

Improved model of  
isoprene emissions  
in Africa

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# Improved model of isoprene emissions in Africa using OMI satellite observations of formaldehyde: implications for oxidants and particulate matter

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## Abstract

We use a 2005–2009 record of isoprene emissions over Africa derived from OMI satellite observations of formaldehyde (HCHO) to better understand the factors controlling isoprene emission on the scale of the continent and evaluate the impact of isoprene emissions on atmospheric composition in Africa. OMI-derived isoprene emissions show large seasonality over savannas driven by temperature and leaf area index (LAI), and much weaker seasonality over equatorial forests driven by temperature. The commonly used MEGAN (version 2.1) global isoprene emission model reproduces this seasonality but is biased high, particularly for equatorial forests, when compared to OMI and relaxed-eddy accumulation measurements. Isoprene emissions in MEGAN are computed as the product of an emission factor  $E_o$ , LAI, and activity factors dependent on environmental variables. We use the OMI-derived emissions to provide improved estimates of  $E_o$  that are in good agreement with direct leaf measurements from field campaigns ( $r = 0.55$ , bias =  $-19\%$ ). The largest downward corrections to MEGAN  $E_o$  values are for equatorial forests and semi-arid environments, and this is consistent with latitudinal transects of isoprene over West Africa from the AMMA aircraft campaign. Total emission of isoprene in Africa is estimated to be  $77 \text{ TgCa}^{-1}$ , compared to  $104 \text{ TgCa}^{-1}$  in MEGAN. Simulations with the GEOS-Chem oxidant-aerosol model suggest that isoprene emissions increase mean surface ozone in West Africa by up to 8 ppbv, and particulate matter by up to  $1.5 \mu\text{g m}^{-3}$ , due to coupling with anthropogenic influences.

## 1 Introduction

Isoprene is the dominant biogenic non-methane volatile organic compound (NMVOC) emitted by vegetation, accounting for about 50 % of global NMVOC emissions in current inventories (Olivier et al., 1996; Guenther et al., 2006). Isoprene affects the oxidative capacity of the atmosphere through reaction with OH (Ren et al., 2008; Lelieveld et al.,

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2008) and as a precursor of O<sub>3</sub> (Trainer et al., 1987). It is also an important precursor for secondary organic aerosols (SOA) (Claeys et al., 2004) and a temporary reservoir for nitrogen oxide radicals (NO<sub>x</sub> ≡ NO + NO<sub>2</sub>) by formation of organic nitrates (Paulot et al., 2012). Isoprene thus has a range of impacts on air quality and climate that need to be included in atmospheric composition models. The widely used MEGAN global emission model (Guenther et al., 2006, 2012) indicates that 80 % of global isoprene emission takes place in the tropics and 25 % in Africa, but there are large uncertainties in these estimates due to lack of data. In previous work we developed a method to estimate isoprene emissions from Africa on the basis of observations of formaldehyde (HCHO) from the OMI satellite instrument. Here we use our OMI-derived isoprene emissions evaluated with local data to better understand the factors controlling isoprene emissions in Africa, improve emission estimates for different African plant functional types (PFTs), and assess the implications for atmospheric oxidants and aerosols.

Isoprene produced in the chloroplasts of plants is released to the atmosphere via the stomata of leaves (Sharkey and Yeh, 2001). Above-canopy emission fluxes  $E_{\text{ISOP}}$  depend on plant species, foliage density, leaf age, temperature, photosynthetically active radiation (PAR), and water stress (Guenther et al., 1995). This is commonly represented in isoprene emission models by multiplying an emission factor  $E_0$  defined for each PFT at standard conditions with an ensemble of coefficients describing the sensitivity to local environmental variables. In the MEGAN (version 2.1) inventory (Guenther et al., 2012) this is given as

$$E_{\text{ISOP}} = E_0 \times C_{\text{CE}} \times \text{LAI} \times \gamma_{\text{PAR}} \times \gamma_T \times \gamma_{\text{AGE}} \times \gamma_{\text{SM}} \quad (1)$$

where LAI is the leaf area index (m<sup>2</sup> leaf surface m<sup>-2</sup> of Earth surface) and the dimensionless activity factors  $\gamma$  describe the sensitivity to above-canopy radiation (PAR), air temperature ( $T$ ), leaf age distribution (AGE), and soil moisture (SM). The coefficient  $C_{\text{CE}}$  (m<sup>2</sup> Earth surface m<sup>-2</sup> leaf surface) enforces  $E_{\text{ISOP}} = E_0$  under standard conditions, which for MEGAN are defined as  $T = 303 \text{ K}$  and  $\text{PAR} = 1000 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ , and a canopy with  $\text{LAI} = 5 \text{ m}^2 \text{ m}^{-2}$ , leaf age distribution of

80 % mature, 10 % growing, and 10 % senescing leaves, and volumetric soil moisture of  $0.3 \text{ m}^3 \text{ m}^{-3}$ . The value of  $C_{CE}$  depends on the radiative transfer model used to describe the attenuation of PAR by the canopy.

Isoprene emission data for African vegetation are very limited, and emission models require extrapolation of data from other continents and across plant species (Guenther et al., 2006, 2012). This can lead to substantial errors, as differences in isoprene fluxes within and across plant species are large. Uncertainty in the distribution of land cover (PFT) in the tropics can further lead to a 5-fold discrepancy in isoprene emissions in equatorial Africa (Pfister et al., 2008).

