Comments "...The iterative method to determine the isomer-specific branching ratios is quite critical to the accuracy of the overall analysis. While I am comfortable with the description of the very careful efforts to understand the sources of calibration errors, the manuscript would be improved with a more extended discussion of this method and the results..."

5

9

More description of the iterative method was added to page 11 (lines 307-311) of the revised
manuscript. To further clarify the iterative method, we have added more detailed description and
a flow chart to the supplement section 3.1.

Comments "...In particular, I am intrigued by Supplement Figure S8, which shows the modeldetermined isomer-specific branching ratios. While it seems to me that this figure contains very
important results, I could not find a reference to this figure in the main manuscript..."

Reference to this figure was added to the revised manuscript page 7 (line 201). Please note
Figure S8 is now Figure S9 because we added a new figure to describe the iterative method.

16 17

18 Comments "...In particular, Figure S8 shows that there are only two significant isomers formed 19 (1,2-IN and 4,3-IN), which is quite a different result than that reported by both the CIMSbased (Paulot et al. 2009) and GC-based (Lockwood et al. 2010) previous laboratory studies. The new 20 finding suggests that the OH-initiated oxidation of isoprene is more regiospecific than these 21 22 previous studies indicated, which is interesting (and simplifying with respect to modeling isoprene oxidation). In particular, both of the previous studies identified the various 1,4- and 4,1-23 24 HNI isomer-producing reaction pathways as having minor, yet significant, branching ratios, but 25 Figure S8 indicates that they are very minor products. It would be helpful to explain what part(s) of the analysis are responsible for the difference as compared to the two previous laboratory 26 27 studies..."

28

Figure S8, which is now Figure S9 in our revised supplement, shows the results of isomer
distribution during the SOAS field study. In comparison, the previous work by Paulot et al.
(2009) and Lockwood et al. (2010) are focused on experimental chamber studies.

32

33 The difference can be explained from two aspects. First, the isoprene nitrates in the field undergo 34 more oxidation degradation than in the reaction chamber. As the 1,4 and 4,1 nitrate isomers have 35 larger OH and O<sub>3</sub> oxidation rate constants than the 1,2 and 4,3 isomers, the 1,4 and 4,1 nitrate are 36 consumed faster in the field, making their contribution to the total IN pool smaller. The second factor is 1,6-H shift of cis-1,4 and cis-4,1 RO<sub>2</sub> isomers and RO<sub>2</sub> interconversion. During the 37 38 SOAS field study, the NO concentration is in the range that allows a significant amount of cis-39 1,4 and cis-4,1 RO<sub>2</sub> to undergo H-shift, which reduces the amount of 1,4- and 4,1- RO<sub>2</sub> available to react with NO and form 1,4- and 4,1-nitrates. In chamber studies, however, NO concentrations 40 are in the range that force all the RO<sub>2</sub> radicals to react with NO, making 1,4- and 4,1-nitrates 41 more important. This is clarified in the revised manuscript on page 14 (lines 387-392). 42

- 43
- 44
- 45
- 46

- p. 17851 line 25: I'm not familiar with the scrubbing technique described here. Please provide a
  reference to the method and indicate what species are being removed.
- 50

- 51 The method and citation are added to the revised manuscript on page 9 (lines 253-256).
- 52

p. 17853 line 3: Since the relative quantity [IN]/([MVK]+[MACR] is the critical quantity being
analyzed, it would be helpful to know how much the absolute quantity ([MVK]+[MACR]) varied
in the SOAS data being analyzed here, as larger variations in the absolute value would have the
potential to cause more uncertainty in the relative value. Additionally, can the absolute variations
in [MVK]+[MACR] be qualitatively explained according to expected changes in transport and

- 58 boundary height?
- 59

60 The [MVK]+[MACR] measurement during SOAS has a 40% measurement uncertainty. This is61 added to page 18 (line 507) of the revised manuscript.

62

63 The absolute diurnal concentrations of MVK+MACR are more consistent with the dynamics of 64 photochemistry, with its peaking concentration at around 1:00 PM when radiation was strong, and [isoprene] is large. The changes in MVK+MACR concentrations are also affected by 65 66 transport and changing boundary layer height. Since we did not apply three-dimensional transport model in this work, we were unable to quantify the influence of transport and boundary 67 layer height on the absolute MVK+MACR concentrations. However, the relative concentrations 68 69 of IN to MVK+MACR can provide some insight into the isoprene oxidation chemistry, since 70 these meteorological factors influence all three compounds. We have discussed in more detail the 71 MVK+MACR diurnal profile on page 13 (lines 361-364) of the revised manuscript.

72 73

p. 17863 line 4: Is the 13% branching ratio OH + HNI pathway to IEPOX under the low NO
conditions a significant source of IEPOX in the afternoon, or does the modeling indicate that
ISOPOOH is still the dominant precursor of IEPOX?

77

During the SOAS study, the 24-hour averaged peak is around 80 ppt for isoprene nitrates and
300 ppt for ISOPOOH. The yield of IEPOX is 13% for isoprene nitrate oxidation and ~ 75% for
ISOPOOH oxidation (St. Clair et al., 2015). Therefore, ISOPOOH is still the dominant precursor

- 81 for IEPOX. This is now discussed in the revised manuscript on page 21 (lines 609-611).
- 82
- 83 Technical corrections
- p. 17848 line 14: What is the solvent used in the IN solutions?
- 85

87

86 Information on the solvent  $(C_2Cl_4)$  is now added on page 5 (line 131).

p. 17851 line 14: change to ". . . , so wall loss correction was not applied to the IN
measurements."

90

91 We have now re-worded the sentence on page 9 (line 240).

p. 17856 equation 2: the alpha term in this equation should be defined immediately as the nitrate 95 96 branching ratio. 97 98 Description of the alpha term is now added on page 14 (line 399). 99 p. 17859 line 24: typo: "efficiency" 100 101 102 Typo has been corrected on page 18 (line 514). 103 p. 17862 line 9: typo: "different" 104 105 106 Typo has been corrected on page 20 (line 584).

107

108

109 Reference

St. Clair, J. M., Rivera, J. C., Crounse, J. D., Knap, H. C., Bates, K. H., Teng, A. P., Jørgensen, S.,

111 Kjaergaard, H. G., Keutsch, F. N., and Wennberg, P. O.: Kinetics and Products of the Reaction of

112 the First-Generation Isoprene Hydroxy Hydroperoxide (ISOPOOH) with OH, The Journal of

113 Physical Chemistry A, 10.1021/acs.jpca.5b06532, 2015.

The authors present a combination of laboratory and field measurements of isoprene nitrates plus related compounds. They use a 0-D chemical model to interpret the data in terms of constraints on isoprene nitrate formation and fate in the atmosphere. This paper is well written and makes a valuable contribution to our developing understanding of isoprene nitrate chemistry in the atmosphere. It should be published. Comments and suggestions are listed below, and are generally minor.

8 Thank you.

9

7

10 17857, 25: You could still have some IN production from NO3 + isoprene in the NBL. NO is not 11 so high at night that it would dominate NO3 loss. The ratio of the rate coefficients 12 k(NO3+NO)/k(NO3+isoprene) is about 37. Fig 6 shows that NOx at midnight was around 0.6 13 ppb, but this is almost all NO2, NO at midnight averaged <10 ppt. On the other hand isoprene 14 was >2ppb at midnight. Based on that, isoprene + NO3 would be 5-6 times faster than NO + 15 NO3 at that time. Would this affect your interpretation of the diurnal IN cycle? Would the CIMS 16 detect any of the NO3 + isoprene nitrates with the daytime products, or no?

17

The CIMS would detect the hydroxynitrates (IN) from NO3 reaction with isoprene near ground.
Our IN raw data during the SOAS study was interpreted with an IN isomer distribution estimated
by a 0D model. Within this model, IN production from NO3 chemistry was also included,
besides IN formation from the OH + isoprene pathway. Therefore, we have already considered
the IN formation form NO3 chemistry in our data analysis. This is clarified on page 16 (lines
456-461) of the revised manuscript.

24

17848, 15-28. Please discuss why you believe it is reasonable to use the sensitivity for 4,3-IN as
a surrogate for all other beta INs (aside from 1,2). And likewise using cis/trans-1,4 for cis/trans4,1.

27 4 28

29 The CIMS sensitivities toward alkyl alcohol and alkyl nitrates are both around 5 orders of magnitude smaller than its sensitivity toward the isoprene hydroxynitrates. Hence, it is the 30 combination of the OH group and the nitrate group, as well as their relative positions that has the 31 dominant influence on the CIMS sensitivity, which will affect how the molecule binds with the 32 iodide ion, while the structure of the carbon backbone would have little effect. For the IN 33 34 isomers, the relative positions of the OH group and the nitrate group are  $\alpha,\beta$  position, trans- $\alpha,\delta$ 35 position and cis- $\alpha$ ,  $\delta$  position. We assume the same sensitivity can be applied to isomers within each structural group, namely  $\beta$ -isomers, trans- $\delta$  isomers and cis- $\delta$  isomers. This assumption is 36 consistent with our observation of identical sensitivity for 1,2-IN and 4,3-IN isomers when water 37 38 is not added to the CIMS. For the case with water addition to CIMS, the smaller sensitivity of 1,2-IN was caused by the smaller amount of 1,2-IN available for detection, as 1,2-IN is lost 39 inside the instrument, rather than from a fundamental difference in the ionization efficiency of 40 1,2-IN. Primary nitrates ( $\delta$ -IN, 3,4-IN, and 2,1-IN) and secondary nitrate (4,3-IN) are not as 41 likely to be affected by water (Hu et al., 2011). We have now clarified these assumptions on 42 page 7 (lines 174-184) of the revised manuscript. 43

44

17854, 5. Some significant assumptions had to be made for the instrumental sensitivities for the
 various IN isomers. Only a couple isomers were actually synthesized and calibrated directly, and

the derived sensitivities for those range over a factor of 7. These were then used to estimate the
sensitivity for other IN isomers. Given that, I find it hard to believe that the reported 15%
uncertainty for IN measurements is even close to realistic.

50

51 While the sensitivities of the IN isomers vary greatly, the relative contribution of each isomer to 52 the total pool of IN is well understood from the theoretical work on the formation of isoprene 53 RO<sub>2</sub> and experimental studies that measured MVK and MACR formed simultaneously with IN 54 in isoprene oxidation. The measurement uncertainty for IN comes from 1) the measurement 55 uncertainty for each individual isomer obtained in our calibrations and 2) the uncertainty in the 56 relative abundance of each isoprene nitrate, caused by uncertainty in the kinetics data of the 57 LIM1 mechanism (Peeters et al., 2014). 15% uncertainty is associated with instrument calibration only, and the IN measurement uncertainty is around 23%, with both factors 58 59 considered. We have provided a more detailed discussion on the estimated IN data uncertainty in 60 supplement section 5 and error propagation is clarified on page 12 (lines 329-339) of the revised 61 manuscript.

- 62
- 63

64 17849, 20-25. Do we know that 1,2-IN is the only isomer affected by humidity in this way? If
65 that's not the case would that affect your interpretation of the field data? Likewise (17854, 13)
66 do we know that 1,2 is the only isomer affected by inlet loss?

67

Hu et al. (2011) suggests that primary and secondary organic nitrates do not undergo hydrolysis under neutral condition, but hydrolysis in neutral pH is efficient for tertiary nitrates. Given the fact that the 1,2-IN is the only tertiary nitrate and the short residence time for the ion-molecule reaction, we do not expect the other nitrates to be affected by humidity. We have discussed the humidity sensitivity of IN isomers on page 7 (lines 184-193) of the revised manuscript.

73

For inlet loss, we have conducted new experiments and found the inlet loss to be around 5% on
average for all the isomers, and our data has been corrected for the 5% inlet loss. This is
described on page 8 (line 213) of the revised manuscript.

77 78

17851, 16-28. Was the sample stream humidified for the field measurements as it was for the labdata? If not, how do you correct for humidity-dependent losses of the various isomers?

81
82 The sample was humidified during the field measurement with the same conditions used in the
83 lab experiments. This information is added on page 9 (line 219) of the revised manuscript.

- 84
- 85

Fig. 7 is well done. The diurnal trend in gamma shows nicely how the fractional importance of
RO2 + NO versus other RO2 sinks varies during the day. It may also be worth pointing out that
the calculated IN production rate (Fig 7b) is not as peaked in the morning as the RO2 loss rate to
RO2 + NO (Fig 7a) or gamma. And that this is because the RO2 concentrations are still quite
low at that time. Right?

92 Right. When the  $\gamma$  value peaks around 6 AM to 8 AM, the isoprene and OH concentrations are 93 relatively low. Therefore, the nitrate production rate is limited by the availability of RO<sub>2</sub>. In the 94 afternoon, while only a fraction of RO<sub>2</sub> will react with NO to form organic nitrates, the total 95 amount of isoprene RO<sub>2</sub> is higher due to higher isoprene and OH concentrations. Therefore, the 96 IN production rate does not have an abrupt decrease in the afternoon as the  $\gamma$  value does. This is 97 further clarified in the revised manuscript on page 14 (lines 407-412).

