

1 Aged boreal biomass burning aerosol size distributions from 2 BORTAS 2011

3

4 K.M. Sakamoto^{1,2}, J. D. Allan³, H. Coe³, J. W. Taylor³, T. J. Duck², J. R. Pierce^{1,2}

5

6 [1] Colorado State University, Fort Collins, CO, USA

7 [2] Department of Physics and Atmospheric Science, Dalhousie University, Halifax, NS, CAN

8 [3] University of Manchester, Manchester, UK

9

10 Corresponding author: J.R. Pierce (jeffrey.pierce@colostate.edu)

11

12 Abstract

13

14 Biomass-burning aerosols contribute to aerosol radiative forcing on the climate system. The magnitude
15 of this effect is partially determined by aerosol size distributions, which are functions of source fire
16 characteristics (e.g. fuel type, MCE) and in-plume microphysical processing. The uncertainties in
17 biomass-burning emission number size-distributions in climate model inventories lead to uncertainties
18 in the CCN concentrations and forcing estimates derived from these models.

19 The BORTAS-B measurement campaign was designed to sample boreal biomass-burning
20 outflow over Eastern Canada in the summer of 2011. Using these BORTAS-B data, we implement
21 plume criteria to isolate the characteristic size-distribution of aged biomass-burning emissions (aged ~
22 1 – 2 days) from boreal wildfires in Northwestern Ontario. The composite median size-distribution
23 yields a single dominant accumulation mode with $D_{pm} = 230$ nm (number-median diameter) and $\sigma =$
24 1.5, which are comparable to literature values of other aged plumes of a similar type. The organic
25 aerosol enhancement ratios ($\Delta OA/\Delta CO$) along the path of Flight b622 show values of 0.09-0.17 $\mu\text{g m}^{-3}$
26 ppbv^{-1} with no significant trend with distance from the source. This lack of enhancement ratio
27 increase/decrease with distance suggests no detectable net OA production/evaporation within the aged
28 plume over the sampling period (plume age: 1-2 days), though does not preclude OA production/loss at
29 earlier stages.

30 A Lagrangian microphysical model was used to determine an estimate of the freshly emitted

31 size distribution corresponding to the BORTAS-B aged size-distributions. The model was restricted to
32 coagulation and dilution processes based on the insignificant net OA production/evaporation derived
33 from the $\Delta\text{OA}/\Delta\text{CO}$ enhancement ratios. We estimate that the young-plume median diameter was in the
34 range of 59-94 nm with modal widths in the range of 1.7-2.8 (the ranges are due to uncertainty in the
35 entrainment rate). Thus, the size of the freshly emitted particles is relatively unconstrained due to the
36 uncertainties in the plume dilution rates.

37

38 **1. Introduction**

39 **1.1 Biomass burning particles**

40 Biomass burning is a significant emission source of carbonaceous aerosols to the global atmosphere
41 (Andreae and Merlet, 2001; Reid et al., 2005). In addition to releasing high levels of greenhouse gases
42 (CO_2 , CO) and volatile organic compounds, biomass burning releases smoke particles that have climate
43 impacts through the direct and indirect aerosol effects. These particles are primarily composed of a
44 mixture of black carbon and organic carbon, with inorganics contributing some mass (Capes et al.,
45 2008; Carrico et al., 2010; Cubison et al., 2011; Hecobian et al., 2011; Hennigan et al., 2011; Hudson et
46 al., 2004; Reid et al., 2005). These particles directly affect the Earth's radiation balance and climate by
47 scattering and absorbing incoming solar radiation (Haywood, 2000; Jacobson, 2001). Biomass
48 burning particles may also act as cloud condensation nuclei (CCN) and affect climate and radiation
49 through modifying cloud albedo and lifetime (Pierce et al., 2007; Spracklen et al., 2011) (indirect
50 aerosol effects). Globally, the direct and indirect climate effects represent the largest uncertainties in
51 radiative forcing as quantified by the recent IPCC report (Myhre et al., 2013), and biomass burning
52 emissions represent significant contributions to each of the effects globally (Alonso-Blanco, 2014; Lee
53 et al., 2013).

54 The size of biomass-burning particles (and all particles in general) can have large impacts on
55 the magnitude of these direct and indirect effects (Lee et al., 2013; Seinfeld and Pandis, 2006;
56 Spracklen et al., 2011). Regarding the direct effect, the mass-scattering and mass-absorption
57 efficiencies (the amount of scattering and absorption per mass of aerosol particles) depend on the size
58 of the particles, so errors in the predicted/assumed values of biomass-burning particle size may lead to
59 errors in simulated direct aerosol climate effects (Seinfeld and Pandis, 2006). Regarding the indirect
60 effect, particles that are larger in diameter and more hygroscopic are more likely to act as CCN (Petters
61 and Kreidenweis, 2007). Typically particles larger than 30-100 nm act as CCN depending on

62 conditions and hygroscopicity (Petters and Kreidenweis, 2007; Petters et al., 2009), though this range
63 may be slightly larger or smaller for fresh biomass-burning particles due to these particles being
64 initially more hydrophobic/hydrophilic (depending on fuel type) than typical ambient aerosol (Carrico
65 et al., 2010; Engelhart et al., 2012; Petters and Kreidenweis, 2007). Furthermore, for constant
66 emissions mass, a factor-of-2 change in diameter leads to a factor-of-8 change in number emissions,
67 which may contribute to significant changes in CCN concentrations (Pierce et al., 2007; Spracklen et
68 al., 2011). Thus, it is important to provide accurate emissions sizes from biomass burning sources to
69 atmospheric aerosol models looking at aerosol-climate interactions. Lee et al. (2013) found that
70 uncertainties in biomass-burning aerosol emission diameter were responsible for large uncertainties in
71 CCN concentrations in the GLOMAP model (third largest CCN sensitivity out of 28 globally).

72 Atmospheric processing causes the physical and chemical properties of biomass-burning (BB)
73 aerosol evolve over time. These processes have an effect on the size and composition of the particles,
74 and thus influence their direct and indirect effects. Coagulation is a driving factor in size-distribution
75 evolution due to the high concentrations of particles within plumes (Andreae and Merlet, 2001; Capes
76 et al., 2008). Production of secondary organic aerosol (SOA) in-plume has been observed in chamber
77 studies (Cubison et al., 2011; Grieshop et al., 2009; Hennigan et al., 2011; Heringa et al., 2011; Ortega
78 et al., 2013) and in the field (DeCarlo et al., 2010; Lee et al., 2008; Reid et al., 1998; Yokelson et al.,
79 2009), and this SOA will condense onto the particles growing them to larger sizes. In addition, the
80 primary organic aerosol (POA) emitted by the fires may evaporate during the dilution of the plume
81 (Huffman et al., 2009; May et al., 2013). Finally, new particle formation in smoke plumes has been
82 observed in smog chamber studies (Hennigan et al., 2012) as well as in the field (Andreae et al., 2001;
83 Hobbs et al., 2003; Rissler et al., 2006).

84 In global and regional modeling of biomass-burning aerosols, mass-based biomass-burning
85 inventories are the standard, and are generally not accompanied by size data (Reid et al., 2009; van der
86 Werf et al., 2010; Wiedinmyer et al., 2011), leaving size-distribution estimates to the individual
87 investigator. Current global and regional atmospheric aerosol models have gridbox spatial scales
88 (10s-100s of kms) much larger than many initial biomass-burning plume widths (<10 km). This means
89 that sub-grid aging of aerosol plumes by microphysical processes (coagulation,
90 condensation/evaporation and nucleation) will lead to changes in the size distribution that the models
91 cannot explicitly resolve. Therefore, the biomass-burning emissions size distributions must be aged
92 distributions that already account for sub-grid processes. Quantifying the natural variations in

93 biomass-burning aerosols are therefore necessary for accurate predictions. Previous studies of field and
94 lab experiments show biomass burning size-distributions vary according to plume age, combustion
95 phase, and fuel type (Adler et al., 2011; Capes et al., 2008; Hobbs et al., 2003; Hosseini et al., 2010;
96 Janhäll et al., 2010; Okoshi et al., 2014). A review of observed size distribution data by Janhäll et al.
97 (2010) shows the differences in modal width and median diameter as a function of fuel type (forest,
98 savannah, grass), modified combustion efficiency, and plume age (fresh versus aged). Combustion
99 chamber experiments in the FLAME lab have demonstrated similar fuel-type differences in fresh BB
100 size-distributions (Levin et al., 2010).

101 Due to the combination of emission and atmospheric processing factors contributing to the
102 evolution of the BB aerosol size-distribution, characterization of observed, aged BB aerosol is valuable.
103 Adding to the database of observations helps constrain the uncertainties associated with aerosol size.
104 Thus, to improve biomass-burning-aerosol/climate interactions in models, there is a need to
105 characterize the size of particles in aging and aged biomass-burning plumes for a range of fire types
106 and atmospheric conditions (Bauer et al., 2010; Chen et al., 2010; Lee et al., 2013; Pierce et al., 2007;
107 Reddington et al., 2011; Spracklen et al., 2011). In this paper, we specifically investigate the size
108 distributions measured in aged plumes (1-2 days) of large boreal forest fires over Canada.

109 In this paper, we analyze size-distribution and organic aerosol data from BORTAS-B flights that
110 sampled highly concentrated smoke plumes over Eastern Canada on July 20-21st, 2011. A brief
111 overview of the BORTAS-B campaign, instrumentation, and source fire conditions are provided in
112 Section 2.1-2.2. A description of the quantitative plume criteria used to determine plume (versus out of
113 plume) sampling periods is found in Section 2.3. In addition to observational data, we use an
114 aerosol-microphysics box model to simulate the microphysical evolution of number size-distributions.
115 This model was employed to estimate the likely young-plume size distribution associated with the
116 source fires sampled by BORTAS-B. A full model description is provided in Section 2.4. We present
117 the BORTAS-B research flight results in Section 3, which include the measured aged size distributions,
118 evidence for/against net OA production, and the aging simulations. Finally, we provide conclusions in
119 Section 4.

120

121 **2. Methods**

122 **2.1 BORTAS overview**

123 The *Quantifying the impact of BOREal forest fires on the Tropospheric oxidants over the Atlantic using*

124 *Aircraft and Satellites* (BORTAS-B) measurement campaign was held in Eastern/Atlantic Canada from
125 July 11 – August 3, 2011 (Palmer, 2013). The goal was to characterize pyrogenic outflow from boreal
126 forest wildfires using a variety of sampling and observational techniques with emphasis on plume
127 photochemical evolution.

128 BORTAS-B incorporated predictive chemical transport modelling (GEOS-Chem), satellite
129 observations, a ground-based in-situ network of sondes (Environment Canada) and ground-base
130 samplers and profilers (Dalhousie Ground Station, DGS), and the UK Facility for Airborne
131 Atmospheric Measurements Airborne Research Aircraft (FAAM-ARA) for inflight sampling. For a
132 complete overview of the BORTAS-B set-up and instrumentation, see (Palmer, 2013). The ARA flew
133 fourteen research flights over the campaign period.

134 The flight paths of the ARA flights that we analyze in this paper can be seen in Figure 1. Flights
135 BAE-b622 and BAE-b623 were research flights between Halifax, NS and Sherbrooke, QB spanning
136 July 20-21, 2011. They flew ascent and descent patterns (ranging ~1-7 km ASL) to sample vertical and
137 horizontal transects in regions forecasted to contain biomass-burning plumes. These flights were
138 selected because they were roughly co-located and back-to-back, increasing the likelihood of sampling
139 similar outflow and allowing for a common plume criteria to be applied across both flights. They also
140 contained the majority of the biomass-burning aerosol sampled during the 14-flight campaign.

141 In addition, Flight b622 sampled along a relatively straight path to/from the fires that allowed
142 for analysis of the evolution of plume aerosol properties (Flight b623 had a much more complicated
143 and compact sampling path so we did not use this flight to determine the evolution of aerosol
144 properties). We have divided these flights into vertical transects by ascent/descent with the midpoints
145 transect represented in Figure 1.

