

1 **Chemical composition and size distribution of summertime**
2 **PM_{2.5} at a high altitude remote location in the northeast of the**
3 **Qinghai-Xizang (Tibet) Plateau: Insights into aerosol sources**
4 **and processing in free troposphere**

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18

19 **Abstract**

20 Aerosol filter samples were collected at a high-elevation mountain observatory (4180 m above
21 sea level) in the northeastern part of the Qinghai-Xizang (Tibet) Plateau (QXP) during summer
22 2012 using a low-volume sampler and a micro-orifice uniform deposit impactor (MOUDI). These
23 samples were analyzed for water-soluble inorganic ions (WSIs), organic carbon (OC), elemental
24 carbon (EC), water soluble organic carbon (WSOC), and total organic nitrogen (TON) to
25 elucidate the size-resolved chemical composition of free tropospheric aerosols in the QXP region.
26 The average mass concentration of the sum of the analyzed species in $PM_{2.5}$ (WSIs + OC + EC
27 +TON) was $3.74 \mu\text{g sm}^{-3}$, 36% of which was sulfate, 18% OC, 17% nitrate, 10% ammonium, 6.6%
28 calcium, 6.4% TON, 2.6% EC, 1.5% sodium, 0.9% chloride, 0.5% magnesium, and 0.3%
29 potassium. The size distributions of sulfate and ammonium peaked in the accumulation mode
30 (0.32–0.56 μm), whereas the size distributions of both nitrate and calcium peaked in the range of
31 1.8–3.2 μm , suggesting the formation of nitrate on mineral dust. OC, EC and TON were also
32 predominantly found in the accumulation mode. The bulk chemical composition and the average
33 oxidation degree of water-soluble organic matter (WSOM) were assessed using a high resolution
34 time-of-flight aerosol mass spectrometer (HR-ToF-AMS). WSOM was found to be highly
35 oxidized in all $PM_{2.5}$ samples with an average oxygen-to-carbon atomic ratio (O/C) of 1.16 and an
36 organic mass-to-carbon ratio (OM/OC) of 2.75. The highly oxidized WSOM was likely related to
37 active cloud processing during upslope air mass transport coupled with strongly oxidizing
38 environments caused by snow/ice photochemistry. High average ratios of OC/EC (7.6) and
39 WSOC/OC (0.79) suggested that organic aerosols were primarily made of secondary species.
40 SOA was estimated on average accounting for 80% (62–96%) of the $PM_{2.5}$, indicating that SOA
41 is an important component of free tropospheric aerosols over the northern QXP.

42 **1 Introduction**

43 The Qinghai-Xizang (Tibet) Plateau (QXP), often called “the Third Pole” (Yao et al., 2012), is
44 one of the most remote and isolated regions in the world. The high altitude of this region has long
45 been recognized as ideal for studying the long-range transported air pollutants. However,
46 measurements from this area have been rare, usually due to the harsh natural condition and
47 logistic difficulties. These restrictions have become less problematic in the last decade because of
48 the development of mountain observatories and improvements in sampling instrumentation (Li et
49 al., 2000; Cong et al., 2007; Bonasoni et al., 2008; Hegde and Kawamura, 2012; Sang et al.,
50 2013).

51 Previous studies in the QXP region have focused on the chemical properties of aerosols and their
52 source signatures due to the important roles of aerosols on climate forcing. A few studies
53 conducted at Himalayas revealed that aerosols in this region are a complex mixture of inorganic
54 and organic compounds (Carrico et al., 2003; Rengarajan et al., 2007; Decesari et al., 2010; Ram
55 et al., 2010). Mineral dust was generally found to be an important constituent of aerosols in the
56 QXP region because of the presence of large arid and semi-arid areas in “High Asia”. A relatively
57 large proportion of carbonaceous aerosols was also observed when the region was influenced by
58 air masses transported from South Asia, where widespread usage of biofuels has led to large
59 emissions of biomass burning aerosols (Engling et al., 2011; Zhao et al., 2013). Indeed, analysis
60 of the chemical compositions of snow pit samples collected from a glacier in the central
61 Himalayas indicated that biomass burning particles were significantly enhanced in snow during
62 the winter – spring periods, due to transport of polluted air masses from northwest India and
63 Nepal (Xu et al., 2013b).

64 Although inorganic species (such as mineral dust) are important aerosol components in the QXP,
65 their chemical process in the atmosphere are reasonably well characterized due to the small

66 number of inorganic species and their relatively simple chemistry. There have been increased
67 interests in the organic constituents of aerosol particles in recent years because of their high
68 abundances, complex chemical processing, and important roles in affecting cloud properties (e.g.,
69 Kanakidou et al., 2005; Jimenez et al., 2009). Organic aerosol (OA) is usually dominated by
70 secondary species in remote regions (Zhang et al., 2007) because of atmospheric aging processes
71 during long-range transport. For example, the average oxygen-to-carbon (O/C) atomic ratio,
72 which is an indicator for the oxidation degree of OA, observed in a remote site in western Canada
73 was 0.83 (Sun et al., 2009), similar to the O/C ratios of highly aged, low-volatility (LV)
74 oxygenated organic aerosol (OOA) in the atmosphere determined via positive matrix factorization
75 (PMF) analysis of the aerosol mass spectrometer (AMS) spectra of OA (Ulbrich et al., 2009;
76 Zhang et al., 2011). However, the extent of OA oxidation and the composition of organic species
77 are closely related to the source characteristics and the aging processes they are involved, which
78 have never been carefully evaluated and determined in the QXP.

79 Most previous studies of aerosol chemistry in the QXP were conducted in the Himalayan regions
80 because of the key roles of Himalayas on regional climate and environment. There have been
81 very few studies on aerosols reported from the northern QXP (Meng et al., 2013), despite the fact
82 that their atmospheric behaviors might be significantly different from those found in the
83 Himalayan regions because of the different climate pattern and aerosol sources in these two
84 regions. For example, aerosols in the northern QXP mainly originate from the inland of China,
85 whereas aerosols in the Himalayas mainly originate in India. Also, fine particles in the northern
86 QXP tend to contain large proportions of sulfate (Xu et al., 2014a; Zhang et al., 2014b), while
87 fine particles in the Himalayas are usually dominated by carbonaceous material.

88 The Qilian Shan Station of Glaciology and Ecologic Environment (QSS) is situated on the
89 northern slope of the western Qilian Shan Mountains, in the northeastern part of the QXP (Fig.
90 S1). Due to its high elevation (4180 m a.s.l) and long distance (200 km) from local pollution

91 sources, the QSS is well suited for sampling background air masses. In addition, the QSS locates
92 near the termini of several glaciers, making this area a unique atmospheric environment where the
93 photochemistry of snow and glaciers has a relatively strong effect.

94 In our previous studies, the seasonal variations of the mass concentration of water soluble ions
95 and the number concentration of particles at the QSS have been characterized (Xu et al., 2013a;
96 Xu et al., 2014a), which presented two maxima in spring and summer, respectively. The first
97 maximum corresponded to the period during which dust storms predominantly occur in North
98 China, while the second peak corresponded to the period when the thermal circulation between
99 areas of high and low elevation is strongest, which the prevailing valley wind in the northeast of
100 the QSS blew the polluted air masses to this region. In the present study, we performed an
101 intensive field measurement study during the summer of 2012 with the aim of determining the
102 chemical characteristics and sources of fine particles at QSS during summer period. Here, we
103 investigated the PM_{2.5} chemical compositions and the properties of the associated organic
104 chemical species, such as elemental ratios, by applying a suite of instruments including a high
105 resolution time-of-flight mass spectrometer (HR-ToF-AMS). The elemental ratios of organic
106 species are valuable in understanding the oxidation state of OA and thus the aging processes that
107 had occurred during its long-range transport. The size distributions of the chemical species were
108 also assessed to understand the sources and chemical processes of aerosol.

