



**Strong atmospheric  
new particle  
formation in winter,  
urban Shanghai,  
China**

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## Abstract

Particle size distributions in the range of 1.34–615.3 nm were recorded from 25 November 2013 to 25 January 2014 in urban Shanghai, using a combination of one nano Condensation Nucleus Counter system (nCNC), one nano-Scanning Mobility Particle Sizer (SMPS), and one long-SMPS. Measurements of sulfur dioxide by an SO<sub>2</sub> analyzer with pulsed UV fluorescence technique allowed calculation of sulfuric acid proxy. In addition, concentrations of ammonia were recorded with a Differential Optical Absorption Spectroscopy (DOAS). During this 62-day campaign, 13 NPF events were identified with strong burst of sub-3 nm particles and subsequent fast growth of newly formed particles. The observed nucleation rate ( $J_{1.34}$ ), formation rate of 3 nm particles ( $J_3$ ), and condensation sink (CS) were 112.4–271.0 cm<sup>-3</sup> s<sup>-1</sup>, 2.3–19.2 cm<sup>-3</sup> s<sup>-1</sup>, and 0.030–0.10 s<sup>-1</sup>, respectively. Subsequent cluster/nanoparticle growth showed a clear size dependence, with average values of GR<sub>1.35~1.39</sub> (from the bin of 1.34–1.37 nm to the bin of 1.37–1.41 nm), GR<sub>1.39~1.46</sub> (from 1.37–1.41 to 1.41–1.52 nm), GR<sub>1.46~1.70</sub> (from 1.41–1.52 to 1.52–1.89 nm), GR<sub>1.70~2.39</sub> (from 1.52–1.89 to 1.89–3.0 nm), GR<sub>2.39~7</sub> (from 1.89–3.0 to 7 nm), and GR<sub>7~20</sub> (from 7 to 20 nm) being  $1.6 \pm 1.0$ ,  $1.4 \pm 2.2$ ,  $7.2 \pm 7.1$ ,  $9.0 \pm 11.4$ ,  $10.9 \pm 9.8$ , and  $11.4 \pm 9.7$  nm h<sup>-1</sup>, respectively. Correlation between nucleation rate ( $J_{1.34}$ ) and sulfuric acid proxy indicates that nucleation rate  $J_{1.34}$  was proportional to a 0.64 power of sulfuric acid proxy. Correlation between nucleation rate ( $J_{1.34}$ ) and gas-phase ammonia suggests that ammonia was associated with NPF events. The calculated sulfuric acid proxy was sufficient to explain the subsequent growth of 1.34–3 nm particles, but insufficient for particles exceeding this size range. Qualitatively, NPF events in urban Shanghai likely occur on days with low levels of PM<sub>2.5</sub>.

## 1 Introduction

Aerosol particles can influence climate directly and indirectly (Andreae and Crutzen, 1997; Haywood and Boucher, 2000; IPCC, 2013), and have adverse impact on human

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health (Dockery et al., 1993; Laden et al., 2006; Pope and Dockery, 2006). Atmospheric nucleation of gas-phase precursors to clusters, and then further to nanoparticles is the largest source of atmospheric aerosol particles (Kulmala et al., 2004; Zhang et al., 2012). This phenomenon has been observed in numerous locations around the world, including areas with a pristine atmosphere, e.g., coastal areas (O'Dowd et al., 2002), Antarctic/Arctic (Park et al., 2004), remote forest (Dal Maso et al., 2005), semi-rural locations with very low pollution levels such as Kent, OH (Kanawade et al., 2012), and heavily polluted cities, such as Mexico City (Dunn et al., 2004).

The exact mechanism for atmospheric nucleation is still under active investigation. Field measurements and laboratory studies have shown that sulfuric acid is a key precursor species for atmospheric nucleation (Weber et al., 1996; Sipila et al., 2010), and that atmospheric nucleation rate can be significantly promoted in presence of other precursors including ammonia (Ball et al., 1999; Benson et al., 2009), amines (Berndt et al., 2010; Zhao et al., 2011), and organic acids (Zhang et al., 2004, 2009). At certain locations, ion-induced nucleation (Yu and Turco, 2001; Lee et al., 2003) or iodine species (O'Dowd et al., 2002) may also help to explain the observed new particle formation. Very recently, with the application of Particle Size Magnifier (PSM) and Chemical Ionization Atmospheric Pressure interface Time-of-Flight (CI-API-ToF) mass spectrometer, progress have been made by combining the Cloud (Cosmics Leaving OUtdoor Droplets) chamber experiments and ambient observations including those at Hyytiälä, Finland, showing that oxidation products of biogenic emissions, together with sulfuric acid, contribute to new particle formation in the atmosphere (Schobesberger et al., 2013; Riccobono et al., 2014).

China suffers severe air pollution especially high atmospheric particle loadings in recent years (Chan and Yao, 2008). Among many potential sources of atmospheric particles, atmospheric nucleation has been suggested to be a significant source of particles (Matsui et al., 2011; Yue et al., 2011). Correspondingly, a number of extensive campaign or long-term observational studies have been carried out in the Beijing area (e.g., Wu et al., 2007; Yue et al., 2009; Zhang et al., 2011; Gao et al., 2012) and Pearl

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River Delta, including Hong Kong (e.g., Guo et al., 2012; Yue et al., 2013). As one of the most industrialized area of China, one of the most populated area in the world, and one of the hotspots for particle pollution, Yangtze River Delta (YRD) has only seen a few research activities on atmospheric nucleation (Du et al., 2012; Herrmann et al., 2014).

5 Among the few studies, measurements at the station for Observing Regional Processes of the Earth System, Nanjing University (SORPES-NJU) offered a first insight for new particle formation in the western part of YRD (Herrmann et al., 2014). On the other hand, atmospheric nucleation research in China is still in its infant stage and the latest experimental techniques are yet to be applied in China. For example, data on freshly  
10 nucleated particles are really sparse, except for those from an air ion spectrometer employed at SORPES-NJU (Herrmann et al., 2014). To the best of our knowledge, the employment of a Particle Size Magnifier (PSM), which is able to measure atmospheric nucleation at the size (mobility diameter) down to  $1.5 \pm 0.4$  nm (Kulmala et al., 2012), has not been reported in a Chinese location in literature. The lack of key information  
15 greatly hinders a better understanding of nucleation mechanisms in China.

In this study, we measured atmospheric nucleation from 25 November 2013 to 25 January 2014 in urban Shanghai with nCNC and two sets of SMPS. Nucleation frequency, nucleation rate ( $J_{1,34}$ ), condensation sink (CS), and growth rates (GR) have been reported and compared with previous studies with similar or dissimilar atmo-  
20 spheric environments. In addition, the potential nucleation mechanism was explored by correlating sulfuric acid proxy calculated from sulfur dioxide precursor and gas-phase ammonia to nucleation rate ( $J_{1,34}$ ). The competition between background particle levels and available condensable sulfuric acid that determines whether an atmospheric nucleation occurs or not has also been discussed.

## 2 Experimental

### 2.1 Nucleation measurements

Nucleation measurements were carried out on the rooftop of a teaching building (31°18' N, 121°30' E) that is about 20 m above ground on the campus of Fudan University between 25 November 2013 and 25 January 2014. This monitoring site is mostly surrounded by commercial properties and residential dwellings. The Middle Ring Road, one of main overhead highways in Shanghai, lies about 100 m to the south of the site. Hence, the Fudan site can be treated as a representative urban site influenced by a wide mixture of emission sources (Wang et al., 2013; Ma et al., 2014).

