

# Uptake and emission of VOCs near ground level below a mixed forest at Borden, Ontario

M. Gordon<sup>1</sup>, A. Vlasenko<sup>1,\*</sup>, R.M. Staebler<sup>1</sup>, C. Stroud<sup>1</sup>, P.A. Makar<sup>1</sup>, J. Liggio<sup>1</sup>, S.-M. Li<sup>1</sup>, S. Brown<sup>2</sup>

[1]{Atmospheric Science and Technology Directorate, Science and Technology Branch, Environment Canada, Toronto, Canada}

[2] Agriculture and Forest Meteorology, Guelph University, Guelph, Canada}

[\*]{now at: Airzone One Ltd., Mississauga, ON, Canada}

Correspondence to: R.M. Staebler (ralf.staebler@ec.gc.ca)

## Abstract

Understanding of the atmosphere/forest canopy exchange of volatile organic compounds (VOCs) requires insight into deposition, emission, and chemical reactions of VOCs below the canopy. Between July 18 and Aug 9, 2009, VOCs were measured with proton-transfer-reaction mass spectrometry (PTR-MS) at 6 heights between 1 and 6 m beneath a 23-m high mixed-forest canopy. Measured VOCs included methanol, isoprene, acetone, methacrolein + methyl vinyl ketone (MACR+MVK), monoterpenes and sesquiterpenes. There are pronounced differences in the behaviour of isoprene and its by-products and that of the terpenes. Non-terpene mixing ratios increase with height, suggesting predominantly downward fluxes. In contrast, the terpene mixing ratios decrease with height, suggesting upward fluxes. A 1-dimensional canopy model was used to compare results to measurements with and without surface deposition of isoprene and MACR+MVK and emissions of monoterpenes and sesquiterpenes. Results suggest deposition velocities of  $2.7 \text{ mm s}^{-1}$  for isoprene and  $1.2 \text{ mm s}^{-1}$  for MACR+MVK and daytime surface emission rates of  $63 \mu\text{g m}^{-2} \text{ h}^{-1}$  for monoterpenes. The modelled isoprene surface deposition is approximately 2% of the canopy top isoprene emissions and the modelled emissions of monoterpenes comprise approximately 15 to 27% of the canopy-top monoterpene emissions to the atmosphere. These results suggest that surface monoterpene emissions are significant for forest

1 canopy/atmosphere exchange for this mixed forest location and surface uptake is relatively  
2 small for all the species measured in this study.

3

#### 4 **1 Introduction**

5 Biogenic volatile organic compounds (BVOCs) can play a significant role in atmospheric  
6 chemistry (Schade et al. 2010) and forests are a significant source of BVOC emissions  
7 (Lappalainen et al. 2009). The emission of BVOCs is the largest terrestrial source of reactive  
8 carbon to the atmosphere and isoprene is the largest contributor (Guenther et al. 1995).  
9 BVOCs are involved in the formation and growth of atmospheric aerosol particles (Tunved et  
10 al., 2006). Holzinger et al. (2005) found a large number of higher molecular weight  
11 compounds (>100 amu) in the air within and above a ponderosa pine plantation. Most  
12 compounds are from reactions between ozone and terpenoids emitted from the forest  
13 (Calogirou et al., 1999) while some are from the reaction of ozone with leaf surfaces  
14 (Fruekilde et al., 1998; Wildt et al., 2003). Methanol is produced in plants and is attributed to  
15 plant cell wall growth and repair (Kreuzwieser et al., 2000). Isoprene, monoterpene, and  
16 methanol emissions are controlled by air temperature (Tingey et al., 1980) and light intensity  
17 (Guenther et al., 1991, Folkers et al, 2008). However, direct correlations are difficult to  
18 measure, as plants acclimate to the environment following cues from previous hours, days, or  
19 even seasons (Oquist and Huner, 2003, Mäkelä et al, 2004).

20

21 The loss of BVOCs to reactions within the canopy is a poorly understood process. Makar et  
22 al. (1999) found a 40% loss of isoprene due to in-canopy chemistry using a one-dimensional  
23 canopy model, while the isoprene loss modelled by Karl and Guenther (2005) was between 2  
24 and 5%. It is unknown how production and loss of VOCs are connected to reactions at the  
25 forest surface. Leaf litter and forage have been identified as potentially significant sources of  
26 VOCs to the atmosphere (Kirstine et al. 1998, Warneke et al. 1999, de Gouw et al. 1999,  
27 Shade et al, 1999, Leff and Fierer, 2008). Leaf litter may be a strong source of methanol,  
28 possibly accounting for  $40 \times 10^{12}$  g yr<sup>-1</sup> of global emissions (Warneke et al. 1999). Leff and  
29 Fierer (2008) found that nonmethane VOC production rates were higher in leaf litter samples  
30 than soil and the rates were correlated with microbial biomass. Uptake of methane in soils is  
31 generally microbially mediated (Schade et al. 1999) and recent research suggests this may

1 also be the case for methanol (Stacheter et al. 2013). Compounds such as methanol,  
2 acetaldehyde, and acetone can also be released from decaying plant material (Warneke et al.  
3 1999).

4

5 In a recent study by Stroud et al. (2005), a one-dimensional model was compared to VOC  
6 measurements made at a pine plantation in order to determine escape efficiencies for  
7 terpenoid emissions. The model output was compared to sub-canopy measurements of  
8 isoprene, pinene, and methyl vinyl ketone and methacrolein (MVK+MACR). To improve  
9 model accuracy near the forest floor, surface deposition of isoprene and MVK+MACR (both  
10  $2 \text{ mm s}^{-1}$ ) and emission of  $\alpha,\beta$ -pinene ( $69 \mu\text{g m}^{-2} \text{ h}^{-1}$ ) were added to the model. Although the  
11 surface deposition of isoprene and MVK+MACR is a negligibly small fraction of the canopy-  
12 top fluxes, the surface emission of  $\alpha,\beta$ -pinene, which is attributed to decaying pine needles, is  
13 10% of the canopy-top flux.

14

15 Measurements to verify these deposition and emission rates are limited. In the same pine  
16 forest of the Stroud et al. (2005) study, Karl et al. (2005) used an inverse-Lagrangian model  
17 with VOC profile measurements to demonstrate a surface uptake of methanol, acetone and  
18 MVK+MARC, and emission of C3/C4 carbonyls. These results verified the Stroud et al.  
19 approximations. A previous report of the same study (Karl and Guenther, 2005) also shows a  
20 surface uptake of isoprene and emission of monoterpenes. In a tropical rainforest, Karl et al.  
21 (2004) used an inverse-Lagrangian model with VOC profile measurements to demonstrate a  
22 surface uptake of methanol, acetone, and isoprene. Aaltonen et al. (2011) used sample  
23 chambers in a boreal pine forest and measured surface emissions of isoprene, monoterpenes,  
24 and sesquiterpenes ( $0.05$ ,  $5.04$ , and  $0.04 \mu\text{g m}^{-2} \text{ h}^{-1}$  respectively). Hence there is a large  
25 amount of variation in both the direction and magnitude of surface flux measurements.