109 **2 Sample collection and analysis**

110 **2.1. Aerosol sampling**

111 The samples were collected at the QSS atmospheric chemistry observatory (39.50 N, 96.51 E;
112 4180 m a.s.l., Fig. S1). The sampling site and the QSS were described in detail by Xu et al.,
113 (2013). The summer climate at the QSS is dominated by the East Asian monsoon, which brings

114 about half of the annual precipitation (360 mm during 2008–2010). This study was conducted
115 from 11 July to 6 September 2012. PM_{2.5} samples were collected using a low-volume (16.7 L/min
116 flow rate) aerosol sampler (BGI, USA, model PQ 200), powered by solar cells. The instrument
117 was regularly calibrated using a TetraCal[®] calibrator (BGI, USA). All samples were collected on
118 47 mm quartz fiber filters (Whatman, Maidstone, England). The flow rate was measured at 5 min
119 intervals by an internal volume flow meter, and the recorded flow data were used to calculate the
120 volume of air sampled. Meteorological parameters including wind speed, wind direction,
121 temperature, precipitation, and relative humidity were recorded at 30 min intervals at the
122 meteorological station at the QSS, and the recorded data were used to calculate the air volume
123 sampled at standard temperature and pressure (STP: 1013 hPa and 273 K). Collection of each
124 sample started in the morning and continued for 3 d, and a total of 19 aerosol filter samples were
125 obtained. Three procedure blanks were collected in the field to assess potential contamination
126 occurred during sampling preparation, transportation, and storage. The sampling volume ranged
127 from 42 to 44 m³ under STP (i.e., sm³), with a mean ($\pm 1\sigma$) value of 43.2 ± 0.5 sm³. Noted that
128 the mass concentrations reported here are all based on STP (e.g., $\mu\text{g sm}^{-3}$).

129 In addition, measurements of the size distribution of chemical species were made at the QSS from
130 July 21 to September 4, 2012. A 10-stage multi-nozzle micro orifice uniform deposit impactor
131 (MOUDI, Model R110, MSP Corp., Shoreview, MN) was used to sample particles at a flow rate
132 of 30 L/min over a size range of 0.056–18 μm with nominal cut offsets of 0.056, 0.10, 0.18, 0.32,
133 0.56, 1.0, 1.8, 2.5, 5.6, 10, and 18 μm . The air pump was calibrated before each sample was
134 collected and the pump was closely monitored to identify any changes of the flow rate. The
135 collection substrates were 47 mm quartz fiber filters which had been heated at 500 $^{\circ}\text{C}$ for 8 h
136 prior to sample collection to remove adsorbed organic material. The sampling time varied
137 between ~20 h and ~120 h depending on weather conditions. A total of four sets of filters were

138 obtained. All the filters were placed in individual aluminum-lined plastic boxes and stored at
139 $-15\text{ }^{\circ}\text{C}$ prior to analysis.

140 **2.2 Chemical Analysis**

141 The samples were analyzed to characterize their chemical compositions using a series of
142 instruments, namely ion chromatography (IC) instruments, a total water soluble organic
143 carbon/total nitrogen (TOC/TN) analyzer, a HR-ToF-AMS, and an organic carbon/element
144 carbon (OC/EC) analyzer. For each OC/EC analysis, a piece of filter measuring at 0.526 cm^2 was
145 punched from a sample filter and analyzed directly using the instrument. The rest of that filter
146 was extracted by sonication in 15 mL deionized water for 30 min at $\sim 0\text{ }^{\circ}\text{C}$, and the extract was
147 immediately filtered using a $0.45\text{ }\mu\text{m}$ Acrodisc syringe filters (Pall Life Sciences, Ann Arbor, MI,
148 USA).

149 **IC analysis.** Eight ionic species (Na^+ , NH_4^+ , K^+ , Ca^{2+} , Mg^{2+} , Cl^- , NO_3^- , and SO_4^{2-}) were
150 determined using two IC system (881 Compact IC Pro, Metrohm, Herisau, Switzerland). One of
151 the IC systems was used to determine cations, and was equipped with a Metrosep C4 guard/2.0
152 column and Metrosep C4 250/2.0 column (Metrohm), which were kept at $30\text{ }^{\circ}\text{C}$ during
153 measurements. The other IC system, equipped with a Metrosep RP2 guard/3.6 column and a
154 Metrosep A Supp15 250/4.0 column (Metrohm), and kept at $45\text{ }^{\circ}\text{C}$ during measurements, was
155 used to determine anions. The mobile phase in the cation IC system was 1.75 mM nitric acid
156 (made from 70% nitric acid, Sigma-Aldrich, St Louis, MO, USA) and 0.75 mM dipicolinic acid
157 (made from $\geq 99.5\%$ pure dipicolinic acid, Sigma-Aldrich), and eluted at a flow rate of 0.3 mL
158 min^{-1} . The eluent of anion IC system was 5 mM sodium carbonate and 0.3 mM sodium hydroxide
159 (made from $\geq 98\%$ pure sodium hydroxide, Sigma-Aldrich), and was used at a flow rate of 0.8 mL
160 min^{-1} . The instruments were calibrated using standard cation and anion solutions (Dionex, CA,
161 USA). The IC analysis results were evaluated in terms of the reproducibilities of peak retention

162 times, peak heights, and the linearity of each calibration curve. More details on the IC analysis
163 methods are given in Ge et al. (2014).

164 **TOC and TN analysis.** An aliquot of each sample was analyzed for TOC and TN contents using
165 a high-sensitivity TOC/TN analyzer (TOC-V_{CPH} with a TNM-1unit, Shimadzu, Kyoto, Japan).
166 The measurements were carried out using the total carbon (TC) and inorganic carbon (IC) method.
167 The TC was determined by combusting the sample at 720 °C in a combustion tube filled with
168 oxidation catalyst, which converted all carbon-containing components into CO₂. The CO₂ was
169 detected by a non-dispersive infrared (NDIR) gas analyzer. The IC was defined as the carbon in
170 carbonates and dissolved CO₂ in the sample. The carbonates were transformed into CO₂ by
171 treating the sample with 25% (by weight) phosphoric acid (H₃PO₄) in an IC reaction vessel. The
172 CO₂ was then volatilized using a sparging procedure and detected by the NDIR analyzer. The
173 TOC content was calculated by subtracting the inorganic carbon (IC) content from total carbon
174 (TC) content. In the TN analysis the nitrogen-containing species were decomposed to NO in the
175 combustion tube at 720 °C, then the sample was cooled and dehumidified using an electronic
176 dehumidifier, and the NO was measured using a chemiluminescence gas analyzer. The total
177 organic nitrogen (TON) was determined by subtracting the inorganic nitrogen (IN, which
178 included ammonium and nitrate) content, quantified by IC, from the TN content. The samples
179 were contained in well-sealed, pre-cleaned glass vials during analysis, and the instruments
180 automatically withdrew a sample aliquot after piercing the vial seal. The calibrations were carried
181 out using a potassium hydrogen phthalate standard for the TC determination, a sodium hydrogen
182 carbonate for the IC determination, and potassium hydrogen phthalate and potassium nitrate
183 standards for the TN determination.