Ambient particle size distributions in the range of 1.34–615.3 nm were measured using a combination of one nano Condensation Nucleus Counter system (model A11, Airmodus, Finland), one nano-SMPS (consisting of one DMA3085 and one CPC3776, TSI, USA), and one long-SMPS (consisting of one DMA3081 and one CPC3775, TSI, USA). The instruments were continuously running except for maintenance and minor instrument breakdown during the campaign.

Ambient air was drawn into a stainless steel manifold of 5.0 m length and 4 inch inner diameter at a flow rate of 153 CFM using a low-volume blower (Model DJT10U- 25M, NUSSUN, China). From this main manifold, 1.75 lpm ambient air was drawn through a 1/4 inch inner diameter stainless tube of 18 cm length, and diluted with a zero air flow generated by a zero air generator (Model 111, Thermo, USA) at a ratio of 1 : 1 to reduce the overall relative humidity (RH) and the number of particles entering PSM, since high RH and particle loading had an impact on the saturation of diethylene glycol in PSM and hence data quality. Subsequently 2.5 lpm diluted air was sampled into nCNC. In addition, 30 lpm split flow was drawn from the main manifold through a 1/4 inch inner diameter conductive silica tubing of 50 cm length, and then 0.3 and 1.5 lpm ambient air from the split flow, respectively, were drawn into nano-SMPS and long-SMPS. The calculated diffusion loss for 1.35 nm particles is 29 % with the above setup.

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The nCNC system consists of one PSM (model A10, Airmodus, Finland) and one butanol Condensation Particle Counter (bCPC, model A20, Airmodus, Finland), and was used to measure size distributions of 1.34–3 nm clusters/particles. Briefly, PSM activates the smallest particles using diethylene glycol as a working fluid and condensationally grow nanoparticles up to larger than 90 nm in mobility equivalent diameter, after which an external CPC is used for further growing the particles to optical sizes and counting the grown particles (Vanhanen et al., 2011). In this study, PSM was used in the scanning mode in which the saturator flow rate is changed continuously, giving an activation spectrum of the measured particles to obtain size distribution of sub-3 nm clusters/particles. A scanning cycle of 100 steps between saturator flow rates 0.1–1 lpm and a time resolution of 220 s were chosen. Sub-3 nm clusters/particles were classified into 5 bins, i.e., 1.34–1.37, 1.37–1.41, 1.41–1.52, 1.52–1.89, 1.89–3.0 nm, respectively. Geometric mean values of upper and lower limits of the five bins, i.e., 1.35, 1.39, 1.46, 1.70, and 2.39 nm, respectively, were used to refer to the five bins in the growth rate calculation.

The nano-SMPS measured particles in the size range from 3 to 64 nm and the long-SMPS recorded particles from 14 to 615.3 nm. For both SMPSs, 64 size bins and 5 min time resolution were chosen. The sample flow to sheath flow ratios for both SMPSs were set at 1 : 10. A comparison analysis on the total particle concentrations between 14 and 64 nm measured by both nano-SMPS and long-SMPS showed less than 10 % difference in the size range of 55–64 nm between two SMPSs. Hence, number concentrations of particles in the size range of 3–615 nm,  $N_{3\sim 615}$ , were calculated from the sum of  $N_{3\sim 55}$  obtained from nano-SMPS,  $N_{55\sim 64}$  from the arithmetic average of nano-SMPS and long-SMPS, and  $N_{64\sim 615}$  from long-SMPS.

At the same site, sulfur dioxide ( $\text{SO}_2$ ) was measured by an  $\text{SO}_2$  analyzer with pulsed UV fluorescence technique (Model 43i, Thermo, USA) and calibration of this  $\text{SO}_2$  analyzer was performed every month. A DOAS system was used to measure the integrated concentration of  $\text{NH}_3$  along the optical path between a transmitter telescope using a 35W Deuterium lamp as the light source and a receiver telescope (53 m), and then

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to yield the average concentration of  $\text{NH}_3$  through dividing the integrated concentration by the absorption length (Platt and Stutz, 2008). In this study, the concentration of  $\text{NH}_3$  was determined by fitting the reference spectra to the atmospheric spectra in a given window (205–220 nm) using a nonlinear least-squares method, similarly to a previous measurement of HONO and  $\text{NO}_2$  (Wang et al., 2013). Detection limit of  $\text{NH}_3$  was about 1 ppb with a 3 min integration time. Mass concentrations of  $\text{PM}_{2.5}$  were measured by a Thermo Model 5030 Synchronized Hybrid Ambient Real-time Particulate Monitors equipped with a  $\text{PM}_{2.5}$  aerosol cutter.

Solar radiation intensity measured by a Pyranometer (Kipp & Zonen CMP6, Netherland) was obtained from Shanghai Pudong Environmental Monitoring Centre ( $31^\circ 14' \text{N}$ ,  $121^\circ 32' \text{E}$ , about 8.78 km from the Fudan site).

## 2.2 Date processing

### 2.2.1 Nucleation rate ( $J_{1.34}$ ), formation rate of 3 nm particles ( $J_3$ ), growth rate (GR), and condensation sink (CS)

In this study, PSM allowed measurements of clusters/particles down to 1.34 nm. Hence, atmospheric nucleation rate,  $J_{1.34}$ , defined as the flux of particles growing over 1.34 nm, can be calculated by taking into account the coagulation losses and growth losses (Kulmala et al., 2012),

$$J_{1.34} = \frac{dN_{1.34\sim 3}}{dt} + \text{CoagS}_{d_p=2\text{nm}} \cdot N_{1.34\sim 3} + \frac{1}{1.66\text{nm}} \text{GR}_{1.34\sim 3} \cdot N_{1.34\sim 3} \quad (1)$$

where  $\text{CoagS}_{d_p=2\text{nm}}$  represents coagulation sink of 2 nm particles, an approximation for that in the size interval of 1.34–3 nm; and  $\text{GR}_{1.34\sim 3}$  represents the apparent clusters/particle growth rate between 1.34 and 3 nm.

Formation rate of 3 nm particles was calculated in a similar way (Sihto et al., 2006; Kulmala et al., 2012), providing a comparison with previous studies,

$$J_3 = \frac{dN_{3\sim6}}{dt} + \text{CoagS}_{d_p=4\text{nm}} \cdot N_{3\sim6} + \frac{1}{3\text{nm}} \text{GR}_{3\sim6} \cdot N_{3\sim6} \quad (2)$$

5 where  $\text{CoagS}_{d_p=4\text{nm}}$  represents coagulation sink of 4 nm particles, an approximation for that in the size interval of 3–6 nm.

Growth rate (GR) is defined as the rate of change in the diameter of a growing particle population, using the maximum-concentration method (Kulmala et al., 2012),

$$10 \text{ GR} = \frac{dd_p}{dt} = \frac{\Delta d_p}{\Delta t} = \frac{d_{p_2} - d_{p_1}}{t_2 - t_1} \quad (3)$$

where  $d_{p_1}$  and  $d_{p_2}$  are the representative particle diameters at times  $t_1$  and  $t_2$ , respectively.

Condensation sink (CS) describes the condensing vapor sink caused by the particle population (Kulmala et al., 2012),

$$15 \text{ CS} = 4\pi D \int_0^{d_p \text{ max}} \beta_{m,d_p} d_p N_{d_p} dd_p = 4\pi D \sum_{d_p} \beta_{m,d_p} d_p N_{d_p} \quad (4)$$

where  $D$  is the diffusion coefficient of the condensing vapor, usually assumed to be sulfuric acid ( $0.104 \text{ cm}^2 \text{ s}^{-1}$  used in this study); and  $\beta_{m,d_p}$  is the transitional regime correction factor.

