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2 **A comparison study between CMAQ-simulated and OMI-**
3 **retrieved NO₂ columns over East Asia for evaluation of**
4 **NO_x emission fluxes of INTEX-B, CAPSS, and REAS**
5 **inventories**
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13 **Shortened title:** Tropospheric NO₂ columns over East Asia
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41 **Abstract**

42 Comparison between the CMAQ-calculated and OMI-retrieved tropospheric NO₂ columns
43 was carried out for 2006 over East Asia (100°-150°E; 20°-50°N) to evaluate the bottom-up
44 NO_x emission fluxes of INTEX-B, CAPSS, and REAS v1.11 inventories. The three emission
45 inventories were applied to the CMAQ model simulations for the countries of China, Korea,
46 and Japan, respectively. For the direct comparison between the two NO₂ columns, the
47 averaging kernels (AKs) obtained from the Royal Netherlands Meteorological Institute
48 (KNMI)/DOMINO v2.0 daily product were applied to the CMAQ-simulated data. The
49 analysis showed that the two tropospheric NO₂ columns from the CMAQ model simulations
50 and OMI observations ($\Omega_{\text{CTM,AK}}$ and Ω_{OMI}) had good spatial and seasonal correlation, with
51 correlation coefficients ranging from 0.71 to 0.96. In addition, the normalized mean errors
52 (NMEs) between the $\Omega_{\text{CTM,AK}}$ and Ω_{OMI} were found to range from ~40% to ~63%. The
53 $\Omega_{\text{CTM,AK}}$ were, on annual average, ~28% smaller (in terms of the NMEs) than the Ω_{OMI} ,
54 indicating that the NO_x emissions used were possibly underestimated in East Asia. Large
55 absolute differences between the $\Omega_{\text{CTM,AK}}$ and Ω_{OMI} were found, particularly over Central East
56 China (CEC) during winter (annual averaged mean error of $\sim 4.51 \times 10^{15}$ molecules cm⁻²).
57 Although such differences between the $\Omega_{\text{CTM,AK}}$ and Ω_{OMI} are likely caused by the errors and
58 biases in the NO_x emissions used in the CMAQ model simulations, it can be rather difficult to
59 directly and quantitatively relate the differences to the accuracy of the NO_x emissions,
60 because there are also several uncertain factors in the CMAQ model, satellite-retrieved NO₂
61 columns and AK products, and NO_x and other trace gas emissions. Therefore, in this study
62 three uncertain factors were selected and analyzed with sensitivity runs (monthly variations in
63 NO_x emissions; influences of different NO_x emission fluxes; and reaction probability of N₂O₅
64 radicals). Other uncertain or possible influential factors were also discussed to suggest future
65 direction of the study.

66 **Keywords:** Tropospheric NO₂ columns; Averaging Kernels; OMI sensor; CMAQ model;
67 Bottom-up NO_x emissions
68

69 **1. Introduction**

70 There has been growing public concern about serious smog events in East Asia due to
71 large amounts of anthropogenic pollutants in the atmosphere. Among the pollutants, nitrogen
72 oxides (NO_x \cong NO + NO₂) play a key role in tropospheric chemistry, such as ozone and
73 secondary aerosol formation. Also, in global climate change, atmospheric NO_x is believed to
74 make indirect negative contributions to radiative forcing in the atmosphere (Wild et al., 2001).
75 For example, secondary nitrates (NO₃⁻) formed via the condensation of atmospheric HNO₃,
76 NO₃, and N₂O₅ into particles contribute, on average, 30.7% to aerosol direct radiative forcing
77 (ADRF) in East Asia during the winter season, which cannot be ignored in the estimation of
78 direct radiative forcing in **East Asia** (Park et al., 2014). HNO₃ formation via the reaction of
79 OH + NO₂ during the daytime and heterogeneous nitrate formation via the condensation of
80 N₂O₅ onto atmospheric particles during the nighttime are believed to be the main chemical
81 and physico-chemical processes removing NO_x from the atmosphere (**McConnell and**
82 **McElroy, 1973; Platt et al., 1984; Dentener and Crutzen, 1993; Brown et al., 2006; Han and**
83 **Song, 2012).**

84 Recently, several studies have reported annual increases in NO_x emissions in China
85 (Zhang et al., 2007; Zhang et al., 2009; Kurokawa et al., 2013). For example, **according to the**
86 **Greenhouse gas and Air pollution INteractions and Synergies (GAINS) model simulations,**
87 **China makes the largest contribution to global NO_x emissions, and its contribution was**
88 **estimated to be 25% for 2010 (Cofala et al., 2012). Also, when several emissions scenarios**
89 **are applied to the GAINS simulations, the contribution of China is estimated to increase, to**
90 **~29% in the years between 2015 and 2035 (Cofala et al., 2012). However, large uncertainty in**
91 **bottom-up NO_x emissions over East Asia has been reported (e.g. Streets et al., 2003; Zhang et**

92 al., 2007; Klimont et al., 2009; Xing et al., 2011).

93 In the meantime, several studies have also reported rapid increases in atmospheric
94 **NO₂ columns** over China, based on Global Ozone Monitoring Experiment (GOME), Ozone
95 Monitoring Instrument (OMI), and SCanning Imaging Absorption spectroMeter for
96 Atmospheric **CartograpHY** (SCIAMACHY) observations (Richter et al., 2005; van der A et
97 al., 2006; **Schneider and van der A, 2012; Hilboll et al., 2013; Itahashi et al., 2014**). These
98 satellite observations have provided useful global/regional information on the spatial
99 distributions of **NO₂ columns**, and have also been used to investigate the accuracy of the
100 global and regional NO_x emissions (e.g. Martin et al., 2006; Uno et al., 2007; Wang et al.,
101 2007; Han et al., 2009).

102 However, these satellite **observations** are not “real” or “true” values, having different
103 vertical sensitivities at different altitudes in the atmosphere. To **consider** this vertical
104 sensitivity of the satellite observations, averaging kernels (AKs) **should be** introduced into
105 comparison studies between chemistry-transport model (CTM)-simulated and satellite-
106 retrieved **tropospheric** NO₂ columns (hereafter, denoted as Ω). **The introduction of AKs could**
107 **correct the large systematic errors typically caused by assumed (or unrealistic) NO₂ vertical**
108 **profiles used in the retrieval process of the NO₂ columns** (Rodgers, 2000; Eskes and Boersma,
109 2003). In particular, Eskes and Boersma (2003) reported that the use of AKs is crucial in
110 interpreting the retrieved Ω , because of the low sensitivity of satellite observations of NO₂
111 near the surface areas.