Space-based observations of HCHO, a high-yield oxidation product of isoprene, have been used in a number of studies to infer isoprene emissions and evaluate inventories globally (Shim et al., 2005; Stavrou et al., 2009a, b) and regionally in Southeast Asia (Fu et al., 2007), South America (Barkley et al., 2008), North America (Palmer et al., 2003, 2006; Millet et al., 2008), Europe (Dufour et al., 2009; Curci et al., 2010), and Africa (Marais et al., 2012). These studies have confirmed temperature as the dominant factor controlling month-to-month variability of isoprene emissions across North America (Palmer et al., 2006; Millet et al., 2008) and Amazonia (Barkley et al., 2008). Leaf phenology and PAR were found to be additional important drivers of isoprene emission seasonality in Amazonia (Barkley et al., 2008, 2009). Stavrou et al. (2009a) found that water stress reduces isoprene emissions in southern Africa during the dry season. Here we use our previous work for Africa (Marais et al., 2012) to better understand the factors controlling isoprene emissions across the African continent and evaluate and improve the MEGANv2.1 emission inventory.

## 2 OMI-derived isoprene emissions in Africa

The derivation of isoprene emissions in Africa using OMI HCHO data is described in Marais et al. (2012) and summarized briefly here. OMI is a UV/VIS solar backscatter instrument on the Aura polar sun-synchronous satellite launched in 2004 (Levelt et al.,

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2006). It has a 13 km × 24 km nadir pixel resolution, daily global coverage through cross-track viewing, and 13:30 local time (LT) overpass. HCHO slant columns are obtained from version 2.0 (Collection 3) retrievals for 2005–2009 ([http://disc.sci.gsfc.nasa.gov/Aura/data-holdings/OMI/omhcho\\_v003.shtml](http://disc.sci.gsfc.nasa.gov/Aura/data-holdings/OMI/omhcho_v003.shtml)). They are corrected for instrument drift and converted to vertical columns using local air mass factors (AMF) for the scattering atmosphere (Palmer et al., 2001) with vertical HCHO profiles from the GEOS-Chem chemical transport model v9-01-03 (<http://www.geos-chem.org>) and scattering weights from the LIDORT radiative transfer model (Spurr et al., 2001).

HCHO enhancements over Africa primarily originate from isoprene emission, biomass burning, and fuel combustion. Scenes affected by biomass burning are excluded on the basis of MODIS satellite observations of fire counts and OMI satellite observations of aerosol absorption optical depth (AAOD) (Torres et al., 2007). Scenes affected by gas flaring are excluded on the basis of a specialized hotspot product from the AATSR satellite sensor (Casadio et al., 2012) and this leads to the exclusion of much of Nigeria where that source is particularly large and urban and industrial sources may contribute as well (Marais et al., 2014).

Marais et al. (2012) thus obtained a 2005–2009 monthly data set of vertical HCHO columns with 1° × 1° spatial resolution screened for biomass burning and anthropogenic influences and thus attributable to isoprene emissions (Fig. 1, left panel). They used GEOS-Chem to derived the sensitivity,  $S$ , of column HCHO  $\Omega_{\text{HCHO}}$  to a perturbation  $\Delta$  in isoprene emission  $E_{\text{ISOP}}$  ( $S = \Delta\Omega_{\text{HCHO}}/\Delta E_{\text{ISOP}}$ ). Values of  $S$  are sensitive to  $\text{NO}_x$  concentrations and this was accounted for using concurrent observations of OMI tropospheric  $\text{NO}_2$  columns. Scenes affected by smearing (displacement of HCHO from the isoprene emission source) were diagnosed with anomalously high values of  $S$  and excluded from the data set.

Marais et al. (2012) obtained in this manner a monthly isoprene emission inventory for 2005–2009 on a 1° × 1° grid (Fig. 1, center panel). The OMI overpass is at 13:30 LT and the corresponding isoprene emissions are for 12:00–15:00 LT, typically the diurnal maximum. Also shown in Fig. 1 is the MODIS IGBP land cover map (Friedl et al.,

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2002). Dominant vegetation types in Africa are roughly defined by latitudinal bands, with evergreen (broadleaf) trees along the Equator successively transitioning to the north and south to woody savannas (30–60 % tree coverage), savannas (10–30 %), grasslands, and deserts. The HCHO column data follow this vegetation gradient and so too do the inferred isoprene emissions.

Marais et al. (2012) presented a detailed error characterization of their OMI-derived isoprene emissions. Spectral fitting of the HCHO column has an error standard deviation of  $8 \times 10^{15}$  molecules  $\text{cm}^{-2}$  for individual observations. Relating the fitted slant HCHO columns to isoprene emissions incurs errors in the AMF estimate (20 %), the isoprene oxidation mechanism (15 %), the use of OMI  $\text{NO}_2$  to obtain  $S$  under low- $\text{NO}_x$  conditions (20–40 %), and smearing (30 % for high- $\text{NO}_x$  conditions, 30–70 % for low- $\text{NO}_x$ ). The resulting error in isoprene emission estimates for individual scenes, adding in quadrature all error contributions, is 40 % for high- $\text{NO}_x$  conditions and 40–90 % for low- $\text{NO}_x$  (Marais et al., 2012). A monthly mean estimate for a  $1^\circ \times 1^\circ$  grid square typically averages 3000 individual scenes. Averaging reduces the error though only to the extent that the error components are random.

