

**“HTAP\_v2: a mosaic of regional and global emission gridmaps for 2008 and 2010 to study hemispheric transport of air pollution” by G. Janssens-Maenhout et al., ACPD 15, C2857–C2864, 2015**

*The authors are grateful to Referee #1 for the interest and comments on the paper. We tried to improve the paper as requested with more details and data.*

*The modifications in reply to the comments of referee # 1 are highlighted “yellow” and “blue” in the paper.*




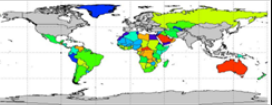
### *Specific Comments*

Page 12871, Line 17: We added to the paper – as suggested - that “The Lamarque et al. (2010) data used a similar methodology of combining country level inventories for most OECD countries with research inventories for Asia and EDGAR for other regions.”

Section 2, general

We added the following overview table specifying the general source and characteristics for the data in each world region.

**Table 1a: Overview of the data sources and their generic characteristics, as used for the different regions in HTAP\_v2.2**

<i>Data source</i>	<b>EMEP - TNO (MACCII)</b>	<b>US EPA + Environ Can</b>	<b>MIX-ASIA (incl REAS2.1)</b>	<b>EDGARv4.3 (prelim.)</b>
<i>type of data source</i>	country inventories + point sources	state inventories + point sources	county inventory for China + country inventories	country inventories
<i>coverage of human activities</i>	all except international shipping/aviation	all except international shipping/aviation	all, except international shipping/aviation and except agricultural waste burning	all inclusive international shipping/aviation
<i>temporal resolution</i>	yearly gridmaps (monthly profiles of EMEP model added)	monthly profiles	monthly gridmaps	monthly profiles (for 3 different latitude bands)
<i>spatial resolution</i>	0.125deg x 0.0625deg, after raster resampling 1/5 x 1/5 and aggregation of 4 x 8 converted into 0.1deg x 0.1deg	0.1deg x 0.1deg and height profiles	0.25deg x 0.25deg, after raster resampling 1/5 x 1/5 and aggregation of 2 x 2 converted into 0.1 deg x 0.1 deg	0.1deg x 0.1deg
<i>substances</i>	CO, NMVOC, NOx, SO2, NH3, PM coarse and fine and BC/OC fractions	CO, NMVOC with speciation profiles, NOx, SO2, NH3, PM10, PM fine, BC and OC	CO, NMVOC, NOx, SO2, NH3, PMcoarse, PM2.5, BC and OC	CO, NMVOC, NOx, SO2, NH3, PM10, PM2.5, BC and OC
<i>in HTAPv2.2 used geo-coverage</i>				

We added the explanation on the re-gridding procedure with special attention to the cells that cover borders between countries at the end of paragraph 2.2.5.

“This replacement took place after the gridmaps were converted into 0.1° x 0.1° using a raster resampling procedure. For EMEP-TNO the resampling implied a 25-fold division to 0.0025°x0.0125° followed by an aggregation of 4x8 gridcells. For the MICS-Asia the resampling needed also a 25th fold division to 0.05°x0.05° followed by an aggregation of

2x2 gridcells. The cells including country borders are split up and allocated to the different countries using the corresponding areal percentage.”

We added – as requested – an additional section 2.3 on the temporal profiles supplementary, in which a comparison has become apparent with Fig. 1c.

Page 12876, section 2.2.1: The authors agree that it is important to detail where lack of data caused not actual but extrapolated data. Even though the 2008 and 2010 are mostly actual data for all data source, unavailability of data lead to few exceptions, which we more explicitly mentioned in the paper:

- “The 2010 data for Canada were missing and as such extrapolated by US EPA based on the 2008 National Emission Inventory of Environment Canada and assuming no trend but using updated point sources (Pouliot et al., 2014).”
- “The EMEP-TNO data were only available for 2006 and 2009. The 2008 data for Europe is based on the EMEP-TNO data for 2009 data and the 2010 data for Europe are based on the same 2009 data but using the trend in EMEP-TNO data between 2006 and 2009.”

The trends between 2008 and 2010 in emissions and in the driving activity data are so small that no significant impact on the implied emission factors is observed.

Page 12877, Line 10: The authors edited the line as suggested. “EMEP-TNO data for country with only partial coverage ...”

Page 12878, Line 6: The EMEP modeling group provided “the monthly profiles, which are with a monthly factor (varying around 1/12) specified for each country and for each sector, with a further compound-specific modulation for the agricultural sector”. This has been added in the text.

Page 12880, Line 6: The paper Balsama et al. (2014) is indeed not describing the EDGARv4.3 gapfilling for HTAP\_v2.2 but analysed the EDGARv4.3 preliminary dataset of EDGAR and its trends. This analysis was useful to identify similarities in the behavior of certain substances and supported the underlying methodology for deriving implied emission factors. The authors agree that it is not here at its correct place (shifted to section 3.6.)

Page 12880, Line 19: We added “EDGAR provides also sector-specific monthly profiles, defined with first-order estimated factors for each of the three different zones: Northern Hemisphere, Equatorial region and Southern Hemisphere (Table S1.2).”

A comparison of the monthly profiles is added in a new section 2.3:

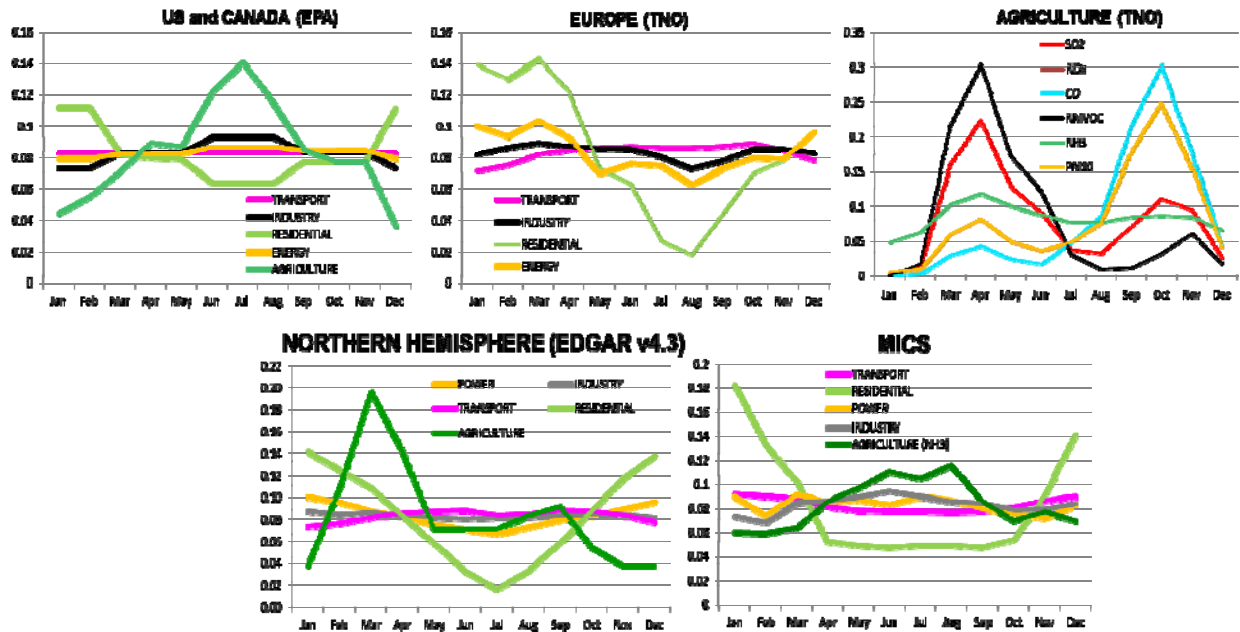
### **2.3 Overview of the temporal profiles used in HTAP\_v2.2**

The modulation of annual emissions over time is necessary in order to provide the modelers emission data consistent with the seasonal pattern and activities. Monthly data were generated for all sectors except for the international shipping and international aviation, which are considered constant over the year. US-EPA, EMEP and EDGAR

provided monthly profiles, but MICS-Asia provided directly and solely monthly emission gridmaps.

Figure 1c summarizes the sector-specific monthly profiles for each of the regional datasets. The temporal profiles are additive and specified with monthly factors modulating around 1/12 for each of the sectors. For the agricultural sector, EMEP provided compound-specific monthly factors, which are characterized by high NMVOC emission in spring and high CO emission in autumn. Agriculture (largely contributing to NH<sub>3</sub> emissions) shows most seasonal variation, which differs also most between the different regions because of region-specific management practices (for e.g. crop cultivation), climate and geographical location and soil composition. The residential sector is characterized by a monthly distribution which is inversely related with the temperature and therefore with the use of heating systems, and in some developed countries with air conditioning (which is boosting emissions in some developed countries during hot summers). The seasonality remains relatively modest in all regions for the transport, industry and energy sectors.

The strongest variation over the year and between regions is observed for the agricultural sector (+215% in the EMEP-TNO profiles but only +45% in the MICS-Asia profiles), followed by the residential sector ([+70%, -75%] in the EMEP-TNO profiles, [+20%, -25%] in the US EPA profiles and [+115%, -40%] in the MICS-Asia profiles).



*Figure 1c – Temporal profiles with relative factors varying around 1/12 and applied on the yearly emissions of the different data sources (US EPA for US and Canada, EMEP-TNO for Europe with compound-specific variation of the agricultural temporal profiles, EDGAR temporal profiles for the Northern hemisphere and MICS profiles for Asia).*

Section 3.1:

Page 12882, Line 20: We reformulated as follows: “The Asian region is still characterized by a relative large contribution of SO<sub>2</sub> from (coal fired) power plants and manufacturing industry.”

Page 12883, Line 3: The authors compared the International shipping emissions with the bottom-up and top-down estimated emissions reported in the "Third IMO GHG Study 2014" in Table 2a. We note that an agreement between the data of HTAP (EDGAR based), and IMO (both top down and bottom up estimates) is obtained for all compounds within 30% except for CO. The CO emission factor showed also in other inventories high uncertainty: the IMO (2009) used a more than twice as high emission factor than the new IMO study (2014). EDGAR shows a 55% and 70% higher estimate for the 2008 and 2010 than the bottom-up values of the IMO (2014) study, which on his turn is 55% respectively 33% higher than the 2008 and 2010 top down estimates of the IMO(2014) study. These observations and the IMO (2014) and IMO (2009) references are taken up in the main text of the paper.

**Table 2a: Comparison of the international shipping emissions: IMO Bottom up (BU) and IMO Top Down (TD) emissions of the IMO(2014) study and the EDGAR emissions of the HTAP\_v2.2 (2015) study.**

kton /yr	BC	CO	NMVOC	NOx	OC	PM10	PM2.5	SO <sub>2</sub>
EDGAR 2008	34	1340	730	13762	458	1376	1376	8348
IMO BU 2008		864	727	20759		1545	1545	11041
IMO TD 2008		553	615	18442		1221	1221	8280
EDGAR 2010	33	1300	720	14000	430	1400	1400	8300
IMO BU 2010		763	593	16708		1332	1332	9895
IMO TD 2010		574	638	19098		1304	1304	9232

### Section 3.2:

Pages 12886 – 12887: The authors consulted several trade databases to provide a quantitative indication of the consumption versus production-based emission inventories for sector 4\_industry. With the World Input-Output Database, Boitier (2012) compared the production-based CO<sub>2</sub> inventory with the consumption-based one and concluded a 14% higher emissions for OECD countries in 2008 (and even 23% for EU27) under the consumption-based approach and a 22% lower emissions for the BRIC countries (20% for China). This range (20% for Germany and 10% for USA whereas -10% for Brasil) matches also with the Global Trade Analysis data of Davis et al (2011). This affects the production-based inventory of air pollutants from the industry sector in a similar way, but probably more than linearly. For the air pollutants there is in addition a considerably lower emission factor of the industry in OECD countries than in developing countries because of an unequal implementation of end-of-pipe measures. Therefore the authors propose the following addition in the paper: “The importance of this consumption- versus production-based approach can be expected in 2008 (and also 2010) to be at least but probably even larger than what Boitier (2012) and Davis et al. (2011) amongst others reported for CO<sub>2</sub>. A consumption-based approach would yield at least 10% higher emissions for industrialised countries whereas 10% lower emissions for developing countries with emerging economy.”

Page 12887: Lines 3-4: Referee #1 points to a substantial difference between the per capita emissions of SO<sub>2</sub> of about 20%. This is indeed worth investigating. We downloaded the EUROSTAT data again and recalculated the per capita emissions. The 11.5 kg SO<sub>2</sub>/cap of Eurostat is valid for 2008 and not for 2010. The 2010 value of EuroSTAT is 8.9 kg SO<sub>2</sub>/cap, which is very close to our estimate of 9.1 kg SO<sub>2</sub>/capita – the 0.2 difference can be due to different years of download (as different reporting years cause small fluctuations) as well as gapfilling by TNO for countries with incomplete time series, but is less than the range we get from using different reporting years. The large decrease of more than 2kg SO<sub>2</sub>/cap between 2008 and 2010 is due to the large emission reduction in the (for some countries coal based) power industry (-26%) and a bit in industrial process industry (-16%).

The authors modified the sentence in the paper accordingly as: "For SO<sub>2</sub> the per capita emission in 2010 for EU-28 of 9.1 kg SO<sub>2</sub>/cap is very close to the reported value of 8.9 kg SO<sub>2</sub>/cap from EuroSTAT (2014) - the 0.2 difference is much less than the 20% higher per capita SO<sub>2</sub> emission in 2008 (11.5 kg SO<sub>2</sub>/cap). EU's 9.1 kg SO<sub>2</sub>/cap is about half the SO<sub>2</sub> per capita for China in 2010 and about one third of the SO<sub>2</sub> per capita for USA."

### Section 3.3:

Page 12888, Line 15 and following: We reformulated the two sentences as follows: "The GDP is subject to heterogeneity (by the different economic activities), to heteroskedasticity (by the time-dependent inflation and currency exchange rates) and to incompleteness (by the not officially reported activities). It is not recommended to use this per unit of GDP emissions indicator for relative small countries with a substantial service sector (e.g. Luxembourg).

### Section 3.4

The authors agreed to provide more details on the calculation of the implied emission factors. In fact, the lack of activity data for all data sources, except for EDGAR induced the following approximation of calculating the denominator of the formula with solely EDGAR activity data for that country and sector while accounting in the numerator the country- and sector-specific emissions as given by the original data source. Moreover the common HTAP sectors aggregated subsectors which are based on activity data with different units. This is mainly the case for the sector 4\_Industry which accounts the combustion emissions (/TJ) and the process emissions (/ton product). With a commonly dominating energy-intensive industry (and a ratio of combustion over process emissions larger than 1), we opted to weigh the industry emissions with the energy needs (and as such partially skewed up the implied emission factor). But also the agricultural sector is skewed up, since we opted to weigh the total emissions of crop cultivation and of livestock with the number of animals elevated (mainly because 85% of the crops is used as animal food). We propose to clarify this in the text by clearly working out the formula for each of the sectors (indicating the use of EDGAR activity for all implied emission factors) and warning for a skewed up implied emission factor. We therefore replaced the single formula with the following:

$$EF_{C,3\_energy}(t, x) = \frac{\sum_{sub\ sector\ j} EM_{C,3\_energy,j}(t, x)}{\sum_{sub\ sector\ j} AD_{C,3\_energy,j}(t)} \Bigg|_{\substack{datasource\ of\ C \\ EDGARv\ 4.3}} [kton / TJ] \quad (1)$$

$$EF_{C,4\_ind.}(t, x) = \frac{\left[ \sum_{comb.\ sub\ sector\ j} EM_{C,4\_ind.,j}(t, x) + \sum_{proc.\ sub\ sector\ k} EM_{C,4\_ind.,k}(t, x) \right]_{datasource\ of\ C}}{\sum_{comb.\ sub\ sector\ j} AD_{C,4\_ind.,j}(t)} \Bigg|_{EDGARv4.3} [kton / TJ] \quad (2)$$

$$EF_{C,5\_ransport}(t, x) = \frac{\sum_{sub\ sector\ j} EM_{C,5\_transport,j}(t, x)}{\sum_{sub\ sector\ j} AD_{C,5\_transport,j}(t)} \Bigg|_{\substack{datasource\ of\ C \\ EDGARv\ 4.3}} [kton / TJ] \quad (3)$$

$$EF_{C,6\_res.}(t, x) = \frac{\left[ \sum_{comb.\ sub\ sector\ j} EM_{C,6\_res.,j}(t, x) + \sum_{waste\ prod.\ sub\ sector\ k} EM_{C,6\_res.,k}(t, x) \right]_{datasource\ of\ C}}{\sum_{comb.\ sub\ sector\ j} AD_{C,6\_res.,j}(t)} \Bigg|_{EDGARv4.3} [kton / TJ] \quad (4)$$

$$EF_{C,8\_agr.}(t, x) = \frac{\left[ \sum_{animal\ sub\ sector\ j} EM_{C,8\_agr.,j}(t, x) + \sum_{crop\ sub\ sector\ k} EM_{C,8\_agr.,k}(t, x) \right]_{datasource\ of\ C}}{\sum_{animal\ sub\ sector\ j} AD_{C,8\_agr.,j}(t)} \Bigg|_{EDGARv4.3} [ton / head] \quad (5)$$

And we added in the main text (and in a footnote in the implied emission factors table): “It should be noted that the implied emission factors of sectors 4\_industry, 6\_residential and 8\_agriculture are slightly skewed up because of an incomplete accounting of activity data which are for these sectors a combination of activities of different nature and as such expressed with different units. The emissions of sector 4\_industry mainly originate from the energy-intensive subsectors and therefore are weighed with the energy needs (in TJ). We omitted the accounting of industrial process emissions, which are calculated per kton product manufactured. In sector 6\_residential the waste is included, although calculated per kton dry or wet waste, which we could not combine with the residential energy consumption in TJ. The emissions of the 8\_agricultural sector are weighed with the number of animals and not with the kton crops cultivated, because the crops serve for 85% as animal food and are therefore considered a justified measure of agricultural activity.”

Results of implied emission factors in figure 4:

The authors recognized that statistics with small numbers are unreliable. Therefore the calculation of robust implied emission factor calculations was only carried out for larger countries with activities in all sectors. As such we left out the following countries:

For CO:

- for the htap\_4\_INDUSTRIY sector: Togo, Eritrea, Congo, Côte d'Ivoire, Kenya, Benin.
- for the htap\_6\_RESIDENTIAL sector: Maldives.
- for the htap\_5\_TRANSPORT sector: North-Korea, Afghanistan, Laos, Tajikistan, Mongolia.

For SO<sub>2</sub>:

- for the htap\_4\_INDUSTRIY sector: Namibia, Laos, Jamaica.

For NO<sub>x</sub>:

- for the htap\_6\_RESIDENTIAL sector: Maldives.
- for the htap\_5\_TRANSPORT sector: Afghanistan, Laos, North-Korea, Tajikistan.

For NMVOC:

- for the htap\_3\_ENERGY sector: Bhutan.
- for the htap\_4\_INDUSTRIY sector: Togo, Eritrea, Côte d'Ivoire, Congo, Cameroon, Kenya, Benin, Aruba, Antigua, Bahamas, Ethiopia, Sudan, Senegal, Equatorial Guinea, Central African Rep., Sri Lanka, Angola, Mozambique, Zambia, Jamaica.
- for the htap\_6\_RESIDENTIAL sector: Am. Samoa, Gum, Maldives, Tonga.
- for the htap\_5\_TRANSPORT sector: Afghanistan, Laos, North-Korea.

For PM<sub>10</sub>:

- for the htap\_4\_INDUSTRIY sector: Togo, Eritrea, Côte d'Ivoir, Congo, Kenya, Benin.
- for the htap\_5\_TRANSPORT sector: Afghanistan.

For PM<sub>2.5</sub>:

- for the htap\_3\_ENERGY sector: Tajikistan, Luxembourg.
- for the htap\_4\_INDUSTRIY sector: Togo and Eritrea.
- for the htap\_5\_TRANSPORT sector: Afghanistan.

For BC:

- for the htap\_3\_ENERGY sector: Nigeria, Malaysia, Belgium, Oman, Finland, Georgia, Vietnam, Canada, Armenia, Tunisia, Jordan, The Netherlands, Trinidad and Tobago, Algeria, Latvia, United Arab Emirates, Brunei, Turkmenistan, Japan, Mozambique, Congo, Qatar, Bahrain, Moldova, Kyrgyzstan, South-Korea, Taiwan, Luxembourg, Bhutan, Tajikistan.
- for the htap\_4\_INDUSTRIY sector: Trinidad and Tobago, Malta.
- for the htap\_5\_TRANSPORT sector: Afghanistan.

For OC:

- for the htap\_3\_ENERGY sector: Tunisia, Jordan, Trinidad and Tobago, Algeria, United Arab Emirates, Brunei, Turkmenistan, Tajikistan, Mozambique, Congo, Qatar, Bahrain, Kyrgyzstan, Taiwan, Myanmar, South-Korea, Vietnam.
- for the htap\_4\_INDUSTRIY sector: Bahrain, Eritrea.
- for the htap\_6\_RESIDENTIAL sector: Greenland, Gibraltar, Faroe Islands, Saint Pierre et Miquelon

- for the htap\_5\_TRANSPORT sector: Afghanistan

For NH3:

- for the htap\_8\_AGRICULTURE sector: Faroe Islands, Tajikistan, Greenland, Falkland Islands, Kyrgyzstan, South-Korea, Brunei, Am. Samoa, Malaysia, Trinidad and Tobago, Bahamas, Saint Pierre et Miquelon, Sri Lanka, Suriname, Réunion, Thailand, Indonesia, Japan, Barbados, Bhutan, Guyana, Costa Rica

The authors propose to mention this list of countries in a footnote on Figure 4.

Page 12889, Line 13-15: We reformulated the text as follows: "It should be noted that emissions, in particularly those reported under country-specific point sources, are allocated to the reporting country solely, also for cells covering country borders. The areal fraction of these cells would incorrectly spread the emissions also to the neighboring country, which yield in the case of e.g. the power emissions for Canada up to 30% increase with the USA emissions along its borders."

Page 12890, Line 13-14: We reformulated the sentence as: "The high SO<sub>2</sub> implied emission factor (from EDGARv4.3) represents the use of lower quality fuels in sea transportation, especially in international waters: 85% of the sea bunker fuel in 2010 consists of residual fuel oil with an emission factor of 1.29 ton SO<sub>2</sub> /TJ."

### Section 3.5

The authors agree that the section should start mentioning where extrapolation in time has been undertaken. This was only done for Canada (US-EPA/Environ Canada) and for Europe (TNO-EMEP). Both regions were affected by the economic crisis of 2008, yielding stagnation and even downwards trends in the following years, mainly in the energy and industry sectors. The latter sectors are constructed for a large share by point source data, which were updated with the real estimates for 2010. As such, the emission gridmaps of 2010 are considered to represent also for Canada and Europe the actual 2010 estimates reasonably well. However every change for each country is not only caused by the change in activity but also and even more by the change in emission factor or implementation of end-of-pipe measures, which were occurring in some developing countries and caused relative large differences.

We propose to add in the beginning of section 3.5 (after the first sentence) the following paragraph: "It should be noted that the data provided for Canada by US-EPA/Environment Canada and for Europe by TNO were actually not representing 2010, but 2008 and 2009, respectively. However updates were undertaken: point source data of 2010 were used and implemented in the gridmaps. Both regions were affected by the economic crisis of 2008, yielding stagnation and even downwards trends in the following years, mainly in the energy and industry sectors. The latter sectors are primarily composed of point sources and as such the gridmaps of 2010 can be considered to represent also for Canada and Europe the actual 2010 situation."