- 98 99
- Fig 8, 17859-17860. You consider the possibility of fast photolysis to explain the day time IN
   decrease. Could heterogeneous losses / deposition be playing a role?
- 102

Aerosol formation may be a small sink for IN, but the C5 hydroxynitrate is unlikely to undergo efficient aerosol uptake (Surratt et al., 2010) due to its small molecular weight. Field observations during SOAS suggest that the isoprene-derived SOA components were associated with IEPOX and more oxidized organic nitrates, not the first-generation hydroxynitrates (Lee et al., 2015; Xu et al., 2015). Therefore, we did not consider aerosol uptake as a major factor for the overestimation of [IN]/([MVK]+[MACR]) ratio in the afternoon. This is discussed on page 22 (lines 663-666) of the revised manuscript.

- 17849, 1-8. It seems this method could be used as a check on the relative sensitivities for all IN
  isomers. Is it just that 4,3 and 1,2 are the only ones detected with enough signal for the purpose?
- The small abundance of the other isomers makes it more difficult to obtain reliable quantification
  through this method. Therefore, the sensitivity was obtained by synthesizing a standard. This
  discussion is added to the revised manuscript on page 6 (line 171-173).
- 117 118

110

17853, 5. Should say "partially cancel the influence of dilution" or "reduce the influence".
Transport could easily affect MVK+MACR differently than INs, just based on differing lifetimes
and concentrations in the residual layer, advected air masses, etc.

- 123 The change has been made on the revised manuscript on page 10 (line 289).
- 124

122

17853, 11-18. Need a description or at least a mention and citation for the other measurementsused (PTR, GC).

127

128 Citation for PTR and GC measurements have been added to the revised manuscript on page 11129 (line 295 and line 297).

130

131 17854, 13-18. Reported 9% IN yield is "in the 4-14% range of IN yields determined from
132 previous experiments (. . .) but is more consistent with determinations from the higher end of the
133 range". Actually it seems to be exactly in the middle.

- 135 We have made the adjustment on the revised manuscript on page 12 (line 342).
- 136

134

- 138 References
- 139

Hu, K. S., Darer, A. I., and Elrod, M. J.: Thermodynamics and kinetics of the hydrolysis of atmospherically relevant organonitrates and organosulfates, Atmospheric Chemistry and Physics, 11, 8307-8320, 10.5194/acp-11-8307-2011, 2011.

Lee, B. H., Mohr, C., Lopez-Hilfiker, F. D., D'Ambro1, E. L., Lutz, A., Hallquist, M., Lee, L., Romer,

- 144 P., Cohen, R. C., Iyer, S., Kurten, T., Hu, W. W., Day, D. A., Campuzano-Jost, P., Jimenez, J. L., Xu,
- L., Ng, N. L., Wild, R. J., Brown, S. S., Koss, A., Gouw, J. d., Olson, K., Goldstein, A. H., Seco, R.,
- 146 Kim, S., McAvey, K., Shepson, P. B., Baumann, K., Edgerton, E. S., Nguyen, T. B., Wennberg, P. O.,
- 147 Liu, J., Shilling, J. E., and Thornton, J. A.: Highly functionalized particle-phase organic nitrates
- observed in the Southeastern U.S.: contribution to secondary organic aerosol and reactivenitrogen budgets, in preparation, 2015.
- Peeters, J., Müller, J.-F., Stavrakou, T., and Nguyen, V. S.: Hydroxyl Radical Recycling in Isoprene
  Oxidation Driven by Hydrogen Bonding and Hydrogen Tunneling: The Upgraded LIM1
  Machanism The Journal of Division Chemistry A 118, 8625, 8642, 10, 1021 (in 5032146, 2014)
- 152 Mechanism, The Journal of Physical Chemistry A, 118, 8625-8643, 10.1021/jp5033146, 2014.
- Surratt, J. D., Chan, A. W. H., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S. P.,
  Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in secondary
  organic aerosol formation from isoprene, Proceedings of the National Academy of Sciences of
- the United States of America, 107, 6640-6645, 10.1073/pnas.0911114107, 2010.
- 157 Xu, L., Suresh, S., Guo, H., Weber, R. J., and Ng, N. L.: Aerosol characterization over the 158 southeastern United States using high-resolution aerosol mass spectrometry: spatial and
- 159 seasonal variation of aerosol composition and sources with a focus on organic nitrates,
- 160 Atmospheric Chemistry and Physics, 15, 7307-7336, 10.5194/acp-15-7307-2015, 2015.

- 1 Observation of Isoprene Hydroxynitrates in the Southeastern
- 2 United States and Implications for the Fate of NO<sub>x</sub>
- 3
- 4 F. Xiong<sup>1</sup>, K. M. McAvey<sup>1</sup>, K. A. Pratt<sup>1,3</sup>, C. J. Groff<sup>1</sup>, M. A.
- 5 Hostetler<sup>1</sup>, M. A. Lipton<sup>1</sup>, T. K. Starn<sup>4</sup>, J. V. Seeley<sup>5</sup>, S. B. Bertman<sup>6</sup>,
- 6 A. P. Teng<sup>7</sup>, J. D. Crounse<sup>8</sup>, T. B. Nguyen<sup>8</sup>, P. O. Wennberg<sup>7,8</sup>, P. K.
- 7 Misztal<sup>9</sup>, A. H. Goldstein<sup>9,10</sup>, A. B. Guenther<sup>11</sup>, A. R. Koss<sup>12,13</sup>, K. F.
- 8 Olson<sup>9</sup>, J. A. de Gouw<sup>12,13</sup>, K. Baumann<sup>14</sup>, E. S. Edgerton<sup>14</sup>, P. A.
- 9~ Feiner^{15}, L. Zhang^{15}, D. O. Miller^{15}, W. H. Brune^{15} and P. B.
- 10 Shepson<sup>1,2</sup>
- 11 [1] Department of Chemistry, Purdue University, West Lafayette, IN
- 12 [2] Department of Earth, Atmospheric and Planetary Sciences, Purdue University, West
- 13 Lafayette, IN
- 14 [3] Department of Chemistry, University of Michigan, Ann Arbor, MI
- [4] Department of Chemistry, West Chester University of Pennsylvania, West Chester,PA
- 17 [5] Department of Chemistry, Oakland University, Rochester, MI
- 18 [6] Department of Chemistry, Western Michigan University, Kalamazoo, MI
- 19 [7] Division of Engineering and Applied Science, California Institute of Technology,
- 20 Pasadena, CA
- 21 [8] Division of Geophysical and Planetary Sciences, California Institute of Technology,
- 22 Pasadena, CA
- 23 [9] Department of Environmental Science, Policy, & Management, University of
- 24 California at Berkeley, Berkeley, CA
- 25 [10] Department of Civil and Environmental Engineering, University of California at
- 26 Berkeley, Berkeley, CA

- 27 [11] Atmospheric Sciences and Global Change Division, Pacific Northwest National
- 28 Laboratory, Richland, WA
- 29 [12] Cooperative Institute for Research in Environmental Sciences, Boulder, CO
- 30 [13] NOAA Earth System Research Laboratory, Boulder, CO
- 31 [14] Atmospheric Research & Analysis, Inc., Cary, NC
- 32 [15] Department of Meteorology, Pennsylvania State University, University Park, PA
- 33 Correspondence to: P. B. Shepson (pshepson@purdue.edu)
- 34

## 35 Abstract

- 36 Isoprene hydroxynitrates (IN) are tracers of the photochemical oxidation of isoprene in
- 37 high NO<sub>x</sub> environments. Production and loss of IN have a significant influence on the
- 38 NO<sub>x</sub> cycle and tropospheric O<sub>3</sub> chemistry. To better understand IN chemistry, a series of
- 39 photochemical reaction chamber experiments wasere conducted to determine the IN yield
- 40 from isoprene photooxidation at high NO concentrations (>100 ppt). By combining
- 41 experimental data and calculated isomer distributions, a total IN yield of 9(+4/-3)% was
- 42 derived. The result was applied in a zero-dimensional model to simulate production and
- 43 loss of ambient IN observed in a temperate forest atmosphere, during the Southern
- 44 Oxidant and Aerosol Study (SOAS) field campaign, from May 27 to July 11, 2013. The 9%
- 45 yield was consistent with the observed IN/(MVK+MACR) ratios observed during SOAS.
- 46 By comparing field observations with model simulations, we identified NO as the
- 47 limiting factor for ambient IN production during SOAS, but vertical mixing at dawn
- 48 might also contribute (~27%) to IN dynamics. A close examination of isoprene's
- 49 oxidation products indicates that its oxidation transitioned from a high-NO dominant
- 50 chemical regime in the morning into a low-NO dominant regime in the afternoon. A
- 51 significant amount of IN produced in the morning high NO regime could be oxidized in
- 52 the low NO regime, and a possible reaction scheme was proposed.

### 53 1 Introduction

54 Isoprene (C5H8) accounts for approximately half of the global non-methane biogenic 55 volatile organic compound (BVOC) emissions (Guenther et al., 2006) and has a significant influence on the budgets of OH, O3 and NOx (Horowitz et al., 2007). Isoprene 56 57 oxidation by OH in the presence of  $NO_x$  can lead to the formation of isoprene 58 hydroxynitrates (IN), as described in Reactions (R1) and (R2). The chain-terminating 59 Reaction (R2a) removes peroxy radicals ( $RO_2$ ) and NO from the atmosphere and 60 decreases tropospheric O<sub>3</sub> production (Carter and Atkinson, 1996). IN serves as a 61 temporary  $NO_x$  reservoir, and the transport and photo-oxidative decomposition of these 62 compounds can further modulate  $NO_x$  and  $O_3$  concentrations (Horowitz et al., 2007; 63 Paulot et al., 2012; Xie et al., 2013). Gas-phase organic nitrates can also partition into the particle phase and undergo hydrolysis, contributing to the growth of secondary organic 64 aerosols (SOA) (Jacobs et al., 2014; Rindelaub et al., 2015). 65

66 isoprene + OH (+ O<sub>2</sub>) 
$$\rightarrow$$
 RO<sub>2</sub> (R1)

$$67 \quad \text{RO}_2 + \text{NO} \rightarrow \text{RONO}_2 \tag{R2a}$$

$$68 \quad \mathrm{RO}_2 + \mathrm{NO} \rightarrow \mathrm{RO} + \mathrm{NO}_2 \tag{R2b}$$

69 The initial OH addition (followed by  $O_2$ ) to isoprene (Reaction R1) produces eight 70 isomeric RO2 radicals. Reaction of these RO2 radicals with NO proceeds primarily via 71 two reaction pathways (Reaction R2a and b). Laboratory studies suggest that the nitrate 72 formation channel (Reaction R2a) is minor compared to the alkoxy radical (RO) 73 formation channel (Reaction R2b), with reported total IN yields ranging from 4% to 14% 74 (Chen et al., 1998; Patchen et al., 2007; Lockwood et al., 2010; Paulot et al., 2009; 75 Sprengnether et al., 2002; Tuazon and Atkinson, 1990). Reaction (R2a) leads to the 76 formation of eight IN isomers, including four  $\beta$ -IN isomers and four  $\delta$ -IN isomers (Table 77 1). The wide range of reported IN yields has led to uncertainty in quantifying isoprene's 78 influence on the NO<sub>x</sub> cycle and  $O_3$  enhancement (Xie et al., 2013; Horowitz et al., 2007; 79 Paulot et al., 2012). Isoprene hydroxynitratesN can also be produced at night through 80  $NO_3$ -initiated isoprene oxidation with a yield around 20%, adding to a total organic nitrate yield of 65-70% (Rollins et al., 2009; Perring et al., 2009; Kwan et al., 2012). The 81 major daytime IN sink is reaction with OH, which leads to a lifetime of 2.5 to 6.5 hours, 82

83 according to a recent kinetics study (Lee et al., 2014b). At night, IN is more susceptible

84 to loss from ozonolysis, and potentially NO<sub>3</sub> oxidation when the NO<sub>x</sub> concentration is

high (Xie et al., 2013). IN have been observed in the ambient environment, primarily in

86 forested areas under the influence of anthropogenic NO<sub>x</sub> plumes (Grossenbacher et al.,

87 2001; Giacopelli et al., 2005; Grossenbacher et al., 2004; Beaver et al., 2012; Lee et al.,

88 2014a). During the BEARPEX 2009 study conducted in the Sierra Nevada Mountains of

89 California, IN constituted 38% of the total organic nitrates (Beaver et al., 2012).

90 Methods to quantify organic nitrates include infrared spectroscopy (IR), thermal 91 dissociation-laser induced fluorescence (TD-LIF) spectroscopy, chemiluminescence, gas 92 chromatography (GC)-based separation and detection techniques, and mass spectrometry 93 (MS) (Rollins et al., 2010; Tuazon and Atkinson, 1990; Sprengnether et al., 2002; Day et al., 2002; O'Brien et al., 1995; Beaver et al., 2012; Lee et al., 2014a; Lockwood et al., 94 95 2010; Paulot et al., 2009; Giacopelli et al., 2005; Grossenbacher et al., 2004; Patchen et 96 al., 2007; Hartsell et al., 1994; Kwan et al., 2012; Teng et al., 2015). IR, TD-LIF, and 97 chemiluminescence can only measure total organic nitrates because they respond solely to the nitrooxy functional group (Day et al., 2002; Rollins et al., 2010; Tuazon and 98 99 Atkinson, 1990; Sprengnether et al., 2002; O'Brien et al., 1995; Hartsell et al., 1994). 100 GC- and MS-based methods can speciate organic nitrates and have been employed 101 previously to quantify IN in both laboratory and field studies (Lockwood et al., 2010; 102 Patchen et al., 2007; Giacopelli et al., 2005; Paulot et al., 2009; Lee et al., 2014a; 103 Grossenbacher et al., 2004; Beaver et al., 2012; Kwan et al., 2012). For MS-based 104 techniques, the fragile O-NO<sub>2</sub> bond in organic nitrates often fragments during ionization 105 (Perring et al., 2009), so soft-ionization methods with reagent ion such as  $H^+(H_2O)_4$ , 106  $CF_3O^-$ , and I<sup>-</sup> are necessary to detect the molecular ion for organic nitrates (Patchen et al., 107 2007; Beaver et al., 2012; Lee et al., 2014a; Crounse et al., 2006).