184 **HR-ToF-AMS analysis.** Each filtered sample extract was aerosolized using argon and
185 dehumidified using a diffusion dryer. The resulting aerosol particles were sampled into a HR-
186 ToF-AMS instrument (Aerodyne Inc., Billerica, MA, USA) through an aerodynamic lens inlet,

187 vaporized at ~600 °C, ionized by 70 eV electrons in an electron impact ionization chamber, and
188 analyzed using the mass spectrometer. The HR-ToF-AMS analysis procedure for aqueous
189 samples and associated data processing are described in detail elsewhere (Sun et al., 2010, 2011;
190 Xu et al., 2013b; Yu et al., 2014). The HR-ToF-AMS was operated in the “V” and “W” ion
191 optical modes alternatively, spending 2.5 min in each mode. In the V-mode the HR-ToF-AMS
192 spent 6 s in the mass spectrum (MS) mode and then 4 s in the particle time-of-flight (PToF) mode,
193 then continuously repeated this 10 s cycle. In the W-mode the instrument used only the MS mode,
194 with 6 s in each cycle. Between every two samples, purified waters (> 18.2 M cm⁻¹, Millipore,
195 USA) were analyzed in the same way to generate analytical blanks. Each sample was measured
196 twice to check the reproducibility of the analysis. Elemental analyses were performed on high-
197 resolution mass spectra (W-mode, with *m/z* up to 120), and these were used to determine the
198 elemental ratios for oxygen to carbon (O/C), hydrogen to carbon (H/C), nitrogen to carbon (N/C),
199 and organic mass-to-carbon (OM/OC) ratio of the DOM (Aiken et al., 2008). The elemental
200 contributions of C, O, H, and N reported in this study are mass-based. The signals of H₂O⁺ and
201 CO⁺ for organic compounds were not directly measured but scaled to that of CO₂⁺ based on the
202 evaluation of the H₂O⁺ and CO⁺ signal in the samples (Fig. S3): H₂O⁺ = 0.94 × CO₂⁺, CO⁺ = 0.46
203 × CO₂⁺.

204 Potential interferences on the CO₂⁺ signals in the aerosol mass spectra caused by the presence of
205 carbonate salts were evaluated by acidifying a sample (QSS1) to pH 4 using sulfuric acid and
206 then analyzing the sample using the HR-ToF-AMS. Meanwhile, the volatility of the sample was
207 investigated by a digitally controlled thermodenuder (TD) system. The sample was aerosolized
208 and passed through a diffusion dryer, switched between the by-pass (BP) mode and TD mode
209 every 5 min, and finally analyzed using the HR-ToF-AMS instrument. The TD system was
210 programmed to cycle through 12 temperature steps (30, 50, 70, 100, 150, 200, 180, 130, 110, 88,
211 66, and then 40 °C). The design and use of the TD system have been described elsewhere (Fierz

212 et al., 2007). The data were processed in a similar way as the normal HR-ToF-AMS data analysis,
213 and the remaining particle mass calculated from the difference between the results of the TD
214 mode and the BP mode analysis were assumed to indicate the volatility of individual aerosol
215 species.

216 **OC and EC analysis.** The samples were analyzed for OC/EC using a Thermal/Optical Carbon
217 Analyzer (DRI Model 2001). The procedure used has been described in detail elsewhere (Cao et
218 al., 2003). Briefly, the system heated the 0.526 cm² punched quartz filter aliquot gradually to
219 120 °C (fraction OC1), 250 °C (fraction OC2), 450 °C (fraction OC3), and 550 °C (fraction OC4)
220 in a non-oxidizing helium atmosphere, and then to 550 °C (fraction EC1), 700 °C (fraction EC2),
221 and 800 °C (fraction EC3) in an oxidizing atmosphere of 2% oxygen in helium. The carbon
222 evolved at each temperature was oxidized to CO₂, then reduced to methane (CH₄) and quantified
223 using a flame ionization detector. Some of the organic carbon was pyrolyzed to form black
224 carbon as temperature increased in the helium atmosphere, resulting in the darkening of the filter.
225 This darkening was monitored by measuring the decrease in the reflectance of the sample using
226 light at 633 nm from a He-Ne laser. The original black carbon and the pyrolyzed black carbon
227 combusted after being exposed to the oxygen-containing atmosphere, and the reflectance
228 increased. The amount of carbon measured after exposure to the oxygen-containing atmosphere
229 until the reflectance reached its original value was reported as the optically detected pyrolyzed
230 carbon (OPC). The eight fractions, OC1, OC2, OC3, OC4, EC1, EC2, EC3, and OPC were
231 reported separately. OC was defined as OC1 + OC2 + OC3 + OC4 + OPC and EC was defined as
232 EC1 + EC2 + EC3 – OPC in the IMPROVE protocol.

233 **Determination of the water-soluble OM and water-insoluble OM concentrations.** The mass
234 concentrations of water-soluble OM (WSOM) and water-insoluble OM (WIOM), and the average
235 OM/OC ratio for the organic matter (WSOM + WIOM) in PM_{2.5} are estimated using equations 1–
236 3,

237
$$\text{WSOM} = \text{WSOC} \times \text{OM/OC}_{\text{WSOM}} \quad (1)$$

238
$$\text{WIOM} = (\text{OC} - \text{WSOC}) \times 1.3 \quad (2)$$

239
$$\text{OM/OC}_{\text{OM}} = (\text{WSOM} + \text{WIOM}) / \text{OC} \quad (3)$$

240 Where WSOC is the water-soluble organic carbon content in the filter extract measured by the
241 TOC/TN analyzer, $\text{OM/OC}_{\text{WSOM}}$ is the OM/OC ratio of the WSOM (determined from the HR-
242 ToF-AMS measurement), OC is taken from the filter measurements using the thermo/optical
243 carbon analyzer, the constant 1.3 in equation 2 is the estimated OM/OC for WIOM (Sun et al.,
244 2011), and OM/OC_{OM} is the average OM/OC ratio of organic matter in $\text{PM}_{2.5}$.

245 **Estimation of the secondary organic aerosol concentrations.** The secondary organic carbon
246 (SOC) content was estimated by determining the primary organic carbon (POC) content using EC
247 as a tracer, and then subtracting the POC from the measured total OC. The primary OA (POA)
248 concentration was estimated based on POC, which was subtracted from the total OM calculated
249 as the product of the measured OC and the OM/OC_{OM} determined in equation 3 to determine the
250 SOA concentration. The equations for these calculations are shown below,

251
$$\text{POC} = (\text{OC/EC})_{\text{pri}} \times \text{EC} \quad (4)$$

252
$$\text{SOC} = \text{OC} - \text{POC} \quad (5)$$

253
$$\text{POA} = \text{OM/OC}_{\text{POA}} \times \text{POC} \quad (6)$$

254
$$\text{SOA} = \text{OC} \times \text{OM/OC}_{\text{OM}} - \text{POA} \quad (7)$$

255 The calculation of POC was based on the hypothesis that OC and EC correlate strongly and stay
256 at a constant ratio in primary particles within a geographical region. Organic aerosol observed at
257 the QSS was mostly originated from lower attitude regions including urban area to east and west
258 of QSS based on the typical diurnal pattern of wind field around the QSS (Xu et al., 2014a).

259 Based on previous studies, the OM/OC for fresh urban organics in northern China is between
260 1.2–1.6 (e.g., Xu et al., 2014b; Zhang et al., 2014a). Since it is expected that organics would be
261 oxidized gradually during transport, we use the ratio of 1.4 for OM/OC of the POA (OM/OC_{POA})
262 at the QSS.

263 **3 Results and discussion**

264 **3.1 Chemical speciation of PM_{2.5}**

265 The meteorological conditions during the measurement period were overall cold and humid. The
266 air temperature (T) ranged from –5.9 to 14.3 °C, with an average of 4.2 °C, and the relative
267 humidity (RH) ranged from 10 to 99%, with an average of 65% (Fig. 1a). Three-day air mass
268 back trajectories originating at ~100 m above ground level were acquired every 6 h during the
269 sampling period and showed that 72% and 23% of the air masses came from west and east of the
270 sampling site, respectively (Fig. S1). Light precipitation occurred frequently between 11 July and
271 19 August (Fig. 1a) because of the topographic effect, but it was relative dry from 19 August to 6
272 September, probably because of the occurrence of a different synoptic-scale weather pattern. The
273 three-day average wind data also showed that the wind speed from west increased during the late
274 part of the sampling period (Fig. S2). Wind direction changed diurnally, with moderate mountain
275 wind (from the southeast at ~2 m/s) during the night and stronger valley wind (from the north at
276 ~4 m/s) during the day (Fig. 1b). The concentrations of various chemical species, including water
277 soluble ionic species (WSIs), OC, EC, and TON, changed significantly according to weather
278 condition throughout the sampling period. No dust storm event was observed during the study
279 even though the QSS is close to the desert regions.