26

27 The goal of this study was to quantify the uptake and emissions of VOCs by a surface beneath  
28 a mixed-deciduous forest canopy. We deployed a proton-transfer-reaction time-of-flight mass  
29 spectrometer (PTR-ToF-MS) at the Borden Forest Research Station to measure VOC profiles  
30 near the forest floor in July and August, 2009. Based on these measurements, the one-  
31 dimensional canopy model of Makar et al. (1999) and Stroud et al. (2005) was modified to

1 include deposition and emissions of VOCs, allowing for the investigation of the relative  
2 importance of VOC uptake and emissions at this sub-canopy surface.

3

## 4 **2 Methods**

### 5 **2.1 Site Description**

6 This study took place at the Borden Forest Research Station (44°19'N 79°56'W) in Ontario,  
7 Canada. The study was part of a larger study investigating aerosol fluxes (Gordon et al.,  
8 2011), isotopes (Santos et al., 2012), and sub-canopy transport (Brown et al., 2013). A 2006  
9 stem count (Teklemariam et al., 2009) gave a forest composition of 52% red maple (*Acer*  
10 *rubum* L.), 14% eastern white pine (*Pinus strobus* L.), 8% large-tooth aspen (*Populus*  
11 *grandidentata* Michx), 7% white ash (*Fraxinus americana* L.), 6% American beech (*Fagus*  
12 *grandifolia*), with the remainder composed of a mixture of aspen, ash, cherry, pine, elm, and  
13 maple. The forest is regrowth on abandoned farm land, with an age of approximately 100  
14 years (Lee et al., 1999). During the 2009 summer, the average canopy height was  
15 approximately 23 m with an understory height of approximately 9 m. The forest soil is loamy  
16 sand (Barr et al., 2002) composed of 91% sand, 6% silt, and 3% clay (Chang, 2011). The  
17 surrounding area is generally flat within a radius of about 4 km. The available forest fetch is  
18 about 4.3 km to the south, and 3 km to the SSW. Outside of this range was predominantly  
19 grass and cropland, with the town of Angus, Ontario less than 4 km to the ENE and the  
20 Canadian Forces Base Borden 5 km to the SE.

21

### 22 **2.2 Instrumentation**

23 The Borden Forest Research Station includes a 44-m tower, which supports an array of  
24 instruments for routine measurements of temperature, wind speed, humidity, and fluxes of  
25 energy, water vapour, and CO<sub>2</sub>. Details of the measurement setup can be found in Lee et al.  
26 (1999), Staebler et al. (2000), Barr et al. (2002), and Teklemariam et al. (2009). Incoming  
27 shortwave radiation (SW) was measured (CNR1, Kipp & Zonen) above the canopy, and  
28 photo-synthetically active radiation (PAR) was measured (LI-191, Licor Inc.) at the forest  
29 floor. Temperature was measured at 12 heights between 1.7 and 41.5 m. A 3D sonic  
30 anemometer (K-Type, Applied Technologies Inc.) was mounted on the tower at 33 m and

1 measured turbulence at 10 Hz. O<sub>3</sub> was measured (49i, Thermo Scientific Inc.) at six heights  
2 between 1 and 41.5 m. NO was also measured (42s, Thermo Scientific Inc.) between July 16  
3 and August 12, 2006 at the same 6 heights as O<sub>3</sub>; however, NO measurements at the tower  
4 were not available during the 2009 study. To estimate the NO values during the 2009 study  
5 measurements at a station 15 km south of the Borden forest (Egbert, ON) were used. The  
6 Egbert station is an open area above the surrounding canopy. During the 2006 measurements  
7 the median hourly values of NO at Borden (at the 41.5 m height) and at Egbert correlated well  
8 ( $r^2 = 0.68$ ) with an average of 0.26 ppb at Egbert and 0.27 ppb at Borden. This suggests that  
9 the Egbert NO measurements are a suitable proxy for above-canopy Borden NO values. To  
10 estimate the sub-canopy NO profiles, average profiles were generated from the 27 days of  
11 2006 Borden tower measurements for each hour of the day. These hourly profile shapes were  
12 normalized with the 41.5 m height value. The normalized profile shapes were then scaled to  
13 the above canopy values estimated from the 2009 Egbert station measurements. Sensitivity of  
14 the model to NO approximation is discussed in Section 4.2.

15

16 A proton-transfer-reaction time-of-flight mass spectrometer (PTR-ToF-MS, Ionicon Inc.),  
17 installed at ground level, was operational from July 18 to Aug 9, 2009. The PTR-ToF-MS  
18 sampled from 6 levels (heights of 1, 2, 3, 4, 5, and 6 m) on a 6-m tower. Samples were drawn  
19 through 12 m of 6.35 mm (0.25 inch) outside diameter Teflon tubing at each level. The  
20 residence time was measured at ~2 s by introducing benzene pulses to each line and timing  
21 the response. An automated valve system switched between each line once per minute. Each  
22 one minute average was background subtracted and normalized by the m/z 21 reagent ion  
23 signal and subsequently converted to mixing ratio using gravimetrically prepared calibration  
24 standards. Due to a lack of calibration standards, the sesquiterpene concentrations were  
25 estimated using the monoterpene standard. Since the sesquiterpene calibration ratio is known  
26 to be linear (Kim et al., 2009), the mixing ratios are presented as relative values without units.  
27 Uncertainty in all the PTR-ToF-MS measurements is estimated at 20%. The 1-min  
28 concentration measurements at the 6 levels were averaged in 30-min intervals. Due to  
29 instrument malfunction, 6.5% of the 30-min profiles from 22 days could not be used. To  
30 ensure adequate fetch and horizontal homogeneity, measurements were filtered to use only  
31 wind direction in the range 20° to 285° following Teklemariam et al. (2009) and Froelich  
32 (personal communication). This resulted in a removal of 17.8% of the remaining data.

## 1 2.3 Canopy Modelling

2 The canopy model is described in detail in Makar et al. (1999) and Stroud et al. (2005). It is a  
3 one-dimensional model which solves the equations

$$4 \frac{\partial C_{i,j}}{\partial t} = E_{i,j} + f_{i,j} + \frac{\partial}{\partial z} \left( K(z_j) \frac{\partial C_{i,j}}{\partial z} \right), \quad (1)$$

5 where  $C_{i,j}$ ,  $E_{i,j}$ , and  $f_{i,j}$  are the concentration, emissions rate, and rate of change due to  
6 chemical reactions, respectively, of the  $i$ th chemical species in the  $j$ th model layer,  $K(z_j)$  is the  
7 eddy diffusivity,  $t$  is time, and  $z$  is height. The model domain is 1001 m with a 1 m vertical  
8 resolution and a 1 min time resolution. Environmental inputs to the model are at a 30 min  
9 time resolution and include temperature ( $T$ ), relative humidity ( $Rh$ ), pressure, PAR, cloud  
10 fraction,  $O_3$  and NO concentrations, and vertical wind variance ( $\sigma_w$ ). Median diurnal  
11 variations of temperature, relative humidity, and incoming solar radiation are shown in Fig. 1.  
12 Leaf Area Index (LAI) profiles were updated for the Borden forest based on 2009  
13 measurements at a 2 m resolution. Eddy diffusivity ( $K$ ) above the canopy was output from  
14 the Global Environmental Multiscale (GEM) Model (Côté et al, 1998) for the Borden location  
15 concurrent with the time period of the study.

