes) of 1.4, 1.6, 1.9, 2.2, 2.4, 2.7 and 3.0 nm cluster/particles on May 15
924 2014 (Type A1 event) and Feb 18 2015 (Type B1 event). Middle: particle size distribution ($dN/d\log D_p$, green square) selected during the two
925 events (9:36 AM and 10:02 AM). Lower: particle growth rates measured during the same time periods (GR_{meas} , red square). Also shown in the
926 figure were H_2SO_4 proxy (black square) and growth rates calculated from the H_2SO_4 proxy ($GR_{H_2SO_4}$, dashed black line), as well as the
927 calculated ELVOC concentration (C_{elvoc} , red square, see Eq.5) during the same time periods. Dashed boxes in the lower panels highlighted the
928 size distributions and growth rates between 1.4 and 3 nm measured with nCNC.

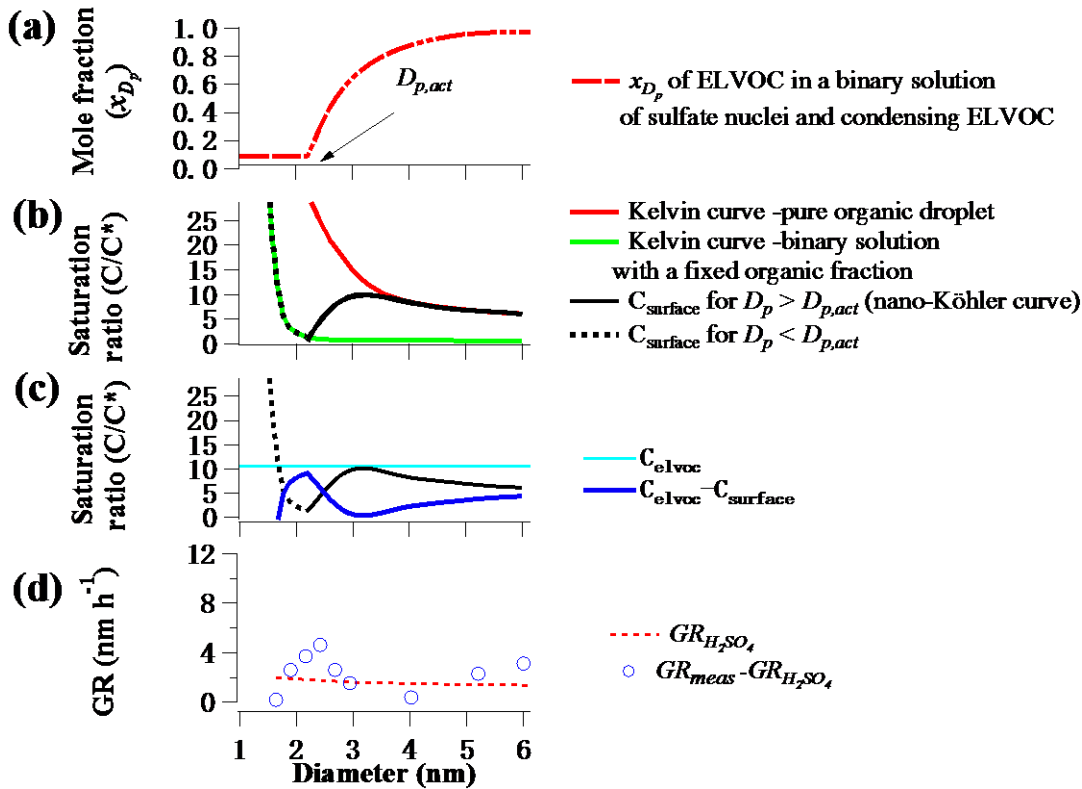


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931 Figure 7. (a) Correlations between $\log J_{1.4}$ and $\log [H_2SO_4]$ for the 8 events. H_2SO_4 proxy was
 932 calculated according to Mikkonen et al. (2011). $J_{1.4}$ and $[H_2SO_4]$ were synchronized to 1 hour that
 933 was the time resolution of solar radiation data. The colored lines showed linear fits to the data of
 934 every single event. (b) The same dataset as (a), but with symbol color to indicate ambient
 935 temperature. Two black lines showed the linear dependences of $J_{1.4}=10^{-4.1} \times [H_2SO_4]$ and $J_{1.4}=10^{-6.3} \times$
 936 $[H_2SO_4]$, between which most of data points located.