### 3 Evaluation with canopy flux measurements

Canopy-scale isoprene flux measurements by relaxed-eddy accumulation (REA) are available from a few African field campaigns. Figure 2 compares OMI-derived isoprene emissions to REA measurements over equatorial evergreen trees (Greenberg et al., 1999; Serça et al., 2001), woody savannas (Greenberg et al., 1999), and savannas (Harley et al., 2003) in central and southern Africa (sites 1–4 in Fig. 1). Also shown are the values calculated using Eq. (1) with MEGAN v2.1 emission factors, combined Terra and Aqua MODIS LAI (Yang et al., 2006), and the Goddard Earth Observing System (GEOS-5) assimilated meteorological data. The Serça et al. (2001) and Harley et al. (2003) measurements (sites 1 and 4) are from walk-up towers with a flux footprint of about 600 m, while the Greenberg et al. (1999) measurements (sites 2 and 3) are

from aircraft with a flux footprint of  $\sim 100\text{ km} \times 100\text{ km}$  at site 2 and  $30\text{ km} \times 30\text{ km}$  at site 3. All values in Fig. 2 are for 12:00–15:00 LT. REA fluxes at sites 2–3, obtained in the morning (09:30–11:30 LT), are increased by a factor of 1.4 as a diurnal correction for temperature and PAR following MEGAN.

5 OMI and MEGAN are sampled for the  $1^\circ \times 1^\circ$  gridsquare coincident with the observation site and for the corresponding months. Interannual variability is of similar magnitude in the OMI-derived and MEGAN data at sites 1 (November), 2 and, 3 where multi-year OMI data are available. The variability is driven in MEGAN predominantly by temperature. At site 3 there are no OMI data in the months of observation  
10 (November–December) because of biomass burning interference and we show instead OMI-derived emissions in September–October, which should be similar to November–December according to MEGAN.

For the equatorial evergreen tree sites in central Africa (sites 1 and 2) OMI-derived isoprene emissions are on average 2 times higher than the REA measurements and  
15 MEGAN is 5 times higher. The flux tower sampled vegetation with a relatively low fraction of isoprene emitters. Nearby landscapes include monodominant stands of the *Gilbertiodendron* trees that have a high isoprene emission factor (Serça et al., 2001). The distribution of this tree species beyond the sampling domain is uncertain and application of its emission factor to land cover in equatorial Africa contributes to the over-  
20 estimate in MEGAN. OMI and MEGAN reproduce the March–November decline at site 1 and this is driven in MEGAN by temperature.

Fluxes at site 2, where the sampling footprint is similar to OMI, have large spatial variability implying that differences in the sampling footprint contribute to discrepancies. Greenberg et al. (1999) applied a positive correction of  $\sim 20\%$  to flux measurements at  
25 sites 2–3 to account for the transport of isoprene that was not accumulated in the two REA reservoirs. A similar negative bias may affect measurements at sites 1 and 4 but the reported values have not been adjusted. An additional negative bias associated with the aircraft REA measurements (sites 2 and 3), but not with the tower sites (1 and 4), is the vertical flux divergence between the height of the aircraft measurement

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and the surface flux. The isoprene vertical flux divergence profiles investigated by Karl et al. (2013) suggest that the fluxes reported by Greenberg et al. (1999) should be increased by a further 25 %.

At the woody savanna site OMI is 2.2 times higher than the REA measurement (1.8 times higher if a 25 % upward correction is applied to the REA measurement), while MEGAN is 8 times higher. At the savanna site OMI is 1.3 times higher than the REA flux measurement while MEGAN is 2.4 times higher. The discrepancy at site 4 is partly due to the low (< 10 %) proportion of isoprene emitting vegetation within the flux tower footprint as compared to ~ 35 % for savannas surveyed at surrounding field sites (Harley et al., 2003).

Overall the REA flux measurements indicate canopy-scale isoprene emissions that are somewhat lower than derived from OMI and much lower than derived from MEGAN. As we will show in Sect. 5, isoprene emission factors obtained with OMI are more consistent with landscape-level field campaign measurements and lead to an improved estimate of isoprene emissions from MEGAN.

## 4 Seasonality of isoprene emissions in Africa

We examine the seasonality of our OMI-derived isoprene emissions in relation to environmental variables, focusing on three seasonally and ecologically coherent regions in Africa where emissions are highest (Fig. 3): (1) equatorial forests dominated by tropical broadleaf evergreen trees, (2) northern savannas (including woody), and (3) southern savannas (including woody). Figure 4 shows the seasonality of OMI-derived isoprene emissions for these three regions together with MODIS LAI and GEOS-5 2 m air temperature. Isoprene emissions and 2 m temperature are for 12:00–15:00 LT and MODIS LAI is the combined Terra and Aqua product (Yang et al., 2006). Multi-year averages (2005–2009) are shown as the regionally averaged interannual variability is small, except over northern savannas in August as discussed below.

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OMI-derived isoprene emissions for equatorial forests are a factor of 2 lower than MEGAN and both show similar weak seasonality, with a decline from March to November that is consistent with the Serça et al. (2001) REA flux measurements (Fig. 2) and is driven in MEGAN principally by temperature. Although LAI exhibits similar seasonality, it remains above  $3.5 \text{ m}^2 \text{ m}^{-2}$  year-round and MEGAN is not sensitive to LAI values above  $2\text{--}3 \text{ m}^2 \text{ m}^{-2}$  due to shading of lower-canopy leaves (Guenther et al., 2006).

Availability of OMI-derived isoprene emission data for the northern savannas is limited to April–November because of pervasive biomass burning influence during the December–March dry season. Emissions are maximum in April, at the beginning of the wet season, and minimum in August when the West African Monsoon (WAM) is fully developed over the continent (Janicot et al., 2008) resulting in cooler temperatures. OMI-derived emissions largely follow temperature over the April–November period. Year-to-year variability in the WAM affects temperature in August, leading to inter-annual variability in August OMI isoprene emissions over the 2005–2009 period that is correlated with temperature ( $r = 0.55$ ).

