

## Response to the reviewers – revision #2

We thank referee #2 for another critical reading of our manuscript, and the editor for the opportunity to revise our work and to add more information in response to the referees' comments during and after the open discussion. Referee #2 stated in his report on the revised version: "The analysis performed in response to the major concerns raised in review are not adequately discussed in the current form of the manuscript. In particular, the authors did not use these analyses to inform their conclusions or interpretation of results." We have addressed this criticism by extending the statistical analysis as described below and by rewriting major parts of sections 3 and 4.

Below, we briefly summarize the major points which the referees have raised in the open discussion and explain how we have addressed them in the newly revised manuscript.

### Quantification of the model performance for each scheme

In response to the reviewer's comments, we have calculated the correlation coefficients ( $R$ ) and normalized mean bias (NMB) between the observed FBAP and simulated fungal spore concentrations for all three parameterization schemes and all stations, as well as for the entire dataset as a whole. In addition, we have now added confidence intervals for the correlation coefficients (95% and 99% confidence level) and the root mean square error (RMSE) into Table 2. The results are now discussed in detail in sections 3.1 for the literature-based emission fluxes and in section 3.2 for the new emission parameterization. Possible reasons for the remaining deficiencies of the simulation are discussed in section 4 (see below).

The rewritten/inserted sections read:

Page 2, lines 15-17 (abstract)

*The new parameterization results in similar root mean square errors and correlation coefficients compared to the FBAP observations as the previously existing fungal spore emission parameterizations, with some improvements in the bias.*

Page 13, line 13- page 14, line 2:

*The correlation coefficient for the entire data set amounts to  $R=0.43$  for the simulation with the Heald and Spracklen (2009) emission ( $F_{H\&S}$ ) and to  $R=-0.05$  for the simulation with the Sesartic and Dallafior (2011) ( $F_{S\&D}$ ) parameterization. For most time periods at Karlsruhe and Hyytiälä, the simulated fungal spore concentrations are smaller than the measured FBAP concentrations (**Fehler! Verweisquelle konnte nicht gefunden werden.**). This difference is highest at Hyytiälä in August 2010. At Hyytiälä in July and at Manchester and Killarney in August,  $F_{H\&S}$  gives median concentration values which agree reasonably well with the measurements. During October, the fungal spores number concentrations based on constant emission fluxes ( $F_{S\&D}$ ) agree best with the measured FBAP concentrations. Taking the whole dataset together, these deviations result in negative values of the normalized mean bias (NMB) of -44% for  $F_{H\&S}$  and of -29% for  $F_{S\&D}$  and root mean square errors of  $26.2\text{ L}^{-1}$  and  $25.6\text{ L}^{-1}$ , respectively (Table 2). Possible causes for the bias of  $F_{H\&S}$  and  $F_{S\&D}$  may come from different assumptions made to determine the*

*fungus spore concentrations in ambient air. On the one hand, the mass size distribution of mannitol, which is used as a chemical tracer for fungus spores by Heald and Spracklen (2009), peaks in their study at particle diameters of  $\sim 5 \mu m$ . Additionally to fungus spores, bacteria, algae, lichens, and plant fragments, can produce mannitol and some of these can contribute to PBAP concentrations at  $\sim 5 \mu m$ . On the other hand, similar assumptions are made for this study by linking FBAP to fungus spores, which can also introduce biases. Furthermore, the treatment of fungus spores as monodisperse particles in the model (while the observed size distribution is rather broad) influences the simulated removal processes and thus the resulting concentrations.*

Page 20, line 25 - page 21, line 10:

*A statistical analysis of the results (Table 3) indicates that normalized mean bias (NMB) improves by the newly introduced emission function  $F_{FBAP}$ , but not the correlation coefficient  $R$ , between simulated fungus spore and observed FBAP concentrations which remains at the value of approximately 0.43 for the overall dataset. For the different time periods and stations,  $R$  varies between -0.17 and 0.66 (negative at only one station). Differences in  $R$  are small between  $N_{H\&S}$  and  $N_{FBAP}$ , because both make use of emission rates as a function of almost the same parameters ( $N_{FBAP}$  includes an additional  $T$  dependence). The bias reduction and similar  $R$  is also visible in Fig. 8. Parameters  $b_1$  and  $b_2$  in eq. (8) are chosen to give fungus spore concentrations matching best with measured FBAP concentrations, thus the reduction in NMB from -44% ( $N_{H\&S}$ ) to -0.4% ( $N_{FBAP}$ ). At the same time, the RMSE decreases slightly from 26.2 L-1 ( $N_{H\&S}$ ) to 23.2 L-1 ( $N_{FBAP}$ ). At  $T = 275.82 K$ ,  $F_{H\&S}$  is equal to  $F_{FBAP}$  for typical values of LAI and  $q_v$ , and temperatures above this threshold (as it is the case for almost all locations) shift  $F_{FBAP}$  to give a larger emission flux. At meteorological conditions present for the selected cases, the second part of eq. (8) dominates over the first part by a factor of  $\sim 4$  and therefore temperature changes have only a secondary influence on the emission flux. Hence,  $R$  is similar for both emission parameterizations.*

### **Distinction between bias and skill, and other potential causes for the bias**

As pointed out in section 4, the new emission parameterization leads to a reduction of the bias between simulated and observed concentrations (as quantified by the NMB in Table 2), but only to very small improvements of the correlation coefficient  $R$  and the RMSE, which can be interpreted as a measure of the model skill. Our wording in the conclusion section reads as follows:

Page 25, lines 14-28:

*The resulting concentrations have on the average a smaller normalized mean bias and a slightly smaller root mean square error compared to the measured FBAP concentrations, but variations in the measurements are not always captured by the simulation. Thus, the correlation coefficient remains low (0.43). Possible reasons include biases due to local conditions at the measurement stations, biases in the input parameters and model errors related to the transport, mixing and sink processes (including boundary layer turbulence and washout by precipitation). Future work including a long-term analysis of FBAP concentrations and environmental conditions may result in a further adjustment of the coefficients or reveal other*

*parameters or functional dependencies driving the emission. Ideally, the observations would be split into a training data set and an evaluation data set, which was not possible until now due to the limited amount of available observations.*

Possible reasons for the lack of improvement in R and RMSE include the following:

- A prerequisite for the derivation of emission fluxes from concentration measurements is that the assumption of local mixing within the boundary layer and an inverse relationship between concentration and boundary layer height holds. This has now been tested statistically (Table 3). The results are discussed in sections 3.1 and 3.2:

Page 15, lines 14-24:

*For a more quantitative analysis, the correlation coefficient between ( $N_f$ ) and  $h_{PBL}^{-1}$  was calculated for the whole dataset. The resulting value is a small positive correlation of  $R=0.11$  with a 95% confidence interval between 0.05 and 0.17. If all data points with simulated boundary layer heights below 10 m (which are considered problematic as the vertical resolution of the model is coarser than 10 m) are omitted from the correlation analysis, the correlation coefficient rises to  $R=0.18$  with a 95% confidence interval between 0.12 and 0.24. The correlation coefficients for the individual time series are listed in Table 3. Three of the stations exhibit negative correlation coefficients, which might be due to perturbations by rain as discussed below.*

Page 18, lines 6-10:

*Eq. (7) is applicable to derive estimates for  $F_{F,c}$  under the assumption that the variation of the boundary layer height explains a significant amount of the variation in  $N_f$ . This condition is fulfilled for four of the individual time series and to a lesser extent for the dataset as a whole (see section 3.1 and Table 3).*

- The boundary layer height  $h_{PBL}$  is taken from the model simulations as it is not continuously available from observations at the stations. In the COSMO model,  $h$  is calculated by a bulk Richardson number method (Szintai and Kaufmann, 2008) with known deficiencies. For example, it was shown for COSMO-2 simulations over the Swiss Plateau that this method overestimates  $h$  for convective boundary layers and strongly underestimates it for stable boundary layers (Collaud Coen et al, 2014).
- The input parameters to the parameterization (LAI,  $T$ ,  $qv$ ) could be biased. This has been investigated by a comparison of the COSMO LAI from the Global Land Cover 2000 database (European Commission, Joint Research Centre, 2003), which was used in the simulations, to the MODIS LAI product which is available at 1 km horizontal and 8-day temporal resolution. No systematic bias has been found.
- Precipitation has two opposing effects on concentrations by (a) scavenging of particles and (b) possibly increasing emissions (Huffman et al, 2013). The latter effect is not taken into account by the emission parameterizations.
- The above points are included in the manuscript in section 3.2:

*Possible reasons for the only small improvements in  $R$  and RMSE may be biases in the input parameters to the parameterizations or model errors in the processes affecting fungal spore concentrations. To investigate this, the leaf area index was compared to independent observations.*

*In Table 5, the LAI from the Global Land Cover 2000 database (European Commission, Joint Research Centre, 2003), which was used in the COSMO simulations, is compared to the MODIS LAI product (MCD14A2, MODIS collection 5, Knyazikhin et al, 1998) which is available at 1 km horizontal and 8-day temporal resolution. Both nearest neighbor pixels and averages over 20x20 km with standard deviations are calculated. For the Karlsruhe station and for Hyytiälä in July and October, a good agreement is obtained. For Hyytiälä and Manchester in August, MODIS indicates an about factor 2 lower LAI than assumed in our simulations. For Killarney, in contrast, the MODIS LAI is about 50% higher. While this possible error in LAI can explain part of the discrepancy between simulated and observed FBAP concentrations for Manchester, the bias in Hyytiälä is in the opposite direction as explainable by an error in LAI and thus has to be related to other error sources. In Killarney, model and measurements agree rather well despite the LAI underprediction.*

*In addition, known deficiencies in the simulated boundary layer height may reduce the quality of the derived fit. The boundary layer height  $h$  had to be taken from the model simulations as it is not continuously available from observations the stations. In the COSMO model,  $h$  is calculated by a bulk Richardson number method (Szintai and Kaufmann, 2008). It was shown for COSMO-2 simulations over the Swiss Plateau that this method overestimates  $h$  for convective boundary layers and strongly underestimates it for stable boundary layers (Collaud Coen et al, 2014).*

*Finally, the role of precipitation on FBAP concentrations is ambiguous, as already discussed in section 3.1. At present, precipitation is not included as argument for the emission functions. Thus, they possibly neglect a driver of the variability.*

- When using local measurements to deduce emission fluxes, it is assumed that transport effects on the concentration are negligible and that the stations are representative for a larger fetch area. This assumption has to be questioned e. g. for Manchester, which is located about 50 km from the coast.

### **Separation of the contributions of meteorology and emissions to variability**

Referee #2 has requested to perform a simulation with constant emissions to separate the effects of meteorology and emissions on the variability in FBAP concentrations. We argue that the simulation with the emission parameterization by Sesartic & Dallafior (2011) already meets this purpose. The emission function depends only on the land cover type, but does not vary in time. The correlation coefficient between the concentrations simulated with the Sesartic & Dallafior (2011) parameterization and observations amounts to  $R=-0.05$  (not significantly different from 0), while the other two parameterizations yield  $R=0.43$  (with a 99% confidence interval between 0.37 and 0.49). Thus, variable emissions can explain more variability in the fungal spore concentrations, although the approaches presented here do not yet capture the full variability in the observations.

## Contribution of mineral dust to FBAP

The analysis of backtrajectories as presented in the replies to referee #1 did not show any indication of influence of Saharan air masses during the investigated time periods. In addition, we present here data from the operational dust forecast with the BSC-DREAM8b model, operated by the Barcelona Supercomputing Center (<http://www.bsc.es/earth-sciences/mineral-dust-forecast-system/>). They clearly indicate that the transport of desert dust across the Alps was negligible for the studied episodes.

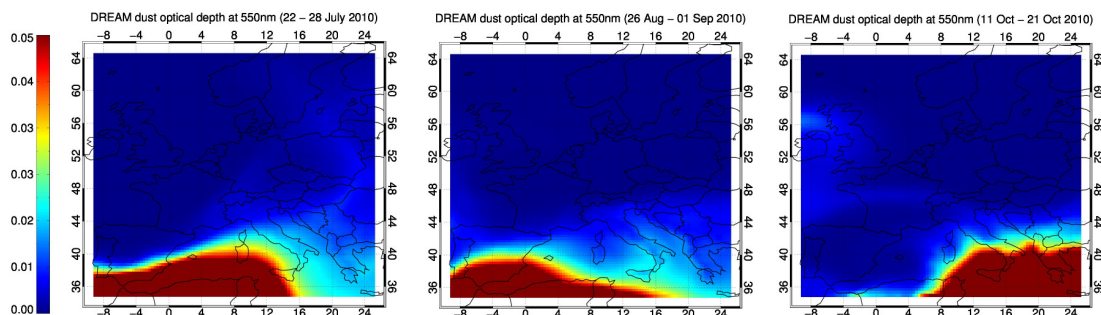


Figure: Dust optical depth as simulated by the BSC-DREAM8b model, operated by the Barcelona Supercomputing Center (<http://www.bsc.es/earth-sciences/mineral-dust-forecast-system/>), averaged over the three studied episodes.

However, a contribution of local dust sources (e.g. from agricultural areas) at the measurement locations cannot be ruled out. Auxiliary aerosol observations are necessary to quantify this contamination of the FBAP observations. Such analyses are outside the scope of this manuscript.

The possible contribution of mineral dust is now discussed in section 4:

Page 26, line 32 – page 27, line 12:

*At the selected time periods, back trajectories with HYSPLIT (Draxler and Rolph, 2013) and the operational dust forecast with the BSC-DREAM8b model, operated by the Barcelona Supercomputing Center (<http://www.bsc.es/earth-sciences/mineral-dust-forecast-system/>) suggest that transport of Sahara dust to the measurement stations into the domain is low (not shown), but influences the aerosol concentrations in Southern Spain. The emission from non-desert dust sources (in particular agricultural areas) is strongly episodic, as it is linked not only to meteorological conditions, but also to human activity (e.g. tillage operations (Funk et al, 2008, Goossens et al, 2001), traffic). The magnitude of the source strength is uncertain and not included in the model, as no validated emission parameterization is available at present. Beuck et al (2011) estimated that mineral dust (including both long-range transport and local sources) corresponds to on average 10% of PM10 concentrations in rural North-West Germany, and 14% of PM10 concentrations in urban areas in this region.*

## Atmospheric/boundary layer residence time of fungal spores

Referee #1 has requested more discussion on the value and interpretation of  $\tau$ . The related discussion in section 3.2 has been reformulated and extended:

Page 18, lines 16-28:

*As a remedy,  $\tau$  is corrected to  $\tau = 4 \frac{3}{4}$  hours which can be understood as a mean time for boundary layer mixing of fungal spores. The deviation from a lifetime of  $3 \mu\text{m}$  particles given in literature may be attributed to the assumption of a constant vertical distribution of fungal spores with increasing altitude until boundary layer height. A typical vertical profile from the model (Fig. 7) shows that this assumption is valid as a first approximation, but that deviations are visible both close to the surface and close to the top of the boundary layer. In the simulations, a mean ratio of approximately 1.75 between surface-level concentrations and mean concentrations within the boundary layer is found, which is still too small to explain the discrepancy between the expected lifetime and the value of  $\tau$  which leads to agreement with the observed concentrations from which the emission flux was derived. The remaining discrepancy may be caused by advection processes, by non-equilibrium conditions, and by the much longer lifetime of spores above the boundary layer.*

### **Further major changes to the manuscript**

Throughout the manuscript, the text has been extensively rewritten in order to clarify our assumptions and results and to address caveats. In particular, the particle number concentrations simulated with the new emission parameterization are now consistently termed as FBAP and not as fungal spores in order to point out the difficulties in the association of FBAP with spores.