112 In this context, several studies have used AKs to evaluate the surface NO_x emissions
113 over several regions (e.g. Herron-Thorpe et al., 2010; Lamsal et al., 2010; **Huijnen et al., 2010;**
114 **Ghude et al., 2013; Zyrichidou et al., 2013**). The **previous studies conducted by Han et al.**
115 **(2009; 2011) also compared the CTM-calculated tropospheric NO₂ columns with GOME-**
116 **retrieved tropospheric NO₂ columns to evaluate the bottom-up NO_x emissions over East Asia,**

117 but without using the AKs. Based on the comparison, Han et al. (2011) concluded that the
118 bottom-up NO_x emissions used in CTM simulations over East Asia may be overestimated.
119 However, such comparison without the application of AKs is like comparing “apples” with
120 “oranges”, and is unreasonable. Therefore, one of the main objectives of this study was to
121 correct our previous conclusions, using the state-of-the-science knowledge and methods,
122 including the application of AKs to the CTM simulations. In this study, we intended to
123 evaluate three bottom-up NO_x emissions of INTEX-B, CAPSS, and REAS v1.11 inventories
124 in East Asia, using OMI-retrieved tropospheric NO_2 columns (Ω_{OMI}) from KNMI/DOMINO
125 v2.0 daily products and the CTM-calculated tropospheric NO_2 columns (Ω_{CTM}). To conduct
126 this investigation, the AKs obtained from the KNMI algorithm were applied, and then direct
127 comparison of the $\Omega_{\text{CTM,AK}}$ with Ω_{OMI} was carried out (refer to Sect. 3.1).

128 However, evaluation of the bottom-up NO_x emissions via comparison between
129 $\Omega_{\text{CTM,AK}}$ and Ω_{OMI} may be hampered by many uncertain factors such as: (i) uncertain temporal
130 variations in NO_x emissions in East Asia; (ii) uncertainty in the magnitudes of emission fluxes
131 of NO_x and other NO_x chemistry-related trace gases; (iii) uncertainty in meteorological fields;
132 (iv) uncertain or missing photo-chemistries in the CTM; and (v) errors in the retrieved NO_2
133 columns and AKs. Because of these errors and uncertainties, it can sometimes be difficult to
134 directly and quantitatively relate the differences between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} to the
135 accuracy of the NO_x emissions in East Asia. Some of these issues are therefore explored with
136 several sensitivity analyses, and other factors are also discussed in Sect. 3.2.

137

138 **2. Experimental Methods**

139 **2.1 Modeling descriptions**

140 First, for the CTM simulations, the US EPA/Models-3 CMAQ (Community Multi-
141 scale Air Quality) v4.7.1 model was used (Byun and Schere, 2006). To drive CMAQ model

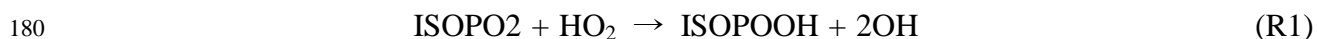
142 simulations, two main drivers are needed: (i) meteorological fields and (ii) emission fields.
143 For the former, PSU/NCAR MM5 (Pennsylvania state University/National Center for
144 Atmospheric Research Meso-scale Model 5) v3.7.1 was used with National Centers for
145 Environmental Prediction (NCEP) reanalyzed data sets (Stauffer and Seaman, 1990; 1994).
146 To prepare more accurate meteorological fields, four-dimensional data assimilation (FDDA)
147 using QuickSCAT 10-m wind data sets was also carried out. For the latter, three
148 anthropogenic emission inventories were used: INTEX-B (Intercontinental Chemical
149 Transport Experiment-Phase B, Zhang et al., 2009), CAPSS (Clean Air Policy Support
150 System, Hong et al., 2008), and REAS v1.11 (Regional Emission Inventory in Asia, Ohara et
151 al., 2007) emission inventories for the year 2006. Annual 0.5°×0.5°-resolved INTEX-B and
152 REAS v1.11 emissions were interpolated into the CMAQ grid cells in China and Japan,
153 respectively. For biogenic emissions, the MEGAN-ECMWF (Model of Emissions of Gases
154 and Aerosols from Nature—European Center for Medium-Range Weather Forecasts) inventory
155 was obtained from the official website, at <http://tropo.aeronomie.be/models/isoprene.htm>
156 (Müller et al., 2008). Biogenic emissions are an important factor during the summer, even in
157 this type of NO_x study, because the mixing ratios of biogenic species can influence the NO₂-
158 to-NO ratios via changing the levels of HO_x and RO₂ radicals (Horowitz et al., 2007; Han et
159 al., 2009). The accuracy of the biogenic emissions used in this study was also evaluated over
160 the same domain, East Asia, in our previous study (Han et al., 2013).

161 Table 1 summarizes base-case simulation and several sensitivity runs for this study.
162 For the base-case simulation, monthly variations of the anthropogenic NO_x emissions from
163 Zhang et al. (2009) were considered for China, while those from Han et al. (2009) were used
164 for Korea and Japan. The monthly factors were applied to the sectors of power generation,
165 residential areas, industry, and transportation. As shown in Fig. 1, data on several monthly
166 variations in NO_x emissions over China were available. Among them, two representative and

167 extreme monthly variations were chosen in this study, which were explored and discussed in
168 Sect. 3.2.1.

169 The modeling period was from January 1 to December 31, 2006. In this study, 2006
170 was chosen for the CMAQ model simulations, because the **INTEX-B inventory was compiled**
171 **for this year (the REAS v1.11 and CAPSS inventories were also chosen for 2006)**. The
172 horizontal domain covers from 100°E to 150°E and from 20°N to 50°N with a grid-resolution
173 of 30 km × 30 km. The vertical domain covers from 1000 hPa to 118 hPa with 14 terrain
174 following σ -coordinates. For considering aerosol dynamics and thermodynamics, the aerosol
175 module of AERO4 was selected (Binkowski and Roselle, 2003).

176 For the consideration of gas-phase chemistry, the SAPRAC-99 (Statewide Air
177 Pollution Research Center-99) mechanism was selected (Carter, 2000). Then, to consider
178 unknown OH radical processes (Lelieveld et al., 2008), the SAPRAC-99 mechanism was
179 modified partly, based on the work of Butler et al. (2008) in the following way (R1):



181 Here, ISOPO₂ and ISOPOOH represent isoprene-derived peroxy radical and peroxide,
182 respectively. Other schemes used in the CMAQ model simulations were the global mass-
183 conserving scheme (YAMO) for horizontal and vertical advection (Yamartino, 1993), the
184 asymmetric convective model (ACM) algorithm for convective cloud mixing, and ACM
185 (ver. 2) for vertical diffusion (Pleim, 2007).