16

17 The atmospheric transport within the canopy model is based on a modified  $K$ -theory of  
18 vertical turbulent diffusion from Raupach (1989),

$$19 \frac{\partial C}{\partial t} = - \frac{\partial}{\partial z} \left( R(\tau/T_L) K \frac{\partial C}{\partial z} \right), \quad (2)$$

20 where the eddy diffusivity ( $K$ ) is modified by the factor  $R$ , which accounts for canopy effects  
21 on turbulence (so-called “near-field” effects). The variable  $R$  is dependent on the ratio of  $\tau /$   
22  $T_L$  as

$$23 R = \frac{[1 - \exp(-\tau/T_L)](\tau/T_L - 1)^{3/2}}{[\tau/T_L - 1 + \exp(-\tau/T_L)]^{3/2}}, \quad \tau/T_L > 1, \quad (3)$$

24 where  $\tau$  is a transport lifetime and  $T_L$  is the Lagrangian timescale. Makar et al. (1999) found  
25 the modelled isoprene measurements above the canopy agreed well with measurement at the  
26 Borden forest location using a transport lifetime of  $\tau = 1.17 T_L$ , while Stroud et al (2005)  
27 found that a value of  $\tau = 4.0 T_L$  improved results at the Duke Forest location. The model

1 version used in the Stroud et al. (2005) study was modified to use a diffusion scheme which  
2 allowed for the inclusion of deposition and emissions at the surface. Stroud et al. (2005)  
3 hypothesized that the difference in transport lifetime could be due to either the change in the  
4 diffusion scheme, or a difference in the canopy structure between the two forests. Initial tests  
5 of the model with the new diffusion scheme at the Borden forest location demonstrated that a  
6 transport lifetime of  $\tau = 4.0 T_L$  gave the best measurement/model comparison, which implies  
7 that the change in mixing timescale is due to the change in the diffusion scheme and not the  
8 canopy structure.

9

10 Makar et al. (1999) modelled an isoprene basal emission rate of  $17.5 \mu\text{g g}^{-1} \text{h}^{-1}$  and an  $\alpha$ -  
11 pinene basal emission rate of  $3.4 \mu\text{g g}^{-1} \text{h}^{-1}$ . The isoprene basal emission rate was generated  
12 to match model results to measurements of isoprene at 30 and 42-m heights, and the  $\alpha$ -pinene  
13 basal emission was based on a parameterization. Stroud et al. (2005) added sesquiterpenes to  
14 the model, represented by  $\beta$ -caryophyllene. In the Stroud et al. (2005) model, the basal  
15 emission rate of sesquiterpenes was set to 1/3 the monoterpene emission rate. They estimated  
16 the dry deposition velocities of  $\text{O}_3$  as  $4 \text{ mm s}^{-1}$ ,  $\text{HNO}_3$  as  $40 \text{ mm s}^{-1}$ ,  $\text{NO}_2$  as  $1 \text{ mm s}^{-1}$ ,  
17 isoprene as  $2 \text{ mm s}^{-1}$ , MACR+MVK as  $2 \text{ mm s}^{-1}$ , and surface emission of monoterpenes as  $69$   
18  $\mu\text{g m}^{-2} \text{h}^{-1}$ . Fluxes into the surface are calculated in the model as the product of deposition  
19 velocity and mixing ratio concentration at a 1-m height. No sesquiterpene surface emissions  
20 were included in the Stroud et al. (2005) model. These are the parameters we use for the first  
21 model run.

22

## 23 **3 Results**

### 24 **3.1 Mixing Ratio Measurements**

25 Diurnal mixing ratios in the study time period are shown in Fig. 2(a-f). Although  
26 sesquiterpenes are not calibrated, the relative measurement provides insight into their  
27 behaviour. All of the measured VOCs increase in mixing ratio at sunrise, peak between 14:00  
28 and 16:00 (local EDT), and decrease into the night, with the exception of monoterpenes and  
29 sesquiterpenes. Monoterpenes and sesquiterpenes decrease at sunrise to a minimum near  
30 13:00, and then increase through the rest of the day and night. Generally speaking, the

1 majority of the VOCs follow the diurnal temperature trend shown in Fig. 1a, while the  
2 terpenes follow a diurnal trend similar to relative humidity (Fig. 1b).

3

4 The range of the median VOC mixing ratios at  $z = 4$  m are compared to measurements of two  
5 other studies in Table 1. In the Karl et al. (2005) study, various VOCs were measured  
6 beneath a sweetgum and pine plantation using a sampling inlet which was continuously  
7 moving between the ground and a height of 24 m. The range of measured values are  
8 estimated from the Karl et al. 2005 study (their Fig. 7) for a height of  $z = 4$  m. The mixing  
9 ratios of methanol and acetone are nearly a factor of 3 higher during the day in the Karl et al.  
10 study compared to our measurements and the mixing ratios of MACR+MVK are more than a  
11 factor of 12 higher. In the Holzinger et al. (2005) study, monoterpene mixing ratios were  
12 measured at 5 heights in a pine plantation. The range of measured values are estimated from  
13 Holzinger et al. 2005 study (their Fig. 2) for a height of  $z = 4$  m. This measured range is  
14 similar to the diurnal range of mixing ratio seen in this study at the same height.

15

16 For turbulent transport of material, neglecting chemical effects, the change in mixing ratio  
17 with height (slope) can be an indication of the direction and relative magnitude of the material  
18 flux. The slope of VOC mixing ratio with height is determined between 1 and 6 m with a  
19 linear least-squares fit to each 30-m profile measurement. The diurnal trend of the slopes is  
20 shown for each VOC in Fig. 1(g-l). As with the mixing ratios, the slopes follow similar  
21 patterns, with the exception of the terpenes. Change in mixing ratio with height at night is  
22 generally near-zero, although positive acetone and negative sesquiterpene slopes are seen.  
23 The daytime slopes suggest an uptake of non-terpene VOCs and an emission of monoterpenes  
24 and sesquiterpenes by the soil and/or forest litter during the day. However, since these are  
25 chemically reactive species which may not be conserved with height, further investigation is  
26 necessary to demonstrate that these slopes are the result of surface deposition or emissions.

27

28 The peak daytime median fluxes from previous studies are compared in Table 1. There have  
29 been few studies which have measured VOC fluxes below the canopy and only a small  
30 number of VOCs have been measured. Measurements in a pine plantation (Schade and  
31 Goldstein, 2001) and a sweetgum and pine plantation (Karl and Guenther, 2005) give varied



1 results, with observations of both emission and uptake of methanol, and emission of acetone,  
2 isoprene, and monoterpenes. Although the change in mixing ratio with height (slope)  
3 suggests an apparent emission of monoterpenes in the Borden forest measurements, the slope  
4 of methanol, acetone and isoprene suggests a predominantly downward flux. This pattern is  
5 similar to the measurements of Karl et al. (2004) beneath a tropical forest, where uptake of  
6 methanol, acetone, and isoprene are also seen. The Karl et al. (2004) and Karl and Guenther  
7 (2005) fluxes were calculated between 0 and 5 m (compared to 1-6 m for this study), while  
8 the heights of the Schade and Goldstein (2001) measurements are not specified.