We also reformulated the second last sentence after having (re) verified the increasing coal use: "For the developing countries (calculated with the EDGARv4.3 data and based on the IEA (2013) fuel statistics), the SO<sub>2</sub> emissions of the energy sector slightly increase from 2008 to 2010 because of the increased coal use mainly in South-East Asia



(as also observed by Weng et al., 2012) and the increased use of heavy fuel oil in the Middle East.”

### Section 3.6

By compiling the dataset with different data sources, it became apparent that at the borders of different datasets, large inconsistencies occur. As an example: the TNO-EMEP and MIX-Asia datasets cover respectively the European and the Asian part of Russia, but were showing ground transport emission differences of one order of magnitude. Even though both emission datasets are compiling a bottom-up inventory with similar methodology, different assumptions on emission factors and end-of-pipe measures can explain this. Therefore we opted to have single countries represented by the same dataset. However, each of the datasets used, calculates the emissions at country or county/province level and makes assumptions at this subregional level, which on its turn can lead to inconsistencies at the borders of each country/county/province.

This is clarified in the paper by modifying the introduction of section 3.6 as follows: “Even though the HTAP\_v2.2 data sources are all bottom-up constructed inventories, they differ considerably in e.g. the assumptions taken on the modelling of technology and end-of-pipe measures and use different emission factors and quite different, and lead to inconsistencies at the borders between two adjacent inventories. On their turn the different bottom-up inventories are constructed with sub-regional (country, state, county or province level) activity data and emission factors. As such, inconsistencies can be expected at each country border and the variation of the emissions at cross-border cells gives already a first indication on the region- and sector-specific emission uncertainty.

### Table 3

Even though the HTAP\_v2.2 mosaic of final emission gridmap products does not allow for a full quantification of the error propagation, the authors agree that more information on the uncertainties can be provided in the main text of the paper. All data sources follow a similar methodology and face similar sources of uncertainty, which resemble the situation of the UNFCCC’s CRF dataset of national inventories. Evaluation of their uncertainties by deterministic error propagation calculations or probabilistic Monte Carlo simulations has been addressed by e.g. Jonas et al (2010) (and references in there) and provides input on an uncertainty analysis of a bottom-up inventory per sector and per region. The GHG inventories are tackling with CO<sub>2</sub> the combustion sectors, with CH<sub>4</sub> also the agricultural (livestock and crops) and waste sectors and with N<sub>2</sub>O the industrial processes and agricultural sectors. The analysis for greenhouse gases is only a starting point, because for the air pollutants the emission factors strongly depend on the technology and end-of-pipe measures. Balsama et al. (2014) evaluated common behaviours between several species in the EDGARv4.2 data and observed that SO<sub>2</sub> and NO<sub>x</sub> belong to the same cluster as CO<sub>2</sub> (all strongly combustion related) and NH<sub>3</sub> belongs to the same cluster as N<sub>2</sub>O.

The approach for assessing the CO<sub>2</sub> uncertainty by Andres et al (2012), grouping countries on the basis of their statistical infrastructure was considered appropriate for the HTAP\_v2.2 global dataset as well. Countries with well maintained statistical

infrastructure are the 24 OECD-1990 countries<sup>1</sup> as well as India - using the British statistical accounting system according to Marland et al. (1999). For the other countries, a larger range in uncertainty is present, for which we refer to Gregg et al. (2008) or Tu (2011) and Olivier (2002). For the annual CO<sub>2</sub> inventory, the biofuel is carbon-neutral and not taken up, which leaves out a relative large source of uncertainty. For the N-related emissions, the division in countries could be based on the common agricultural practices of countries for which we refer to Leip et al (2011) and Rufino et al (2014). This explains the setup of Table 3 with qualitative indication of uncertainty ranges (using the terminology low (L), low medium (LM), upper medium (UM) or high (H)) for the different sectors and species.

In addition to the uncertainty of the activities, the quality and representativeness of the controlled emission factors play a crucial role. The standard range of uncertainty already varies according to the EMEP/EEA (2013) Guidebook's Uncertainties Chapter 5 for the absolute annual total of different pollutants between at least 10% for SO<sub>2</sub>, at least 20% for NO<sub>x</sub> and CO, at least 50% for NMVOC, an order of magnitude for NH<sub>3</sub>, and PM<sub>10</sub>, PM<sub>2.5</sub>, BC and OC. These considerations have been taken into account to indicate qualitatively a range for the different uncertainties (L, LM, UM, H).

For the combustion-related sectors is the uncertainty of the partially abated emission factor for air pollutants and in particular for aerosols larger than the uncertainty on the reported activity data, yielding a relative uncertainty that is larger than for CO<sub>2</sub>. In addition non-reported activities, in particular using non-reported biofuel or even rubbish, fall beyond this assessment and would need for an assessment the use of top-down derived emission estimates.

The Authors propose a shortening of the caption of Table 3 and the following addition in the main text of the paper: "Guidance on evaluation of emission uncertainties can be obtained from the evaluations of the national inventories reported to UNFCCC, addressed by e.g. Jonas et al (2010) (and references in there). With the evaluation of common behaviours between species in EDGARv4.2 of Balsama et al (2014) we propose the same approach of CO<sub>2</sub> uncertainty assessment for SO<sub>2</sub> and NO<sub>x</sub> (all driven by combustion-related activities), and the approach of N<sub>2</sub>O for NH<sub>3</sub>. As such Table 3 follows the grouping of countries by Andres et al (2012) and Marland et al (1999), based on their statistical infrastructure. Countries with well maintained statistical infrastructure are the 24 OECD-1990 countries plus India with a British statistical accounting system. For the other countries, a larger range in uncertainty is present, for which we refer to Gregg et al. (2008) or Tu (2011) and Olivier (2002). For the annual CO<sub>2</sub> inventory, the biofuel is carbon-neutral and not taken up in the national inventories. However, for the air pollutants it is an additional large source of uncertainty, which is often not officially reported and as such missing. For the N-related emissions, the division in countries could be based on common agricultural practices (Leip et al, 2011 and Rufino et al, 2014). In addition to the uncertainty of the activities, the quality and representativeness of the controlled emission factors play a crucial role. The standard range of uncertainty already varies according to the EMEP/EEA (2013) Guidebook's Uncertainties Chapter 5 for the absolute annual total of different pollutants between at least 10% for SO<sub>2</sub>, at least 20%

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<sup>1</sup> Australia, Austria, Belgium, Canada, Switzerland, Germany, Denmark, Spain, Finland, France, United Kingdom, Greece, Ireland, Iceland, Italy, Japan, Luxembourg, The Netherlands, Norway, New Zealand, Portugal, Sweden, Turkey, and the United States

for NO<sub>x</sub> and CO, at least 50% for NMVOC, an order of magnitude for NH<sub>3</sub>, and PM<sub>10</sub>, PM<sub>2.5</sub>, BC and OC. These considerations have been taken into account to indicate qualitatively a range for the different uncertainties (using the terminology low (L), low medium (LM), upper medium (UM) or high (H)) for the different sectors and species.”

Page 12891, Line 14: The HTAP modeling community is not only using the HTAP\_v2.2 emission inventory but will also run the emission scenarios of ECLIPSEv5, which starts in 2010. The starting emission inventory (or base year inventory) of the ECLIPSEv5 scenarios is the important point of reference for all projections. Here we compare the ECLIPSEv5 emission inventory for 2010 with the HTAP\_v2.2 2010 data, in order to evaluate how close the reference point is to the “officially accepted” regional inventories. We agree that the GAINS dataset can not be considered an external independent source of verification. The huge amount of information in GAINS on emission factors and reductions for certain technologies has also been flowing in the TNO-EMEP, MIX-Asia and EDGARv4.3 datasets. We added this to the paper.

Page 12892, Line 15: If for the same region two different data sources provide emission gridmaps for PM<sub>2.5</sub> and PM<sub>10</sub>, it is not guaranteed that for each cell the flux of PM<sub>2.5</sub> emissions is smaller than the flux of PM<sub>10</sub> emissions and with non-compliance of the equation  $mass\_PM_{2.5} \leq mass_{PM_{10}}$ . Another spatial proxy data set with and without point source can create a huge difference. The same applies for different data sources of BC and OC compared to PM<sub>2.5</sub>, for which  $BC+OC \leq PM_{2.5}$  should hold. We reformulated this in the paper as follows: Another type of inconsistency in mass balance at grid cell level occurs when for the same region the data sources of the gridmaps for PM<sub>10</sub> and PM<sub>2.5</sub> or for PM<sub>2.5</sub> and BC/OC are different. Already the application of different spatial proxy datasets (e.g. with and without point sources) results in an inconsistent allocation of multi-pollutant sources to different grid cells.

Page 12892, Line 24 – Page 12893, Line 3 has been rewritten as follows:

“Even though this mosaic inventory can not present the same consistency as one global bottom-up inventory, its extensive evaluation and use helped improving its quality. The evaluation was undertaken in particular in discussion with TNO and with US EPA to identify missing sources or misallocation of point sources. In addition the use of the dataset by global and regional climate and air quality modelers and the modelers’ feedback (personal communications with L. Emmons of 5 November 2013 and D. Henze of 19 November 2013) were most useful and are further encouraged.”

Page 12893, Line 6: The authors refer with the annotation “regionally accepted as reference” to the buy-in of each region for accepting this dataset as reference. The emission inventory for their region has been provided by their own regional inventory compilers. Therefore the dataset has a more official status than any global emission inventory that is not composed of regional inventories.

We propose to modify the sentence as follows: “This paper describes the HTAP global air pollutant reference emission inventory for 2010, which is composed of latest available data from regional inventory compilers.”

Page 12893, Line 15: Indeed the sector-specific emissions are calculated according to the international standards such as IPCC/EMEP guidelines but for the activity data we needed to refer to consistent international statistics. The sentence is modified as follows: “Even though the HTAP\_v2.2 dataset is not a self-consistent bottom-up database with activity data of consistent international statistics, with harmonized emission factors, and with global sets of spatial proxy data, it provides a unique set of emission gridmaps with global coverage and high spatial resolution, including important point sources.”

#### Figure 2

The captions for figure 2 are shortened with one single caption with: “Sector-specific breakdown of regional emission totals (Tg) for 2010: SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOC, PM<sub>10</sub>, PM<sub>2.5</sub>, BC, OC and NH<sub>3</sub>”. The species name is placed within each sub-figure as suggested on top of the center of the Antarctica region.

The sectors in Table 1b and used further in the main text of the paper (incl. the figures) are the same. The authors opted to use abbreviations which contain the names of the sectors as they are used in the figures: 1\_AIR , 2\_SHIPS, 3\_ENERGY, 4\_INDUSTRY, 5\_TRANSPORT, 6\_RESIDENTIAL and 8\_AGRICULTURE. Table 1b and the main text of the paper has been modified accordingly.

**“HTAP\_v2: a mosaic of regional and global emission gridmaps for 2008 and 2010 to study hemispheric transport of air pollution” by G. Janssens-Maenhout et al., ACPD 15, C2857–C2864, 2015**

*The authors are grateful to Referee #2 for his interest and comments on the paper. We tried to improve the paper as requested with more details and data.*

*The modifications in reply to the comments of referee # 2 are highlighted “green” and “blue” in the paper.*

### *Specific Comments*

Page 12870, Lines 5-7 and Page 12890, Line 1-2. Referee # correctly indicates an incomplete wording in the abstract that leads to confusion. The authors meant that the energy and industry emissions of acidifying **gaseous** air pollutants differ strongly between industrialised and developed countries, whereas no such difference is observed in the acidifying gaseous air pollutant emissions from the residential sector. The authors agree that it is needed to mention explicitly SO<sub>2</sub> and NO<sub>x</sub> to avoid confusion. Moreover the authors are happy to take up the suggestion of referee #2 to complete the abstract with the findings on the aerosols, which show almost the opposite effect. Large differences are not present in the energy (and industry) sector, but they are present in the residential sector.

The authors suggest to add in the abstract:

“An analysis of country-specific implied emission factors shows a large difference between industrialised countries and developing countries for acidifying gaseous air pollutant emissions (SO<sub>2</sub> and NO<sub>x</sub>) from the energy and industry sectors. This is not observed for the particulate matter emissions, which show large differences between countries from the residential sector instead.”

Page 12879, Line 13: Raster-resample procedure and country totals in Table S1.1  
At the time of compilation of the HTAP\_v2.2, only monthly emission gridmaps, as the result, were delivered to the EDGAR team. We then applied the EDGAR table which allocates each grid cell to the country or countries it belongs to. Cells containing borders of countries allocate the area to the different countries with a percentage that reflects the areal coverage in the cell. This table works like a complete set of country masks. With the country masks, the EDGAR team derived also the country totals for the countries, which include a given error because of border issues. However, meanwhile the MICS-Asia team was so kind to deliver the original country totals, which have been compared in Table S1.3. This revealed that applying country masks to obtain country totals (as also done by modellers and e.g. in the ECCAD system) is only valid if the total emission value is larger than 0.2% of each of the country totals of the neighbouring countries. Otherwise a derived country-specific sector total that is 50% larger than the bottom-up one is observed, mainly in the energy sector with many point sources which are typically located on waterways or coastal areas, and end up in cross border cells. The latter caused derived sector totals for Kyrgyzstan, Tajikistan, Afghanistan, Laos, Myanmar,

Bangladesh, which deviated with one order of magnitude from the bottom-up totals. However, China shows good agreement between derived totals and bottom-up totals: within 5% for the energy sector, within 1.4% for Industry and Residential sectors, within 2.6% for the Transport and within 0.4% for the Agriculture sector. India idem: below 3% difference, with the exception of SO<sub>2</sub>, which differs 6% respectively 14% for the SO<sub>2</sub> from the energy and transport sectors. All derived emissions are agreeing within 7% for Indonesia and within 12.5% for Thailand. Japan and South Korea show a bit more deviation of maximum 16.0% and 17.3%.

In Table S1.1: we replaced for the MICS-Asia region the previous values with the country totals received from the MICS-Asia team, to make the dataset more consistent. Now all the country totals are real bottom-up country totals and no longer with one part derived using a mask on a gridmap.

We added a more detailed explanation on the raster-resample procedure on Page 12879, Line 13: “As such, countries within the broad area, spanning from 89.875°N to 20.125°S in latitude and from 40.125°E to 179.875°E in longitude were inserted in the 0.1° x 0.1° emission gridmaps after converting the 0.25° x 0.25° with a raster resample procedure – dividing the cells in 5x5 and then aggregating the 0.05°x0.05° cells with 2x2.

The expertise in comparing the derived totals from the gridmaps with the real country bottom-up totals has been added in section 3.6: “It should be noted that derivation of country totals from the 0.1°x0.1° emission gridmaps (as e.g. done in the ECCAD system) is only valid if the country-specific total is larger than 0.2% of each of the totals of the neighbouring countries. Otherwise the derived country-specific sector total can be 50% larger than the bottom-up one, mainly in the energy sector with many point sources which are typically located on waterways or coastal areas and as such in cross border cells. Table S1.3 illustrates the deviations of derived country-specific sector totals to the bottom-up ones for the Asian region. The latter caused derived sector totals for Kyrgyzstan, Tajikistan, Afghanistan, Laos, Myanmar, Bangladesh, which deviated with one order of magnitude from the bottom-up totals. However, the relative small differences for China (≤5%), India (≤3% for all except for SO<sub>2</sub> from energy where it is 14%), Indonesia (≤7%) and Thailand (≤12.5%), Japan (≤16.0%) and South Korea (≤17.3%) show a good agreement for the top 6 Asian emitters.” (Table S1.3 is added in the supplementary.)

Table 1: names of the sectors and consistent use throughout the paper.

Referee # 2 correctly indicates an inconsistency in the naming of the sectors throughout the paper, which needs correction. The sectors in Table 1 and used further in the paper and in the figures are the same. The authors opted to shorten the name of the sectors in Table 1, and use the same names as used in the figures: 1\_AIR , 2\_SHIPS, 3\_ENERGY, 4\_INDUSTRY, 5\_TRANSPORT, 6\_RESIDENTIAL and 8\_AGRICULTURE.

Page 12887, Lines 3-4: Referee #2 points to a substantial difference between the per capita emission of SO<sub>2</sub> of about 20%. This is indeed worth investigating. We downloaded the EUROSTAT data again and recalculated the per capita emissions. The 11.5 kg SO<sub>2</sub>/cap of Eurostat is valid for 2008 and not for 2010. The 2010 value of EuroSTAT is 8.9 kg SO<sub>2</sub>/cap, which is very close to our estimate of 9.1 kg SO<sub>2</sub>/capita – the 0.2 difference can be due to different years of download (as different reporting years

cause small fluctuations) as well as gapfilling by TNO for countries with incomplete timeseries but is less than the range we get from using different reporting years. The large decrease of more than 2kg SO<sub>2</sub>/cap between 2008 and 2010 is due to the large emission reduction in the (for some countries coal based) power industry (-26%) and a bit in industrial process industry (-16%).

The authors modified the sentence in the paper accordingly as: "For SO<sub>2</sub> the per capita emission in 2010 for EU-28 of 9.1 kg SO<sub>2</sub>/cap is very close to the reported value of 8.9 kg SO<sub>2</sub>/cap from EuroSTAT (2014) - the 0.2 difference is much less than the 20% higher per capita SO<sub>2</sub> emission in 2008 (11.5 kg SO<sub>2</sub>/cap). EU's 9.1 kg SO<sub>2</sub>/cap is about half the SO<sub>2</sub> per capita for China in 2010 and about one third of the SO<sub>2</sub> per capita for USA."

Table 2b: ranking of USA, Germany and China

The list of USA, Germany, China was based on the selection of the top CO<sub>2</sub> emitters in 2010 of each of the three continents in the northern hemisphere. The ranking is a combination of the per capita activity and the level of implementation of end-of-pipe measurement technology. The activity level is best reflected by the per capita CO<sub>2</sub> emissions, which is highest for USA explaining the high air pollutant emissions per capita. However China with lowest CO<sub>2</sub> per capita is not having the lowest per capita air pollutant emissions, because of the level of technology and end-of-pipe implementation. To measure the latter we apply a kind of surrogate variable: the Human Development Indicator (2010) from UNDP(2015). This shows that Germany is more advanced and therefore having lower emissions per capita than China. In order to provide a more complete picture, the authors agreed to include the top 6 world CO<sub>2</sub> emitters: China, USA, India, Russia, Germany and Japan.

For the paper we propose an extension of Table 2b with the CO<sub>2</sub>/cap and the HDI. Moreover we added the countries India, Russia and Japan.

Substance	USA	Germany	China	India	Russia	Japan
ton CO <sub>2</sub> (long cycle C) /yr/cap	17.6	9.9	6.4	1.5	11.9	9.7
HDI	0.91	0.9	0.7	0.57	0.77	0.88
kg SO <sub>x</sub> /yr/cap	32.6	5.2	21	8.0	31.9	5.2
kg NO <sub>x</sub> /yr/cap	43.6	14.2	20.8	7.9	25.1	14.5
kg VOC/yr/cap	43.1	11.9	16.9	14.0	26.9	9.1
kg CO/yr/cap	148.3	35.6	125.6	56.0	52.8	33.1
kg NH <sub>3</sub> /yr/cap	11.6	7.3	6.7	8.2	6.3	3.7
kg PM <sub>2.5</sub> /yr/cap	5.25	1.08	8.93	5.19	2.18	0.62
kg BC/yr/cap	0.95	0.20	1.29	0.85	0.29	0.16

In the main text of the paper, the findings of Table 2b are summarized in section 3.2 as follows:

The level of per capita air pollution results from a combination of the per capita activity and the level of implementation of end-of-pipe measurement technology. The activity level can be reflected by the per capita CO<sub>2</sub> emissions, which is highest for USA explaining the high air pollutant emissions per capita. However not India with lowest CO<sub>2</sub> per capita, but Japan and Germany are having the lowest per capita air pollutant emissions, because of the level of technology and end-of-pipe implementation. To measure the latter we apply a kind of surrogate variable: the Human Development Indicator (2010) from UNDP(2015). This shows that Germany and Japan are more advanced and have therefore lower emissions per capita for all air pollutants (except NH<sub>3</sub> for Germany) and for the PM. We observe that the PM emissions per capita of Japan (0.16 kgPM<sub>2.5</sub>/yr/cap) are only 60% of those of Germany and Germany's one are about one fifth of the per capita emissions of the USA, which are on their turn only 60% of the per capita PM<sub>2.5</sub> for China. Table S3 indicates that developing countries, in particular those with emerging economies but not yet fully penetrated clean technologies and end-of-pipe measures, have enhanced PM per capita emissions (China – 8.2 kgPM<sub>2.5</sub>/yr/cap, India – 5.2 kgPM<sub>2.5</sub>/yr/cap, Brasil – 3.1 kgPM<sub>2.5</sub>/yr/cap). Russia has relatively high per capita PM emissions (2.2 kg PM<sub>2.5</sub>/yr/cap because of fossil fuel production and consumption in the power sector, but much less than Canada (7.4 kg PM<sub>2.5</sub>/yr/cap), a much less populated country but with important fossil fuel production industry for export. Both countries, with important contribution in the Arctic region, show relatively high NMVOC and SO<sub>2</sub> emissions (50.9 kg VOC/yr/cap and 48.7 kg SO<sub>2</sub>/yr/cap for Canada respectively 26.8 kg NMVOC/yr/cap and 31.9 kg SO<sub>2</sub>/yr/cap for Russia) due to their significant inland waterway transport using heavy residual fuel oil or diesel.

Air pollutant emissions per unit of GDP: extra Table 2c

Substance	USA	Germany	China	India	Russia	Japan
kg CO <sub>2</sub> (long cycle C) /yr/USD	339.71	287.79	240.88	136.6	644.58	267.08
GDP/cap	49307	39668	9230	4638	21663	34561
g SO <sub>x</sub> /yr/USD	0.668	0.132	2.310	1.719	1.482	0.150
g NO <sub>x</sub> /yr/USD	0.892	0.363	2.295	1.714	1.166	0.419
g VOC/yr/USD	0.882	0.305	1.863	3.013	1.249	0.263
g CO/yr/USD	3.036	0.910	13.830	12.069	2.449	0.957
g NH <sub>3</sub> /yr/USDP	0.236	0.187	0.735	1.770	0.291	0.108
g PM <sub>2.5</sub> /yr/USD	0.108	0.028	0.984	1.119	0.101	0.018
g BC/yr/USD	0.019	0.005	0.143	0.183	0.013	0.004

India's carbonaceous particulate matter emissions per unit of GDP are indeed higher than those of China, because of the per capita relative low GDP per capita and the use of less clean technologies. Those countries with relative high GDP per capita and



implementation of clean technology that score lowest are Germany and Japan with only 0.005 g BC per invested unit of GDP (USD PPP corrected in 2010).

This Table 2c and the following explanation are added to the paper: “In analogy with Table 2b, Table 2c provides for the world top 6 CO<sub>2</sub> emitters a comparison of the air pollutants per unit of GDP, which are linked to the country’s economic activity (in GDP per capita) and CO<sub>2</sub> per unit of GDP (measuring the energy intensive industry). It is directly apparent that again Germany and Japan are having high economic activity, with still important energy intensive industry but low air pollutant emissions per unit of GDP because of the investment in clean technology. On the other side, India has still much lower economic activity but nevertheless a much higher particulate matter emission per unit of GDP.”

More specific comments with the request for supporting information

Page 12882, Line 24-26: Based on the bottom-up inventory of MICS-Asia per sector and country, we observe that although India’s SO<sub>2</sub> emissions are only 32% of the Chinese one, the energy sector emits 67% of what the complete energy sector in China emits in SO<sub>2</sub>. We modified the text as follows: “High annual SO<sub>2</sub> emissions are also observed for India, to which the energy sector contributes 59% and the energy-intensive manufacturing industry (iron & steel) 32%, both using also coking and bituminous coal according to IEA (2013).”