108 Here we present a comprehensive laboratory and field study of the formation of IN from

109 isoprene reaction with OH. In the summer of 2013, we quantified ambient IN in rural

110 Alabama for 6 weeks during the Southern Oxidant and Aerosol Studies (SOAS,

111 http://soas2013.rutgers.edu/). In parallel with the field study, laboratory experiments were

112 conducted to determine the yield of IN from isoprene oxidation. For laboratory

113 experiments, we synthesized authentic standards for the quantification of IN, using

114 multiple calibration techniques. The IN yield obtained from lab experiments was applied

115 in a zero-dimensional model to simulate IN production and loss in the atmosphere, which

116 was then compared with the measurements from SOAS, to examine our understanding of

117 atmospheric IN chemistry.

118 2 Experiment

# 119 2.1 CIMS IN calibration

120 A chemical ionization mass spectrometer (CIMS) was used to measure IN concentrations

121 during the chamber experiments and the SOAS field study. The instrument is similar to

122 the one described by Liao et al. (2011), which uses  $I(H_2O)_n$  to form iodide clusters with

123 the analyte compounds.

Two authentic standards, 4,3-IN and 1,4-IN (a mixture of *trans-* and *cis*-1,4-IN), were synthesized to determine the sensitivity of CIMS toward IN isomers. 1,4-IN was prepared

using the nitrification method described by Lee et al. (2014b), and the sample was used

127 after flash column chromatography without further purification to separate the *trans* and

128 *cis* isomers. 4,3-IN was prepared by nitrification of (1-methylethenyl)oxirane, and the

129 epoxide was synthesized following Harwood et al. (1990).

130 The IN gas-phase sample for CIMS calibration was prepared by evaporating an 131 IN/CH2Cl4\_-standard solution of known volume into 50 L of clean air. The IN 132 concentration in the standard solution was determined with NMR and FTIR, and the 133 results from the two methods were consistent within 15%. Multiple CIMS calibrations for 134 4,3-IN were performed, and the results did not deviate more than 15% after 1.5 years 135 (Supplement Sect. 1). The average sensitivity of 4,3-IN normalized to the reagent ion 136 signal was  $2.3(\pm 0.3) \times 10^{-3}$  ppt<sup>-1</sup> and was used as the surrogate sensitivity for the  $\beta$ -IN 137 isomers except 1,2-IN.

The 1,4-IN calibration was conducted following the same procedures. Since the 1,4-IN standard contained a mixture of *trans-* and *cis-*1,4-IN, the measured sensitivity was a weighted average of both isomers. The relative abundance of the *trans-* and *cis-*1,4-IN

141 isomers was obtained from NMR, and their individual sensitivities were estimated using

142 a least-squares method (Supplement Sect. 2). The sensitivity for trans 1,4 IN was

143  $3(\pm 2) \times 10^{4}$  ppt<sup>-1</sup>, and it was used as surrogate sensitivity for *trans* 4,1 IN. The sensitivity

144 for cis-1,4-IN was  $1.3(\pm 0.3) \times 10^{-3}$  ppt<sup>-1</sup>, and it was used as surrogate sensitivity for cis-4,1-

145 IN. The CIMS sensitivity was  $3(\pm 2) \times 10^{-4}$  ppt<sup>-1</sup> for *trans*-1,4-IN and  $1.3(\pm 0.3) \times 10^{-3}$  ppt<sup>-1</sup>

146 <u>for cis-4,1-IN.</u>

147 As we were unable to synthesize the 1,2-IN standard in the condensed phase, a relative 148 method was used, where the CIMS was interfaced with a GC equipped with an electron 149 capture detector (ECD, Figure 1) to determine the CIMS sensitivity of 1,2-IN relative to 150 4,3-IN. A mixture of the eight IN isomers was generated by irradiation of a mixture of 151 isoprene, isopropyl nitrite, and NO. The IN mixture was cryo-focused at the head of a 5-4152 m Rtx-1701 column that separated the IN isomers, and the effluent was split into two 153 fused-silica deactivated transfer columns, directed simultaneously to the CIMS and the 154 ECD.

155 As the CIMS was operated with water addition to the sample gas before ionization, the 156 GC-ECD/CIMS setup enabled direct observation of the influence of water vapor to the 157 sensitivity of the two dominant IN isomers. Figure 2 shows the GC-ECD/CIMS 158 chromatograms with and without water added to the CIMS. The change in retention time 159 was the result of change in initial oven temperature setting, which had little influence on 160 the elution temperature of IN. 1,2-IN and 4,3-IN were the dominant IN isomers and 1,2-161 IN eluted before 4,3-IN, according to a recent study using the same stationary phase 162 (Nguyen et al., 2014b). 1,2-IN and 4,3-IN are expected to have the same ECD sensitivity, 163 because the ECD has similar response to all mono-nitrates and the hydroxyl group in 164 hydroxynitrate has no influence on ECD sensitivity (Hao et al., 1994). Therefore, the 165 CIMS sensitivity of 1,2-IN relative to 4,3-IN was calculated as the ratio of the CIMS 166 signal intensity to the corresponding ECD signal intensity, for the pair of isomers. The 167 calculated relative CIMS sensitivity was  $0.37(\pm 0.06)$  with water and  $0.95(\pm 0.06)$  without 168 water added, determined as the average of three trials for each setup. The result indicated 169 that water addition to the sample air lowered the CIMS sensitivity of to the 1,2-IN isomer. 170 Hence, 1,2-IN has to be considered separately from the other B-IN isomers when 171 calculating the IN calibration factors. The small abundance of the other isomers makes it 172 difficult to obtain reliable quantification through this method. Therefore, the sensitivities 173 for cis- and trans-1,4-IN were obtained with a synthesized standard.

174	The CIMS sensitivities toward alkyl alcohols and alkyl nitrates are both around 5 orders
175	of magnitude smaller than its sensitivity toward the isoprene hydroxynitrates. Hence, it is
176	the combination of the OH group and the NO3 group, as well as their relative positions
177	that has the dominant influence on the CIMS sensitivity, which will affect how the
178	molecule binds with the iodide ion, while the structure of the carbon backbone would
179	have little effect. For the IN isomers, the relative positions of the OH group and the
180	nitrate group are $\alpha,\beta$ position, trans- $\alpha,\delta$ position and cis- $\alpha,\delta$ position. We assume the
181	same sensitivity can be applied to isomers within each structural group, namely β-
182	isomers, trans-b isomers and cis-b isomers. This assumption is consistent with our
183	observation of identical sensitivity for 1,2-IN and 4,3-IN isomers when water is not
184	added to the CIMS. For the case with water addition to CIMS, the smaller sensitivity of
185	the 1,2-IN was caused by the smaller amount of 1,2-IN available for detection, as 1,2-IN
186	is lost inside the instrument, rather than from a fundamental difference in the ionization
187	efficiency of 1,2-IN. Primary nitrates (δ-IN, 3,4-IN, and 2,1-IN) and secondary nitrates
188	(4,3-IN) are not as likely to be affected by water (Hu et al., 2011). As a result, eis-1,4-IN
189	was used as a surrogate for tris-4,1-IN, and trans-1,4-IN was used a surrogate for trans-
190	<u>4,1-IN. For the <math>\beta</math>-IN isomers, 1,2-IN had to be considered separately due to its loss inside</u>
191	the instrument, but 4,3-IN was used as a surrogate for 3,4- and 2,1-IN isomers. Our
192	assignment of CIMS sensitivitiesy for IN isomers is consistent with reports from Lee et al.
193	(2014a) <u>.</u>
194	Given the significant difference in sensitivity for different IN isomers, the CIMS IN data
195	have to be interpreted with the knowledge of relative IN isomer distribution, which
196	depends on both IN production and IN-loss. Since the IN isomer distribution was not
197	measured in either the laboratory or the field studies, model simulation was used to
198	estimate the relative abundance of IN isomers. The distribution of IN isomers during the
199	chamber experiments was estimated using an iterative method (Supplement Sect. 3.1).

For IN measurement during SOAS, a diurnal average of the changing IN isomerdistribution (Figure S9) was estimated and applied to calibrate IN data for each individual

202 day. The isomer-weighted IN sensitivity changed by less than 20% throughout the day

203 (Supplement Sect. 3.2).

Formatted: Font color: Text 1

Formatted: Font: Italic, Font color: Text 1 Formatted: Font color: Text 1 Formatted: Font: Italic, Font color: Text 1 Formatted: Font color: Text 1

-	Formatted: Font: Italic
-	Formatted: Font: Italic
1	Formatted: Font: Italic
Y	Formatted: Font: Italic

### 204 **2.2** Isoprene chamber experiments

Seven experiments were conducted in the 5500 L Purdue photochemical reaction chamber (Chen et al., 1998) to determine the yield of IN from OH-initiated isoprene oxidation in the presence of  $NO_x$ . OH was generated from the photolysis of isopropyl nitrite. The starting conditions for the experiments are listed in Table 2. Each experiment was initiated by switching on the UV lamps and was considered complete when half of the isoprene was consumed or the NO concentration dropped to around 5 ppb.

211 The IN concentration was measured continuously during each experiment with the CIMS. 212 Chamber air was sampled through a 5.2 m long inlet, made of 0.8 cm ID heated (constant 213 50 °C) FEP Teflon tubing. A total flow of 5 liters per minute (lpm) was pulled through 214 the inlet into a custom-built three-way valve system (Liao et al., 2011), where 2 lpm was 215 sub-sampled into the CIMS through a 0.51 mm orifice. Water vapor was added 216 downstream of the orifice to humidify the sample air to reduce the influence that 217 variations in ambient RH and temperature have on the distribution of I(H<sub>2</sub>O)<sub>n</sub> clusters. 218 Laboratory tests showed that with constant H<sub>2</sub>O addition, the CIMS sensitivity is not 219 dependent on ambient air humidity (Supplement Sect. 6). The fractional loss inside the 50 220 °C sampling inlet was measured to be 5 % for a mixture of the eight IN isomers. The IN 221 loss in the 50 °C sampling inlet was measured to be 5% for the 4,3 IN isomer. However, 222 for the 1.2-IN isomer, it is more susceptible loss under high temperature. This is reflected 223 in the chromatograms in Figure 2a and 2b, as the ECD peaks of 1,2-IN and 4,3-IN have 224 approximately the same height, while OH would preferentially add to C1 than C4 225 (Peeters et al., 2007), leading to a theoretical relative yield of 1.8 2 for 1.2 IN to 4.3 IN. 226 Therefore, it is likely 1,2 IN decomposed in the GC oven with an elution temperature of 227 80 °C. The loss of 1,2 IN is unclear as it flows through the 50 °C inlet with a residence 228 time of 3s, so an upper limit of 50% was estimated, based on the loss of 1,2-IN inside the 229 GC system.

Isoprene and its oxidation products, methyl vinyl ketone (MVK) and methacrolein (MACR), were quantified with a proton-transfer reaction linear ion trap mass spectrometer (PTR-LIT MS), with measurement precision of 3 ppb and accuracy of  $\pm 17\%$ (Mielke et al., 2010). MVK and MACR were observed as the same nominal mass without further differentiation for relative isomeric abundance. The NO concentration was measured through chemiluminescence using the total reactive nitrogen instrument (TRENI) (Lockwood et al., 2010), and the addition of isopropyl nitrite did not cause any interference signal for TRENI during the chamber experiments.

One wall loss experiment was conducted by keeping the IN isomers produced from
isoprene oxidation in the dark chamber and sampling the chamber air with CIMS
periodically for four hours. No significant IN loss was observed, so no wall loss
correction was not corrected applied for IN measurement.

## 242 2.3 CIMS SOAS measurement

243 During SOAS, the CIMS was used to measure ambient IN concentrations continuously 244 from May 26st to July 11th, 2013 at the Centerville (CTR) site (32.90° N, 87.25° W). The 245 CTR site is located about 50 miles south of Birmingham and Tuscaloosa near the 246 Talladega National Forest, a region abundant with pine and oak trees. The CIMS was 247 operated under the same conditions as those during the chamber experiments, in terms of 248 voltage setting, gas flow and sample humidification. Air was sampled from 5.3 m above 249 the ground, with the same inlet (heated to constant 50 °C) and valve system that were 250 used for chamber experiments. The CIMS three-way valve system was used to allow 251 automated background measurement and in-situ Br<sub>2</sub> calibration to monitor instrument 252 stability. The background was determined by passing ambient air through nylon wool 253 coated with sodium bicarbonate for 2 minutes every 15 minutes (Crounse et al., 2006). 254 Laboratory tests suggested that the scrubber could-removes isoprene-derived organic 255 nitrates, including hydroxynitrates, carbonyl nitrates and hydroxyperoxy nitrates, and 256 acids such as nitric acid and formic acid. Br2 calibration was performed hourly by adding 257 a 30 sccm Br<sub>2</sub>/N<sub>2</sub> flow from a Br<sub>2</sub> permeation device to the ambient air being sampled 258 into the CIMS for 2 min. The CIMS sensitivity to IN was calibrated relative to the Br<sub>2</sub> sensitivity, which were both normalized to the reagent ion signal I(H2<sup>18</sup>O)<sup>-</sup>. The Br2 259 260 output rate from the permeation device was determined daily with the optical absorption 261 method following Liao et al. (2011). The averaged Br2 output of the permeation source 262 throughout the campaign was  $60(\pm 8)$  ng/min, which was  $1.8(\pm 0.2)$  ppb when diluted with 263 ambient air.