186 In the CMAQ modeling, initial conditions (ICs) were prepared from 1 week-long
187 spin-up model simulations, and boundary conditions (BCs) were obtained from global CTM
188 simulations, MOZART (Model for OZone And Related chemical Tracers) (Emmons et al.,
189 2010). The MOZART model simulation data for the BCs were obtained from
190 <http://www.acd.ucar.edu/wrf-chem/mozart.shtml>. Other details about the **model setup** were
191 reported by Han et al. (2013).

192 For synchronization with the Ω_{OMI} , the Ω_{CMAQ} data were collected and then averaged
193 between 13:00 and 14:00 local time (LT), because the OMI sensor scans the atmosphere over
194 East Asia approximately at 13:45 LT. For further detailed analyses, eight highly-populated
195 focus regions were defined in this study, and are presented in Fig. 2.

196

197 **2.2 OMI-retrieved NO₂ columns and AKs**

198 The OMI instrument on board the NASA/EOS–Aura satellite, a nadir-viewing
199 imaging spectrometer, provides information on the properties of aerosols and clouds as well
200 as global levels of atmospheric species such as ozone, NO₂, SO₂, OCIO, BrO, and HCHO on
201 a daily basis via observing backscattered UV-VIS radiances from 270 to 550 nm (Levelt et al.,
202 2006). Two-dimensional charge-coupled device (CCD) detectors equipped in the OMI
203 instrument observe the atmosphere with a spatial resolution of 13 km × 24 km **at the nadir**.
204 CCD1 covers the UV channel of 270-310 nm and 310-365 nm. The visible channel, ranging
205 from 365 to 500 nm, is covered by CCD2 to observe NO₂.

206 In this study, daily levels of OMI-retrieved tropospheric NO₂ columns from
207 KNMI/DOMINO v2.0 products were used (Boersma et al., 2007; 2011a). The
208 KNMI/DOMINO v2.0 algorithm (hereafter, KNMI algorithm) for retrieving the tropospheric
209 NO₂ columns from the OMI radiance data proceeds in the following sequence. First, a slant
210 NO₂ column density was determined from spectral fitting, using the differential optical
211 absorption spectroscopy (DOAS) method. Second, the stratospheric NO₂ contribution was
212 removed by subtracting the stratospheric portions of slant NO₂ columns from the total slant
213 NO₂ columns. **The stratospheric NO₂ slant columns were calculated by data assimilation of**
214 **OMI-observed slant NO₂ columns in the global CTM (TM4) (Boersma et al., 2007).** Finally,
215 the tropospheric slant NO₂ columns were converted into vertical NO₂ columns, using the air
216 mass factor (AMF), defined as the ratio of the measured slant column to the vertical column.

217 This AMF is a function of several factors, such as the satellite viewing geometry, surface
218 albedo, surface pressure, and vertical distributions of clouds, aerosols, and trace gases.

219 In this study, to reduce retrieval errors, measured scenes with surface albedo values
220 larger than 0.3 were excluded, as suggested by Boersma et al. (2011b). The surface albedo
221 data was also obtained from the OMI observations (Kleipool et al., 2008). Also, observed
222 pixels with cloud radiance fractions (CRF) larger than 50% were filtered out, which are
223 approximately equivalent to cloud fractions (CF) smaller than 20% (van der A et al., 2006).
224 Thus, OMI-retrieved tropospheric NO₂ columns under almost “cloud-free” conditions were
225 used in this study.

226 Errors in the retrieval of the Ω_{OMI} can mainly be caused by calculations of the AMFs.
227 Boersma et al. (2011a) reported that errors of the Ω_{OMI} mostly due to calculations of the
228 AMFs in KNMI/DOMINO v2.0 products were approximated to be $\sim 1.0 \times 10^{15}$ molecules cm⁻²,
229 with a relative error of 25%. The other errors in the products were from the spectral fitting
230 ($\sim 0.7 \times 10^{15}$ molecules cm⁻²) and the stratospheric slant column ($\sim 0.25 \times 10^{15}$ molecules cm⁻²).

231 The AKs were also applied to the CMAQ model simulations. The AKs are
232 analytically expressed in Eq. (1) (Rodgers, 2000; Eskes and Boersma, 2003):

$$\begin{aligned} AK &= G_y K_x \\ &= \frac{\partial R}{\partial y} \frac{\partial F}{\partial x} \\ &= \frac{\partial \hat{x}}{\partial x} \end{aligned} \quad (1)$$

234 where G_y and K_x represent the sensitivities of the retrieval (R) to the measurement (y) and the
235 forward model (F) to the state (x), respectively. Also, K_x is known as a weighting function or
236 Jacobian matrix. Thus, as shown in Eq. (1), the AKs represent the sensitivity of the retrieved
237 quantities (here, vertical NO₂ column, \hat{x}) to the true atmospheric state (x). Using the AKs, the
238 retrieved quantity (\hat{x}) can be expressed by Eq. (2):

239
$$\hat{x} - \hat{x}_a = AK(x - x_a) + \varepsilon \quad (2)$$

240 where x_a and ε represent *a priori* estimate and total error in measured signal relative to the
241 forward model, respectively. Information on the AKs and retrieved quantity are included in
242 the daily KNMI products (http://www.temis.nl/airpollution/no2col/no2regioomi_v2.php).

243 Fig. 3 presents the vertical distributions of the seasonally-averaged AKs retrieved
244 from the KNMI algorithms over Central East China (CEC) and other regions (defined in Fig.
245 2). As shown in Fig. 3, the AKs are strongly altitude-dependent in the troposphere. For
246 example, near the surface, the AKs are smaller than unity, ranging between 0.2 and 0.7 (based
247 on seasonal averaged values). In contrast, in the upper troposphere, the AKs are larger than
248 unity, ranging between 1.1 and 2.1 (an AK of unity means that the OMI instruments can
249 directly measure the true NO₂ column densities). Additionally, the AKs are generally lower in
250 warm seasons than in cold seasons. These lower values in the AKs during the summer are
251 probably related to lower surface albedos, lower concentrations of aerosols, and large
252 uncertainty in cloud retrieval during the summer (Eskes and Boersma, 2003).