146 The sampled wildfire plumes originated from intense regional fires near the Northwestern
147 Ontario-Manitoba border (centred 52° N, 93° W). The MODIS hotspots in Figure 1 show a number of
148 intense fires (fire radiative power >100 MW) in northwestern Ontario for the three days prior to the
149 analyzed flights (June 17-20, 2011). According to the Ontario Ministry of Natural Resources, Ontario
150 experienced one of its worst fire seasons in terms of burned area with 635,374 hectares burned in 2011.
151 The abundance of individual fires in a relatively large source region led to mixed combustion phases
152 and dominant hotspots over the course of the campaign. A combination of flaming and smouldering
153 phases were reported by Natural Resources Canada with primary fuels consisting of jack pine (*pinus*
154 *banksiana*) and black spruce (*picea mariana*) throughout the fire region (Ontario ministry of natural

155 resources: 2011 forest summary).

156 The dominant west-east climatological meteorology during the BORTAS-B campaign allowed
157 the biomass-burning emissions from these fires to be transported downwind over the ground-base
158 (DGS) in Halifax, NS (44.5° N, 63.1° W). The plumes intersected by flights b622 and b623 had a
159 physical transport age estimated through HYSPLIT backtrajectories of between 1-2 days as
160 summarized in Table 1. The backtrajectory analysis (not shown) shows air masses passing over the
161 biomass-burning region later being intersected by the flight paths at varying altitudes. The estimated
162 photochemical age of the plumes, calculated by Palmer et al. (2013) (by non-methane hydrocarbon
163 analysis; Parrish et al. (2007)), were 1-5 days for b622 and 2-4 days for b623. These estimates may be
164 longer than the physical transport ages due to the entrainment of background air (which is more
165 photochemically aged) into the plumes.

166

167 **2.2 ARA Instrumentation**

168 The ARA aircraft was outfitted with instruments designed for sampling chemical and physical
169 characteristics of biomass-burning outflow. Gaseous and particulate in-flight sampling was
170 accomplished across a suite of instruments; the relevant instruments for this study are described below.
171 A full description of all payload instruments can be found in (Palmer, 2013).

172 The suite of instruments on the ARA included measurements of multiple gaseous
173 biomass-burning tracers. Carbon monoxide (CO) mole fraction was measured via VUV Fast
174 fluorescence CO analyzer averaged over 1s (3% estimated accuracy). Acetonitrile (CH₃CN), a
175 biomass-burning marker VOC associated with plant pyrolysis, was measured along with a number of
176 other VOCs with a proton-transfer-reaction mass spectrometer (PTR-MS) system (co: University of
177 East Anglia). The PTR-MS concentrations were averaged over 1s with an estimated precision of ± 37
178 ppt (Palmer, 2013).

179 Aerosol composition measurements used here were taken by i) refractory black carbon (BC)
180 mass and number measurements from a Single Particle Soot Photometer (SP2) (accuracy 20%,
181 precision 5%, 5s averaging time) and ii) non-refractory organic aerosol (OA) via an aerosol mass
182 spectrometer (precision ~ 15-150 ng m⁻³) both operated by the University of Manchester (Jolleys et al.
183 2014; Taylor et al., 2014). The number concentrations of the combined aerosol particles was measured
184 by Scanning Mobility Particle Sizer (SMPS) with 26 lognormally-spaced diameter bins ranging from
185 20-330 nm and corrected to STP. A full scan takes 60s (Palmer, 2013). The SMPS data was inverted

186 using the commonly-used Wiedensohler (1988) parameterisation, however recent work has suggested
187 that this may be quantitatively unreliable for applications to aircraft data due to variations in the
188 charging efficiency with pressure (López-Yglesias and Flagan, 2013; Leppä et al., 2014). While this
189 may have affected the magnitude of the number concentrations, no altitude dependency was noted in
190 the sizing data, so the conclusions of this paper regarding particle size are unaffected.

191 The combination of gas and particle tracer measurements listed above were used to identify
192 flight periods of biomass-burning plume sampling, determine if SOA formation or OA evaporation may
193 have occurred in the plume, and characterize the size-distribution of aerosols within the plume.

194

195 **2.3 Plume Criteria**

196 We determine if measurements are in-plume versus out-of-plume using threshold plume criteria. We
197 designate sampling periods as in-plume if pre-specified threshold values of four tracer species: CO,
198 CH₃CN, BC, and OA, were exceeded. For out-of-plume conditions, we determine “background values”
199 for each tracer by averaging the tracers over the out-of-plume periods.

200 Carbon monoxide ($\tau_{\text{co}} \sim$ months (Staudt et al., 2001)) and acetonitrile ($\tau_{\text{ace}} \sim$ 6 months
201 (Holzinger et al., 2005)) were used in conjunction as gaseous tracers due to their high mixing ratios in
202 biomass-burning plumes relative to the background and long atmospheric lifetimes relative to the
203 estimated plume transport times. The background CO levels were 80-120 ppbv with an overall average
204 of 100 ppbv. The threshold CO value was set to 150 ppbv (1.5 x [background]), with some CO
205 concentrations in-plume reaching ten times background concentrations (1000 ppbv). The threshold
206 CH₃CN level was 200 pptv (background \sim 100 pptv).

207 The particulate matter thresholds (BC number, OA mass) were introduced to ensure
208 high-enough aerosol contributions to the plume to analyze size-distributions. This ensured high-gas,
209 low-aerosol sampling periods were not included in the size-distribution analysis. At least one case of
210 this situation in BORTAS-B has been attributed by Franklin et al. (2014) to aerosol rainout during
211 transport. The mean background concentrations for both BC number and OA mass were minimal (< 20
212 cm^{-3} and $2 \mu\text{g m}^{-3}$ respectively). The threshold values were set to 50 cm^{-3} for BC number and $20 \mu\text{g m}^{-3}$
213 for OA mass. These thresholds for particles are higher relative to background than CO and CH₃CN
214 because we wanted to exclude a higher-elevation plume that had undergone aerosol wet deposition (see
215 Section 3.2).

216 The selected CO, CH₃CN, aerosol data, and flight altitude time series for Flight b622 is shown

217 in Figure 2. The flight is divided into transects (labeled 1-9 and colored) as seen in the altitude plot
218 (Figure 2, bottom). We use these in-plume time periods to differentiate between in-plume and
219 background aerosols throughout the paper.

220

221 **2.4 Model Description**

222 We use a Lagrangian box model to simulate the evolution of the biomass-burning size distribution due
223 to coagulation. The model has fifty-five logarithmically distributed size bins that correspond to the size
224 bins of the SMPS on the ARA and extend to both larger and smaller diameters. The model includes
225 coagulation and dilution as the only physical processes, with no chemistry or speciation of the aerosol
226 (we show in Section 3.2 that we cannot see evidence of net OA condensation/evaporation in the plume
227 over the sampling period). The model distributions are therefore limited by the lack of condensational
228 growth known to occur in BB plumes during the first few hours of aging (e.g. Reid et al. 1998).

229 We use an inverse method to estimate the initial young (~ 1 hour) size-distributions by
230 successively running the model from initial conditions to BORTAS observation forward in time and
231 changing the initial size distribution until the model most closely matches the observed aged size
232 distribution. This method estimates the initial distribution assuming that coagulation was the only
233 physical processes affecting the in-plume particles. The box model does not include any cloud
234 interaction chemistry, which could have influenced the distribution considerably depending on
235 meteorological conditions, notably through wet deposition and aqueous chemistry.

236 Each model forward simulation requires the young size-distribution input as a single lognormal
237 mode with parameters: median diameter (D_{pm}), modal width (σ), and particle number (N_0). For
238 coagulation, we use the brownian coagulation kernel of Fuchs (1964). Dilution of the plume in
239 transport was modelled using a simple e-folding volume mixing time, τ_{dil} . This parameter controlled the
240 entrainment each timestep between the in-plume and background aerosol. The rate of plume dilution
241 may significantly affect the rate of coagulation throughout the simulation (the coagulation rate is
242 proportional to N^2). Different values of τ_{dil} were tested to account for a range of entrainment rates as the
243 dilution rate in the plume is relatively unconstrained. We test τ_{dil} values of 24, 36 and 48 hours. The 36
244 hr dilution timescale was calculated as the mean timescale for dilution from Gaussian plume equations
245 with an initial plume width of 10 km in a neutral stability environment (Klug, 1969) (note, however,
246 that expansion occurs at faster timescales early in the plume aging, and this timescale slows with time).
247 The range (24 – 48 hrs) accounts for atmospheric stability and plume width variations in the BORTAS

248 source region. The model simulation time is 48 hours based on the upper age limits shown in Table 1.

249 To determine the best estimate for initial conditions, we simulate a range of young plume
250 parameters: median diameter, D_{pm} , modal width, σ , and number, N_0 . The input median diameter range
251 was between 60-120 nm (increment = 1 nm), with σ ranging from 1.0-2.5 (increment = 0.1) and N_0
252 ranging from 5,000- 150,000 cm^{-3} (increment = 500 cm^{-3}). The parameter space was optimized by
253 brute force (i.e. every combination of input parameters was simulations) for each set dilution time and
254 the final modelled size-distribution was compared to the observed in-plume size distribution by an
255 equally weighted objective function. The objective function used was the sum of the absolute residual
256 across the SMPS range. Modelled data outside of the SMPS size range was not used in the objective
257 function.

258

259 **3. Results**

260 **3.1 Observed size distributions**

261 Observed SMPS size-distributions for individual plume transects showed highly elevated particle
262 counts with little variation between transects and flights. The transect-divided data for Flight b622 are
263 shown in Figure 3. Transects 2-6 and 9, show a clearly elevated accumulation mode within the plume,
264 with peak median diameters of 180-240 nm. Transects 1, 7 and 8 have significantly less data (< 3 data
265 points per size bin) due to the lesser in-plume sampling periods (incomplete SMPS scans).

266 Those transects with sufficient plume data (> 3 data points per bin) are plotted against their
267 accumulation mode median diameter D_{pm} in Figure 4. We do not observe any discernible trend in
268 size-distribution with the distance from the source fires in either median diameter or number
269 concentration. This lack of a trend suggests that the microphysical processing during the range of
270 distances sampled has smaller effects on the size distribution than the variability between plumes for
271 Flight b622. Similarly small inter-transect variation was seen for Flight b623 (not shown). The median
272 size-distributions show no bias based on altitude or ascent/descent rate as an artefact of SMPS flow rate
273 fluctuations from altitude changes (not shown).

274 The composite median distribution across all plume sampling periods and both flights is shown
275 in Figure 5a. This characteristic size distribution is presented as a median value, minimizing the
276 contributions of outlying data. Figure 5b shows the same composite distribution normalized by CO
277 concentration to attempt to account for differences in the amount of emissions from the source. The
278 plume particle size-distribution shows the median size distribution highlighted in black, with the 25th

279 and 75th percentiles outlined in red. A clearly defined accumulation mode was identified centred at D_{pm}
280 = 230 nm and with a modal width of 1.5, based on a single lognormal mode fit. Normalizing the plume
281 distribution by CO mixing ratio produced a very similar pattern shown in Figure 5b (accumulation
282 mode: $D_{pm} = 230$ nm, $\sigma = 1.4$). The composite background aerosol size-distribution (sampling periods
283 that failed the in-plume criteria) are seen in black (with 25th and 75th percentiles shown in gray) in
284 Figure 5a. It shows relatively constant $dN/d\log D_p$ concentrations across the SMPS range and is lacking
285 the concentrated accumulation mode found in-plume.

286 The aged composite size-distribution and associated lognormal parameters are similar to those
287 found in other field studies of aged biomass-burning emissions. Aged biomass-burning size
288 distributions compiled by Janhäll et al. (2010) for all different fuel types show a similar D_{pm} to modal
289 width ratio ($D_{pm} = 175 - 300$ nm, $\sigma = 1.7 - 1.3$). Capes et al. (2008) show a similar aged BB size
290 distribution median diameter over West Africa during the DABEX campaign ($D_{pm} = 240$ nm). The
291 ARCTAS-B campaign over Northern Canada sampled similar Boreal pyrogenic outflow and collected
292 very similar aged distributions of BC and OC constituents ($D_{pm} = 224 \pm 14$ nm, $\sigma = 1.33 \pm 0.05$)
293 (Kondo et al., 2011).