### 264 2.4 0D Model for IN Data Interpretation

265 A zero-dimensional (0D) model based on the Master Chemical Mechanism (MCMv3.2) 266 (Jenkin et al., 1997; Saunders et al., 2003) was used to investigate the production and loss 267 of IN in the chamber and in the SOAS field study. The mechanism was updated for 268 recent experimental and theoretical studies of isoprene chemistry, including the 269 interconversion of isomeric isoprene RO<sub>2</sub> radicals (LIM1) (Peeters et al., 2014), IN 270 reaction rate constants for OH and  $O_3$  (Lee et al., 2014b), IEPOX reaction rate constants 271 for OH (Bates et al., 2014) and the branching ratio for NO<sub>3</sub> addition to isoprene (Fan and 272 Zhang, 2004).

273 For the IN observations during SOAS, our analysis is focused on the production and loss 274 of IN. Therefore, the 0D model for the SOAS data analysis was constrained to the 275 observed concentrations of the major species involved in the IN chemistry, including 276 isoprene, HO<sub>x</sub>, O<sub>3</sub>, NO<sub>x</sub>, α-pinene, β-pinene and limonene. The NO<sub>2</sub> photolysis frequency 277 in the 0D model (J<sub>NO2</sub>) was calculated using the Tropospheric Ultraviolet & Visible 278 (TUV) Radiation Model (Madronich and Flocke, 1998) for clear sky conditions with 300 279 DU ozone, and the model input was scaled relative to observed radiation. The photolysis frequencies for all the other species were scaled relative to J<sub>NO2</sub> at zero-degree solar 280 zenith angle. 281

282 Because the 0D model does not take into account the changes in IN concentration as IN 283 was transported to and out of the measurement site both vertically and horizontally, the 284 ratio of total IN concentration to the sum of methyl vinyl ketone (MVK) and 285 methacrolein (MACR) was used to compare the model results with observations. Major 286 sources of MVK and MACR include isoprene ozonolysis (Grosjean et al., 1993) and OH-287 initiated isoprene oxidation (Liu et al., 2013). As IN, MVK and MACR are produced 288 simultaneously in the isoprene photochemical oxidation process, the ratio 289 [IN]/([MVK]+[MACR]) may cancel-reduceout the influence of dilution caused by 290 vertical mixing and changing boundary layer height, making results from the 0D model 291 comparable to ambient observations. Besides chemical loss to reaction with OH,  $O_3$  and 292 NO<sub>3</sub>, the model also included loss for dry deposition for IN, MVK and MACR, with

averaged daytime deposition velocities of 1.5cm/s, 0.7cm/s and 0.4cm/s (Nguyen et al.,
2015; Zhang et al., 2002).

295 Isoprene data from the PTR-ToF-MS measurement\_(Misztal et al., In preparation) were 296 used to constrain the model and its MVK+MACR data were used for model-observation 297 comparison for most days. The MVK and MACR data from the GC-MS measurement 298 (Gilman et al., 2010) were used when knowledge of the relative abundance of MVK and 299 MACR was required to calculate their initial concentrations in the model and when PTR-300 ToF-MS data were unavailable. The PTR-ToF-MS data were used primarily because of 301 its higher time resolution. Model constraints on  $\alpha$ -pinene,  $\beta$ -pinene and limonene 302 concentrations were based on measurements from GC-MS, and 2D-GC when GC-MS 303 data were unavailable.

# 304 **3 Results**

### **305 3.1 IN yield from chamber experiments**

306 The IN yield was calculated from the production of IN relative to the loss of isoprene, 307 using data obtained in the photochemical reaction chamber experiments. The isomer-308 weighted IN sensitivity is expected to change during each experiment, as IN isomers are 309 lost to OH consumption with different reaction rate constants. - An To account for the 310 change in IN isomer distribution during each experiment, an iterative method was applied 311 to derive a self-consistent set of total IN yield, IN isomeric distribution and isomer-312 weighted IN sensitivity (Supplement 3.1). IN loss by OH oxidation was corrected 313 (Atkinson et al., 1982) with an isomer-weighted rate constant to account for the 314 difference in OH reactivity for different isomers (Lee et al., 2014b). The correction factor 315 was around 25% by the end of each experiment. Figure 3 shows the results from the IN 316 yield chamber experiments. The average IN yield was 9%, based on the slope of  $\Delta$ IN vs. 317 (-\Delta isoprene). We note that the yield has no apparent [NO] dependence with [NO] varying in the range from 125 ppb to 2400 ppb. 318

- 319 The relative measurement uncertainties were 15% for IN and 17% for isoprene, based on
- 320 the instrument calibrations. Uncertainty in the reported rate constants for IN oxidation
- 321 could cause 20% error when IN data was corrected for OH consumption. The uncertainty

323 based on sensitivity tests (Supplement Sect. 5). Therefore, the overall relative uncertainty 324 based on error propagation for our experimentally derived IN yield is 36%. When 325 uncertainties in the inlet loss for the 1,2-IN isomer was taken into consideration, a 326 maximum total yield of 12.6% was derived, based on the assumed limit of 50% 1,2-IN 327 loss. Therefore, we report our total IN yield to be 9(+4/-3)%, with the asymmetric 328 uncertainty to account for possible underestimated losses. 329 The relative uncertainty for isoprene concentrations is 17% based on instrument 330 calibration. The uncertainty for IN concentrations is derived caused by both the 331 uncertainty in the CIMS sensitivity for each IN isomer and the uncertainty in the relative 332 abundance of the IN isomers. Through a sensitivity test on the RO<sub>2</sub> interconversion rate 333 constants of the LIM1 mechanism (Supplement Sect. 5), the IN measurement uncertainty 334 was estimated to be +23%/-20%. The fractional loss for the CIMS inlet was  $4(\pm 6)\%$ , 335 making the IN measurement uncertainty to be +24%/-20%. The uncertainty in the 336 reported rate constants for IN oxidation could cause 20 % error when IN data was 337 corrected for OH consumption. Therefore, the overall relative uncertainty in our IN yield 338 is +36%/-33% and we report our total IN yield to be 9(+4/-3)% to encompass all the 339 measurement uncertainties. This result lies in the middle of the 4-14% range of IN yields 340 determined from previous experiments (Chen et al., 1998; Patchen et al., 2007; 341 Lockwood et al., 2010; Paulot et al., 2009; Sprengnether et al., 2002; Tuazon and 342 Atkinson, 1990), but is more consistent with determinations from the higher end of the 343 range. Previous IN studies conducted in our group using GC methods consistently 344 resulted in lower IN yields (Chen et al., 1998; Lockwood et al., 2010). We partially 345 attribute the discrepancy of our previous and current work to the possible loss of the 1,2-

in the IN isomer distribution derived from the LIM1 mechanism was 20%, estimated

346 IN isomer in the GC column and metal sample injection system. This work employed MS 347 to quantify IN during the chamber experiments to circumvent these problems. The 348 current yield result will be applied in the 0D model to simulate IN concentrations during 349 SOAS. The model-measurement agreement offers a metric to evaluate the validity of the

350 laboratory-derived IN yield

### 351 3.2 Observation of IN during SOAS

Figure 4 shows the temporal profile of total IN mixing ratio observed during the SOAS field study with an averaging 10-minute time resolution. In general, fast IN production was observed after sunrise. On average, the concentration rose to peak around 70 ppt at 10:00 AM (Figure 5) and then decreased to a minimum around 10 ppt by 6:00 AM the next day, as a result of vertical mixing, boundary layer expansion, dry deposition and further oxidation. IN concentrations were significantly lower from Jul 4 to Jul 8, due to wet deposition and less photochemical reactivity caused by continuous rain events.

359 In contrast to the IN average diurnal profile (Figure 5), the diurnal profiles for isoprene,

 $\beta$ 60 OH, and NO<sub>x</sub> and MVK+MACR, each peaked at different times of the day (Figure 6).

While IN and MVK + MACR are products of the parallel RO<sub>2</sub> + NO Reactions (R2a and

b), the diurnal MVK + MACR concentrations are more consistent with the temporal

profiles of isoprene, OH and O<sub>3</sub> with peak concentration around 1:00 PM when radiation

364 was strong. The decrease in IN, and continued increase of MVK and MACR around

10:00 AM can be attributed to the competition among the four RO<sub>2</sub> loss channels (R2,
R3, R4, and R5).

- $367 \quad \text{RO}_2 + \text{HO}_2 \rightarrow \text{ROOH} + \text{O}_2 \tag{R3}$
- $368 \quad \mathrm{RO}_2 + \mathrm{RO}_2 \rightarrow 2\mathrm{RO} + \mathrm{O}_2 \tag{R4}$
- $369 \quad \text{RO}_2 \rightarrow \text{ products}$  (R5)

370 The fraction of RO<sub>2</sub> loss to NO reaction is defined as  $\gamma$ , which is calculated with the 371 following equation.

372 
$$\gamma = \frac{k_{RO2+NO}[NO]}{k_{RO2+NO}[NO] + k_{RO2+HO2}[HO_2] + k_{RO2+RO2}[RO_2] + k_{isomerize}}$$
(1)

373 Isoprene RO<sub>2</sub> loss to permutation reactions R4 was calculated assuming [RO<sub>2</sub>]=[HO<sub>2</sub>], 374 and the rate constant  $1.6 \times 10^{-13}$  cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> was used (Jenkin et al., 1997). Isoprene 375 RO<sub>2</sub> loss rates for reaction with NO and HO<sub>2</sub> (R2 and R3) were calculated based on 376 observed NO and HO<sub>2</sub> concentrations, using rate constants  $k_{RO2+NO} = 9 \times 10^{-12}$  cm<sup>3</sup> 377 molecule<sup>-1</sup> s<sup>-1</sup> and  $k_{RO2+HO2} = 1.61 \times 10^{-11}$  cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> (Saunders et al., 2003; Stevens 378 et al., 1999). The sum of the first-order RO<sub>2</sub> loss rate for reaction with NO, HO<sub>2</sub> and RO<sub>2</sub> 379 was 0.01-0.07 s<sup>-1</sup> (Figure 7a). Therefore, contribution from 1,5-H shift for  $\beta$ -RO<sub>2</sub> was negligible, due to the small isomerization rate constant (Peeters et al., 2014). However, 380 381 for isoprene  $cis-\delta$ -RO<sub>2</sub>, the 1,6-H shift rate constant is on the order of 0.1-1s<sup>-1</sup> (Peeters et 382 al., 2009; Crounse et al., 2011; Peeters et al., 2014). This fast isomerization depletes cis-383 δ-RO<sub>2</sub> radicals rapidly to form closed-shell products, e.g. hydroperoxy aldehyde (HPALD), and makes the relative abundance of cis-\delta-RO2 radicals very small (~1%, 384 385 Supplement Sect. 4). For this reason,  $k_{isomerize}$  was omitted from the calculation of  $\gamma$ , but 386 the yield of total RO2 was incorporated when estimating the production rate of total IN, to 387 account for rapid loss of cis-\delta-RO<sub>2</sub> to 1,6-H shift. In addition, the fast 1,6-H 388 isomerization for  $cis-\delta$ -RO<sub>2</sub> decreased the production rates of  $\delta$ -IN among the IN 389 isomers. With this smaller production rates and their faster loss rates to OH and O3 (Lee 390 et al., 2014b), the relative abundance of  $\delta$ -IN during this field study was much smaller 391 than what have been observed in laboratory studies (Lockwood et al., 2010; Paulot et al., 392 2009).

393 The calculated diurnal average of the  $\gamma$  value is shown in Figure 7b. For RO<sub>2</sub> radicals that 394 were lost to reaction with NO or HO<sub>2</sub>, the RO<sub>2</sub>+NO reaction was the sole contributor in 395 the early morning, but it was surpassed by RO<sub>2</sub>+HO<sub>2</sub> reaction before noon, as the NO 396 concentration decreased steadily throughout the day. The IN production rate was 397 calculated with the following equation.

398  $P_{IN} = k_{ISOP+OH}[OH][ISOP] \bullet \Phi \bullet \gamma \bullet \alpha$ 

(2)

399  $\alpha$  is the branching ratio  $(=k_{2a}/(k_{2a} + k_{2b}))$  for<del>of</del> isoprene RO<sub>2</sub> + NO reaction for nitrate 400 formation.  $\Phi$  is the yield of total RO<sub>2</sub> from OH addition to isoprene that are available to 401 react with NO, HO<sub>2</sub> and RO<sub>2</sub>, with an RO<sub>2</sub> lifetime in the range of 10~20s. The 402 calculated  $\Phi$  is 0.83 (Supplement Sect. 4), with contribution from  $\beta$ -RO<sub>2</sub> being 0.81, *cis*-403 δ-RO<sub>2</sub> being 0.01 and trans-δ-RO<sub>2</sub> being 0.02, and the remaining 17% products from 404 isoprene OH oxidation are closed-shell species such as HPALD. The IN production rate 405 decreased with the declining y value (Figure 7b), even though the concentrations of OH 406 and isoprene each peaked at ~1:00 PM.