253 Once the CMAQ model simulations were done, all the vertically-resolved NO₂
254 mixing ratios were interpolated into the OMI grid cells on a daily basis. The AKs are
255 sometimes significantly sensitive to allocation into small spatial scales, particularly when they
256 are interpolated into small model grid cells (Boersma et al., 2011b). This is why this direction
257 of allocation was chosen in this study. In other words, the CMAQ-calculated NO₂ data was
258 interpolated into the OMI footprint cells. After this, the AKs under almost cloud-free
259 conditions were applied to the NO₂ mixing ratios at different layers, and were then integrated
260 from surface to tropopause in order to calculate $\Omega_{\text{CMAQ,AK}}$. Meanwhile, the tropospheric NO₂
261 columns were retrieved from the OMI observations via the KNMI algorithms. A direct
262 comparison study was then made between the two Ω products (i.e. Ω_{OMI} vs. $\Omega_{\text{CMAQ,AK}}$). Fig. 4
263 illustrates the main processes of the comparison study.

264 For the purpose of this study, the seasonal average values of Ω_{OMI} and $\Omega_{\text{CMAQ,AK}}$ were
265 calculated (in case of the $\Omega_{\text{CMAQ,AK}}$, daily AK applications were first conducted and then
266 seasonal average values were calculated). Seasonal averaging was carried out to reduce the
267 “random errors” in the NO_2 retrieval process typically caused by instrument signal noise,
268 fitting errors, and uncertainty in cloud information. It has been suggested and demonstrated
269 that the random errors can be diminished by both temporal and/or spatial averaging (Fioletov
270 et al., 2002; Monaghan et al., 2006; Johnson et al., 2007; Richter et al., 2011; Clarisse et al.,
271 2013). In this study, temporal (i.e. seasonal) averaging was selected, since the grid size of the
272 CMAQ modeling ($30 \text{ km} \times 30 \text{ km}$) was similar to the cell size of the OMI footprint ($13 \text{ km} \times$
273 24 km).

274 On the other hand, the application of AKs can reduce “smoothing errors” in the NO_2
275 retrieval process, which are mainly caused by bias in the *a priori* vertical NO_2 profiles. As
276 mentioned previously, TM4-derived *a priori* profiles were used in the OMI NO_2 retrieval
277 process, which can sometimes cause serious smoothing errors. In order to correct such errors,
278 AKs were applied to the CMAQ model simulations in this study (Rodgers, 2000; Eskes and
279 Boersma, 2003). After the application of AKs, *a priori* information from TM4 did not
280 influence the comparison between Ω_{OMI} and $\Omega_{\text{CMAQ,AK}}$.

281

282 **3. Results and Discussions**

283 The objective of this study was to evaluate the NO_x emissions of the INTEX-B,
284 CAPSS, and REAS v1.11 inventories over East Asia by comparing two Ω obtained from the
285 CMAQ model simulations and OMI observations (Sect. 3.1). In addition, several sensitivity
286 analyses were also conducted to examine the influences of the uncertainty factors on the
287 discrepancies between $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} (Sect. 3.2). Obviously, not all the influential
288 factors can be explored within the framework of this study. Thus, several selected issues that

289 may be important are also discussed further in Sect. 3.2.4.

290 **3.1. Comparison between CMAQ-estimated and OMI-retrieved NO₂ columns: Case 1**

291 **3.1.1. CMAQ-calculated vs. OMI-retrieved NO₂ columns**

292 In this study, the analyses were conducted for four seasons: (i) Spring (March–May,
293 2006), (ii) Summer (June–August, 2006), (iii) Fall (September–November, 2006), and (iv)
294 Winter (January–February, 2006 and December, 2006). For more detailed analyses, eight
295 focus regions were also defined: (i) Central East China (CEC), (ii) Central East China 2
296 (CEC2), (iii) South China (SC), (iv) Sichuan Basin (SB), (v) South Korea (SK), (vi) the
297 western part of Japan (JP1), (vii) the eastern part of Japan (JP2), and (viii) the entire domain
298 (DM) (refer to Fig. 2 regarding the domains).

299 Fig. 5 presents the comparison analysis between the Ω_{CMAQ} and Ω_{OMI} for the four
300 seasons over East Asia before and after the applications of the AKs. As shown in Fig. 5, the
301 CMAQ model simulations (the first and second columns) show spatially and seasonally
302 consistent patterns with OMI observations (the third column). For example, the high values of
303 the Ω_{OMI} over the densely populated and economically developed mega-city regions such as
304 Beijing, Shanghai, Hong Kong, Seoul, and Tokyo (refer to Fig. 2 regarding their locations) are
305 well captured by the CMAQ model simulations. The levels of the Ω during the winter are
306 distinctly high. Also, the low values of the Ω_{CMAQ} from the CMAQ model simulation during
307 the summer are well matched with those from the OMI observations. The low levels of the Ω
308 during the summer are **mainly** caused by active NO_x chemical losses via the reaction of NO₂
309 with OH radicals (McConnell and McElroy, 1973; Atkinson et al., 2004; Boersma et al., 2009;
310 Han et al., 2009; Stavrakou et al., 2013). The uncertainties and unknown factors **related to this**
311 **reaction** will be discussed further in Sect. 3.2.4.

312 When panels (a) and (c) in Fig.5 are compared, it can be seen that the Ω_{CMAQ} is in
313 general greatly larger than the Ω_{OMI} over **the regions with strong NO_x emission. This was also**

314 presented in Han et al. (2011). The large differences between the two NO₂ columns can be
315 confirmed again in panel (d) of Fig. 5. However, such a comparison *without* applying the AKs
316 is like comparing apples and oranges, and is not reasonable. Such studies have been
317 conducted over East Asia, with misleading conclusions (e.g. Ma et al., 2006; He et al., 2007;
318 Uno et al., 2007; Shi et al., 2008; Han et al., 2009; 2011). In this context, we now wish to
319 correct our previous conclusions (Han et al., 2011) here, applying the AKs to the CMAQ
320 model simulations, using the linear relationship presented in Eq. (2).

321 After the application of the AKs to the CMAQ model simulations, the comparison
322 becomes independent of *a priori* profile shape used in the NO₂ retrieval process (Eskes and
323 Boersma, 2003). In this study, when the panels (b) and (c) in Fig. 5 are compared, it can be
324 seen that the CMAQ-calculated NO₂ columns considering the AKs are much more
325 comparable to the OMI-retrieved NO₂ columns, possibly indicating that the bottom-up NO_x
326 emission used in the CMAQ model simulations would not be very greatly overestimated,
327 unlike the previous conclusion drawn by Han et al. (2011). Figs. 5(d) and 5(e) more directly
328 show the effects of the application of the AKs. When the AKs are applied, the differences are
329 greatly diminished, and are even negative, particularly over the CEC regions. The $\Omega_{\text{CMAQ,AK}}$
330 becomes smaller than the Ω_{OMI} over the CEC, SC, SK, JP1, and JP2 regions. Also, possible
331 overestimations of the bottom-up NO_x emissions were found in the CEC2 and SB regions,
332 particularly during the winter. Possible underestimations over the CEC and SC regions and
333 overestimations over the SB and CEC2 regions were also presented in the study of Lin (2012).
334 In Lin (2012), the $\Omega_{\text{GOES-CHEM,AK}}$ values were found to be about 20% and 36% lower than the
335 Ω_{OMI} over eastern China in summer and winter, respectively, whereas in the calculations
336 herein, the respective $\Omega_{\text{CMAQ,AK}}$ values were about 57% and 5% lower than the Ω_{OMI} over
337 eastern China. These differences would be caused by the constant NO_x emission fluxes and
338 relatively coarse horizontal resolutions ($0.67^\circ \times 0.5^\circ$) used in the GEOS-CHEM simulations

339 performed by Lin (2012).