280 The total mass concentrations of the measured species (WSIs + OC + EC + TON) throughout the
281 sampling period were in the range of 1.8–8.0 µg sm⁻³ with the average ($\pm 1\sigma$) at 3.7 ± 1.9 µg sm⁻³

282 (Fig. 2a), which were lower than that was measured in 2010 ($2.7 \mu\text{g sm}^{-3}$ in 2012 vs. $5.4 \mu\text{g sm}^{-3}$
283 in 2010 for the same WSIs in July and August) at the QSS (Xu et al., 2014a), probably because of
284 the low frequency of dust storm events in 2012. Indeed, mass concentration of calcium was more
285 than 4 times lower in 2012 ($0.27 \mu\text{g sm}^{-3}$) than that in 2010 ($1.2 \mu\text{g sm}^{-3}$). Overall, sulfate was
286 the main contributor (on average 36%) to the aerosol mass concentrations during the observation
287 period (Fig. 2a), similar to previous observations in the northern QXP (Li et al., 2013; Xu et al.,
288 2014a; Zhang et al., 2014b). The mass concentration of sulfate was also lower in 2012 ($1.4 \mu\text{g}$
289 sm^{-3}) than that in 2010 ($2.7 \mu\text{g sm}^{-3}$). Major cations (ammonium (10%), calcium (6.6%), sodium
290 (1.5%), magnesium (0.5%), and potassium (0.3%)) and anions (sulfate (36.2%), nitrate (16.9%),
291 chloride (0.9%)) together accounted for 45–88% (mean = 73%) of the total aerosol mass. The ion
292 balance, expressed as the ratio of the equivalent concentration ($\mu\text{eq sm}^{-3}$) cation to that of anion
293 (C/A), is shown in Fig. 2b. The mean C/A ratio was 0.97, which was close to 1, further indicating
294 that the contribution of mineral dust was negligible since carbonate and bicarbonate were not
295 measured in the IC analyses. The average EC concentration was $0.09 \mu\text{g sm}^{-3}$ during the
296 observation period, and its mean contribution to the total $\text{PM}_{2.5}$ mass was 2.6%. The average (\pm
297 1σ) concentrations of OC and TON were $0.66 (\pm 0.43)$ and $0.24 (\pm 0.16) \mu\text{g sm}^{-3}$, respectively.
298 The OC and EC concentrations found in this study were lower than those found in the summer of
299 2010 at Qinghai Lake (1.6 and $0.4 \mu\text{g sm}^{-3}$), which is also in the northern part of the QXP. The
300 concentrations were different probably because Qinghai Lake is located at a lower elevation
301 (3200 m a.s.l. (Li et al., 2013)) than the QSS (4180 m a.s.l.) and is subjected to more emissions
302 from the boundary layer.

303 The correlation coefficients (r) between all of the chemical species are shown in Table 1, and the
304 strong correlations ($r \geq 0.75$) are shown in bold. In general, strong correlations were found
305 between Na^+ and Ca^{2+} , Mg^{2+} , and Cl^- , which are representative of primary species in mineral salts,
306 and among secondary ions such as SO_4^{2-} , NO_3^- , and NH_4^+ . In addition, the good correlations

307 between secondary ions (SO_4^{2-} and NO_3^-) and K^+ and Mg^{2+} might arise from acid replacement on
308 mineral particles. The facts that WSOC and OC tightly correlate ($r = 0.97$) and that WSOC
309 accounts for 79% of the OC indicate that a majority of OC in fine particles was secondary at QSS
310 (more discussions are given in section 3.3).

311 **3.2 Chemically-resolved size distributions**

312 The size distributions of all species are shown in Fig. 3, and the sum of the species presents a
313 prominent accumulation mode peaking at MOUDI stage of 0.32–0.56 μm and a coarse mode
314 peaking at 1.8–3.2 μm . In the accumulation mode, sulfate dominated (39%) the composition of
315 particles in the size range of 0.18–1 μm , with OC and EC accounting for 24% and 5.0%,
316 respectively, followed by TON (9.4%), NH_4^+ (7.9%), NO_3^- (3.9%), Cl^- (2.8%), K^+ (1.3%), and
317 Ca^{2+} (0.6%). However, OC dominated (36%) the composition of particles smaller than 0.18 μm
318 with sulfate contributing 26%. In the coarse mode, nitrate was the main contributor, accounting
319 for 23% of the particle mass in the size range 1.8–5.6 μm , followed by OC (17%), sulfate (16%),
320 TON (12%), Ca^{2+} (11%), Cl^- (8.4%), and EC (2.2%); The rest of species totally accounted for 3.7%
321 at this size range. TON and OC were important contributors of particle mass over the whole size
322 range (0.056–18 μm).

323 The different size distributions of different species suggested that they had different sources
324 and/or have undergone atmospheric transformation processes. The species that were relatively
325 abundant in the accumulation mode aerosols were mainly secondary species, such as ammonium,
326 sulfate and OC, while the species that were relatively abundant in the coarse mode aerosols were
327 mainly primary mineral ionic species, such as Ca^{2+} , Na^+ , and Cl^- . Nitrate was closely associated
328 with dust particles as a result of its formation through the reactions of HNO_3 gas with carbonate
329 salts (such as calcite and dolomite) (Sullivan et al., 2009) which could form $\text{Ca}(\text{NO}_3)_2$ and
330 $\text{Mg}(\text{NO}_3)_2$ (Li and Shao, 2009). As shown in Fig. 4, the equivalent balances of water-soluble

331 species in different size modes indicate that the accumulation mode particles were somewhat
332 acidic (with the linear regression slope of $[\text{NH}_4^+ + \text{Ca}^{2+} + \text{Mg}^+ + \text{K}^+]$ vs $[\text{SO}_4^{2-} + \text{NO}_3^-]$ being 0.6)
333 and that the coarse mode particles were almost neutral (the slope was 0.999), similar to the results
334 observed at Mt. Hua in 2009 (Wang et al., 2013). The sizes of WSIs in this study were smaller
335 than observed at other sites, e.g., Hong Kong, where the accumulation mode was at 1–1.8 μm and
336 the coarse mode was at 3.2–5.6 μm (Xue et al., 2014). The smaller mode size at the QSS was
337 probably because of the lower specific humidity at QSS.

338 The OC and TON species in the coarse mode probably came from soil organic matter or were
339 formed by the condensation of volatile organic gases on mineral dust. The EC and OC species
340 reaching maximum concentrations in the accumulation mode, consistent with other results,
341 indicating the occurrence of aging and aqueous processing of particles (Yu and Yu, 2009).

342 **3.3 Relationship between OC, EC, and WSOC**

343 The relationship between OC and EC concentrations can provide useful insights into the origin of
344 carbonaceous aerosols, because particles from different sources have different OC/EC ratios. The
345 correlation between EC and OC concentrations at the QSS was statistically significant ($r^2 = 0.4$,
346 $n=19$), with the slope of 3.29 and the intercept of 0.23 (Fig. 5a and b). The OC/EC ratios ranged
347 from 2.8 to 26.4 with an average of 7.6, which is higher than those observed from Chinese urban
348 sites (1 to 4) (Cao et al., 2003). However, the OC/EC ratio at the QSS was similar to the ratios
349 found at remote sites in Western China (Fig. S1) such as Qinghai Lake (6.0 ± 3.9 in $\text{PM}_{2.5}$) during
350 the summer of 2010 (Li et al., 2013), Muztagh Ata Mountain (11.9 in total suspended particles)
351 during the summers of 2004-2005 (Cao et al., 2009), and Akdala (12.2 in PM_{10}) during July 2004
352 and March 2005 (Qu et al., 2009). The high OC/EC ratios observed at the QSS suggest the

353 formation of secondary OC during the long range transport. The chemical characteristics of
354 particulate organics can be further evaluated using the WSOC/OC ratio, which can be used to
355 assess the aging of organic species. High WSOC/OC ratios (> 0.4) have been found in aged
356 aerosols because a significant portion of the OC can be oxidized to WSOC (Ram et al., 2010). A
357 tight correlation between WSOC and OOA was observed previously in Tokyo, indicating that
358 OOA and WSOC have very similar chemical characteristics (Kondo et al., 2007). Field studies
359 have shown that the OOA factors derived from multivariate analysis AMS organic aerosol mass
360 spectra are generally representative of SOA (Zhang et al., 2005, 2011). The WSOC and OC
361 concentrations in the QSS were strongly positively correlated ($r^2 = 0.97$) with a slope of 0.79.
362 This slope is higher than those in Chinese urban sites (0.3–0.6) (Pathak et al., 2011) and at remote
363 sites in Western China, such as Qinghai Lake (0.42) during the summer of 2012 (Li et al., 2013)
364 and Himalayas (0.26–0.51) (Rengarajan et al., 2007; Ram et al., 2010; Shrestha et al., 2010).
365 However, it is similar to the average ($\pm 1\sigma$) ratio of WSOC to OOA (0.88 ± 0.29) in Tokyo
366 (Kondo et al., 2007). The high OC/EC ratios, tight WSOC and OC correlation, and the high
367 WSOC/OC ratios found in the aerosol particles from the QSS can be regarded as a solid evidence
368 for the formation of SOA in the QXP region.