294 Of note in the BORTAS-B plume size distribution is the elevated number concentrations of
295 small diameter particles (20-90 nm), which form an elevated small-diameter 'tail' of the distribution.
296 These higher concentrations were not expected due to the high rate of removal of small particles by
297 coagulation with the larger particles in the accumulation mode. We calculated first-order
298 coagulation-loss timescales to investigate the timescale of the removal of these small particles by the
299 larger plume particles. If these small particles were brought into the plume by entrainment of
300 background air, there would be an associated amount of time before these particles were lost by
301 coagulation. For the calculation, we assume brownian coagulation of entrained background aerosol
302 (bin range 20-90 nm) with the observed in-plume SMPS data (90-333 nm) and with artificial
303 large-diameter bins from 330 nm – 1 μ m (6 bins). These artificial bin concentrations were based on the
304 accumulation mode lognormal fit and account for those particle concentrations at sizes larger than
305 those measured by the SMPS but that nonetheless contribute to the coagulation scavenging of the
306 small-diameter particles. Particles with diameters $> 1 \mu$ m were ignored since their relative scarcity
307 relative to the large number of accumulation-mode particles causes a negligible impact on the
308 number-concentration driven coagulation process.

309 The predicted concentrations of background aerosol remaining after 24, 36 and 48 hours are

310 shown in Figure 6a. These times are within the estimated physical transport age ranges of the transects.
311 After 12 hours, coagulation alone has already caused a significant decrease in the concentrations of the
312 smallest measured particles, reducing them to levels well below the concentrations observed in plume
313 (red line). This deficit increases with time ($t=36$ hrs, $t=48$ hrs). The coagulation lifetimes of the
314 particles in this diameter range (30-90 nm) are seen in Figure 6b and extend into the tens of hours. Note
315 that the concentrations of these small-diameter particles are similar in the plume compared to the
316 background. This means that the entrainment rate of background air into the plume would need to be
317 much faster than the coagulation loss timescales (~ 5 hours for 20 nm particles) in order for
318 entrainment to sustain the number of small particles. If entrainment timescales were significantly
319 shorter than 5 hours, the plume would almost completely disperse into the background within 1 day.

320 There are a number of mechanisms other than entrainment that could explain the higher tail
321 concentrations found in plume despite the short coagulation lifetimes. In-plume nucleation and
322 subsequent growth to SMPS-detectable sizes could also partially account for sustained elevated small
323 particle concentrations. Hennigan et al. (2012) showed with the FLAME-III chamber studies that
324 in-plume nucleation was possible as a result of photochemical aging and SOA production in smoke
325 plumes. Nucleation modes in association with smoke plumes have also been observed previously in
326 field studies (Hobbs et al., 2003; Rissler et al., 2006). We attempted to determine the nucleation and
327 growth rates required to sustain the observed concentration of small particles; however, the necessary
328 condensational growth rates that were required to fit the observed data were unrealistically high, which
329 we see as evidence against nucleation/growth being the primary source of the small particles. Thus, we
330 are unsure of the source of these particles.

331

332 **3.2 Net production/loss of organic aerosol with time**

333 Enhancement ratios are a way of characterizing plume chemistry as a ratio of a specific species to a
334 reference species. This was done for the sampled BORTAS pyrogenic outflow by taking the excess
335 (background concentration removed) of the AMS organic aerosol normalized to the excess CO ($\Delta OA /$
336 ΔCO). Only those data which were in excess of the mean background ($CO = 100$ ppbv, $OA = 2 \mu g m^{-3}$)
337 were compiled. The characteristic $\Delta OA / \Delta CO$ ratio can be used as a comparison value between fires of
338 different fuel type, phases or photochemical ages.

339 Since CO has a sufficiently long lifetime and is co-emitted with OA in abundance at the source,
340 any changes in the organic aerosol enhancement ratio over the lifetime of the plume are attributed to

341 in-plume chemistry. Entrainment of background air into the most concentrated sections of the plumes is
342 slow (timescale > 10 hours) and we consider any change in $\Delta\text{OA}/\Delta\text{CO}$ ratios to be attributable to
343 in-plume processes only. The formation of secondary organic aerosol is possible within the plume by
344 oxidation of organic vapors to lower-volatility products. Evaporation of less-volatile POA during
345 plume dilution competes with the SOA condensation. The net OA production is therefore: $\Delta\text{OA}_{\text{net}} =$
346 $\text{SOA}_{\text{prod}} - \text{OA}_{\text{evap}}$. Changes in the $\Delta\text{OA}/\Delta\text{CO}$ ratio over time can therefore indicate which of the two
347 processes is dominant.

348 The organic aerosol enhancement ratios for Flight b622 are shown in Figure 7. There is a fairly
349 pronounced altitude dependence as seen in Figure 7a, with several high altitude (~ 7 km) samples
350 having fairly low excess organic aerosol, but significant ΔCO (300 ppbv). This trend is featured in
351 Franklin et al. (2014) where the high-altitude plume showed evidence of an aerosol rainout event
352 causing low $\Delta\text{OA}/\Delta\text{CO}$ within the plume transected at those high altitudes.

353 We will focus on the lower-altitude plume where the aerosol was not rained out, so we employ a
354 height cutoff of 4.6 km to restrict the enhancement ratio calculations to lower-altitude, OA-rich plumes
355 least likely to have seen significant reduction in organic aerosol from wet deposition. The mean
356 enhancement ratios by transect are seen in Figure 7b (for transect locations see Figure 1). Only those
357 sampling periods that passed the OA and CO plume criteria (detailed above) are shown. The
358 lower-altitude plume enhancement ratio show correlations of $R^2 > 0.5$ for each transect with the
359 exception of Transect 8 ($R^2 = 0.26$).

360 Figure 8 shows $\Delta\text{OA}/\Delta\text{CO}$ as a function of the distance from the source fires (horizontal error
361 bars correspond to error due to the radius of the Ontario fire region, vertical error bars are calculated
362 from transect data scatter). Compared across transects, the enhancement ratios show no significant
363 trend (to P-value = 0.55). The average enhancement ratio is 0.14 ± 0.01 [$\mu\text{g m}^{-3} \text{ ppbv}^{-1}$] and can be
364 considered characteristic of the aged boreal plume during these BORTAS flights.

365 The lack of trend in Figure 8 suggests that we cannot determine if there was any net
366 production/evaporation of OA happening inside the plume over this sampling period (plume ages: 1-2
367 days). Any SOA produced photochemically inside the plume is either being accompanied by an
368 opposing loss of POA or at such a rate that is below the observational variability over the sampled time
369 period. The statistically invariant $\Delta\text{OA}/\Delta\text{CO}$ does not discount evaporation-condensational cycling of
370 POA and SOA, or the effects such recondensation would have on the size-distribution (although there
371 was no apparent trend in the size distribution either [Figure 4]). No increase in normalized excess OA

372 fraction means significant levels of excess SOA were not likely driving condensational growth,
373 ensuring that coagulation was dominating the size-distribution evolution during the period of aging
374 between 1 day and 2 days since emission.

375

376 Since no significant trend was found in size-distribution D_{pm} with distance from the source in
377 the observations (> 1000 km; see Figure 4), any effect of POA-SOA cycling on the shape of the
378 distribution cannot be isolated above the noise. However, it does not preclude that there was significant
379 net OA production/evaporation that occurred prior to or after this observed period as has been observed
380 in other BB field studies (e.g. Akagi et al. (2012); Yokelson et al. 2009). Thus, although evidence of
381 photo-oxidation and chemical processing was observed in-plume by Parrington et al. (2013), any
382 chemical composition impact on the size-distribution at the ages observed here seems negligible.

383

384 **3.2 Estimation of the young biomass burning size distribution**

385 In this section, we test the parameter space of our microphysical model to estimate the young (aged
386 ~ 1 -3 hr) plume size-distribution emitted from the source fires. We allowed the young biomass-burning
387 size distribution to evolve for 48 hours and compared the result to the observed SMPS plume
388 composite distribution to isolate the optimal young plume size-distribution parameters. These were
389 then compared to observed young BB size distributions for context.

390 The Lagrangian microphysical model was run for 48 hours with fixed entrainment coefficients
391 of $\tau_{dil}=24$ hrs, $\tau_{dil} = 36$ hours and $\tau_{dil}=48$ hours. Figure 9a shows the optimal young plume distribution
392 parameters that were obtained for each tested entrainment rate (with particle number concentration per
393 co-emitted CO above background). Figure 9b shows the modelled aged distributions plotted with the
394 measured distribution. None of the model runs can capture the elevated concentrations in the tail
395 particles in the SMPS data, though this is expected due to the coagulation-dominant aging in the model
396 (discussed above) and adds further uncertainty to the existence of this small tail.

397 The young plume size-distributions are unimodal with median diameters of 94 nm, 67 nm, and
398 59 nm for $\tau_{dil}=24$ hrs, $\tau_{dil}=36$ hrs and $\tau_{dil}=48$ hrs, respectively ($\sigma = 1.7, 2.1, 2.8$, respectively). The
399 higher entrainment rate of background aerosol requires the young plume distribution to be narrower
400 (lower σ) and have an initial median diameter closer to the final diameter ($D_{pfinal} = 230$ nm). The initial
401 number concentrations in the young plume were found to be optimized at 62,500, 80,000, and 115,000
402 [cm^{-3}] for $\tau_{dil}=48$ hrs, $\tau_{dil}=36$ hrs and $\tau_{dil}=24$ hrs, respectively. The initial higher concentrations, narrower

403 modal width and larger median diameter are required for the higher entrainment rates to account for the
404 more rapid plume dilution and subsequently the slowing of the coagulation rates. Normalized to
405 estimated freshly emitted excess CO, the young plume number concentrations are 37, 53, and 60 cm⁻³
406 ppbv⁻¹ (for $\tau_{\text{dil}}=24$ hrs, $\tau_{\text{dil}}=36$ hrs and $\tau_{\text{dil}}=48$ hrs respectively; see Figure 9a). The similar magnitudes of
407 these normalized size-distributions indicate a relatively robust particle/CO ratio regardless of the
408 dilution parameter, though in the absence of the source fire fuel densities, a direct comparison to
409 emission factors (kg⁻¹) cannot be made.

410 As the exact aging time and dilution profiles are unknown in addition to uncertainties in the
411 plume age, we cannot say with certainty which of these estimates is best; however, these results
412 compare to the field observations presented in Janhäll et al. (2010) for fresh plume smoke (range: $D_{\text{pm}} =$
413 100-150 nm) and to small-scale lab experiments measuring fresh smoke (range: $D_{\text{pm}} = 30-90$ nm)
414 (Hosseini et al., 2010). Capes et al. (2008) conducted a similar fresh-plume size-distribution estimate
415 from their observed DABEX aged African smoke data using a coagulation box-model without dilution.
416 Their estimates for very fresh smoke have a much smaller D_{pm} (~ 30 nm).

417 The plume size distributions modelled here are very sensitive to microphysical processes
418 directly after emission. Very close to the source, rapid dilution and condensation (due to cooling) may
419 occur, which are not captured by the coagulation/dilution model we have developed. Thus the
420 modelled plumes are better categorized as 'young' rather than freshly emitted. The young-plume
421 distributions in this study neglect any immediate effects of condensation and/or evaporation of OA on
422 the size-distribution during cooling and dilution respectively, and focus on the effects of coagulation
423 which shape the size-distribution over a longer timescale (~ 10hrs). There is therefore an associated
424 uncertainty in the young distributions due to the exclusion of condensational growth from the model,
425 despite evidence of its effects on BB particle sizes especially during the first hours of aging (e.g. Reid
426 et al. 1998).

427 Figure 10 shows a time series of the optimal modelled size distribution for $\tau_{\text{dil}} = 36$ hrs over the
428 48 hr period. The median diameter growth (black line) occurs more rapidly during the early stages of
429 the plume due to the higher particle concentrations before significant dilution. Eighty percent of the
430 final median diameter is achieved within 10 hrs of coagulation processing. Less drastic but similar
431 rapid growth by coagulation was seen by Capes et al. (2008) in their coagulation box model. This quick
432 size distribution evolution within the early plume stages suggests that large grid box models (global,
433 regional) should be using aged biomass-burning size-distributions as input. Figure 10 also shows the

434 size-distribution growth slowing considerably as the particle concentrations decrease and perhaps can
435 provide an explanation for the lack of strong trend in the observed BORTAS D_{pm} across transects
436 (Figure 4), which were already aged between 1-2 days when they were sampled.