- 407  $\gamma$  value peaked around 6:00 AM to 8:00 AM when the isoprene and OH concentrations
- 408 were relatively low. During this period, the IN production rate was limited by the

Formatted: Font: (Default) Times New Roman

409	availability of RO2. In the afternoon, when isoprene RO2 was more abundant with higher
410	isoprene and OH concentrations, the IN production rate was limited by the availability of
411	NO, and decreased with the declining $\gamma$ value (Fig. 7b) This could The declining $\gamma$ value
412	could lead to IN loss from OH oxidation to exceed IN production, making IN peak
413	around the time when HO_2 reaction became the major RO_2 loss channel ( $\gamma\!\!<\!\!0.5$ ). In this
414	relatively clean environment, MVK and MACR production continued through isoprene
415	ozonolysis (Grosjean et al., 1993) and OH oxidation in the low NO regime (Liu et al.,
416	2013). The MVK+MACR production rate was calculated using the following equation.

- =  $k_{ISOP+OH}[OH][ISOP] \bullet \Phi_{\beta} \bullet \gamma \bullet (1-\alpha)$ 417  $P_{MVK+MACR} \\$
- 418 +  $k_{ISOP+OH}[OH][ISOP] \bullet \Phi \bullet (1-\gamma) \bullet 0.06$
- 419  $+ k_{ISOP+O3}[O_3][ISOP] \bullet 0.61$

(3)

420  $\Phi_{\beta}$  denotes the yield of isoprene  $\beta$ -RO<sub>2</sub>, the precursors for MVK+MACR, and the 421 calculated  $\Phi_{\beta}$  was 0.81 (Supplement Sect. 4). The term k<sub>ISOP+OH</sub>[OH][ISOP]• $\Phi_{\beta}$ • $\gamma$ •(1- $\alpha$ ) is 422 the production rate of MVK+MACR with the isoprene  $\beta$ -RO<sub>2</sub> undergoing the RO<sub>2</sub>+NO 423 reaction pathway. The term  $k_{ISOP+OH}[OH][ISOP] \bullet \Phi \bullet (1-\gamma) \bullet 0.06$  is the production rate of 424 MVK+MACR when the isoprene RO2 proceeds via HO2-+and-RO2 reaction pathways to 425 form MVK+MACR with an overall yield of 6% (Liu et al., 2013). The term 426 kISOP+O3[O3][ISOP]•0.61 is the production rate of MVK+MACR from isoprene 427 ozonolysis, with a total yield of 61% (Grosjean et al., 1993).

428 As shown in Figure 7b, the production rates of IN and MVK+MACR both plateaued 429 around 10:00 AM. For MVK+MACR, the decrease was later compensated with 430 production from the HO2 RHO2 and O3 pathway, and the production rate peaked around 431 2:00 PM when radiation was strong. For IN, however, its production rate did not increase 432 with radiation due the limited availability of NO (small y value). Therefore, the change in 433 the relative importance of the two RO<sub>2</sub> sinks, RO<sub>2</sub>+NO and RO<sub>2</sub>+HO<sub>2</sub>, is likely one of the 434 reasons that the IN concentration peaked earlier than the MVK+MACR concentration 435

- during SOAS.
- The early morning increase in IN concentration could imply significant contribution from 436
- 437 downward mixing of accumulated IN in the residual layer (RL), as the inversion is

broken up after dawn (Hastie et al., 1993). When the earth surface cools in the evening, 438 439 the remnants of the upper daytime boundary layer are isolated from the lower region near 440 the ground, and the RL forms. The RL contains the same amount of isoprene, IN, and 441 NO<sub>x</sub> as is found near the ground around sunset, thus serving as an IN reservoir at night (Neu et al., 1994). While IN in the nighttime boundary layer (NBL) is slowly lost to dry 442 deposition, IN in the RL, which is isolated from the ground, is better preserved. In 443 444 addition, IN production from reaction of isoprene with NO3 may also contribute to RL 445 IN, but this process is not as important in the NBL, <u>because the NO<sub>3</sub> + NO reaction near</u> 446 the ground makes the NO<sub>3</sub> concentrations smaller because NO<sub>3</sub> is likely to be titrated by 447 reaction with NO near ground (Stutz, 2004). As a result, the IN concentration in the RL is 448 expected to be higher than that in the NBL before dawn. Perhaps more importantly, the 449 relative volume fraction of NBL vs. RL is small (~0.1), and thus surface level nighttime 450 chemistry cannot contribute significantly to the surface IN increase at ~10am. Figure 7 451 shows that the surface (NBL) IN production rate is very low at night, though non-452 negligible.-During sunrise, IN in the RL can mix downward, which in combination with 453 photochemical IN production leads to an increase in IN near the ground. The relative 454 importance of these two processes will be assessed with our 0D model in the following 455 section.

456 It is worth mentioning that the nighttime ground-level IN production from  $NO_3 + \bullet$ 

457 isoprene can afford a different IN isomer distribution, which can influence the isomer-

458 weighted IN sensitivity. However, the 0D model simulation of IN isomer distribution has

459 <u>included IN formation from the NO<sub>3</sub> + isoprene pathway. Therefore, our interpretation of</u>

- 460 the SOAS IN measurement data has considered the changing IN isomer distribution from
- 461 <u>both the OH- and the NO<sub>3</sub>-initiated IN production near ground.</u>
- 462 <u>alterCIMS (2004)</u>

# 463 **3.3 0D model for IN photochemistry during SOAS**

464 Due to limited availability of overlapping data for model input from multiple instruments,

- 465 ambient data for the following 12 days were used: Jun 14, Jun 16, Jun 22-Jun 23, Jun 25-
- 466 Jul 1 and Jul 3. For each day, only the daytime chemistry (5:00 AM 7:00 PM) was

Formatted: Normal
Formatted: Subscript

simulated, when photochemical reactivity was high. The observed IN and MVK+MACR 467 468 concentrations at 5:00 AM were used as the initial concentrations for 469 modelingsimulations. For isoprene, α-pinene, β-pinene, limonene, NO, NO<sub>2</sub>, OH, HO<sub>2</sub>, 470 and O<sub>3</sub>, the model concentrations were constrained to observations for the entire 471 modeling period. The branching ratio for IN formation resulting from the isoprene RO<sub>2</sub> 472 with NO reaction was set to 0.09 for all isomers, which is based on our measured 9% 473 yield from the chamber experiments. As mentioned above, to avoid the complication in 474 the simulation of the absolute concentration variability froms caused by transport and 475 changing boundary layer height, we compared the simulated and observed [IN]/([MVK]+[MACR]) ratio to evaluate the model. 476

477 Figure 8a shows the temporal profiles of the modeled and observed 478 [IN]/([MVK]+[MACR]) ratio for the selected 12 days. To gain a statistical overview of 479 the model and observation comparison, the 12-day average was calculated (Figure 8b). 480 The 0D model generally agrees with the observed ratio, lending support to the IN branching ratio determined in the chamber experiments. However, the morning increase 481 482 was underestimated by the model on certain days (Jun 14, Jun 16, Jun 29, Jul 1 and Jul 3), 483 while on other days (Jun 23 and Jun 25-Jun 27), the decrease rate for the 484 [IN]/([MVK]+[MACR]) ratio was underestimated in the afternoon. Since the IN yield 485 applied in the 0D model has  $\pm 36\%/-33\%$  uncertainty, a sensitivity test on the yield 486 was performed. As shown in Figure 8c, the simulated [IN]/([MVK]+[MACR]) ratio is 487 highly sensitive to the yield used in the model-as expected. The 6% yield significantly underestimated the ratio in the morning, and the 12% yield significantly overestimated 488 489 the ratio in the afternoon.

490 4 Discussion

### 491 4.1 Model-observation comparison for SOAS data

As shown in Figure 8c, the modeled results deviated from observations from 10:00 AM
to 12:00 PM for all the three yields applied, During this period, as the simulated
[IN]/([MVK]+[MACR]) ratio decreased slowly, but while the observed ratio dropped
rapidly. The fast decrease in the [IN]/([MVK]+[MACR]) ratio implies either fast

production of MVK+MACR, or fast consumption of IN. In terms of fast production of 496 497 MVK+MACR, the formation of MVK+MACR from OH and O<sub>3</sub> have been characterized 498 in the model, and the model was capable of simulating MVK+MACR concentration to 499 within measurement uncertainty for the chamber experiments (Supplement 3.1). 500 Therefore, the discrepancy between model and observation is more likelypotentially 501 associated with underestimated loss rate of IN. The model results with the 6% yield were 502 lower than observations, despite potential underestimated IN loss rate, so a higher yield 503 (9-12%) may be more accurate to describe the branching ratio for isoprene RO<sub>2</sub>+NO 504 reaction.

505 The model overestimation in the afternoon can be caused collectively by measurement 506 uncertainties for model input, uncertainties in the IN loss rates for OH oxidation and 507 deposition, uncertainties in ambient IN (25%) and MVK+MACR (40%) measurement, 508 and other missing IN loss processes. A recent study found that isoprene hydroperoxide 509 (ISOPOOH) could interfere with MVK and MACR measurement when standard PTR-510 MS and GC methods are used (Rivera-Rios et al., 2014). We found that the model 511 appeared to agree better with observations in the afternoon, if the ISOPOOH+IEPOX 512 concentration was subtracted from the MVK+MACR measurement data (Figure 8d). 513 However, the exact influence of ISOPOOH+IEPOX on the observations of 514 MVK+MACR is unclear, as the ISOPOOH conversion efficiency is highly dependent on 515 instrumental sampling configuration, and the interferece of IEPOX has not been 516 characterized.

517 We also considered that an underestimated IN photolysis rate could be one of the reasons 518 for the model-observation discrepancy. The photolysis rate for IN was set to be identical 519 as to the photolysis rate for alkyl nitrates in MCMv3.2, but IN isomers have double bonds 520 and hydroxyl groups, which could increase the IN absorption cross section and enhance 521 the photolytic reactivity for IN. When the IN photolysis rate was increased by 5 times for 522 the 9% yield, or 12.5 times for the 12% yield, the simulated [IN]/([MVK]+[MACR]) 523 ratio was brought closer to observation in the afternoon, but the IN loss rate still appeared 524 underestimated between 10:00 AM and 12:00 PM (Figure 8e). When the higher 525 photolysis rates were applied, the simulated IN loss to photolysis could contribute up to 526 30% (9% yield case) or 50% (12% yield case) of total IN loss. Since no experimental

527 data were available on the absorption cross spectrum and quantum yield for IN, large 528 photo-dissociation rate coefficients are purely hypothetical. While photolysis may be a 529 significant IN sink in <u>the</u> ambient environment, its contribution to IN loss in chamber 530 experiments is negligible, as the lamp radiation was  $\simeq 10\%$  of solar radiation and the 531 durations of the chamber experiments were short. Therefore, no correction for the 532 photolytic loss <u>will bewas</u> made for the IN measurement performed in chamber 533 experiments.

534 Despite the discrepancy in absolute values, the simulated [IN]/([MVK]+[MACR]) ratios all peaked in the morning, consistent with observation. The peak signifies the time when 535 536 the IN loss rate started to exceed the IN production rate. As the OH-loss lifetime of IN 537 decreased from 8:00 AM to 1:00 PM, the IN production rate (Figure 7b) remained 538 constant during this time. Although isoprene and OH concentrations were both greater 539 after the noontime, the IN production rate did not increase, due to the small  $\gamma$  value. 540 Therefore, the morning IN peak can be attributed to the combined effects of slow IN 541 production and fast IN consumption in the afternoon, with NO<sub>x</sub> being the limiting factor 542 for IN production during this field study.

Although the simulated [IN]/([MVK]+[MACR]) ratios all peaked in the morning, they 543 544 peaked one hour later than the observed ratio (Figure 8c). In addition, the modeled ratio 545 had a smaller growth rate than the observed ratio between 7:00 AM and 9:00 AM, 546 regardless of the IN yield and IN loss rate applied (Figure 8c and 8d). This 547 underprediction implies an unknown source of IN, and we can hypothesize that it was caused by downward mixing of the RL IN, as the fast morning increase of IN coincided 548 549 with inversion breakup. By combining the observations of IN and MVK+MACR during 550 SOAS and the results from the 0D model, we can calculate the growth rate of ambient IN 551 concentration caused solely by isoprene photochemistry in the daytime (Supplement Sect. 552 7). This photochemical IN growth rate was compared with the observed IN growth rate, 553 and from that we estimate that downward mixing can contribute to  $27(\pm 16)\%$  of the fast 554 IN increase in the morning, where the large uncertainty originates from the 30% 555 uncertainty in the IN yield.