### 9 **3.2 Model Comparison**

10 In order to investigate whether or not the measured mixing ratio slopes are due to surface  
11 emissions and deposition, the 1-dimensional canopy model was run with and without surface  
12 deposition and fluxes. To best match measurements, the basal emission rates were modified  
13 to give  $21.2 \mu\text{g g}^{-1} \text{h}^{-1}$  for isoprene,  $2.3 \mu\text{g g}^{-1} \text{h}^{-1}$  for monoterpenes, and  $0.33 \mu\text{g g}^{-1} \text{h}^{-1}$  for  
14 sesquiterpenes (although this is only an estimate based on uncalibrated sesquiterpene  
15 measurements). This represents an increase of 20% and a decrease of 11.5% from the Makar  
16 et al. (1999) isoprene and monoterpene basal emission rates, respectively, and a factor of 3.5  
17 decrease from the Stroud et al. (2005) sesquiterpene basal emission rate. In the case of the  
18 Stroud et al. (2005) study, these differences are likely due to the different forest  
19 environments. The differences between the Makar et al. (1999) study and this one (which  
20 takes place in the same location) could be due to the model constraint to below canopy  
21 measurements at a height of 6 m, as opposed to the 30 and 42 m measurement constraints  
22 used by Makar et al. Because turbulent transport within the canopy, chemistry, and light  
23 penetration into the canopy can affect the vertical distribution of VOCs, the 20% increase in  
24 isoprene basal emission rate may be due to errors in the vertical distribution of isoprene.  
25 However, basal emission rate changes between the studies could also be due to changing  
26 forest composition between 1995 and 2009 and different temperature and moisture histories.

27

28 The model was run with two scenarios: a base case with no surface emissions or deposition,  
29 and an active surface case which included deposition of isoprene and MACR+MVK, and  
30 emissions of monoterpenes and sesquiterpenes. Based on initial test runs, day-time surface  
31 deposition velocities were modified to  $2.7 \text{ mm s}^{-1}$  for isoprene and  $1.2 \text{ mm s}^{-1}$  for

1 MACR+MVK. Surface emission rates were modified to  $63 \mu\text{g m}^{-2} \text{h}^{-1}$  for monoterpenes and  
2  $0.86 \mu\text{g m}^{-2} \text{h}^{-1}$  for sesquiterpenes. This represents a 35% increase for isoprene, a 40%  
3 decrease of MACR+MVK, a 7% decrease for monoterpenes, and 19% decrease for  
4 sesquiterpenes from the Stroud et al. (2005) deposition velocities and emission rates at a pine  
5 plantation.

6

7 The resulting modelled mixing ratios at a height of 6 m are compared to the measurements in  
8 Fig. 3(a-d). The active surface has a negligible effect on the mixing ratios compared to the  
9 base case scenario. Generally, the model output of isoprene and monoterpene mixing ratios  
10 are with the quartiles of measured values. MACR+MVK are over-predicted by the model in  
11 the late afternoon. Both monoterpenes and sesquiterpenes decrease to background levels in  
12 the model between 6:00 and 8:00 (sunrise), while the measurements reduce more slowly,  
13 between 6:00 and 11:00. Model statistics for the base case are compared in Table 2. The  
14 model demonstrates relatively good agreement for isoprene ( $r^2 = 0.5$  and 65% of modelled  
15 values between 50% and 150% of the observations), moderate agreement for MACR+MVK  
16 and monoterpenes ( $r^2 \sim 0.3$ ), and poor agreement for sesquiterpenes ( $r^2 = 0.1$ ).

17

18 The slopes of mixing ratio with height were calculated from model results between 1 and 6 m  
19 from a linear least-squares fit. The slopes for the daytime hours are compared in Fig. 3(e-h).  
20 In the base model case, the slopes of isoprene are negative, compared to the positive slopes  
21 seen in the measurements, and the modelled slopes of MACR+MVK, monoterpenes, and  
22 sesquiterpenes are negligible relative to the measurements. Inclusion of isoprene deposition  
23 in the model gives slopes similar to the observed values. Deposition of MACR+MVK gives  
24 slopes which underestimate the observed slopes in the morning and overestimate the observed  
25 slopes in the afternoon. Emission of monoterpene and sesquiterpenes from the surface by the  
26 model produces slopes which are similar to the observed values.

27

28 There is good agreement between the measured and modelled slopes for isoprene and  
29 MACR+MVK at night between 20:00 and 8:00 (not shown). During this time period the  
30 mixing ratios and slopes of isoprene and MACR+MVK are near zero, as shown in Fig.  
31 2(b,d,h,j). Because the ground fluxes in the active surface model case are input as deposition

1 velocities for isoprene and MACR+MVK, the near-zero concentrations at night result in low  
2 deposition fluxes (from the definition of deposition velocity as the ratio of flux to  
3 concentration). For monoterpenes and sesquiterpenes, which are input as surface fluxes in the  
4 active surface model case, the modelled slopes are non-zero at night, which is in disagreement  
5 with the observed slopes, as shown in Fig. 4(k,l). Due to the difficulty in determining fluxes  
6 at night and the disagreement between measured and modelled slopes during this period, the  
7 actual behaviour of terpene emissions during the night remains unclear.

8

## 9 **4 Discussion**

### 10 **4.1 Canopy-top Fluxes**

11 Based on these model results, the average fluxes from the surface due to deposition of  
12 isoprene ( $2.7 \text{ mm s}^{-1}$ ) and MACR+MVK ( $1.2 \text{ mm s}^{-1}$ ) and emissions of monoterpenes and  
13 sesquiterpenes are calculated for the 22 day period (after filtering for wind direction). These  
14 are compared to the average modelled fluxes out the canopy top (upward positive) in Table 3  
15 for the active surface case model run. Due to the uncertainty in the night-time emissions of  
16 terpenes, as discussed in Section 3.2, the monoterpene and sesquiterpene surface emissions  
17 are presented as a range between zero emissions at night and constant emissions for the full  
18 24 hour period. The average emissions seen here are much larger than those measured by  
19 Aaltonen et al. (2011) in a boreal Scots pine forest in southern Finland. Aaltonen et al.  
20 measured yearly average emissions of  $0.05 \mu\text{g m}^{-2} \text{ h}^{-1}$  for isoprene,  $5 \mu\text{g m}^{-2} \text{ h}^{-1}$  for  
21 monoterpenes, and  $0.04 \mu\text{g m}^{-2} \text{ h}^{-1}$  for sesquiterpenes. These large differences are likely due  
22 differences in the climates, tree compositions and densities, and seasons. For example, in the  
23 Aaltonen et al. (2011) study,  $\alpha,\beta$ -pinene emissions increased by a factor between 3 and 5 in  
24 June and October. Similar seasonal difference could change results at this mixed forest  
25 location for different time periods.

26

27 As shown in Table 3, the model suggests that the amount of isoprene that is emitted from the  
28 forest to the atmosphere is reduced by approximately 2% due to surface uptake. Although  
29 this is a relatively small fraction of the total, on a regional scale this could have a significant  
30 effect on the total isoprene released from forests. With the inclusion of MACR+MVK  
31 deposition, there is negligible release of MACR+MVK to the atmosphere from the canopy

1 over the period of the study and any MACR+MVK that is created as a by-product of chemical  
2 reactions is, on average, absorbed at the forest surface. This is consistent with the small  
3 canopy-top fluxes for MACR+MVK seen in other studies (Karl et al. 2004; Spirig et al.  
4 2005).