The complete seasonality simulated by MEGAN in northern savannas shows low isoprene emissions in the December–March dry season when LAI is less than  $1 \text{ m}^2 \text{ m}^{-2}$ , and a broad maximum in the April–November wet season as the August minimum in temperature is compensated by a corresponding maximum in LAI. MODIS LAI in the northern savannas is less than  $2.5 \text{ m}^2 \text{ m}^{-2}$  year-round, sufficiently low that the MEGAN dependence on LAI does not saturate (Guenther et al., 2012). However, MODIS may underestimate LAI in West Africa during the wet season because of cloud contamination (Gessner et al., 2013).

OMI-derived emissions for southern savannas are in close agreement with MEGAN, featuring a minimum in the winter dry season and a maximum in the summer wet season. The seasonal minimum follows that of temperature (June–July) with a 1 month lag that reflects the very dry conditions in July–September. We find that LAI and temperature are both important for driving the seasonality of isoprene emissions in southern savannas.

## 5 Satellite-derived isoprene emission factors for Africa

The general ability of MEGAN to reproduce the seasonal variation of OMI-derived isoprene emissions suggests that the MEGAN activity factors ( $\gamma$  in Eq. 1) are appropriate for conditions in Africa, with temperature and LAI as the principal drivers. The large MEGAN bias over equatorial forests and northern savannas can therefore be attributed to an overestimate of emission factors ( $E_o$  in Eq. 1) that results from uncertainties in PFT distribution and limited emission data for plant species in these regions.

The emission factors in MEGAN are isoprene fluxes for a canopy with leaves at standard conditions of air temperature ( $T = 303\text{ K}$ ) and light ( $\text{PAR} = 1000\ \mu\text{mol photons m}^{-2}\ \text{s}^{-1}$ ) gridded to a detailed regional land cover map for Africa south of the Equator (Otter et al., 2003) and the Olson et al. (2001) global ecoregion data in the north (Guenther et al., 2006). We can infer them from the OMI-derived canopy-level isoprene emission data by using Eq. (1). In so doing we only consider individual activity factors  $\gamma$  within the range 0.5–1.5.

Figure 5 shows the resulting distributions of MEGAN and OMI emission factors  $E_o$  over Africa together with observations from field campaigns. The latter are at the landscape-level and were obtained by scaling measured leaf-level isoprene fluxes for representative plant species with foliage density and species distribution data. Leaf-level measurements at standard conditions were obtained by using enclosure measurements with controlled temperature and PAR (Otter et al., 2002; Serça et al., 2001), or adjusting to standard conditions with MEGAN activity factors for temperature and PAR (Guenther et al., 1996). For the former an upward correction applied to the Serça et al. (2001) landscape emission factor accounts for isoprene fluxes obtained from shade-adapted leaves that have lower emissions at standard conditions than sunlit leaves (Guenther et al., 1999). Leaf-level fluxes of Klinger et al. (1998) were determined to be at standard conditions with coincident measurements of temperature and PAR, but we exclude the data from shaded leaves at dense forest sites.

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We find from Fig. 5 remarkable agreement between OMI-derived emission factors and the field data ( $r = 0.55$ , OMI normalized mean bias =  $-19\%$ ). Woody savannas in Zambia and savannas in South Africa have large variability in  $E_o$  ( $0.5\text{--}4.5\text{ mg C m}^{-2}\text{ h}^{-1}$ ) that is reproduced by OMI. The two sites in Botswana have low emission factors as the site to the north is dominated by monoterpene emitting mopane vegetation, while the site to the south is predominantly shrubland (Otter et al., 2002).

Differences between OMI and MEGAN emission factors are largest for equatorial forests, and the field enclosure observations are in good agreement with OMI and much lower than MEGAN. The equatorial forest enclosure measurements are used in MEGAN to estimate emission factors there, but a large positive correction is applied to account for leaf enclosure measurements of shade-adapted leaves. Our OMI-derived emission factors do not support a large positive correction, and this is also reflected in the MEGAN overestimate of REA flux measurements (Fig. 2). OMI emission factors for equatorial forests are larger in the west than east and this may result from differences in the proportion of isoprene emitting species. The west is dominated by dry tropical forests, while the east is dominated by permanently or seasonally flooded forests (White, 1983).

Table 1 shows mean isoprene emission factors for individual PFTs as obtained by mapping the data from Fig. 5 onto the MODIS IGBP land map (Friedl et al., 2002) and the Global Land Cover (GLC) 2000 land map (Mayaux et al., 2006). The distribution of MODIS IGBP woody savannas is spatially consistent with GLC 2000 broadleaf trees and MODIS IGBP savannas correspond to GLC 2000 shrubs interspersed with broadleaf trees and cultivated land. The GLC 2000 classification scheme is more consistent with the PFTs of MEGANv2.1 (Guenther et al., 2012). OMI gives higher emission factors for forested vegetation than for grasslands but the difference is not as large as MEGAN and more consistent with the field enclosure observations. The largest differences between OMI and MEGAN emission factors are for broadleaf evergreen trees and for semi-arid vegetation (shrubs and herbs).