Page 12883, Line 8-10: Based on the data in Table S1.1, we observe a relative high contribution of the residential + industry sector for the total NO<sub>x</sub> in Canada, but also The Netherlands and Norway. All are according to IEA(2013) characterized by a high percentage of natural gas in their fuel consumption for these sectors.

We reformulated the paragraph on NO<sub>x</sub> with some more quantitative information as follows: “In Central and South America major emissions are attributed to the transportation sector and just to a minor extent to the energy sector (e.g. in Mexico 65% of the NO<sub>x</sub> emissions originate from road transport). Those industrialised countries with a large share of natural gas as fuel for heating houses and commercial centres and for industry (such as Canada, the Netherlands, Norway) show relatively high emissions of NO<sub>x</sub>: the share of the residential and industry NO<sub>x</sub> emissions is around 30% of the total NO<sub>x</sub>, whereas in USA this is only 20%.”

Page 12884, Line 7-9: Based on the data in Table S1.1 we addressed the observations on NMVOC with a quantification of the share. We also used underlying fuel statistics from IEA(2013), in particular to address the biofuel use and the charcoal production.

For the latter we summarized the data of 2008 and 2010 production for the top 3 charcoal producers in the table underneath, but which we feel that these fuel statistics fall outside the scope of the paper. In the Table underneath referee #2 can see that Brasil, Thailand and Kenya are (with distance from other producers) the world top 3. REAS2.1 however is not modeling charcoal production and therefore this emission source is missing for Thailand in HTAP\_v2.2. In addition it is interesting that Brasil reduced considerably (to 46%) its charcoal production activity, whereas the other two countries kept a constant production.

The paragraph on NMVOC has been modified with a more balanced and quantitative description as follows: “In the Middle East NMVOC sources include oil production: the industry sector in Saudi-Arabia contributes 75% to its total NMVOC emissions. In China, particular high emissions are originating from industry (62%) and residential (27%), the latter also associated with the high use of solvents in paints. In Brazil particular high use of biogasoline is present resulting in a 52% NMVOC contribution of the transport sector. Also the production of charcoal is emitting strongly NMVOC and the world top 3 emitters (IEA, 2013) are Brasil, Thailand<sup>2</sup> and Kenya, which explains that their industry sector is contributing to the NMVOC total with respectively 35%, 37% and 80% in 2010.”

Table: TJ charcoal produced by the countries, which contribute more than 1% to the world total charcoal production (IEA, 2013)

TJ charcoal produced	Y_2008	Y_2010	share 2008	share 2010
Brasil	267549	122671.5	22.8%	11.8%
Thailand	137861	133779	11.7%	12.9%
Kenya	91168	96003.5	7.8%	9.2%
Sudan	53116	55135.5	4.5%	5.3%
South Africa	50204	51312	4.3%	4.9%
Tanzania	47340	48324.5	4.0%	4.6%
Ethiopia	35358	37360	3.0%	3.6%
Cote d'Ivoir	33664	35820.5	2.9%	3.4%
Nigeria	33264	34804	2.8%	3.3%
Angola	32894	34604	2.8%	3.3%
Zambia	31160	32708	2.7%	3.1%
Philippines	29221	29785.5	2.5%	2.9%
Ghana	21468	22330.5	1.8%	2.1%
Congo	20975	22391.5	1.8%	2.2%
Paraguay	20945	10659.5	1.8%	1.0%
Indonesia	20451	19911.5	1.7%	1.9%
Vietnam	17648	18079.5	1.5%	1.7%
Togo	16724	17525.5	1.4%	1.7%
Malaysia	15585	16139	1.3%	1.6%
Columbia	14815	14815	1.3%	1.4%
Senegal	14502	17276	1.2%	1.7%
Mozambique	13298	13899.5	1.1%	1.3%
Dominican Rep.	12104	12104	1.0%	1.2%

Page 12885, Line 11-16. We quantified the paragraph on text further as follows: “A decreasing trend from 2008 to 2010 is observed for Brazil due to decreases in emissions from charcoal production (with 23% share in the world production in 2008 and 12% in 2010, according to IEA, 2013). Emissions from charcoal production are also important for some African countries (Kenya, Sudan, South Africa, Tanzania, Ethiopia), with

<sup>2</sup> The charcoal production emissions for Thailand are missing because REAS2.1 is not accounting for this source.

country-specific shares in world production varying between 1.3% and 12.910% according to IEA (2013).”

Page 12885, Line 23-24: Indeed the coarse sector breakdown in fig. 2g does only show that the transport sector is mostly contributing. However the BC (controlled) emission factor is two orders of magnitude larger for diesel than for petrol (see Table underneath). Therefore the authors were confident to mention that these BC emissions are caused by the diesel transport.

Table: Emission factors for petrol and diesel vehicles (light duty, passenger car) with different types of end-of-pipe measures (as present in a European fleet).

fuel	vehicle	EOP			2008-2010	fuel	vehicle	EOP			2008-2010	EFpetrol/ EFdiesel
diesel	light duty	EU1	BC	kg/TJ	10.59147	petrol	light duty	EU1	BC	kg/TJ	0.05745	0.5%
diesel	light duty	EU2	BC	kg/TJ	9.35451	petrol	light duty	EU2	BC	kg/TJ	0.04532	0.5%
diesel	light duty	EU3	BC	kg/TJ	8.5041	petrol	light duty	EU3	BC	kg/TJ	0.03732	0.4%
diesel	light duty	EU4	BC	kg/TJ	4.25205	petrol	light duty	EU4	BC	kg/TJ	0.03732	0.9%
diesel	light duty	EU5	BC	kg/TJ	4.25205	petrol	light duty	EU5	BC	kg/TJ	0.03732	0.9%
diesel	light duty	no control	BC	kg/TJ	38.655	petrol	light duty	no control	BC	kg/TJ	0.07774	0.2%
diesel	light duty	pre EU	BC	kg/TJ	38.655	petrol	light duty	pre EU	BC	kg/TJ	0.07774	0.2%
diesel	passenger cars	EU1	BC	kg/TJ	14.813288	petrol	passenger cars	EU1	BC	kg/TJ	0.12405	0.8%
diesel	passenger cars	EU2	BC	kg/TJ	10.097136	petrol	passenger cars	EU2	BC	kg/TJ	0.08373	0.8%
diesel	passenger cars	EU3	BC	kg/TJ	6.95996	petrol	passenger cars	EU3	BC	kg/TJ	0.05681	0.8%
diesel	passenger cars	EU4	BC	kg/TJ	6.357456	petrol	passenger cars	EU4	BC	kg/TJ	0.05681	0.9%
diesel	passenger cars	EU5	BC	kg/TJ	6.357456	petrol	passenger cars	EU5	BC	kg/TJ	0.05681	0.9%
diesel	passenger cars	EU6	BC	kg/TJ	6.357456	petrol	passenger cars	EU6	BC	kg/TJ	0.05681	0.9%
diesel	passenger cars	no control	BC	kg/TJ	41.552	petrol	passenger cars	no control	BC	kg/TJ	0.12405	0.3%
diesel	passenger cars	pre EU	BC	kg/TJ	41.552	petrol	passenger cars	pre EU	BC	kg/TJ	0.12405	0.3%

We modified the sentence as follows: “Fig.2g shows that the largest contributing sector for BC in North America, Europe and the Middle East is road transport, which should be mainly from diesel vehicles given the much higher BC emission factor for diesel than for petrol.”

Page 12886, Line 1-3: We quantified the shares of BC emissions of the industry and residential sector in China and India, and compared these with the shares in USA and Germany. We consulted the IEA (2013) fuel statistics and understood that the (bituminous) coal use in power plants, coke ovens, non-metallic minerals (cement) and even in the residential sector are causing this for China and the use of coal but also of solid biomass is causing the same high share in India.

Page 12886, Line 4-7. We quantified the shares of BC emissions of the residential sector in China and Russia. The emissions for Russia are calculated with EDGARv4.3 and all details are known. Therefore we comment the contribution of the different fuels in the residential sector in more detail, as taken from the EDGARv4.3 BC emissions of the Russian residential sector in the Table below. We consider it out of balance to include this detailed table in the paper but we updated the paragraph as follows: “The residential sector in China accounts for more than half (52%) of its BC total. Russia shows a similar high share of the residential sector (46%) to its total BC. Most important sources calculated in EDGARv4.3 for heating buildings in Russia include bituminous coal (57%), solid biomass (30%), lignite (6%) and industrial waste (3%) burning in the residential

sector (for domestic housing as well as commercial services) (EC-JRC/PBL, 2011 and IEA, 2013).”

Table: Fuel-specific breakdown of the BC emissions from the residential sector of Russia in 2010 from EDGARv4.3 (EC-JRC/PBL, 2011)

type of building	fuel type	kton BC in 2010	
Farms	bituminous coal	1.14E-01	0.60%
Farms	diesel	6.66E-02	0.35%
Farms	industrial waste	8.57E-02	0.45%
Farms	lignite	1.08E-01	0.57%
Farms	LPG	2.07E-03	0.01%
Farms	natural gas	3.25E-03	0.02%
Farms	peat	2.86E-03	0.02%
Farms	solid biomass	8.80E-01	4.62%
Commercial services	BKB	8.79E-03	0.05%
Commercial services	bituminous coal	6.55E+00	34.38%
Commercial services	diesel	2.59E-02	0.14%
Commercial services	residual fuel oil	8.15E-03	0.04%
Commercial services	industrial waste	4.18E-01	2.19%
Commercial services	lignite	1.15E+00	6.04%
Commercial services	LPG	4.95E-03	0.03%
Commercial services	natural gas	1.01E-02	0.05%
Commercial services	Oven coke	1.73E-01	0.91%
Commercial services	solid biomass	2.27E+00	11.90%
Fisheries	bituminous coal	4.99E-03	0.03%
Fisheries	diesel	7.63E-03	0.04%
Fisheries	heavy residual fuel oil	1.57E-02	0.08%
Domestic housing	bituminous coal	4.23E+00	22.19%
Domestic housing	diesel	5.51E-03	0.03%
Domestic housing	heavy residual fuel oil	4.72E-03	0.02%
Domestic housing	lignite	1.65E-01	0.87%
Domestic housing	LPG	8.43E-02	0.44%
Domestic housing	natural gas	1.64E-01	0.86%
Domestic housing	peat	6.31E-03	0.03%
Domestic housing	solid biomass	2.48E+00	13.04%
	<b>Totals</b>	<b>1.90E+01</b>	

Page 12891, Line 2-3: The authors agree that a decrease in PM emissions from 2008 to 2010 in developing countries results from a combination of reduced activity and penetration of abatement technology. Only for the developing countries calculated with the EDGARv4.3 emissions database these two causes can be decoupled. Largest reductions over these two years were observed for Brasil, Mexico, Columbia, Venezuela, Kazakhstan, Cuba etc. in the industry (fuel transformation), energy and road transport sector. We added a table with the (sub)sectors contributing mostly to the reduction, demonstrating the relative reduction in activity and in (controlled) emission factor, as

modeled in EDGARv4.3. For the paper, we consider it most appropriate to mention only the two largest countries Kazakhstan and Brasil.

Table: Reductions in activity and in emission factor for some developing countries between 2008 and 2010 from EDGARv4.3 (EC-JRC/PBL,2011).

developing country	activity	(2010-2008)/2008	
		activity reduction	emission factor reduction
Brasil	charcoal production	-54.1%	-2.4%
Mexico	energy (bit. coal)	-13.9%	-2.5%
Columbia	energy (bit.coal)	6.3%	-9.0%
	charcoal production	0.0%	-1.0%
Kazakhstan	energy for coal mining	-44.9%	-3.9%
	power with lignite/coal	-11.1%	-30.9%
Venezuela	road transport	2.3%	-2.5%
Cuba	road transport	-43.0%	-2.5%
	energy (crude oil)	-5.9%	0%

In the paper we reformulated the last two sentence in this section with some more information as: “For the other developing countries (calculated with the EDGARv4.3 data and based on the IEA(2013) fuel statistics), the SO<sub>2</sub> emissions of the energy sector slightly increase from 2008 to 2010 because of the increased coal use (as also observed by Weng et al., 2012) and the increased use of heavy fuel oil in the Middle East. The PM emissions from the energy and industry of some other developing countries show a decrease from 2008 to 2010, mainly due to the activity reduction but also in some cases due to the modelled decrease in controlled emission factor in EDGARv4.3. Largest reductions were seen for Brazil (with 54% reduction of its 2008 charcoal production) and Kazakhstan (11% reduction in coal power generation, which is modelled with a 31% decreasing BC emission factor).”

Reformulation of some sentence were undertaken, as suggested by referee #2 and resulted in:

Page 12870, Line 7-11: “The per capita emissions of all world countries, classified from low to high income, reveal an increase in level and in variation for gaseous acidifying pollutants, but not for aerosols. For aerosols an opposite trend is apparent with higher per capita emissions of particulate matter for low income countries.”

Page 12871, Line 5-9, “Responsibility of providing emission inventories to several international bodies is often distributed within a particular country: e.g. the methane inventory of some Annex I countries is provided by different national institutions. Although they represent the same region, they might be different, which is often the case and leads to confusion (Janssens-Maenhout et al., 2012).”

Page 12871, Line 24-27: “For example, the atmospheric modelling groups, which contributed to the HTAP multi-model experiments described in HTAP (2010), used their own best estimates for emissions for the year 2001, obtaining in some cases comparable

global emissions (e.g. for NO<sub>x</sub> and SO<sub>2</sub> model input), and sometimes getting larger differences in the model input (e.g. for NMVOC emissions).”

*Minor comments*

The authors made the typographic corrections as suggested on Page 12872 (Line 5), Page 12875, Line 7, Page 12884, Line 20, Page 12885, Line 25, Page 12889, Line 7, Page 12890, Line 25, Page 12891, Line 17-18, Page 12904, Fig.1.

We prefer not to change the labeling of the agriculture with number 7, to avoid confusion with the former HTAP definitions. Agriculture was always number 8 but the former 7 and 4 are converted to 4. Therefore number 7 is no longer existing now.

The authors took the decision to refer to this database unambiguously as “HTAP\_v2.2” and corrected this as such through the paper.

1 **HTAP\_v2.2: a mosaic of regional and global emission**  
2 **gridmaps for 2008 and 2010 to study hemispheric transport**  
3 **of air pollution**

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5 **G. Janssens-Maenhout<sup>1,\*</sup>, M. Crippa<sup>1</sup>, D. Guizzardi<sup>1</sup>, F. Dentener<sup>1</sup>, M. Muntean<sup>1</sup>,**  
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## 6 **Abstract**

7 The mandate of the Task Force Hemispheric Transport of Air Pollution (HTAP) under the  
8 Convention on Long-Range Transboundary Air Pollution (CLRTAP) is to improve the  
9 scientific understanding of the intercontinental air pollution transport, to quantify impacts on  
10 human health, vegetation and climate, to identify emission mitigation options across the  
11 regions of the Northern Hemisphere, and to guide future policies on these aspects.

12 The harmonization and improvement of regional emission inventories is imperative to obtain  
13 consolidated estimates on the formation of global-scale air pollution. An emissions dataset  
14 has been constructed using regional emission gridmaps (annual and monthly) for SO<sub>2</sub>, NO<sub>x</sub>,  
15 CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>, BC and OC for the years 2008 and 2010, with the  
16 purpose of providing consistent information to global and regional scale modelling efforts.

17 This compilation of different regional gridded inventories, including the Environmental  
18 Protection Agency (EPA)'s for USA, EPA and Environment Canada's for Canada, the  
19 European Monitoring and Evaluation Programme (EMEP) and Netherlands Organisation for  
20 Applied Scientific Research (TNO)'s for Europe, and the Model Inter-comparison Study ~~for~~  
21 Asia (MICS-Asia [III](#))'s for China, India and other Asian countries, was gap-filled with the  
22 emission gridmaps of the Emissions Database for Global Atmospheric Research  
23 (EDGARv4.3) for the rest of the world (mainly South-America, Africa, Russia and Oceania).  
24 Emissions from seven main categories of human activities (power, industry, residential,  
25 agriculture, ground transport, aviation and shipping) were estimated and spatially distributed  
26 on a common grid of 0.1° × 0.1° longitude-latitude, to yield monthly, global, sector-specific  
27 gridmaps for each substance and year.

28 The HTAP\_v2.2 air pollutant gridmaps are considered to combine latest available regional  
29 information within a complete global dataset. The disaggregation by sectors, high spatial and  
30 temporal resolution and detailed information on the data sources and references used will  
31 provide the user the required transparency. Because HTAP\_v2.2 contains primarily official



1 and/or widely used regional emission gridmaps, it can be recommended as a global baseline  
2 emission inventory, which is regionally accepted as a reference and from which different  
3 scenarios assessing emission reduction policies at a global scale could start.

4 An analysis of country-specific implied emission factors shows a large difference between  
5 industrialised countries and developing countries for acidifying gaseous all-air pollutant  
6 emissions (SO<sub>2</sub> and NO<sub>x</sub>) from the energy and industry sectors. This is ~~but not observed for~~  
7 ~~the particulate matter emissions (PM<sub>10</sub>, PM<sub>2.5</sub>), which show large differences between~~  
8 ~~countries in from the residential sector insteadone. A comparison of the population-weighted~~  
9 ~~emissions for all world countries, grouped into four classes of similar income, reveals that the~~  
10 ~~per capita emissions are, with increasing income group of countries, increasing in level but~~  
11 ~~also in variation for all air pollutants but not for aerosols.~~ The per capita emissions of all  
12 world countries, classified from low to high income, reveal an increase in level and in  
13 variation for gaseous acidifying pollutants, but not for aerosols. For aerosols an opposite trend  
14 is apparent with higher per capita emissions of particulate matter for low income countries.

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## 16 1 Introduction

17 Intercontinental transport of air pollution occurs on timescales of days to weeks and,  
18 depending on the specific type of pollutant, may contribute substantially to local scale  
19 pollution episodes (HTAP, 2010). Common international understanding of global air pollution  
20 and its influence on human health, vegetation and climate, is imperative for providing a basis  
21 for future international policies and is a prime objective for the Task Force Hemispheric  
22 Transport of Air Pollution (TF HTAP)<sup>1</sup>. While nowadays many countries and regions report  
23 their air pollutant emissions, these estimates may not be readily accessible, or may be difficult  
24 to interpret without additional information, and their quality may differ widely, having  
25 various degrees of detail and being presented in different formats.

26 The UN Framework Convention on Climate Change (UNFCCC) requires official inventory  
27 reporting that complies with the TACCC principles of quality aiming at Transparency,  
28 Accuracy, Consistency, Comparability and Completeness<sup>2</sup>, reviewed by UNFCCC roster

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<sup>1</sup>More info on [www.htap.org](http://www.htap.org).

<sup>2</sup>Timeliness is recently also considered.

1 experts and made available at their website (UNFCCC, 2013). Under the CLRTAP the parties  
2 need to report emissions to the EMEP Centre for Emission Inventories and Projections  
3 (CEIP), which also reviews data on completeness and consistency. ~~Responsibility of~~  
4 ~~providing emission inventories to several international bodies is often distributed within a~~  
5 ~~particular country and so an inventory for, for example, methane can be provided by different~~  
6 ~~organisations and although they represent the same region they might be different, in fact~~  
7 ~~often are, leading to confusion we need to work with (Janssens-Maenhout et al., 2012).~~  
8 ~~Responsibility of providing emission inventories to several international bodies is often~~  
9 ~~distributed within a particular country; e.g. the methane inventory of some Annex I countries~~  
10 ~~is provided by different national institutions. Although they represent the same region, they~~  
11 ~~might be different, which is often the case and leads to confusion (Janssens-Maenhout et al.,~~  
12 ~~2012).~~

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13 Currently available emission inventories differ in spatial and temporal resolution  
14 (“consistency”), in coverage of geographical area, time period and list of compounds  
15 (“completeness”) and in the sector-specific details of the source calculation (“transparency”).  
16 Moreover the official inventories submitted by countries have at least one year time lag, are  
17 updated with different frequency and with or without review of the historical time series. ~~The~~  
18 ~~work of Lamarque et al. (2010) provides a unique example of a comprehensive ‘composite’~~  
19 ~~historical emissions dataset spanning from 1850 to 2000, mainly based on scientific estimates~~  
20 ~~using a similar methodology of combining country level inventories for most OECD countries~~  
21 ~~with research inventories for Asia and EDGAR for other regions.~~ The dataset also provided  
22 harmonized base-year (2000) emissions that were used as a starting point for the development  
23 of the so-called RCP (Representative Concentration Pathways) emission scenarios (e.g. Moss  
24 et al., 2010; van Vuuren et al., 2011). For other years and specific model domains covering  
25 multiple regions, atmospheric modellers often compile their own emission inputs drawing  
26 upon different pieces of the available inventories. These compilations involve sometimes  
27 arbitrary choices, and are often not clearly described or evaluated. For example, the  
28 atmospheric modelling groups, which contributed to the HTAP multi-model experiments  
29 described in HTAP (2010), used their own best estimates for emissions for the year 2001,  
30 obtaining in some cases comparable global emissions (e.g. for NO<sub>x</sub> and SO<sub>2</sub> model input),  
31 and sometimes getting larger differences in the model input (e.g. for NMVOC emissions).  
32 Moreover, Streets et al. (2010) evaluated the consistency of the emissions used in the various  
33 models and nationally reported emissions. For a follow-up study in HTAP Phase 2, it was

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1 recommended to provide a harmonised emissions dataset for the years 2008 and 2010 in line  
2 with the following 4 major objectives:

- 3 1) To facilitate development   mitigation policies by making use of well documented  
4 national inventories;
- 5 2) To identify missing (anthropogenic) sources and gap-fill them with scientific  
6 inventories for a more complete picture at global scale;
- 7 3) To provide a reference dataset for further emission compilation activities  
8 (benchmarking or scenario exercises);
- 9 4) To provide a single entry point for consistent global and regional modelling activities  
10 focusing on the contribution of long-range (intercontinental) air pollution to  
11 regional air quality issues.

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12 A harmonized global, gridded, air pollution emission dataset has been compiled with  
13 officially reported, gridded inventories at the national scale, to the extent possible and  
14 complemented with science-based inventories for regions and sectors where nationally  
15 reported data were not available.

16 Whereas for a preceding dataset<sup>3</sup> of EDGAR-HTAP\_v1 the nationally reported emissions,  
17 combined with regional scientific inventories and gapfilled with the global set originating  
18 from EDGARv4.2 were all gridded with geospatial data from EDGAR (Janssens-Maenhout et  
19 al., 2012), this time we used regional gridded emissions, which are officially accepted and  
20 complemented with EDGARv4.3 gridmaps (Janssens-Maenhout et al., 2013) for countries or  
21 sectors without reported data.