5

6 According to model results, the emission of monoterpenes from the surface represents a  
7 significant amount (15 to 27%) of what is released from the canopy-top to the atmosphere.  
8 This is a much larger fraction than the 10% which was modelled by Stroud et al. (2005) for a  
9 pine plantation using the same canopy model. This is an important result for the  
10 interpretation of measured canopy top emissions relating to basal tree emission rates.  
11 Although a large fraction of monoterpene emissions may relate to leaf respiration, a  
12 significant amount may also be due to leaf litter or other unknown surface emissions. Hence,  
13 basal tree emission rates based on canopy top emission measurements may be overestimated.

14

15 Faiola et al. (2014) measured VOC emissions from soil and leaf litter samples collected from  
16 a pine and larch forest in Idaho. They measured peak monoterpene emission rates from the  
17 samples of  $258 \mu\text{g m}^{-2} \text{h}^{-1}$ . Comparing these measurements to model results, they conclude  
18 that surface monoterpene emissions could account for between 12 and 136% of the canopy  
19 emissions in the spring and fall.

20

21 Although sesquiterpene measurements in this study were uncalibrated, the model output  
22 provides some insight into the relative surface flux of sesquiterpenes. According to the model  
23 estimates, approximately 2 to 5% of sesquiterpenes emitted from canopy top are due to  
24 surface emissions. Although we are not able to confirm the basal tree and surface emission  
25 rates used in the model, sesquiterpene emissions from canopies are typically more than an  
26 order of magnitude lower than they are for primary VOCs such as isoprene, suggesting that  
27 the emissions of sesquiterpenes from the sub-canopy surface is not significant.

28

## 1 4.2 Model Sensitivity and Uncertainty

2 To estimate the error in the model due to the estimation of model input NO from nearby  
3 measurements, two 2-day model runs were done with modified NO levels. The active surface  
4 case was used with a 50% increase in NO levels and a 50% reduction in NO. The resulting  
5 change in the average mixing ratio and daytime slope of isoprene, MACR+MVK,  
6 monoterpenes, and sesquiterpenes are listed in Table 4. Changes in average mixing ratios and  
7 slopes for these compounds are generally less than or near  $\pm 50\%$ .

8

9 As discussed in Section 2.2, the Egbert NO measurements used as a proxy for the Borden  
10 measurements correlate well ( $r^2 = 0.68$ ) with an average of 0.26 ppb at Egbert and 0.27 ppb at  
11 Borden. However, there is a high root-mean-square error ( $E_{rms} = 0.66$  ppb) over the same  
12 period, primarily due to high NO values seen at only one site during non-aligned wind  
13 directions. In the model sensitivity runs, the average VOC mixing ratio and slope changes  
14 due to modified NO levels are consistently smaller than the 50% increase and decrease in NO.  
15 Hence, it would be expected that over the 22 days of the study, the average error due to the  
16 use of proxy NO levels should be near the 6% average difference in NO between Egbert and  
17 Borden; however hourly errors may be much higher, as suggested by the relatively high root-  
18 mean-square error.

19

20 Further sensitivity tests were run for the same 2-day period with modified transport lifetime ( $\tau$   
21 in Eqs. 2 and 3), O<sub>3</sub> levels, and solar flux ( $I$ , photons m<sup>-2</sup> s<sup>-1</sup> nm<sup>-1</sup>). The modifications result in  
22 average isoprene mixing ratio changes between -17% and 53% and average isoprene slope  
23 changes between -27% and 108%. The isoprene slope is very sensitive to light conditions, as  
24 demonstrated by the reduced solar flux run ( $0.5I$ ), which results in a 108% increase in average  
25 slope. Monoterpenes are generally much less sensitive to the modifications, with average  
26 mixing ratio changes between -12% and 15% and average slope changes between -6% and  
27 21%. Although it is difficult to quantify the uncertainty in input variables such as  $\tau$ , O<sub>3</sub>, and  $I$ ,  
28 these tests demonstrate that the model error should be on the same scale as these uncertainties.  
29 For example, the 20% increase in isoprene basal emissions rates relative to the Makar et al.  
30 (1999) study could be due to an uncertainty of the same scale introduced by the below canopy  
31 (6-m) model constraint. A change in light conditions in the subcanopy will change the

1 amount of isoprene and necessitate a change in basal emission rate to compensate. Hence the  
2 error in isoprene slope (and hence the derived deposition) may be as much as 40%, while  
3 error in monoterpene emissions are likely less.

#### 4 **4.3 A Comparison to Profile Method Fluxes**

5 It is generally considered incorrect to derive flux-profile relationships below the canopy, as  
6 the assumption of uniform mixing is violated by canopy structures (e.g. Garratt, 1994).  
7 However, a recent study (Vickers and Thomas, 2014) demonstrates accurate measurements of  
8 heat flux beneath a canopy using the flux-profile relationship. To investigate the utility of this  
9 approach here, surface fluxes are derived from the mixing ratio measurements between the  
10 heights of 1 and 6 m. The derivation of the fluxes is outlined in the supplementary material.  
11 Figure 4 compares the day-time profile fluxes (derived from both the measurements and the  
12 model output between heights of 1 and 6 m) and the surface exchange (deposition and ground  
13 emissions) specified in the active surface model run. The fluxes derived from the profile  
14 method systematically overestimate the specified surface exchange fluxes. However, the  
15 specified surface exchange fluxes are generally within the upper and lower quartiles of the  
16 profile method fluxes. Although the profiles may not be quantitatively accurate, they are  
17 directionally consistent. This suggests that the positive slopes of methanol and acetone (Fig.  
18 2) demonstrate a deposition of these VOCs to the surface, in contrast to measured emissions  
19 from litter samples (Warneke et al., 1999).

20

#### 21 **5 Conclusions**

22 Direct measurements of VOC mixing ratios between 1 and 6 m demonstrate a trend of  
23 increasing methanol, isoprene, acetone, and MACR+MVK during the day, and increasing  
24 monoterpenes and sesquiterpenes during the night for this forest location. Mixing ratios of  
25 methanol, acetone, and monoterpenes compare well to other near-surface measurements made  
26 at pine plantations and plantations with both sweetgum and pine.

27

28 The change in mixing ratio with height suggests a downward flux of methanol, isoprene,  
29 acetone, and MACR+MVK during the day, and an upward flux of monoterpenes and  
30 sesquiterpenes during the day. Slopes through the night are generally near zero. There is a

1 large amount of variation in observed fluxes of VOCs near the surface in previous studies;  
2 however directional consistency is seen between measurements at the Borden forest and the  
3 deposition of methanol, isoprene, and acetone in a tropical forest (Karl et al., 2004). There is  
4 also directional consistency with the emission of monoterpenes from a Scots pine forest  
5 (Aaltonen et al., 2011).