369 **3.4 Spectra characteristics of the WSOM**

370 The average MS for the WSOM in $PM_{2.5}$ is shown in Fig. 6. The major spectral features are the
371 high mass fraction of m/z 44 (f_{44}) (mainly CO_2^+ , 94.6%), m/z 18 (H_2O^+), and m/z 28 (CO^+). The
372 f_{44} peaks were almost identical ($\sim 20\%$) in all filter samples, and in two samples, QSS3 and
373 QSS18, the contributions of f_{44} peaks were particularly high, at 27.4% and 27.3%, respectively.

374 It has previously been suggested that CO_2^+ ion in the HR-ToF-AMS MS is typically associated
375 with the presence of carboxylic acids (Takegawa et al., 2007), which can be the oxidation
376 products of organic species through heterogeneous and homogeneous chemical processes
377 (fragmentation) (Jimenez et al., 2009). This can be supported by the low intensity fragments in
378 the higher m/z range ($m/z > 50$), which were probably caused by the fragmentation of organic
379 species during oxidation and conversion into small organic acids (increased f_{44}). The similarities
380 between the mass spectra of the acidified and untreated QSS1 sample ($r^2 = 0.98$; Fig. S4) ruled
381 out the potential influence of carbonate salts on the intensity of the CO_2^+ peak. The O/C ratio, an
382 oxidation degree index, has a mean value of 1.16 and a range of 0.93–1.66 (Fig. 6a). The mean
383 OM/OC ratio of WSOM in our filter samples was 2.75, which contrasts strongly with the range of
384 1.6–2.2 found in other studies, indicating that secondary organic aerosol made important
385 contributions to the aerosols in our study. The H/C ratio was also relatively high, at 1.89. The
386 high O/C and H/C ratios were caused not only by the high contribution of CO_2^+ ions but also by
387 the high contribution from H_2O^+ , which can be produced by the fragmentation of acidic species.
388 Carbon oxidation state (OSc) values are more robust and less variable than measured H/C and
389 O/C ratios, so we calculated the OSc values ($2 \times \text{O/C} - \text{H/C}$). The mean OSc for the filter samples
390 was 0.4 ranging between 0.1 and 1.2, similar to the OSc values of diacids and multifunctional
391 acids (Canagaratna et al., 2015). The elemental composition of the WSOM in the filter samples
392 was C (36%), O (56%), H (6%), and N (2%). The mass spectrum was composed of H_yO_1^+ (25%),
393 $\text{C}_x\text{H}_y\text{O}_1^+$ (22%), $\text{C}_x\text{H}_y\text{O}_2^+$ (23%), and C_xH_y^+ (25%) ions, indicating that oxygenated functional
394 groups were predominant in the WSOM. The high contribution from H_yO_1^+ ions could be
395 generated from diacids and alcohols (Canagaratna et al., 2015).

396 The high oxidation state of organics at the remote area of northern QXP has been suggested in a
397 previous study at Qinghai Lake in 2010 (Li et al., 2013). We sought further evidence for the high
398 oxidation state of organics in our samples by checking the volatility of OC, which normally has a

399 reverse relationship with the degree of oxidation (Huffman et al., 2009). The OC mass fractions
400 that evaporated at different temperature step were 12% at 120 °C (OC1), 26% at 250 °C (OC2),
401 39% at 450 °C (OC3), and 23% at 550 °C (OC4), respectively. The large fractions of OC at high
402 temperature suggested its low volatility. Indeed, the volatility measurements performed using the
403 TD system on the QSS1 sample indicated that the organic species were less volatile than
404 ammonium and nitrate and more volatile than sulfate (Fig. 7a). However the remaining fraction of
405 organics was higher than sulfate at 180–200 °C (Fig. 7b). The volatility distribution of the
406 WSOM of the QSS1 sample was similar to that of LV-OOA in samples from urban sites
407 (Huffman et al., 2009). Note that the thermal profile of sulfate in our sample is similar to that of
408 ammonium sulfate but nitrate appears to be less volatile than ammonium nitrate based on the
409 laboratory study of Huffman et al., (2009). A possible reason is that the nitrate in filter aerosol
410 was main present in the form of organic nitrate or metal nitrate. The ratio of NO^+ vs NO_2^+ in mass
411 spectrum of QSS filter samples indeed show higher values than that of ammonium nitrate which
412 is usually resulted from the organic nitrate or metal nitrate (Farmer et al., 2010) (Fig. S6). This is
413 consistent with the fact that a significant fraction of nitrate in these samples was likely associated
414 with metals such as calcium and sodium (Fig. 2).

415 The AMS spectra showed that organic species in $\text{PM}_{2.5}$ collected at QSS were on average more
416 oxidized than low-volatility oxygenated organic aerosol factors (e.g., LV-OOA, which had an
417 O/C ratio of 0.5–1) determined using positive matrix factorization analyses of urban aerosols
418 (Aiken et al., 2008; Ng et al., 2010). The higher level of oxidation of organic species in our
419 samples was probably caused by intense photo-chemistry because of the stronger solar radiation
420 in free troposphere and/or aqueous-phase reactions in cloud droplets and particulate water phase
421 during the upslope transport of air mass from the lowlands. Indeed, the highly oxidized organic
422 MS observed in this study is similar to that of oxidized organic matter reported in the study of
423 Lee et al. (2012) ($r^2 = 0.95$, Fig. S5), during which filter samples collected at a mountain site were

424 oxidized in the laboratory using a photochemical reactor. The potentially strong oxidizing
425 environment at the QSS was another important factor in producing the highly oxidized organic
426 species. For example, the photochemical reactions on snow/ice are suspected to the release of
427 reactive gaseous species, including H₂O₂, HONO, and OH (Grannas et al., 2007).

428 **3.5 Estimation of secondary organic aerosol concentrations**

429 The EC-tracer method, which estimates SOC concentration based on the relationship of POC and
430 EC, has been widely used, although significant uncertainty may arise due to the usage of assumed
431 (OC/EC)_{primary} ratios. In this study, a value of 2.0 is used to represent the (OC/EC)_{primary} ratio at
432 QSS, which is commonly used to estimate SOA mass (Chow et al., 1996). In addition, this value
433 is similar to the minimum OC/EC ratios that have been used for estimation of SOA concentration
434 at remote sites in China, such as Mount Heng (2.2) (Zhou et al., 2012) and Mount Tai (2.19)
435 (Wang et al., 2012). The minimum OC/EC (2.8) in the PM_{2.5} samples at QSS is higher than this
436 value, which suggests the filter samples were likely always a mixture of POA and SOA due to the
437 fact that each filter was collected for 3 days.