340 In Table 2, we summarize the seasonal average tropospheric NO₂ columns and
341 normalized mean errors (NMEs, defined in Table A1) with and without considering the AKs
342 for the eight focus regions. It can be seen that the NMEs (with AKs applied) ranged from 40.3%
343 to 63.2% over the entire domain in Table 2. Although the differences between $\Omega_{\text{CMAQ,AK}}$ and
344 Ω_{OMI} were the smallest during the summer, as shown in Fig. 5, the NMEs showed the largest
345 values during summer. The reasons for this are discussed in detail in Sect. 3.1.2.

346 Collectively, the seasonal and regional (spatial) characteristics observed from the
347 OMI sensor were found to be captured well by the CMAQ model simulations using the
348 INTEX-B, CAPSS, and REAS emission inventories. However, some regional discrepancies
349 between the two NO₂ columns were also found, particularly during winter, indicating possible
350 underestimation of the NO_x emissions over the CEC and SC regions as well as overestimation
351 over the CEC2 and SB regions in the CMAQ model simulations. To further investigate the
352 eight regions of interest, scatter plots and statistical analyses were carried out in Sect. 3.1.2.

353 3.1.2. Scatter plots and statistical analyses

354 Fig. 6 presents the seasonal scatter plot analysis between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} for
355 the eight focus regions defined in Fig. 2. The statistical analysis related to the scatter plots
356 was also conducted in terms of the Pearson correlation coefficient (R), linear regression slope
357 (S), and y-intercept (Y-I). As mentioned in Sect. 2.2, seasonal average of the daily Ω was
358 taken to reduce the random errors which have occurred during the NO₂ measurement and
359 retrieval processes (Fioletov et al., 2002; Monaghan et al., 2006; Johnson et al., 2007; Richter
360 et al., 2011; Clarisse et al., 2013). The use of seasonally-averaged data improved the
361 correlation coefficients from 0.49–0.63 to 0.78–0.88 over the entire domain (DM) (regarding
362 this issue, readers can compare Fig. 6 with Fig. S1). Although the correlation coefficients
363 were sometimes lower than 0.7 in Fig. 6, the two NO₂ columns correlated well, with R values

364 between 0.71 and 0.96 (also, refer to the ‘R’ values colored in Fig. 7). Slopes lower than 1.0
365 (see dashed lines in Fig. 6) were also found in the “blue” regions in Fig. 5(e) such as the CEC,
366 SC, SB, JP1, and JP2 regions. These low slopes indicate the possible “underestimation” of the
367 bottom-up NO_x emissions used in the CMAQ model simulations, as discussed in Sect. 3.1.1.

368 Further statistical analyses were conducted. For absolute differences, Mean Error
369 (ME) and Mean Bias (MB) were utilized. For relative differences, Mean Normalized Gross
370 Error (MNGE), Mean Normalized Bias (MNB), Normalized Mean Error (NME), Normalized
371 Mean Bias (NMB), Mean Fractional Error (MFE), and Mean Fractional Bias (MFB) were
372 used. The Pearson correlation coefficient (R) and index of agreement (IOA) were also
373 analyzed to assess the degrees of correlations and agreement, respectively. These 10
374 performance metrics are defined and described in Table A1 (see Appendix A).

375 Fig. 7 summarizes the seasonal statistical analyses for 8 focus regions. Light colors
376 were used to indicate good agreements, while dark colors marked poor agreements. As shown,
377 the IOAs (as a measure of the degree of model prediction errors, Willmott, 1981) showed high
378 values, between 0.78 and 0.93, over the entire domain. However, the IOAs sometimes showed
379 relatively low values during the summer over several regions where large relative differences
380 were found (e.g. SB, JP1, and JP2 regions during the summer), because the IOA decreased
381 with the large difference between $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} . As shown in Fig. 7, large MEs were
382 found over the CEC region (2.03×10^{15} to 4.51×10^{15} molecules cm⁻²) and MBs mostly ranges
383 between -1.78×10^{15} and 1.88×10^{15} molecules cm⁻² in East Asia, except in CEC. Again, the
384 negative values of the MBs in Fig. 7 indicate that the NO_x emissions used were possibly
385 underestimated.

386 In the seasonal perspective, all statistical parameters of the relative differences (i.e.
387 MNGE, MNB, NME, NMB, MFE, and MFB) showed large values for the summer in all the
388 regions, because the OMI-retrieved quantity in the denominator of the equations (see Table

389 A1) for the summer were relatively small versus the values of the absolute differences in the
390 numerator. In this study, the $\Omega_{\text{CMAQ,AK}}$ values over the entire domain were 7.3% and 59.7%
391 smaller than Ω_{OMI} in terms of the NMB during the summer and winter seasons, respectively.
392 In the regional perspective, the relative differences showed large values in the SC, SB, JP1,
393 and JP2 regions, where the Ω were relatively low (i.e. the same reason leading to larger
394 relative errors and biases in the summer). In this study, the $\Omega_{\text{CMAQ,AK}}$ values during winter
395 were found to be 21.8% smaller than the Ω_{OMI} over CEC, but 32.3% and 54.7% larger over
396 CEC2 and SB, respectively. Collectively, the statistical analyses showed that the $\Omega_{\text{CMAQ,AK}}$
397 were, on annual average, ~28% (from 7% to 60% with seasonal variation) smaller than the
398 Ω_{OMI} , indicating that the NO_x emissions for East Asia were possibly underestimated.