437

438 **4. Conclusions**

439 The BORTAS-B campaign provided the opportunity to collect numerous gaseous and aerosol
440 measurements from aged North-American biomass-burning plumes in July, 2011. The boreal fire
441 emissions in northwestern Ontario were transported (1-2 days) downwind to where they were sampled
442 by the FAAM BAE-146 research aircraft. We analyzed the plume data from two research flights (b622
443 and b623) and found little variation in size-distributions between transects.

444 A characteristic size-distribution consistent between flights and transects was dominated by the
445 accumulation mode with $D_{pm} = 230$ nm and with $\sigma = 1.5$. This unimodal result is consistent with aged
446 biomass-burning observations found globally in the previous field studies (Capes et al., 2008; Janhäll et
447 al., 2010; Kondo et al., 2011).

448 We also found elevated concentrations of small-diameter particles in the plume contrary to their
449 coagulation lifetimes associated with the biomass-burning-associated accumulation mode. We were not
450 able to explain such concentrations by entrainment of background aerosol alone. The presence of such
451 concentrations in the size-distribution tail remains inconclusive.

452 The $\Delta OA / \Delta CO$ enhancement ratios across Flight b622 show a strong linear correlation below
453 4.6 km ($R^2 > 0.50$) with values between $(0.09 - 0.17) \pm 0.01 \mu\text{g m}^{-3} \text{ppbv}^{-1}$. We found no trend in transect
454 enhancement ratios with distance from the source, indicating no significant net SOA production
455 in-plume over the sampling period, though this does not preclude OA production or loss during earlier
456 stages of aging.

457 We used a microphysical model to estimate the young plume size distribution associated with
458 the BORTAS-B observations. Optimizing lognormal parameters for different assumed dilution
459 coefficients ($\tau_{dil} = 24, 36, 48$ hrs), the young plume size distribution had $D_{pm} = 59-94$ nm, $\sigma = 2.8-1.7$,
460 and $N_0 = 62,500 - 115,000 \text{ cm}^{-3}$ ($37 - 60 \text{ cm}^{-3} \text{ppbv}^{-1}$ normalized by initial co-emitted excess CO).
461 Though the model lacks condensation and chemical considerations, processing through coagulation and
462 dilution alone led to 80% of the observed 48-hour median-diameter growth within the first 10 hrs. This
463 suggests that global climate models should be using coagulation-aged BB size distribution inputs to
464 account for the rapid evolution in plume particle size occurring on scales smaller than the gridbox

465 length.

466

467 **5. References**

- 468 Adler, G., Flores, J. M., Abo Riziq, a., Borrmann, S. and Rudich, Y.: Chemical, physical, and optical
469 evolution of biomass burning aerosols: a case study, *Atmos. Chem. Phys.*, 11(4), 1491–1503,
470 doi:10.5194/acp-11-1491-2011, 2011.
- 471 Akagi, S. K., Craven, J. S., Taylor, J. W., McMeeking, G. R., Yokelson, R. J., Burling, I. R., Urbanski,
472 S.P., Wold, C. E., Seinfeld, J. H., Coe, H., Alvarado, M. J. and Weise, D. R.: Evolution of trace
473 gases and particles emitted by a chaparral fire in California, *Atmos. Chem. Phys.*, 12(3),
474 1397–1421, doi:10.5194/acp-12-1397-2012, 2012
- 475 Alonso-Blanco, E.: Impact of Biomass Burning on Aerosol Size Distribution, Aerosol Optical
476 Properties and Associated Radiative Forcing, *Aerosol Air Qual. Res.*, 006, 708–724,
477 doi:10.4209/aaqr.2013.05.0163, 2014.
- 478 Andreae, M. O. and Merlet, P.: Emission of trace gases and aerosols from biomass burning, *Global*
479 *Biogeochem. Cycles*, 15(4), 955–966, doi:10.1029/2000GB001382, 2001.
- 480 Andreae, M. O., Artaxo, P., Fischer, H., Freitas, S. R., Grégoire, J.-M., Hansel, A., Hoor, P., Kormann,
481 R., Krejci, R., Lange, L., Lelieveld, J., Lindinger, W., Longo, K., Peters, W., de Reus, M.,
482 Scheeren, B., Silva Dias, M. A. F., Ström, J., van Velthoven, P. F. J. and Williams, J.: Transport of
483 biomass burning smoke to the upper troposphere by deep convection in the equatorial region,
484 *Geophys. Res. Lett.*, 28(6), 951–954, doi:10.1029/2000GL012391, 2001.
- 485 Bauer, S. E., Menon, S., Koch, D., Bond, T. C. and Tsigaridis, K.: A global modeling study on
486 carbonaceous aerosol microphysical characteristics and radiative effects, *Atmos. Chem. Phys.*,
487 10(15), 7439–7456, doi:10.5194/acp-10-7439-2010, 2010.
- 488 Capes, G., Johnson, B., McFiggans, G., Williams, P. I., Haywood, J. and Coe, H.: Aging of biomass
489 burning aerosols over West Africa: Aircraft measurements of chemical composition, microphysical
490 properties, and emission ratios, *J. Geophys. Res.*, 113, D00C15, doi:10.1029/2008JD009845,
491 2008.
- 492 Carrico, C. M., Petters, M. D., Kreidenweis, S. M., Sullivan, A. P., McMeeking, G. R., Levin, E. J. T.,
493 Engling, G., Malm, W. C. and Collett, J. L.: Water uptake and chemical composition of fresh
494 aerosols generated in open burning of biomass, *Atmos. Chem. Phys.*, 10(11), 5165–5178,
495 doi:10.5194/acp-10-5165-2010, 2010.
- 496 Chen, W.T., Lee, Y. H., Adams, P. J., Nenes, A. and Seinfeld, J. H.: Will black carbon mitigation
497 dampen aerosol indirect forcing?, *Geophys. Res. Lett.*, 37(9), L09801,
498 doi:10.1029/2010GL042886, 2010.
- 499 Cubison, M. J., Ortega, A. M., Hayes, P. L., Farmer, D. K., Day, D., Lechner, M. J., Brune, W. H., Apel,
500 E., Diskin, G. S., Fisher, J. A., Fuelberg, H. E., Hecobian, A., Knapp, D. J., Mikoviny, T., Riemer,
501 D., Sachse, G. W., Sessions, W., Weber, R. J., Weinheimer, A. J., Wisthaler, A. and Jimenez, J. L.:
502 Effects of aging on organic aerosol from open biomass burning smoke in aircraft and laboratory
503 studies, *Atmos. Chem. Phys.*, 11(23), 12049–12064, doi:10.5194/acp-11-12049-2011, 2011.
- 504 DeCarlo, P. F., Ulbrich, I. M., Crounse, J., de Foy, B., Dunlea, E. J., Aiken, A. C., Knapp, D.,

505 Weinheimer, A. J., Campos, T., Wennberg, P. O. and Jimenez, J. L.: Investigation of the sources
506 and processing of organic aerosol over the Central Mexican Plateau from aircraft measurements
507 during MILAGRO, *Atmos. Chem. Phys.*, 10(12), 5257–5280, doi:10.5194/acp-10-5257-2010,
508 2010.

509 Engelhart, G. J., Hennigan, C. J., Miracolo, M. A., Robinson, A. L. and Pandis, S. N.: Cloud
510 condensation nuclei activity of fresh primary and aged biomass burning aerosol, *Atmos. Chem.*
511 *Phys.*, 12(15), 7285–7293, doi:10.5194/acp-12-7285-2012, 2012.

512 Franklin, J. E., Drummond, J. R., Griffin, D., Pierce, J. R., Waugh, D. L., Palmer, P. I., Parrington, M.,
513 Lee, J. D., Lewis, A. C., Rickard, A. R., Taylor, J. W., Allan, J. D., Coe, H., Walker, K. A.,
514 Chisholm, L., Duck, T. J., Hopper, J. T., Blanchard, Y., Gibson, M. D., Curry, K. R., Sakamoto, K.
515 M., Lesins, G., Dan, L., Kliever, J. and Saha, A.: A case study of aerosol scavenging in a biomass
516 burning plume over eastern Canada during the 2011 BORTAS field experiment, *Atmos. Chem.*
517 *Phys.*, 14(16), 8449–8460, doi:10.5194/acp-14-8449-2014, 2014.

518 Fuchs, N. A.: *The Mechanics of Aerosols*, Oxford: Pergamon Press., 1964.

519 Grieshop, A. P., Logue, J. M., Donahue, N. M. and Robinson, A. L.: Laboratory investigation of
520 photochemical oxidation of organic aerosol from wood fires 1: measurement and simulation of
521 organic aerosol evolution, *Atmos. Chem. Phys.*, 9(4), 1263–1277, doi:10.5194/acp-9-1263-2009,
522 2009.

523 Haywood, J. and Boucher, O.: Estimates of the direct and indirect radiative forcing due to tropospheric
524 aerosols: a review, *Rev. Geophys.*, 38, 513–543, doi:10.1029/1999RG000078, 2000.

525 Hecobian, A., Liu, Z., Hennigan, C. J., Huey, L. G., Jimenez, J. L., Cubison, M. J., Vay, S., Diskin, G.
526 S., Sachse, G. W., Wisthaler, A., Mikoviny, T., Weinheimer, A. J., Liao, J., Knapp, D. J.,
527 Wennberg, P. O., Kürten, A., Crounse, J. D., Clair, J. St., Wang, Y. and Weber, R. J.: Comparison
528 of chemical characteristics of 495 biomass burning plumes intercepted by the NASA DC-8 aircraft
529 during the ARCTAS/CARB-2008 field campaign, *Atmos. Chem. Phys.*, 11(24), 13325–13337,
530 doi:10.5194/acp-11-13325-2011, 2011.

531 Hennigan, C. J., Miracolo, M. A., Engelhart, G. J., May, A. A., Presto, A. A., Lee, T., Sullivan, A. P.,
532 McMeeking, G. R., Coe, H., Wold, C. E., Hao, W.-M., Gilman, J. B., Kuster, W. C., de Gouw, J.,
533 Schichtel, B. A., Kreidenweis, S. M. and Robinson, A. L.: Chemical and physical transformations
534 of organic aerosol from the photo-oxidation of open biomass burning emissions in an
535 environmental chamber, *Atmos. Chem. Phys.*, 11(15), 7669–7686, doi:10.5194/acp-11-7669-2011,
536 2011.

537 Hennigan, C. J., Westervelt, D. M., Riipinen, I., Engelhart, G. J., Lee, T., Collett, J. L., Pandis, S. N.,
538 Adams, P. J. and Robinson, A. L.: New particle formation and growth in biomass burning plumes:
539 An important source of cloud condensation nuclei, *Geophys. Res. Lett.*, 39(9), n/a–n/a,
540 doi:10.1029/2012GL050930, 2012.

541 Heringa, M. F., DeCarlo, P. F., Chirico, R., Tritscher, T., Dommen, J., Weingartner, E., Richter, R.,
542 Wehrle, G., Prévôt, A. S. H. and Baltensperger, U.: Investigations of primary and secondary
543 particulate matter of different wood combustion appliances with a high-resolution time-of-flight
544 aerosol mass spectrometer, *Atmos. Chem. Phys.*, 11(12), 5945–5957,
545 doi:10.5194/acp-11-5945-2011, 2011.

546 Hobbs, P. V., Sinha, P., Yokelson, R. J., Christian, T. J., Blake, D. R., Gao, S., Kirchstetter, T. W.,

547 Novakov, T. and Pilewskie, P.: Evolution of gases and particles from a savanna fire in South
548 Africa, *J. Geophys. Res.*, 108(D13), doi:10.1029/2002JD002352, 2003.