22 The resulting dataset, named HTAP\_v2.2, is a compilation of annual and monthly gridmaps  
23 of anthropogenic air pollution emissions (with a 0.1°×0.1° grid resolution). It contains region-  
24 specific information on human activity (concerning intensity and geospatial distribution) and

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<sup>3</sup> EDGAR-HTAP\_v1 completed in October 2010 comprises sector-specific annual gridmaps for the six years from 2000 to 2005 and covers air pollutants (CH<sub>4</sub>, CO, NH<sub>3</sub>, NMVOC, SO<sub>2</sub> and NO<sub>x</sub>) and particulate matter with its carbonaceous speciation (PM<sub>10</sub>, PM<sub>2.5</sub>, BC and OC). The annual gridmaps of 0.1°×0.1° resolution are made available via [http://edgar.jrc.ec.europa.eu/national\\_reported\\_data/htap.php](http://edgar.jrc.ec.europa.eu/national_reported_data/htap.php) and the CIERA and ECCAD servers. Documentation is available in the HTAP\_v1 EUR25229EN report of Janssens-Maenhout et al (2012) ([http://edgar.jrc.ec.europa.eu/htap/EDGAR-HTAP\\_v1\\_final\\_jan2012.pdf](http://edgar.jrc.ec.europa.eu/htap/EDGAR-HTAP_v1_final_jan2012.pdf) )

1 on fuel-, technology- and process-dependent emission factors and end-of-pipe abatement, but  
2 it is not as consistent as a globally consistent emission inventory using international statistics  
3 and global geospatial distributions. With the perspective of being used in chemical transport  
4 models, this inventory includes the atmospheric gaseous pollutants (SO<sub>2</sub>, NO<sub>x</sub>, CO,  
5 NMVOC<sup>4</sup>, NH<sub>3</sub>) and particulate matter with carbonaceous speciation (PM<sub>10</sub>, PM<sub>2.5</sub>, BC and  
6 OC)<sup>5</sup>.

7 This paper provides a detailed description of the datasets and of the methodology used to  
8 compute the 0.1°×0.1° gridmaps for 2008 and 2010, which are delivered via the EDGAR JRC  
9 website (see Section 4). Section 2 defines the considered emitting sectors and presents the  
10 original data sources: a) the officially accepted regional/national gridded emission  
11 inventories, which were mainly provided by national and international institutions, and b)  
12 EDGAR\_v4.3 for gap-filling the remaining regions and/or sectors for some substances. In the  
13 HTAP\_v2.2 database, gridmaps were merged together with a “collage/mosaic” approach  
14 instead of gridding the global emission inventory with one single proxy dataset, as done in for  
15 the EDGAR-HTAP\_v1 dataset compilation (Janssens-Maenhout et al., 2012). The  
16 HTAP\_v2.2 inventory aims to obtain more local accuracy on the location of single point  
17 sources compared to the previous HTAP\_v1, but the downside is that a consistent single  
18 location of a specific source of multi-pollutants is no longer ensured, when data originated  
19 from different sources, possibly leading to spurious chemical reactions involving non-linear  
20 chemistry in the air quality models. Section 3 discusses the resulting gridmaps and addresses  
21 the contents of the HTAP\_v2.2 compilation methodology, the assumptions, dataflows and  
22 consistency of the data used to create the global gridmaps. Whereas HTAP\_v2.2 uses more  
23 regional bottom-up data (local information on emission factors, on assumed penetration of

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<sup>4</sup> The non-methane volatile organic compounds (NMVOC) of HTAP\_v2.2 are defined as the total sum of Alkanols, Ethane, Propane, Butanes, Pentanes, Hexanes and higher, Ethene, Propene, Ethyne, Isoprenes, Monoterpenes, Other alk(adi)enes/alkynes, Benzene, Methylbenzene, Dimethylbenzenes, Trimethylbenzenes, Other aromatics, Esters, Ethers, Chlorinated hydrocarbons, Methanal, Other alkanals, Alkanones, Acids, Other Aromatics, all expressed in their full weight, not just C.

<sup>5</sup> Whereas PM<sub>10</sub> is defined as primary emitted aerosols with aerodynamic diameter up to 10 micrometer, PM<sub>2.5</sub> is a subset with aerodynamic diameter up to 2.5 micrometer, including elemental carbon (BC), organic carbon (OC), SO<sub>4</sub><sup>2-</sup>, NO<sub>3</sub><sup>1-</sup>, crustal material, metal and other dust particles. Note that BC and OC are additive to each other but not to PM<sub>2.5</sub> ( $\{BC,OC\} \subset \{PM_{2.5}\}$  and  $\{PM_{2.5}\} \subset \{PM_{10}\}$ ).

1 technology and end-of-pipe control measures in the facilities), the higher spatial accuracy is  
2 sometimes overshadowed by artefacts at borders- at least when graphically displaying the  
3 data. This is followed with an evaluation of the HTAP\_v2.2 by comparing per capita  
4 emissions, emissions per unit of GDP and implied emission factors for the different countries.  
5 The concluding section 4 summarises the purposes, content and access to this dataset that is  
6 currently in use by the HTAP modellers community.

7

## 8 **2 Methods**

### 9 **2.1 Defining the sector-specific breakdown**

10 An overview of the data sources used is given in Table 1a. For the development of  
11 HTAP\_v2.2, a detailed cross-walk table of the US EPA, EDGAR and EMEP (sub)sector-  
12 specific activities has been setup, using all human activities defined in detail by IPCC (1996)  
13 and applied for the reporting under the UNFCCC. The US EPA and the contributing dataset  
14 from Environment Canada, provided the most detailed cross-walk matrix between the  
15 categories used in their national inventory and the full-fledged set of all IPCC categories.  
16 However, a higher level of aggregation was needed to find a common basis with the Asian  
17 emission inventories, which led to the establishment of the 7 categories: Aircraft,  
18 International Shipping, Power Industry, Industry, Ground Transport, Residential and  
19 Agriculture (described in Table 1b underneath).

20 HTAP\_v2.2 focusses only on anthropogenic emissions, in a comprehensive way, but excludes  
21 large-scale biomass burning (forest fires, peat fires and their decay) and agricultural waste or  
22 field burning. We refer to inventories such as GFED3 (van der Werf, 2010) for the forest,  
23 grassland and Savannah fires (IPCC categories 5A+C+4E) and to the 1°x1° gridmaps of  
24 Yevich ~~et al~~ Logan (2003) or the 0.1°x0.1° EDGARv4.2 gridmaps (EC-JRC/PBL, 2011) for  
25 the agricultural waste burning (4F). Moreover, only NH<sub>3</sub> emissions from the agricultural  
26 sector were taken up in the htap\_8\_AGRICULTURE sector of HTAP\_v2.2 inventory, so that  
27 the occasionally reported NO<sub>x</sub> from agricultural waste burning or from biological N-fixation  
28 and crop residues (which is typically considered under S10 for Europe) are excluded.

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## 2.2 Gridded input datasets for HTAP\_v2.2

As explained earlier, the goal of the HTAP\_v2.2 inventory is to provide consistent and highly resolved information (see Fig. 1a) to global and regional modelling. It is important to realize that in the HTAP modelling exercise both global and regional models are participating. The HTAP global modelling is coordinated with the regional modelling exercise of Air Quality Model Evaluation International Initiative AQMEII (Galmarini et al., 2012 and 2015) that manages regional scale activities for Europe and North America, and the regional modelling exercise of the Model Intercomparison study for Asia MICS-Asia (Carmichael et al., 2008) that manages the regional modeling over Asia. Hence, the regional inventories used for HTAP\_v2.2 are constructed and used in accordance with these regional activities.

### 2.2.1 USA and Canada: EPA and Environment Canada gridmaps and EPA temporal profiles

EPA (2013) provides the 2008 and 2010 areal and point source emissions for the complete North American domain at 0.1°x0.1° resolution, covering USA with a grid ranging from 180°W-63°W in longitude and 75°N-15°N in latitude and covering Canada with a grid from 142°W-47.8°W in longitude and 85°N-41°N in latitude. Mexico is not covered by these latitudes and it is gapfilled with EDGARv4.3 data (see section 2.2.4). For the northern latitudes above 45°N, Environment Canada provided the 2008 basis and an update of the point sources for 2010, from which US EPA prepared the full set of detailed gridmaps also for 2010. The 2010 data for Canada were missing and as such extrapolated by US EPA based on the 2008 National Emission Inventory of Environment Canada and assuming no trend but using updated point sources (Pouliot et al., 2014). The temporal profiles of US EPA were applied for USA and Canada with identical monthly distributions per sector for 2008 and 2010. More details about the US inventory are given by Pouliot et al. (2014) and (2015).

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### 2.2.2 Europe: TNO gridmaps and EMEP temporal trends

Countries that are parties to the CLRTAP (<http://www.unece.org/env/lrtap>) need to report anthropogenic emissions of air pollutants and particulate matter, but neither BC nor OC. These reported/official inventories are reported on the national level to EMEP-CEIP<sup>6</sup> which

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<sup>6</sup> More info on [www.ceip.at](http://www.ceip.at).

1 provides the annual emission inventory data for CO, NH<sub>3</sub>, NMVOC, NO<sub>x</sub>, SO<sub>x</sub>, PM<sub>10</sub> and  
2 PM<sub>2.5</sub> (not BC and not OC). However, the currently used EMEP grid uses a polar-  
3 stereographic projection with about 50km x 50km grid cells centered over the European  
4 region and converting to a Mercator projection implied a loss of spatial accuracy. These  
5 reported data are incomplete according to the CEIP annual report of Mareckova et al. (2013)  
6 and for evaluation with the EMEP unified model further gapfilling is needed, resulting in a  
7 semi-official emission dataset. To overcome the problems of inconsistent emissions time  
8 series and fulfil the need for a higher spatial resolution to support AQ modelling in Europe in  
9 the European FP7 project Monitoring Atmospheric Composition and Climate (MACC), TNO  
10 established a scientifically complete and widely accepted dataset, which is fully documented  
11 by Kuenen et al (2014). This so-called TNO-MACC-II inventory of Kuenen et al (2014)  
12 covers the same European domain with areal and point source emission gridmaps at 1/8° x  
13 1/16° resolution for SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub>, PM<sub>2.5</sub> with point sources  
14 allocated to their exact location. The grid-domain ranges from 30°W-60°E in longitude and  
15 72°N-30°N in latitude. The geographical area covered all EU-28 countries, Switzerland,  
16 Norway, Iceland and Liechtenstein, Albania, Bosnia-Herzegovina, Serbia, Macedonia, 6  
17 Newly Independent States (Armenia, Azerbaijan, Belarus, Georgia, Moldova, Ukraine) and  
18 Turkey. [EMEP-TNO data for E-countries](#) with [only](#) partial coverage (Russia, Turkmenistan,  
19 Kazakhstan and Uzbekistan) were not used in the [HTAP\\_v2](#) inventory because of  
20 inconsistencies with other datasets (see section 2.2.4). Sector-specific data (given by SNAP-  
21 code, see Table 1b) are used for all countries with complete coverage of their territory and for  
22 each substance the contribution from each sector is compared to EMEP and EDGARv4.3  
23 estimates. Standard re-sampling is applied to obtain gridmaps at the common resolution of  
24 0.1°x0.1°. Point-source, ground-level airport emissions in the transport sector (under SNAP 8)  
25 were taken out, in order to avoid a double counting with the aviation sector (HTAP-1 AIR),  
26 for which the same geospatial dataset taken from EDGAR\_v4.3 was used globally.

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27 [The EMEP-TNO data were only available for 2006 and 2009. The 2008 data for Europe is](#)  
28 [based on the EMEP-TNO data for 2009 data and the 2010 data for Europe are based on the](#)  
29 [same 2009 data but using the trend in EMEP-TNO data between 2006 and 2009.](#)

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1 For NH<sub>3</sub>, the reporting of emissions from the energy, industry and residential sectors was  
2 apparently negligible for some countries<sup>7</sup> compared to the agricultural emissions and was  
3 therefore not gapfilled by EMEP and/or TNO.

4 BC and OC emission data are not available as emission gridmaps within the MACC-II  
5 dataset, but the PM gridmaps are accompanied by a recommendation on the PM composition  
6 describing the carbonaceous profiles per SNAP code and country. This so-called PM split  
7 table (per SNAP code and country) of TNO (TNO, 2009) is used to derive the BC and OC  
8 from PM<sub>10</sub> and PM<sub>2.5</sub> emission gridmaps (see Kuenen et al. (2014) for details).

9 ~~Finally, to derive the monthly gridmaps the EMEP modelling group provided we used the~~  
10 ~~country-specific and sector-specific data the monthly profiles, which are with a monthly~~  
11 ~~factors varying around 0.08334/42 specified for each country and for each sector, with a~~  
12 ~~further substance-specific variation for the agricultural sector per substance for the EMEP~~  
13 ~~model~~ (personal communication with M. Schulz of 27 May 2013 and A. Nyiri of 4 June  
14 2013).

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### 15 2.2.3 Asia: monthly gridmaps from MIX

16 For Asia, a different challenge is faced, because no countries except Japan are legally required  
17 to yearly report detailed emission inventories under the LRTAP, UNFCCC or similar  
18 conventions. However, in Asia many scientific efforts aimed at establishing a detailed  
19 emission inventory, accepted by the different regions, using official or semi-official statistics  
20 collected at county level (by provinces for China). Under the Model Inter-comparison Study  
21 for Asia Phase III (MICS-Asia III), a mosaic Asian anthropogenic emission inventory was  
22 developed for 2008 and 2010 (Li et al., 2015). The mosaic inventory, named MIX,  
23 incorporated several local emission inventories including the Multi-resolution Emission  
24 Inventory for China (MEIC), NH<sub>3</sub> emission inventory from Peking University (Huang et al.,  
25 2012), Korean emissions from the Clean Air Policy Support System (CAPSS) (Lee et al.,  
26 2011), Indian emissions from the Argonne National Laboratory (Lu et al, 2011), and fill the

---

<sup>7</sup>No NH<sub>3</sub> emissions are reported in the energy sector: for the countries Albania, Bosnia-Herzegovina, Cyprus, Estonia, Greece, Ireland, Iceland, Luxembourg, Latvia, FRY Macedonia, Malta, Norway, Poland, Romania, Slovakia, and Slovenia; in the industry sector for the countries Albania, Bosnia-Herzegovina, Greece, Ireland, Iceland, and FRY Macedonia; and in the residential sector for the countries Greece, Iceland and Slovenia.



1 gap where local emission data are not available using REAS2.1<sup>8</sup> developed by Kurokawa et  
2 al. (2013).

3 MEIC is developed by Tsinghua University under an open-access model framework that  
4 provides model-ready emission data over China to support chemical transport models and  
5 climate models at different spatial resolution and time scale. In the MIX inventory, the MEIC  
6 v.1.0 data was used which contains the anthropogenic emissions of China for SO<sub>2</sub>, NO<sub>x</sub>, CO,  
7 NMVOC, NH<sub>3</sub>, CO<sub>2</sub>, PM<sub>2.5</sub>, PM<sub>coarse</sub>, BC, and OC for the years 2008 and 2010 with  
8 monthly temporal variation at 0.25° x 0.25°. For India, MIX used the Indian emission  
9 inventory provided by ANL for SO<sub>2</sub>, BC, and OC and REAS2.1 for other species. With the  
10 input from different regions, the MIX inventory provided harmonized emission data at 0.25° x  
11 0.25° grid resolution with monthly variation for both 2008 and 2010. The detailed mosaic  
12 process of the MIX inventory is documented in Li et al. (2015). Reported emissions from  
13 countries which are only partly covered by the MIX, like Russia, Turkmenistan, Uzbekistan  
14 and Kazakhstan were not taken up in the HTAP inventory and instead gap-filling by  
15 EDGARv4.3 was used (see section 2.2.4).

16 As such, countries within the a broad area, spanningranging from 89.875°N to 20.125°S in  
17 latitude and from 40.125°E to 179.875°E in longitude were inserted in the as-after-a-raster  
18 resample procedure covered by 0.1° x 0.1° emission gridmaps after converting the 0.25° x  
19 0.25° with a raster resample procedure – dividing the cells in 5x5 and then aggregating the  
20 0.05°x0.05° cells 2x2. Monthly gridmap results (without distinction between point and areal  
21 sources and without temporal profiles) are given per sector (energy, industry, residential,  
22 transport, and agriculture only for NH<sub>3</sub>).

### 23 2.2.4 Rest of the world covered by EDGARv4.3

24 The Emission Database for Global Atmospheric Research (EDGAR) of EC-JRC/PBL (2011)  
25 provides historical (1970-2008) global anthropogenic emissions of greenhouse gases<sup>9</sup> CO<sub>2</sub>,

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<sup>8</sup> The REAS2.1 inventory for Japan includes the data developed by Ministry of the Environment of Japan (MOEJ, 2009) for NMVOC evaporative emissions from stationary sources, the database developed by the Ocean Policy Research Foundation (OPRF, 2012) for the maritime sector, and the Japan Auto-Oil Program Emission Inventory-Data Base (JEI-DB) developed by Japan Petroleum Energy Center (JPEC, 2012a, b, c) for other sources.

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1 CH<sub>4</sub>, N<sub>2</sub>O, HFCs, PFCs and SF<sub>6</sub>, of precursor gases, such as CO, NO<sub>x</sub>, NMVOC and SO<sub>2</sub>  
2 and of aerosols (PM<sub>10</sub>), including PM<sub>10</sub>, PM<sub>2.5</sub>, BC and OC per source category at country  
3 level on 0.1° x 0.1° gridmaps. ~~This dataset is in the version EDGARv4.3 extended to with the~~  
4 ~~years 2009 and 2010 in EDGARv4.3 and covering with the carbonaceous species also the~~  
5 ~~substances PM<sub>2.5</sub>, BC and OC. For HTAP v2.2 a preliminary version of the EDGARv4.3~~  
6 ~~(JRC-EC/PBL, 2015) is used.~~ Emissions are calculated by taking into account human activity  
7 data of IEA (2013) for fuel consumption and of FAO (2012) for agriculture, different  
8 technologies with installed abatement measures, uncontrolled emission factors (IPCC, 2006)  
9 and emission reduction effects of control measures (EMEP/EEA, 2013). Anthropogenic  
10 emissions calculations are extended till 2010 for all 246 world countries for the emission  
11 source (sub)groups; (i) combustion/conversion in energy industry, manufacturing industry,  
12 transport and residential sectors, (ii) industrial processes, (iii) solvents and other product use,  
13 (iv) agriculture, (v) large scale biomass burning, (vi) waste and (vii) miscellaneous sources. ~~A~~  
14 ~~detailed overview of the EDGAR emissions database and how it can be used for gapfilling~~  
15 ~~can be found in Balsama et al (2014).~~

16 The EDGAR emission data are spatially distributed using an extensive set of global proxy  
17 data, which are representative for major source sectors and documented in the EDGAR  
18 gridding manual of Janssens-Maenhout et al. (2013). For HTAP\_v2.2, the EDGARv4.3  
19 database provides yearly emission gridmaps with a resolution of 0.1x0.1 degree for the “rest  
20 of the world” countries of Table AS1.2 of Annex I in the Supplement for all pollutants (SO<sub>2</sub>,  
21 NO<sub>x</sub>, CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>, OC, BC) and HTAP sectors for the years 2008 and  
22 2010. The htap\_2 SHIPS data are provided for the entire world, while the htap\_1  
23 AIRAviation data are provided for the entire world for the international aviation part and for  
24 the world excluding USA and Canada for the domestic aviation. EDGAR provides also  
25 sector-specific monthly profiles, defined with first order, ~~gu~~ estimated factors for each of the  
26 three different zones: Northern Hemisphere, Equatorial region and Southern Hemisphere  
27 (Table SA1.2). A reverse profile is applied for the two Hemispheres from the EDGAR v4.3  
28 database, while no seasonal pattern is used for the Equatorial regions. Monthly emissions  
29 gridmaps are generated from the annual emission data per HTAP sector using these EDGAR

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<sup>9</sup> The methodology for the greenhouse gas emission time series applied in EDGARv4.2 is detailed in Olivier and Janssens-Maenhout (2012).

1 monthly factors, which resemble most to the EMEP-TNO profiles (see section 2.3) defined  
2 for the three different zones: Northern Hemisphere, Equatorial region and Southern  
3 Hemisphere (Table A1.2).

4 The countries with partial geo-spatial coverage under the MACC-II and MIX inventories (see  
5 sections 2.2.2 and 2.2.3) are completely replaced with EDGARv4.3 data to avoid  
6 inconsistencies and artefacts at the border between two datasets within one country (such as  
7 Russia, Kazakhstan, Turkmenistan and Uzbekistan). This replacement took place after the  
8 gridmaps were converted into  $0.1^\circ \times 0.1^\circ$  using a raster resampling procedure. For EMEP-  
9 TNO the resampling implied a 25-fold division to  $0.0025^\circ \times 0.0125^\circ$  followed by an  
10 aggregation of  $4 \times 8$  gridcells. For the MICS-AsiaMIX the resampling needed also a 25<sup>th</sup> fold  
11 division to  $0.05^\circ \times 0.05^\circ$  followed by an aggregation of  $2 \times 2$  gridcells. The cells including  
12 country borders are split up and allocated to the different countries using the corresponding  
13 areal percentage.

### 14 **2.3 Overview of the temporal profiles used in HTAP v2.2**

15 The modulation of annual emissions over time is necessary in order to provide the modelers  
16 emission data consistent with the seasonal pattern and activities. Monthly data were generated  
17 for all sectors except for the international shipping and international aviation, which are  
18 considered constant over the year. US-EPA, EMEP and EDGAR provided monthly profiles,  
19 but MICS-Asia provided directly and solely monthly emission gridmaps.

20 Figure 1c summarizes the sector-specific monthly profiles for each of the regional datasets.  
21 The temporal profiles are additive and specified with monthly factors modulating around 1/12  
22 for each of the sectors. For the agricultural sector, EMEP provided compound-specific  
23 monthly factors, which characterise high NMVOC emission in spring and high CO emission  
24 in autumn. Agriculture (largely contributing to NH<sub>3</sub> emissions) shows most seasonal  
25 variation, which differs also most between the different regions because of region-specific  
26 management practices (for e.g. crop cultivation), climate and geographical location and soil  
27 composition. The residential sector is characterized by a monthly distribution which is  
28 inversely related with the temperature and therefore with the use of heating systems, and in  
29 some developed countries with air conditioning. In some developed countries with hot  
30 summers, the air conditioning is again boosting emissions during the summer. The seasonality  
31 remains relatively modest in all regions for the sectors transport, industry and energy.

The strongest variation over the year and between regions is observed for the agricultural sector (+215% in the EMEP-TNO profiles but only +45% in the MIXMICS-Asia profiles), followed by the residential sector ([+70%, -75%] in the EMEP-TNO profiles, [+20%, -25%] in the US EPA profiles and [+115%, -40%] in the MIXMICS-Asia profiles).

### 3 Results

Monthly global gridmaps were produced for 2008 and 2010 and are available per htap sector and substance at [http://edgar.jrc.ec.europa.eu/htap\\_v2/index.php?SECURE=123](http://edgar.jrc.ec.europa.eu/htap_v2/index.php?SECURE=123). —We describe major characteristics of the gridmaps in section 3.1. We focus on 2010 but the observations remain valid for 2008 (in the same period of recession). A summary graph of the emission totals and their sector-specific composition is given in Fig. 1b. In sections 3.2 and 3.3 we put the country totals (given bottom-up except for the MICS-Asia regions, where we derived the totals from the gridmaps) in perspective with a comparative analysis of the emissions per capita and emissions per GDP for low, lower middle, upper middle and high income country groups. To estimate how polluting the activities are in the different regions, section 3.4 addresses the implied emission factors. Finally, we address the difference in emissions 2008 to 2010 in section 3.5 and we conclude with a qualitative assessment of the uncertainty of the gridmaps in 3.6.