## 20 2.2.2 Sulfuric acid

Sulfuric acid has been accepted as a key gas-phase precursor for atmospheric nucleation and contributes to the subsequent growth of newly-formed particles (Weber

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et al., 1996; Sipila et al., 2010). The accurate measurement of gas-phase sulfuric acid requires application of chemical ionization mass spectrometry using nitrates as reagent ions (Eisele and Tanner, 1993), which is not possessed by this research group during this study. Instead, the sulfuric acid proxy  $[H_2SO_4]$  was estimated based on local solar radiation level Radiation,  $SO_2$  concentration  $[SO_2]$ , condensation sink CS, and relatively humidity (Mikkonen et al., 2011),

$$[H_2SO_4] = 8.21 \times 10^{-3} \cdot k \cdot \text{Radiation} \cdot [SO_2]^{0.62} \cdot (\text{CS} \cdot \text{RH})^{-0.13} \quad (5)$$

where  $k$  is the temperature dependent reaction rate constant (Mikkonen et al., 2011).

Condensation of sulfuric acid contributes to the growth of newly-formed particles. The growth of clusters/particles due to condensation of sulfuric acid,  $GR_{H_2SO_4}$ , can be estimated by the following equations (Nieminen et al., 2010),

$$GR_{H_2SO_4} = \frac{\gamma}{2\rho_v} \left(1 + \frac{d_v}{d_p}\right)^2 \left(\frac{8kT}{\pi}\right)^{1/2} \left(\frac{1}{m_p} + \frac{1}{m_v}\right)^{1/2} m_v [H_2SO_4] \quad (6)$$

and

$$\gamma = \frac{4}{3} \cdot Kn \cdot \beta_{m,d_p} \quad (7)$$

where  $\rho_v$  and  $d_v$  are the condensed phase density and molecule diameter of  $H_2SO_4$ , respectively;  $m_p$  and  $m_v$  are particle and  $H_2SO_4$  vapor molecule masses, respectively;  $Kn$  is the Knudsen number (Lehtinen and Kulmala, 2003). For this calculation,  $\rho_p = 1.83 \text{ g cm}^{-3}$ ,  $d_v = 0.56 \text{ nm}$ , and  $m_v = 98 \text{ amu}$  were used.

### 3 Results and discussion

#### 3.1 Classification of new particle formation (NPF) events

Figure 1 presents a contour plot for particle size distributions of 3–615.3 nm and a number concentration plot of sub-3 nm clusters/particles  $N_{1,34\sim 3}$  during 25 November 2013–

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25 January 2014. Data were occasionally missing because of maintenance and minor breakdown of instruments. From the figure, frequent bursts of sub-3 nm particles were evident, with concentrations up to  $8.0 \times 10^4 \text{ cm}^{-3}$  around noon time. However, similarly to previous studies (Kulmala et al., 2007, 2013; Yu et al., 2014), not all sub-3 nm particles eventually underwent a continuous growth to larger sizes. In this study, we define an observation day with appearances of sub-3 nm clusters/particles over a time span of hours and their subsequent growth to larger sizes for a few hours that presents a banana-shaped contour plot of particle size distributions obtained from SMPS (Dal Maso et al., 2005) as a NPF event day, and focus on characteristics and potential mechanisms of these events.

According to the classification, there were 13 event days during the 62-day campaign, as illustrated by the shadow in Fig. 1. Although nCNC data were partially unavailable on 26 December 2013 and completely unavailable on 24 January 2014, these two days are still defined as NPF days since a distinctive banana-shaped contour plot for particle distributions between 3–615.3 nm existed. 18 December 2013 was not regarded as a NPF day since  $N_{1.34\sim 3}$  was not recorded and the growth of 3 ~ 20 nm particles was relatively short-lived.

Among these NPF events, 5 NPF events occurred in November, 3 in December, and 5 in January. The averaged frequency for NPF events was 21.0% during the 62-day campaign. Our NPF frequency at Shanghai is larger than the average value in winter 1996–2003, SMEAR II station, Hyytiälä, Finland (Dal Maso et al., 2005), likely because nucleation events at Hyytiälä have recently been related to oxidation products of biogenic emissions (Kulmala et al., 1998; Schobesberger et al., 2013; Riccobono et al., 2014) and photochemistry of volatile organic compounds is less intensive in winter months. This frequency is also higher than that in winter, semi-rural Kent, OH, during which transport of sulfur dioxide from the east-southeast power plant to Kent is not favored (Kanawade et al., 2012). NPF events occurred at a frequency of around 40% during November–December 2004 in Beijing (Wu et al., 2007), much more often than in Shanghai. On the other hand, NPF frequency in Shanghai is remarkably close

to that measured in winter 2012, Nanjing, which is also located in Yangtze River delta (Herrmann et al., 2014).

Number concentrations of particles in different size ranges, i.e.,  $N_{1.34\sim3}$ ,  $N_{3\sim7}$ , and  $N_{7\sim30}$ , respectively, on a NPF day (11 December 2013) and an obvious non-NPF day (7 January 2014) are further examined to illustrate features of a NPF event, as shown in Fig. 2. On the NPF day, 1.34–3 nm particles appeared as early as 7 a.m. in the morning that was right after sunrise (6:42 a.m. on 11 December 2013), reached its maximum just before noontime, and spanned for almost the whole daytime (sunset at 4:52 p.m., 11 December 2013). This size distribution of atmospheric neutral and charged clusters/particles by a scanning PSM is identical to that measured at Hyytiälä, Finland (Kulmala et al., 2013), suggesting that photochemistry products likely contribute to formation of smallest particles. On the same NPF day, 3–7 nm and 7–30 nm particles appeared much later, resembling previous findings only with SMPS measurements (e.g., Yue et al., 2010). The lag in peaking times of  $N_{1.34\sim3}$ ,  $N_{3\sim7}$ , and  $N_{7\sim30}$  on the NPF day clearly indicated the continuous growth of clusters/particles, the loss of particles due to coagulation during the growth, and the diverse sources of particles in the size range of 7–30 nm. In contrast,  $N_{1.34\sim3}$  and  $N_{3\sim7}$  showed a flat profile on the non-NPF day. The minor enhancement in  $N_{7\sim30}$  between 10 a.m. and 5 p.m. on the non-NPF day was not due to growth of newly formed particles. Instead, direct emission of 7–30 nm particles from transportation activity likely explained their appearance.

### 3.2 General characteristics of NPF events

Table 1 summaries characteristics of each NPF event observed in this campaign, including  $J_{1.34}$ ,  $J_3$ ,  $GR_{1.35\sim2.39}$  (from the bin of 1.34–1.37 nm to the bin of 1.89–3.0 nm),  $GR_{2.39\sim7}$ ,  $GR_{7\sim20}$ , **CS**,  $[H_2SO_4]$ ,  $N_{1.34\sim3}$ , and total number concentrations of particles  $N_{1.34\sim615}$ , and compares the mean values to those in selected other studies. Nucleation rate  $J_{1.34}$  and formation rate of 3 nm particles  $J_3$  were 112.4–271.0 and 2.3–19.2  $cm^{-3} s^{-1}$ , respectively. Nucleation rate  $J_{1.34}$  at Shanghai is obviously significantly

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larger than  $1.4 \text{ cm}^{-3} \text{ s}^{-1}$  at Hyytiälä, Finland with a pristine atmosphere (Kulmala et al., 2012) and  $1.3 \text{ cm}^{-3} \text{ s}^{-1}$  at Kent, OH with relatively lower levels of pollutants (Yu et al., 2014). Direct comparison of our nucleation rate with that in a Chinese location is not feasible because no previous reports are available. However, Herrmann et al. (2014) reported a  $J_2$  of  $33.2 \text{ cm}^{-3} \text{ s}^{-1}$  at the SORPES-NJU station, Nanjing China. Together with their results, we conclude that strong nucleation events occur geographically widely in the YRD region.