The former section 3.2 is now subdivided into two sections to enhance readability:  
3.2. *Development of a new FBAP Emission Parameterization by Adaptation to FBAP Measurements*

3.3. *Results of simulations with the new emission parameterization*

M. Stengel has provided the comparisons of the GLC2000 leaf area index to MODIS observations and has been added as a coauthor.

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# Regional-scale Simulations of Fungal Spore Aerosols Using an Emission Parameterization Adapted to Local Measurements of Fluorescent Biological Aerosol Particles

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## Abstract

Fungal spores as a prominent type of primary biological aerosol particles (PBAP) have been  
incorporated into the COSMO-ART regional atmospheric model. ~~using and comparing three~~  
~~different emission parameterizations.~~ Two literature-based emission rates [for fungal spores](#)



derived from fungal spore colony counts and chemical tracer measurements were used as a parameterization baseline for this study. A third, new emission parameterization [for fluorescent biological aerosol particles \(FBAP\)](#) was adapted to field measurements ~~of fluorescent biological aerosol particles (FBAP)~~ from four locations across [Northern Europe](#). FBAP concentrations can be regarded as a lower estimate of total PBAP concentrations. Size distributions of FBAP often show a distinct mode at approx. 3  $\mu\text{m}$ , corresponding to a diameter range characteristic for many fungal spores. Previous studies [for several locations](#) have suggested that ~~the majority of FBAP in several locations~~ are [in many cases](#) dominated by fungal spores. Thus, we suggest that simulated [FBAP and](#) fungal spore concentrations obtained from the [three different](#) emission parameterizations can be compared [to FBAP measurements](#). ~~to the sum of total FBAP concentrations. A~~ The comparison reveals that ~~parameterized estimates of simulated~~ fungal spore concentrations based on literature ~~numbers~~ [emission parameterizations underestimate are lower than](#) measured FBAP concentrations. In agreement with [the measurements data](#), the model results show a diurnal cycle in simulated fungal spore concentrations, which may develop partially as a consequence of a varying boundary layer height between day and night. ~~Measured FBAP and simulated fungal spore concentrations also correlate similarly with simulated temperature and humidity. These meteorological variables~~ [Temperature and specific humidity](#), together with leaf area index, were chosen ~~to drive the new emission parameterization discussed here~~ [which is fitted to the FBAP observations. The new parameterization results in similar root mean square errors and correlation coefficients compared to the FBAP observations as the previously existing fungal spore emission parameterizations, with some improvements in the bias.](#) Using the new emission parameterization on a model domain covering Western Europe, ~~FBAP fungal spores~~ in the lowest model layer comprise a fraction of 15% of the total aerosol mass over land and reach average number concentrations of  $26 \text{ L}^{-1}$ . The results confirm that fungal spores and biological particles may account for a major fraction of supermicron aerosol particle number and mass concentration over vegetated continental regions and should thus be explicitly considered in air quality and climate studies.

## 1. Introduction

Particles emitted from biological sources are a miscellaneous and omnipresent group of the Earth's atmospheric aerosols (Elbert et al., 2007; Després et al., 2012). These primary

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1 biological aerosol particles (PBAP) can be transported over large distances and their impacts  
 2 are studied by various fields of research, such as atmospheric science, agricultural research,  
 3 biogeography and public health (Burrows et al., 2009). PBAP are solid airborne particles of  
 4 biological origin and include microorganisms or reproductive units (e.g. bacteria, fungi,  
 5 spores, pollen or viruses) as well as excretions and fragments of biological organisms (e.g.  
 6 detritus, microbial fragments or leaf debris) (Després et al., 2012). Typical sizes range from  
 7  $< 0.3 \mu\text{m}$  for viruses to diameters of single bacteria ( $0.25 - 3 \mu\text{m}$ ), bacteria agglomerates  
 8 ( $3 - 8 \mu\text{m}$ ), fungal spores ( $1 - 30 \mu\text{m}$ ), and up to  $10 - 100 \mu\text{m}$  for airborne pollen (Jones and  
 9 Harrison, 2004; Shaffer and Lighthart, 1997; Després et al., 2012).

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10 The share of atmospheric aerosol composition belonging to PBAP is large and possibly  
 11 underestimated (Jaenicke et al., 2007), but is also very uncertain. Estimates of relative PBAP  
 12 fraction from global models and local measurements reveal large differences between reports.

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13 On one hand, the calculated ~~global~~ number concentration of PBAP (zonal annual mean  
 14 surface concentrations of  $10^{-2} - 10^{-1} \text{cm}^{-3}$ ) is less than that of ~~below~~ mineral dust ( $65 \text{cm}^{-3}$ ) or  
 15 soot ( $1000 \text{cm}^{-3}$ ) concentrations by several orders of magnitude (Hoose et al., 2010b).

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16 Modeling studies have yielded global source strengths of  $\sim 10 \text{Tg/yr}$  (plant debris and fungal  
 17 spores, Winiwarter et al., 2009),  $56 \text{Tg/yr}$  (all PBAP types, Penner, 1995),  $78 \text{Tg/yr}$  (bacteria,  
 18 fungal spores and pollen, Hoose et al., 2010a),  $164 \text{Tg/yr}$  (all PBAP types, Mahowald et al.,  
 19 2008) and  $312 \text{Tg/yr}$  (bacteria, fungal spores and pollen, Jacobson and Streets, 2009) ~~for~~  
 20 different PBAP components. On the other hand, measurements of continental boundary layer  
 21 air in remote vegetated regions indicate that the mass fraction of PBAP in the coarse particle  
 22 size range can be as high as  $\sim 30\%$  ( $> 0.2 \mu\text{m}$ , Siberia, Matthias-Maser et al., 2000) or  $65-85\%$   
 23 ( $> 1 \mu\text{m}$ , Amazonia, Martin et al., 2010; Pöschl et al., 2010; Huffman et al., 2012).

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24 Like all other aerosol particles, PBAP can influence the Earth's climate by forcing the  
 25 radiation budget directly (by absorbing or scattering radiation) and indirectly (by affecting  
 26 cloud microphysics) (Forster et al., 2007). The direct PBAP effect on climate is difficult to

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27 estimate, because ~~evaluations of the~~ atmospheric PBAP concentrations can vary by several  
 28 orders of magnitude depending on time and location. ~~when taking spatial and temporal~~  
 29 ~~divergences into account~~. Describing the radiative properties of PBAP is complicated,  
 30 because their size ranges from fine to coarse (up to  $100 \mu\text{m}$  in diameter) and in many cases  
 31 their shapes are non-spherical and not accurately known. Hence, the applicability of Mie  
 32 scattering theory is limited (Després et al., 2012). However, the direct PBAP effect on global

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1 and regional climate is generally assumed to be small due to low average concentrations, in  
2 contrast to the numbers of sub-micron absorbing and scattering aerosols. The indirect PBAP  
3 effect on climate is caused by PBAP that act as cloud condensation nuclei (CCN) and/or as  
4 ice nuclei (IN). Generally, changing aerosol populations (e.g. by increasing nuclei  
5 concentrations or behavior) can alter the microphysical properties of clouds, thus influencing  
6 the climate system (Forster et al., 2007). Most PBAP are assumed to be good CCN, because  
7 their surface area is large compared to most other aerosol species (Petters and Kreidenweis,  
8 2007; Ariya et al., 2009) and thus may act as so-called giant CCN (Pöschl et al., 2010). For  
9 these particles~~Here~~, the Kelvin effect can be neglected when describing water vapor  
10 condensation, and thus activation and growth proceeds quickly (Pope, 2010). Some particles  
11 of biological origin (e.g. *P. syringae* bacteria and some fungal species) have been found to  
12 efficiently nucleate ice growth at relatively high temperatures (Després et al., 2012; Murray et  
13 al., 2012; Hoose and Möhler, 2012; Morris et al., 2004; Morris et al., 2013; Haga et al., 2013).  
14 Biological particles have been observed ubiquitously in precipitation, fog, and snowpack (e.g.  
15 Christner et al., 2008) and, in clouds by~~from~~ airborne measurements (e.g. Prenni et al., 2009;  
16 DeLeon-Rodriguez et al., 2013) and have been shown to be important fractions of IN  
17 measured at ground level (e.g. Huffman et al., 2013; Prenni et al., 2013). These bio-IN may  
18 be important for ice nucleation in mixed-phase clouds at temperatures higher than -15°C  
19 (DeMott and Prenni, 2010). In regimes colder than that, mineral dust particles and other ice  
20 nucleators are also active and the relative atmospheric abundance of PBAP is probably too  
21 small to contribute significantly to formation and evolution of these colder clouds. Previous  
22 modelling studies suggest that bio-IN concentrations are several orders of magnitude lower  
23 than IN concentrations from mineral dust or soot and hence the influence of bio-IN on  
24 precipitation is limited on the global scale (Hoose et al., 2010a; Sesartic et al., 2012;  
25 Spracklen and Heald, 2013). In-situ analyses of insoluble cloud ice and precipitation residuals  
26 meanwhile highlight the contribution of bio-IN to precipitation, and back trajectories indicate  
27 that they can be transported over large distances (Creamean et al., 2013). These bio-IN may be  
28 important to ice nucleation in mixed-phase clouds at temperatures warmer than -15°C (DeMott  
29 and Prenni, 2010; Hoose et al., 2010a; Creamean et al., 2013), and therefore could influence  
30 atmospheric radiative properties on up to regional scales. In regimes colder than that, mineral  
31 dust particles and other ice nucleators are also active and the relative atmospheric abundance of  
32 PBAP is probably too small to contribute significantly to formation and evolution of these  
33 colder clouds.

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1 The methods for identifying and detecting PBAP are challenging and many different PBAP  
 2 can introduce significant detection biases. Particle diameter often plays heavily into PBAP  
 3 detection and characterization, and it should be noted that large discrepancies can exist  
 4 between physical and aerodynamic diameter measurements (Huffman et al., 2010; Reponen et  
 5 al., 2001). PBAP concentrations can be obtained either by online techniques, in which  
 6 samples are analyzed by advanced instrumentation in real-time, or by offline measurement  
 7 techniques. If measured offline, samples of airborne biological particles are stored under  
 8 refrigeration and common methods include analysis by microscopy (stained or unmodified),  
 9 by cultivation of the sample on growth media, and by amplification and detection of genetic  
 10 material by sequencing or electrophoretic separation. Chemical and optical properties of  
 11 PBAP samples or their tracers can be monitored in real time by: chromatography,  
 12 massspectrometry, fluorescence spectrophotometry, LIDAR, and flow cytometry. Short  
 13 overviews of PBAP analysis techniques have been given by Caruana et al. (2011) and  
 14 Després et al. (2012).

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15 This paper focuses on the mesoscale simulation of atmospheric concentrations of fungal  
 16 spores. ~~The COSMO-ART~~ART limited-area model is used for the simulations and the setup  
 17 includes a model domain covering most parts of Europe with a horizontal resolution of 14 km.  
 18 Two different fungal spore emission parameterizations (Heald and Spracklen, 2009; Sesartic  
 19 and Dallafior, 2011) are tested by comparing their number concentrations to online laser-  
 20 induced fluorescence (LIF) measurements of airborne fluorescent biological particles.  
 21 Additionally, a new emission parameterization adapted to these measurements is introduced.  
 22 Field data used here comes from a real-time measurement technique that detects the intrinsic  
 23 (i.e. unstained) fluorescence signal, after UV excitation, of fluorophores commonly present in  
 24 most biological materials (e.g. free proteins, fungal spores, bacteria, and leaf fragments).  
 25 Detected particles are categorized as fluorescent biological aerosol particles (FBAP), which  
 26 may broadly be considered a lower limit for the abundance of PBAP (Huffman et al., 2010;  
 27 Pöhlker et al., 2012; Healy et al., 2014). FBAP were measured at four different locations  
 28 (Table 1Table 1) concurrently during three focus periods in summer 2010 and fall 2010. The  
 29 resulting FBAP size distribution is usually dominated by particles in the range from 2 µm to  
 30 4 µm, which is consistent with the size of fungal spores (Huffman et al., 2010; Pöschl et al.,  
 31 2010; Huffman et al., 2012; Healy et al., 2012a; Toprak and Schnaiter, 2013; Huffman et al.,  
 32 2013). Further, the concentration of FBAP in a given air-mass is generally considered to  
 33 underestimate PBAP concentration due to biological particles that exhibit very low levels of

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1 fluorescent emission (Huffman et al., 2012). To some extent, non-biological aerosol  
2 components can also be part of the fluorescence signal for fine particles ( $\sim 1 \mu\text{m}$ ) (Huffman et  
3 al., 2010; Toprak and Schnaiter, 2013). These factors contribute uncertainty to the [evaluation](#)  
4 [of the](#) parameterizations discussed here, however the overall ability of LIF techniques to  
5 provide real-time FBAP measurements allows first approximation measurements that can be  
6 enlightening.

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## 2. Methodology

### 2.1. Model Description

The COSMO-ART ([Consortium for Small-scale Modelling – Aerosols and Reactive Trace gases](#)) atmospheric model system is based on the forecast model of the German weather service, combined with an online coupled module for simulating the spatial and temporal distribution of reactive gaseous and particulate components (Vogel et al., 2009). Additionally, fungal spores are incorporated as an independent, monodisperse particle class ( $d_p = 3 \mu\text{m}$ ).

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Parameterizations for emission, sedimentation, and washout, which were originally developed for pollen dispersal, are included for this particle class (Helbig et al., 2004). Fungal spores are treated independently, as no interactions with other aerosols or gases (coagulation or condensation) are considered. The temporal development of the fungal spore number concentration is calculated by:

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$$\rho \frac{d\Psi}{dt} = -\nabla \cdot \vec{F}_T - \frac{\partial}{\partial z} F_S - \lambda \Psi - \frac{1}{N} \frac{\partial}{\partial z} F_E \quad (1)$$

with the number mixing ratio of fungal spores being

$$\Psi = \frac{N_f}{N}, \quad (2)$$

and the number concentration of fungal spores  $N_f$ , the total number of particles and air molecules  $N$  per  $\text{m}^3$ , the air density  $\rho$ , the turbulent flux  $\vec{F}_T$ , the sedimentation flux  $F_S$ , a washout coefficient  $\lambda$  and a vertical emission flux  $F_E$  (Vogel et al., 2008). The turbulent flux is calculated by  $\vec{F}_T = \overline{\rho v' \Psi'}$ , incorporating the turbulent fluctuations of wind speed  $v'$  and fungal spore number mixing ratio  $\Psi'$ . Fungal spore sedimentation is calculated by  $F_S = \rho \Psi v_s$ . The fungal spore settling velocity  $v_s$  is calculated by applying the volume-equivalent particle diameter  $d_e = 2 \sqrt[3]{a^2 b}$ , with  $a = 2.1 \mu\text{m}$  and  $b = 5 \mu\text{m}$  (Yamamoto et al., 2012) being the major and minor radius of a prolate spheroid. This results in:

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$$v_s^2 = \frac{4 \rho_p d_e g}{3 \rho c_d} \quad (3)$$

where  $\rho_p = 1 \text{ g/cm}^3$  is the spore density (Trail et al., 2005; Gregory, 1961) and  $c_d$  the drag coefficient (Aylor, 2002). The calculation of the washout coefficient is based on the assumption of raindrops being much larger than aerosol particles and having a much higher terminal fall velocity. It yields:

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$$\lambda(d_p) = \int_0^{\infty} \frac{\pi}{4} D_D^2 v_t(D_D) E(d_p, D_D) n(D_D) dD_D \quad (4)$$

(Rinke, 2008).  $D_D$  and  $d_p$  are the diameters of raindrops and particles, respectively,  $v_t(D_D)$  is the terminal fall velocity,  $E$  is a collision efficiency and  $n(D_D)$  is the size distribution of the raindrop number concentration. For fungal spores with a spherical diameter of 3  $\mu\text{m}$ , the collision efficiency  $E$  with 0.1 mm and 1 mm droplets is approximately 0.085 and 0.3, respectively.