#### 3.1 Spatial distribution of global emissions per sector

An overview on the region-specific totals and the composition per region and sector is given in the 9 maps of Fig. 2a-i for the different substances for the year 2010. The sector-specific country-totals are given in Table AS1.1 and the totals for each of the 16 HTAP source region, as defined for the source-receptor calculations of the HTAP modelling community and described in Table SA2.1 are given in Table SA2.2 of Annex II in the Supplement. Before focusing on the emissions over land surface, we assess the global shipping emissions. Table 2a. compares the international shipping emissions with the bottom-up and top-down estimated emissions reported by IMO (2014). We note that an agreement between the data of HTAP (EDGAR based), and IMO (both top down and bottom up estimates) is obtained for all compounds within 30%, except for CO. For the latter EDGAR shows a 55% and 70% higher

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1 estimate for the 2008 and 2010 bottom-up values of the IMO (2014) study, which on his turn  
2 is 55% respectively 33% higher than the 2008 and 2010 top down estimates of the IMO(2014)  
3 study. It is worth mentioning that a 250% downscaling of the CO emission factor was  
4 undertaken in IMO (2014) compared to the previous study of IMO (2009).

5 Developing countries contribute from 70% to more than 90% to the current global  
6 anthropogenic pollutant emissions, depending on the considered compound and Asian  
7 countries are the major emitters, contributing from 40% to 70%. Among these countries,  
8 China and India represent two densely populated regions, producing together more than two  
9 thirds of the total Asian emissions. On the contrary, developed regions (like North America  
10 and Europe) produce much lower emissions, representing overall from 30% down to 10% of  
11 the total annual global anthropogenic emissions. Since the rest of the world group of countries  
12 includes a variety of regions, differing in population, human activities, types of industries,  
13 etc., it is crucial to disaggregate it into its components. In particular for PM2.5 and somewhat  
14 less for NOx, Asia strongly contributes to the global emissions compared to the contribution  
15 of North America and Europe.

16 Generally, higher emissions are observed for populated areas and coastal regions, but specific  
17 features can be highlighted depending on the pollutant and activity for specific countries per  
18 substance. The differences of the figures 2a-2i in the sector-specific composition (pie charts)  
19 of the emission sources for world regions (represented by the color scale) vary strongly  
20 between compounds. Some of the factors include:

- 21 • For SO<sub>2</sub> the emissions will depend on the importance of coal used in the industry and  
22 residential sectors and the degree of flue gas desulphurization. In some regions non-  
23 ferrous metals industry will be of great importance.
- 24 • For NO<sub>x</sub> emissions industrial combustion and transport are key and with increasing  
25 level of activity the application of end-of-pipe controls, including catalytic reduction  
26 of flue gases, is playing an ever increasing role.
- 27 • CO and NMVOC emissions are dominated by incomplete combustion (cooking and  
28 heating stoves) and transport, especially in absence of advanced controls. For  
29 NMVOC additionally evaporative losses from solvent use and oil industry are of high  
30 relevance.

- Finally for PM, incomplete combustion (stoves) and in developing countries poor efficiency of filters installed on industrial boilers can be a source of large emissions while more recently transport emissions from diesel engines became of concern.

## SO<sub>2</sub>

The Asian region ~~keeps suffering~~ is still characterised by ~~from~~ a relative large contribution of SO<sub>2</sub> from (coal fired) power plants and manufacturing industry. Most of the SO<sub>2</sub> emitted in North America and Europe comes from coal power plants. However, in Europe Fig. 2a shows that SO<sub>2</sub> is also emitted from the residential and waste disposal sector. Residential (heating and cooking) and waste disposal sources are particularly relevant in Africa. High annual SO<sub>2</sub> emissions are also observed for India, to which the energy sector contributes 59% and the energy-intensive manufacturing industry (iron & steel) 32% ~~and correspond to high contributions from the industrial combustion, both using also coking and bituminous coal in the power and iron & steel industry~~ according to IEA (2013). Finally, international shipping contributes ~10% to the global SO<sub>2</sub> emissions. SO<sub>2</sub> gridmaps clearly show the ship emission tracks connecting Asia and Europe with Africa and America.

## NO<sub>x</sub>

Figure 2b shows that the major sources of NO<sub>x</sub> are ground transport and power generation and these source contributions show a rather uniform feature for all the considered regions. In Central and South America major emissions are attributed to the transportation sector and just to a minor extent to the energy sector (e.g. in Mexico 65% of the NO<sub>x</sub> emissions originate from road transport). Those industrialised countries with a large share of natural gas as fuel for heating houses and commercial centres and for industry (such as Canada, the Netherlands, Norway) show relatively high emissions of NO<sub>x</sub>: the share of the residential and industry NO<sub>x</sub> emissions is around 30% of the total NO<sub>x</sub>, whereas in USA this is only 20%. International shipping and, in particular, aviation contribute together more than 10% of global NO<sub>x</sub> emissions.

## CO

CO is a product of incomplete combustion, which can therefore be emitted by any fuel combustion (ground transport, industrial processes involving combustion, as well as domestic heating). As presented in Fig. 2c, the power generation sector emits less CO than the residential one because of higher combustion efficiency and higher temperatures compared to

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1 domestic burners. In Africa, there are large emissions of CO from the residential sector,  
2 mainly due to the use of wood and charcoal for cooking activities. As shown in Fig. 2c, some  
3 industrial activities emit CO, like the production of non-metallic minerals and crude steel and  
4 iron, which is particularly relevant for India and China, while non-ferrous metal and iron and  
5 steel production are dominant in Oceania.

## 6 **NMVOOC**

7 NMVOCs (non-methane volatile organic compounds) are emitted from chemical and  
8 manufacturing industries, as well as fuel transformation processes, the production of primary  
9 fuels, the use of solvents and from the residential sector, inclusive waste (Fig.2d). Important  
10 sources of NMVOCs include also evaporative emissions from road transport, specifically  
11 gasoline engines and the use of biofuels. Major emission sectors in the USA emitting  
12 NMVOCs include oil refineries, oil and gas production, several industrial processes and  
13 motor vehicles. Most of the NMVOC emissions in Europe are due to solvent use, road  
14 transport, and the use of primary solid biomass in the residential sector. **In the Middle East**  
15 **NMVOC sources include oil production: the industry sector in Saudi-Arabia contributes 75%**  
16 **to its total NMVOC emissions, and in South-Eastern Asia charcoal production.** In China,  
17 particular high emissions are **originating from industry (62%) and residential (27%), the latter**  
18 **also associated with the high use of solvents in paints.** ~~and~~ In Brazil **particular high use of**  
19 **biogasoline is present resulting in a 52% NMVOC contribution of the transport sector with the**  
20 **use of biofuels. Also the production of charcoal is emitting strongly NMVOC and the world**  
21 **top 3 emitters (IEA, 2013) are Brasil, Thailand<sup>10</sup>, and Kenya, which explains that their**  
22 **industry sector is contributing to the NMVOC total with respectively 35%, 37% and 80% in**  
23 **2010.** NMVOC speciation is not provided by the HTAP\_v2.2 emission database; however  
24 TNO has produced a breakdown into 23 **NMVOC** species, which has been used for the  
25 RETRO project and the RCP scenarios of IPCC AR5. Recommendations for the NMVOC  
26 splits are given on the HTAP wiki site <http://iek8wikis.iek.fz-juelich.de/HTAPWiki/WP1.1>.

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## 27 **NH3**

28 NH3 is mainly emitted by the agricultural sector, including management of manure and  
29 agricultural soils (application of nitrogen fertilizers, incl. animal waste), as Fig. 2i shows,

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<sup>10</sup> No charcoal production emissions are accounted for in the REAS2.1 inventory, which is a shortcoming mainly for Thailand.

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1 while a relatively small amount is emitted by the deployment of catalysts in gasoline cars.  
2 Minor contributions are also observed for Asian countries from the residential sector due to  
3 dung and vegetable waste burning and coal combustion. For industrialized regions, especially  
4 for countries using low sulphur fuel, Mejía-Centeneo et al. (2007) reported that the  
5 deployment of catalytic converters in gasoline cars enhanced the NH<sub>3</sub> emissions from this  
6 source since mid-2000. This is also observed by the larger NH<sub>3</sub> with increased transport  
7 activity and corresponding increased consumption of low sulphur fuels. In the USA gasoline  
8 vehicle catalysts represent ca 6% of total NH<sub>3</sub> emissions, while a lower contribution is found  
9 for Europe due to the high deployment of diesel vehicles.

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### 10 **PM10 and PM2.5**

11 Particulate matter (PM), both in the fine and coarse fraction, is mainly emitted by biomass  
12 and fossil fuel combustion in domestic and industrial activities (Figs. 2e and 2f). On the  
13 contrary, ground transportation contributes ~5% to total PM emissions (excluding non-  
14 exhaust road abrasion dust and tyre wear emissions). As depicted in Fig. 1b, developed  
15 countries (like USA and EU) represent ~10% of global emissions of PM and its components,  
16 while much higher contributions derive from developing countries where less strict legislation  
17 is applied in the industrial sector and in road transport. Figs. 2e and 2f show a similar  
18 composition of the contributing sectors to PM10 and PM2.5 globally. PM10 and PM2.5  
19 gridmaps point out the enhanced PM emissions in Asian countries, due to industrial processes  
20 and the residential sector. A decreasing trend from 2008 to 2010 is observed for Brazil due to  
21 decreases in emissions from charcoal production (with 23% share in the world production in  
22 2008 and 12% in 2010, according to IEA, 2013). Emissions from charcoal production are also  
23 important for some South-Eastern Asia (Thailand, Philippines, Indonesia, Vietnam, Malaysia)  
24 and some African countries (Kenya, Sudan, South Africa, Tanzania, Ethiopia), with country-  
25 specific shares in world production varying between 1.3% and 12.940% according to IEA  
26 (2013). Western Africa generally emits more PM than the Eastern part because of more  
27 industrial activities.

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### 28 **BC and OC**

29 Black carbon (BC), the light-absorbing component of the carbonaceous part of PM, and  
30 organic carbon (OC) are emitted from incomplete combustion. Major emission sources are  
31 residential cooking and heating (fossil fuel and biomass combustion) and for BC also ground  
32 transport (especially diesel engines). Very low emissions originate from the energy sector due



1 to higher process efficiencies and high combustion temperatures. Fig.2g shows that the largest  
2 contributing sector for BC in North America, Europe and the Middle East is road transport,  
3 ~~which can be allocated, should be~~ mainly from ~~to~~ diesel vehicles ~~given the much higher BC~~  
4 ~~emission factor for diesel than for petrol.~~ Heavy duty and light duty vehicles in these regions,  
5 ~~as well as~~ ~~but~~ also diesel passenger cars in Europe and the Middle East, cause this relatively  
6 large contribution despite the use of particle filters, which have not yet fully penetrated the  
7 fleet. For Asia, Oceania, Africa and Central- and South-America, the residential sector is the  
8 main contributor of BC emissions. ~~In China and India the industry and residential sectors~~  
9 ~~contribute to respectively 84% and 91% of their total BC emissions, while this share in USA~~  
10 ~~or in Germany is only 42% respectively 36%.~~ ~~emit more BC than Western industrialized~~  
11 ~~countries from.~~ With the IEA (2003) data this indicates to the combination of high use of coal  
12 ~~(mainly in China) and of biomass (mainly for India) in power plants, the coke ovens and non-~~  
13 ~~metallic mineral industries, as well as the residential from domestic heating activities~~  
14 ~~involving the combustion of solid biomass and bituminous coal and charcoal production. The~~  
15 ~~residential sector in China accounts for more than half (52%) of its BC total. Russia shows a~~  
16 ~~similar high share of the residential sector (46%) to its total BC.~~ Most important sources  
17 ~~calculated in EDGARv4.3 for heating buildings in Russia include bituminous coal (57%) and~~  
18 ~~primary solid biomass (30%), lignite (6%) and industrial waste (3%) burning in the residential~~  
19 ~~sector (for domestic housing as well as commercial services) and other bituminous coal~~  
20 ~~combustion in the commercial sector and in the cogeneration and heat plants (EC-JRC/PBL,~~  
21 ~~2011 and IEA, 2013).~~ A different situation is observed for Africa, where in addition to  
22 emissions from traffic and oil production, an important role is played by charcoal production  
23 and the use of primary solid biomass and charcoal in the residential sector. Nigeria has high  
24 flaring emissions from oil and gas production and Kenya and Sudan suffer from large  
25 charcoal production activities. For OC (Fig. 2h), all regions except the Middle East show that  
26 the largest emission contribution comes from the residential sector (combustion of charcoal  
27 and solid biomass). For the Middle East a relatively large contribution from industrial  
28 activities (fuel production) is observed.

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### 30 3.2 Per capita emissions

31 To compare emissions from worldwide countries characterized by different degrees of  
32 development and numbers of inhabitants, per capita emissions were calculated. Country-

1 specific per capita total emissions are given in Table SA3.1 of Annex III in the Supplement.  
2 ~~In Table 2b we compare –and an example– for the world top 6 CO2 emitters three selected~~  
3 ~~countries, China, USA, India, Russia, Japan and Germany –and China– the per capita air~~  
4 ~~pollutant emissions while making the link with the country’s activity level and level of clean~~  
5 ~~technologies development. is given in Table 2 below.~~ Country total population data were

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6 obtained from the United Nations Population Division (UNDP, 2013). This approach  
7 allocates the emissions from industrial production to a country without taking into account  
8 exports. No life cycle assessment of products at the point of consumption is considered here.  
9 This production-based approach has limitations as moving heavy industry from industrialized  
10 to developing countries under this production-based approach puts a large burden on countries  
11 (in particular those with small populations and mining/manufacturing activities for export).  
12 For example mining for export is having a growing impact in Oceania (with low population)  
13 and industrial production in China for international markets became increasingly important  
14 since 2002 when China entered the World Trade Organisation. ~~The importance of this~~

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15 ~~consumption- versus production-based approach can be expected in 2008 (and also 2010) to~~  
16 ~~be at least but probably even larger than what Boitier (2012) and Davis et al. (2011) amongst~~  
17 ~~others reported for CO<sub>2</sub>. A consumption-based approach would yield at least 10% higher~~  
18 ~~emissions for industrialised countries whereas 10% lower emissions for developing countries~~  
19 ~~with emerging economy.~~

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20 ~~For SO<sub>2</sub> the per capita emission in 2010 for EU-28 of 9.1 kg SO<sub>2</sub>/cap is slightly lower~~  
21 ~~than very close to the reported value of 11.58.9 kg SO<sub>2</sub>/cap from EuroSTAT (2014) - the 0.2~~  
22 ~~difference is much less than the 20% higher per capita SO<sub>2</sub> emission in 2008 (11.5 kg~~  
23 ~~SO<sub>2</sub>/cap). EU’s ~~this~~ 9.1 kg SO<sub>2</sub>/cap is about half the SO<sub>2</sub> per capita for China in 2010 and~~  
24 about one third of the SO<sub>2</sub> per capita for USA. Significant reductions of the Chinese SO<sub>2</sub> per  
25 capita emissions started due to the introduction of very strict emission limits followed by  
26 ambitious flue gas desulfurization programs in power plants (Lu et al. 2011; Klimont et al.  
27 2013; Wang et al., 2014). China is expected to follow the European example, where the SO<sub>2</sub>  
28 per capita decreased from 1995 to 2005 with 65% of the decrease occurring in Germany and  
29 UK according to Ramanathan & Feng (2009).

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30 For NO<sub>x</sub> and NMVOC, China is similar to the European per capita levels. North America and  
31 Oceania double the level of European and Asian per capita emissions of NO<sub>x</sub> and NMVOC  
32 for industrial combustion and transport mainly due to their larger fuel consumptions in the

1 industry (Olivier et al., 2013) and road transport (Anderson et al., 2011) sectors, while having  
2 similar abatement technologies.

3 ~~The level of per capita air pollution results from a combination of the per capita activity and~~  
4 ~~the level of implementation of end-of-pipe measurement technology. The activity level can be~~  
5 ~~reflected by the per capita CO<sub>2</sub> emissions, which is highest for USA explaining the high air~~  
6 ~~pollutant emissions per capita. However not India with lowest CO<sub>2</sub> per capita, but Japan and~~  
7 ~~Germany are having the lowest per capita air pollutant emissions, because of the level of~~  
8 ~~technology and end-of-pipe implementation. To measure the latter we apply a kind of~~  
9 ~~surrogate variable: the Human Development Indicator (2010) from UNDP(2015). This shows~~  
10 ~~that Germany and Japan are more advanced and have therefore lower emissions per capita for~~  
11 ~~all air pollutants (except NH<sub>3</sub> for Germany) and for the PM. Whereas for NO<sub>x</sub>, NMVOC, CO~~  
12 ~~and NH<sub>3</sub>, the ranking from lowest to highest per capita emitters in Table 2 is Germany—~~  
13 ~~China—USA, for PM this order is not present. We observe that the PM emissions per capita~~  
14 ~~of Japan (0.16 kgPM<sub>2.5</sub>/yr/cap) are only 60% of those of Germany and Germany's one are~~  
15 ~~about one fifth of the per capita emissions of the USA, which are on their turn only 60% of~~  
16 ~~and the per capita PM<sub>2.5</sub> for China is more than double the per capita PM<sub>2.5</sub> of the USA.~~  
17 ~~Table SA3.1 indicates that developing countries, in particular those with emerging economies~~  
18 ~~but not yet fully penetrated clean technologies and end-of-pipe measures, have enhanced PM~~  
19 ~~per capita emissions (China – 8.2 kgPM<sub>2.5</sub>/yr/cap, India – 5.2 kgPM<sub>2.5</sub>/yr/cap, Brasil – 3.1~~  
20 ~~kgPM<sub>2.5</sub>/yr/cap). Russia has relatively high per capita PM emissions (2.2 kg PM<sub>2.5</sub>/yr/cap~~  
21 ~~because of for fossil fuel production and consumption in the power generation sector, but~~  
22 ~~much less than, while Canada (7.4 kg PM<sub>2.5</sub>/yr/cap), a much less populated country but with~~  
23 ~~important fossil fuel production industry for export shows high values for non power~~  
24 ~~industry. Both countries, with important contribution in the Arctic region, show relatively~~  
25 ~~high NMVOC and SO<sub>2</sub> emissions (50.9 kg NMVOC/yr/cap and 48.7 kg SO<sub>2</sub>/yr/cap for~~  
26 ~~Canada respectively 26.8 kg NMVOC/yr/cap and 31.9 kg SO<sub>2</sub>/yr/cap for Russia) due to~~  
27 ~~their significant inland waterway transport using heavy residual fuel oil or diesel.~~

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28 Fig. 3 gives an overview of the per capita emissions for high, upper and lower middle and low  
29 income countries, as defined for the WGIII of AR5 of IPCC (2014). The largest variation  
30 between the different groups of countries is observed for SO<sub>2</sub> and NO<sub>x</sub>, which represent the  
31 presence of industry. The median of per capita SO<sub>2</sub> and NO<sub>x</sub> emissions are higher for high  
32 and upper middle income countries than for low or lower middle income countries. The

1 median of per capita CO and NMVOC is not strongly dependent on the income of the  
2 countries, whereas the median of per capita PM (and BC and OC) are definitely lower for  
3 high income countries than for low income countries.

4

### 5 3.3 Per GDP emissions

6 Another indicator of emission intensity of a country is the ratio of emissions and Gross  
7 Domestic Product (GDP) in USD, in constant Purchasing Power Parity (PPP), as given in  
8 Table SA3.2 of Annex III and shown in Fig. 3b. The GDP 2010 data for the different  
9 countries were obtained from World Bank (2014) and IMF (2014). This indicator is much  
10 more uncertain than the per capita emissions because the GDP is subject to heterogeneity (by  
11 the different more difficult to cover with the various inhomogeneous economic activities), to  
12 heteroskedasticity (which are also influenced by time-dependent inflation and currency  
13 exchange rates) and to incompleteness (by the not officially reported which are incomplete  
14 with the unrecorded unofficial activities). It is not recommended to use this per unit of GDP  
15 emissions indicator only for comparing levels because the correlation between emissions and  
16 GDP can be for relative small for countries with a substantially contributing service sector  
17 (e.g. Luxembourg).

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18 For 2010 Fig. 3b shows that EU and USA have similar low emissions per unit of GDP for all  
19 substances, except NO<sub>x</sub> where EU's emission per unit of GDP is still significantly lower than  
20 in USA. China's emissions of SO<sub>2</sub> and NO<sub>x</sub> per unit of GDP are at the high end, whereas for  
21 NH<sub>3</sub> and the carbonaceous particulate matter China is bypassed by India, which shows even  
22 higher emissions per unit of GDP. In analogy with Table 2b, Table 2c provides for the world  
23 top 6 CO<sub>2</sub> emitters a comparison of the air pollutants per unit of GDP, which are linked to the  
24 country's economic activity (in GDP per capita) and CO<sub>2</sub> per unit of GDP (measuring the  
25 energy intensive industry). It is directly apparent that again Germany and Japan are having  
26 high economic activity, with still important energy intensive industry but low air pollutant  
27 emissions per unit of GDP because of the investment in clean technology. On the other side,  
28 India has still much lower economic activity but nevertheless a much higher particulate matter  
29 emission per unit of GDP

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30

### 1 3.4 Implied emission factors

2 Energy-intensity is a widely used indicator to assess the fuel efficiency of manufacturing  
 3 processes. Analogous to energy-intensity, we analyse in this section air pollution emission-  
 4 intensity for all world countries. Emission intensity of economic activities for a given region  
 5 are determined by- implied emission factors. **The region-specific implied emission factors**  
 6 **(EF) present the emissions per unit of activity (per TJ energy consumed for all combustion-**  
 7 **related activities inclusive, per kg product for industrial processes and solvents or per 1000**  
 8 **head of animalsha cultivated land for agricultural related activities) and are defined for a**  
 9 **substance x with speciation l at year time t due to activities AD in activity subsectors j,k of**  
 10 **each of the main HTAP sectors (htap\_3 ENERGY, htap\_4 INDUSTRY,**  
 11 **htap\_5 TRANSPORT, htap\_6 RESIDENTIAL, htap\_8 AGRICULTURE)with technologies**  
 12 **j and end-of pipe measures k, in a country C as:**

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$$13 \quad EF_{C,i}(t, x_l) = \frac{\sum_{j,k} [AD_{C,i}(t) * TECH_{C,i,j}(t) * EOP_{C,i,j,k}(t) * EF_{C,i,j}(t, x_l) * (1 - RED_{C,i,j,k}(t, x_l)) * f_{C,i,j}(x_l)]}{\sum_{j,k} [AD_{C,i}(t) * TECH_{C,i,j}(t) * EOP_{C,i,j,k}(t)]}$$

16 with TECH representing the technologies, EOP the end-of pipe measures, EF the emission  
 17 factors and RED the emission reduction due to end-of pipe control measures.