Formation rate of 3 nm particles  $J_3$  has been more routinely reported. Similarly to  $J_{1.34}$ ,  $J_3$  at Shanghai is significantly larger than  $0.61 \text{ cm}^{-3} \text{ s}^{-1}$  at Hyytiälä, Finland (Kulmala et al., 2012), and generally comparable to  $3.3\text{--}81.4$  and  $1.1\text{--}22.4 \text{ cm}^{-3} \text{ s}^{-1}$  at Beijing (Wu et al., 2007; Yue et al., 2009),  $3.6\text{--}6.9 \text{ cm}^{-3} \text{ s}^{-1}$  at Hong Kong (Guo et al., 2012), and  $2.4\text{--}4.0 \text{ cm}^{-3} \text{ s}^{-1}$  in a back-garden rural site of Pearl River Delta (Yue et al., 2013). The fast reduction from  $J_{1.34}$  to  $J_3$  was likely due to the presence of a large background particle number as shown in Table 1.

The large background particle number concentrations corresponded to the high condensation sink (CS of  $0.030\text{--}0.10 \text{ s}^{-1}$ ) observed during the campaign. As shown in Table 1, CS at Shanghai is much larger than  $(0.05\text{--}0.35) \times 10^{-2} \text{ s}^{-1}$  at Hyytiälä, Finland (Kulmala et al., 2012) and  $0.8 \times 10^{-2} \text{ s}^{-1}$  at Kent, OH (Yu et al., 2014), but comparable to  $(0.58\text{--}8.4) \times 10^{-2} \text{ s}^{-1}$  at Beijing (Wu et al., 2007; Yue et al., 2009; Zhang et al., 2011; Gao et al., 2012),  $(1.0\text{--}6.2) \times 10^{-2} \text{ s}^{-1}$  at Hong Kong (Guo et al., 2012),  $2.4 \times 10^{-2} \text{ s}^{-1}$  at Nanjing (Herrmann et al., 2014), and  $(3.5\text{--}4.6) \times 10^{-2} \text{ s}^{-1}$  in an urban site of Pearl River Delta (Yue et al., 2013). High sulfuric acid proxy ( $[\text{H}_2\text{SO}_4]$  of  $(2.3\text{--}6.4) \times 10^7 \text{ molecules cm}^{-3}$ ) existed to promote the NPF events. Measurements of gas-phase sulfuric acid by a chemical ionization mass spectrometer during the CARE-Beijing 2008 Campaign, a time period with strict air quality control regulations, reported peak concentrations of sulfuric acid up to the order of  $10^7 \text{ molecules cm}^{-3}$  (Zheng et al., 2011), smaller than our sulfuric acid proxy. Uncertainty may well exist for our sulfuric acid proxy that was calculated from the concentrations of sulfur dioxide and radiation intensity. However, Judging from CS and  $[\text{H}_2\text{SO}_4]$  together, it is clear that the con-

densable vapor at Shanghai is likely a main impetus for observed strong new particle formation events.

$GR_{1.35\sim 2.39}$ ,  $GR_{2.39\sim 7}$ , and  $GR_{7\sim 20}$  were in the range of 0.49–8.1, 3.1–35.7, 4.5–38.3  $\text{nm h}^{-1}$ , respectively. The arithmetic average values of  $GR_{1.35\sim 2.39}$ ,  $GR_{2.39\sim 7}$ , and  $GR_{7\sim 20}$  were  $2.0 \pm 2.7$  (one SD),  $10.9 \pm 9.8$  and  $11.4 \pm 9.7$   $\text{nm h}^{-1}$ , respectively. A closer examination of growth rates by dividing  $GR_{1.35\sim 2.39}$  into growth of clusters/particles from one bin to another, i.e.,  $GR_{1.35\sim 1.39}$  ( $1.6 \pm 1.0$   $\text{nm h}^{-1}$  from the bin of 1.34–1.37 nm to the bin of 1.37–1.41 nm),  $GR_{1.39\sim 1.46}$  ( $1.4 \pm 2.2$   $\text{nm h}^{-1}$  from 1.37–1.41 to 1.41–1.52 nm),  $GR_{1.46\sim 1.70}$  ( $7.2 \pm 7.1$   $\text{nm h}^{-1}$  from 1.41–1.52 to 1.52–1.89 nm), and  $GR_{1.70\sim 2.39}$  ( $9.0 \pm 11.4$   $\text{nm h}^{-1}$  from 1.52–1.89 to 1.89–3.0 nm), shows a clear size-depend particle growth (Fig. 3), owing to the nano-Köhler activation (Kulmala et al., 2004), Kelvin effect, and surface or volume-controlled reaction corrected for the Kelvin effect on surface or volume concentrations (Kuang et al., 2012). Similar observations have been reported in previous studies using nCNC (Kulmala et al., 2013) and DEG UCPC (Kuang et al., 2012), respectively. Our  $GR_{2.39\sim 7}$  is larger than 6.3  $\text{nm h}^{-1}$  at Nanjing (Herrmann et al., 2014), and our  $GR_{7\sim 20}$  is close to the upper bound of those at urban Beijing (Wu et al., 2007; Yue et al., 2009; Zhang et al., 2011; Gao et al., 2012), and generally larger than 1.5–8.4  $\text{nm h}^{-1}$  in Hong Kong (Guo et al., 2012), indicating that the condensable vapor was intense. In addition, our growth rates suggest that the smallest clusters (the bin of 1.34–1.37 nm), if not scavenged by larger particles, will grow to 3 nm within  $\sim 12$  min, and to 20 nm within  $\sim 2$  h.

### 3.3 Potential mechanisms for NPF events

As shown in Table 1, nucleation rate ( $J_{1.34}$ ) in this study is typically larger than  $100 \text{ cm}^{-3} \text{ s}^{-1}$ , suggesting that the ion-induced nucleation was not a main mechanism for observed fast nucleation (Hirsikko et al., 2011; Riccobono et al., 2014). The 2012 winter study at the SORPES-NJU station that is also located at YRD shows that the ratio of  $J_2$  between ions and total particles (ions plus neutral particles) was 0.002, also

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indicating the minor role of ion-induced nucleation (Herrmann et al., 2014). Hence, it is likely that nucleation of neutral precursor molecules actually largely determined the observed NPF events.

Correlations between  $\log J_{1,34}$  and  $\log [\text{H}_2\text{SO}_4]$  (Fig. 4) and between  $\log J_{1,34}$  and  $\log [\text{NH}_3]$  (Fig. 5) have been examined to elucidate potential mechanisms for our NPF events. Since  $J_{1,34}$  could not be accurately determined on some of the NPF days, the data points in both figures are less than the actual number of NPF events that have been observed. Daily daytime (6 a.m.–6 p.m.) averages of sulfuric acid proxy and ammonia were used as approximations for their effective concentrations on a NPF day. The correlation ( $R^2 = 0.67$ ) between  $\log J_{1,34}$  and  $\log [\text{NH}_3]$  is better than that ( $R^2 = 0.23$ ) between  $\log J_{1,34}$  and  $\log [\text{H}_2\text{SO}_4]$ , and slopes are 0.58 and 0.64, respectively.