438 The average concentration of SOC at the QSS during the summer was $0.44 \pm 0.37 \mu\text{g sm}^{-3}$, on
439 average accounting for 64% (29 – 92%) of the total OC (Fig. 8a and b). The average
440 concentration of SOA was $1.27 \pm 0.88 \mu\text{g sm}^{-3}$, on average accounting for 80% (62 – 96%) of the
441 total OA (Fig. 8c and d). These results indicate that secondary aerosols were dominant at the QSS
442 during the summer. The SOC and SOA contributed relatively small to PM_{2.5} mass during the late
443 sampling period, and this is consistent with drier condition and thus less chances of aqueous
444 processing during this period. The SOC contribution at the QSS is consistent with those (36 –
445 52%) at Himalayas during summer. For example, Ram et al. (2008) found, using the EC-tracer
446 method, that SOC contributed 52% of the aerosols at Manora Peak during summer. The estimated
447 SOC contribution is also consistent with the results at Mount Heng (54% of the total OC) and at

448 Mount Tai (57% and 71% of total OC in spring and summer, respectively) (Wang et al., 2012;
449 Zhou et al., 2012).

450 **4 Conclusions**

451 An intensive study was conducted at a high elevation remote site (QSS) in the northern part of the
452 QXP to characterize the chemical compositions of PM_{2.5} during the summer of 2012 – a period
453 that was influenced by strong exchange of air masses between boundary layer and free
454 troposphere. The average $\pm 1\sigma$ mass concentration of PM_{2.5} species, which include WSI (SO₄²⁻,
455 NO₃⁻, NH₄⁺, Ca²⁺, Na⁺, Cl⁻, Mg²⁺, and K⁺), OC, EC, and TON, was $3.7 \pm 1.9 \mu\text{g sm}^{-3}$. Mineral
456 dust appeared to be a minor PM_{2.5} component during this study. SO₄²⁻ was a main contributor to
457 the PM_{2.5} mass (36%), followed by OC (18%), NO₃⁻ (17%), NH₄⁺ (10%), Ca²⁺ (6.6%), TON
458 (6.4%), and EC (2.6%). The size distribution of the particles presented a bimodal distribution
459 with a prominent accumulation mode (0.32–0.56 μm) and a coarse mode (1.8–3.2 μm). Sulfate,
460 OC, and EC dominated the accumulation mode (contributing ~70% of the mass), while nitrate
461 was a main contributor to the coarse mode (contributing 23%), followed by OC (17%), sulfate
462 (16%), TON (12%), Ca²⁺ (11%), Cl⁻ (8.4%), and EC (2.2%). Stoichiometry analysis indicated
463 that submicrometer particles were on average acidic whereas coarse particles were mostly neutral.
464 The facts that OC/EC ratios were high (2.8 – 26.4) and that a major fraction of OC was water
465 soluble (79%) suggest an important contribution of secondary OC to PM_{2.5} composition at QSS.
466 Indeed, chemical characterization using HR-ToF-AMS showed that WSOM was mainly
467 composed of highly oxygenated organic species. For example, the average AMS spectrum of
468 WSOM was dominated by H_yO₁⁺ (24%), C_xH_yO₁⁺ (22%), C_xH_yO₂⁺ (22%), and C_xH_y⁺ (26%) ions
469 and its O/C and OM/OC ratios were 1.16 and 2.75, respectively. These results suggest that
470 organic species became highly oxidized during long-range transport from lowland to elevated
471 mountain areas and/or locally in the northeastern region of QXP through intense photochemical

472 and aqueous-phase processing in the free troposphere. The estimated SOA on average accounted
473 for 80% (range 62–96%) of the PM_{2.5} mass, which is higher than those reported from Himalayas
474 previously. Given that dry and wet deposition of aerosol particles strongly influences the
475 chemical composition of snow and glaciers, our results may shed lights on the coupling between
476 atmospheric chemistry and cryospheric chemistry in the northern QXP region.

477 **Acknowledgements**

478 This research was supported by grants from the Hundred Talents Program of Chinese Academy
479 of Sciences, the Science Fund for Creative Research Groups of the National Natural Science
480 Foundation of China (NSFC) (41121001), the Scientific Research Foundation of the Key
481 Laboratory of Cryospheric Sciences (SKLCS-ZZ-2013-01-04), and the Changjiang Scholars
482 program of the Chinese Ministry of Education.

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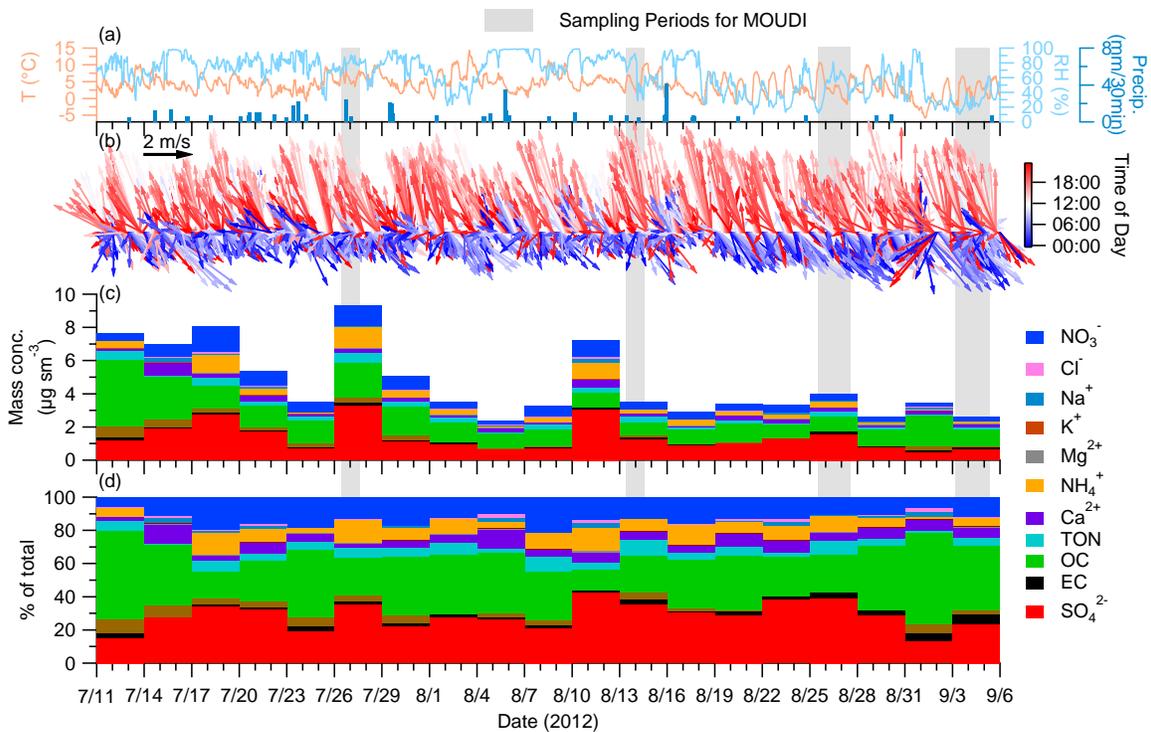
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710 Table 1. Correlation coefficients (Pearson's r) between the water soluble inorganic ions, , organic
 711 carbon (OC), elemental carbon (EC), water soluble organic carbon (WSOC), and total water
 712 soluble nitrogen (TON) concentrations (n=19). Values that indicate a strong correlation (i.e., $r \geq$
 713 0.75) are in bold.