6

7 A 1-dimensional canopy model was used to determine if the change in VOC mixing ratios  
8 with height were due to deposition and emissions of VOCs to the sub-canopy surface. Model  
9 results suggest a deposition of  $2.7 \text{ mm s}^{-1}$  for isoprene which results in average downward  
10 (negative) surface flux of  $4.9 \mu\text{g m}^{-2} \text{ h}^{-1}$  for the duration of the study. This represents 2% of  
11 the magnitude of the canopy top emissions of isoprene to the atmosphere. Model results  
12 suggest a deposition of  $1.2 \text{ mm s}^{-1}$  for MACR+MVK which results in an average downward  
13 (negative) surface flux of  $1.4 \mu\text{g m}^{-2} \text{ h}^{-1}$ . This compensates for the formation of  
14 MACR+MVK in the canopy resulting in negligible emissions of MACR+MVK from the  
15 canopy top into the atmosphere. Results suggest a surface emission of  $63 \mu\text{g m}^{-2} \text{ h}^{-1}$  for  
16 monoterpenes, which comprises 15 to 27% of the total emissions from the canopy-top into the  
17 atmosphere. This represents a significant fraction of the emitted monoterpenes, suggesting  
18 that forest surface emissions at this location are comparable in scale to the emissions from  
19 tree foliage and should be taken into account in canopy modelling. These results are similar  
20 to the results of Faiola et al. (2014), who suggest that surface monoterpene emissions could  
21 account for between 12 and 136% of the canopy emissions in the spring and fall. Results  
22 were less conclusive for sesquiterpenes, owing in part to a lack of calibration standard.  
23 However, it appears that the emissions of sesquiterpenes from the sub-canopy surface are  
24 generally not significant.

25

26 This study represents explorative research into VOC deposition and emissions from the sub-  
27 canopy surface of a mixed forest location. Further study is necessary in order to study the  
28 variation of surface emissions seasonally and the behaviour of terpene emissions during the  
29 night, and to quantify VOC concentrations for the entire forest height, which would allow  
30 verification of model results for the full height of the canopy. The VOC exchange to and  
31 from soil and litter at the surface is also poorly constrained (e.g. Stacheter et al., 2013; Faiola

1 et al., 2014; Leff and Fierer, 2008) and further investigation is required to identify the sources  
2 of emission and deposition.

3

#### 4 **Acknowledgements**

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6 Canada and was funded by the Science and Technology Branch, Environment Canada. The  
7 authors thank the Canadian Forces Base Borden for hosting the Flux Tower site.



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- 11

1 Table 1. A comparison of understory minimum and maximum median mixing ratios and peak  
 2 fluxes. Karl et al., 2005 (K05) and Karl and Guenther, 2005 (KG05) were in a sweetgum and  
 3 pine plantation. Holzinger et al., 2005 (H05) and Schade and Goldstein, 2001 (SG01) were in  
 4 a pine plantation. Karl et al., 2004 (K04) was in a tropical forest

| VOC          | Min-Max MR at 4 m [ppb] |           |         | Peak Fluxes [ $\mu\text{g m}^{-2} \text{h}^{-1}$ ] |      |      |
|--------------|-------------------------|-----------|---------|--|------|------|
|              | This Study              | K05       | H05     | KG05   | K04  | SG01 |
| Methanol     | 0.5 - 2.4               | 0.5 - 7.0 |         | -30  | -70  | 250  |
| Isoprene     | 0.1 - 0.6               |           |         | 30   | -100 |      |
| Acetone      | 0.7 - 1.6               | 1.0 - 4.0 |         |  | -10  | 50   |
| MACR+MVK     | 0.02 - 0.2              | ~0 - 2.5  |         |  |      |      |
| Monoterpenes | 0.4 - 1.9               |           | 0.5 - 2 | 50   |      |      |

5

6

1 Table 2. A comparison of mean mixing ratios from observation ( $\mu_O$ ) and model output ( $\mu_M$ )  
 2 for the 22 day period. Coefficient of correlation ( $r^2$ ), root-mean-square error ( $E_{rms}$ ) and the  
 3 fraction of modelled 30-min averages between 50% and 150% of the observed 30-min  
 4 averages are listed.

|                | $\mu_O$ | $\mu_M$ | $r^2$ | $E_{rms}$ | $\pm 50\%$ |
|----------------|---------|---------|-------|-----------|------------|
|                | ppb     | ppb     |       | ppb       | %          |
| Isoprene       | 0.30    | 0.32    | 0.51  | 0.25      | 64.6       |
| MACR+MVK       | 0.16    | 0.10    | 0.26  | 0.19      | 39.6       |
| Monoterpenes   | 0.94    | 1.27    | 0.27  | 1.09      | 42.0       |
| Sesquiterpenes | 0.01    | 0.01    | 0.11  | 0.01      | 37.4       |

5

6

1 Table 3. The average modelled fluxes (positive upward) from the surface ( $F_S$ ) and at the  
 2 canopy top ( $F_C$ ) in units of  $\mu\text{g m}^{-2} \text{h}^{-1}$ . A range is given for terpenes with emissions from  
 3 8:00-20:00 EDT only (no night-time emissions) and emissions from 0:00-24:00 (the full day).

|                | $F_S$ |       | $F_C$ | $F_S / F_C$ |       |
|----------------|-------|-------|-------|-------------|-------|
|                | 8-20h | 0-24h |       | 8-20h       | 0-24h |
| Isoprene       | -5.2  |       | 243.5 | -2.1%       |       |
| MACR+MVK       | -1.5  |       | -0.01 |             |       |
| Monoterpenes   | 31.4  | 62.9  | 234.7 | 15.4%       | 26.8% |
| Sesquiterpenes | 0.4   | 0.9   | 18.7  | 2.3%        | 4.7%  |

4

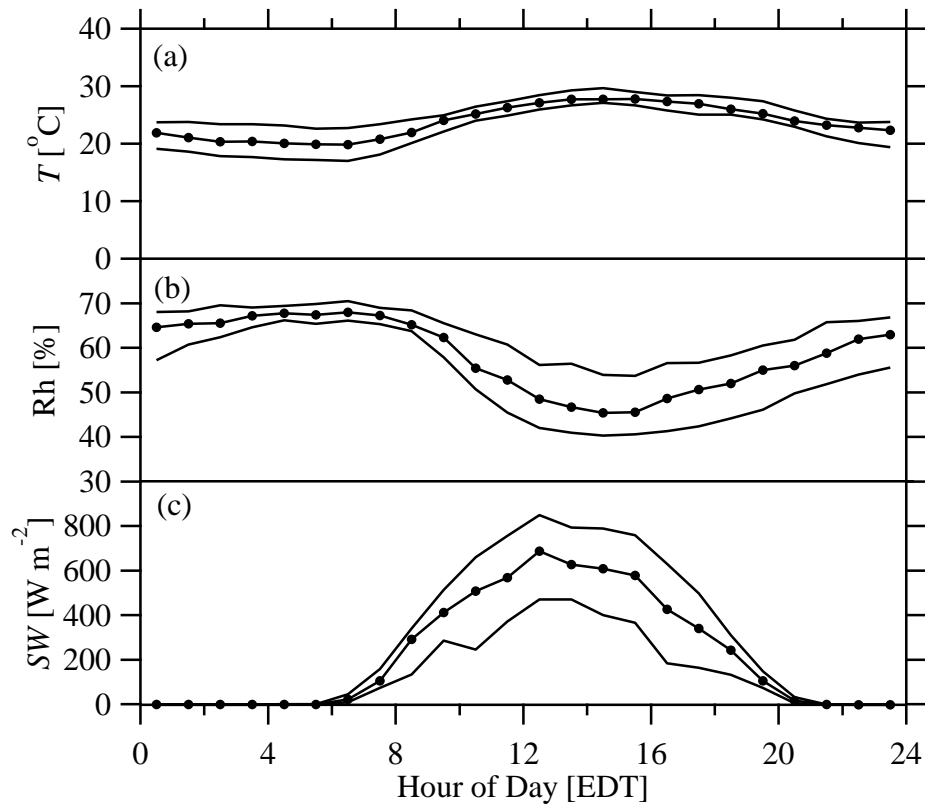
5



1 Table 4. Percentage change in average mixing ratio and slope for a 2-day model run with  
 2 modified: input NO; transport lifetime ( $\tau$ ), input O<sub>3</sub>; and input solar flux. Changes are relative  
 3 to the “active surface” model run.