549 Holzinger, R., Williams, J., Salisbury, G., Klüpfel, T., de Reus, M., Traub, M., Crutzen, P. J. and
550 Lelieveld, J.: Oxygenated compounds in aged biomass burning plumes over the Eastern
551 Mediterranean: evidence for strong secondary production of methanol and acetone, *Atmos. Chem.*
552 *Phys.*, 5(1), 39–46, doi:10.5194/acp-5-39-2005, 2005.

553 Hosseini, S., Li, Q., Cocker, D., Weise, D., Miller, A., Shrivastava, M., Miller, J. W., Mahalingam, S.,
554 Princevac, M. and Jung, H.: Particle size distributions from laboratory-scale biomass fires using
555 fast response instruments, *Atmos. Chem. Phys.*, 10(16), 8065–8076,
556 doi:10.5194/acp-10-8065-2010, 2010.

557 Hudson, P. K., Murphy, D. M., Cziczo, D. J., Thomson, D. S., de Gouw, J. A., Warneke, C., Holloway,
558 J., Jost, H.-J. and Hübner, G.: Biomass-burning particle measurements: Characteristic composition
559 and chemical processing, *J. Geophys. Res.*, 109(D23), D23S27, doi:10.1029/2003JD004398,
560 2004.

561 Huffman, J. A., Docherty, K. S., Mohr, C., Cubison, M. J., Ulbrich, I. M., Ziemann, P. J., Onasch, T. B.
562 and Jimenez, J. L.: Chemically-Resolved Volatility Measurements of Organic Aerosol from
563 Different Sources, *Environ. Sci. Technol.*, 43(14), 5351–5357, doi:10.1021/es803539d, 2009.

564 Jacobson, M. Z.: Strong radiative heating due to the mixing state of black carbon in atmospheric
565 aerosols, *Nature*, 409(6821), 695–697, 2001.

566 Janhäll, S., Andreae, M. O. and Pöschl, U.: Biomass burning aerosol emissions from vegetation fires :
567 particle number and mass emission factors and size distributions, *Atmos. Chem. Phys.*, 1427–1439,
568 2010.

569 Klug, W. : A method for determining diffusion conditions from synoptic observations, *Staub-Reinhalt.*
570 *Luft*, 29, 14–20, 1969.
571

572 Kondo, Y., Matsui, H., Moteki, N., Sahu, L., Takegawa, N., Kajino, M., Zhao, Y., Cubison, M. J.,
573 Jimenez, J. L., Vay, S., Diskin, G. S., Anderson, B., Wisthaler, A., Mikoviny, T., Fuelberg, H. E.,
574 Blake, D. R., Huey, G., Weinheimer, A. J., Knapp, D. J. and Brune, W. H.: Emissions of black
575 carbon, organic, and inorganic aerosols from biomass burning in North America and Asia in 2008,
576 *J. Geophys. Res.*, 116(D8), D08204, doi:10.1029/2010JD015152, 2011.

577 Jolleys, M. D., Coe, H., McFiggans, G., Taylor, J. W., O’Shea, S. J., Le Breton, M., Bauguitte, S. J.-B.,
578 Moller, S., Di Carlo, P., Aruffo, E., Palmer, P. I. and Lee, J. D.: Properties and evolution of
579 biomass burning organic aerosol from Canadian boreal forest fires, *Atmos. Chem. Phys. Discuss.*,
580 14(18), 25095–25138, doi:10.5194/acpd-14-25095-2014, 2014.

581 Lee, L. A., Pringle, K. J., Reddington, C. L., Mann, G. W., Stier, P., Spracklen, D. V., Pierce, J. R. and
582 Carslaw, K. S.: The magnitude and causes of uncertainty in global model simulations of cloud
583 condensation nuclei, *Atmos. Chem. Phys.*, 13(17), 8879–8914, doi:10.5194/acp-13-8879-2013,
584 2013.

585 Lee, S., Kim, H. K., Yan, B., Cobb, C. E., Hennigan, C., Nichols, S., Chamber, M., Edgerton, E. S.,
586 Jansen, J. J., Hu, Y., Zheng, M., Weber, R. J. and Russell, A. G.: Diagnosis of aged prescribed
587 burning plumes impacting an urban area., *Environ. Sci. Technol.*, 42(5), 1438–44, 2008.

588 Leppä, J., Mui, W., Grantz, A. M. and Flagan, R. C.: OP10-5, Importance of charger ion properties in

589 size distribution measurements, International Aerosol Conference, August 28 - September 2,
590 BEXCO, Busan, Korea, 2014.

591 Levin, E. J. T., McMeeking, G. R., Carrico, C. M., Mack, L. E., Kreidenweis, S. M., Wold, C. E.,
592 Moosmüller, H., Arnott, W. P., Hao, W. M., Collett, J. L. and Malm, W. C.: Biomass burning
593 smoke aerosol properties measured during Fire Laboratory at Missoula Experiments (FLAME), J.
594 Geophys. Res., 115(D18), D18210, doi:10.1029/2009JD013601, 2010.

595 López-Yglesias, X., and Flagan, R. C.: Ion–Aerosol Flux Coefficients and the Steady-State Charge
596 Distribution of Aerosols in a Bipolar Ion Environment, Aerosol Sci. Technol., 47, 688-704,
597 10.1080/02786826.2013.783684, 2013.

598 May, A. A., Levin, E. J. T., Hennigan, C. J., Riipinen, I., Lee, T., Collett, J. L., Jimenez, J. L.,
599 Kreidenweis, S. M. and Robinson, A. L.: Gas-particle partitioning of primary organic aerosol
600 emissions: 3. Biomass burning, J. Geophys. Res. Atmos., 118(19), 11,327–11,338,
601 doi:10.1002/jgrd.50828, 2013.

602 McMeeking, G. R., Kreidenweis, S. M., Baker, S., Carrico, C. M., Chow, J. C., Collett, J. L., Hao, W.
603 M., Holden, A. S., Kirchstetter, T. W., Malm, W. C., Moosmüller, H., Sullivan, A. P. and Wold, C.
604 E.: Emissions of trace gases and aerosols during the open combustion of biomass in the laboratory,
605 J. Geophys. Res., 114(D19), D19210, doi:10.1029/2009JD011836, 2009.

606 Myhre, G., D. Shindell, F.-M. Bréon, W. Collins, J. Fuglestedt, J. Huang, D. Koch, J.-F. Lamarque, D.
607 Lee, B. Mendoza, T. Nakajima, A. Robock, G. Stephens, T. Takemura and H. Zhang:
608 Anthropogenic and Natural Radiative Forcing. In: *Climate Change 2013: The Physical Science*
609 *Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental*
610 *Panel on Climate Change* [Stocker, T.F., D. Qin, G.-K. Plattner, M. Tignor, S.K. Allen, J.
611 Boschung, A. Nauels, Y. Xia, V. Bex and P.M. Midgley (eds.)]. Cambridge University Press,
612 Cambridge, United Kingdom and New York, NY, USA, 659-740, 2013.

613 Okoshi, R., Rasheed, A., Chen Reddy, G., McCrowey, C. J. and Curtis, D. B.: Size and mass
614 distributions of ground-level sub-micrometer biomass burning aerosol from small wildfires,
615 Atmos. Environ., 89, 392–402, doi:10.1016/j.atmosenv.2014.01.024, 2014.

616 Ontario Ministry of Natural Resources and Forestry. 2011 Forest Fire Summary.
617 <https://www.ontario.ca/law-and-safety/forest-fires>. Accessed June 18, 2012.

618 Ortega, A. M., Day, D. A., Cubison, M. J., Brune, W. H., Bon, D., de Gouw, J. A. and Jimenez, J. L.:
619 Secondary organic aerosol formation and primary organic aerosol oxidation from biomass-burning
620 smoke in a flow reactor during FLAME-3, Atmos. Chem. Phys., 13(22), 11551–11571,
621 doi:10.5194/acp-13-11551-2013, 2013.

622 Palmer, P. I., Parrington, M., Lee, J. D., Lewis, A. C., Rickard, A. R., Bernath, P. F., Duck, T. J., Waugh,
623 D. L., Tarasick, D. W., Andrews, S., Aruffo, E., Bailey, L. J., Barrett, E., Bauguitte, S. J.-B., Curry,
624 K. R., Di Carlo, P., Chisholm, L., Dan, L., Forster, G., Franklin, J. E., Gibson, M. D., Griffin, D.,
625 Helmig, D., Hopkins, J. R., Hopper, J. T., Jenkin, M. E., Kindred, D., Kliever, J., Le Breton, M.,
626 Matthiesen, S., Maurice, M., Moller, S., Moore, D. P., Oram, D. E., O'Shea, S. J., Owen, R. C.,
627 Pagniello, C. M. L. S., Pawson, S., Percival, C. J., Pierce, J. R., Punjabi, S., Purvis, R. M.,
628 Remedios, J. J., Rotermund, K. M., Sakamoto, K. M., da Silva, A. M., Strawbridge, K. B., Strong,
629 K., Taylor, J., Trigwell, R., Tereszchuk, K. A., Walker, K. A., Weaver, D., Whaley, C., and Young,
630 J. C.: Quantifying the impact of BOREal forest fires on Tropospheric oxidants over the Atlantic
631 using Aircraft and Satellites (BORTAS) experiment: design, execution and science overview,

- 632 Atmos. Chem. Phys., 13, 6239-6261, doi:10.5194/acp-13-6239-2013, 2013.
- 633 Parrington, M., Palmer, P. I., Lewis, A. C., Lee, J. D., Rickard, A. R., Di Carlo, P., Taylor, J. W.,
634 Hopkins, J. R., Punjabi, S., Oram, D. E., Forster, G., Aruffo, E., Moller, S. J., Bauguitte, S. J.-B.,
635 Allan, J. D., Coe, H. and Leigh, R. J.: Ozone photochemistry in boreal biomass burning
636 plumes Parrington, M., Palmer, P. I., Lewis, A. C., Lee, J. D., Rickard, A. R., Di Carlo, P., ...
637 Leigh, R. J. (2013). Ozone photochemistry in boreal biomass burning plumes. Atmospheric
638 Chemistry and Physics, 13, Atmos. Chem. Phys., 13(15), 7321–7341,
639 doi:10.5194/acp-13-7321-2013, 2013.
- 640 Parrish, D. D., Stohl, A., Forster, C., Atlas, E. L., Blake, D. R., Goldan, P. D., Kuster, W. C. and de
641 Gouw, J. A.: Effects of mixing on evolution of hydrocarbon ratios in the troposphere, J. Geophys.
642 Res., 112(D10), D10S34, doi:10.1029/2006JD007583, 2007.
- 643 Petters, M. D. and Kreidenweis, S. M.: A single parameter representation of hygroscopic growth and
644 cloud condensation nucleus activity, Atmos. Chem. Phys., 7(8), 1961–1971,
645 doi:10.5194/acp-7-1961-2007, 2007.
- 646 Petters, M. D., Carrico, C. M., Kreidenweis, S. M., Prenni, A. J., DeMott, P. J., Collett, J. L. and
647 Moosmüller, H.: Cloud condensation nucleation activity of biomass burning aerosol, J. Geophys.
648 Res., 114(D22), D22205, doi:10.1029/2009JD012353, 2009.
- 649 Pierce, J. R., Chen, K. and Adams, P. J.: Contribution of primary carbonaceous aerosol to cloud
650 condensation nuclei: processes and uncertainties evaluated with a global aerosol microphysics
651 model, Atmos. Chem. Phys., 7, 5447–5466, 2007.
- 652 Reddington, C. L., Carslaw, K. S., Spracklen, D. V., Frontoso, M. G., Collins, L., Merikanto, J.,
653 Minikin, A., Hamburger, T., Coe, H., Kulmala, M., Aalto, P., Flentje, H., Plass-Dülmer, C.,
654 Birmili, W., Wiedensohler, A., Wehner, B., Tuch, T., Sonntag, A., O'Dowd, C. D., Jennings, S. G.,
655 Dupuy, R., Baltensperger, U., Weingartner, E., Hansson, H.-C., Tunved, P., Laj, P., Sellegri, K.,
656 Boulon, J., Putaud, J.-P., Gruening, C., Swietlicki, E., Roldin, P., Henzing, J. S., Moerman, M.,
657 Mihalopoulos, N., Kouvarakis, G., Ždímal, V., Zíková, N., Marinoni, A., Bonasoni, P. and Duchi,
658 R.: Primary versus secondary contributions to particle number concentrations in the European
659 boundary layer, Atmos. Chem. Phys., 11(23), 12007–12036, doi:10.5194/acp-11-12007-2011,
660 2011.
- 661 Reid, J. S., Hobbs, P. and Ferek, R.: Physical, chemical, and optical properties of regional hazes
662 dominated by smoke in Brazil, J. Geophys. Res., 103(98), 32059–32080, DOI:
663 10.1029/98JD00458, 1998.
- 664 Reid, J. S., Hyer, E. J., Prins, E. M., Westphal, D. L., Zhang, J., Wang, J., Christopher, S. A., Curtis, C.
665 A., Schmidt, C. C., Eleuterio, D. P., Richardson, K. A. and Hoffman, J. P.: Global Monitoring and
666 Forecasting of Biomass-Burning Smoke: Description of and Lessons From the Fire Locating and
667 Modeling of Burning Emissions (FLAMBE) Program, IEEE J. Sel. Top. Appl. Earth Obs. Remote
668 Sens., 2(3), 144–162, doi:10.1109/JSTARS.2009.2027443, 2009.
- 669 Reid, J. S., Koppmann, R., Eck, T. F. and Eleuterio, D. P.: A review of biomass burning emissions part
670 II: intensive physical properties of biomass burning particles, Atmos. Chem. Phys., 5(3), 799–825,
671 doi:10.5194/acp-5-799-2005, 2005.
- 672 Rissler, J., Vestin, A., Swietlicki, E., Fisch, G., Zhou, J., Artaxo, P. and Andreae, M. O.: Size
673 distribution and hygroscopic properties of aerosol particles from dry-season biomass burning in