The residual layer IN concentration before mixing (6:00 AM) was estimated with the 0D 556 557 model, using the same initial input as the ground-level observation on the previous day at 558 8:00 PM. The chemical processes involved are IN production from isoprene oxidation by 559 NO3 (R5 and R6b) and IN consumption by OH, O3 and NO3. Based on our model 560 calculation, the steady-state NO<sub>3</sub> concentration at night was on the order of  $1 \times 10^6$ molecules cm<sup>-3</sup>. Nighttime OH was generated through BVOC ozonolysis, and the 561 562 simulated concentration was on the order of  $5 \times 10^4$  molecules cm<sup>-3</sup>. Even though the OH 563 concentration was very low at night, it was still the dominant IN loss pathway, as a 564 resultbecause of the fast IN+OH reaction rate constants. It is worth noting that  $RO_2$ produced from isoprene + NO<sub>3</sub> (R6) also has competing loss channels for reaction with 565  $RO_2$  (R7) and with HO<sub>2</sub> (R8). Therefore, only a fraction of the isoprene nitrooxy peroxy 566 567 radicals (ONO<sub>2</sub>RO<sub>2</sub>) can react with other peroxy radicals to produce IN through reaction 568 R7b.

569	$C_5H_8$ +	$NO_3 (+ O_2) \rightarrow$	$ONO_2RO_2$	(R6)

$$570 \quad ONO_2RO_2 + RO_2 \rightarrow ONO_2RO + RO + O_2$$
(R7a)

571  $ONO_2RO_2 + RO_2 \rightarrow ONO_2ROH + R'CHO + O_2$  (R7b)

(R8)

572  $ONO_2RO_2 + HO_2 \rightarrow ONO_2ROOH$ 

573 Figure 9 shows the simulated IN concentration in the RL and IN observed near ground 574 before dawn, assuming the RL was completely stable at night with no depositional loss 575 for IN from the RL. The simulated IN concentration in the RL before dawn was greater 576 than the concentrations measured at ground level by up to one order of magnitude, 577 indicating the IN stored in the RL overnight may be a significant ground level IN source 578 during inversion breakup. This high IN concentration above the NBL is the result of IN 579 produced during the previous day, which is present with the high concentration in the RL 580 as it is formed, and zero deposition removal overnight. The NO<sub>2</sub> concentration is low when the RL is formed at sunset, so contribution from IN production through NO<sub>3</sub> 581 582 chemistry is small (1-10ppt), a minor fraction compared with the concentration of IN 583 already present in the RL in the evening. Isoprene-NO<sub>3</sub> chemistry can generate IN isomers with a differently isomeric distribution. Since IN production from this reaction 584

585 scheme is small, no sensitivity correction was performed to account for the changes in 586 isomer distribution when RL IN mixed with ground-level IN in the morning.

587 The calculated residual layer IN does not take into account the changingaltitude-588 dependent IN concentration with altitude caused by OH oxidation, as well as possible IN 589 concentration change caused by advection. Therefore, the actual IN concentration may be 590 very different from the calculated results. This is reflected in a comparison of the large 591 RL IN excess relative to surface IN on Jun 26 and 27 (Figure 9), -with simultaneous 592 model overprediction of daytime IN on these two days (Figure 8a). Hence, detailed three-593 dimensional chemical transport models are needed to fully elucidate the production and 594 storage mechanisms of IN in the ambient environment.

# 595 4.2 High-NO<sub>x</sub> and low-NO<sub>x</sub> chemistry during SOAS

596 OH oxidation was the most important daytime sink for BVOCs during SOAS. As the  $\gamma$ 597 value decreased from 0.95 at 7:00 AM to 0.3 at 1:00 PM (Figure 7b), the BVOC-derived 598 RO<sub>2</sub> radicals are expected to undergo both NO (high NO<sub>x</sub>) and HO<sub>2</sub> (low NO<sub>x</sub>) reaction 599 pathways throughout the day. For isoprene, the presence of the two reaction schemes was 500 signified by the oxidation products, with IN peaking in the morning and ISOPOOH and 501 IEPOX peaking in the afternoon (Figure 10).

602 As IN was consumed by OH, it would also undertake both NO and  $HO_2$  reaction 603 pathways. Since the highest OH concentrations (1:00 PM) were accompanied with a

604 small γ value (~0.3, Figure 7b), significant amount of IN would be oxidized following the 605 HO<sub>2</sub> pathway. A possible reaction scheme is illustrated in Figure 11 with 1,2-IN as an

606 example.

607 Experimental studies by Jacobs et al. (2014) suggest that OH addition to IN can invoke

608 IEPOX formation with a yield of 13% at atmospheric pressure, which simultaneously

for releases NO2. Although IEPOX can be produced from IN oxidation, the ISOPOOH

bio pathway was still the dominant IEOPOX precursor during this study, due to the higher

611 concentrations of ISOPOOH and its higher yield for IEPOX (~70-80%) (St. Clair et al.,

612 2015). For RO<sub>2</sub> radicals produced from OH addition to IN, 30% will react with NO and

613 70% will react with HO<sub>2</sub> for a  $\gamma$  value of 0.3 at 1:00 PM.

For the RO<sub>2</sub>+NO reaction, Lee et al. (2014b) observed the formation of dinitrate for δ isomers of IN and estimated a branching ratio of less than 18% for β-4,3-IN based on missing carbon in the gas phase. The RO radicals from the RO<sub>2</sub>+NO reaction will dissociate to make either MACR nitrate or lose NO<sub>2</sub> to form hydroxyacetone and glycoaldehyde. Both Jacobs et al. (2014) and Lee et al. (2014b) reported MACR nitrate being the dominant product with an overall yield of 70%, thus making the corresponding branching ratio for the RO radical to be around 80%.

621 The RO<sub>2</sub>+HO<sub>2</sub> products from IN oxidation are less understood. Alkyl peroxy radical reaction with HO<sub>2</sub> can undergo two reaction channels to afford either hydroperoxide or 622 623 RO radical and OH. The branching ratio is highly structure dependent. Simple alkyl 624 peroxy radicals, such as CH<sub>3</sub>CH<sub>2</sub>O<sub>2</sub>, can form hydroperoxide with almost unity yield 625 (Hasson et al., 2004). However, for peroxy radicals with  $\beta$  carbonyl groups, such as 626  $RC(O)CH_2O_2$ , the branching ratio for the OH formation pathway is more than 60% 627 (Hasson et al., 2012; Hasson et al., 2004). The  $\beta$  carbonyl oxygen can stabilize the 628 reaction intermediate through internal hydrogen bonding, thus making the reaction favor the formation of OH and RO (Hasson et al., 2005). The RO<sub>2</sub> from IN oxidation has a β-629 630 OH group and a β-NO<sub>3</sub> group, both capable of forming internal hydrogen bonding with 631 the hydrogen of HO<sub>2</sub>. Therefore, formation of OH and RO radicals may be a significant 632 reaction channel when the RO<sub>2</sub> radicals derived from IN react with HO<sub>2</sub>. The closed-shell 633 product from the RO<sub>2</sub>+HO<sub>2</sub> reaction is dihydroxy hydroperoxy nitrate (DHHPN). This 634 compound has not been identified in any laboratory studies. However, Lee et al. (2015) 635 found a significant amount of compounds with the corresponding molecular formula of  $C_{5}H_{11}O_{7}N$  in the aerosol phase during SOAS, which suggests that hydroperoxide 636

637 formation and aerosol uptake could be an important sink for IN.

638 A range can be estimated for the  $NO_2$  recycling efficiency for IN oxidation, as the 639 detailed  $RO_2$ +HO<sub>2</sub> reaction mechanism is unclear. If  $RO_2$ +HO<sub>2</sub> reaction forms only

- 640 hydroperoxide, the NO<sub>2</sub> yield from IN oxidation will be 17%. If RO<sub>2</sub>+HO<sub>2</sub> reaction only
- 641 undergoes the radical formation channel, the NO<sub>2</sub> yield will be 30%, and the major
- 642 products of IN oxidation are highly oxidized secondary nitrates.

#### 643 5 Summary and atmospheric implications

644 Our chamber experiments indicate a 9(+4/-3)% nitrate yield from isoprene 645 hydroxyperoxy radical reaction with NO. The product yield provides a more reliable 646 groundwork for future modeling studies on the interplay of isoprene oxidation, NO<sub>x</sub> 647 cycling, and tropospheric O<sub>3</sub> production.

Our field measurements and model simulations suggest that in the southeast US, 648 formation of organic nitrates in the boundary layer is controlled by the availability of 649 650 NO<sub>x</sub>. During the SOAS field study, when isoprene was oxidized by OH addition, the NO 651 peak in the morning on average consumed 95% of the isoprene  $RO_2$  to form high  $NO_x$ photooxidation products such as IN, MVK and MACR. As the NO<sub>x</sub> concentration 652 653 decreased during the day, the  $RO_2$ +HO<sub>2</sub> reactions became more important, and by ~1:00 PM only 30% of the RO2 react with NO, and thus only 2.7% of the RO2 would form 654 655 organic nitrates. The high NOx concentration in the early morning caused an early IN 656 maximum at 10:00 AM, a combined result of slow afternoon IN production with limited NO<sub>x</sub>, and fast IN consumption due to peak radiation and fast OH production in the 657 658 afternoon. By comparing simulation results with observations, we estimate the inversion 659 breakup after sunrise may contribute to  $27(\pm 16)\%$  of the rapid IN increase in the morning. The observed daytime IN loss can be approximated with the current 660 661 understanding of IN oxidation reactions and dry deposition, but some discrepancies still 662 exist, which could be caused by other less studied loss pathways, such as nitrate 663 photolysis. Aerosol uptake could also be an IN sink, but the contribution is expected to be 664 small (Surratt et al., 2010b). Observations during SOAS suggest that the isoprene-derived 665 SOA components were associated with IEPOX and more oxidized organic nitrates, not 666 the first-generation hydroxynitrates (Xu et al., 2015b; Lee et al., 2015). While IN were produced and destroyed in the morning through high NOx chemistry, a 667

major portion of the afternoon IN oxidation process involved low NO<sub>x</sub> chemistry, which could yield products such as the highly oxidized hydroperoxide DHHPN. DHHPN is expected to have very low vapor pressure and undergo fast dry deposition and aerosol partitioning, possibly followed by hydrolysis and formation of NO<sub>3</sub><sup>-</sup> and trihydroxy hydroperoxide. This process removes NO<sub>x</sub> from the atmosphere and helps to shift the Field Code Changed

Formatted: Font color: Text 1

- 673 photochemical processes further toward the low NO<sub>x</sub> regime, forming a positive feedback
- 674 mechanism to reduce the atmospheric NO<sub>x</sub> concentration. However, more experimental
- studies are required to elucidate the detailed mechanism for the RO<sub>2</sub>+HO<sub>2</sub> reactions.
- 676 During the past 15 years,  $NO_x$  emissions in the southeastern US have decreased by more 677 than 50% (Hidy et al., 2014). As more effort is devoted to controlling anthropogenic 678 emissions, the BVOC oxidation processes will start to shift further toward the low  $NO_x$ 679 regime. Isoprene products resulting from oxidation in the low NO<sub>x</sub> condition, such as 680 IEPOX, are more prone to reactive uptake and thus contribute more effectively to the 681 growth of SOA than IN (Xu et al., 2015a; Surratt et al., 2010a; Nguyen et al., 2014a), 682 indicating potentially higher SOA burdens from isoprene chemistry in the future. The low NO<sub>x</sub> photochemistry is often complicated by radical reactions including intramolecular 683 684 H-shift and autoxidation (So et al., 2014; Peeters et al., 2014; Savee et al., 2015; Crounse 685 et al., 2013), so more theoretical and experimental studies of the fundamental reaction kinetics are needed to unravel the complete BVOC oxidation mechanism. The 686 687 photochemical reactions that involve both the high NO<sub>x</sub> and low NO<sub>x</sub> pathways can yield new highly-oxidized multi-functional nitrate products. Identification, quantification and 688 689 study of the chemistry of these organic nitrates is essential to understand the fate of NOx. 690 As the highly-oxidized compounds, such as DHHPN and dinitrate, tend to partition into 691 the aerosol phase, it will be a challenge for the development of analytical techniques to 692 investigate their aging process in the particle phase and their role in the NO<sub>x</sub> cycle.

# 693 Acknowledgement

We thank the organizers of the SOAS study, especially Dr. Ann Marie Carlton. We appreciate help from Dr. Jozef Peeters at University of Leuven in elucidating the uncertainties associated with the current LIM1 mechanism. We acknowledge funding from the National Science Foundation (NSF) grant 1228496 and US Environmental Protection Agency (EPA) STAR grant 83540901.

#### 699 References

- 700 Atkinson, R., Aschmann, S. M., Carter, W. P. L., Winer, A. M., and Pitts, J. N.: Alkyl nitrate
- 701 formation from the nitrogen oxide (NOx)-air photooxidations of C2-C8 n-alkanes, The Journal of
- 702 Physical Chemistry, 86, 4563-4569, 10.1021/j100220a022, 1982.

703Bates, K. H., Crounse, J. D., St Clair, J. M., Bennett, N. B., Nguyen, T. B., Seinfeld, J. H., Stoltz, B.704M., and Wennberg, P. O.: Gas phase production and loss of isoprene epoxydiols, The journal of

705 physical chemistry. A, 118, 1237-1246, 10.1021/jp4107958, 2014.