399 **3.2. Sensitivity analyses**

400 After the application of the AKs, both the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} became much more
401 comparable with each other as shown in Fig. 5. Even so, this comparison study still has
402 several uncertainties. Because of the uncertainties, it is a bit difficult to directly relate the
403 differences between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} to under- or over-estimations in the NO_x
404 emissions. Therefore, examination of the uncertainty issues was carried out herein. The issues
405 selected for examination in this study were as follows: (i) the monthly variation in NO_x
406 emissions; (ii) influences of the different magnitude of NO_x emissions; and (iii) different
407 parameterizations of the reaction probability of N_2O_5 onto aerosols in the CMAQ model
408 simulations. These three issues were selected for the following reasons: (i) the emission flux
409 in East Asia is believed to be one of the most uncertain factors, and its magnitude can vary
410 greatly depending on monthly variation as well as methodology and activity data used to
411 estimate the emission fluxes (Cases 2 and 3) (Wang et al., 2007; Zhang et al., 2007; Han et al.,
412 2009; Klimont et al., 2009; Zhang et al., 2009; Xing et al., 2011); and (ii) although the
413 condensation of N_2O_5 radicals is a major NO_x loss processes during the winter and thus may

414 significantly influence the tropospheric NO₂ columns, the magnitudes of γ_{N2O5} remain highly
415 uncertain, ranging between 0.1 and 0.001 (Case 4) (Dentener and Crutzen, 1993; Jacob, 2000;
416 Brown et al., 2006; Davis et al., 2008; Macintyre and Evans, 2010). Sect. 3.2 is therefore
417 devoted to these issues, which are addressed with sensitivity analyses.

418 3.2.1. Monthly variation in NO_x emissions: Case 2

419 First, the monthly variations of NO_x emissions over China were investigated,
420 choosing different monthly variations from the base-case emission. In this sensitivity run (see
421 Table 1), we applied a more drastic/extreme monthly variation of the NO_x emissions (thick-
422 black line in Fig. 1) (Han et al., 2009) to the CMAQ model simulation over China (i.e. all the
423 monthly factors from Han et al. (2009) were applied to China, Korea, and Japan) in this one-
424 year run. The main reason we did this is that, as shown in Figs. 5 and 6, the $\Omega_{\text{CMAQ,AK}}$ was
425 smaller than the Ω_{OMI} , over several main regions (such as CEC and main mega-city areas like
426 Hong Kong and Shanghai) in China, particularly during the “cold months”. It should be noted
427 that during the cold months, the NO_x emission fluxes reported in Han et al. (2009) for China
428 were 1.20 times larger than those from the INTEX-B inventory.

429 The results are presented in Fig. 8. The spatial distributions of the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI}
430 are shown in Fig. 8 for the four seasons. As indicated in Table S1 in the supplementary
431 materials, the application of the AKs again greatly reduced the errors and biases between the
432 two tropospheric NO₂ columns in this sensitivity test. As expected, the $\Omega_{\text{CMAQ,AK}}$ in Fig. 8 (a)
433 generally increased for the spring and winter, whereas it decreased for the summer and fall,
434 compared with the values in Fig. 5 (b). These increases in the $\Omega_{\text{CMAQ,AK}}$ for the winter
435 produced better agreement with the Ω_{OMI} , particularly over the CEC region, showing that the
436 MBs over CEC during the winter decreased from -3.10×10^{15} molecule cm⁻² to -7.42×10^{14}
437 molecule cm⁻² (see the average NO₂ columns and NMEs in Tables 2 and S1). However, as
438 shown in Tables 2 and S1, the situations became worse, except for the CEC region, showing

439 significant increases in NMEs, compared with the NMEs in cases using the monthly variation
440 of the INTEX-B inventory taken from Zhang et al. (2009). Even larger (more serious)
441 differences between the two NO₂ columns in Fig. 8 (c) were found over other regions of
442 China (CEC2, SC, and SB) than those shown in Fig. 5 (e) in terms of errors and biases. For
443 example, the MBs during the winter increased from 2.74×10^{15} , -2.92×10^{13} , and 1.88×10^{15}
444 molecule cm⁻² to 5.26×10^{15} , 7.10×10^{15} , and 5.35×10^{15} molecule cm⁻² over CEC2, SC, and SB,
445 respectively.

446 Further detailed analyses over the eight focus regions were carried out, and the scatter
447 plots and statistical analyses are presented in Figs. S2 and S3 of the supplementary materials.
448 Collectively, the sensitivity test showed that the monthly variations of the OMI observations
449 were better captured by the CMAQ model simulations using the monthly variations of the
450 INTEX-B inventory than those from Han et al. (2009), although the monthly variations in the
451 NO_x emission of the INTEX-B inventory still remain uncertain in China, particularly over the
452 CEC region.

453 3.2.2. Another NO_x emission inventory (REAS v1.11): Case 3

454 There is another NO_x emission inventory available in China: the REAS v1.11
455 emission inventory for 2006 (Ohara et al., 2007). Thus, in this section, the REAS emission
456 inventory, a frequently used bottom-up inventory established by the National Institute of
457 Environmental Studies (NIES) in Japan, was tested over China for January (a cold month) in
458 order to determine the influence of different NO_x emissions on the tropospheric NO₂ columns.
459 Because the REAS v1.11 inventory does not include monthly variation, the same monthly
460 variation of the INTEX-B inventory was also applied to this sensitivity study. The NO_x
461 emissions between the INTEX-B and REAS inventories differed greatly over China. For
462 example, the annual NO_x emissions from the INTEX-B inventory were 2.48, 2.22, 1.60, and
463 0.57 Tg N yr⁻¹ over the CEC, CEC2, SC, and SB regions, respectively, whereas those from the

464 REAS inventory were 1.93, 1.56, 1.40, and 0.40 Tg N yr⁻¹, respectively, over the same regions.

465 The results are presented in Fig. 9 and Table S2. The application of the AKs to the
466 CMAQ model simulations were also taken into account in this comparison (see Table S2). As
467 expected, the $\Omega_{\text{CMAQ,AK}}$ decreased significantly over China, when the REAS NO_x emissions
468 are used (refer to Table S2 in the supplementary materials). Although the absolute differences
469 between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} became smaller over the CEC2 and SB regions, much large
470 underestimates were found over the CEC region, compared with the case of the INTEX-B
471 inventory as shown in Fig. 9.

472 Collectively, our results indicate that (i) the NO_x emission fluxes from the REAS
473 inventory were also underestimated over China (particularly, over the CEC region), (ii) both
474 NO_x emission inventories (INTEX-B and REAS) showed underestimation over the CEC
475 region and the Hong Kong area, and (iii) accurate spatial distributions of NO_x emissions and
476 the magnitude of NO_x emissions were important factors to reduce the degree of disagreement
477 between the CTM-estimated and satellite-retrieved NO₂ columns. For better agreement
478 between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} over China, a combination of the two emission inventories
479 may be a good practical attempt in the CMAQ model simulations over East Asia, based on
480 this result. That is, the INTEX-B NO_x emissions data tended to produce better results over the
481 CEC region, whereas the REAS NO_x emissions data tended to generate better results over the
482 CEC2 and SB regions. However, this issue (i.e., the combination of the two emission
483 inventories) needs to be examined using a more sophisticated approach, and should be
484 investigated further.