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$$18 \quad EF_{C,3\_energy}(t, x) = \frac{\sum_{\text{subsector } j} EM_{C,3\_energy,j}(t, x)}{\sum_{\text{subsector } j} AD_{C,3\_energy,j}(t)} \Bigg|_{\substack{\text{datasource of } C \\ \text{EDGARv4.3}}} \quad [kton/TJ] \quad (1)$$

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$$19 \quad EF_{C,4\_ind.}(t, x) = \frac{\left[ \sum_{\text{comb. subsector } j} EM_{C,4\_ind.,j}(t, x) + \sum_{\text{proc. subsector } k} EM_{C,4\_ind.,k}(t, x) \right]}{\sum_{\text{comb. subsector } j} AD_{C,4\_ind.,j}(t)} \Bigg|_{\substack{\text{datasource of } C \\ \text{EDGARv4.3}}} \quad [kton/TJ] \quad (2)$$

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$$20 \quad EF_{C,5\_transport}(t, x) = \frac{\sum_{\text{subsector } j} EM_{C,5\_transport,j}(t, x)}{\sum_{\text{subsector } j} AD_{C,5\_transport,j}(t)} \Bigg|_{\substack{\text{datasource of } C \\ \text{EDGARv4.3}}} \quad [kton/TJ] \quad (3)$$

$$EF_{C,6\_res.}(t, x) = \frac{\left[ \sum_{comb. \ sub \ sector \ j} EM_{C,6\_res.,j}(t, x) + \sum_{waste \ prod. \ sub \ sector \ k} EM_{C,6\_res.,k}(t, x) \right]_{datasource \ of \ C}}{\sum_{comb. \ sub \ sector \ j} AD_{C,6\_res.,j}(t)}_{EDGARv4.3} \quad [kton/TJ] \quad (4)$$

$$EF_{C,8\_agr.}(t, x) = \frac{\left[ \sum_{animal \ sub \ sector \ j} EM_{C,8\_agr.,j}(t, x) + \sum_{crop \ sub \ sector \ k} EM_{C,8\_agr.,k}(t, x) \right]_{datasource \ of \ C}}{\sum_{animal \ sub \ sector \ j} AD_{C,8\_agr.,j}(t)}_{EDGARv4.3} \quad [ton/head] \quad (5)$$

It should be noted that the implied emission factors of sectors htap 4 INDUSTRY and htap 8 AGRICULTURE are slightly skewed up because of an incomplete accounting of the activity data which are for these sectors a combination of activities of different nature and as such expressed with different units. The emissions of sector htap 4 INDUSTRY mainly originate from the energy-intensive subsectors and therefore are weighted with the energy needs (in TJ). We omitted the accounting of industrial process emissions, which are calculated per kton product manufactured. In sector htap 6 RESIDENTIAL, the waste is included, although calculated per kton dry or wet waste, which we could not combine with the residential energy consumption in TJ. The emissions of the htap 8 AGRICULTURE sector are weighted with the number of animals and not with the kton crops cultivated, because the crops serve for 85% as animal food and are therefore considered a justified measure of agricultural activity.

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Thereto, emissions of sector-specific gridmaps for 2010 have been aggregated to country level and divided with the activity data for that sector in that country from EDGARv4.3, which are for energy-related activities based on IEA (2013) statistics and for agricultural-related activities on FAO (2012) statistics. It should be noted that emissions in particularly those reported under country-specific point sources are allocated to the reporting country solely, also for cells covering country borders. The areal fraction of these cells would incorrectly spread the emissions also to the neighbouring country, which yield in the case of e.g. the power emissions for Canada up to 30% increase with the USA emissions along its borders. The aggregation of the country cells, taking into account the relative areal fraction of that country in cross border cells, needed to be corrected with country specific reporting, in order to allocate point sources (e.g. power plants) at borders (e.g. waterways) to the responsible country. The implied emission factor results are given for all world countries and for 2010 in the Table SA4 of Annex IV in the Supplement.

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1 Fig. 4 gives an overview per sector of the range of different implied emission factors for each  
2 country with the maximum/minimum, the percentiles and the median. In addition the position  
3 in this range of EU27, USA, China and India is indicated to evaluate the level of emission-  
4 intensity of the different activities. EU 27 and USA show very similar implied emission  
5 factors for the energy and industry sectors, which are much lower than the median for all  
6 pollutants. China also shows implied emission factors for energy and industry that are lower  
7 than the medians, but still larger than USA and EU 27. India shows much higher implied  
8 emission factors for energy and industry, which are for CO, PM2.5, BC, and OC above the  
9 median. In the case of the residential sector, the range of variation of the implied emission  
10 factors is the smallest for SO2 and NOx, but the largest for PM2.5 and BC. For the transport  
11 sector a relatively large variation is present for CO, with an implied emission factor for China  
12 that is above the median. For agriculture it is remarkable that China and India, ~~but~~ ~~always well~~  
13 ~~in~~ the USA and EU 27, have implied emission factors that are above the median, with China  
14 reaching the maximum compared to all other world countries.

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15 Even though only implied emissions factors for country emissions are presented in Fig. 3b,  
16 the implied emission factors were also calculated for the international bunker fuel and  
17 indicated that the implied emission factors are at the high end of the range for SO2 (0.98 ton  
18 SO2/TJ similar to the road transport emission factor of Laos or Panama), NOx (with 1.65 ton  
19 NOx/TJ similar as for transport in Bangladesh or Myanmar), PM2.5 (with 0.17 ton PM2.5/TJ  
20 similar as for transport in China), but are relatively low for CO, NMVOC and BC. The high  
21 SO2 implied emission factor might indicate the use of lower quality fuels in sea  
22 transportation, especially in international waters. The high SO2 implied emission factor (from  
23 EDGARv4.3) represents the use of lower quality fuels in sea transportation, especially in  
24 international waters: -85% of the sea bunker fuel in 2010 consists of residual fuel oil with an  
25 emission factor of 1.29 ton SO2 /TJ.

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### 27 3.5 Emission changes 2008-2010

28 The emission change from 2008 to 2010 is given in Table SA2.3 of Annex II. It should be  
29 noted that the data provided for Canada by US-EPA/Environment Canada and for Europe by  
30 TNO were actually not representing 2010, but 2008 respectively 2009. However updates were  
31 undertaken: point source data of 2010 were used and implemented in the gridmaps. Both  
32 regions were affected by the economic crisis of 2008, yielding stagnation and even

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1 downwards trends in the following years, mainly in the energy and industry sectors. The latter  
2 sectors are primarily composed of point sources, and therefore, the gridmaps of 2010 can be  
3 considered to represent also for Canada and Europe the actual 2010 situation. For the

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4 developed countries in North America and Europe the decline of emissions between 2008 and  
5 2010 for most of the pollutants are driven mostly by continued implementation of emission  
6 reduction technologies. In some cases this also leads to increases in sectorial emissions,  
7 although insignificant for the total, as is estimated for NH<sub>3</sub> in the energy and transport  
8 sectors, due to the use of catalysts.

9 For the MICS-Asia region, the emissions are mostly increasing except for the energy sector,  
10 where the SO<sub>2</sub> and PM emissions are reduced in 2010 due to the wide deployment of flue-gas  
11 desulfurization (FGD) and particulate matter filters in the power plants, consistent with Wang  
12 et al. (2014). For the other developing countries (calculated with the EDGARv4.3 data and  
13 based on the IEA(2013) fuel statistics), the SO<sub>2</sub> emissions of the energy sector slightly  
14 increase from 2008 to 2010 in the energy sector, possibly due to the impact because of the  
15 increased coal use (as also observed by Weng et al., 2012) and the increased use of even  
16 heavy fuel oil (in the Middle East power sector according to IEA (2013) activity data). The  
17 PM emissions from the energy and industry of some the other developing countries show a  
18 decrease from 2008 to 2010, mainly due to the activity reduction and but also in some cases  
19 due to the modelled decrease in controlled emission factor in EDGARv4.3. Largest reductions  
20 were seen for Brazil (with 54% reduction of its 2008 charcoal production) and Kazakhstan  
21 (11% reduction in coal power generation, which is modelled with a 31% decreasing BC  
22 emission factor) indicating slow penetration of end-of-pipe abatement.

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### 24 3.6 Qualitative assessment of the uncertainty of emission gridmaps

25 Even though the HTAP\_v2.2 data sources are all bottom-up constructed inventories, they  
26 differ considerably in e.g. the assumptions taken on the modelling of technology and end-or-  
27 pipe measures and use different emission factors, which quite different, and lead to  
28 inconsistencies at the borders between two adjacent inventories. On their turn the different  
29 bottom-up inventories are constructed with sub-regional (country, state, county or province  
30 level) activity data and emission factors. As such, inconsistencies can be expected at each  
31 country border and the variation of the emissions at cross-border cells gives already a first  
32 indication on the region- and sector-specific emission uncertainty over borders, a bottom-up

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1 ~~methodology with activity data and emission factors is applied to calculate emission totals~~  
2 ~~and distribute these on the grid.~~ The propagation of uncertainty is given by the effect of  
3 variables' uncertainties (or errors) on the uncertainty, i.e. the variance of the activity data and  
4 that of the emission factor. Table 3 provides some insight in the estimation of the uncertainty  
5 range, however the approach followed in HTAP v2.2 inhibits an overall consistent uncertainty  
6 assessment because it is not one single bottom-up inventory.

7 Guidance on evaluation of emission uncertainties can be obtained from the evaluations of the  
8 national inventories reported to UNFCCC, addressed by e.g. Jonas et al (2010) (and  
9 references in there). With the evaluation of common behaviours between species in  
10 EDGARv4.2 of Balsama et al (2014) we propose the same approach of CO2 uncertainty  
11 assessment for SO2 and NOx (all driven by combustion-related activities), and the approach  
12 of N2O for NH3. As such Table 3 follows the grouping of countries by Andres et al (2012)  
13 and Marland et al (1999), based on their statistical infrastructure. Countries with well  
14 maintained statistical infrastructure are the 24 OECD-1990 countries plus India with a British  
15 statistical accounting system. For the other countries, a larger range in uncertainty is present,  
16 for which we refer to Gregg et al. (2008) or Tu (2011) and Olivier (2002). For the annual CO2  
17 inventory, the biofuel is carbon-neutral and not taken up in the national inventories. However,  
18 for the air pollutants it is an additional large source of uncertainty, which is often not  
19 officially reported and as such missing. For the N-related emissions, the division in countries  
20 could be based on common agricultural practices (Leip et al, 2011 and Rufino et al, 2014).

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21 In addition to the uncertainty of the activities, the quality and representativeness of the  
22 controlled emission factors play a crucial role. The standard range of uncertainty already  
23 varies according to the EMEP/EEA (2013) Guidebook's Uncertainties Chapter 5 for the  
24 absolute annual total of different pollutants between at least 10% for SO2, at least 20% for  
25 NOx and CO, at least 50% for NMVOC, an order of magnitude for NH3, and PM10, PM2.5,  
26 BC and OC. These considerations have been taken into account to indicate qualitatively a  
27 range for the different uncertainties (using the terminology low (L), low medium (LM), upper  
28 medium (UM) or high (H)) for the different sectors and species.

29 The HTAP modelling community is expected to run in addition to the actual 2008 and 2010  
30 simulations with the HTAP v2.2 emission inventory also the emission scenarios of  
31 ECLIPSEv5 We can only compare the HTAP v2.2 with the ECLIPSEv5 dataset of (Klimont  
32 et al., (in preparation 2015). ECLIPSEv5 starts with a 2010 emission inventory (or, which is a

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1 ~~fully consistently built global bottom up inventory and serves as base year inventory~~, that  
2 serves also as reference point for all projections. Here we compare the ECLIPSEv5 emission  
3 inventory for 2010 with the HTAP v2.2 2010 data, in order to evaluate how close the  
4 reference point is to the “officially accepted” regional inventories of HTAP v2.2. ~~for the~~  
5 ~~HTAP scenarios~~. At global level, a relatively good agreement is found with small relative  
6 emission differences  $(ECLIPSEv5 - HTAPv2.2) / HTAPv2.2$  for the aggregated sectors in  
7 2010. ~~It should be noted that the GAINS dataset, another bottom inventory, can not be~~  
8 ~~considered an external independent source of verification, because similar information on~~  
9 ~~emission factors and reductions for certain technologies have been applied in the TNO-~~  
10 ~~EMEP, MIX-Asia and EDGARv4.3 datasets~~. The relative difference for NO<sub>x</sub> and CO is only  
11 -4% respectively +5%. For SO<sub>2</sub> a larger difference of -8% reflects the recent important S-  
12 reductions for the non-ferrous metal smelters in ECLIPSEv5 (Klimont et al., 2013). For NH<sub>3</sub>  
13 a relative difference of +17% is acceptable because of the larger uncertainty in emission  
14 factors driven by lack of information about manure management practices and also by  
15 incomplete data on the agricultural activities. For NMVOC a difference of -27% stems  
16 primarily from the assumptions about emissions from solvent use. The information about  
17 activity levels is scarce and even less is known about the emission factors for some important  
18 sources. Both regional inventory compilers and modellers often make assumptions about per  
19 capita or per GDP solvent use NMVOC emissions from particular sectors. Here assumptions  
20 employed in the ECLIPSEv5 lead to lower emissions from these activities. As anticipated  
21 (and reflected in Table 3) larger differences of 48% and 29% are present for PM<sub>2.5</sub> and BC,  
22 respectively. While for PM<sub>2.5</sub>, assumptions about penetration and efficiency of filters in  
23 industrial and small-scale residential boilers as well as emission factors and activity data for  
24 biomass used in cooking stoves play a key role, for BC assumptions about coal consumption  
25 in East Asia are of relevance since ECLIPSEv5 relied on provincial statistics for China which  
26 results in higher coal consumption than reported in national statistics and IEA. Additionally,  
27 ECLIPSEv5 includes emissions from kerosene wick lamps, especially relevant for South Asia  
28 and parts of Africa according to Lam et al. (2012), gas flaring and high emitting vehicles,  
29 which together result in about 30% higher emissions.

30 In addition, the spatial allocation is subject to other types of errors, with a spatial variance for  
31 point sources and a more important systematic error when a spatial proxy is used to distribute  
32 the emissions. Geo-spatial consistency is lower in the HTAP\_v2.2 database than if the  
33 national totals would have been spatially redistributed with one harmonised spatial proxy

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1 dataset. It should be also noted that derivation of country totals from the 0.1°x0.1° emission  
2 gridmaps (as e.g. done in the ECCAD system) is only valid if the country-specific total is  
3 larger than 0.2% of each of the totals of the neighbouring countries. Otherwise the derived  
4 country-specific sector total can be 50% larger than the bottom-up one, mainly in the energy  
5 sector with many point sources which are typically located on waterways or coastal areas and  
6 as such in cross border cells. Table SA1.3 illustrates the deviations of derived country-  
7 specific sector totals to the bottom-up ones for the Asian region. The latter caused derived  
8 sector totals for Kyrgyzstan, Tajikistan, Afghanistan, Laos, Myanmar, Bangladesh, which  
9 deviated with one order of magnitude from the bottom-up totals. However, the relative small  
10 differences for China (<5%), India (<3% for all except for SO2 from energy where it is 14%),  
11 Indonesia (<7%) and Thailand (<12.5%), Japan (<16.0%) and South Korea (<17.3%) show a  
12 good agreement for the top 6 Asian emitters.

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13 Another type of inconsistency in mass balance at grid cell level occurs when for the same  
14 region the data sources providing the emission gridmaps for PM10 and PM2.5 or for PM2.5  
15 and BC/OC are different. Already the application of different spatial proxy datasets (e.g. with  
16 and without point sources) result in an inconsistent allocation of multi-pollutant sources to  
17 different grid cells. speeiation of a substance is done with gridmaps of different data sources.

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18 This was another reason not to use the PM gridmaps of EMEP, as no BC and OC speciation is  
19 available from the same EMEP data source. Instead we used the gridmaps of TNO for all PM  
20 components (PM10 and PM2.5) and the TNO speciation file for BC and OC. In addition a  
21 check was performed to ensure that the sum of BC and OC emissions in every grid cell is  
22 smaller than the PM2.5 emission in that grid cell. Thereto a re-allocation of the emissions of  
23 some point sources (industrial facilities) was needed within Europe (e.g. Poland) and  
24 performed in consultation with TNO.

25 Another check was to estimate per grid cell the change in emission from 2008 to 2010 and  
26 allowed to find missing sources. However, Even though this mosaic inventory can not present  
27 the same global consistency as one global bottom-up inventory, its extensive evaluation and  
28 use helped improving its quality. cannot be guaranteed and a comparison of different  
29 countries or of different years cannot be conclusive. The evaluation was undertaken in  
30 particular in discussion with TNO and with US EPA to identify missing sources or  
31 misallocation of point sources. In particular point sources are very important input, but their  
32 strengths and locations are subject to input errors with larger consequences and cannot be

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1 extrapolated in time. (Closure of power plants as large point sources can change the emission  
2 distribution pattern from one year to another.) In addition the use of the dataset by global and  
3 regional climate and air quality modellers and the modellers' feedback from users of the  
4 emission dataset has already helped to improve its quality (personal communications with L.  
5 Emmons of 5 November 2013 and D. Henze of 19 November 2013) were most useful and  
6 isare further encouraged.

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#### 8 **4 Conclusions and recommendations**

9 This paper describes the HTAP global air pollutant baseline reference emission inventory for  
10 2010, which is composed of latest available data from regional inventory compilers also  
11 regionally accepted as reference. It assures a consistent input for both regional and global  
12 modelling as required by the HTAP modelling exercise. The HTAP\_v2.2 emission database  
13 makes use of consolidated estimates of official and latest available regional information with  
14 air pollutant gridmaps from US EPA and EnvironCanada for North America, EMEP-TNO for  
15 Europe, MIX for Asia, and the EDGARv4.3 database for the rest of the world. The mosaic of  
16 gridmaps provides comprehensive local information on the emission of air pollutants, because  
17 it results from the collection of point sources and national emission gridmaps at 0.1° (for  
18 some regions 0.25°) resolution. Even though the HTAP\_v2.2 dataset is not a self-consistent  
19 bottom-up database, with activity data of consistent defined according to international  
20 statistics standards, with harmonized emission factors, and with global sets of spatial emissions  
21 gridded with global proxy data, it provides a unique set of emission gridmaps with global  
22 coverage and high spatial resolution, including in particular important point sources. The  
23 compilation of implied emission factors and per capita emissions for the different world  
24 regions using multiple sources provides the regional and national emission inventory  
25 compilers with a valuable asset for comparison with their own data for cross checking and  
26 analysis which may lead to identification of future improvement options.

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27 This dataset was prepared as emission input for the HTAP community of modellers and its  
28 preparation has involved outreach to global and regional climate and air quality modellers  
29 (collaborating also within the AQMEII and MICS-Asia modelling exercises). The TF HTAP  
30 needed an emission inventory that was suitable for simultaneous and comparable modelling of  
31 air quality at the regional scale and at the global scale to deliver consistent policy support at  
32 both scales. The HTAP-v2.2 emission inventory presented in this paper is tailor-made to

1 allow the TF HTAP to fulfil its prime objectives and contribute to a common international  
2 understanding of global and regional air pollution and its influence on human health,  
3 vegetation and climate. The use of the HTAPv2.2 inventory will substantially help to provide  
4 a basis for future international policies because it combines and is consistent with the  
5 inventories that are used for regional (EU, US Canada, China) policy analysis and support.

6

#### 7 **Access to the data**

8 The 0.1° x 0.1° emission gridmaps can be downloaded from the EDGAR website on  
9 [http://edgar.jrc.ec.europa.eu/htap\\_v2/index.php?SECURE=123](http://edgar.jrc.ec.europa.eu/htap_v2/index.php?SECURE=123) per year, per substance and  
10 per sector either in the format of netcdf-files or .txt files. The emissions in the netcdf-files are  
11 expressed in kg substance/m<sup>2</sup>/s but the emissions in the .txt are in ton substance / gridcell. For  
12 the NMVOC speciated gridmaps we refer to the link on the ECCAD data portal:  
13 <http://eccad2.sedoo.fr/eccad2/mapdisplay.xhtml?faces-redirect=true>.

14

15

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8

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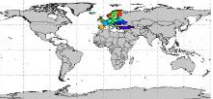


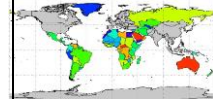
1 **Tables**

2

3 **Table 1a-- Overview of the data sources and their generic characteristics, as used for the**  
 4 **different regions in HTAP\_v2.2**

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Data source	EMEP-TNO (MACCII)	US EPA _ Environ Can	MICS-Asia (+ REAS2.1)	EDGARv4.3 (prelim.)
Type of data source	Country inventories + point sources	State inventories + point sources	County inventory for China + country inventories from CAPSS & REAS 2.1	Country inventories from the preliminary version of EDGARv4.3
Coverage of human activities	All except international shipping and except international aviation	All except international shipping and except international aviation	All except international shipping, international aviation and agricultural waste burning	All inclusive international shipping and international aviation
Temporal resolution	Yearly gridmaps (monthly profiles of EMEP model added)	Monthly profiles	Monthly gridmaps	Monthly profiles (for 3 different latitude bands)
Spatial resolution	0.125° x 0.0625° converted to 0.1°x0.1° by raster resampling with factor 1/5x1/5 and aggregation of 4x4	0.1° x 0.1° and height profiles	0.25° x 0.25° converted to 0.1° x 0.1° by raster resampling 1/5x1/5 and aggregation of 2x2	0.1° x 0.1°
Substances	CO, NMVOC, NOx, SO2, NH3, PM coarse and fine and BC/OC fractions	CO, NMVOC with speciation, NOx, SO2, NH3, PM10, PMfine, BC and OC	CO, NMVOC, NOx, SO2, NH3, PM10, PM2.5, BC and OC	CO, NMVOC, NOx, SO2, NH3, PM10, PM2.5, BC and OC
Geocoverage used in HTAP_v2.2				

5

6

7 **Table 1b-- Sectors in the HTAP\_v2.2 inventory (only anthropogenic sources are included)**  
 8 **and the corresponding Nomenclature for Reporting (NFR) and the Selected Nomenclature for**  
 9 **Sources of Air Pollution (SNAP) codes as spelled out in the EMEP (2002) Reporting**  
 10 **Guidelines.**

Tag	Description	IPCC level (NFR code)	EMEP SNAP code
htap_1_AIR/craft	International and domestic aviation	1.A.3a(i)+(ii)	S8(*)
htap_2_SHIPS/International Shipping	International shipping	1.A.3d(ii)	
htap_3_ENERGY/Power industry	Power generation	1.A.1a	S1
htap_4_INDUSTRY/Industry	industrial non-power but large-scale combustion emissions and emissions of industrial processes (**) and product use	1.A.1b+c, 1.A.2, 1.B.1+2, 2.A+B+C+D+G, 3	S3 + S4 + S5 + S6 (***)

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inclusive solvents.

<p>htap_5_TRANSPORT transport</p>	<p>Ground t Ground t</p>	<p>Transport by road, railway, inland waterways, pipeline and other ground transport of mobile machinery (#). Htap_5 does not include re-suspended dust from pavements or tyre and brake wear.</p>	<p>1.A.3b+c+d(ii)+e  6.A+B+C+D</p>	<p>S71 + S72 + S73 + S74 + S75 + S8 (##)</p>
---------------------------------------	------------------------------	--	--	--

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<p>htap_6_RESIDENTIAL Residential</p>	<p>Small-scale combustion, including heating, cooling, lighting, cooking and auxiliary engines to equip (###) residential, commercial buildings, service institutes, and agricultural facilities and fisheries; solid waste (landfills/ incineration) and wastewater treatment.</p>	<p>1.A.4+5 6.A+B+C+D</p>	<p>S2 + S9</p>
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<p>htap_8_AGRICULTURE Agriculture</p>	<p>Agricultural emissions from livestock, crop cultivation but not from agricultural waste burning and not including Savannah burning</p>	<p>4.A+B+C+D</p>	<p>S10</p>
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Notes: (\*) S8 (point source) includes local emissions of aircrafts around the airport only below 3000ft.

(\*\*) Product testing by the manufacturer inside is not considered an emission of the building (htap\_6) but taken up under the industry (htap\_4). The oil production sector is completely covered in htap\_6 and includes the fugitive (evaporative) emissions (mainly NMVOC) during the oil & gas exploration and production and transmission. As such, there are NMVOC emissions along the oil tanker tracks visible under the htap\_4 sector).

(\*\*\*) Note that S34=S3+ S4 in the TNO-MACC-II inventory (Kuenen et al., 2014). Fuel transformation processes (and refineries) are included here.

(#) The pipeline transport does not include transmission of natural gas and crude oil, because the latter is included in the oil and gas production industry under htap\_4 but it does include the transport of refined products (motorgasoline, diesel, liquefied petroleum gas) or goods. The other ground transport includes all mobile (non-stationary) machinery (as used in the agriculture, forestry or construction sector).