706 Beaver, M. R., St Clair, J. M., Paulot, F., Spencer, K. M., Crounse, J. D., LaFranchi, B. W., Min, K. E.,

Pusede, S. E., Wooldridge, P. J., Schade, G. W., Park, C., Cohen, R. C., and Wennberg, P. O.:
 Importance of biogenic precursors to the budget of organic nitrates: observations of
 multifunctional organic nitrates by CIMS and TD-LIF during BEARPEX 2009, Atmospheric

710 Chemistry and Physics, 12, 5773-5785, 10.5194/acp-12-5773-2012, 2012.

711 Carter, W. P. L., and Atkinson, R.: Development and evaluation of a detailed mechanism for the 712 atmospheric reactions of isoprene and NOx, International Journal of Chemical Kinetics, 28, 497-

713 530, 10.1002/(SICI)1097-4601(1996)28:7<497::AID-KIN4>3.0.CO;2-Q, 1996.

714 Chen, X., Hulbert, D., and Shepson, P. B.: Measurement of the organic nitrate yield from OH 715 reaction with isoprene, Journal of Geophysical Research, 103, 25563, 10.1029/98jd01483, 1998.

Crounse, J. D., McKinney, K. A., Kwan, A. J., and Wennberg, P. O.: Measurement of Gas-Phase
Hydroperoxides by Chemical Ionization Mass Spectrometry, Analytical chemistry, 78, 6726-6732,
10.1021/ac0604235, 2006.

Crounse, J. D., Paulot, F., Kjaergaard, H. G., and Wennberg, P. O.: Peroxy radical isomerization in
the oxidation of isoprene, Physical chemistry chemical physics : PCCP, 13, 13607-13613,
10.1039/c1cp21330j, 2011.

Crounse, J. D., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation
of Organic Compounds in the Atmosphere, The Journal of Physical Chemistry Letters, 4, 35133520, 10.1021/jz4019207, 2013.

Day, D. A., Wooldridge, P. J., Dillon, M. B., Thornton, J. A., and Cohen, R. C.: A thermal
dissociation laser-induced fluorescence instrument for in situ detection of NO2, peroxy nitrates,
alkyl nitrates, and HNO3, Journal of Geophysical Research: Atmospheres, 107, ACH 4-1-ACH 414, 10.1029/2001JD000779, 2002.

Fan, J., and Zhang, R.: Atmospheric Oxidation Mechanism of Isoprene, Environmental Chemistry,
1, 140, 10.1071/en04045, 2004.

Giacopelli, P., Ford, K., Espada, C., and Shepson, P. B.: Comparison of the measured and
simulated isoprene nitrate distributions above a forest canopy, Journal of Geophysical Research,
110, D01304, 10.1029/2004jd005123, 2005.

734 Gilman, J. B., Burkhart, J. F., Lerner, B. M., Williams, E. J., Kuster, W. C., Goldan, P. D., Murphy, P.

- C., Warneke, C., Fowler, C., Montzka, S. A., Miller, B. R., Miller, L., Oltmans, S. J., Ryerson, T. B.,
  Cooper, O. R., Stohl, A., and de Gouw, J. A.: Ozone variability and halogen oxidation within the
  Arctic and sub-Arctic springtime boundary layer, Atmos. Chem. Phys., 10, 10223-10236,
- 738 10.5194/acp-10-10223-2010, 2010.

739 Grosjean, D., Williams, E. L., and Grosjean, E.: Atmospheric chemistry of isoprene and of its 740 carbonyl products, Environmental science & technology, 27, 830-840, 10.1021/es00042a004,

740 carbor 741 1993.

742 Grossenbacher, J. W., Couch, T., Shepson, P. B., Thornberry, T., Witmer-Rich, M., Carroll, M. A.,

743 Faloona, I., Tan, D., Brune, W., Ostling, K., and Bertman, S.: Measurements of isoprene nitrates

- 744above a forest canopy, Journal of Geophysical Research, 106, 24429, 10.1029/2001jd900029,7452001.
- Grossenbacher, J. W., Barket Jr, D. J., Shepson, P. B., Carroll, M. A., Olszyna, K., and Apel, E.: A
  comparison of isoprene nitrate concentrations at two forest-impacted sites, Journal of
  Geophysical Research: Atmospheres, 109, D11311, 10.1029/2003JD003966, 2004.
- Guenther, A., Karl, T., Harley, P., Wiedinmyer, C., Palmer, P. I., and Geron, C.: Estimates of global
  terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from
  Nature), Atmos. Chem. Phys., 6, 3181-3210, 10.5194/acp-6-3181-2006, 2006.
- Hao, C., Shepson, P. B., Drummond, J. W., and Muthuramu, K.: Gas Chromatographic Detector
- for Selective and Sensitive Detection of Atmospheric Organic Nitrates, Analytical chemistry, 66,
   3737-3743, 10.1021/ac00093a032, 1994.
- Hartsell, B. E., Aneja, V. P., and Lonneman, W. A.: Relationships between peroxyacetyl nitrate,
  O3, and NOy at the rural Southern Oxidants Study site in central Piedmont, North Carolina, site
  SONIA, Journal of Geophysical Research: Atmospheres, 99, 21033-21041, 10.1029/94JD01021,
  1994.
- Harwood, L. M., Casy, G., and Sherlock, J.: A Simple Laboratory Procedure for Preparation of (1Methylethenyl)oxirane (3,4-Epoxyisoprene), Synthetic Communications, 20, 1287-1292,
  10.1080/00397919008052839, 1990.
- Hasson, A. S., Tyndall, G. S., and Orlando, J. J.: A Product Yield Study of the Reaction of HO2
  Radicals with Ethyl Peroxy (C2H5O2), Acetyl Peroxy (CH3C(O)O2), and Acetonyl Peroxy
  (CH3C(O)CH2O2) Radicals, The Journal of Physical Chemistry A, 108, 5979-5989,
  10.1021/jp048873t, 2004.
- Hasson, A. S., Kuwata, K. T., Arroyo, M. C., and Petersen, E. B.: Theoretical studies of the
  reaction of hydroperoxy radicals (HO2) with ethyl peroxy (CH3CH2O2), acetyl peroxy
  (CH3C(O)O2), and acetonyl peroxy (CH3C(O)CH2O2) radicals, Journal of Photochemistry and
  Photobiology A: Chemistry, 176, 218-230, <u>http://dx.doi.org/10.1016/j.jphotochem.2005.08.012</u>,
  2005.
- Hasson, A. S., Tyndall, G. S., Orlando, J. J., Singh, S., Hernandez, S. Q., Campbell, S., and Ibarra, Y.:
  Branching ratios for the reaction of selected carbonyl-containing peroxy radicals with
  hydroperoxy radicals, The journal of physical chemistry. A, 116, 6264-6281, 10.1021/jp211799c,
  2012.
- Hastie, D. R., Shepson, P. B., Sharma, S., and Schiff, H. I.: The influence of the nocturnal boundary layer on secondary trace species in the atmosphere at Dorset, Ontario, Atmospheric Environment. Part A. General Topics, 27, 533-541, <u>http://dx.doi.org/10.1016/0960-</u> 1686(93)90210-P, 1993.
- Hidy, G. M., Blanchard, C. L., Baumann, K., Edgerton, E., Tanenbaum, S., Shaw, S., Knipping, E.,
  Tombach, I., Jansen, J., and Walters, J.: Chemical climatology of the southeastern United States,
  1999–2013, Atmospheric Chemistry and Physics, 14, 11893-11914, 10.5194/acp-1411893-2014, 2014.
- Horowitz, L. W., Fiore, A. M., Milly, G. P., Cohen, R. C., Perring, A., Wooldridge, P. J., Hess, P. G.,
  Emmons, L. K., and Lamarque, J.-F.: Observational constraints on the chemistry of isoprene
  nitrates over the eastern United States, Journal of Geophysical Research, 112, D12S08,
  10.1029/2006jd007747, 2007.
  - 26

- Hu, K. S., Darer, A. I., and Elrod, M. J.: Thermodynamics and kinetics of the hydrolysis of
   atmospherically relevant organonitrates and organosulfates, Atmospheric Chemistry and
   Physics, 11, 8307-8320, 10.5194/acp-11-8307-2011, 2011.
- Jacobs, M. I., Burke, W. J., and Elrod, M. J.: Kinetics of the reactions of isoprene-derived
  hydroxynitrates: gas phase epoxide formation and solution phase hydrolysis, Atmospheric
  Chemistry and Physics, 14, 8933-8946, 10.5194/acp-14-8933-2014, 2014.
- Jenkin, M. E., Saunders, S. M., and Pilling, M. J.: The tropospheric degradation of volatile organic
   compounds: a protocol for mechanism development, Atmospheric Environment, 31, 81-104,
   http://dx.doi.org/10.1016/S1352-2310(96)00105-7, 1997.
- 796 Kwan, A. J., Chan, A. W. H., Ng, N. L., Kjaergaard, H. G., Seinfeld, J. H., and Wennberg, P. O.:
- Peroxy radical chemistry and OH radical production during the NO<sub>3</sub>-initiated
  oxidation of isoprene, Atmospheric Chemistry and Physics, 12, 7499-7515, 10.5194/acp-127499-2012, 2012.
- Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., Kurtén, T., Worsnop, D. R., and Thornton, J. A.: An
  Iodide-Adduct High-Resolution Time-of-Flight Chemical-Ionization Mass Spectrometer:
  Application to Atmospheric Inorganic and Organic Compounds, Environmental science &
  technology, 48, 6309-6317, 10.1021/es500362a, 2014a.
- 804 Lee, B. H., Mohr, C., Lopez-Hilfiker, F. D., D'Ambro1, E. L., Lutz, A., Hallquist, M., Lee, L., Romer,
- 805 P., Cohen, R. C., Iyer, S., Kurten, T., Hu, W. W., Day, D. A., Campuzano-Jost, P., Jimenez, J. L., Xu,
- 806 L., Ng, N. L., Wild, R. J., Brown, S. S., Koss, A., Gouw, J. d., Olson, K., Goldstein, A. H., Seco, R.,
- 807 Kim, S., McAvey, K., Shepson, P. B., Baumann, K., Edgerton, E. S., Nguyen, T. B., Wennberg, P. O.,
- 808 Liu, J., Shilling, J. E., and Thornton, J. A.: Highly functionalized particle-phase organic nitrates 809 observed in the Southeastern U.S.: contribution to secondary organic aerosol and reactive
- 809 observed in the Southeastern U.S.: contribution to secondary organic aerosol and reactive810 nitrogen budgets, in preparation, 2015.
- Lee, L., Teng, A. P., Wennberg, P. O., Crounse, J. D., and Cohen, R. C.: On rates and mechanisms
  of OH and O3 reactions with isoprene-derived hydroxy nitrates, The journal of physical
  chemistry. A, 118, 1622-1637, 10.1021/ip4107603, 2014b.
- 814 Liao, J., Sihler, H., Huey, L. G., Neuman, J. A., Tanner, D. J., Friess, U., Platt, U., Flocke, F. M.,
- Orlando, J. J., Shepson, P. B., Beine, H. J., Weinheimer, A. J., Sjostedt, S. J., Nowak, J. B., Knapp,
  D. J., Staebler, R. M., Zheng, W., Sander, R., Hall, S. R., and Ullmann, K.: A comparison of Arctic
  BrO measurements by chemical ionization mass spectrometry and long path-differential optical
  absorption spectroscopy, Journal of Geophysical Research, 116, D00R02,
  10.1029/2010jd014788, 2011.
- Liu, Y. J., Herdlinger-Blatt, I., McKinney, K. A., and Martin, S. T.: Production of methyl vinyl
  ketone and methacrolein via the hydroperoxyl pathway of isoprene oxidation, Atmospheric
  Chemistry and Physics, 13, 5715-5730, 10.5194/acp-13-5715-2013, 2013.
- Lockwood, A. L., Shepson, P. B., Fiddler, M. N., and Alaghmand, M.: Isoprene nitrates:
  preparation, separation, identification, yields, and atmospheric chemistry, Atmospheric
  Chemistry and Physics, 10, 6169-6178, 10.5194/acp-10-6169-2010, 2010.
- 826Madronich, S., and Flocke, S.: The role of solar radiation in atmospheric chemistry, in: Handbook827of Environmental Chemistry, edited by: Boule, P., Springer-Verlag, Heidelberg, 1-26, 1998.
- 828 Mielke, L. H., Pratt, K. A., Shepson, P. B., McLuckey, S. A., Wisthaler, A., and Hansel, A.: 829 Quantitative Determination of Biogenic Volatile Organic Compounds in the Atmosphere Using