Most ambient studies showed that nucleation rate  $J$  is proportional to the first or second power of the concentration of gas-phase sulfuric acid, i.e.,  $J = A \cdot [\text{H}_2\text{SO}_4]^P$ , where  $P = 1$  or  $2$ , and  $A$  is a pre-exponential factor (McMurry et al., 2005; Sihto et al., 2006; Erupe et al., 2010). Our  $P$  of 0.64 might be influenced by the uncertainty during the calculation of sulfuric acid proxy  $[\text{H}_2\text{SO}_4]$ . This  $P$  value is conventionally interpreted as the number of sulfuric acid molecules in the critical nucleus. If rounded to an integral number, i.e.,  $P = 1$ , our linear correlation between  $\log J_{1,34}$  and  $\log [\text{H}_2\text{SO}_4]$  suggests that nucleation occurs after activation of clusters containing one molecule of sulfuric acid, with subsequent growth involving other species (Kulmala et al., 2006). However, Kupiainen-Määttä et al. (2014) recently reported that the number of molecules in a critical cluster cannot be determined by a slope analysis in atmospherically relevant applications, underscoring the need to further explore the exact nucleation mechanism. Herrmann et al. (2014) also calculated the sulfuric acid proxy, related it to observed nucleation rates, and speculated that the sulfuric acid exponent might be well over 2, which underscores the reliability of calculation methods in a Chinese location. Hence, our preliminary result should be further tested with actual measurements of gas-phase sulfuric acid, but does indicate the key role of sulfuric acid in NPF events. On the other hand, linear correlation between  $\log J$  and  $\log [\text{NH}_3]$  was observed in a previous nu-

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5 cleation study in Atlanta, GA, with a slope of 1.17 (McMurry et al., 2005), but a clear relationship was not perceived at Kent, OH (Erupe et al., 2010). This discrepancy may come from the level of ammonia that has been predicted to have a saturation threshold (Napari et al., 2002) and/or the co-occurring sulfuric acid concentration (Benson et al., 2009). Nevertheless, our correlation between  $\log J$  and  $\log[\text{NH}_3]$  suggests that ammonia also participated in the nucleation.

10 The subsequent growth of newly formed particles can be partially attributed to condensation of sulfuric acid. The theoretical maximum growth rate of 1.34–3 nm clusters/particles due to condensation of sulfuric acid ( $\text{GR}_{\text{H}_2\text{SO}_4(1.34\sim 3)}$ ), calculated according to Eqs. (6) and (7), was  $2.5 \pm 0.82 \text{ nm h}^{-1}$  on average. This rate is larger than the observed growth rates of clusters/particles from the bin of 1.34–1.37 nm to the bin of 1.89–3.0 nm ( $\text{GR}_{1.35\sim 2.39}$ ), being  $2.0 \pm 2.7 \text{ nm h}^{-1}$ , indicating that sulfuric acid proxy was enough to explain the observed growth for particles under 3 nm. Similar calculation of  $\text{GR}_{\text{H}_2\text{SO}_4(3\sim 7)}$  and  $\text{GR}_{\text{H}_2\text{SO}_4(7\sim 20)}$  yielded  $1.9 \pm 0.62$  and  $1.6 \pm 0.54 \text{ nm h}^{-1}$ , respectively. In Fig. 6, relative contributions of sulfuric acid to growth of particles in the range of 3–7 and 7–20 nm, respectively, on each NPF day is presented. Since 7 nm particles reached their maximum earlier than 3 nm particles on 9 and 15 January 2014, there was no calculated  $\text{GR}_{3\sim 7}$  and hence no ratios available on these two days. In addition, condensation of sulfuric acid was more prominent for 3–7 nm particles on 6 NPF days (25, 26, 30 November, 10, 11, and 12 December 2013), whereas it was more significant for 7–20 nm particles on the other 5 NPF days (28, 29 November 2013, 13, 21, and 24 January 2014). On average, condensation of gas-phase sulfuric acid explained 26.5 % of  $\text{GR}_{2.39\sim 7}$ , and 20.9 % of  $\text{GR}_{7\sim 20}$ , respectively. The rest of growth might be largely attributed to condensation of extremely low volatility organic compounds (Ehn et al., 2014), and potentially heterogeneous reactions of organics on nanoparticle surface (Wang et al., 2010, 2011).



### 3.4 NPF and background PM<sub>2.5</sub>

NPF events in urban environment are of special interests since the pre-existing particle surface may significantly scavenge the newly formed particles and change the probability of NPF. We plot number concentrations of 1.34–10 nm particles ( $N_{1.34\sim 10}$ ), sulfuric acid proxy ( $[H_2SO_4]$ ), ammonia ( $[NH_3]$ ), and mass concentrations of PM<sub>2.5</sub> with shadowed blocks representing NPF events in Fig. 7. Note that  $N_{1.34\sim 10}$  was used as an approximation for nucleation and subsequent growth while excluding particles from direct emission. The average daytime (6 a.m.–6 p.m.)  $N_{1.34\sim 10}$  on NPF days was  $(2.7 \pm 2.1) \times 10^4 \text{ cm}^{-3}$ , much larger than  $(1.5 \pm 1.0) \times 10^4 \text{ cm}^{-3}$  on the rest days of the campaign, indicating that a stronger input of particles from nucleation processes on NPF days. However, daytime  $[H_2SO_4]$  did not show an apparent difference between on NPF days ( $(3.7 \pm 1.2) \times 10^7 \text{ molecules cm}^{-3}$ ) and on the rest days ( $(3.9 \pm 2.5) \times 10^7 \text{ molecules cm}^{-3}$ ). For example, an episode with high daily sulfuric acid proxy during 19 December 2013 and 16 January 2014 did not lead to any NPF events. Instead, observed NPF events occurred on days with low PM<sub>2.5</sub> levels and moderate  $[H_2SO_4]$ . During our campaign, NPF days were characterized with low concentrations of PM<sub>2.5</sub> ( $62.4 \pm 38.8 \mu\text{g m}^{-3}$ ), whereas the average was  $110.9 \pm 94.2 \mu\text{g m}^{-3}$  on the rest days of the campaign. Ammonia varied dramatically, even within a single day. NPF events occurred on days with around 10-fold difference in ammonia concentrations. According to ammonia's profile and its positive correlation with  $J_{1.34}$ , we speculate that ammonia was involved in nucleation but it is not the driving force.

Examination of these parameters from 12 to 14 January 2014 was performed since the three days were characterized with similar meteorological conditions. The average daytime concentration of sulfuric acid proxy was 2.8, 2.3, and  $1.0 \times 10^7 \text{ molecules cm}^{-3}$  on 12, 13 and 14 January 2014, respectively. No NPF event was observed on 14 January at least partially because of the low sulfuric acid proxy. Appearance of a NPF event on 13 January and non-appearance on 12 January could be explained by the high PM<sub>2.5</sub> concentration on 12 January, with maximum mass concentration up to

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214.8  $\mu\text{g m}^{-3}$ , although similar sulfuric acid proxies existed between the two days. Hence, we conclude that, qualitatively, NPF processes in urban Shanghai occurred with low levels of  $\text{PM}_{2.5}$  and that high sulfuric acid favored NPF events when  $\text{PM}_{2.5}$  was low. This conclusion is identical to that drawn from a Mexico City study where NPF events observed in the city correlated with elevated concentrations of sulfur dioxide and low particulate matter mass concentrations in the afternoon hours (Dunn et al., 2004).