(##) For the split-up of SNAP7 into S71 S72, S73, S74 and S75 we refer to the definitions used for the TNO-MACCII inventory documented in (Kuenen et al., 2014)

(###) In particular industrial, commercial and/or agricultural buildings can be more extensively equipped with auxiliary stationary (non-mobile) infrastructure in and around the building (e.g. lifting devices).

Table 2a - Comparison of the international shipping emissions: IMO Bottom up (BU) and IMO Top Down (TD) emissions of the IMO(2014) study and the EDGAR emissions of the HTAP v2.2 (2015) study

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kton /yr	BC	CO	NMVOC	NOx	OC	PM10	PM2.5	SO2
EDGAR 2008	34	1340	730	13762	458	1376	1376	8348
IMO BU 2008		864	727	20759		1545	1545	11041
IMO TD 2008		553	615	18442		1221	1221	8280
EDGAR 2010	33	1300	720	14000	430	1400	1400	8300
IMO BU 2010		763	593	16708		1332	1332	9895
IMO TD 2010		574	638	19098		1304	1304	9232

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Table 2b - Comparison of per capita emissions in 2010 for USA, Germany, and China, India, Russia and Japan from HTAP\_v2.2

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Substance	USA	Germany	China
kg SOx/yr/cap	32.6	5.2	20.9
kg NOx/yr/cap	43.6	14.2	20.8
kg VOC/yr/cap	43.1	11.9	16.7
kg CO/yr/cap	148.3	35.6	125.6
kg NH3/yr/cap	11.6	7.3	6.7
kg PM2.5/yr/cap	5.3	1.1	12.2
kg BC/yr/cap	0.9	0.2	1.3

Substance	USA	Germany	China	India	Russia	Japan
ton CO2(long cycle C) /yr/cap	17.6	9.9	6.4	1.5	11.9	9.7
HDI	0.91	0.9	0.7	0.57	0.77	0.88
kg SOx/yr/cap	32.6	5.2	21	8.0	31.9	5.2
kg NOx/yr/cap	43.6	14.2	20.8	7.9	25.1	14.5
kg VOC/yr/cap	43.1	11.9	16.9	14.0	26.9	9.1
kg CO/yr/cap	148.3	35.6	125.6	56.0	52.8	33.1
kg NH3/yr/cap	11.6	7.3	6.7	8.2	6.3	3.7
kg PM2.5/yr/cap	5.25	1.08	8.93	5.19	2.18	0.62
kg BC/yr/cap	0.95	0.20	1.29	0.85	0.29	0.16

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1 **Table 2c - Comparison of emissions per unit of GDP in 2010 for USA, Germany, China,**  
 2 **India, Russia and Japan from HTAP v2.3**

Substance	USA	Germany	China	India	Russia	Japan
kg CO <sub>2</sub> (long cycle C) /yr/USD	339.71	287.79	240.88	136.6	644.58	267.08
GDP/cap	49307	39668	9230	4638	21663	34561
g SO <sub>x</sub> /yr/USD	0.668	0.132	2.310	1.719	1.482	0.150
g NO <sub>x</sub> /yr/USD	0.892	0.363	2.295	1.714	1.166	0.419
g VOC/yr/USD	0.882	0.305	1.863	3.013	1.249	0.263
g CO/yr/USD	3.036	0.910	13.830	12.069	2.449	0.957
g NH <sub>3</sub> /yr/USDP	0.236	0.187	0.735	1.770	0.291	0.108
g PM <sub>2.5</sub> /yr/USD	0.108	0.028	0.984	1.119	0.101	0.018
g BC/yr/USD	0.019	0.005	0.143	0.183	0.013	0.004

3

1 Table 3. Variables' uncertainties for sector- and country-specific totals per region with  
 2 qualitative classification using the abbreviations Low (L), Low-Medium (LM), Upper-  
 3 Medium, and High (H). The legend provides an interpretation of the level Low, Low-  
 4 Medium, Upper-Medium and High, which is indicatively specified for two groups of  
 5 countries with two different statistical infrastructures.

	SO2	NOx	CO	NM/VOC	NH3	PM	BC/OC	With legend:	
htap1_AIRair	L	LM	LM	UM	LM	UM	UM	countries with well	Countries with
htap2_SHIPShip	L	LM	LM	UM	LM	H	H	maintained	poorly maintained
htap3_ENERGYenergy	L	LM	LM	UM	LM	UM	UM	statistical	statistical
htap4_INDUSTRYindustry	LM	LM	LM	UM	UM	LM	LM	infrastructure	infrastructure
htap5_TRANSPORTground transport	LM	UM	UM	UM	H	H	H	L < 15%	L < 35%
htap6_RESIDENTIALresidential	LM	UM	UM	UM	H	H	H	15% ≤ LM < 50%	35% ≤ LM < 70%
htap8_AGRICULTUREagriculture	UM	UM	UM	UM	H	H	H	50% ≤ UM < 100%	70% ≤ UM < 150%
								100% ≤ H	150% ≤ H

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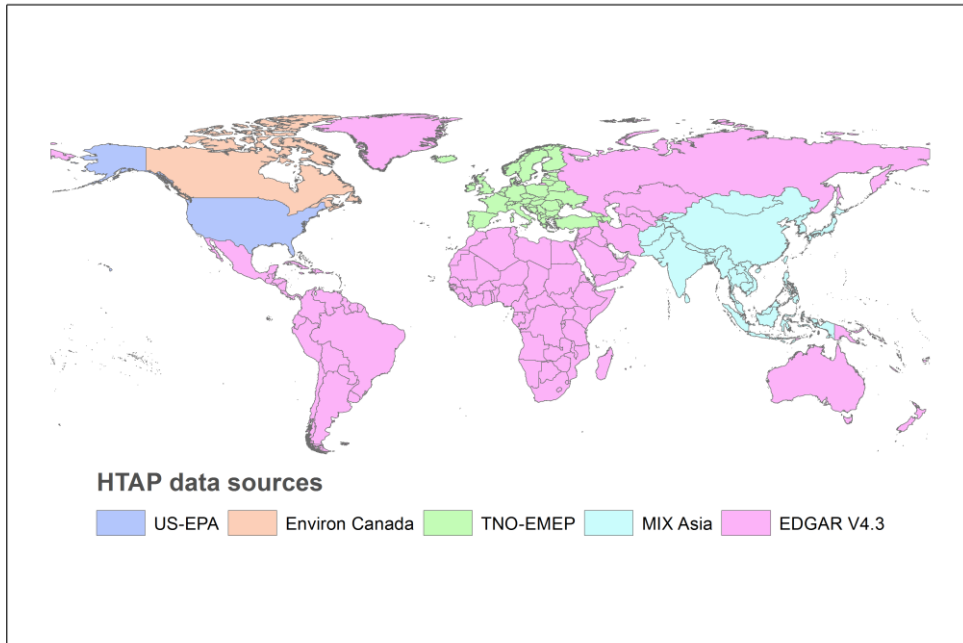
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6 Note: The statistical infrastructure of a country determines the uncertainty of the country's emission inventory. Andres et al. (2012) consider  
 7 under the countries with well maintained statistical infrastructure: the 24 OECD-1990 countries (Australia, Austria, Belgium, Canada,  
 8 Switzerland, Germany, Denmark, Spain, Finland, France, United Kingdom, Greece, Ireland, Iceland, Italy, Japan, Luxembourg, The  
 9 Netherlands, Norway, New Zealand, Portugal, Sweden, Turkey, and the United States) as well as India (using the British statistical  
 10 accounting system according to Marland et al. (1999). For the other countries, a larger range in uncertainty is present, for which we refer to  
 11 Gregg et al. (2008) or Tu (2011) and Olivier (2002). The sector-specific uncertainty of the activity and the quality and representativeness of  
 12 the controlled emission factors play an important role. The standard range of uncertainty already varies according to (The EMEP/EEA (2013)  
 13 Guidebook's Uncertainties Chapter 5 for the absolute annual total of different pollutants between at least 10% for SO2, at least 20% for NOx  
 14 and CO, at least 50% for NMVOC, an order of magnitude for NH3, and PM10, PM2.5, BC and OC. These considerations have been taken  
 15 into account to qualitatively indicate a low (L), low medium (LM), upper medium (UM) or high (H) uncertainty for the different sectors and  
 16 species substances. Countries with well maintained infrastructure are mainly the 24 OECD(1990) countries and India. Other countries are  
 17 considered to have a relative poorly maintained statistical infrastructure. —  
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1 **Figures**

2



3

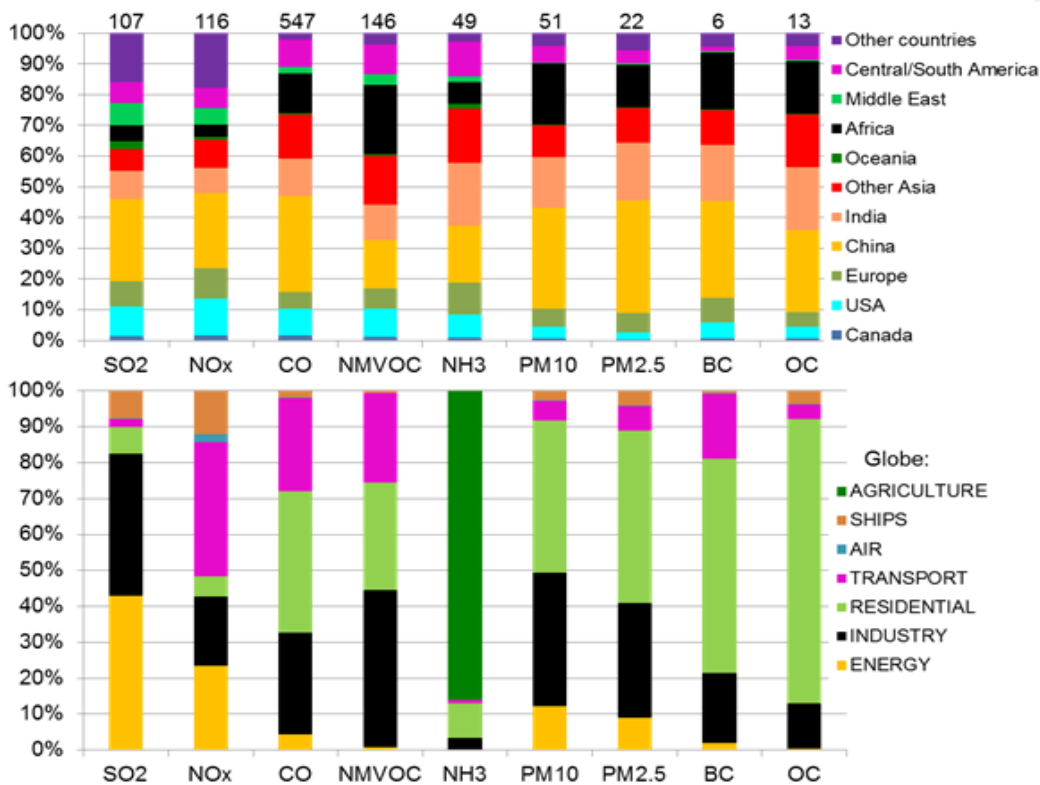
4 **Figure 1a – Collection of regional emission inventories (US-EPA, Environ Canada,**  
5 **TNO-EMEP, MIX/MICS-Asia III), EDGARv4.3 for the global air pollutants and**  
6 **their use for world countries in dataset HTAP v2.2**

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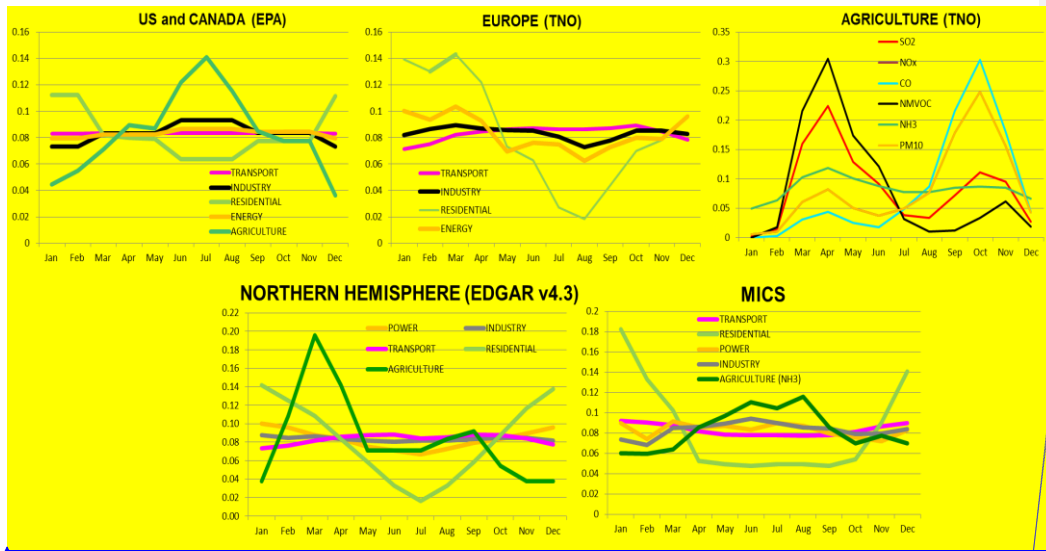
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1  
 2 **Figure 1b - Regional relative contribution to 2010 pollutant emissions (upper panel).**  
 3 **Asian emissions have been divided into China, India and other Asia fractions from the**  
 4 **MIXCS database. The region “rest of the world” has been disaggregated into Oceania,**  
 5 **Africa, Middle East, Central/South America and other countries making use of the**  
 6 **EDGAR v4.3 inventory. Global sector-specific anthropogenic emissions of gaseous**  
 7 **pollutants and particulate matter components for the year 2010 (lower panel). Global**  
 8 **absolute emissions are reported on top of each bar in Tg species per year. Large scale**  
 9 **open-biomass burning is not included in the analysis.**

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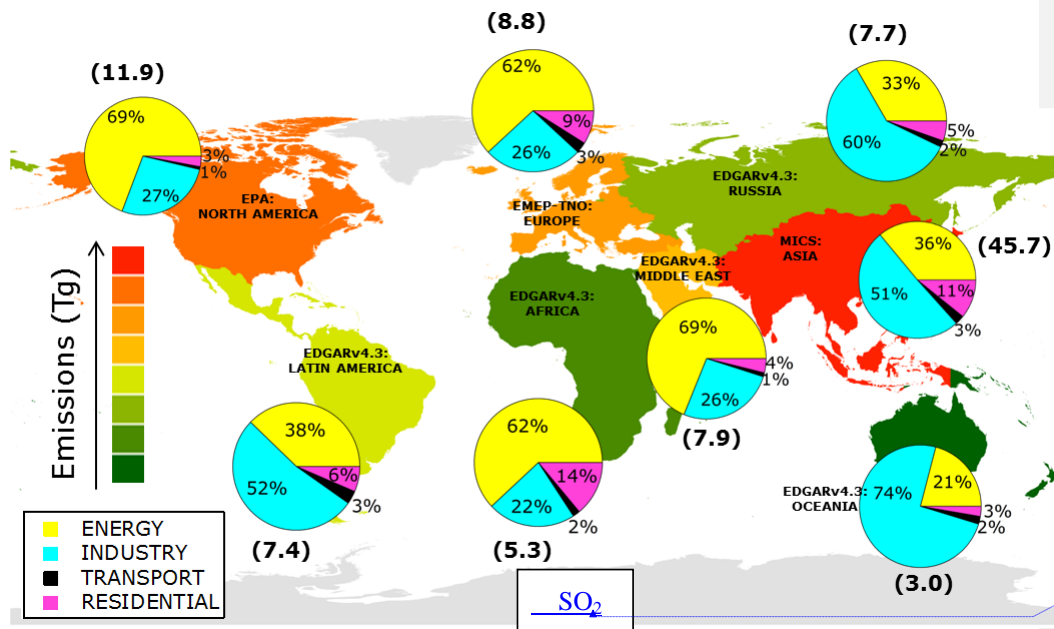
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Figure 1c – Temporal profiles with relative factors varying around 1/12 and applied on the yearly emissions of the different data sources (US–EPA for US and Canada, EMEP-TNO for Europe with compound-specific variation of the agricultural temporal profiles, EDGAR temporal profiles for the Northern hemisphere and MICS profiles for Asia).

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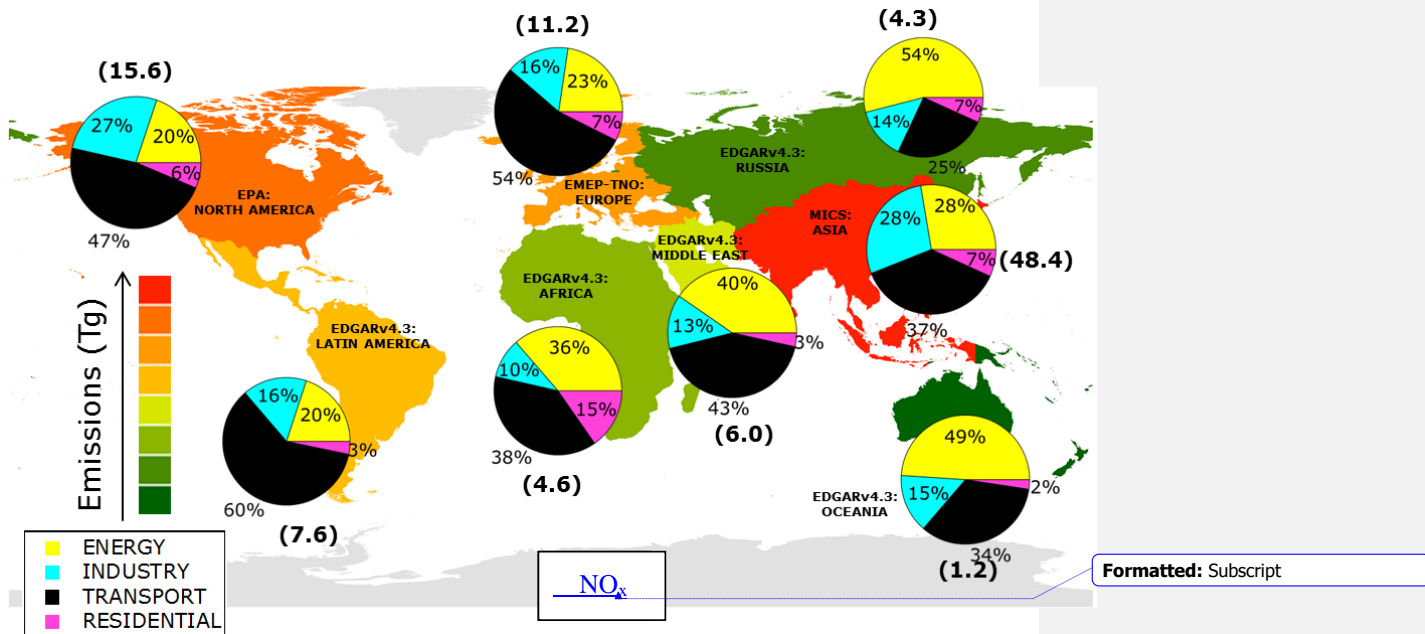
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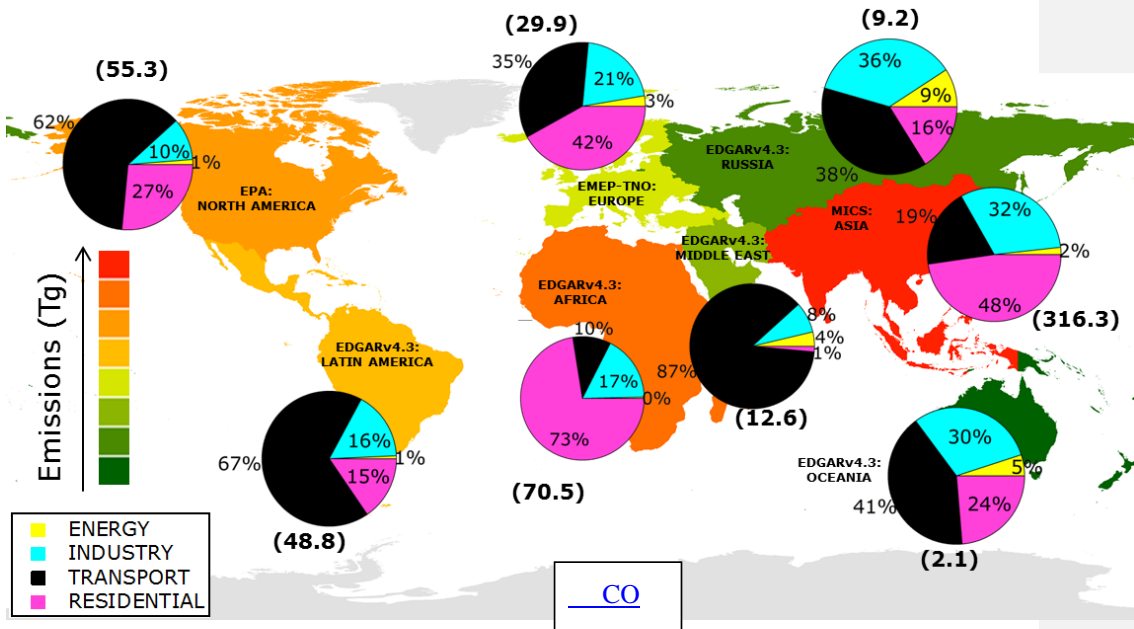


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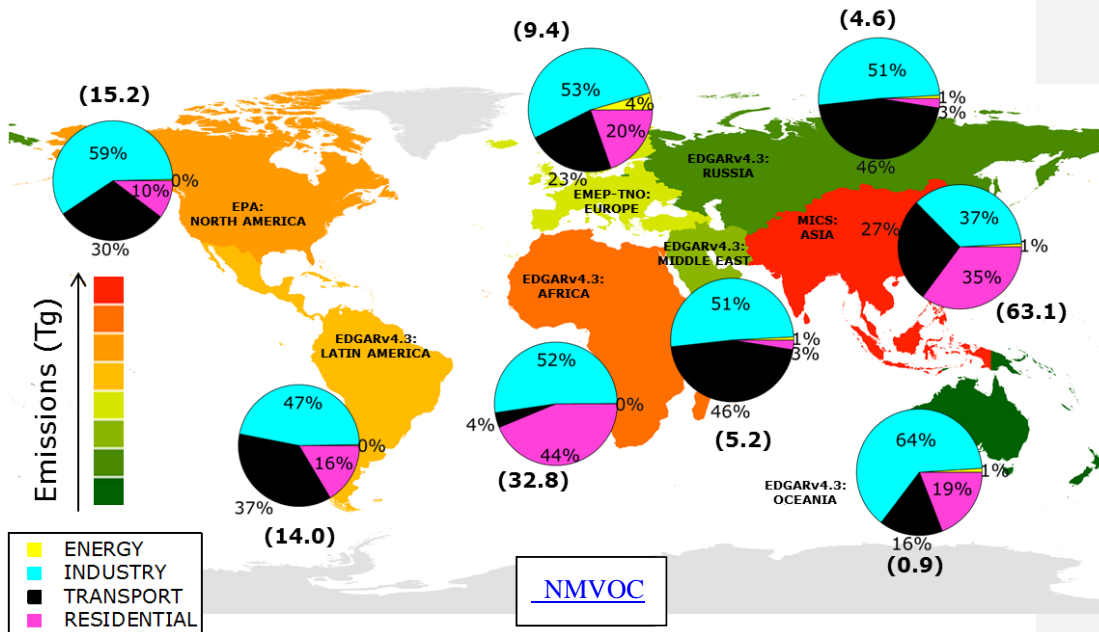
1 **Figure 2a—Total Tg SO<sub>2</sub> emissions for 2010 (in brackets) and sector-specific composition**  
 2 **for world regions.**