Adapting the model for simulations of fungal spores requires inclusion of an emission flux  $F_E$  in the source term of eq. (1) by means of an emission parameterization which will be described in the next section.

Together with fungal spore simulations COSMO-ART is used to compute the mass concentration of major atmospheric aerosol components. Hence, the proportion of fungal spores with respect to the dry aerosol mass can be estimated (section 3.4.3). In addition to primary aerosol emissions, further gaseous emissions ~~given by the EMPA emission dataset~~ (section 2.3) are taken into account. Partitioning of inorganic aerosol components between the gases ~~at and~~ particulate phase is simulated by the ISORROPIA II module (Fountoukis and Nenes, 2007). Condensation on fungal spore aerosols is not included. The contribution of secondary organic aerosols (SOA) to the particles is handled by condensation of oxidized volatile organic compounds as described by Schell et al. (2001). When soot aerosols are not involved as a solid nuclei enabling condensation, clusters build by gas-to-particle conversion via binary nucleation of sulfuric acid and water. They are computed as an individual particle mode. All aerosol particles including these chemical compounds are assumed to be internally mixed. A soot mode without mixing of other chemical compounds is included as particles that are emitted directly into the atmosphere. Anthropogenic primary aerosols (aPA) in the coarse size range ( $<10 \mu\text{m}$ ) are treated as a separate mode. Detailed descriptions are given in Vogel et al. (2009). Furthermore, sea salt is included in the model simulation and its emission is related to sea water temperature and wind speed (Lundgren et al., 2013). No desert dust emissions are included, as the model domain does not cover the corresponding emission regions and no transport into the model domain is taken into account.

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## 2.2. Emission Parameterization of Fungal Spores

In ~~literature~~previous studies, a constant emission rate was used as input of a global chemical transport model to represent the magnitude and range of measured concentrations of mannitol as a molecular tracer for basidiospores (Elbert et al., 2007). Broad geographical differences can be included in the emission flux by distinguishing between ecosystems. While reviewing the measured data available on measured fungal spore concentrations, Sesartic and Dallafior (2011) calculated number fluxes of fungal spore emissions for six different ecosystems (defined by Olson et al., 2001). Four of these emission fluxes were included into COSMO-ART, and coupled to ecosystem definitions by the GLC2000 (Global Landcover 2000 Database) (forest and shrubs) and Ramankutty et al. (2008) (grassland and crops). The sum of these fluxes, as defined by Sesartic and Dallafior (2011), are emitted from the land area fraction  $E_i$  of each ecosystem  $i$  ( $\sum_i^n E_i = 1$  for  $n$  number of ecosystems), gives the total emission flux  $F_E = F_{S\&D}$  in  $\text{m}^{-2}\text{s}^{-1}$  of eq. (1) for fungal spores:

$$F_{S\&D} = 214 \text{ m}^{-2}\text{s}^{-1}E_{forest} + 1203 \text{ m}^{-2}\text{s}^{-1}E_{shrub} + 165 \text{ m}^{-2}\text{s}^{-1}E_{grassland} + 2509 \text{ m}^{-2}\text{s}^{-1}E_{crop} \quad (5)$$

Additionally, a second emission parameterization was tested, which varies as a function of meteorological and surface conditions. Jones and Harrison (2004) reviewed the relations determined when analyzing the observed fungal spore concentrations and atmospheric factors. Seasonal variations can be explained by changes in the leaf area index ( $LAI$ ). This was verified by correlation to the observed mannitol concentrations. Among the drivers of day-to-day variations, specific humidity ( $q_v$ ) correlates best with the mannitol concentrations (Heald and Spracklen, 2009). It was argued that though other atmospheric factors (e.g. temperature) may actually drive the correlation, this does not change correlation results and thus parameterizations can proceed without having information about the root drivers of fungal spore release. ~~A constant~~The emission rate ~~was used here by~~is linearly scaled eding with  $LAI$  and  $q_v$  in order to give global fungal spore concentrations matching the mean mannitol concentrations (Heald and Spracklen, 2009). In order to ~~fit to rescale~~ the emission flux specified in Hoose et al. (2010a) ~~from~~a spore diameter of 5  $\mu\text{m}$  as in Hoose et al (2010a) to a spore diameter of 3  $\mu\text{m}$  as in this study, the prefactor ~~a constant~~  $c$  is set to  $c = 2315 \text{ m}^{-2} \text{ s}^{-1}$  ~~to be appropriate for fungal spores with 3  $\mu\text{m}$  in diameter~~. Based on the emission flux in eq. (1) ~~(4)~~, this gives an alternative source  $F_E = F_{H\&S}$  of fungal spores in  $\text{m}^{-2}\text{s}^{-1}$ :

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$$F_{H\&S} = c \frac{LAI}{5 \text{ m}^2 \text{ m}^{-2}} \frac{q_v}{1.5 \cdot 10^{-2} \text{ kg kg}^{-1}} \quad (6)$$

~~LAI is the leaf area index,  $q_v$  is the specific humidity at the surface, and their scaling factors adapted from tropical rain forest conditions are assumed to be  $LAI_{max} = 5 \text{ m}^2 \text{ m}^{-2}$  and  $q_{v,max} = 1.5 \times 10^{-2} \text{ kg kg}^{-1}$ .~~ In the COSMO-ART simulations LAI is horizontally distributed according to GLC2000 containing monthly variation and  $q_v$  is provided by the model as a meteorological variable.

### 2.3. Model Domain and Input Data

The COSMO-ART mesoscale model system is driven by initial and boundary data for meteorological ~~and aerosol and chemistry~~ conditions. ~~The meteorological conditions~~ They are updated every six hours and result from interpolation of the coarse grid operational atmospheric model analysis of the ECMWF (European Centre for Medium-Range Weather Forecasts). No initial and boundary concentrations are predefined for aerosols or gases. Therefore, all gaseous species are set to a climatological, homogeneously distributed initial concentration. ~~The and~~ emission rates for chemical compounds included in the ART part are updated hourly. They are provided by EMPA (Swiss Federal Laboratories for Materials Science and Technology) based on the TNO/MACC (Monitoring Atmospheric Composition and Climate) inventory (Kuenen et al., 2011). The treatment of emissions for COSMO-ART can be found in Knote et al. (2011). Homogeneously distributed mass densities for each aerosol are used as initial conditions, together with the aerosol size distribution and particle density. Primary particle emissions are included as parameterizations based on meteorological and surface conditions. Land use data and constant surface properties are derived from the GLC2000 database (Bartholomé and Belward, 2005). All parameters are post-processed to the rotated spherical coordinate system of COSMO-ART (Doms and Schättler, 2002). For the purpose of this paper, the model domain covers most parts of Western Europe from mainland Portugal to northern Finland, the longitudinal extension being 2849 km the latitudinal extension being 3803 km with a horizontal spacing of  $0.125^\circ$  ( $\triangleq 14 \text{ km}$ ) on a rotated grid. In vertical direction the model reaches up to an altitude of about 24 km distributed over 40 terrain-following levels. The timestepping of the Runge-Kutta dynamical core is set to 30 s.

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## 2.4. Auto-fluorescence Measurements

Ambient aerosols can be roughly classified as biological or not by interrogating particles at characteristic wavelengths of excitation and measuring the resultant emission in a process called ultraviolet light-induced fluorescence (UV-LIF) (e.g. Hairston et al., 1997; Pan et al., 1999). In particular, the region of fluorescent excitation near 360 nm is often used as characteristic of certain cell metabolites present in all living cells, including riboflavin and reduced pyridine nucleotides (e.g. NAD(P)H). The region of excitation near 270 nm includes certain amino acids (e.g. tryptophan) contained in ~~all~~most proteins. However, many other biological fluorophores exist and the relationship between the measured fluorescence of complex biological particles and fluorophore assignment is very complex (Pöhlker et al., 2012; Pöhlker et al., 2013).

Two instrument types were utilized at four locations for the comparison discussed in this paper. The ultraviolet aerodynamic particle sizer (UV-APS; TSI, Inc., Shoreview, MN, USA) measures particle size aerodynamically, excites individual particles using a single Nd:YAG laser pulse at 355 nm, and detects integrated fluorescent emission (non-dispersed) in a single wavelength region between 420 nm and 575 nm (Hairston et al., 1997; Brosseau et al., 2000; Huffman et al., 2010). The Waveband Integrated Bioaerosol Sensor (WIBS, versions 3 and 4; University of Hertfordshire, UK) measures particle size optically and excites individual particles via two sequential pulses from a Xe-flash lamp, at 280 nm and 370 nm, respectively (e.g. Kaye et al., 2005; Foot et al., 2008). Fluorescence for each particle is then measured in one of two wavelength regions, resulting in three measured fluorescence parameters for each WIBS instrument named FL1\_280, FL2\_280, and FL3\_370. See Gabey et al. (2010) and Robinson et al. (2013) for more details, including slight differences in WIBS-3 and WIBS-4 models. The number concentration of FBAP can be written as  $N_{F,c}$  with subscripts referring to fluorescent and coarse particle size. The differences in the pairs of wavelengths used for fluorescence, as well as the possible differences in sensitivity between instruments, suggest that the term “FBAP” as determined by each instrument is not rigorously interchangeable, and it is critical to understand the method of analysis when comparing datasets. For example, the ambient FBAP number concentration as determined by UV-APS has been shown to be qualitatively consistent with the number concentration of particles that fluorescence in the WIBS FL3\_370 channel, while the  $N_{F,c}$  comparison between UV-APS and WIBS FL1\_280 channel is relatively poor (Healy et al., 2014). Here we use the term FBAP

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1 from WIBS data to mean particles that exhibit fluorescence simultaneously in both channels  
2 FL1\_280 and FL3\_370.

3 Particle size can aide differentiation between biological particles classes observed, however  
4 the selectivity based on size alone is very uncertain. For example, and to a rough first  
5 approximation, it may be true that many FBAP ~1 µm are single bacterial particles and that  
6 many FBAP 2 - 6 µm may be fungal spores or bacterial agglomerates (Shaffer and Lighthart,  
7 1997). [A comparison between WIBS-4 and a Burkard volumetric impactor reported by](#)  
8 [O'Connor et al. \(2014\) from a yew forest site showed excellent correlation with  \$R^2 > 0.9\$](#)   
9 [demonstrating the real-time counting capability of WIBS for pollen.](#) However, biological  
10 species can vary widely, and other FBAP classes (e.g. fragments of larger PBAP, internal  
11 components of burst pollen, the presence of other biological species) confound the simple  
12 assignment of FBAP based on size (Després et al., 2012).

13 Further, at least a fraction of fluorescent, supermicron particles are likely to come from non-  
14 biological sources, and thus could be counted as FBAP. These non-biological process include  
15 anthropogenic sources (e.g. polycyclic aromatic hydrocarbon particles from combustion and  
16 cigarette smoke), present most often in submicron particles (Huffman et al., 2010), select  
17 oxidized organic aerosol particles (e.g. absorbing brown carbon particles) (Bones et al., 2010;  
18 Lee et al., 2013), and some humic-like substances (Gabey et al., 2013). For example, at the  
19 rural, elevated site of Puy de Dôme, France, WIBS-3FBAP measurements were compared to  
20 results from fluorescence microscopy paired with staining of fungal spores and bacteria.  
21 These results suggest that the real-time UV-LIF measurements indeed track the diurnal cycle  
22 of the bacteria concentration, but that non-biological particles still contributed significantly to  
23 fluorescent particle number (Gabey et al., 2013).

24 Virtually every ambient measurement study performed with the WIBS or UV-APS to date has  
25 shown a dominant FBAP mode peaking at 2 - 4 µm in size (Huffman et al., 2010; Huffman et  
26 al., 2012; Huffman et al., 2013; Gabey et al., 2010; Toprak and Schnaiter, 2013; Healy et al.,  
27 2014). For example, the FBAP size distributions measured at each of the four sampling  
28 locations discussed here is shown in [Figure 1](#)~~Figure 1~~, highlighting the common presence of  
29 the 2 - 4 µm peak. It has been proposed that fungal spores and bacteria agglomerates are the  
30 most dominant biological aerosols in this size range (Jones and Harrison, 2004; Després et al.,  
31 2012; Fang et al., 2008) and that the FBAP signal in this size range is typically dominated by  
32 fungal spores. This was corroborated in more detail for a remote Amazonian site using FBAP

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1 analysis along with fluorescence microscopy of stained filter samples (Huffman et al., 2012),  
2 but has not yet been rigorously tested in other environments. At the costal site of Killarney,  
3 results of fluorescence and optical microscopy of impacted biological particles reveal that  
4 some PBAP, e.g. sSpores of *Cladosporium spp.*, which have been frequently observed in  
5 many environments, were not correlated to the FBAP concentration (Healy et al., 2014).  
6 However, particle size modes of WIBS channel FL2 280 correlate with the concentration of  
7 airborne fungal spores commonly observed at the sampling site (Healy et al., 2014). Other  
8 microscopy and DNA-based studies have suggested that fungal spores constitute the largest  
9 fraction of PBAP in the 2 – 4 µm size (e.g. Graham et al., 2003; Lin and Li, 1996; Burch and  
10 Levetin, 2002). Bauer et al. (2008) showed that fungal spores account for an average of 60%  
11 of the organic content in the particulate matter in a size range of 2 - 10 µm in rural and urban  
12 areas of Vienna, Austria.