485

486 3.2.3. Reaction probability of N₂O₅: Case 4

487 We explored the issue of reaction probability of N₂O₅ ($\gamma_{\text{N}_2\text{O}_5}$) onto aerosols, because a
488 relatively large discrepancy between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} was found, particularly during

489 the winter season. During the winter season, the condensation of N_2O_5 into atmospheric
490 particles is an important NO_x loss process (Dentener and Crutzen, 1993; Brown et al., 2004;
491 2006). Thus, it can affect the CMAQ-simulated NO_2 columns ($\Omega_{\text{CMAQ,AK}}$). Although it is an
492 important physico-chemical NO_x loss process during the winter, the magnitude of $\gamma_{\text{N}_2\text{O}_5}$ has
493 been a controversial issue. In this study, five $\Omega_{\text{CMAQ,AK}}$ from the CMAQ model simulations
494 with five different $\gamma_{\text{N}_2\text{O}_5}$ parameterizations were compared with the Ω_{OMI} over East Asia.
495 These five parameterizations are from the works of: (i) Dentener and Crutzen (1993), (ii)
496 Riemer et al. (2003), (iii) a combination of Riemer et al. (2003) and Evans and Jacob (2005),
497 (iv) Davis et al. (2008), and (v) Brown et al., (2006). The mathematical expressions for these
498 parameterizations are summarized briefly in Table 3. In the Dentener and Crutzen's
499 parameterization (1993), they used a fixed value of $\gamma_{\text{N}_2\text{O}_5}$ of 0.1 in their global CTM
500 simulation (Scheme I in Table 3). In Riemer et al.'s parameterization (2003), $\gamma_{\text{N}_2\text{O}_5}$ is a main
501 function of the acidity of the particles (Scheme II). In the combined parameterization of Evans
502 and Jacob (2006) and Riemer et al. (2003), $\gamma_{\text{N}_2\text{O}_5}$ is a function of relative humidity (RH),
503 temperature, and the acidity of the particles (Scheme III, standard scheme). In Davis et al.
504 (2008)'s parameterization, $\gamma_{\text{N}_2\text{O}_5}$ is a function of all the factors, such as RH, temperature, the
505 acidity of the particles, and the mixing state (Scheme IV). Finally, for Brown et al. (2006)'s
506 parameterization, we used a fixed minimum value of $\gamma_{\text{N}_2\text{O}_5}$ of 10^{-3} in the CMAQ model
507 simulation (Scheme V).

508 The comparison results are presented in Fig. 10. As shown in Fig. 10 and Table 4, the
509 $\Omega_{\text{CMAQ,AK}}$ with the Brown et al. (2006) parameterization were ~19% larger than those with the
510 standard Scheme (III) over East Asia. This indicates that Brown et al.'s parameterization
511 resulted in the smallest NO_x loss rates (or nitrate formation rates) via this physico-chemical
512 reaction pathway.

513 In contrast, the application of the Dentener and Crutzen's parameterization to the

514 CMAQ model simulation produced the smallest $\Omega_{\text{CMAQ,AK}}$ in East Asia, indicating the fastest
515 NO_x loss rates, due to the large $\gamma_{\text{N}_2\text{O}_5}$. These results suggest that Brown et al.'s $\gamma_{\text{N}_2\text{O}_5}$ (= 0.001)
516 may be smaller than the real value, while Dentener and Crutzen's $\gamma_{\text{N}_2\text{O}_5}$ (= 0.1) is probably
517 larger. Other than Brown et al.'s and Dentener and Crutzen's parameterizations, it was found
518 that there was almost no significant or practical difference in the $\Omega_{\text{CMAQ,AK}}$ among the other
519 three Schemes, II, III, and IV (also, refer to Table 4).

520 As shown in Fig. 10 and Table 4, Schemes II, III, and IV tended to produce better
521 $\Omega_{\text{CMAQ,AK}}$ data over East Asia than Schemes I and V, compared with Ω_{OMI} . More recently,
522 Brown et al. (2009) and Bertram et al. (2009) also discussed that the $\gamma_{\text{N}_2\text{O}_5}$ values being used
523 currently in regional/global CTMs were generally larger than those from their observed $\gamma_{\text{N}_2\text{O}_5}$.
524 In addition to the issue of $\gamma_{\text{N}_2\text{O}_5}$, it should be noted that the aerosol surface density (A) is
525 another uncertain factor that can influence the $\Omega_{\text{CMAQ,AK}}$, because the rate constant ($k_{\text{N}_2\text{O}_5}$) of
526 the physico-chemical reaction also depends on the aerosol surface density (refer to the
527 Schwartz formula, $k_{\text{N}_2\text{O}_5} = \frac{A \cdot C_{\text{mean}} \gamma_{\text{N}_2\text{O}_5}}{4}$). Although all of these issues are arguable, our
528 results show that the $\gamma_{\text{N}_2\text{O}_5}$ parameterizations can certainly influence the levels of Ω_{NO_2} in East
529 Asia, particularly during the winter season.

530 3.2.4. More uncertainties and outlooks

531 As mentioned previously, in this type of analysis all types of temporal variation are
532 potentially important and should therefore be taken into account. A sensitivity analysis on the
533 monthly variation in the NO_x emissions in China was performed in Sect. 3.2.1, showing that
534 the monthly variations in NO_x emissions were an important factor. In contrast, there is only
535 limited information on other temporal variation, such as daily and weekly variation in NO_x
536 emissions in East Asia. Unfortunately, no emission inventory in East Asia can provide us with
537 this level of information. Regarding the issue of the temporal variation, the future Korean

538 Geostationary Environmental Monitoring Spectrometer (GEMS) sensor, which is planned to
539 be launched in 2018, will be able to help to obtain such information on daily and weekly
540 variation in the NO_x emissions over East Asia (Kim, 2012).