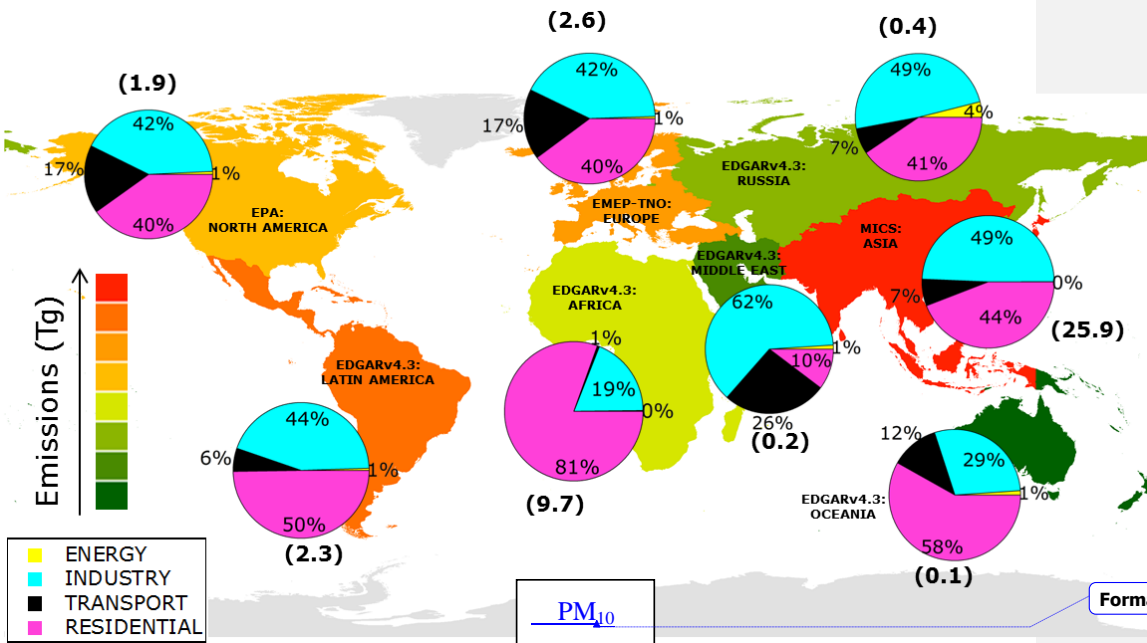
3  
 4 **Figure 2b—Total Tg NO<sub>x</sub> emissions for 2010 (in brackets) and sector-specific**  
 5 **composition for world regions.**



1 **Figure 2c—Total Tg CO emissions for 2010 (in brackets) and sector-specific composition**  
 2 **for world regions.**

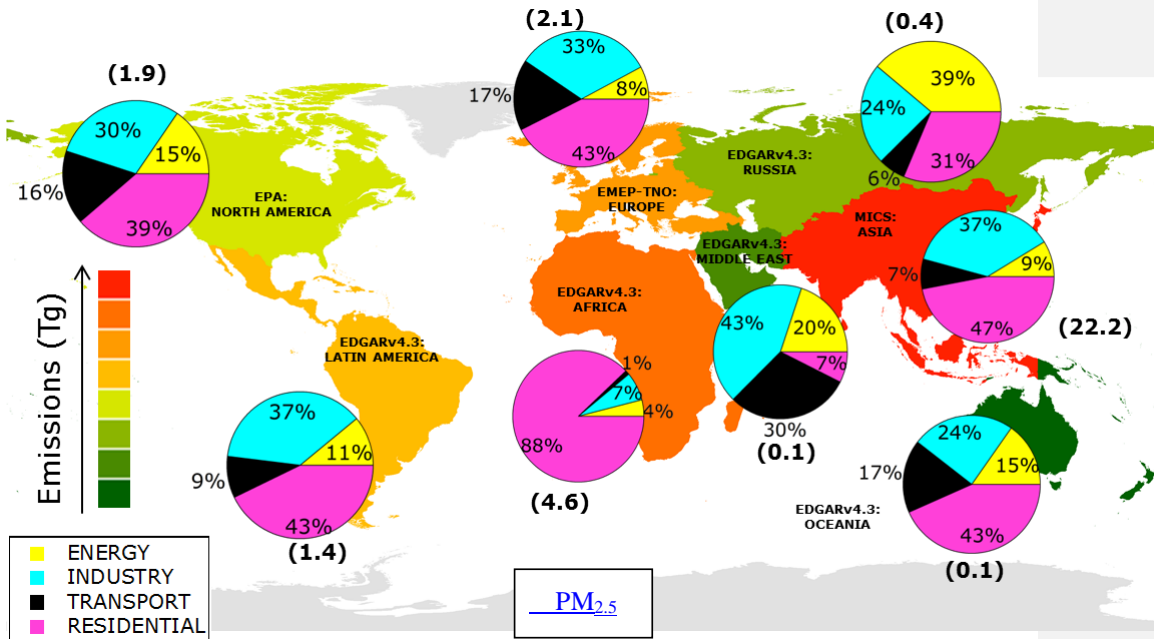


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 4 **Figure 2d—Total Tg NMVOC emissions for 2010 (in brackets) and sector-specific**  
 5 **composition for world regions.**

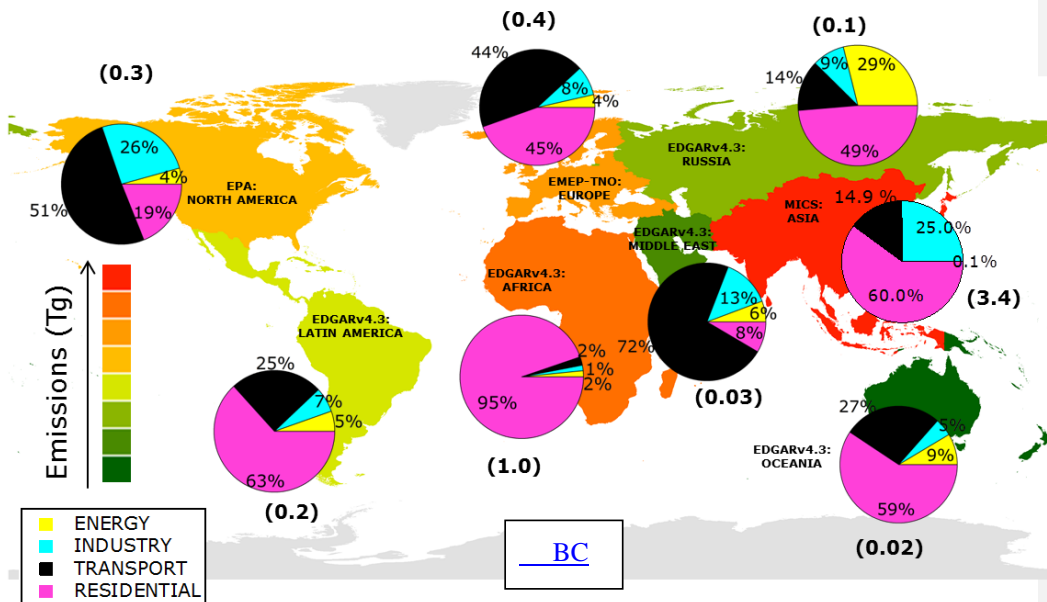


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1 **Figure 2e— Total Tg PM<sub>10</sub> emissions for 2010 (in brackets) and sector-specific**  
 2 **composition for world regions.**

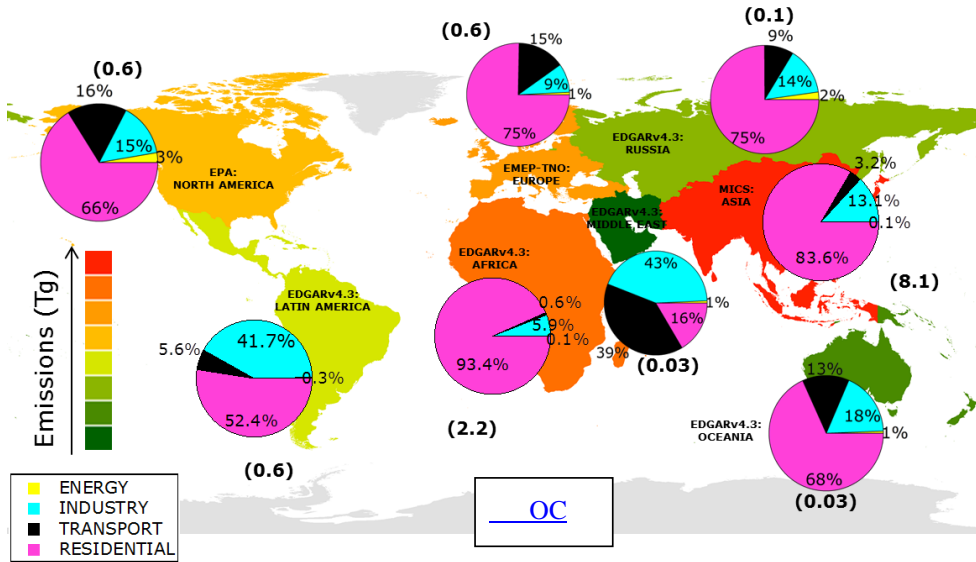


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 4 **Figure 2f— Total Tg PM<sub>2.5</sub> emissions for 2010 (in brackets) and sector-specific**  
 5 **composition for world regions.**

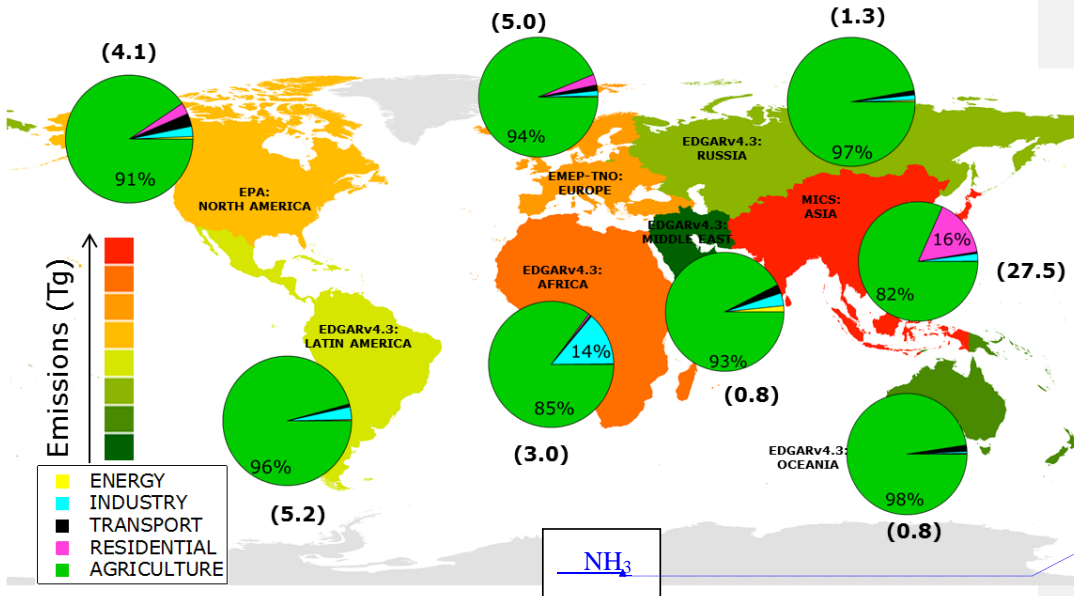




1 **Figure 2g** Total Tg BC emissions for 2010 (in brackets) and sector-specific composition  
 2 for world regions.  
 3

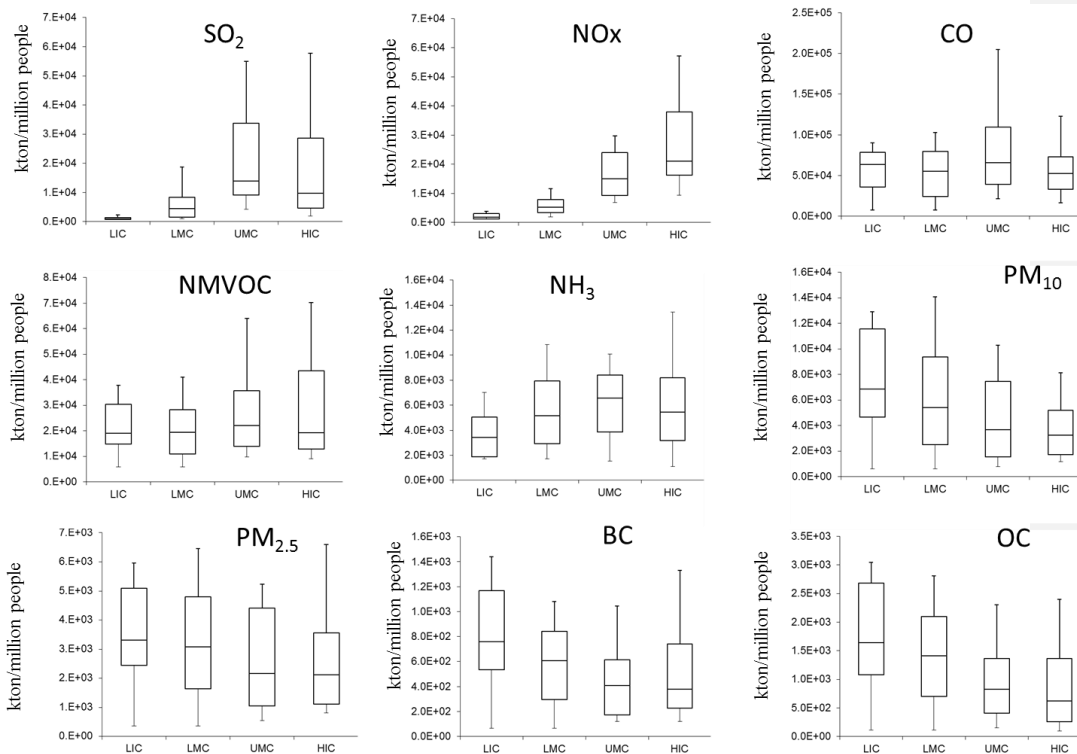


5 **Figure 2h** Total Tg OC emissions for 2010 (in brackets) and sector-specific composition  
 6 for world regions.



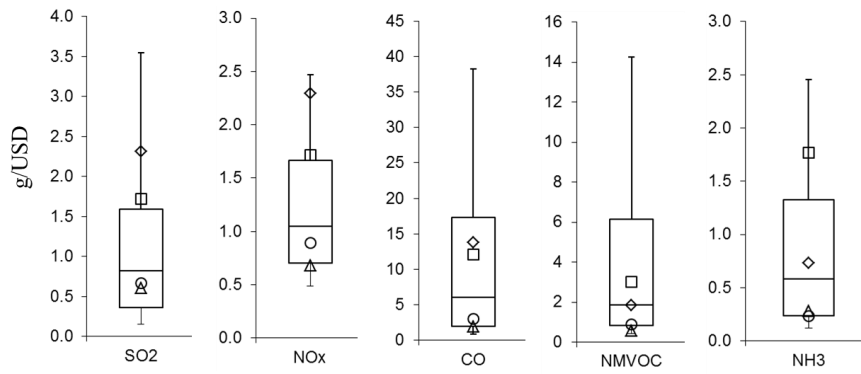
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**Figure 2: Sector-specific breakdown of regional emission totals (Tg) for 2010: SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOC, PM<sub>10</sub>, PM<sub>2.5</sub>, BC, OC and NH<sub>3</sub>—Total Tg NH<sub>3</sub> emissions in 2010 (in brackets) and sector-specific composition for world regions.**

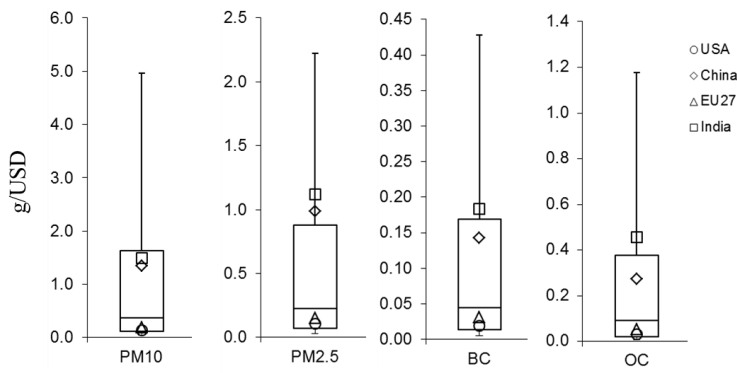


**Figure 3a 2010 per capita emissions per substance and per group of countries: low income (LIC), lower middle income (LMC), upper middle income (UMC) and high income (HIC) with the maximum, and minimum and the percentiles reported in the box plot (10°, 50°, 90°) and the maximum and minimum in each group of countries.**

1

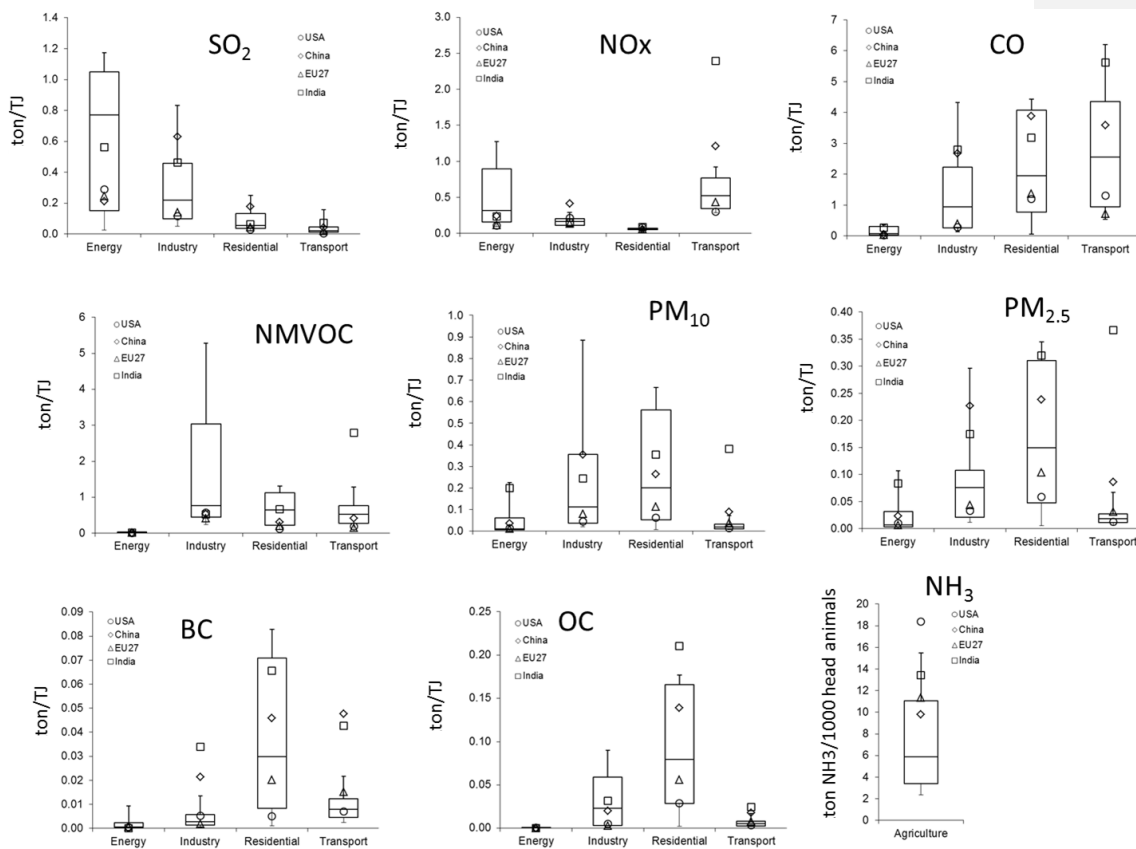


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3 **Figure 3b – Pollutant specific emissions divided by GDP (g/USD) for the year 2010.**  
4 **Percentiles are reported in the box plots (10°, 25°, 50°, 75°, 90°) together with**  
5 **emission/GDP for specific regions (EU27, USA, China and India).**

6



1  
2 **Figure 4 - Sector specific implied emissions (ton/TJ) for the year 2010. Percentiles are reported in the box**  
3 **plots (10°, 25°, 50°, 75°, 90°) together with implied emission factors for specific regions (EU27, USA,**  
4 **China and India). For the percentiles the following countries are left out:**  
5 **For CO: for the INDUSTRY sector: Togo, Eritrea, Congo, Côte d'Ivoire, Kenya, Benin; for the RESIDENTIAL**  
6 **sector: Maldives; for the TRANSPORT sector: North-Korea, Afghanistan, Laos, Tajikistan, Mongolia.**  
7 **For SO2: for the INDUSTRY sector: Namibia, Laos, Jamaica.**  
8 **For NOx: for THE RESIDENTIAL sector: Maldives; for the TRANSPORT sector: Afghanistan, Laos, North-**  
9 **Korea, Tajikistan.**  
10 **For NMVOC: for the ENERGY sector: Bhutan; for the INDUSTRY sector: Togo, Eritrea, Côte d'Ivoire, Congo,**  
11 **Cameroon, Kenya, Benin, Aruba, Antigua, Bahamas, Ethiopia, Sudan, Senegal, Equatorial Guinea, Central**  
12 **African Rep., Sri Lanka, Angola, Mozambique, Zambia, Jamaica; for the RESIDENTIAL sector: Am. Samoa,**  
13 **Gum, Maldives, Tonga; for the TRANSPORT sector: Afghanistan, Laos, North-Korea.**  
14 **For PM10: for the INDUSTRY sector: Togo, Eritrea, Côte d'Ivoir, Congo, Kenya, Benin, for the TRANSPORT**  
15 **sector: Afghanistan.**

1 For PM2.5: for the ENERGY sector: Tajikistan, Luxembourg; for the INDUSTRY sector: Togo and Eritrea; for  
2 the TRANSPORT sector: Afghanistan.

3 For BC: for the ENERGY sector: Nigeria, Malaysia, Belgium, Oman, Finland, Georgia, Vietnam, Canada,  
4 Armenia, Tunisia, Jordan, The Netherlands, Trinidad and Tobago, Algeria, Latvia, United Arab Emirates,  
5 Brunei, Turkmenistan, Japan, Mozambique, Congo, Qatar, Bahrain, Moldova, Kyrgyzstan, South-Korea,  
6 Taiwan, Luxembourg, Bhutan, Tajikistan; for the INDUSTRY: Trinidad and Tobago, Malta; for the  
7 TRANSPORT sector: Afghanistan.

8 For OC: for the ENERGY sector: Tunisia, Jordan, Trinidad and Tobago, Algeria, United Arab Emirates, Brunei,  
9 Turkmenistan, Tajikistan, Mozambique, Congo, Qatar, Bahrain, Kyrgyzstan, Taiwan, Myanmar, South-Korea,  
10 Vietnam; for the INDUSTRY sector: Bahrain, Eritrea; for the RESIDENTIAL sector: Greenland, Gibraltar,  
11 Faroe Islands, Saint Pierre et Miquelon; for the TRANSPORT sector: Afghanistan

12 For NH3: for the AGRICULTURE sector: Faroe Islands, Tajikistan, Greenland, Falkland Islands, Kyrgyzstan,  
13 South-Korea, Brunei, Am. Samoa, Malaysia, Trinidad and Tobago, Bahamas, Saint Pierre et Miquelon, Sri  
14 Lanka, Suriname, Réunion, Thailand, Indonesia, Japan, Barbados, Bhutan, Guyana, Costa Rica.

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