Concentrations of  $\text{PM}_{2.5}$  showed correlations with particle growth rates in our study. Figure 8 denotes particle growth rates of particles in various size ranges and the real-time hourly-averaged mass concentrations of  $\text{PM}_{2.5}$  when particles grew to the corresponding size.  $\text{GR}_{1.35\sim 2.39}$  and  $\text{GR}_{2.39\sim 7}$  were anti-correlated with  $\text{PM}_{2.5}$  as expected, since preexisting particles coagulate with smallest new clusters/particles and scavenge them.  $\text{GR}_{7\sim 20}$  was positively correlated with  $\text{PM}_{2.5}$ . Although this positive correlation is not exactly understood, we speculate that high  $\text{PM}_{2.5}$  concentrations in Shanghai were at least partially associated with secondary processes in the atmosphere leading to low vapor pressure organic and inorganic species, which could increase condensation rate or react with larger particles therefore increasing PM.

#### 4 Summary and conclusions

Atmospheric new particle formation is a significant source of atmospheric aerosol particles. Understanding NPF under the current levels of air pollution in China is of special scientific interests because the exact nucleation mechanism remains elusive. From 25 November 2013 to 25 January 2014, a combination of one nano-CNC, one nano-SPMS, and one long-SPMS has been utilized to investigate atmospheric nucleation by measuring particle size distributions in the range of 1.34–615.3 nm at urban Shanghai, located in the east Yangtze River Delta. During this 62-day campaign, 13 NPF events were identified with strong burst of sub-3 nm particles and subsequent fast growth of newly formed particles. Together with nucleation frequency (21 %), the obtained nucleation rate  $J_{1.34}$  (112.4–271.0  $\text{cm}^{-3} \text{s}^{-1}$ ), condensation sink CS (0.030–



first measurement of sub-3 nm particles in urban Shanghai, and provides some of the preliminary mechanisms for NPF events in China.

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**Table 1.** Nucleation rate ( $J_{1,34}$ ), formation rate of 3 nm particles ( $J_3$ ), particle growth rates ( $GR_{1.35\sim 2.39}$ ,  $GR_{2.39\sim 7}$ , and  $GR_{7\sim 20}$ ), condensation sink (CS), sulfuric acid proxy ( $[H_2SO_4]$ ), number concentrations of 1.34–3 nm clusters/particles ( $N_{1,34\sim 3}$ ), and total number concentrations of particles  $N_{1,34\sim 615}$ , of each NPF event during this campaign.

Date	$J_{1,34}$ ( $cm^{-3} s^{-1}$ )	$J_3$ ( $cm^{-3} s^{-1}$ )	$GR_{1.35\sim 2.39}$ ( $nm h^{-1}$ )	$GR_{2.39\sim 7}$ ( $nm h^{-1}$ )	$GR_{7\sim 20}$ ( $nm h^{-1}$ )	CS <sup>a</sup> ( $10^{-2} s^{-1}$ )	$[H_2SO_4]$ <sup>a</sup> ( $10^7 cm^{-3}$ )	$N_{1,34\sim 3}$ <sup>f</sup> ( $10^4 cm^{-3}$ )	$N_{1,34\sim 615}$ <sup>f</sup> ( $10^4 cm^{-3}$ )	Ref.
25 Nov 2013	n.a. <sup>a</sup>	10.6	n.a.	12.4 <sup>b</sup>	38.3	4.7	2.7	3.3 <sup>g</sup>	6.3	this study
26 Nov 2013	n.a.	2.3	n.a.	0.32 <sup>b</sup>	n.a.	5.9	2.6	n.a.	n.a.	this study
28 Nov 2013	185.1	13.4	0.94	35.7	4.6	5.7	3.6	1.6	4.2	this study
29 Nov 2013	271.0	3.9	1.7	10.6	4.5	6.3	4.3	2.1	4.5	this study
30 Nov 2013	n.a.	n.a.	0.82	3.4	10.2	n.a.	3.1	1.5	n.a.	this study
10 Dec 2013	268.4	10.0	0.49	18.6	21.0	9.9	5.5	1.4	4.6	this study
11 Dec 2013	219.0	19.2	n.a.	5.1	9.6	10.2	6.4	1.1	4.5	this study
12 Dec 2013	190.3	7.6	n.a.	3.1	12.3	8.8	4.5	1.1	4.1	this study
9 Jan 2014	136.2	n.a.	8.1	n.a.	9.5	3.7	2.3	1.6	3.8	this study
13 Jan 2014	n.a.	2.7	n.a.	6.3	1.9	3.0	2.3	1.5	3.4	this study
15 Jan 2014	121.9	n.a.	0.56	n.a.	9.7	4.2	4.1	1.5	4.3	this study
21 Jan 2014	112.4	9.2	1.5	11.9	7.5	4.9	3.7	1.1	3.9	this study
24 Jan 2014	n.a.	8.1	n.a.	12.2 <sup>b</sup>	7.8	4.7	3.4	1.7	4.2	this study
Mean	188.0	8.7	2.0	10.9	11.4	6.0	3.7	1.5	4.4	this study
Finland	1.4	0.61	1.4	3.9	4.9	0.05–0.35				Kulmala et al. (2012)
USA <sup>h</sup>	1.3					0.8		0.9		Yu et al. (2014)
Beijing					1.2–8.0 <sup>c</sup>	2.4–3.6				Gao et al. (2012)
Beijing		3.3–81.4			0.1–11.2 <sup>c</sup>	0.58–4.3				Wu et al. (2007)
Beijing					2.7–13.9 <sup>d</sup>	0.6–8.4				Zhang et al. (2011)
Beijing		1.1–22.4			1.2–5.6 <sup>e</sup>	0.9–5.3				Yue et al. (2009)
Nanjing	33.2 <sup>j</sup>	1.1 <sup>l</sup>		6.3	8 <sup>d</sup>	2.4				Herrmann et al. (2014)
Hong Kong		3.6–6.9			1.5–8.4 <sup>c</sup>	1.0–6.2				Guo et al. (2012)
Pearl River Delta		2.4 ~ 4.0 (rural)			4.0 ~ 22.7 (rural)	2.3 ~ 3.3 (rural)				Yue et al. (2013)
					10.1 ~ 18.9 (urban)	3.5 ~ 4.6 (urban)				

<sup>a</sup> Data were not available or could not be accurately determined.

<sup>b</sup> Result were calculated from nano-SMPS data.

<sup>c</sup> Shown here is  $GR_{3\sim 30}$ .

<sup>d</sup> Shown here is  $GR_{7\sim 30}$ .

<sup>e</sup> Daytime average (from 6.00 a.m. to 6.00 p.m.).

<sup>f</sup> 24 h-average.

<sup>g</sup> Average values between 10 a.m. and 4 p.m.