	Na ⁺	NH ₄ ⁺	K ⁺	Mg ²⁺	Ca ²⁺	Cl ⁻	SO ₄ ²⁻	NO ₃ ⁻	EC	WSOC	OC
NH ₄ ⁺	0.28										
K ⁺	0.38	0.87									
Mg ²⁺	0.84	0.51	0.63								
Ca ²⁺	0.80	0.04	0.23	0.79							
Cl ⁻	0.85	0.32	0.41	0.83	0.77						
SO ₄ ²⁻	0.57	0.88	0.86	0.77	0.47	0.58					
NO ₃ ⁻	0.41	0.81	0.77	0.67	0.35	0.42	0.83				
EC	-0.22	0.23	0.22	-0.26	-0.28	-0.22	0.10	0.03			
WSOC	-0.04	0.17	0.12	-0.10	0.10	-0.03	0.16	0.17	0.59		
OC	-0.01	0.16	0.14	-0.05	0.17	-0.01	0.18	0.23	0.56	0.97	
TON	-0.10	0.81	0.61	0.16	-0.16	-0.08	0.65	0.67	0.51	0.45	0.43

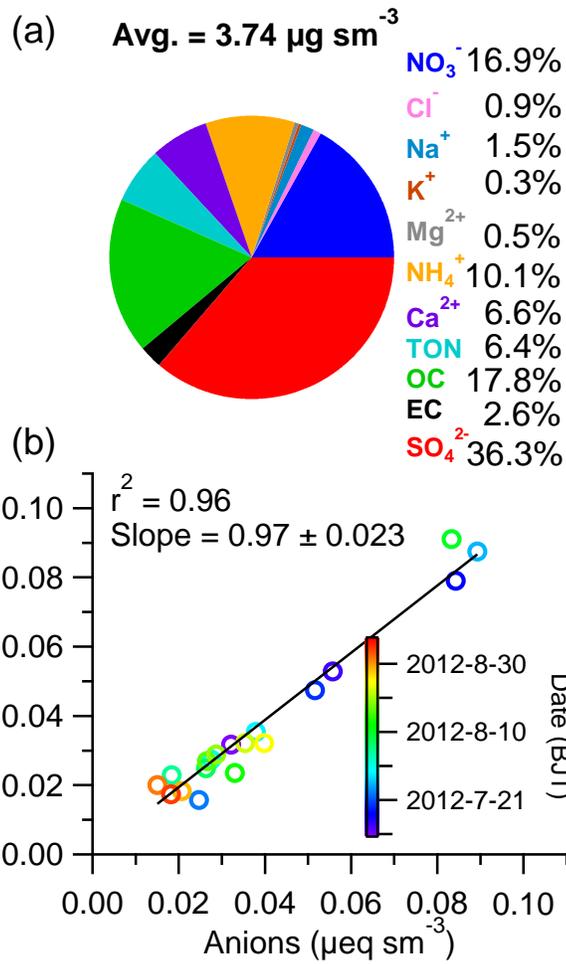
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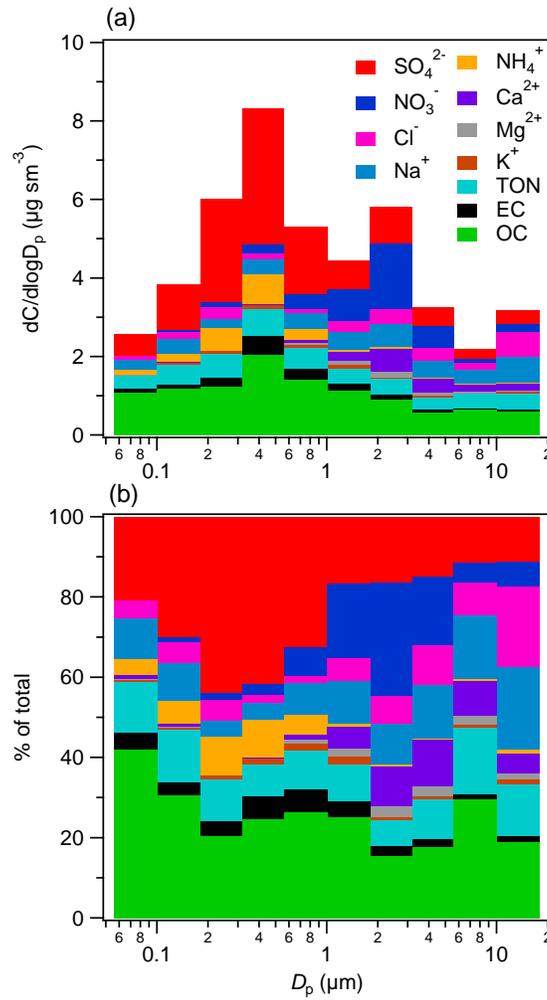
717 Fig. 1. Time series of (a) meteorological data, (b) wind speed and wind direction colored by time
 718 of day (BJT), (c) mass concentrations of water soluble ions, organic carbon (OC), elemental
 719 carbon (EC), and total organic nitrogen (TON) in $\text{PM}_{2.5}$, and (d) percent contributions of various
 720 species to total mass.



721

722 Fig. 2. (a) The average composition of the species analyzed and (b) the charge balance between

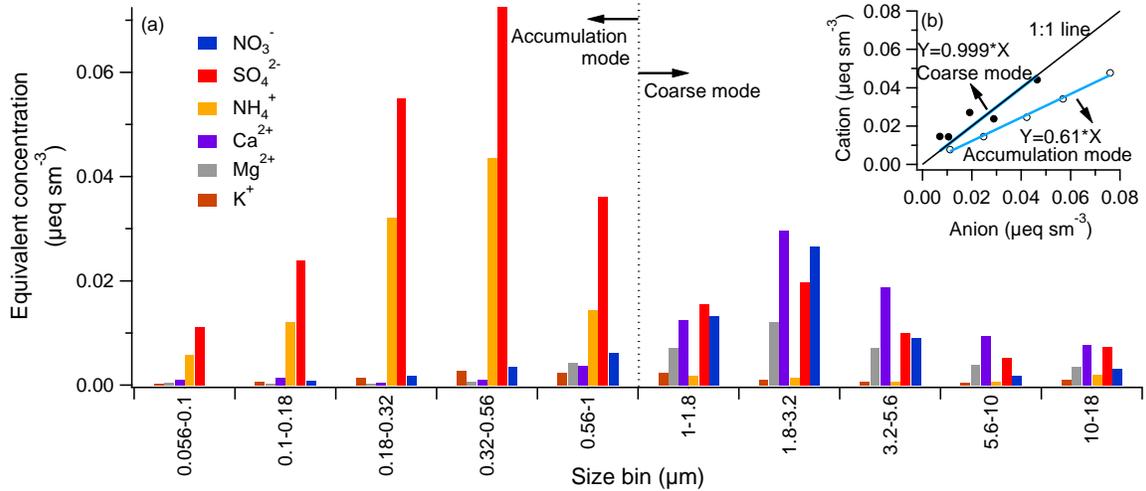
723 the cations ($\text{Na}^+ + \text{NH}_4^+ + \text{K}^+ + \text{Mg}^{2+} + \text{Ca}^{2+}$) and anions ($\text{Cl}^- + \text{SO}_4^{2-} + \text{NO}_3^-$).



724

725 Fig. 3. (a) Average size distributions of the mass concentrations of water soluble inorganic ions
 726 (SO_4^{2-} , NO_3^- , Cl^- , NH_4^+ , Ca^{2+} , Mg^{2+} , and K^+), Total Organic Nitrogen (TON), elemental carbon
 727 (EC), and organic carbon (OC). (b) The fractional contributions of individual species to total
 728 mass in different size bins.