| Mixing Ratios  | 1.5NO | 0.5NO | 1.5 $\tau$ | 0.5 $\tau$ | 1.5O <sub>3</sub> | 0.5O <sub>3</sub> | 1.5I  | 0.5I  |
|----------------|-------|-------|------------|------------|-------------------|-------------------|-------|-------|
| Isoprene       | 13.7  | -16.8 | -0.5       | 7.8        | 5.9               | -9.5              | -17.1 | 53.0  |
| MACR+MVK       | 7.9   | -14.0 | -0.3       | 0.3        | 5.4               | -11.3             | -20.8 | 12.9  |
| Monoterpenes   | 8.3   | -12.9 | -0.7       | 10.0       | -11.8             | 14.5              | -3.3  | 6.3   |
| Sesquiterpenes | 25.5  | -30.1 | -0.7       | 8.6        | -18.2             | 40.1              | -1.0  | 1.6   |
| Slopes         | 1.5NO | 0.5NO | 1.5 $\tau$ | 0.5 $\tau$ | 1.5O <sub>3</sub> | 0.5O <sub>3</sub> | 1.5I  | 0.5I  |
| Isoprene       | 24.9  | -27.0 | -3.0       | 46.2       | 13.9              | -19.3             | -31.7 | 107.5 |
| MACR+MVK       | 12.9  | -22.4 | -2.3       | 28.8       | 7.8               | -16.4             | -29.5 | 22.5  |
| Monoterpenes   | -2.3  | 2.5   | -1.6       | 21.0       | 1.3               | -2.7              | 2.7   | -6.0  |
| Sesquiterpenes | 2.1   | -2.9  | -1.9       | 29.6       | -12.5             | 16.8              | -1.2  | 0.7   |

4

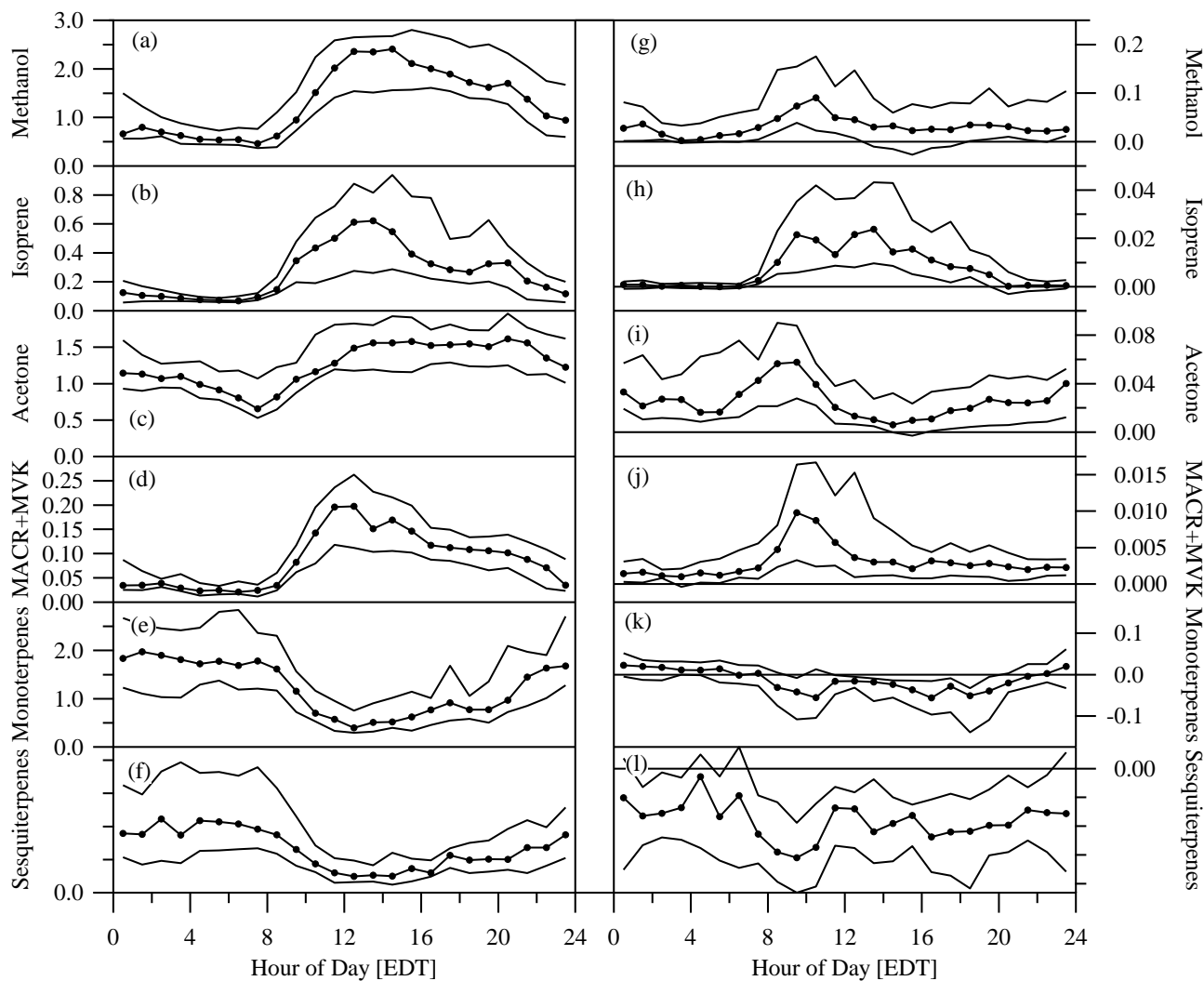


1

2 Figure 1. Hourly median and quartiles of (a) temperature,  $T$ , (b) relative humidity, Rh, and  
 3 (c) incoming solar radiation,  $SW$ , all at a height of 33 m, for the 22 days of the study.