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### 3. Results

#### 3.1. Comparison of Time Series of Measured FBAP and Simulated Fungal Spores

Fungal spore concentrations simulated using the emission fluxes given in eqs. (5) and (6) according to Sesartic and Dallafior (2011) and Heald and Spracklen (2009) were first compared to FBAP measurements without further adjustment. An overview of time series for all measurements and simulations discussed here is shown in Figure 2 by a box-and-whiskers plot. Statistical parameters of the correlation between observations and simulations are presented in Table 2. The time periods for each of three case studies (Table 1) were chosen as exemplary periods when UV-LIF instruments were operating simultaneously at a minimum of two locations, with no requirements applied with respect to environmental conditions. For the statistical analysis, FBAP measurements were averaged over one hour periods in order to be consistent to the model output time steps. The correlation coefficient for the entire data set amounts to  $R=0.43$  for the simulation with the Heald and Spracklen (2009) emission ( $F_{H\&S}$ ) and to  $R=-0.05$  for the simulation with the Sesartic and Dallafior (2011) ( $F_{S\&D}$ ) parameterization. For most time periods at Karlsruhe and Hyytiälä, the simulated fungal spore concentrations are smaller than the measured FBAP concentrations (Figure 2). This difference is highest at Hyytiälä in August 2010. At Hyytiälä in July and at Manchester and Killarney in August,  $F_{H\&S}$  gives median concentration values which agree reasonably well with the measurements. During October, the fungal spores number concentrations based on constant emission fluxes ( $F_{S\&D}$ ) agree best with the measured FBAP concentrations. Taking the whole dataset together, these deviations result in negative values of the normalized mean bias (NMB) of -44% for  $F_{H\&S}$  and of -29% for  $F_{S\&D}$  and root mean square errors of  $26.2 \text{ L}^{-1}$  and  $25.6 \text{ L}^{-1}$ , respectively (Table 2). Possible causes for the bias of  $F_{H\&S}$  and  $F_{S\&D}$  may come from different assumptions made to determine the fungal spore concentrations in ambient air. On the one hand, the mass size distribution of mannitol, which is used as a chemical tracer for fungal spores by Heald and Spracklen (2009), peaks in their study at particle diameters of  $\sim 5 \mu\text{m}$ . Additionally to fungal spores, bacteria, algae, lichens, and plant fragments, can produce mannitol and some of these can contribute to PBAP concentrations at  $\sim 5 \mu\text{m}$ . On the other hand, similar assumptions are made for this study by linking FBAP to fungal spores, which can also introduce biases. Furthermore, the treatment of fungal spores as monodisperse particles in the model (while the observed size distribution is

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rather broad) influences the simulated removal processes and thus the resulting concentrations.

Long-term analysis of FBAP measurements, including periods at the Karlsruhe (Toprak and Schnaiter, 2013) and Hyytiälä site (Schumacher et al., 2013) discussed here, shows an annual cycle of average FBAP number concentrations peaking in summer and lowest in winter. Thus, a simulation based on a constant emission flux (such as  $F_{S\&D}$ ) may not be appropriate to reproduce the seasonal variation in FBAP concentrations, contributing to the low value of  $R_{\text{f}}$ .

Figures 3 to 6 show a series of one-week long case studies, each representing two measurement sites. The plots show comparisons between simulation and measurement time series for each station. The simulated fungal spore number concentration is given for the model grid point closest to the measuring site. Due to model spin-up, the first six hours of the simulated fungal spore concentrations are removed from the figures and are not included in the analysis. The total precipitation calculated by the model is shown by gray bars with the ordinate on the right hand side of the figure. The simulated boundary layer height is also included at the bottom of each panel in the figures.

Measured FBAP number concentrations often exhibit distinct diel (24-h) cycles with a maximum in the morning hours or around midnight and a minimum around noon. These features have been consistently reported by most studies discussing temporal behavior of FBAP (Gabey et al., 2010; Huffman et al., 2010; Huffman et al., 2012; Toprak and Schnaiter, 2013). Here, a similar diel cycle is frequently obtained from simulations, and the simulated fungal spore concentrations often anti-correlate with the simulated boundary layer height ( $h_{PBL}$ ) (Figures 3 to 6). The measured FBAP concentrations often qualitatively track the general pattern of simulated  $h_{PBL}$ , however the magnitude of concentration change and the timing is often not consistent. For example, on 24 and 25 July at the Karlsruhe site (Figure 3a) a boundary layer compression during the night leads to an increase in the simulated fungal spore concentrations by a factor of  $\sim 4$ , and during day the concentrations decreases as the boundary layer rises again. In this case, the measured FBAP concentrations are in relatively good agreement with the simulated fungal spore numbers, with  $N_{F,C}$  dropping slowly during the day on 24 and 27 July, and to a rate closer to the simulations on 25 July. This suggests that FBAP concentrations were likely influenced, at least partially, by the changing boundary layer height, though diel changes in biological emission are also likely to

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influence diel FBAP patterns. A similar temporal pattern in simulated fungal spore concentrations is shown in [Figure 4](#)~~Figure 4~~a, where a maximum in  $h_{PBL}$  at 12 and 13 October occurs approximately coincident with a minimum in the simulated number concentration. In this case, however, the measured FBAP concentrations do not reflect the diel pattern of the simulations. On 31 August ([Figure 5](#)~~Figure 5~~a), measured FBAP and simulated fungal spore number concentration increase simultaneously and parallel to the boundary layer compression, but the increase is more intense for FBAP measurements than for spore simulations. Additionally, at Manchester between 31 August and 1 September ([Figure 6](#)~~Figure 6~~a) measured and simulated concentrations are in good agreement. Distinct minima and maxima clearly anti-correlate with the minima and maxima of the boundary layer height. ~~In contrast~~Similarly, during the same time period in Killarney ([Figure 6](#)~~Figure 6~~b), ~~small~~ changes in the boundary layer height were simulated along with ~~minor coincident~~ changes in fungal spore concentrations. ~~In this case, measured FBAP concentrations qualitatively reflect the same temporal pattern of number concentration, but show poor trend consistency with  $h_{PBL}$ . The magnitude of diel FBAP concentration change was similar throughout the week shown, whereas  $h_{PBL}$  showed large diel variations between 26 and 30 August and relatively no change in  $h_{PBL}$  from 30 August to 1 September. In contrast, Figures 4a3b, 4b, and 5b also~~ show diel FBAP concentration changes that correlate poorly with simulated  $h_{PBL}^{-1}$ .

~~For a more quantitative analysis, the correlation coefficient between  $N_f$  and  $h_{PBL}^{-1}$  was calculated for the whole dataset (Table 3). The resulting value is a small positive correlation of  $R=0.11$  with a 95% confidence interval between 0.05 and 0.17. If all data points with simulated boundary layer heights below 10 m (which are considered problematic as the vertical resolution of the model is coarser than 10 m) are omitted from the correlation analysis, the correlation coefficient rises to  $R=0.18$  with a 95% confidence interval between 0.12 and 0.24. The correlation coefficients for the individual time series are also listed in Table 3. Note that a large positive correlation cannot be expected, as variations in the emission flux, deposition processes and transport all lead to a reduction of the correlation between  $N_f$  and  $h_{PBL}^{-1}$ . Three of the stations exhibit negative correlation coefficients, which might be due to perturbations by rain as discussed below.~~

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We conclude that (i) simulated fungal spore concentrations are sensitive to changes in the simulated boundary layer height, by extension, that (ii) diel cycles of FBAP concentrations are likely to be partially influenced by diel cycles of boundary layer height, but that (iii) the development of the FBAP concentration is in addition influenced by daily cycles in biological emission processes, including those of fungal spores and other PBAP classes, and atmospheric transport and sink processes. These competing effects are impossible to separate by this analysis when only point measurements are available.

A comparison of measured FBAP and simulated fungal spore number concentrations for July 2010 is shown in ~~Figure 3~~ Figure 3. At the measurement site of Karlsruhe, diel cycles were found in the simulated and measured time series, with constantly lower concentrations being obtained from simulations based on emission parameterizations given in literature. When precipitation occurs in the simulation, the simulated fungal spore concentrations decrease due to washout and the diel development of the concentration is interrupted. Afterwards, the simulated concentrations quickly return to the previous baseline. At Hyytiälä a strong decrease in simulated fungal spore concentration on 24 July precisely overlaps with the simulation of precipitation. After hitting a minimum value during simulated precipitation, the simulated fungal spore concentration increases steadily for two days as a result of a post-frontal shift in wind direction and decrease in wind speed. The increase is also reflected in the measured FBAP concentrations. However, the simulated precipitation values do not always coincide with precipitation at the site, as was the case in this instance. As a result of no rain falling at the site on 24 July, the measured FBAP concentration was not affected by washout as in the simulation ~~the simulated rain~~. ~~While this example shows that uncertainty in local errors in the simulated meteorology contribute to uncertainty to the aerosol output of the model, in the simulated aerosol concentrations, washout from precipitation remains an important modeled process for estimating FBAP concentrations.~~ Additionally, other dynamic processes are known to affect FBAP concentrations. For example, FBAP has been shown to increase dramatically during rainfall, a process reported recently for both a site in Colorado (Huffman et al., 2013) and also at the Hyytiälä site (Schumacher et al., 2013). The reasons for this FBAP increase are unclear, but are thought to be related to mechanical ejection from terrestrial surfaces as a result of rain droplet splash (Huffman et al., 2013). These effects are known to be dependent on the local geography and ecology, however, and are outside the scope of the presented emission parameterizations.

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During the simulation period of October 2010, the simulated fungal spore number concentrations  $F_{H\&S}$  are consistently below the measured FBAP concentrations at the sites of Karlsruhe and Hyytiälä, whereas  $F_{S\&D}$  matches the relative magnitude of the measurements more closely in both cases (Figure 4). At Karlsruhe, concentrations simulated by each emission parameterization follow a distinct diel cycle and increase slightly through the week, reaching concentration maxima on 15 October. The measured FBAP concentration develops differently, with only very weak diel cycle present from 11 to 14 October, and showing little relationship to the simulated  $h_{PBL}$ , as discussed above.

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At the end of August 2010, four different measurement series were available for a comparison to fungal spore simulations (Figure 5 and 6). The measured time series of FBAP number concentrations generally exhibit diel cycles, as discussed. The absolute FBAP concentration at Hyytiälä was consistently highest, when comparing all four sites. This trend is even more obvious when comparing the median concentrations on a linear scale (Figure 2). As a result, concentrations simulated from the literature-based parameterizations under-predict measurements by the greatest margin at Hyytiälä. This under-prediction is likely a result of particle washout due to the persistent precipitation simulated by the model and is an indication that precipitation has a stronger influence on the simulated concentrations than changes in the boundary layer height. Measured rainfall during this period at Hyytiälä was less consistent than the model predicts, but occurred with episodic peaks. In all other August case studies, simulated fungal spore concentrations show relatively good agreement with FBAP measurements.

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### 3.2. Development of a new FBAPa-Fungal Spore Emission Parameterization by Adaptation to FBAP Measurements

Direct comparison between simulated fungal spores and measured FBAP reveals that in general the simulated concentrations systematically underestimate are lower than the measured concentrations (Figure 8a). This difference is most distinct at Hyytiälä during the August case study and at Karlsruhe in the July and October case studies. Here we suggest a an improved new parameterization, including meteorological and surface parameters identified earlier, as drivers of fungal spore emissions of FBAP, which may in many cases be comparable to fungal spores. In the model, FBAP are treated identically to fungal spores as

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described in section 2.1. Additionally, new parameters driving ~~FBAPfungal-spore~~ emissions have been investigated. The emission flux depends on these parameters and their fitting coefficients obtained from a regression analysis of the FBAP measurements as described below. The new parameterization for ~~FBAP-fungal-spore~~ emissions has been incorporated into COSMO-ART and the resulting concentrations are included in Figures 3 to 6.

The adjusted parameterization for the emission flux ~~from the~~ is based on a regression analysis ~~is adjusted of to~~ an emission flux  $F_{F,c}$  estimated from the FBAP number concentration. For this, it is assumed that particles are evenly distributed throughout the planetary boundary layer and that the simulated ~~FBAPfungal-spore~~ concentration ~~negatively~~ correlates with  $h_{PBL}$ . This assumption is expected to be fulfilled best for days with a cloud-free convective boundary layer. Together with a steady-state condition and neglecting horizontal exchanges with the surrounding air, the balance holds between the number concentration ( $N_f$ ) and the emission rate ( $F_{F,c}$ ) together with the atmospheric lifetime of ~~fungal spores~~FBAP ( $\tau$ ):

$$N_f = \frac{F_{F,c}\tau}{h_{PBL}} \quad (7)$$

(Seinfeld and Pandis, 2006). The boundary layer height at the measurement site needs to be taken from the model simulation as it is not measured consistently. Eq. (7) is applicable to derive estimates for  $F_{F,c}$  under the assumption that the variation of the simulated boundary layer height explains a significant amount of the variation in  $N_f$ . This condition is fulfilled for four of the individual time series and to a lesser extent for the dataset as a whole (see section 3.1 and Table 3).

In eq. (7), Here, ~~the fungal-spore~~FBAP lifetime  $\tau$  represents a boundary layer mixing time and is not identical to an atmospheric residence time. For an initial test simulation,  $\tau$  is assumed to be constant and is estimated with an initial value of one day, as given in literature for atmospheric lifetimes of aerosol particles with 3  $\mu\text{m}$  in diameter (Jaenicke, 1978). A new simulation ofIn this case, the ~~FBAPfungal-spore~~ concentrations with the initial value of atmospheric-spore lifetime reveals an underestimation compared to the FBAP measurements. As a remedy,  $\tau$  is corrected to  $\tau = 4 \frac{3}{4}$  hours which can be understood as a mean time for boundary layer mixing of FBAPfungal spores. The deviation from a lifetime of 3  $\mu\text{m}$  particles given in literature may be attributed to the assumption of a constant vertical distribution of fungal spores with increasing altitude until boundary layer height. A typical vertical profile

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from the model (Figure 7) shows that this assumption is valid as a first approximation, but that deviations are visible both close to the surface and close to the top of the boundary layer. In the simulations, a mean ratio of approximately 1.75 between surface-level concentrations and mean concentrations within the boundary layer is found. ~~–which–This ratio is still~~ too small to explain the discrepancy between the expected lifetime and the value of  $\tau = 4 \frac{3}{4}$  hours which leads to agreement with the observed concentrations from which the emission flux was derived. The remaining discrepancy may be caused by the broad size distribution of FBAP particles (see Figure 1), by advection processes, by non-equilibrium conditions, and by the much longer lifetime of spores above the boundary layer. ~~and fungal spores above boundary layer may stay in atmosphere for longer times in lifetime. All discrepancies between this boundary layer mixing time and a typical atmospheric lifetime are caused by assumptions which are done for eq. (7). This difference may be caused by deviations from a well mixed constant concentration profile within the boundary layer, because source and removal processes in the simulation are not in equilibrium and fungal spores are continuously removed at the model boundaries.~~

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Two types of instruments operating with different numbers of channels and detecting fluorescence at different wavelengths are used here for deriving an emission parameterization appropriate for fungal spores. The technical difference may lead to slightly deviating ~~significantly different~~ FBAP concentrations (Healy et al., 2014), because the WIBS instrument only counts particles as FBAP when a signal exceeds a threshold in both channels (Pöhlker et al., 2012; Gabey et al., 2010). Some fungal spores most abundant in the Earth's atmosphere and very common for fungal spores of 2 - 4  $\mu\text{m}$  (*Cladosporium sp.*, *Aspergillus versicolor*, *Penicillium solitum*) (Fröhlich-Nowoisky et al., 2012; Hameed and Khodr, 2001) only show a weak signal in the emission wavelength of 310 nm to 400 nm (Saari et al., 2013; Healy et al., 2014). This difference needs to be taken into account when comparing absolute concentrations of fungal spores and FBAP. During the time periods shown here, the WIBS indicate slightly lower FBAP concentrations than the UV-APS when comparing to the model results. In general, this feature is not always valid and detailed side-by-side comparisons between the two types of instruments are required to determine their behavior in terms of FBAP detection and estimation of the PBAP concentration. In the attempt to factor out the

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technical difference between the instruments, we assume that the FBAP concentration can be multiplied by a constant factor for the concentration values to match each other. The amount of FBAP given by the UV-APS may be represented best by WIBS channel FL3\_370 (section 2.4). The FBAP concentration given by the UV-APS is therefore reduced by a factor derived from the WIBS instrument as the mean ratio between channel FL3\_370 and the total FBAP concentration  $N_{F,c}$  (channels FL1\_280 and FL3\_370). The ~~factor is estimated to be 2.2~~ average ratio between FL3\_370 and FL1\_280+FL3\_370 obtained and identical from the WIBS data at Karlsruhe and Manchester during the study periods is 2.2. ~~This factor difference~~ is not taken into account in the comparison of the time series in section 3.1, but corrected before applying eq. ~~(7)(7)~~ to the UV-APS data.