674 Amazonia, *Atmos. Chem. Phys.*, 6, 471–491, 2006.

675 Seinfeld, J. H. and Pandis, S. N.: Atmospheric chemistry and physics: from air pollution to climate
676 change, Hoboken, N. J., Wiley, 2006.

677 Spracklen, D. V., Carslaw, K. S., Pöschl, U., Rap, A. and Forster, P. M.: Global cloud condensation
678 nuclei influenced by carbonaceous combustion aerosol, *Atmos. Chem. Phys.*, 11(17), 9067–9087,
679 doi:10.5194/acp-11-9067-2011, 2011.

680 Staudt, A. C., Jacob, D. J., Logan, J. A. and Sachse, G. W.: Continental sources , transoceanic
681 transport , and interhemispheric exchange of carbon, *J. Geophys. Res.*, 106, 32571–32589, DOI:
682 10.1029/2001JD900078, 2001.

683 Taylor, J. W., Allan, J. D., Allen, G., Coe, H., Williams, P. I., Flynn, M. J., Le Breton, M., Muller, J. B.
684 A., Percival, C. J., Oram, D., Forster, G., Lee, J. D., Rickard, A. R. and Palmer, P. I.:
685 Size-dependent wet removal of black carbon in Canadian biomass burning plumes, *Atmos. Chem.*
686 *Phys. Discuss.*, 14(13), 19469–19513, doi:10.5194/acpd-14-19469-2014, 2014.

687 Van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M., Kasibhatla, P. S., Morton, D.
688 C., DeFries, R. S., Jin, Y. and van Leeuwen, T. T.: Global fire emissions and the contribution of
689 deforestation, savanna, forest, agricultural, and peat fires (1997–2009), *Atmos. Chem. Phys.*,
690 10(23), 11707–11735, doi:10.5194/acp-10-11707-2010, 2010.

691 Wiedensohler, A.: An Approximation of the Bipolar Charge-Distribution for Particles in the
692 Sub-Micron Size Range, *J. Aerosol. Sci.*, 19, 387-389, 1988.

693 Wiedinmyer, C., Akagi, S. K., Yokelson, R. J., Emmons, L. K., Al-Saadi, J. A., Orlando, J. J. and Soja,
694 A. J.: The Fire INventory from NCAR (FINN): a high resolution global model to estimate the
695 emissions from open burning, *Geosci. Model Dev.*, 4(3), 625–641, doi:10.5194/gmd-4-625-2011,
696 2011.

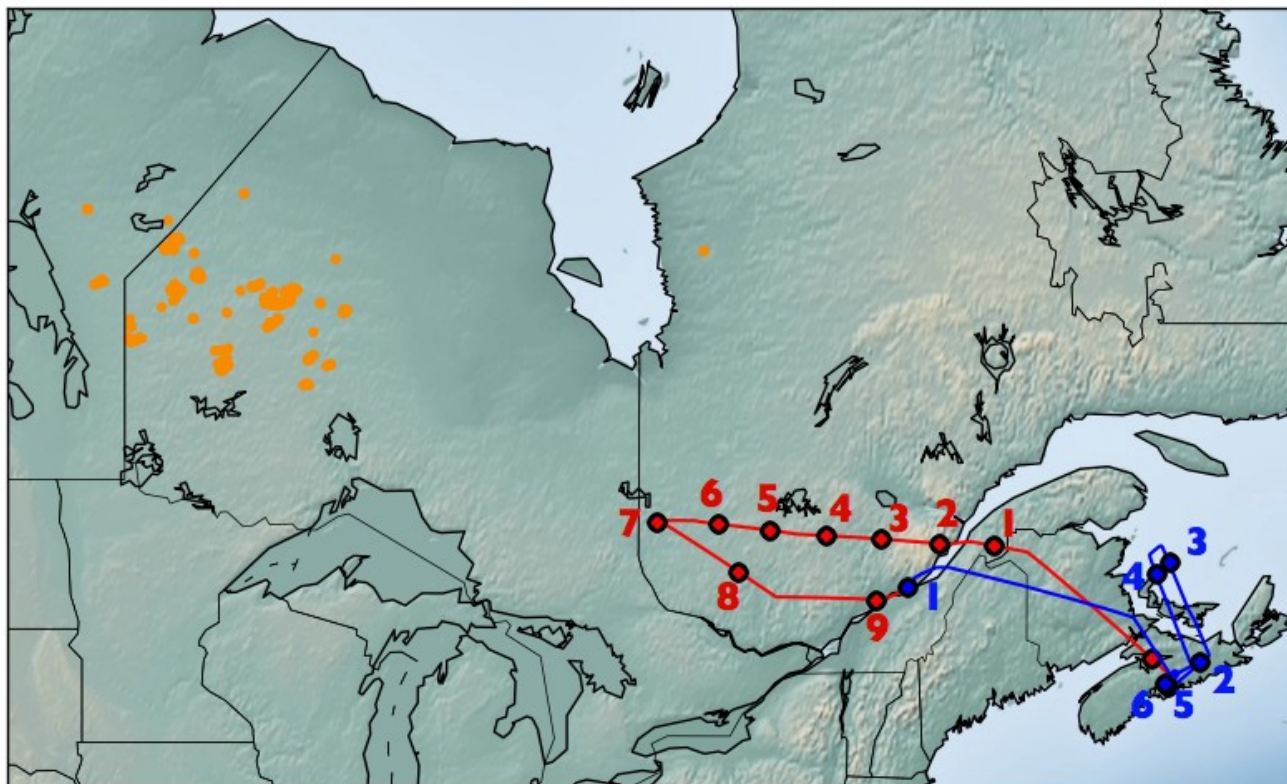
697 Yokelson, R. J., Crounse, J. D., Decarlo, P. F., Karl, T., Urbanski, S., Atlas, E., Campos, T. and
698 Shinozuka, Y.: Emissions from biomass burning in the Yucatan, *Atmos. Chem. Phys.*, 5785–5812,
699 2009.

700

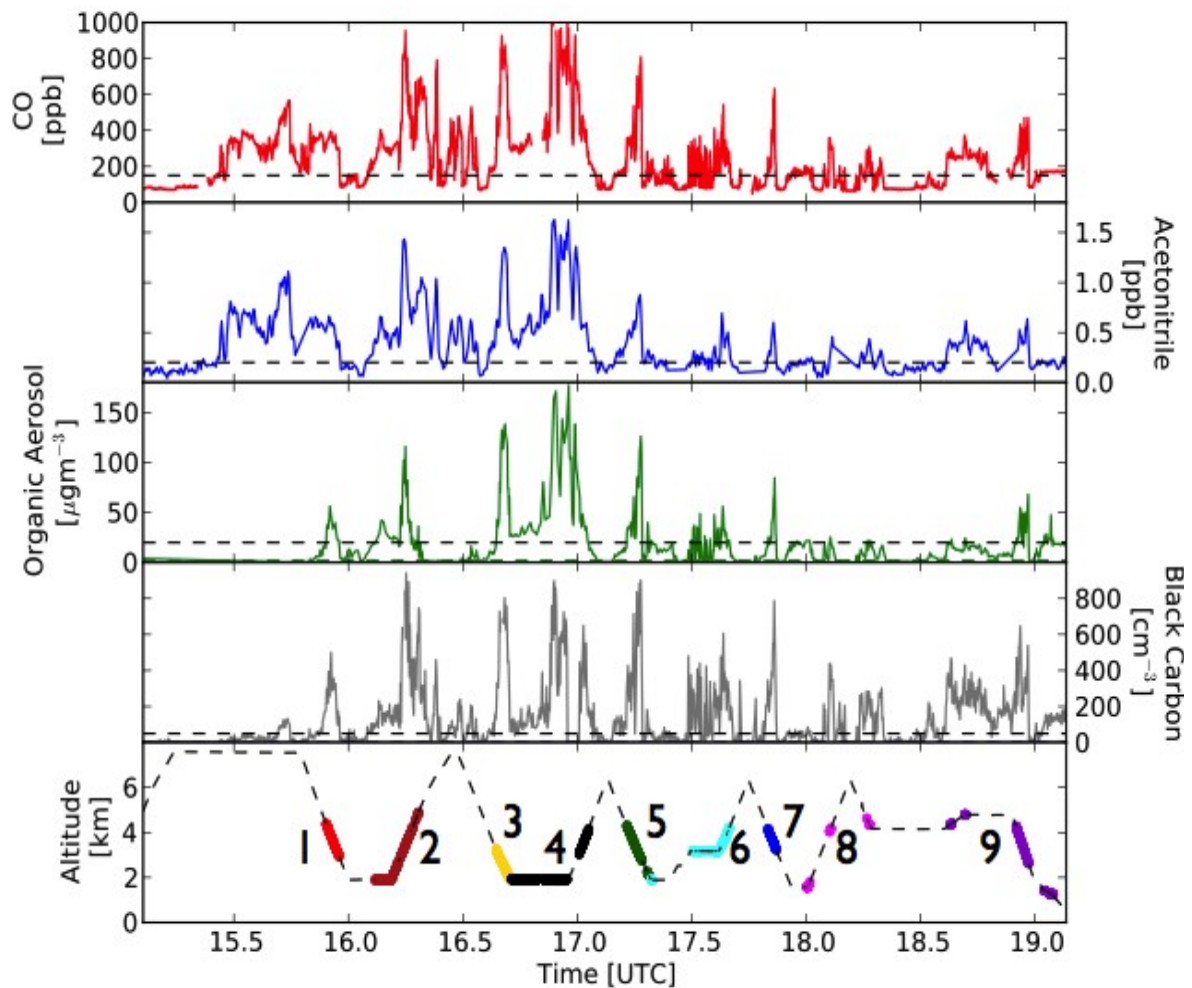
701 **Table 1.** Approximate physical transport age and distance of numbered flight transect midpoints from
 702 source fires. Ages were estimated by HYSPLIT back trajectories. The large ranges in the determined
 703 values are due to the large extent of the source fire region and variability in fire conditions. The
 704 distances are given from transect midpoints to the source fire region (± 150 km).
 705