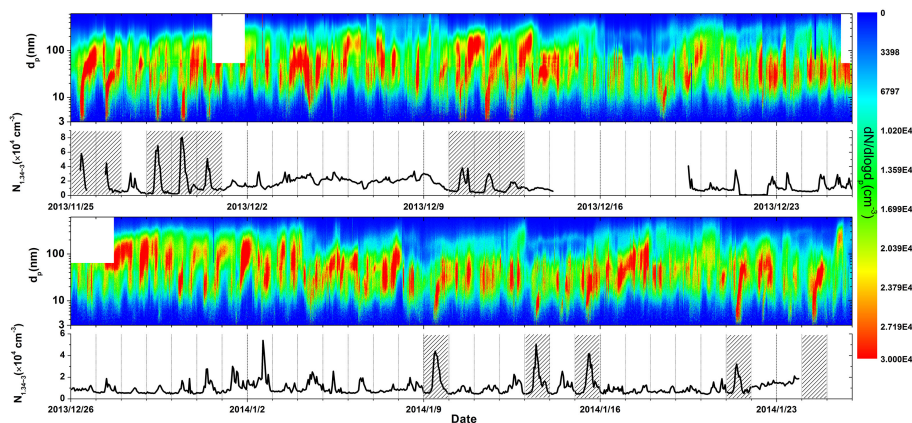
<sup>h</sup> Shown here are data in Kent, OH.

<sup>i</sup> Shown here is  $J_3$ .

<sup>j</sup> Shown here is  $J_6$ .

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**Figure 1.** Contour plot for particle size distributions of 3–615.3 nm and plot of number concentrations of sub-3 nm clusters/particles ( $N_{1.34\sim 3}$ ) during 25 November 2013–25 January 2014. Data were occasionally missing because of the maintenance and minor breakdown of instruments. NPF events are illustrated with shadows.

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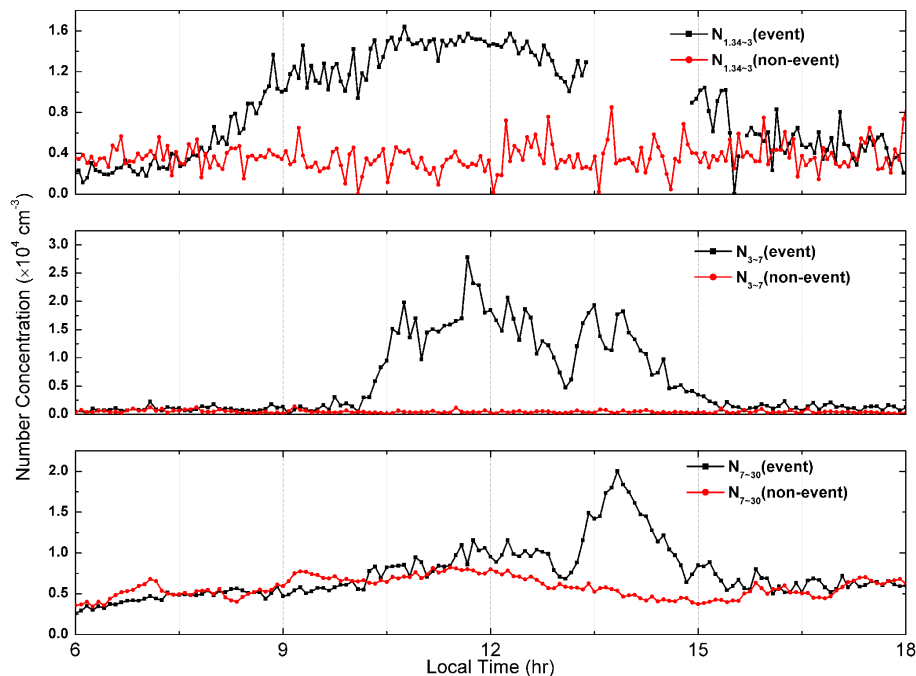
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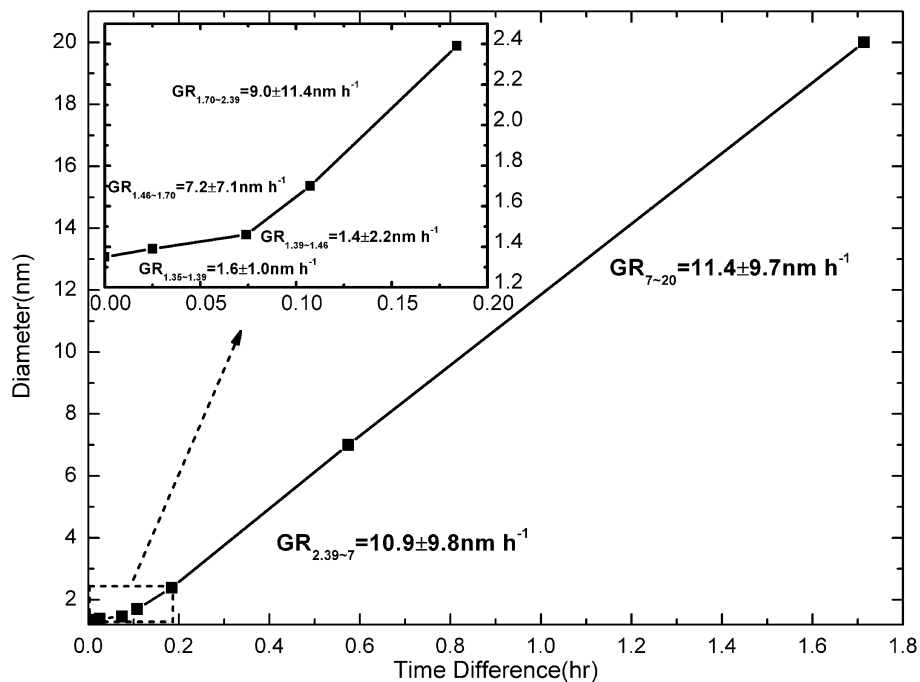
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**Figure 2.** Profiles of  $N_{1,34\sim 3}$ ,  $N_{3\sim 7}$ , and  $N_{7\sim 30}$  from 6 a.m. to 6 p.m. on a NPF day (11 December 2013) and a non-NPF day (7 January 2014), respectively.

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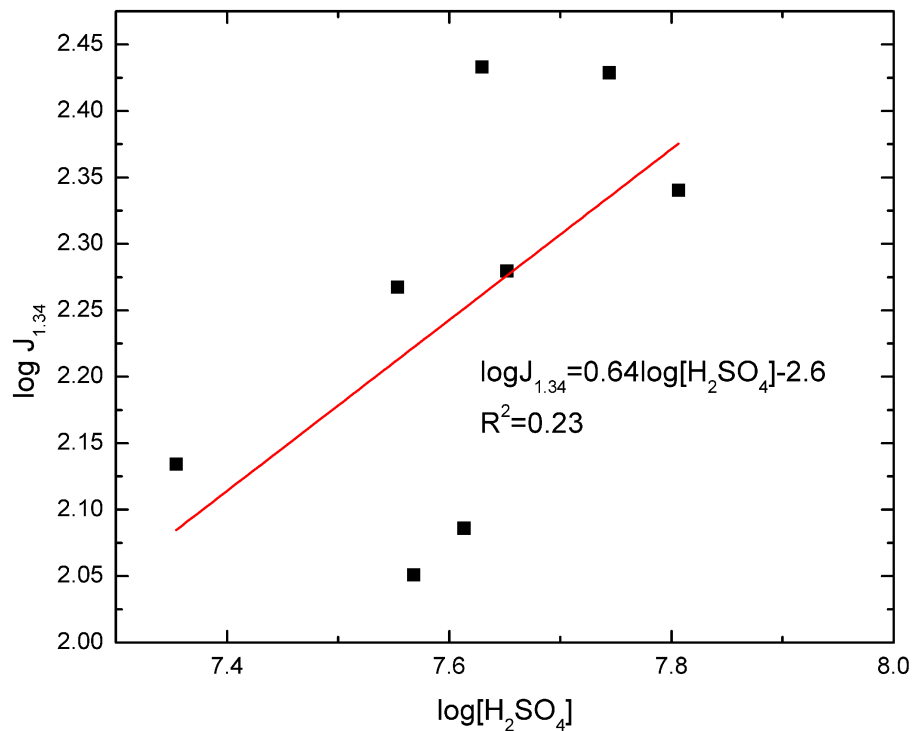
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**Figure 3.** Averaged particle size evolution on NPF days. Arithmetic mean of particle growth rates are given with one SD.