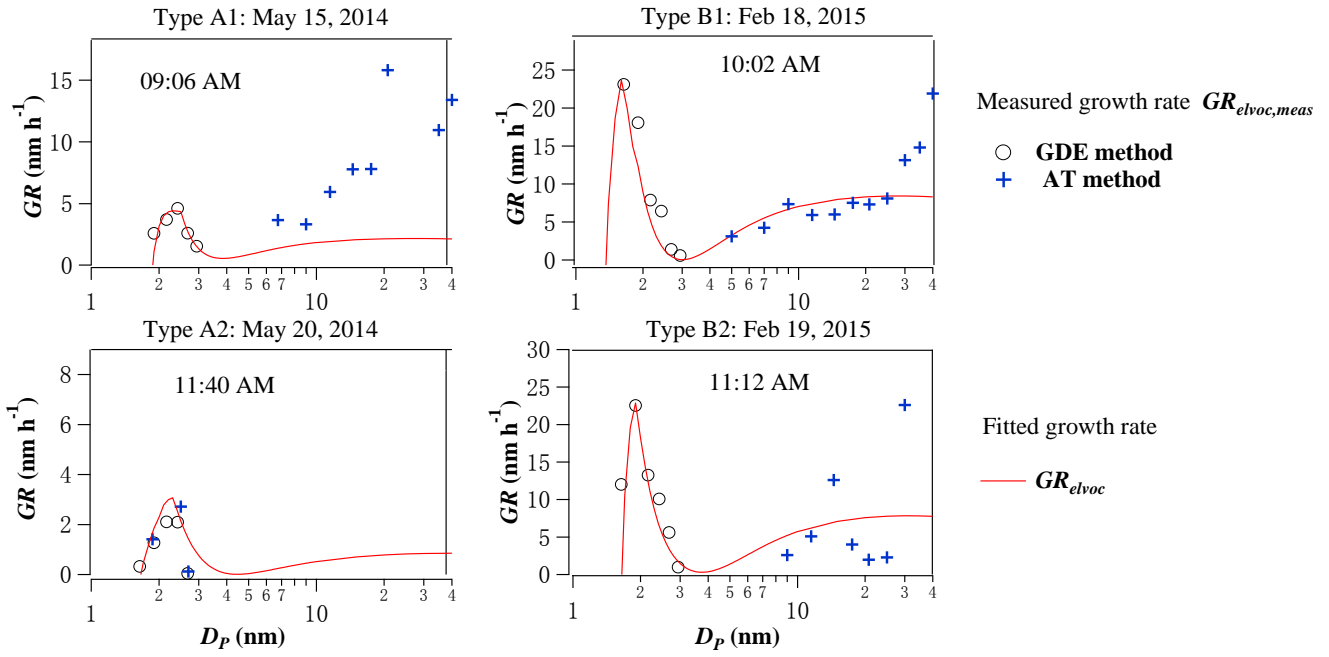
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938 Figure 8. (a) Mole fraction of organics (x_{D_p}) in a binary solution of sulfate nuclei and activating
 939 organics (ELVOC) in a new particle. Nuclei activation diameter $D_{p,act}$ is the size where ELVOC
 940 begins to dilute the nuclei. (b) Kelvin equilibrium curves over a pure organic droplet (red line) and a
 941 binary solution with a fixed organic fraction (green line), nano Köhler curve for $D_p > D_{p,act}$ (black
 942 solid line) and surface concentration $C_{surface}$ for $D_p < D_{p,act}$ (black dashed line). (c) gas phase
 943 concentration of the organic vapor (C_{elvoc} , cyan line), surface concentration C_{∞} (black line), and
 944 $C_{elvoc} - C_{surface}$ (blue line). (d) Growth rate $GR_{H_2SO_4}$ due to H_2SO_4 (Mikkonen et al. 2011 proxy)
 945 and growth rate due to organic vapor $GR_{elvoc,meas}$, calculated as $GR_{meas} - GR_{H_2SO_4}$.

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951 Figure 9. Comparisons of measured ($GR_{elvoc,meas}$, black circle) and fitted (GR_{elvoc} , red line) growth

952 rates from Eq. (4) for typical Type A1, A2, B1, and B2 events. Also shown were growth rates

953 calculated from appearance time method (blue cross) for sub-3 nm particles when growth rate was

954 relatively small or for larger particles with large size intervals.

955