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The OMI-derived emission factors in Fig. 5 can be used to improve the MEGAN isoprene emission estimates as computed from Eq. (1). Figure 6 compares the resulting isoprene concentrations simulated by GEOS-Chem with a latitudinal profile of isoprene concentration measurements below 1 km across West Africa during the AMMA wet season aircraft campaign in July–August 2006 (Murphy et al., 2010). There is a strong vegetation gradient along the AMMA flight track from the Gulf of Guinea to Benin woodlands to arid conditions in the north that is reflected in the isoprene data. Simulated isoprene in GEOS-Chem is a factor of 2 too low over Benin woodlands likely due to a seasonal low bias in MODIS LAI over West Africa from cloud contamination (Gessner et al., 2013). Isoprene emissions over the AMMA domain are not only sensitive to LAI, but also MEGAN emission factors (Ferreira et al., 2010). The OMI-derived emission factors are much better able to reproduce the latitudinal gradient than the original MEGAN emission factors, including in particular the decline to the north associated with increased aridity. Throughout Africa MEGAN emission factors are too high for semi-arid PFTs, as indicated by the overestimate in MEGAN for GLC 2000 sparse herb/shrub cover in Table 1.

## 6 Implications for oxidants and aerosols

We use the GEOS-Chem chemical transport model to (1) evaluate the change in atmospheric composition that results from replacing MEGAN emission factors with those obtained using OMI, and (2) determine the impact of isoprene emissions using OMI-derived emission factors on aerosols and oxidants. GEOS-Chem includes the standard representation of oxidant-aerosol chemistry as described for example by Mao et al. (2010) with updates to the isoprene oxidation mechanism (Paulot et al., 2009a, b).

Total annual isoprene emissions in Africa averaged over 2005–2009 using OMI-derived emission factors are  $77 \text{ TgCa}^{-1}$ , as compared to  $104 \text{ TgCa}^{-1}$  in MEGAN v2.1. The difference is mainly for the equatorial evergreen forest PFT in central Africa, where

OMI-derived isoprene emissions are 2–3 times lower than MEGAN. GEOS-Chem using OMI-derived isoprene emissions indicates a factor of 4 increase in boundary layer OH concentrations over central Africa relative to MEGAN, a 4 ppbv increase in surface  $O_3$ , and 8 ppbv decrease in surface isoprene.

Figure 7 shows the effect of African isoprene emissions on surface concentrations of daily maximum 8 h average (MDA8)  $O_3$ , particulate matter (PM),  $NO_x$ , and OH. We calculated this effect by difference between a GEOS-Chem simulation with OMI-derived isoprene emission factors and a simulation with no isoprene emission. The largest effect on  $O_3$  is over West Africa because of high anthropogenic, soil, and biomass burning  $NO_x$  emissions (Marais et al., 2014). The largest effect on PM is also over West Africa and reflects the availability of high pre-existing primary PM from combustion (biomass burning and fuel) on which isoprene oxidation products can condense.  $NO_x$  declines in West Africa and the tropics due to formation of isoprene nitrates. Loss of OH from reaction with isoprene is highest in the tropics where low levels of  $NO_x$  limit the recycling of  $HO_x$  radicals.

## 7 Conclusions

We used a 2005–2009 data set of monthly isoprene emissions in Africa derived from OMI satellite observations of formaldehyde (HCHO) to study the factors controlling these emissions in different areas of the continent. Our goal was to achieve a better representation of isoprene emission in chemical transport models (CTMs), in part through evaluation and improvement of the commonly used MEGAN emission inventory, and to examine the implications for oxidants and aerosols over the continent.

We began by evaluating the OMI-derived isoprene emissions with relaxed-eddy accumulation flux (REA) measurements obtained from above-canopy towers and aircraft during African field campaigns. OMI-derived isoprene emissions are on average 2 times higher than REA measurements over the equatorial forest and woody savan-

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nas but this could reflect biases in the measurements. MEGAN emissions are 5–10 times higher.

We subdivided Africa into 3 seasonally and ecologically coherent regions to examine the seasonality in OMI-derived isoprene emissions, and compare to the seasonality in MEGAN and in driving environmental variables. Equatorial forests exhibit weak seasonality that is driven predominantly by temperature, while seasonality in savannas is driven by both temperature and LAI, in a manner consistent with MEGAN.

Isoprene emissions in MEGAN are computed as the product of (1) an emission factor  $E_o$  characteristic of the PFT, (2) the LAI, and (3) activity factors dependent on local environmental variables. We applied the LAI and MEGAN activity factors to our OMI-derived isoprene emissions to obtain emission factors representative of different PFTs. These agree well with the ensemble of leaf-level flux measurements in Africa and imply large downward corrections to MEGAN emission factors for equatorial forests and semi-arid vegetation. Such corrections are consistent with the latitudinal gradient of isoprene across West Africa measured in the AMMA field campaign.

The OMI-derived emission factors can be incorporated into the MEGAN formalism (Eq. 1) to improve modeling of isoprene emissions in Africa. The resulting isoprene emissions for the continent are  $77 \text{ TgCa}^{-1}$ , as compared to  $104 \text{ TgCa}^{-1}$  in the standard MEGAN inventory. Most of the difference is over equatorial Africa. We conducted GEOS-Chem simulations with and without African isoprene emissions (using OMI-derived emission factors) to examine the impact on regional particulate matter (PM) and oxidants. The largest effect of isoprene emissions on surface  $\text{O}_3$  is over West Africa where  $\text{NO}_x$  is high, and the largest effect on PM is also over West Africa because of pre-existing high concentrations of primary PM from combustion.

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**Table 1.** Isoprene emission factors for African plant functional types<sup>a</sup>.