Transect #	Approx. physical Age [hrs]	Approx. distance from source [km]
<i>Flight b622</i>		
1	27 - 32	1450 - 1750
2	27 - 32	1350 - 1650
3	27 - 32	1250 - 1550
4	27 - 32	1150 - 1450
5	18 - 25	1050 - 1350
6	18 - 25	850 - 1150
7	18 - 25	850 - 1150
8	18 - 25	1050 - 1350
9	24 - 30	1350 - 1650
<i>Flight b623</i>		
1	24 - 36	1350 - 1650
2	28 - 36	2050 - 2350
3 - 4	28 - 36	1850 - 2150
5 - 6	28 - 36	1950 - 2250

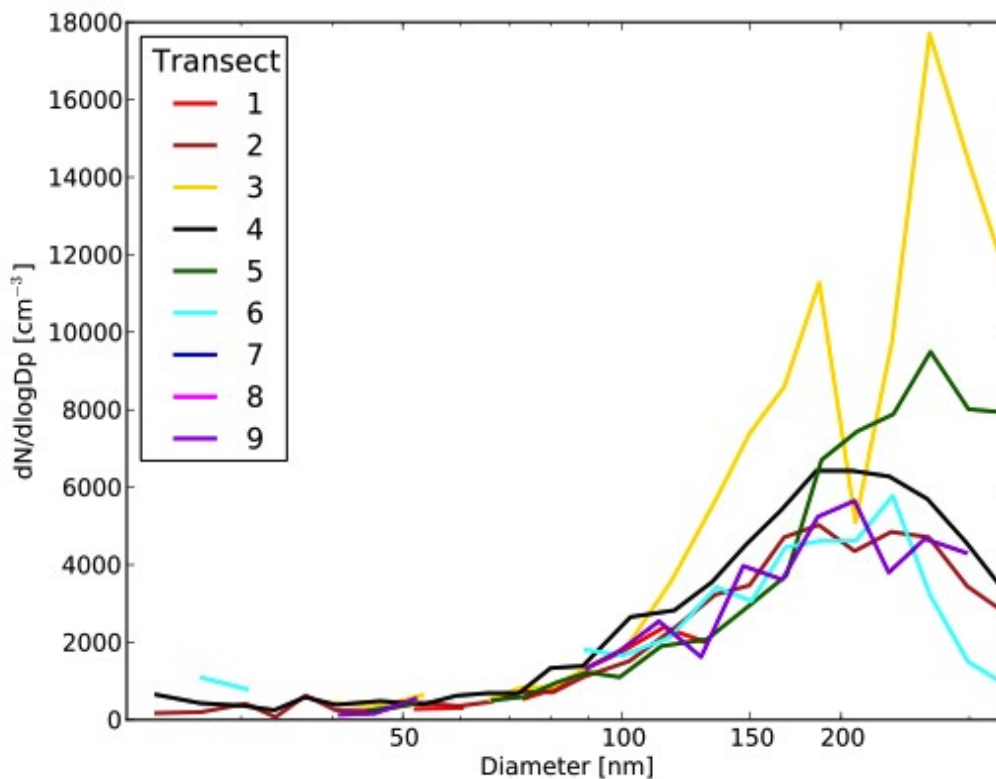
706



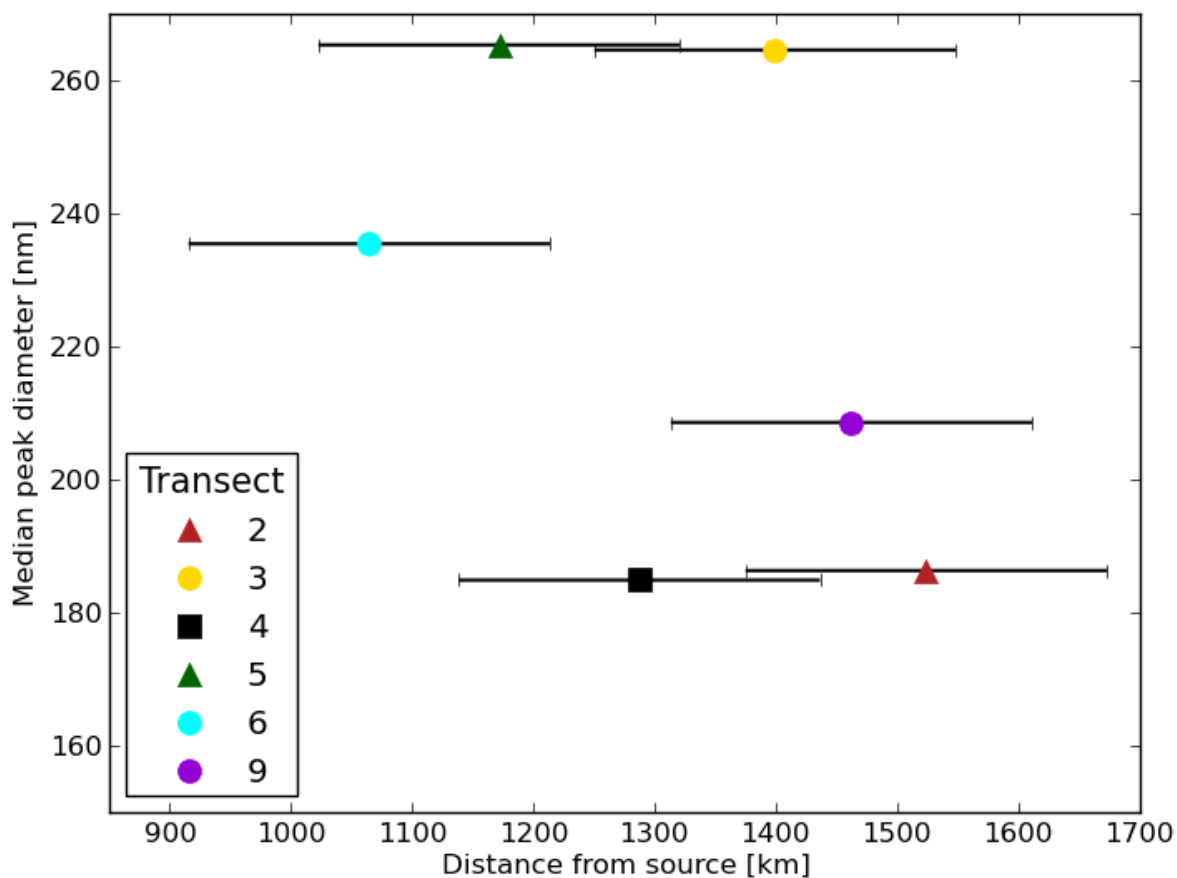
708 **Figure 1.** BORTAS-B ARA research flights b622 (red) from Nova Scotia to Quebec, and the return
709 flight b623 (blue) both on July 20-21, 2011. Circles represent midpoints of ascent/descent transects
710 along the flight paths. The ARA flew through biomass-burning emissions originating from fires in
711 Northwestern Ontario. The July 17-20, 2011 MODIS hotspot fires (fire radiative power >100 MW) are
712 plotted in orange.



714 **Figure 2.** Time series of BORTAS-B aircraft measurements of biomass-burning tracer species for
 715 Flight b622. Threshold values (dashed black lines) were used across four species as plume criteria: i)
 716 CO (red, threshold = 150 ppb), ii) Acetonitrile (blue, threshold = 200 pptv), iii) Organic aerosols
 717 (green, mass threshold= 20 $\mu\text{g m}^{-3}$, at STP), iv) Black carbon (grey, number threshold = 50 cm^{-3} , at
 718 STP). The bottom panel shows flight altitude with plume sampling periods coloured. The plume data is
 719 further divided into transects (1-9 in red-violet).



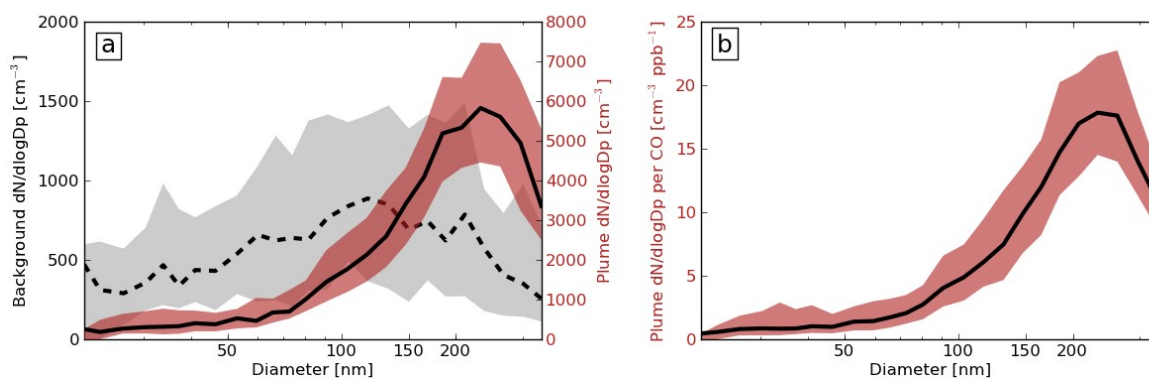
721 **Figure 3.** Median plume number size distributions (corrected to cm^{-3} at STP) divided by transect for
 722 Flight b622. All size distributions show a consistent accumulation mode with $D_{\text{pm}} \sim 220$ nm. Size bins
 723 with less than three data points in any transect are not shown, limiting the contributions from transects
 724 1, 7 and 8. The composite plume size-distribution for both Flight b622 and b623 is seen in Figure 5.



725 **Figure 4.** Accumulation mode peak diameter by transect (2-6, 9) showing no significant trend with
 726 plume transport distance. All colours are the same as in Figure 3. Distance from fire sources was
 727 estimated using transect midpoints and approximate source region area. Transects 1, 7 and 8 have
 728 insufficient accumulation mode plume data and have been omitted. The uncertainty bars show
 729 uncertainty in the distance from the source (± 150 km).