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Analyzing the meteorological and surface parameters of the model output, it was found that a better correlation with the measured FBAP concentrations is achieved for specific humidity rather than relative humidity, as it was reported for previous field measurements (Gabey et al., 2010; Toprak and Schnaiter, 2013; Di Filippo et al., 2013). During the time period in July 2010, the measured FBAP concentrations vary in a narrow range of specific humidity, which is not reproduced by the literature-based simulation. For this reason, the July case study was removed from the regression analysis. A dependence on the *LAI* is ~~assumed~~ required in order to take the seasonal change into account and to distinguish among various regions. A combination of *LAI* and specific humidity in the regression has the advantage of reducing the fitting parameters. The same relation was chosen by Heald and Spracklen (2009) for the previously discussed fungal spore emission parameterization. Additionally, surface temperature dependence as suggested by Di Filippo et al. (2013) is indicated by the time series and factored in a regression analysis. The parameters ( $b_1 = 20.426$  and  $b_2 = 3.93 \times 10^4$ ) are estimated by minimizing the sum of all squared residuals and result in a multiple linear regression giving an emission flux  $F_E = F_{FBAP}$  in  $\text{m}^2\text{s}^{-1}$  ~~for fungal spores~~ fitted to FBAP measurements:

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$$F_{FBAP} = b_1(T - 275.82 \text{ K}) + b_2 q_v LAI \quad (8)$$

where  $T$  is the surface temperature in K,  $q_v$  the specific humidity in  $\text{kg kg}^{-1}$ , and  $LAI$  the leaf area index in  $\text{m}^2 \text{m}^{-2}$ . The parameter inside the parentheses is related to an emission offset of the regression and covers unknown influences. The coefficient  $b_2$  is approximately the same as the constants in the Heald and Spracklen (2009) emission for a particle diameter of  $3 \mu\text{m}$  given in eq. ~~(6)(6)~~. The additional temperature dependence in eq. ~~(8)(8)~~ increases the fungal

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spore emission for temperatures above 275.8 K and lowers the emission for temperatures below this value.

The multiple linear regression (eq. (8)) yields a coefficient of determination of  $R^2 = 0.4$ .

~~By comparing the simulated concentrations (based on  $F_{FBAP}$ ) to the measured FBAP concentrations, it is found that they distribute more evenly along the 1:1 line (Figure 8b). The statistical overview (Figure 2) shows a better agreement between the median concentrations of simulation and measurement for the new emission parameterization than for the literature-based emissions, which is most obvious at Karlsruhe in August and at Hyytiälä in July. The new emission parameterization only slightly reduces the underestimations found for Hyytiälä during August.~~

### 3.3. Results of simulations with the new emission parameterization

$F_{FBAP}$  was implemented into COSMO-ART and applied in simulations for all three episodes. Note that the extrapolation of the emission parameterization to regions especially in Southern Europe, where ecosystems are different than at the stations where FBAP measurements were available, is uncertain and only valid under the assumption that temperature, specific humidity and LAI are universal proxies for the biological and meteorological drivers of the emission strength. This hypothesis should be tested by additional observations in the future.

~~Figure 9~~ Figure 9 shows the emission flux for late August 2010, following the new parameterization, horizontally distributed over a model domain covering Europe. Here, averaged over land areas of the domain,  $F_{FBAP}$  gives  $1.03 \times 10^3 \text{ m}^2 \text{ s}^{-1}$ . During July and October, the average flux is shifted to  $1.4 \times 10^3 \text{ m}^2 \text{ s}^{-1}$  and  $0.4 \times 10^3 \text{ m}^2 \text{ s}^{-1}$ , respectively, mainly as a result of seasonal changes of LAI and  $T$  (not shown).

-When analyzing the temporal development of the simulated fungal spore/FBAP concentrations for each time series,  $F_{FBAP}$  mostly results in a slightly higher number concentration than  $F_{H\&S}$  or  $F_{S\&D}$  (Figures 3 to 6). This is not the case for October 2010, where the  $F_{FBAP}$ -concentrations are in the range of the literature-based concentrations. A sharp decrease on 15 October at Hyytiälä, which is not reflected by ~~the literature-based simulations with literature-based emission fluxes~~, is caused by a rapid temperature change. ~~Comparison for the August case study show that simulated  $F_{FBAP}$ -concentrations agree well with measured FBAP concentrations without overestimating the measurement at Manchester and Killarney, where literature-based simulations and measurements already correspond to each other. Only~~

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a slight overestimation can be found at Manchester, which might be due to an urban measuring site that is not represented accurately by the model setup with its broad resolution.

For a statistical analysis of the results, the correlation coefficient  $R$  and its confidence intervals, the normalized mean bias  $NMB$ , and the root mean square error have been calculated for the simulations with the three different emission functions (Table 3). The results indicate that the bias (given by the  $NMB$ ) improves by the newly introduced emission function  $F_{FBAP}$ , but not the correlation  $R$  between simulated FBAP and observed FBAP concentrations which remains at the value of approximately 0.43 for the overall dataset. For the different time periods and stations,  $R$  varies between -0.17 and 0.66 (negative at only one station). Differences in  $R$  are small between  $N_{H\&S}$  and  $N_{FBAP}$ , because both make use of emission rates as a function of almost the same parameters ( $N_{FBAP}$  includes an additional  $T$ -dependence). The bias reduction and similar  $R$  is also visible in Fig. 8. Parameters  $b_1$  and  $b_2$  in eq. (8)(8) are chosen to give FBAP fungal spore concentrations matching best with measured FBAP concentrations, thus the reduction in  $NMB$  from -44% ( $N_{H\&S}$ ) to -0.4% ( $N_{FBAP}$ ). However, the bias reduction for the dataset as a whole is coincident with larger positive and negative biases for 4 out of the 8 episodes. Therefore, at the same time, the  $RMSE$  decreases only slightly from  $26.2 \text{ L}^{-1}$  ( $N_{H\&S}$ ) to  $23.2 \text{ L}^{-1}$  ( $N_{FBAP}$ ). At  $T = 275.82 \text{ K}$ ,  $F_{H\&S}$  is equal to  $F_{FBAP}$  for typical values of  $LAI$  and  $q_w$  and temperatures above this threshold (as it is the case for almost all locations) shift  $F_{FBAP}$  to give a larger emission flux. At meteorological conditions present for the selected cases, the second part of eq. (8)(8) dominates over the first part by a factor of  $\sim 4$  and therefore temperature changes have only a secondary influence on the emission flux. Hence,  $R^2$  is similar for both emission parameterizations.

Possible reasons for the only small improvements in  $R$  and  $RMSE$  may be biases in the input parameters to the parameterizations or model errors in the processes affecting fungal spore concentrations. To investigate this, the leaf area index was compared to independent observations.

In Table 4, the  $LAI$  from the Global Land Cover 2000 database (European Commission, Joint Research Centre, 2003), which was used in the COSMO simulations, is compared to the MODIS  $LAI$  product (MCD14A2, MODIS collection 5, Knyazikhin et al, 1998) which is available at 1 km horizontal and 8-day temporal resolution. Both nearest neighbor pixels and

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averages over 20x20 km with standard deviations are calculated. For the Karlsruhe station and for Hyytiälä in July and October, a good agreement is obtained. For Hyytiälä and Manchester in August, MODIS indicates a factor of approximately 2 lower LAI than assumed in our simulations. For Killarney, in contrast, the MODIS LAI is about 50% higher. While this possible error in LAI can explain part of the discrepancy between simulated and observed FBAP concentrations for Manchester, the bias in Hyytiälä is in the opposite direction as explainable by an error in LAI and thus has to be related to other error sources. In Killarney, model and measurements agree rather well despite the LAI underprediction.

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In addition, known deficiencies in the simulated boundary layer height may reduce the quality of the derived fit. The boundary layer height  $h_{PBL}$  had to be taken from the model simulations as it is not continuously available from observations at the stations. In the COSMO model,  $h_{PBL}$  is calculated by a bulk Richardson number method (Szintai and Kaufmann, 2008). It was shown for COSMO-2 simulations over the Swiss Plateau that this method overestimates  $h_{PBL}$  for convective boundary layers and strongly underestimates it for stable boundary layers (Collaud Coen et al, 2014).

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Finally, the role of precipitation on FBAP concentrations is ambiguous, as already discussed in section 3.1. At present, precipitation is not included as argument for the emission functions. Thus, they possibly neglect a driver of the variability.

~~Possible causes for the bias of  $F_{H\&S}$  and  $F_{S\&D}$  may come from different assumptions made to determine the fungal spore concentrations in ambient air. The mass size distribution of mannitol, which is used as a chemical tracer for fungal spores by Heald and Spracklen (2009), peaks in their study at particle diameters of  $\sim 5 \mu m$ . Additionally to fungal spores, bacteria, algae, lichens, and plant fragments, can produce mannitol and some of these can contribute to PBAP concentrations at  $\sim 5 \mu m$ . Similar assumptions are made for this study by linking FBAP to fungal spores, but chemical tracers vary between both studies. Furthermore, both literature-based emission fluxes compare local measurements to concentrations simulated on a global scale. Additional biases may arise when using these fluxes on a regional scale.~~

#### 3.3.3.4. Contribution of Fungal Spores to Near-surface Aerosol Composition

For a comparison of simulated [FBAP particles \(in the model treated as fungal spores\)](#) to the dry aerosol chemical composition, the fungal spore mass concentration is ~~calculated~~[estimated](#) from the number concentration assuming monodisperse and spherical particles ( ~~$d=3\mu\text{m}$~~ [,  \$d\_p=1\text{ g/cm}^3\$](#) ; section 2.1). The horizontally distributed near-surface (approximately 10 m above ground) [FBAP/fungal spore number concentration using  \$F\_{FBAP}\$](#)  is shown in [Figure 10](#)~~Figure 10~~. Concentrations simulated at the measurement locations are considerably lower than the high surface concentrations in the southern part of the model domain.

The simulated mass concentrations of each chemical aerosol compound are averaged over the land areas of the model domain and the time period of late August 2010 (~~Figure 11~~[Figure 14](#)).

The total aerosol mass concentration is approximately  $2.5\text{ }\mu\text{g/m}^3$ . Fungal spores distribute in the domain with an average number concentration of  $26\text{ L}^{-1}$  over land. This corresponds to an average mass concentration of fungal spores of  $0.37\text{ }\mu\text{g/m}^3$  which accounts for 15.4% of the total simulated aerosol mass. The total aerosol mass excludes mineral dust as one of the main contributors to the chemical aerosol composition, which might lower the fraction of fungal spore mass ~~ss considerably~~. A list of mass concentrations of the simulated chemical aerosol compounds, including fungal spores, ~~occurring~~ at the [four](#) measurement sites, is given in [Table 5](#)~~Table 3~~. The [mass](#) fraction of fungal spores [compared to the total aerosol](#) simulated for these sites varies between 9% and ~~320%~~. ~~FBAP mass concentrations calculated from the measured FBAP number concentrations are also listed in Table 35. Here, the same spherical particle diameter and particle density as for the fungal sporesimulation is assumed. The FBAP number concentrations are averaged over the same time period as covered by the aerosol simulation. Their share of the aerosol mass ranges from 5% at Manchester up to 64% at Hyytiälä. For Karlsruhe and Killarney, the fractions calculated from the measurements are in good agreement with the fractions resulting from the simulated fungal spore mass concentrations.~~

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#### 4. Discussion and Conclusions

FBAP measurements from four locations in ~~Northern~~ Europe were compared with simulated ~~fungal spore~~ concentrations of fungal spores and FBAP. ~~Fluorescent particles are often observed to be most abundant~~ in the diameter range of 2 - 4  $\mu\text{m}$  ~~are highest in number concentration of FBAP measurements at the rural site near Karlsruhe, Germany~~ (Huffman et al., 2010; Pöschl et al., 2010; Huffman et al., 2012; Healy et al., 2012a; Toprak and Schnaiter, 2013; Huffman et al., 2013). ~~This~~ diameter range for peak FBAP concentration matches closely with the modal size of many species of fungal spores known to be present in airborne concentrations the atmospheric aerosol. ~~Simulated fungal spores have been adjusted to match this diameter.~~ Contrary to that, an increase in number concentration towards small particles has been reported for some FBAP measurement series, although but only a small fraction of FBAP particles could be identified ~~counted~~ as bacteria cells (Gabey et al., 2011; Huffman et al., 2010). Therefore, FBAP cannot be equated with fungal spores, although the concentrations of these may agree well in many cases. This complicates the interpretation of the comparisons conducted in this study.

Comparison of simulations and measurements at four locations and the correlation of FBAP concentrations to meteorological and surface conditions are expected to be most robust when applying identical methods and conditions at all locations. These conditions were not fulfilled in our study. On one hand, site characteristics vary between the stations, which will ~~may~~ influence the sensitivity of PBAP emission to surrounding conditions. On the other hand, the measurements are made with different instruments. The measurement series at Karlsruhe, Germany, are done with a WIBS-4 instrument which includes technical improvements compared to the WIBS-3 used at Manchester, UK and Cork, Ireland (Gabey et al., 2010; Healy et al., 2012b). At Hyytiälä, Finland, and Killarney, Ireland, the UV-APS is used to determine the FBAP concentration. This variation may lead to different estimation of the FBAP concentration and within this case study WIBS may report FBAP at lower concentrations than UV-APS at different locations but similar meteorological conditions.