Plant functional type	Emission factor [ $\text{mgCm}^{-2}\text{h}^{-1}$ ]	
	MEGAN	OMI
MODIS IGBP classification <sup>b</sup>		
Evergreen broadleaf trees	$4.3 \pm 2.0$	$2.7 \pm 1.0$
Deciduous broadleaf trees	$4.4 \pm 1.7$	$2.9 \pm 0.2$
Woody savannas	$3.2 \pm 1.3$	$2.6 \pm 1.0$
Savannas	$2.9 \pm 1.2$	$2.3 \pm 0.8$
Shrubs	$3.0 \pm 1.3$	$1.6 \pm 0.8$
Grasses	$1.8 \pm 0.9$	$1.6 \pm 0.9$
Crops	$1.4 \pm 0.9$	$1.6 \pm 0.6$
Mosaic of crops and natural vegetation	$2.3 \pm 1.1$	$2.5 \pm 1.0$
GLC 2000 classification <sup>c</sup>		
Evergreen broadleaf trees	$4.4 \pm 1.9$	$2.5 \pm 1.0$
Deciduous broadleaf trees	$3.0 \pm 1.3$	$2.7 \pm 0.9$
Shrubs	$3.1 \pm 1.6$	$2.2 \pm 1.0$
Herbs	$2.4 \pm 1.3$	$1.9 \pm 0.9$
Sparse herbs or shrubs	$2.4 \pm 1.2$	$1.5 \pm 0.7$
Cultivated land	$1.8 \pm 1.1$	$2.2 \pm 0.9$
Mosaic of crops and natural vegetation	$3.0 \pm 1.3$	$2.7 \pm 0.7$
Mosaic of crops and shrubs or grasses	$2.6 \pm 1.0$	$1.9 \pm 0.9$

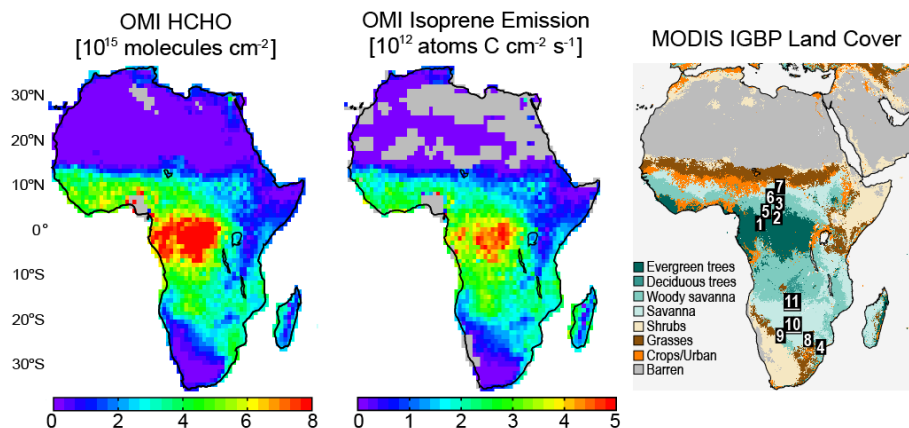
<sup>a</sup> Isoprene emission factor  $E_o$  in Eq. (1) at standard conditions of air temperature (303 K) and photosynthetically active radiation ( $1000 \mu\text{mol photons m}^{-2} \text{s}^{-1}$ ). Values are means and standard deviations obtained by mapping the  $E_o$  data from Fig. 5 onto the MODIS IGBP and GLC 2000 land maps. Plant functional type classifications are as given by each land map.

<sup>b</sup> Friedl et al. (2002) and shown in Fig. 1.

<sup>c</sup> Mayaux et al. (2006).

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**Fig. 1.** Annual mean (2005–2009) OMI HCHO vertical columns at  $1^\circ \times 1^\circ$  horizontal resolution screened for biomass burning and anthropogenic HCHO (left), and resulting OMI-derived isoprene emissions (center), as derived by Marais et al. (2012) and summarized in the text. The OMI observations are at 13:30 LT and the OMI-derived isoprene emissions are for 12:00–15:00 LT. The right panel is a MODIS IGBP land cover map (Friedl et al., 2002) with numbers showing the location of isoprene flux measurements used to evaluate the OMI-derived isoprene emissions. These include 1. Serça et al. (2001) for March and November 1996; 2. Klinger et al. (1998) for 1995–1996 (Fig. 5) and Greenberg et al. (1999) for November–December 1996 (Fig. 2), 3. Greenberg et al. (1999) for November–December 1996, 4. Guenther et al. (1996) for December 1992 (Fig. 5) and Harley et al. (2003) for February 2001 (Fig. 2), 5–7. Klinger et al. (1998) for 1995–1996, 8–11. Otter et al. (2002) for February–March 2001.