730
 731
 732
 733
 734
 735
 736
 737
 738
 739
 740
 741
 742
 743
 744

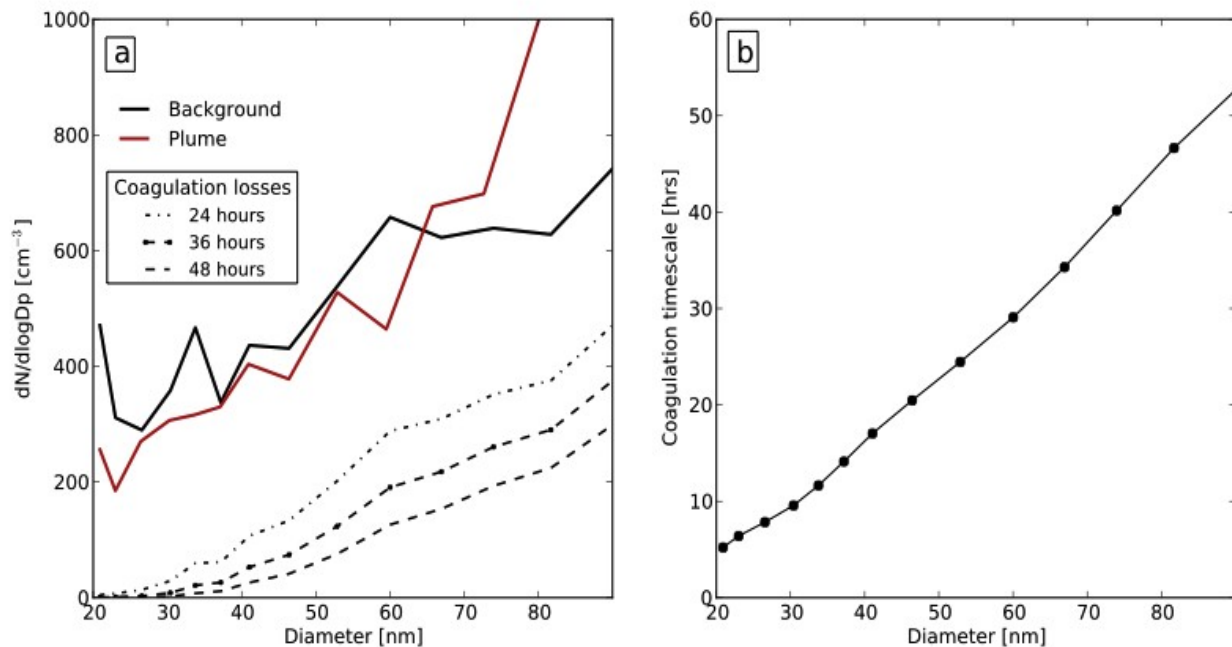
745



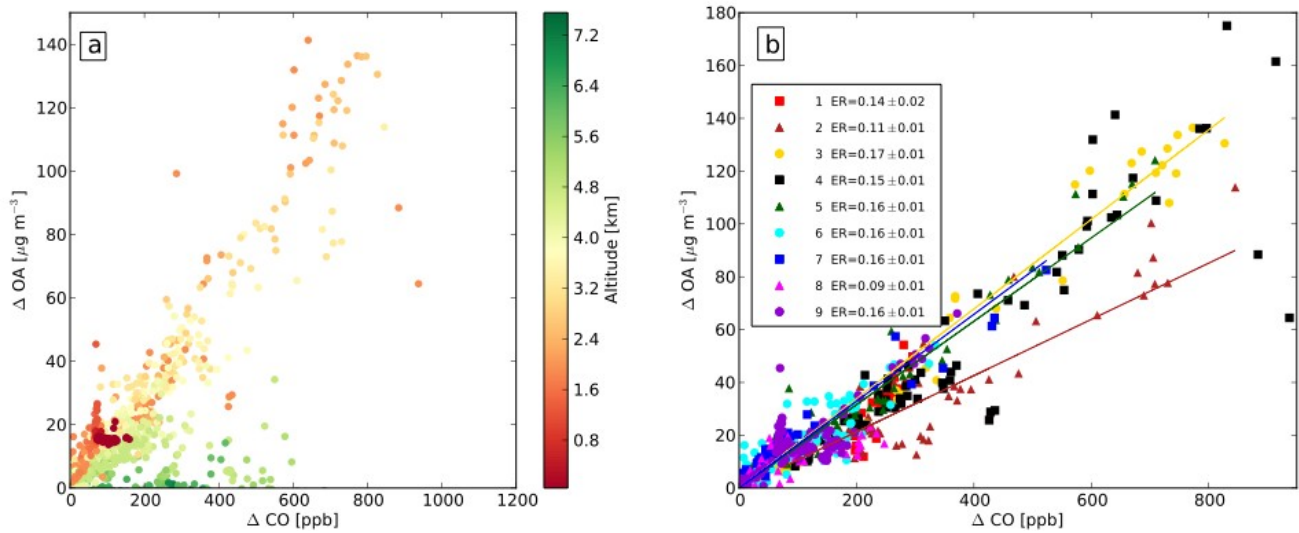
747

748

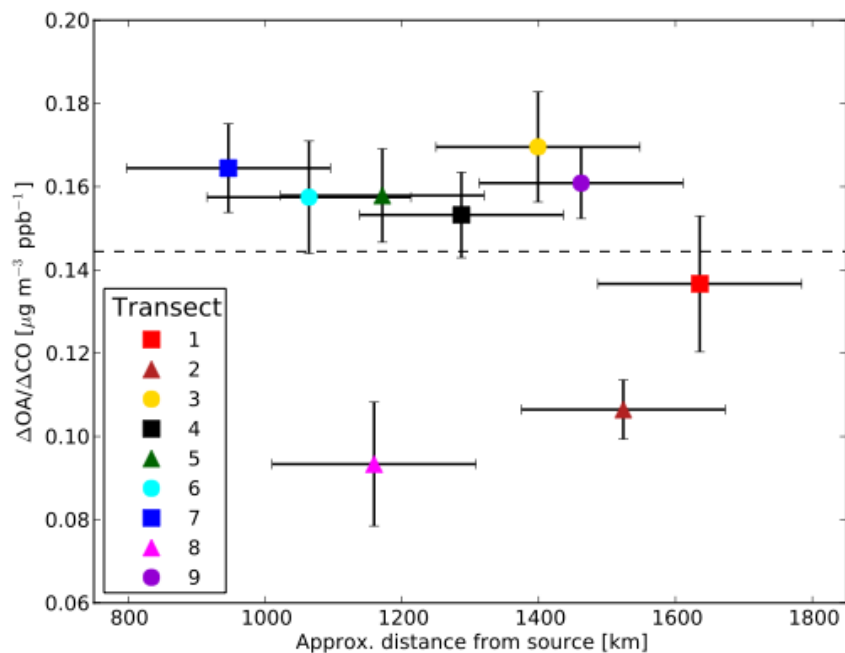
749 **Figure 5.** Composite median number size distributions for Flights b622 and b623 (cm⁻³ at STP). The
750 in-plume (red) and background (grey) air distributions are shown as absolute concentrations (5a). The
751 in-plume distributions are also normalized by CO mixing ratio (5b). The black lines are the median
752 with the 25th and 75th percentiles overlain. The plume distributions have $D_{pm}=230$ nm.



754 **Figure 6.** Figure 6a shows background (black solid line) and plume (red line) median concentrations
 755 for small particle diameters (20-90 nm). The black dashed lines are the number distributions after 24,
 756 36 and 48 hours of coagulation losses by the plume accumulation mode (Figure 5a) from the
 757 background level concentrations. These calculated concentrations are much lower than those found in
 758 plume. Figure 6b shows the coagulation lifetime as a function of particle diameter (on the order of 10s
 759 of hours in this diameter range).

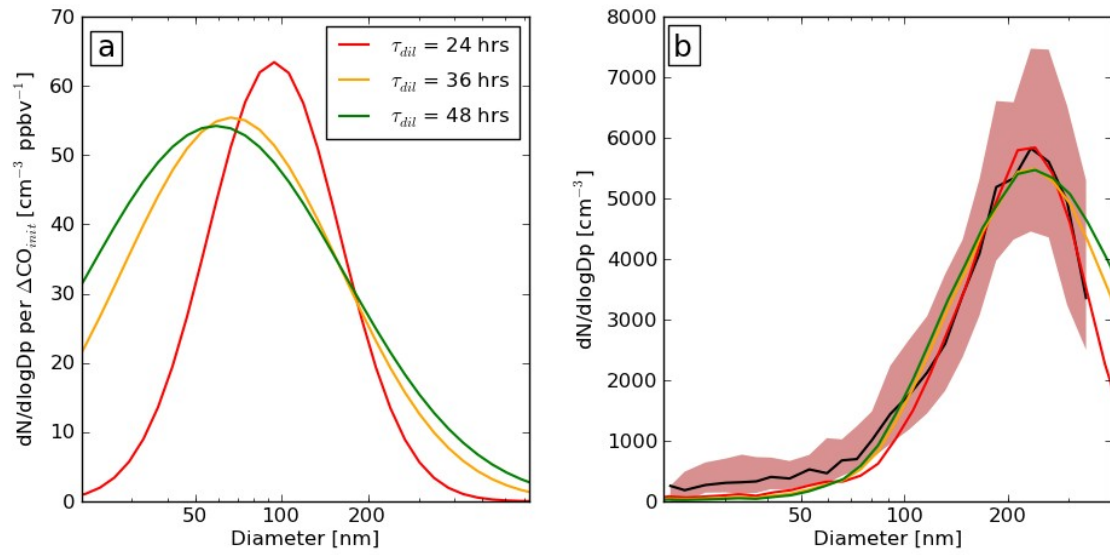


761 **Figure 7.** Enhancement ratios of $\Delta OA/\Delta CO$ for Flight b622. Figure 7a is coloured by altitude showing
 762 potential aerosol washout in the high-altitude plume (>4.6 km). Figure 7b shows the ERs separated by
 763 flight transect showing individual enhancement ratios of between 0.09 - $0.17 \pm 0.01 \mu g m^{-3} ppb^{-1}$ with
 764 generally high correlation coefficients ($R^2 > 0.7$) for the majority. The data points collected at altitudes
 765 greater than 4.6 km have been removed (as per 7a).

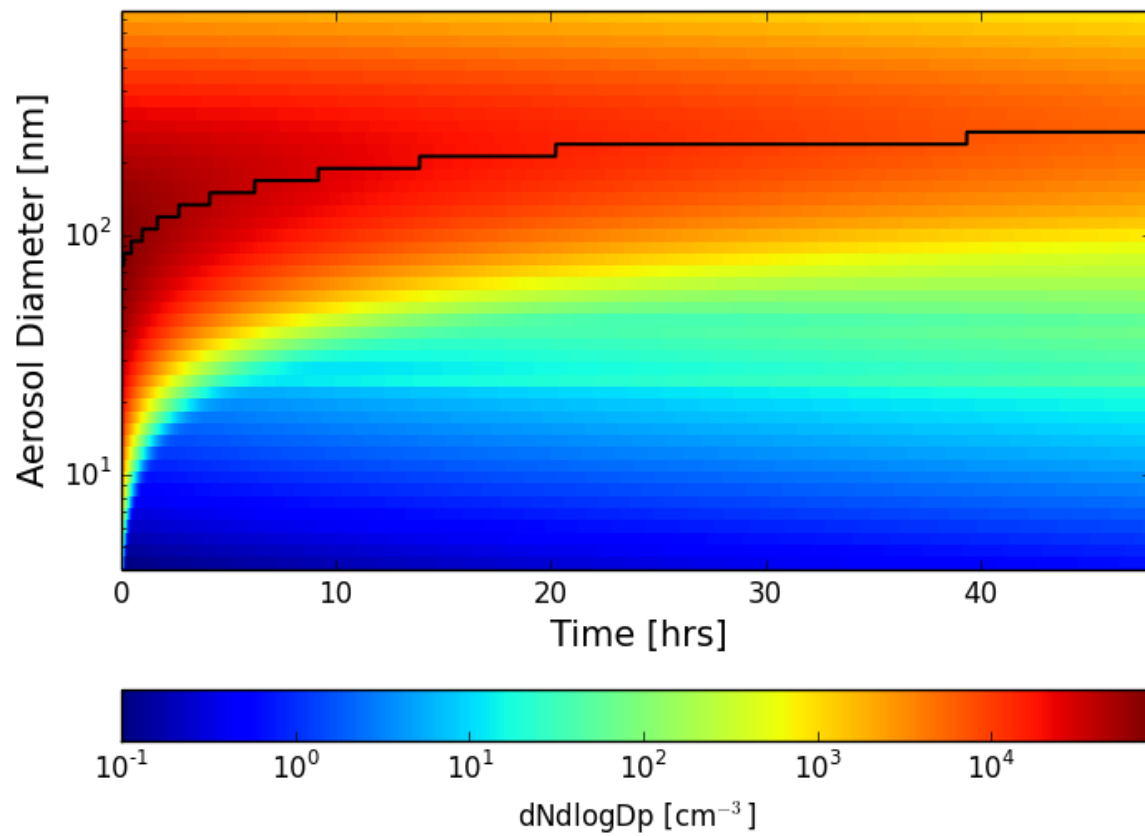


791
792
793
794
795
796
797
798
799
800
801
802
803
804
805
806
807
808

809 **Figure 8.** Transect $\Delta\text{OA}/\Delta\text{CO}$ enhancement ratios for Flight b622 as a function of the distance from the
810 source fire region. The average ER is represented by the dashed black line ($0.144 \mu\text{g m}^{-3} \text{ppb}^{-1}$). There
811 is no discernible trend in ΔOA enhancement either by distance (x-axis) or time (colours). The
812 uncertainty bars display the uncertainty in distance and in fitted enhancement ratios.



814 **Figure 9.** Figure 9a shows the optimized young-plume size-distributions for entrainment parameters
 815 $\tau_{dil} = 24, 36, 48$ hrs as particle concentration per ΔCO_{init} [cm^{-3} ppbv $^{-1}$]. Figure 9b shows the final
 816 modelled size-distributions compared to the measured aged plume size-distribution (black median, red
 817 quartiles).



819 **Figure 10.** Plot of modelled size-distribution evolution for $\tau_{dil} = 36$ hrs. The black line shows the peak
 820 diameter at each timestep ($\Delta t = 10$ s). The young-plume size-distribution has optimal initial parameters:
 821 $D_{pm} = 67$ nm, $\sigma = 2.4$, $N_0 = 80,000$ cm^{-3} .