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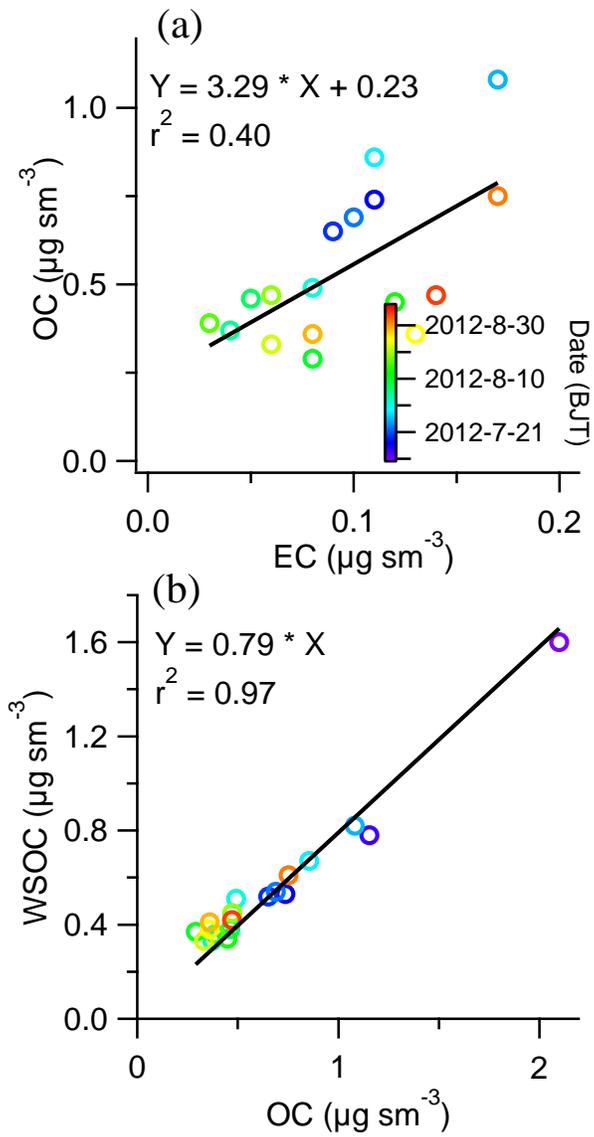
731 Fig. 4. (a) Average equivalent concentrations water soluble inorganic species in individual size

732 bins. The vertical dashed line indicates the boundary between the accumulation mode and the

733 coarse mode. (b) Scatter plot that compares the equivalent concentrations of cations (i.e., NH_4^+ +

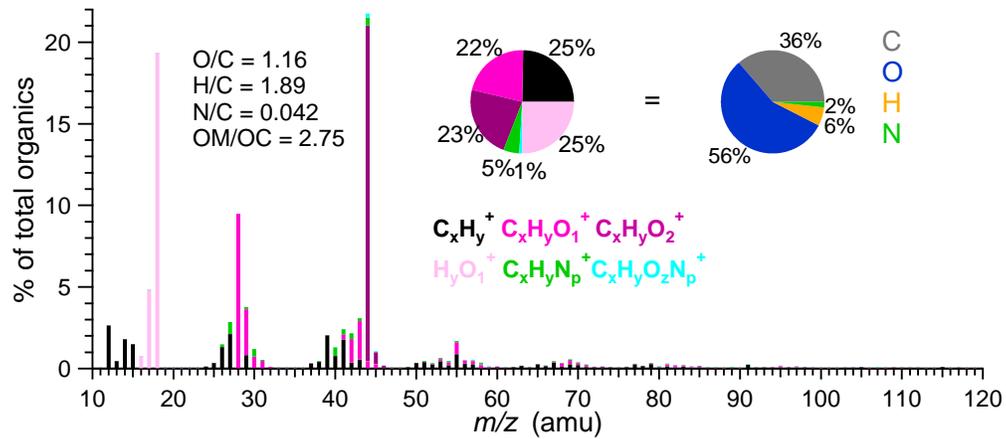
734 Ca^{2+} + Mg^{2+} + K^+) and the equivalent concentrations of anions (i.e., SO_4^{2-} + NO_3^-) in the

735 accumulation mode and coarse mode particles, respectively.



736

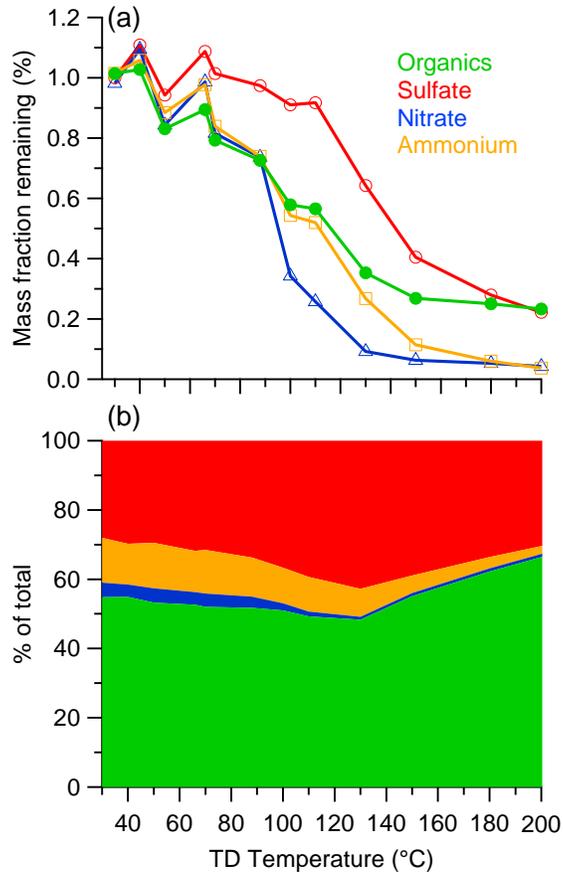
737 Fig. 5. Scatter plots of (a) the organic carbon (OC) concentration against the elemental carbon
 738 (EC) concentration and (b) the water soluble organic carbon concentration (WSOC) against the
 739 OC colored by the sampling date.



740

741 Fig. 6. Average water soluble organic carbon (WSOM) mass spectrum and mass concentration
 742 fraction (pie charts) colored by the contributions of six ion categories and elements (C, O, H, and
 743 N) for PM_{2.5} filter samples.

744



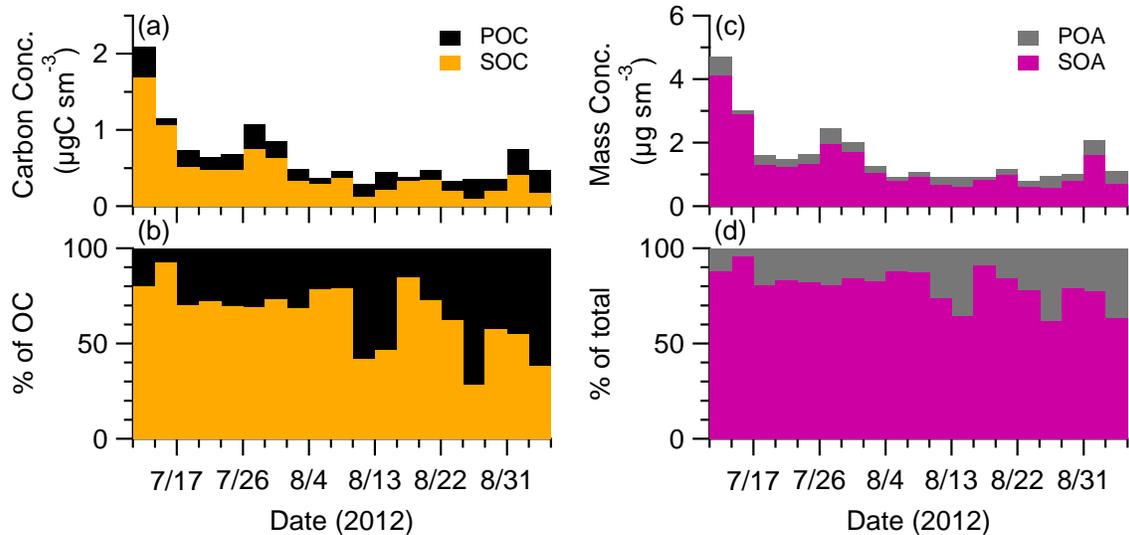
745

746 Fig. 7. Thermal profiles of organic aerosol, sulfate, nitrate, and ammonium in PM_{2.5} from QSS

747 based on (a) the mass fraction of remaining aerosol compared to ambient temperature (b) the

748 composition of non-refractory aerosol materials at given TD temperatures.

749



750

751 Fig. 8. (a) and (b) Estimated concentrations of primary and secondary organic carbon and their
 752 percent contributions to total OC in PM_{2.5} from QSS; (c) and (d) estimated primary and secondary
 753 organic aerosol mass concentrations and percent contributions to total OA mass.

754

755