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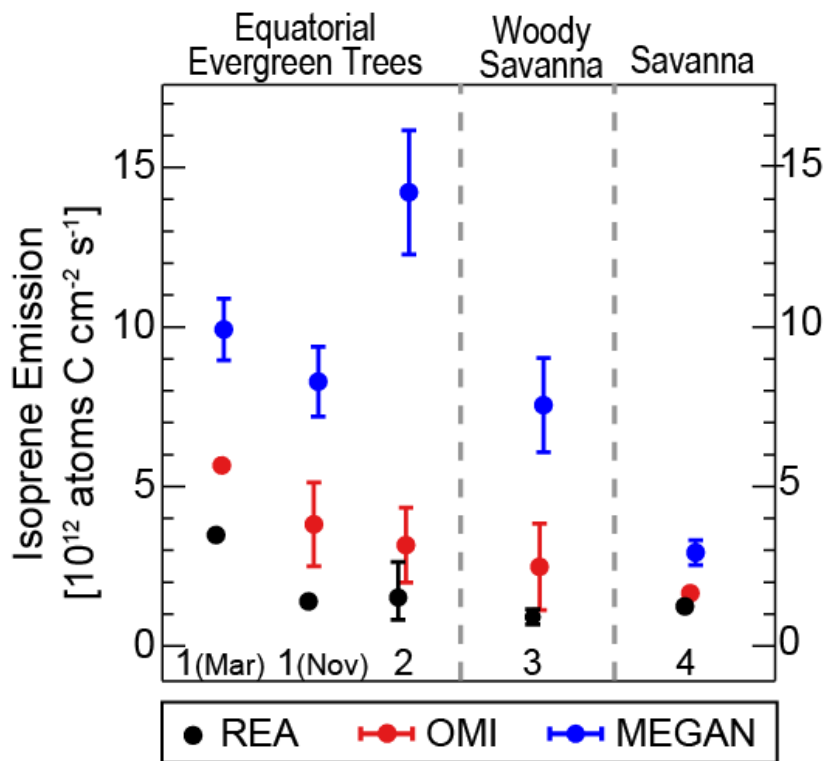
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**Fig. 2.** Mean canopy-scale isoprene emissions at African sites 1–4 (see Fig. 1) measured by relaxed-eddy accumulation (REA), and comparison to OMI-derived and MEGAN values. All values are for 12:00–15:00 LT, with diurnal correction for REA measurements at sites 2 and 3 (see text). Vertical bars on the REA measurements for sites 2–3 are the interquartile ranges over the aircraft sampling domain given in Greenberg et al. (1999). OMI-derived and MEGAN values are 2005–2009 monthly averages for the site locations and observation times, with interannual standard deviations shown as vertical bars. Mar=March, Nov=November.

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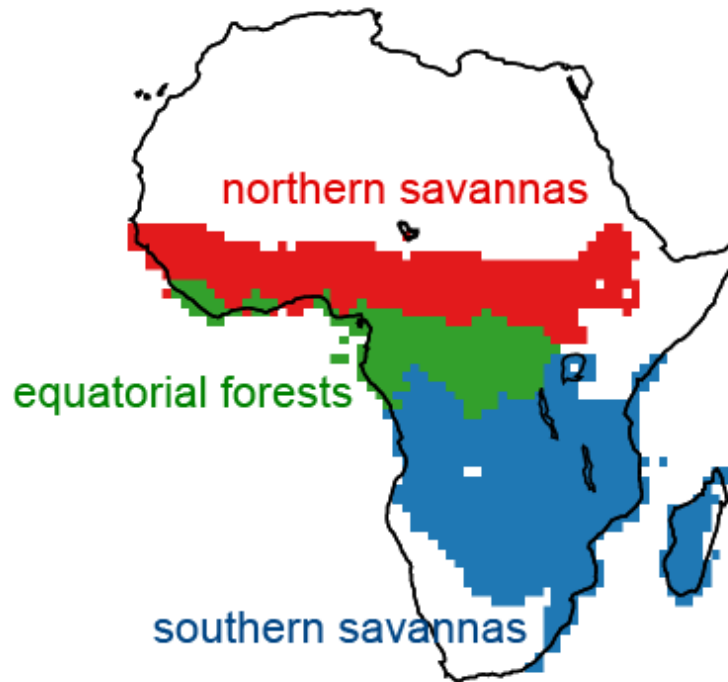
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**Fig. 3.** Coherent regions used for analysis of the factors controlling OMI-derived isoprene emissions: equatorial forests (green), northern savannas (red), and southern savannas (blue). Land cover definitions are from the MODIS IGBP map (Fig. 1).

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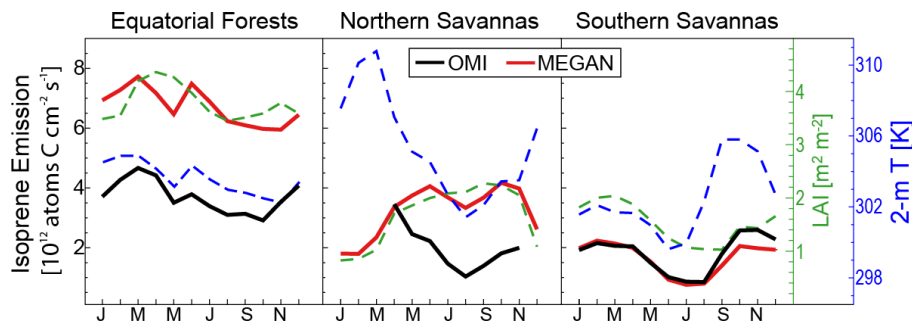
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**Fig. 4.** Seasonality of isoprene emissions and environmental variables averaged over the coherent African regions of Fig. 3. Monthly mean OMI-derived (black) and MEGAN (red) isoprene emissions are shown together with GEOS-5 2 m temperature (blue) and MODIS LAI (green). OMI-derived isoprene emissions are not available for northern savannas in December–March because of biomass burning interference. Emissions and 2 m temperature are for 12:00–15:00 LT. LAI is the combined Terra and Aqua product (Yang et al., 2006). All data are means for 2005–2009.