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**Figure 4.** Correlation between  $\log J_{1.34}$  and  $\log[\text{H}_2\text{SO}_4]$ . Daily daytime average of sulfuric acid proxy was used as an approximation for its effective concentration on a NPF day.

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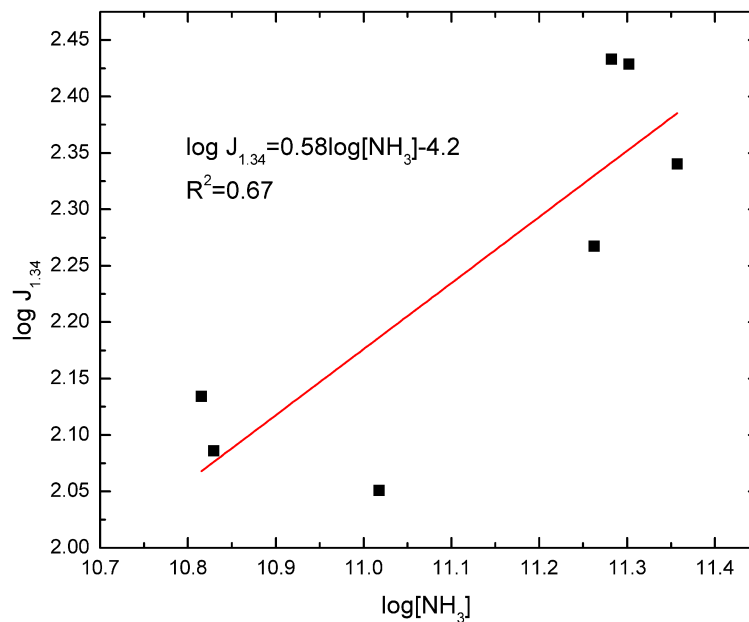
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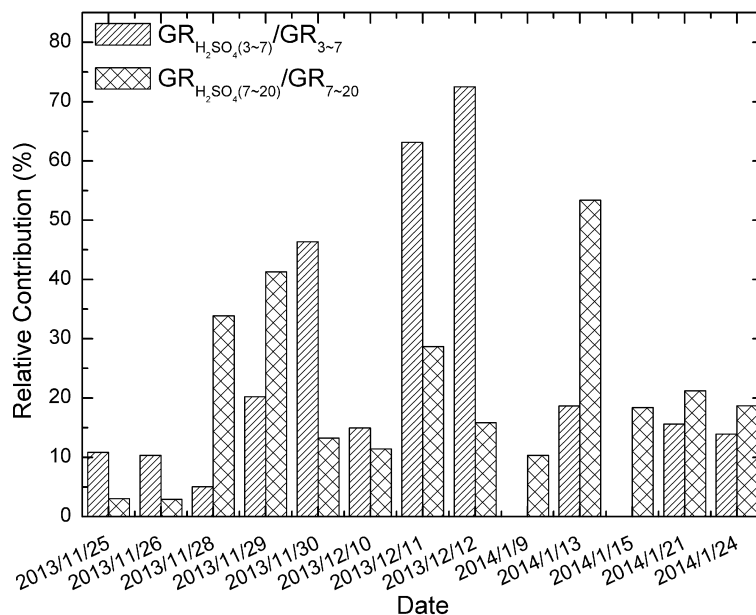


**Figure 5.** Correlation between  $\log J_{1,34}$  and  $\log[\text{NH}_3]$ . Daily daytime average of ammonia was used as an approximation for its effective concentration on a NPF day.

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**Figure 6.** Relative contribution of sulfuric acid to growth of particles in the range of 3–7 and 7–20 nm, respectively, on each NPF day.

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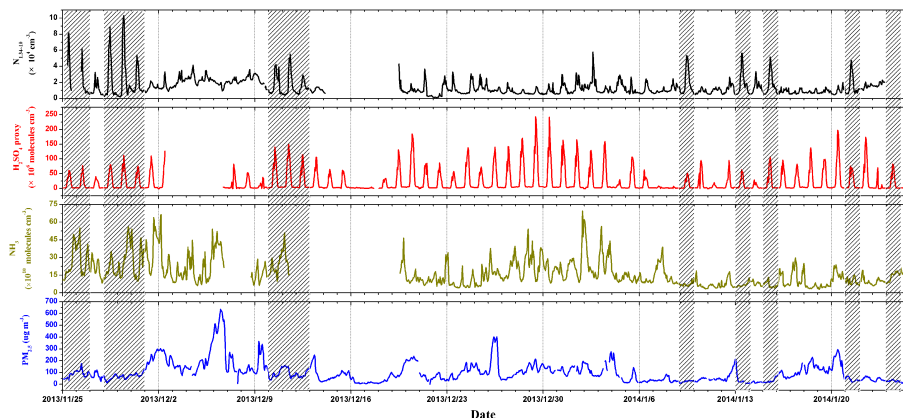
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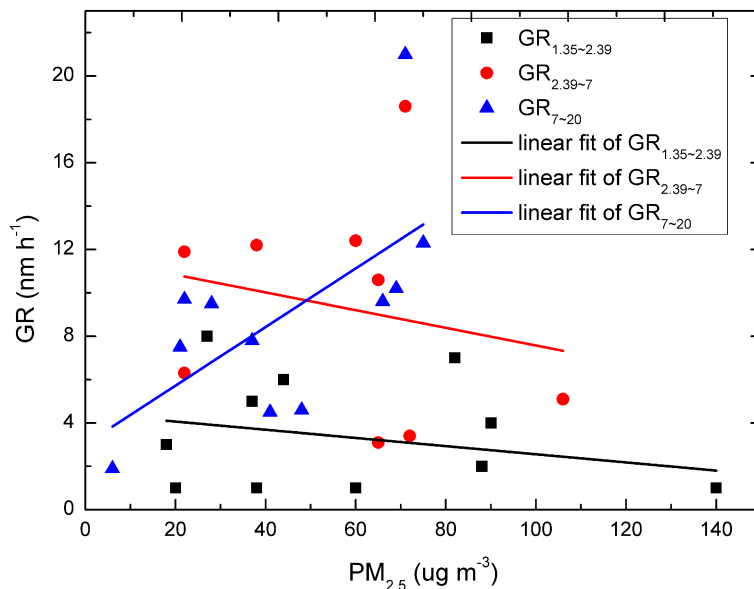


**Figure 7.** Number concentrations of 1.34–10 nm particles ( $N_{1.34\sim 10}$ ), sulfuric acid proxy ( $\text{[H}_2\text{SO}_4\text{]}$ ), concentrations of ammonia, and mass concentrations of  $\text{PM}_{2.5}$  during the campaign. NPF events are illustrated with shadows.

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**Figure 8.** Correlations between particle growth rates and real-time hourly-averaged mass concentrations of  $PM_{2.5}$  when particles grew to the corresponding size.

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