In this paper, fungal spore concentrations are calculated with the COSMO-ART atmospheric model by using literature-based emission parameterizations which are based on the adaptation of simulated global atmospheric concentration to mannitol measurements or spore colony counts (Heald and Spracklen, 2009; Sesartic and Dallafior, 2011). Although mannitol concentration can include contributions from other PBAP (e.g. insects, bacteria, and algae)

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1 and from lower plants, the association to fungal spore concentration is reasonable (Di Filippo  
2 et al., 2013). ~~Some differences in the comparison may occur from the usage of FBAP~~  
3 ~~concentrations as a representative for fungal spores.~~ Overall, ~~T~~he temporal development of  
4 the literature-based simulated fungal spore concentrations calculated by COSMO-ART  
5 approximately reproduces the measured FBAP concentrations. ~~Some differences in the~~  
6 ~~comparison between simulated fungal spore and observed FBAP concentrations may be~~  
7 ~~explainable by occur from the usage of FBAP concentrations as a representative for fungal~~  
8 ~~spores.~~

9 By using a time-independent (but spatially varying) emission flux  $F_{S\&D}$ , every development in  
10 the local temporal pattern arises from meteorological influences. A similar diurnal cycle  
11 develops between simulations with constant ( $F_{S\&D}$ ) and time-dependent ( $F_{H\&S}$  and  $F_{FBAP}$ )  
12 fungal spore emissions, but the diurnal amplitude differs to varying extent. ~~Therefore and~~  
13 ~~from by visually comparing to a correlation analysis with the inverse of the simulated~~  
14 ~~boundary layer height, a~~ diurnal cycle in the simulated fungal spore concentrations with a  
15 maximum between midnight and sunrise is probably shown to be influenced at least partly by  
16 boundary layer compression at night. However, measured FBAP concentrations are in some  
17 time periods not consistent with the simulated  $h_{PBL}$ , which suggests that  $N_{F,c}$  is additionally  
18 influenced e. g. by possible increases in biological emission at night.

19 ~~In this work, The purpose of the work reported here was to develop~~ a new emission  
20 parameterization for FBAP particles is developed by a regression analysis to observations. for  
21 ~~fungal spores, because literature-based emissions for fungal spores have been found to~~  
22 ~~significantly underestimate measured FBAP concentrations. The parameterization is therefore~~  
23 ~~adjusted to the FBAP concentrations (section 3.2).~~ As was formulated by Heald and  
24 Spracklen (2009), it depends on the specific humidity and the leaf area index, but is extended  
25 by temperature. The resulting concentrations have on the average a smaller normalized mean  
26 bias and a slightly smaller root mean square error compared to the measured FBAP  
27 concentrations than the previously used fungal spore emission parameterizations, but  
28 variations in the measurements are not always captured by the simulation. Thus, the  
29 correlation coefficient remains low (0.43). Possible reasons include biases due to local  
30 conditions at the measurement stations, biases in the input parameters and model errors  
31 related to the transport, mixing and sink processes (including boundary layer turbulence and  
32 washout by precipitation). Future work including a long-term analysis of FBAP

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concentrations and [environmental](#) conditions may result in a further adjustment of the [coefficients](#) or reveal [other parameters or functional dependencies](#) driving the emission. [Ideally, the observations would be split into a training data set and an evaluation data set, which was not possible until now due to the limited amount of available observations.](#)

Using the new emission parameterization on a model domain for Europe, [FBAP](#)~~fungal spore~~ emission fluxes are extrapolated from northern parts of the domain, where UV-LIF measurements were located, also to Southern Europe. There, much higher emission fluxes occur in the simulation, partially caused by higher specific humidity, which is also the case for  $F_{H\&S}$ , as well by temperature dependence in  $F_{FBAP}$ . This extrapolation is without [support from](#) local Southern European measurements, however, and thus further UV-LIF measurements are recommended for this region in Southern Europe where fungal spore emission fluxes are potentially ~~greater~~[larger](#).

As a result of the relatively low [model](#) horizontal resolution [of 14 km](#), small-scale variations influencing fungal spore [and FBAP](#) emission at the measurement sites may not be resolved. Influences on a small scale might be due to an increased amount of fungi for the given vegetation type. When taking the leaf area index as a surrogate for the vegetation type, uncertainties may result from an insufficient relation to the presence of fungi or additional surrounding factors favoring fungi growth. Furthermore, variations in precipitation may not be captured by the model, which then may lead to ~~improper biases in the~~ fungal spore concentrations. The same holds for small wind gusts and convective cells which may have a strong influence on spore dispersion, but are not captured well in the model. An increase in [FBAP and](#) fungal spore concentration during or shortly after rain events (Huffman et al., 2013) could not be reproduced by the simulations, as this effect is not included in the emission parameterizations to an adequate extent.

The module calculating the dispersion of [FBAP](#)/fungal spores does not include all processes of aerosol dynamics and cloud physics. Of the processes not included, only breaking up of spores can enhance their number concentration. Coagulation is neglected, as in most cases the [FBAP](#)/fungal spore number concentration is low and, hence, their collision is highly improbable. A coagulation of spores with other aerosol particles is more likely to happen, but not included in the simulations. Not much is known about the role of fungal spores [and other biological particles](#) in clouds and their ability to act as cloud condensation nuclei.

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The simulations presented in this paper highlight the importance of PBAP to the composition of atmospheric aerosol. Fungal spores, the focus of this paper, are among the main contributors to PBAP and therefore exert significant influence on aerosol loading. In this study, COSMO-ART is used<sup>up</sup> to simulate all major chemical aerosol compounds except for mineral dust in a domain covering Western Europe. When averaging the mass concentration horizontally across the land-covered part of the model domain and over all time steps of the simulation, fungal spores (assumed to be represented by FBAP) are among the major mass components (Figure 11~~Figure 14~~). However, the mass fraction of fungal spores might be overestimated here, as another major aerosol component, mineral dust, is not included, because the domain does not include any desert dust source areas. At the selected time periods, back trajectories with HYSPLIT (Draxler and Rolph, 2013) and the operational dust forecast with the BSC-DREAM8b model, operated by the Barcelona Supercomputing Center (<http://www.bsc.es/earth-sciences/mineral-dust-forecast-system/>) —suggest that transport of Sahara dust to the measurement stations into the domain is low (not shown), but influences the aerosol concentrations in Southern Spain. The emission from non-desert dust sources (in particular agricultural areas) is strongly episodic, as it is linked not only to meteorological conditions, but also to human activity (e. g. tillage operations (Funk et al, 2008, Goossens et al, 2001), traffic). The magnitude of the local dust source strength is uncertain and not included in the model, as no validated emission parameterization is available at present. Beuck et al (2011) estimated that mineral dust (including both long-range transport and local sources) corresponds to on average 10% of PM10 concentrations in rural North-West Germany, and 14% of PM10 concentrations in urban areas in this region.

The ~~FBAP—fungal spore~~ mass concentration, ~~estimated from measured FBAP number concentrations ( $d_p = 3 \mu\text{m}$ ;  $\rho_p = 1 \text{ g/cm}^3$ ),~~ may reach up to ~~30~~64% of simulated near-surface chemical aerosol mass components in rural areas of Finland (~~Table 3~~Table 5). In comparison, ratios of PBAP to total aerosol concentrations given in literature can assume similar values. The volume fraction of biological particles among~~of~~ particles larger than  $0.2 \mu\text{m}$  during one year of measurements at a remote site in Siberia reaches 28% on the average and at Mainz, Germany the volume fraction amounts to 22% (Matthias-Maser et al., 2000). Both of these fractions agree well with simulated mass fractions of this study for comparable locations, although simulated concentrations given in this study are much lower than total PBAP number concentrations given in Matthias-Maser et al. (2000). In contrast, the number and mass fractions in the Amazonian basin are above 80% (Pöschl et al., 2010) and therefore

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much higher than in the ~~highlighted~~ urban and remote areas ~~in this study (Pöschl et al., 2010), but here the absolute concentrations are smaller and therefore in the order of magnitude given by the simulation of this study.~~

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PBAP and especially fungal spores might account for a major part of the aerosol loading. Locally, ~~a correlations~~ between ~~increasing~~ FBAP and ice nuclei number concentration (Tobias et al., 2013) shows that future model studies of PBAP impacts on clouds are needed to determine their relevance to atmospheric ice nucleation.

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Table 1. Overview of the measurement sites, including their geographical location and the types of instrument used ( $d_p$  corresponds to the optical particle diameter and  $d_a$  to the aerodynamic particle diameter). The sections below show the simulation periods and the availability of data at this site (filled dot). Mean values for the simulated meteorological and surface conditions used for the new emission parameterization (section 3.2) at the measurement site during the corresponding time periods are added to each section.

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location	Karlsruhe, Germany	Hyytiälä, Finland	Manchester, UK	Killarney, Ireland
coordinates	49° 5' 43.6" N 8° 25' 45.0" E	61° 50' 41.0" N 24° 17' 17.4" E	53° 27' 57.0" N 2° 13' 56.0" W	52° 3' 28.0" N 9° 30' 16.4" W
altitude	111 m a.s.l.	152 m a.s.l.	45 m a.s.l.	34 m a.s.l.
instrument	WIBS-4	UV-APS	WIBS-3	UV-APS
size range	$0.8 \leq d_p \leq 16 \mu\text{m}$	$1 < d_a \leq 20 \mu\text{m}$	$0.8 \leq d_p \leq 20 \mu\text{m}$	$1 < d_a \leq 20 \mu\text{m}$
22 July 2010 - 28 July 2010	•	•	○	○
LAI ( $\text{m}^2/\text{m}^2$ )	3.18	3.72	-	-
mean T ( $^{\circ}\text{C}$ )	17.3	16.2	-	-
mean $q_v$ (kg/kg)	0.0088	0.0108	-	-
26 August 2010 -01 September 2010	•	•	•	•
LAI ( $\text{m}^2/\text{m}^2$ )	2.94	3.4	2.87	2.06
mean T ( $^{\circ}\text{C}$ )	16.6	8.5	11.6	11.1
mean $q_v$ (kg/kg)	0.0099	0.0067	0.0073	0.0072
11 October 2010 - 21 October 2010	•	•	○	○
LAI ( $\text{m}^2/\text{m}^2$ )	1.49	1.27	-	-
mean T ( $^{\circ}\text{C}$ )	6.5	-0.6	-	-
mean $q_v$ (kg/kg)	0.0055	0.0034	-	-

1 Table 2. Correlation coefficient (R), root mean square error (RMSE) and normalized mean bias (NMB) for  
2 correlations between simulated fungal spore/FBAP and measured FBAP concentrations at different locations and  
3 three different time periods.

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	N <sub>H&amp;S</sub>			N <sub>S&amp;D</sub>			N <sub>FBAP</sub>			Confidence intervals for R	
	R	RMSE [L <sup>-1</sup> ]	NMB [%]	R	RMSE [L <sup>-1</sup> ]	NMB [%]	R	RMSE [L <sup>-1</sup> ]	NMB [%]	95%	99%
<b>Karlsruhe, Jul10</b>	0.07	45	-58	0.11	48	-66	0.07	38	-32	±0.17	±0.22
<b>Karlsruhe, Aug10</b>	0.11	25	-36	-0.07	28	-40	0.11	30	1.4	±0.16	±0.20
<b>Karlsruhe, Oct10</b>	-0.11	22	-64	0.01	21	20	-0.17	17	-35	±0.18	±0.23
<b>Hyytiälä, Jul10</b>	0.57	17	3.2	0.58	24	-67	0.57	23	51	±0.11	±0.15
<b>Hyytiälä, Aug10</b>	0.04	20	-62	0.19	22	-76	0.00	18	-46	±0.16	±0.21
<b>Hyytiälä, Oct10</b>	0.01	4.7	-58	-0.10	5.1	35	0.23	4.6	-59	±0.17	±0.23
<b>Manchester, Aug10</b>	0.68	15	2.7	0.65	19	38	0.66	22	63	±0.09	±0.12
<b>Killarney, Aug10</b>	0.49	6.6	84	0.32	15	200	0.45	15	214	±0.13	±0.17
<b>all</b>	<b>0.428</b>	<b>26.2</b>	<b>-44.04</b>	<b>-0.05</b>	<b>25.6</b>	<b>-28.74</b>	<b>0.433</b>	<b>23.2</b>	<b>-0.43</b>	<b>±0.05</b>	<b>±0.06</b>

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1 Table 23. Correlation coefficient (R) and confidence intervals for correlations between observed FBAP  
 2 concentrations and the inverse boundary layer height at different locations and three different time periods.

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	<i><b>R</b></i>	<i><b>95% confidence interval</b></i>	<i><b>99% confidence interval</b></i>
Karlsruhe, Jul10	0.11	±0.16	±0.21
Karlsruhe, Aug10	-0.03	±0.16	±0.21
Karlsruhe, Oct10	-0.12	±0.15	±0.19
Hyytiälä, Jul10	0.35	±0.15	±0.19
Hyytiälä, Aug10	0.32	±0.14	±0.18
Hyytiälä, Oct10	-0.20	±0.18	±0.23
Manchester, Aug10	0.52	±0.12	±0.15
Killarney, Aug10	0.29	±0.14	±0.19
<b>All</b>	0.11	±0.06	±0.07
<b>all, without h&lt;10 m</b>	<b>0.18</b>	<b>±0.06</b>	<b>±0.07</b>

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1 Table 4. Leaf area index ( $\text{m}^2/\text{m}^2$ ) from the climatological GLC2000 dataset as used in the simulations compared  
2 to the nearest neighbor pixel of the 8-day, 1 km resolution MODIS LAI and to an average of the 1 km resolution  
3 MODIS LAI over a 20x20 km area centered on the station gridpoint. In addition, the standard deviation within  
4 that 20x20 km area is given. For the Manchester nearest neighbour 1km gridpoint, no data are available for the  
5 given time period.

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		Karlsruhe	Hyytiälä	Manchester	Killarney
22 Jul–28 Jul 2010	GLC2000	3.18	3.72	-	-
	MODIS 1km	2.30	4.80	-	-
	MODIS 20km average	2.70	3.77	-	-
	MODIS 20km std dev	1.64	1.34	-	-
26 Aug–1 Sep 2010	GLC2000	2.94	3.40	2.87	2.06
	MODIS 1km	3.46	1.59	n.d.	3.20
	MODIS 20km average	2.62	1.65	1.28	3.15
	MODIS 20km std dev	1.48	1.10	0.66	1.97
11 Oct–21 Oct 2010	GLC2000	1.49	1.27	-	-
	MODIS 1km	1.62	0.80	-	-
	MODIS 20km average	1.41	1.32	-	-
	MODIS 20km std dev	0.93	0.96	-	-

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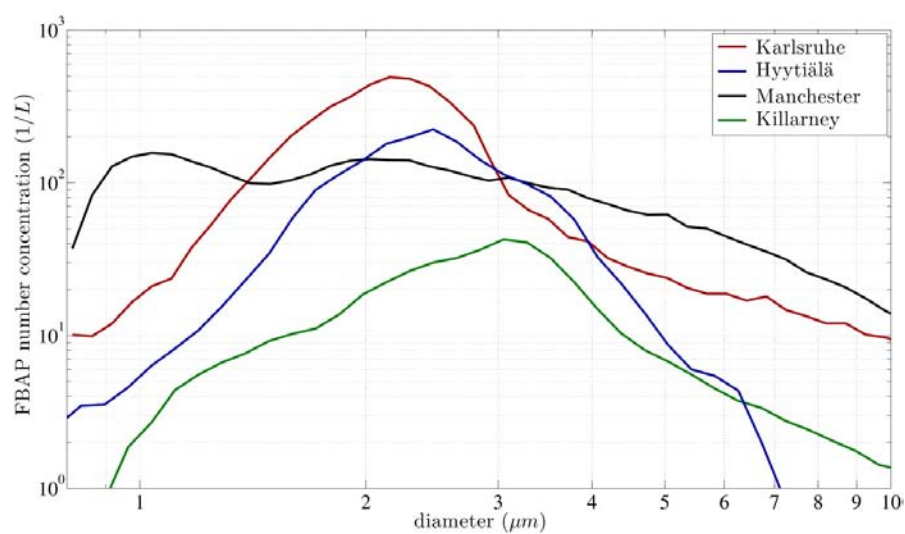


1 Table 35. Simulated aerosol mass concentrations for aerosol chemical components, including fungal spores,  
2 ~~together with measured FBAP values~~ in  $\mu\text{g}/\text{m}^3$  at the measuring sites as averages over the time period during  
3 August 2010

Particle Mass ( $\mu\text{g}/\text{m}^3$ )	Karlsruhe, Germany	Hyytiälä, Finland	Manchester, UK	Killarney, Ireland
<del>measured FBAP</del>	<del>0.46</del>	<del>0.81</del>	<del>0.19</del>	<del>0.19</del>
simulated fungal spores	0.41	0.20	0.35	0.28
sea salt	0.44	0.01	1.62	1.11
soot	0.19	0.06	0.42	0.04
$\text{SO}_4^{2-}$	0.18	0.01	0.11	0.05
$\text{NH}_4^+$	0.44	0.01	0.14	0.07
$\text{NO}_3^-$	1.29	0.01	0.34	0.18
SOA	0.41	0.24	0.14	0.04
aPOA	0.67	0.13	0.85	0.11

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3 Figure 1. Average FBAP size distributions derived from UV-LIF measurements during case studies in August  
 4 2010 at Karlsruhe (Germany) [with WIBS-4](#), Hyytiälä (Finland) [with UV-APS](#), Manchester (UK) [with WIBS-3](#),  
 5 and Killarney (Ireland) [with UV-APS](#).