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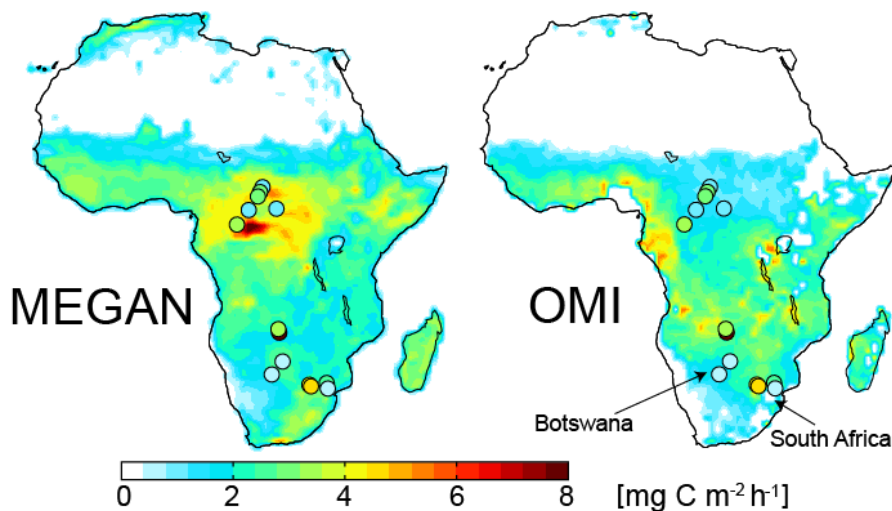
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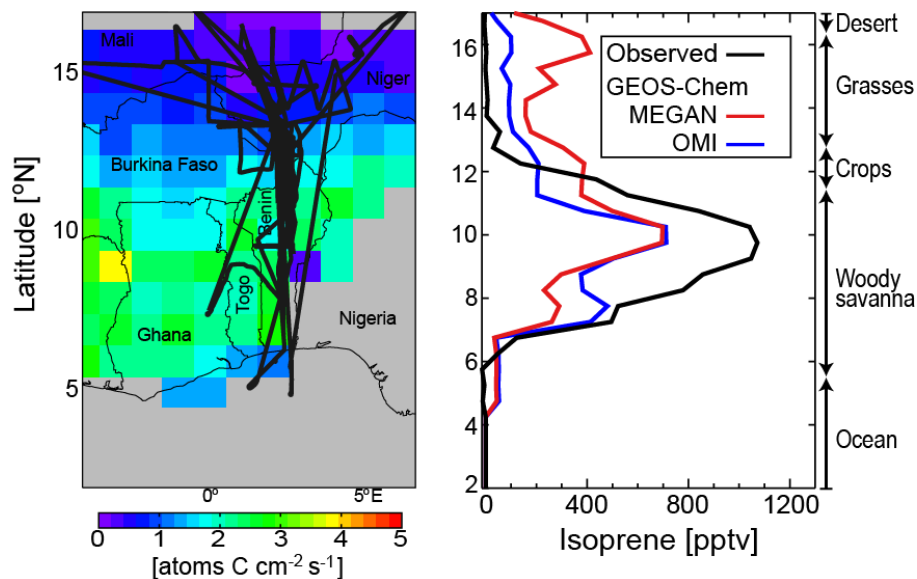


**Fig. 5.** Isoprene emission factors ( $E_o$  in Eq. 1) representing the emission flux under standard conditions. Measured landscape-level emission factors from field sites (circles; references given in Fig. 1) are compared to those used in MEGAN (left) and obtained with OMI (right). OMI data missing in Nigeria are due to interference from anthropogenic HCHO, and in East Africa and Namibia/South Africa due to excessive departure from standard conditions (see text for details).

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**Fig. 6.** Latitudinal variability of isoprene in West Africa. Left panel shows the July–August 2006 AMMA flight tracks superimposed on July–August 2005–2009 OMI-derived isoprene emissions (Marais et al., 2012). Right panel shows boundary-layer (< 1 km) isoprene concentrations along the AMMA flight tracks; observations averaged over  $0.5^\circ$  latitude bands are compared to GEOS-Chem model results using either MEGAN or OMI-derived isoprene emission factors  $E_0$ . Dominant MODIS IGBP biomes along the flight tracks are indicated.

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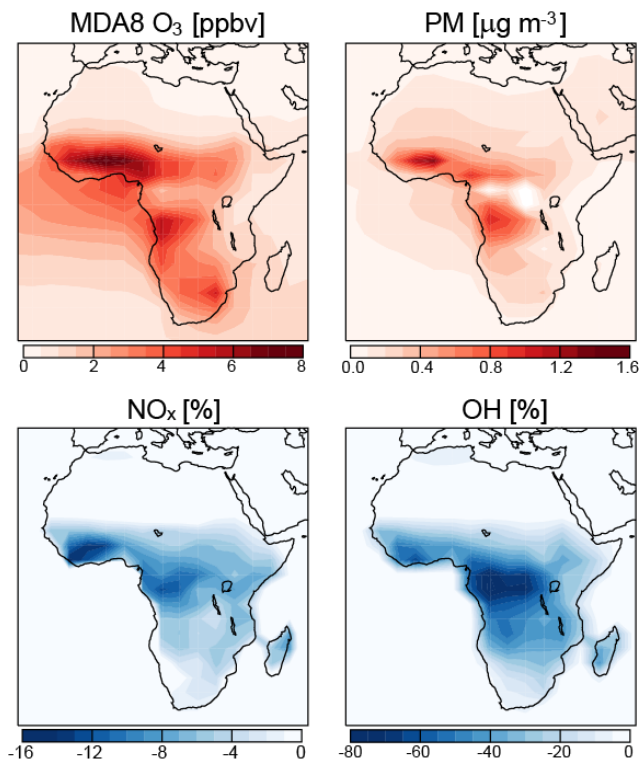
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**Fig. 7.** Effect of African isoprene emissions on regional oxidant chemistry and particulate matter in surface air. Shown are the annual mean differences between GEOS-Chem simulations with and without African isoprene emissions for surface daily maximum 8 h average (MDA8) O<sub>3</sub>, particulate matter (PM), NO<sub>x</sub>, and OH. MDA8 O<sub>3</sub> and PM are absolute differences and NO<sub>x</sub> and OH are relative differences. The simulation of isoprene emission uses OMI-derived emission factors.