- Proton-Transfer Reaction Linear Ion Trap Mass Spectrometry, Analytical chemistry, 82, 79527957, 10.1021/ac1014244, 2010.
- 832 Misztal, P. K., Guenther, A., and Goldstein, A. H.: Flux observations of isoprene oxidation 833 products above forests point to potential role of leaf-surface reactions, In preparation.
- Neu, U., Künzle, T., and Wanner, H.: On the relation between ozone storage in the residual layer
  and daily variation in near-surface ozone concentration A case study, Boundary-Layer
  Meteorol, 69, 221-247, 10.1007/BF00708857, 1994.
- Nguyen, T. B., Coggon, M. M., Bates, K. H., Zhang, X., Schwantes, R. H., Schilling, K. A., Loza, C. L.,
  Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Organic aerosol formation from the reactive
- uptake of isoprene epoxydiols (IEPOX) onto non-acidified inorganic seeds, Atmospheric
   Chemistry and Physics, 14, 3497-3510, 10.5194/acp-14-3497-2014, 2014a.
- Nguyen, T. B., Crounse, J. D., Schwantes, R. H., Teng, A. P., Bates, K. H., Zhang, X., St. Clair, J. M.,
  Brune, W. H., Tyndall, G. S., Keutsch, F. N., Seinfeld, J. H., and Wennberg, P. O.: Overview of the
  Focused Isoprene eXperiment at the California Institute of Technology (FIXCIT): mechanistic
  chamber studies on the oxidation of biogenic compounds, Atmospheric Chemistry and Physics,
- 845 14, 13531-13549, 10.5194/acp-14-13531-2014, 2014b.
- Nguyen, T. B., Crounse, J. D., Teng, A. P., St Clair, J. M., Paulot, F., Wolfe, G. M., and Wennberg,
  P. O.: Rapid deposition of oxidized biogenic compounds to a temperate forest, Proceedings of
  the National Academy of Sciences of the United States of America, 10.1073/pnas.1418702112,
- 849 2015.
- O'Brien, J. M., Shepson, P. B., Muthuramu, K., Hao, C., Niki, H., Hastie, D. R., Taylor, R., and
  Roussel, P. B.: Measurements of alkyl and multifunctional organic nitrates at a rural site in
  Ontario, Journal of Geophysical Research: Atmospheres, 100, 22795-22804,
  10.1029/94JD03247, 1995.
- Patchen, A. K., Pennino, M. J., Kiep, A. C., and Elrod, M. J.: Direct kinetics study of the product forming channels of the reaction of isoprene-derived hydroxyperoxy radicals with NO,
   International Journal of Chemical Kinetics, 39, 353-361, 10.1002/kin.20248, 2007.
- Paulot, F., Crounse, J. D., Kjaergaard, H. G., Kroll, J. H., Seinfeld, J. H., and Wennberg, P. O.:
  Isoprene photooxidation: new insights into the production of acids and organic nitrates, Atmos.
  Chem. Phys., 9, 1479-1501, 10.5194/acp-9-1479-2009, 2009.
- Paulot, F., Henze, D. K., and Wennberg, P. O.: Impact of the isoprene photochemical cascade on
  tropical ozone, Atmospheric Chemistry and Physics, 12, 1307-1325, 10.5194/acp-12-1307-2012,
  2012.
- 863 Peeters, J., Nguyen, T. L., and Vereecken, L.: HOx radical regeneration in the oxidation of 864 isoprene, Physical chemistry chemical physics : PCCP, 11, 5935-5939, 10.1039/b908511d, 2009.
- Peeters, J., Müller, J.-F., Stavrakou, T., and Nguyen, V. S.: Hydroxyl Radical Recycling in Isoprene
  Oxidation Driven by Hydrogen Bonding and Hydrogen Tunneling: The Upgraded LIM1
  Mechanism, The Journal of Physical Chemistry A, 118, 8625-8643, 10.1021/jp5033146, 2014.
- 868 Perring, A. E., Wisthaler, A., Graus, M., Wooldridge, P. J., Lockwood, A. L., Mielke, L. H., Shepson,
- P. B., Hansel, A., and Cohen, R. C.: A product study of the isoprene+NO3 reaction, Atmos. Chem.
   Phys., 9, 4945-4956, 10.5194/acp-9-4945-2009, 2009.

- 871 Rindelaub, J. D., McAvey, K. M., and Shepson, P. B.: The photochemical production of organic
- 872 nitrates from  $\alpha$ -pinene and loss via acid-dependent particle phase hydrolysis, Atmospheric 873 Environment, 100, 193-201, http://dx.doi.org/10.1016/j.atmosenv.2014.11.010, 2015.
- 874
- Rivera-Rios, J. C., Nguyen, T. B., Crounse, J. D., Jud, W., St. Clair, J. M., Mikoviny, T., Gilman, J. B., 875 Lerner, B. M., Kaiser, J. B., de Gouw, J., Wisthaler, A., Hansel, A., Wennberg, P. O., Seinfeld, J. H.,
- 876 and Keutsch, F. N.: Conversion of hydroperoxides to carbonyls in field and laboratory
- 877 instrumentation: Observational bias in diagnosing pristine versus anthropogenically controlled
- 878 atmospheric chemistry, Geophysical Research Letters, n/a-n/a, 10.1002/2014gl061919, 2014.
- 879 Rollins, A. W., Kiendler-Scharr, A., Fry, J. L., Brauers, T., Brown, S. S., Dorn, H. P., Dubé, W. P.,
- 880 Fuchs, H., Mensah, A., Mentel, T. F., Rohrer, F., Tillmann, R., Wegener, R., Wooldridge, P. J., and 881 Cohen, R. C.: Isoprene oxidation by nitrate radical: alkyl nitrate and secondary organic aerosol 882 yields, Atmos. Chem. Phys., 9, 6685-6703, 10.5194/acp-9-6685-2009, 2009.
- 883 Rollins, A. W., Smith, J. D., Wilson, K. R., and Cohen, R. C.: Real Time In Situ Detection of Organic 884 Nitrates in Atmospheric Aerosols, Environmental science & technology, 44, 5540-5545, 885 10.1021/es100926x, 2010.
- 886 Saunders, S. M., Jenkin, M. E., Derwent, R. G., and Pilling, M. J.: Protocol for the development of 887 the Master Chemical Mechanism, MCM v3 (Part A): tropospheric degradation of non-aromatic 888 volatile organic compounds, Atmos. Chem. Phys., 3, 161-180, 10.5194/acp-3-161-2003, 2003.
- 889 Savee, J. D., Papajak, E., Rotavera, B., Huang, H., Eskola, A. J., Welz, O., Sheps, L., Taatjes, C. A., 890
- Zádor, J., and Osborn, D. L.: Direct observation and kinetics of a hydroperoxyalkyl radical 891 (QOOH), Science, 347, 643-646, 10.1126/science.aaa1495, 2015.
- 892 So, S., Kirk, B. B., Trevitt, A. J., Wille, U., Blanksby, S. J., and da Silva, G.: Unimolecular reaction 893 chemistry of a charge-tagged beta-hydroxyperoxyl radical, Physical Chemistry Chemical Physics, 894 16, 24954-24964, 10.1039/C4CP02981J, 2014.
- Sprengnether, M., Demerjian, K. L., Donahue, N. M., and Anderson, J. G.: Product analysis of the 895 896 OH oxidation of isoprene and 1,3-butadiene in the presence of NO, Journal of Geophysical 897 Research, 107, 10.1029/2001jd000716, 2002.
- 898 St. Clair, J. M., Rivera, J. C., Crounse, J. D., Knap, H. C., Bates, K. H., Teng, A. P., Jørgensen, S., 899 Kjaergaard, H. G., Keutsch, F. N., and Wennberg, P. O.: Kinetics and Products of the Reaction of 900 the First-Generation Isoprene Hydroxy Hydroperoxide (ISOPOOH) with OH, The Journal of 901 Physical Chemistry A, 10.1021/acs.jpca.5b06532, 2015.
- 902 Stevens, P., L'Esperance, D., Chuong, B., and Martin, G.: Measurements of the kinetics of the 903 OH-initiated oxidation of isoprene: Radical propagation in the OH + isoprene + O2 + NO reaction 904 system, International Journal of Chemical Kinetics, 31, 637-643, 10.1002/(SICI)1097-
- 905 4601(1999)31:9<637::AID-KIN5>3.0.CO;2-O, 1999.
- 906 Stutz, J.: Vertical profiles of NO3, N2O5, O3, and NOxin the nocturnal boundary layer: 1. 907 Observations during the Texas Air Quality Study 2000, Journal of Geophysical Research, 109, 908 D12306, 10.1029/2003jd004209, 2004.
- 909 Surratt, J. D., Chan, A. W., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S. P.,
- 910 Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in secondary
- 911 organic aerosol formation from isoprene. Proceedings of the National Academy of Sciences of
- 912 the United States of America, 107, 6640-6645, 10.1073/pnas.0911114107, 2010a.

- 913 Surratt, J. D., Chan, A. W. H., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S. P.,
- 914 Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in secondary
- 915 organic aerosol formation from isoprene, Proceedings of the National Academy of Sciences of
- 916 the United States of America, 107, 6640-6645, 10.1073/pnas.0911114107, 2010b.
- 917 Teng, A. P., Crounse, J. D., Lee, L., St. Clair, J. M., Cohen, R. C., and Wennberg, P. O.: Hydroxy 918 nitrate production in the OH-initiated oxidation of alkenes, Atmos. Chem. Phys., 15, 4297-4316,
- 919 10.5194/acp-15-4297-2015, 2015.
- 920 Tuazon, E. C., and Atkinson, R.: A product study of the gas-phase reaction of Isoprene with the 921 OH radical in the presence of NOx, International Journal of Chemical Kinetics, 22, 1221-1236, 922 10.1002/kin.550221202, 1990.
- 923 Xie, Y., Paulot, F., Carter, W. P. L., Nolte, C. G., Luecken, D. J., Hutzell, W. T., Wennberg, P. O.,
- 924 Cohen, R. C., and Pinder, R. W.: Understanding the impact of recent advances in isoprene 925 photooxidation on simulations of regional air quality, Atmospheric Chemistry and Physics, 13, 926 8439-8455, 10.5194/acp-13-8439-2013, 2013.
- 927
- Xu, L., Guo, H., Boyd, C. M., Klein, M., Bougiatioti, A., Cerully, K. M., Hite, J. R., Isaacman-928 VanWertz, G., Kreisberg, N. M., Knote, C., Olson, K., Koss, A., Goldstein, A. H., Hering, S. V., de
- 929 Gouw, J., Baumann, K., Lee, S. H., Nenes, A., Weber, R. J., and Ng, N. L.: Effects of anthropogenic
- 930 emissions on aerosol formation from isoprene and monoterpenes in the southeastern United
- 931 States, Proceedings of the National Academy of Sciences of the United States of America, 112,
- 932 37-42, 10.1073/pnas.1417609112, 2015a.
- 933 Xu, L., Suresh, S., Guo, H., Weber, R. J., and Ng, N. L.: Aerosol characterization over the 934 southeastern United States using high-resolution aerosol mass spectrometry: spatial and 935 seasonal variation of aerosol composition and sources with a focus on organic nitrates, 936 Atmospheric Chemistry and Physics, 15, 7307-7336, 10.5194/acp-15-7307-2015, 2015b.
- 937 Zhang, L., Moran, M. D., Makar, P. A., Brook, J. R., and Gong, S.: Modelling gaseous dry 938 deposition in AURAMS: a unified regional air-quality modelling system, Atmospheric 939 Environment, 36, 537-560, http://dx.doi.org/10.1016/S1352-2310(01)00447-2, 2002.

940

943

944



# 946 Table 1. Hydroxynitrates from OH-initiated isoprene oxidation (high $NO_x$ ).

# 949 Table 2. Initial conditions for IN yield experiments.

Expt. number	Isoprene (ppb)	Isopropyl nitrite (ppb)	NO (ppb)	Expt. duration (min)
1	140	180	160	16
2	80	180	130	15
3	70	180	130	12
4	120	180	125	14
5	90	180	220	14
6	75	180	210	12
7	85	180	2400	54



953 Figure 1. GC-ECD/CIMS setup for the CIMS sensitivity of 1,2-IN relative to 4,3-IN.



956 Figure 2. GC-ECD/CIMS chromatogram with water (a) and without water (b) added to

 $\,$  the CIMS. The ECD signal is in black and the CIMS signal is in red.



Figure 3. IN and isoprene data for chamber experiments. An average yield of 9% wasobtained from data of the seven experiments.





964 Figure 4. IN observed during SOAS.



968 Figure 5. IN diurnal average from May 28 to Jul 11. The blue shade indicates day-to-day
969 variation (1σ). The abrupt drop of concentration at 7 PM is caused by instrument
970 fluctuation during its daily maintenance.



973 Figure 6. Diurnal average of OH (a, Jun 13 - Jul 3), isoprene (b, Jun 16 - Jul 11), NO<sub>x</sub> (c,

974 Jun 1 - Jul 15) and sum of MVK and MACR (d, Jun 16 - Jul 11).



Figure 7. (a) Diurnal average of RO<sub>2</sub> loss rates for reaction with NO, HO<sub>2</sub> and RO<sub>2</sub> from Jun 22 to Jul 7. (b) Diurnal average of  $\gamma$  value and production rates of IN and MVK+MACR from Jun 22 to Jul 7. For MVK+MACR, production from the three reaction channels are shown in different colors.



Figure 8. Simulated and observed [IN]/([MVK]+[MACR]) ratio. (a) Results for each
selected days. (b) Averaged results over the 12 days. The error bars represent day-to-day
variation. (c) Sensitivity test with IN yield set as 6%, 9% and 12% in the model. (d)
MVK+MACR data was adjusted by subtracting observed IEPOX+ISOPOOH
concentration from observed MVK+MACR concentration. (e) Results with enhanced IN
photolysis rate.



Figure 9 Modeled IN in the residual layer and IN observed near ground before dawn the
next day. The model includes IN production from isoprene oxidation by NO<sub>3</sub> and IN
consumption by reaction with OH, O<sub>3</sub> and NO<sub>3</sub>. The modeled IN may be biased, as
concentration change caused by transport is not considered.



Figure 10 Diurnal averages of IN and ISOPOOH+IEPOX from May 30 to July 4.

1000



1002Figure 11 Possible oxidation mechanism for 1,2-IN with  $\gamma$ =0.3.1003