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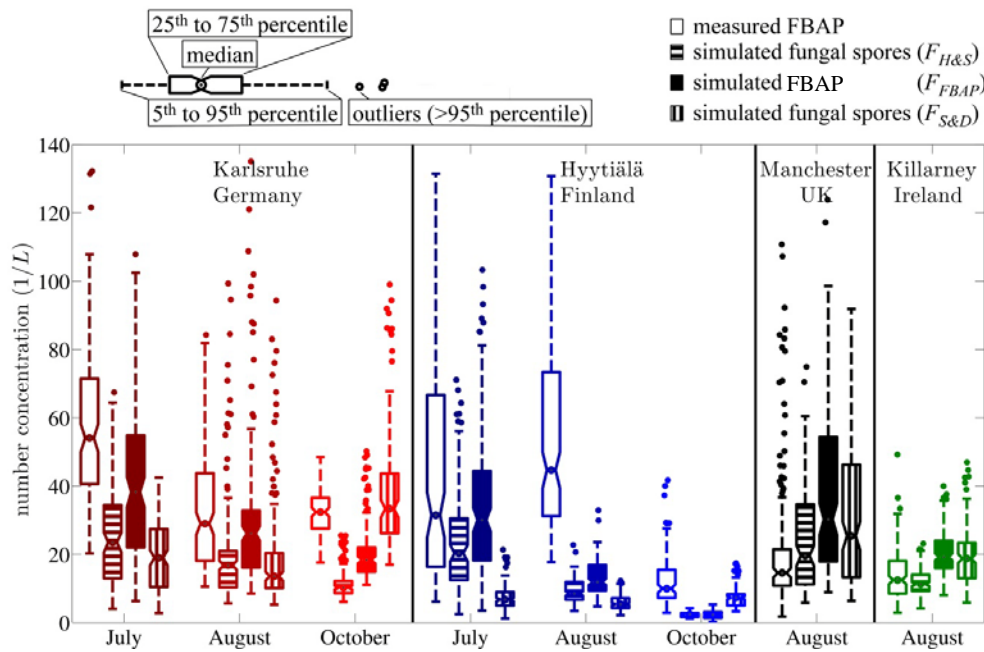


Figure 2. Box-whisker plots for all case studies of: measured hourly FBAP concentration (open boxes), simulated fungal spore concentration with three different emission parameterizations:  $F_{H\&S}$  from Heald and Spracklen (2009, horizontally hatched boxes);  $F_{FBAP}$  from this study (filled boxes);  $F_{S\&D}$  from Sesartic and Dallafior (2011, vertically hatched boxes), emitted by  $F_{H\&S}$  (horizontally hatched boxes),  $F_{FBAP}$  (filled boxes), and  $F_{S\&D}$  (vertically hatched boxes) for all case studies. The central mark of each box shows the median, its edges the 25<sup>th</sup> and 75<sup>th</sup> percentiles, and whiskers show 5<sup>th</sup> and 95<sup>th</sup> percentiles. Dots above whisker show outliers (>95<sup>th</sup> percentile).

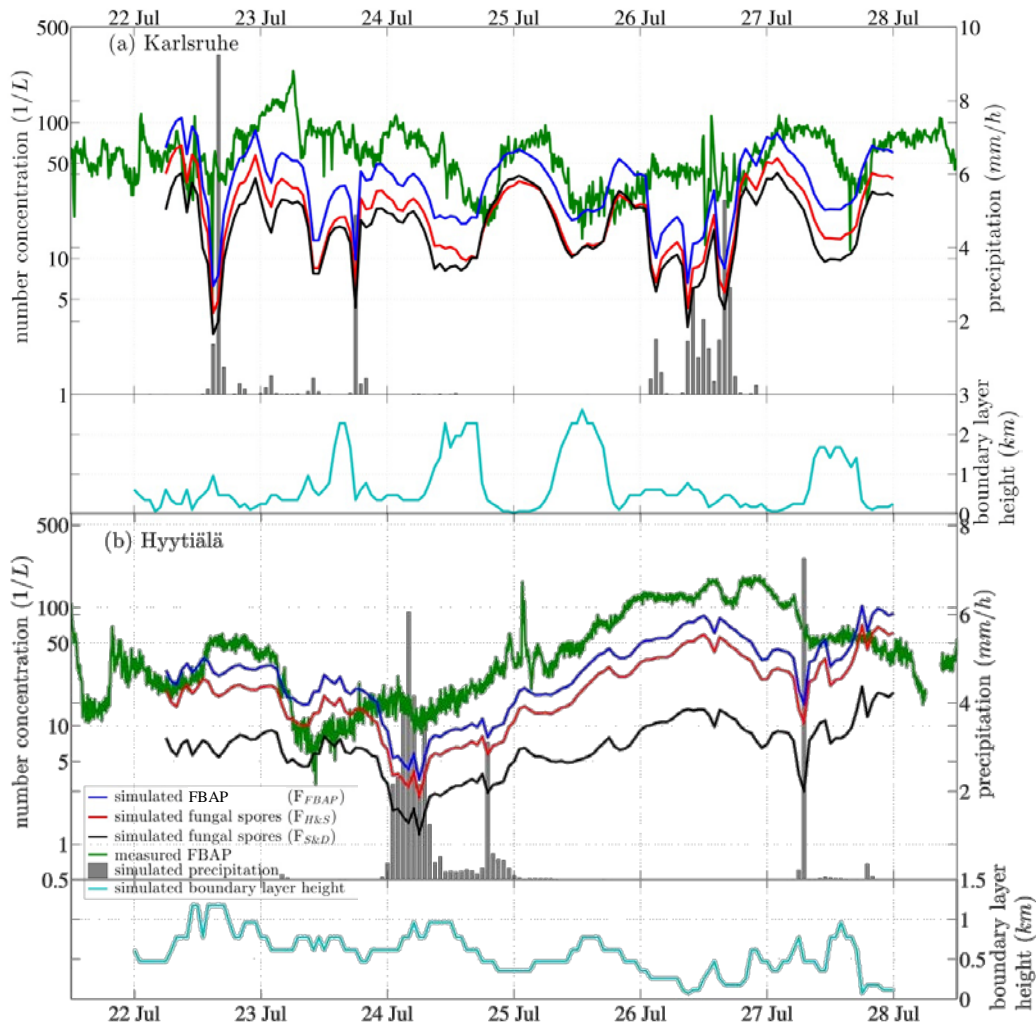


Figure 3. Time series of measured FBAP and simulated fungal spore/FBAP number concentrations in 1/L together with simulated precipitation in mm/h (right axis) and simulated boundary layer height in km (right axis) during the case study from 22 July to 28 July 2010 at (a) Karlsruhe, Germany and (b) Hyytiälä, Finland. Simulations were performed with three different emission parameterizations:  $F_{H&S}$  from Heald and Spracklen (2009);  $F_{S&D}$  from Sesartic and Dallafior (2011);  $F_{FBAP}$  from this study.

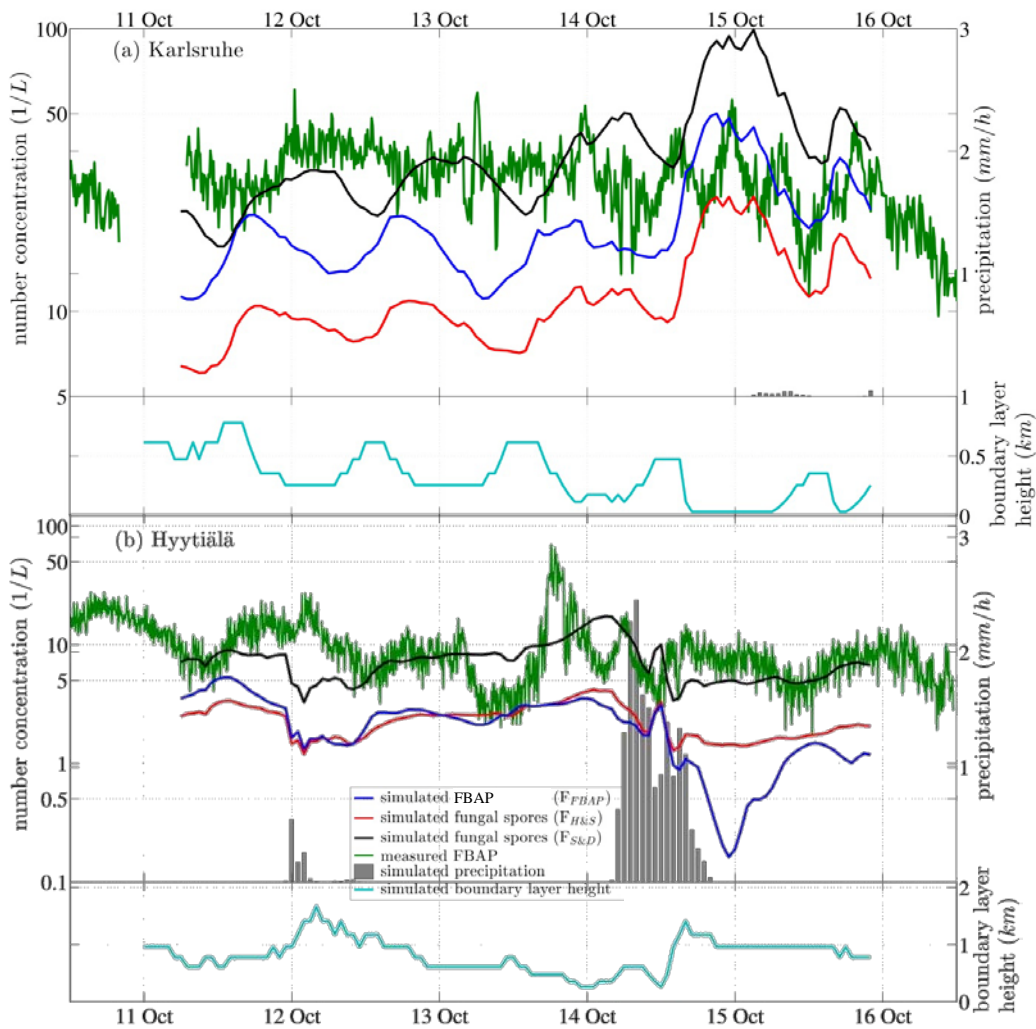


Figure 4. Time series of measured FBAP and simulated fungal spore/FBAP number concentrations in  $1/L$  together with simulated precipitation in  $mm/h$  (right axis) and simulated boundary layer height in  $km$  (right axis) during the case study from 11 October 2010 to 21 October 2010 at (a) Karlsruhe, Germany and (b) Hyytiälä, Finland. Simulations were performed with three different emission parameterizations:  $F_{H\&S}$  from Heald and Spracklen (2009);  $F_{S\&D}$  from Sesartic and Dallafior (2011);  $F_{FBAP}$  from this study.

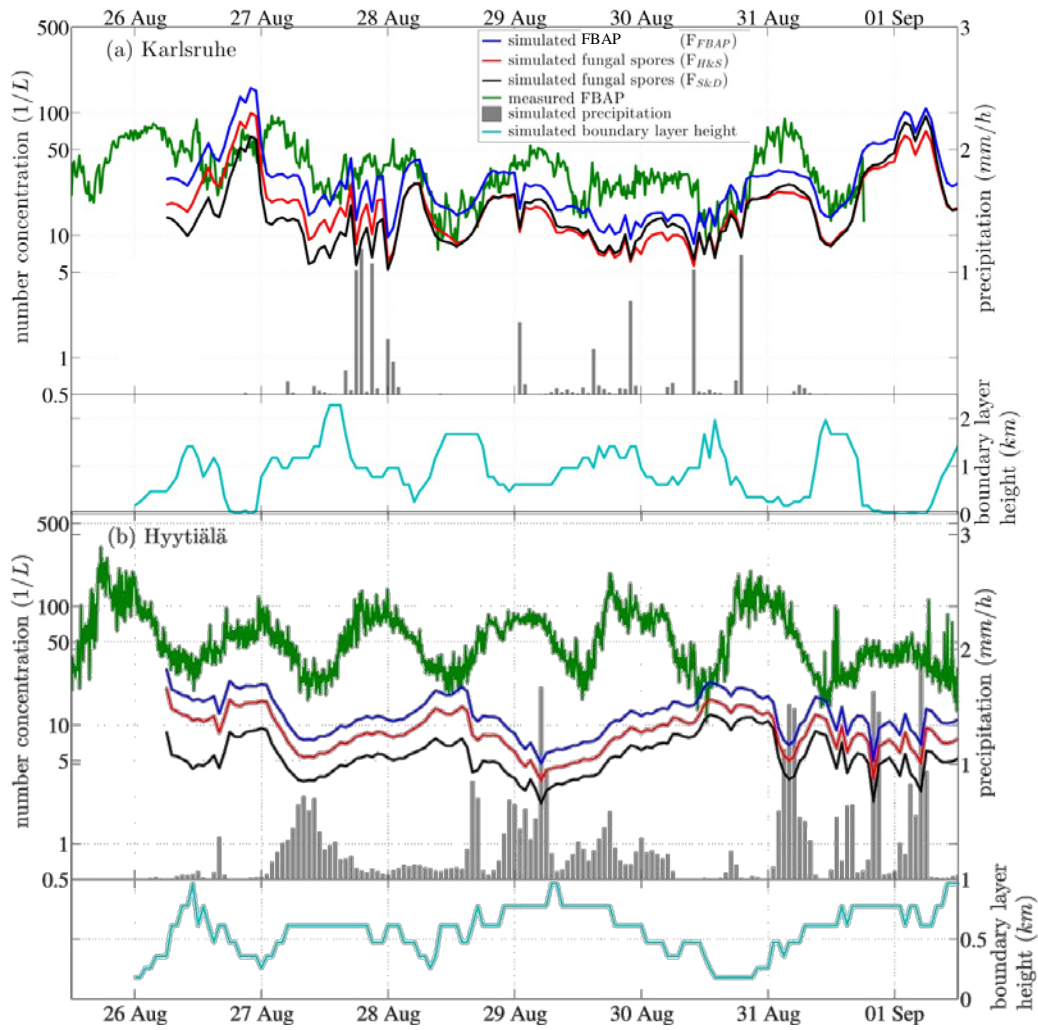


Figure 5. Time series of measured FBAP and simulated fungal spore/FBAP number concentrations in 1/L together with simulated precipitation in mm/h (right axis) and simulated boundary layer height in km (right axis) during the case study from 26 August 2010 to 01 September 2010 at (a) Karlsruhe, Germany. (b) Hyytiälä, Finland. Simulations were performed with three different emission parameterizations:  $F_{H\&S}$  from Heald and Spracklen (2009);  $F_{S\&D}$  from Sesartic and Dallafior (2011);  $F_{FBAP}$  from this study.