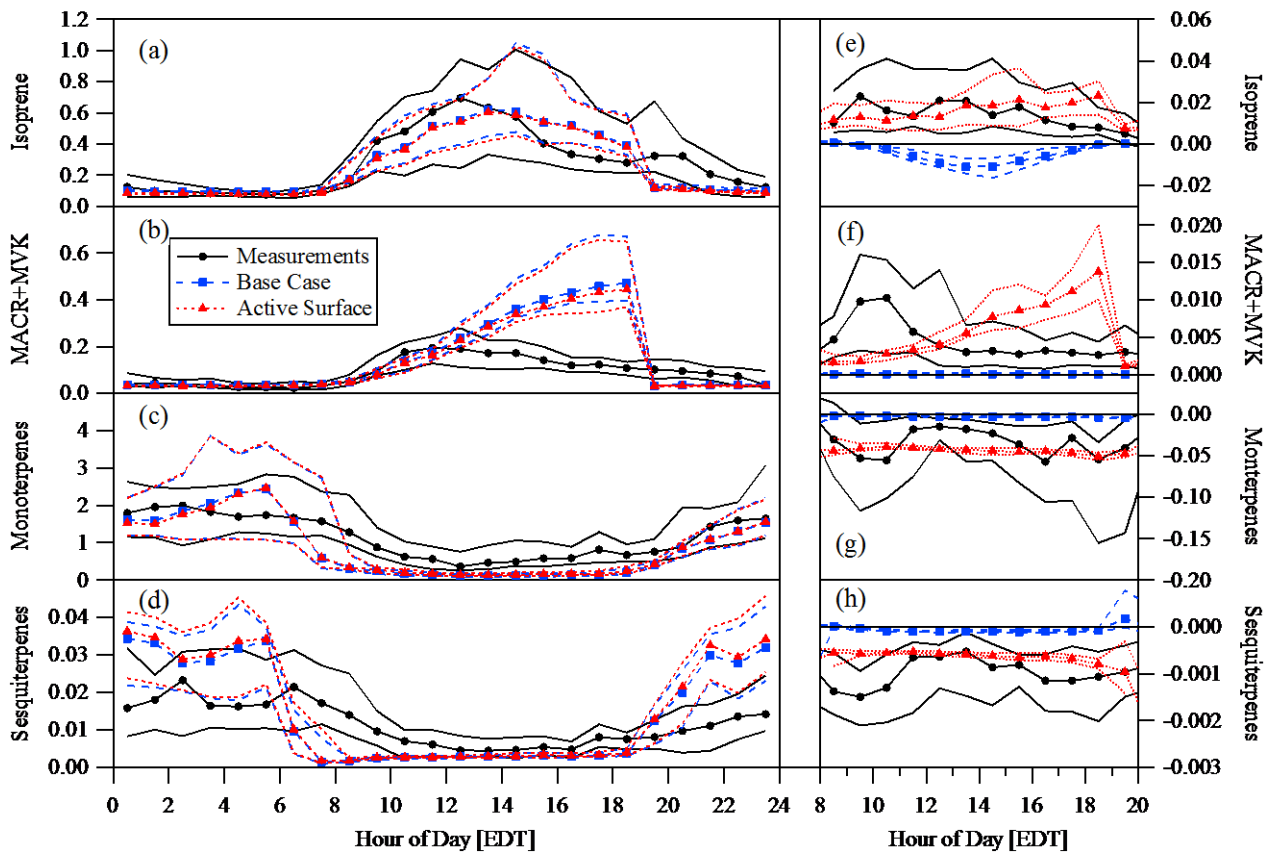
4

5



1  
 2 Figure 2. Hourly medians and quartiles of VOC mixing ratios (a-f) and slopes (g-l) over the  
 3 22 day measurement period (after filtering for wind direction). Mixing ratios are at the 2-m  
 4 level in units of ppb. Slopes are determined as a least-squares fit between 1 and 6 m and are  
 5 in units of ppb m<sup>-1</sup>. Sesquiterpene mixing ratio and slope (f, l) are not calibrated and are  
 6 presented as relative units only.

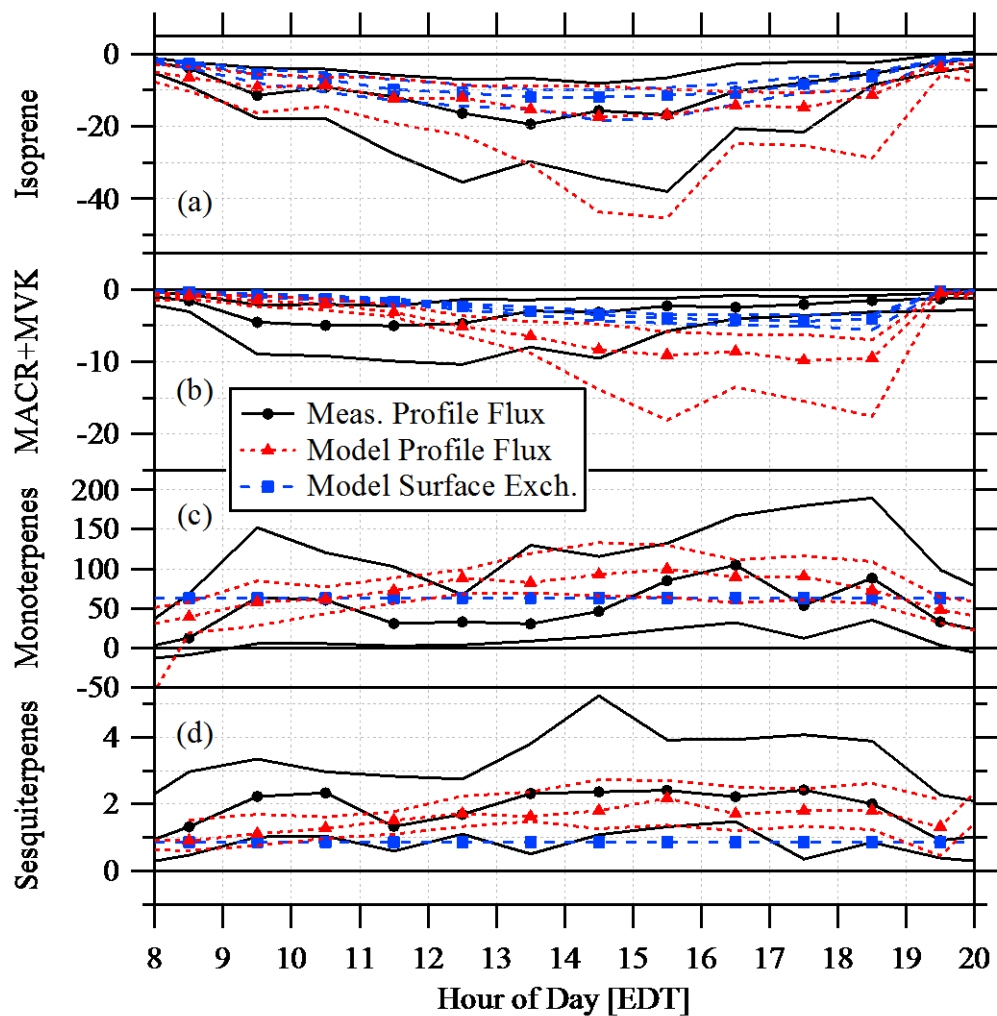
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2 Figure 3. Hourly medians and quartiles of measured and modelled mixing ratios (a-d) and  
 3 daytime slopes (e-h). Mixing ratios are at the 6-m level in units of ppb. Slopes are in units of  
 4 ppb m<sup>-1</sup>. Measurements (black circles), the base model case (blue squares), and the active  
 5 surface case (red triangles) are compared. Sesquiterpene measured mixing ratios and slopes  
 6 (d,h) are not calibrated and are presented with units in order to compare to model output.

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 2 Figure 4. Hourly medians and quartiles of fluxes as determined from the profile method and  
 3 compared to the modelled surface exchange fluxes (blue squares). The profile method is used  
 4 to estimate surface fluxes with both the measurements (black circles) and the model results  
 5 (red triangles) between 1 and 6-m heights. Fluxes are in units of  $\mu\text{g m}^{-2} \text{h}^{-1}$ .