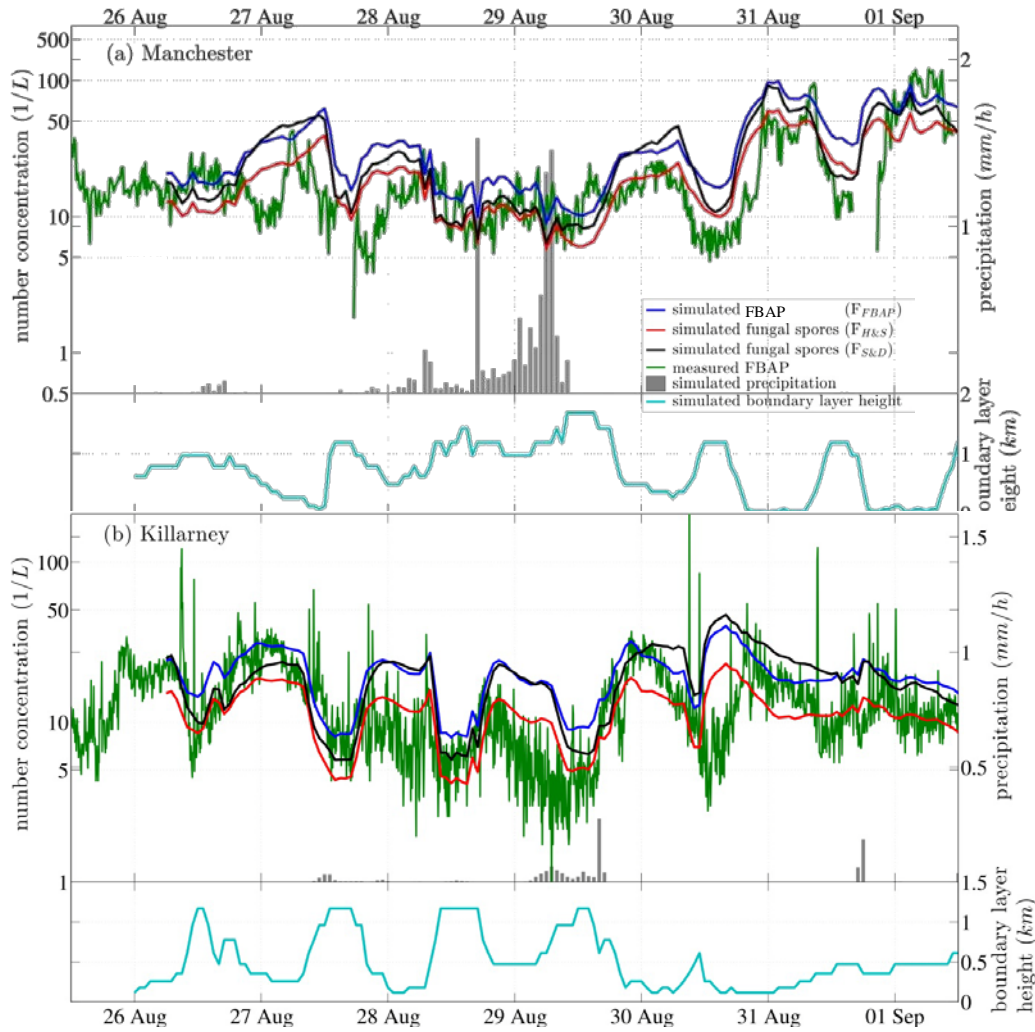
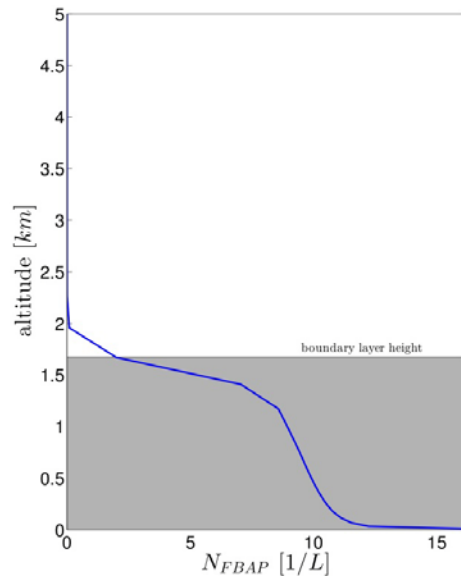


Figure 6. Time series of measured FBAP and simulated fungal spore number concentrations in 1/L together with simulated precipitation in mm/h (right axis) and simulated boundary layer height in km (right axis) during the case study from 26 August 2010 to 01 September 2010 at (a) Manchester, UK and (b) Killarney, Ireland. Simulations were performed with three different emission parameterizations:  $F_{H&S}$  from Heald and Spracklen (2009);  $F_{S&D}$  from Sesartic and Dallafior (2011);  $F_{FBAP}$  from this study.

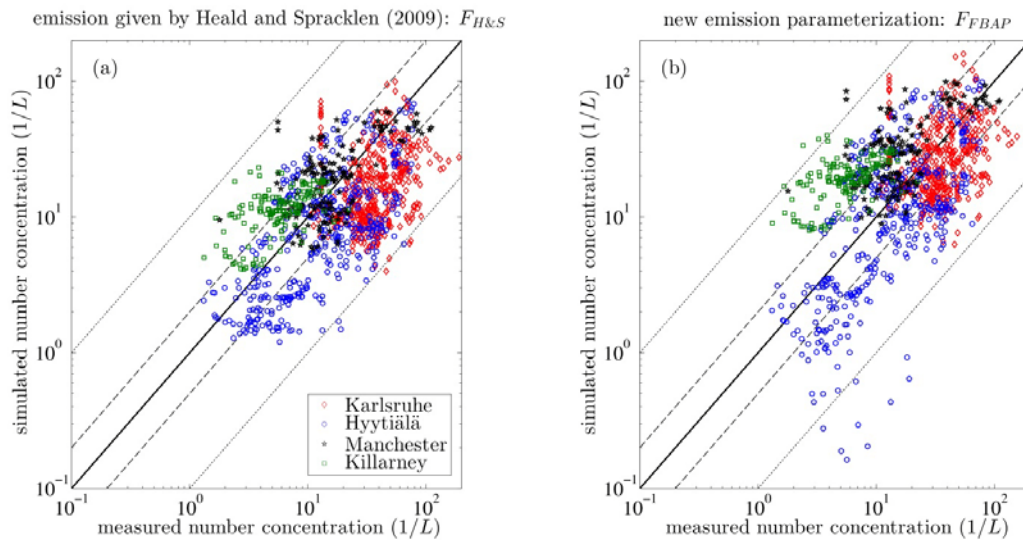


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2 | Figure 7. Exemplary vertical profile of simulated [fungal spore FBAP](#) concentration within and above the  
3 planetary boundary layer for Karlsruhe at 28 Aug 2010 14 UTC.

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3 Figure 8. Comparison for all case studies: Measured FBAP number concentrations plotted versus simulated  
 4 [fungal spore](#) number concentrations of (a) [fungal spores](#) based on [the](#) Heald and Spracklen (2009) emission flux  
 5 and (b) [FBAP](#) based on [the new](#) emission parameterization derived from a multiple linear regression to FBAP  
 6 concentrations. Solid black lines represent the 1:1-line, dashed lines the 1:2-line and dotted lines the 1:10-line.

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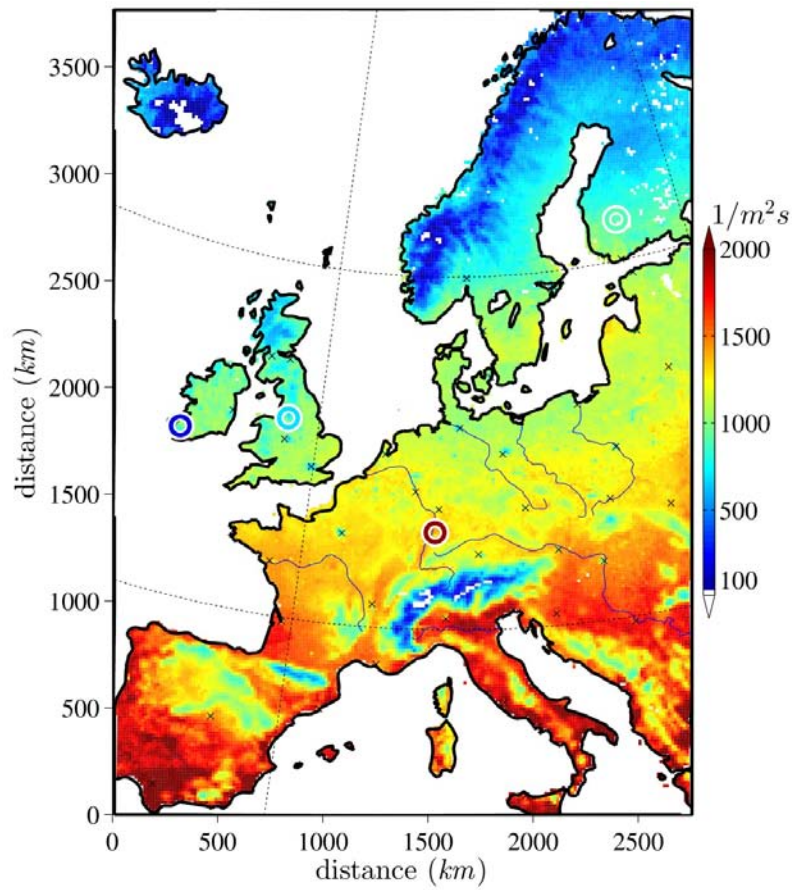


Figure 9. Average simulated ~~fungal spore~~FBAP emission flux ( $F_{FBAP}$ ) in  $\text{m}^2\text{s}^{-1}$  from 26 August to 01 September 2010, (excluding a spin-up period of 6 hours). ~~White~~Circles indicate the locations of the different FBAP measurement time series and the color within the white circles represents the mean emission flux calculated from FBAP measurements at each location.

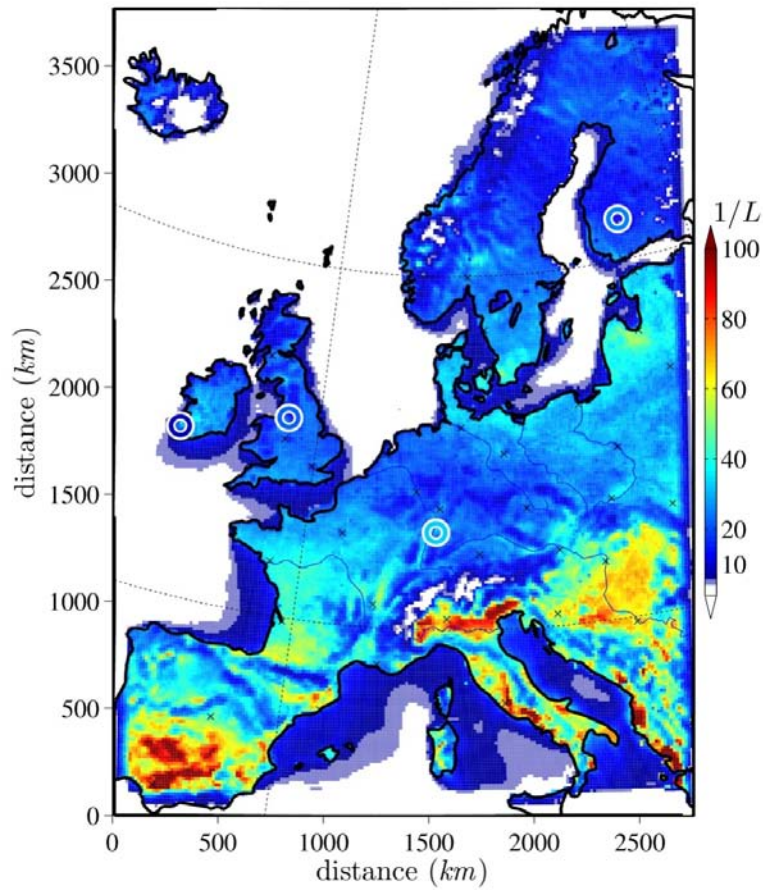
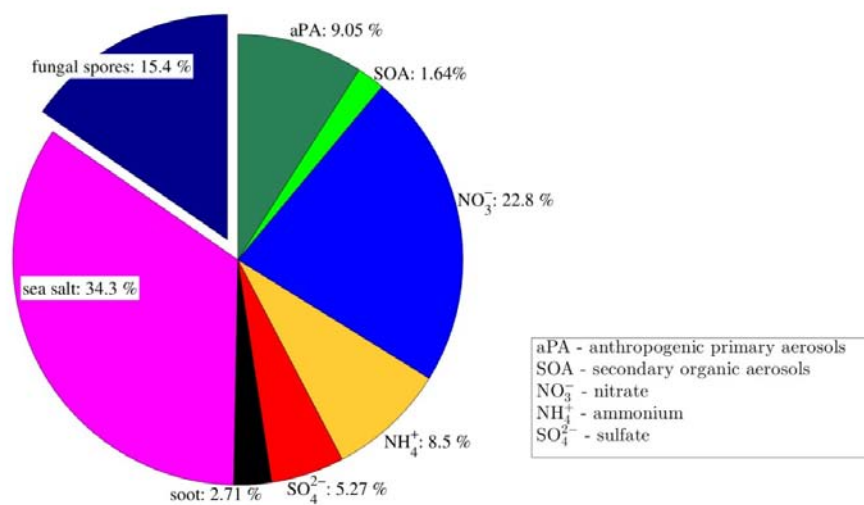


Figure 10. Horizontally distributed  $FBAP$ /fungal spore concentration in  $1/L$ , emitted by  $F_{FBAP}$ , in the lowest model layer, averaged from 26 August to 01 September 2010 (excluding a spin-up period of 6 hours). White circles indicate the locations of the different  $FBAP$  measurement time series and the color within the white circles represents the mean  $FBAP$  number concentration measured at each location.

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Figure 11. Near-surface chemical aerosol mass composition simulated by COSMO-ART, horizontally averaged over the land area in the model domain and temporally averaged from 26 August to 01 September 2010 (excluding a spin-up period of 6 hours)