541 There is also some level of uncertainty in the NO₂-to-NO ratios, as discussed
542 previously by Richter et al. (2005) and Han et al. (2009). This factor may be important,
543 because every satellite remote-sensor monitors only NO₂ columns, not NO_x columns. The
544 NO₂-to-NO ratios are affected seriously by anthropogenic and biogenic VOC (AVOC and
545 BVOC) emissions and their mixing ratios. For example, if we assume a photo-stationary state,
546 the NO₂-to-NO ratios can be influenced by the mixing ratios of ozone and HO₂, CH₃O₂, and
547 RO₂ radicals, as shown in the following formula:

$$548 \quad \frac{[NO_2]}{[NO]} = \frac{k_1[O_3] + k_2[HO_2] + k_3[CH_3O_2] + k_4[RO_2]}{J_1} \quad (3)$$

549 where J_1 is the NO₂ photolysis rate constant (s⁻¹) and k_1 (=1.81×10⁻¹⁴ at 298 K), k_2 (=8.41×10⁻
550 ¹² at 298 K), k_3 (=7.29×10⁻¹² at 298 K), and k_4 (=9.04×10⁻¹² – 2.80×10⁻¹¹ at 298 K) are the
551 reaction rate constants (cm³ molecules⁻¹ s⁻¹) for NO+O₃, NO+HO₂, NO+CH₃O₂, and NO+RO₂
552 reactions, respectively. Although k_1 is the smallest among the 4 reaction rate constants, the
553 NO₂ to-NO ratio tends to be determined by the NO+O₃ reaction, together with the photolysis
554 of NO₂ (J_1), because ambient O₃ mixing ratios usually occur in several tens of ppb. However,
555 the NO+HO₂ and NO+RO₂ reactions during summer have almost equivalent (non-negligible)
556 contribution to the NO₂-to-NO ratios, for example, over the SC region where BVOC
557 emissions are active. In addition, the mixing ratios of ozone, HO₂, CH₃O₂, and RO₂ in Eq. (3)
558 can be affected by AVOC and BVOC emissions and their mixing ratios, which are believed to
559 be highly uncertain in East Asia (Fu et al., 2007; Lin et al., 2012; Han et al., 2013).

560 Third, as also discussed by Han et al. (2009), there is large uncertainty in the NO_x
561 loss rates (or NO_x lifetime) in global/regional CTMs. Many groups have reported that the

562 uncertainty in the NO_x loss rate is related to several factors (Lin et al., 2012; Stavrou et al.,
563 2013), such as nitric acid formation **via the NO_2+OH reaction** (Atkinson et al., 2004; Mollner
564 et al., 2010; Sander et al., 2011; Henderson et al., 2012) **and $\text{NO}+\text{HO}_2$ reaction** (Butkovskaya
565 et al., 2005; 2009), isoprene chemistry (e.g. OH regeneration) during the summer months
566 (Butler et al., 2008; Lelieveld et al., 2008; Archibald et al., 2010; Kubistin et al., 2010; Pugh
567 et al., 2010), alkyl nitrate formation (Browne and Cohen, 2012; Browne et al., 2013),
568 “daytime” HONO chemistry (Harris et al., 1982; Svennson et al., 1987; Rondon and
569 Sanhueza, 1989; Pagsberg et al., 1997; Stemmler et al., 2006; Sörgel et al., 2011; Zhou et al.,
570 2011), **inclusion of in-plume photochemistry** (Karamchandani et al., 2000; Song et al., 2003;
571 **Kim et al., 2009; Song et al., 2010**), and peroxyacetyl nitrate (PAN) formation (Robert et al.,
572 **2002**).

573 **Recently, modeling uncertainties including meteorological parameters were discussed**
574 **comprehensively by Lin et al. (2012). They reported that when tropospheric NO_2 columns**
575 **from several sensitivity simulations were compared with those from standard simulations, the**
576 **largest impact on the tropospheric NO_2 columns was caused by modifying the reaction**
577 **probability of HO_2 onto aerosols (i.e. γ_{HO_2}), followed by the modifications of cloud optical**
578 **depth, HNO_3 formation rate via NO_2+OH , $\gamma_{\text{N}_2\text{O}_5}$, and aromatic species emissions. It was also**
579 **reported in their study that modification of all the parameters could increase the tropospheric**
580 **NO_2 columns by 18% during July and by 8% during January. Although the results herein can**
581 **be complementary to those reported by Lin et al. (2012), all of these issues are on-going and**
582 **open questions.**

583 In addition to the issues mentioned above, in the CTM simulations there are
584 additional uncertainties in **biological NO_x emissions from soil** and pyrogenic NO_x emissions
585 (e.g., biomass burning NO_x emissions) (Bertram et al., 2005; Jaeglé et al., 2005; Hudman et
586 al., 2010; Lin, 2012). However, for example, the **biological NO_x emissions from soil** are

587 usually more active during the summer. During the summer, the NO_x loss rates are so fast that
588 considerations of additional NO_x emissions would hardly change the CTM-calculated NO_2
589 columns (Boersma et al., 2009; Han et al., 2009). The same is true for the issues of OH
590 recycling and isoprene-derived alkyl nitrate formation mentioned above. There are
591 uncertainties and unknown chemistry related to isoprene, but, due to the fast NO_x loss rates
592 during the summer, it has been found that these factors do not greatly affect the $\Omega_{\text{CMAQ,AK}}$
593 during the summer in our test runs (data not shown).

594 On the other hand, in the view of satellite observations, there are errors and
595 uncertainties in the retrievals of the NO_2 vertical columns and the AKs. There are also several
596 NO_2 vertical column products from different sensors (e.g. GOME, OMI, SCIAMACHY, and
597 GOME-2) and from different algorithms (e.g. KNMI, Bremen, BIRA, Harvard Smithsonian,
598 and NASA). For example, the different NO_2 products sometimes show considerable
599 differences (Herron-Thorpe et al., 2010). Overall, different combinations of these sensors and
600 algorithms can produce different NO_2 column products. Thus, in this type of comparison
601 analysis, all the uncertainty factors mentioned above should be taken into account cautiously.

602 **4. Summary and conclusions**

604 The accuracy of bottom-up NO_x emission fluxes from the INTEX-B, CAPSS, and
605 REAS emission inventories were investigated through comparisons between the $\Omega_{\text{CMAQ,AK}}$
606 and Ω_{OMI} in East Asia. For the comparison study, the CMAQ model simulations were carried
607 out over 12 months in 2006 over East Asia. Also, for the direct comparison between the
608 Ω_{CMAQ} and Ω_{OMI} , we applied the AKs to the CMAQ model simulations. **This study showed**
609 **that the seasonal and regional/spatial characteristics from the OMI observations were captured**
610 **well by the CMAQ model simulations using the INTEX-B, CAPSS, and REAS v1.11**
611 **emission inventories over East Asia.** It was also found that the normalized mean errors

612 (NMEs) between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} for the data from East Asia decreased, from ~80% to
613 ~46%, from ~79% to ~44%, and from ~98% to ~40% during the spring, fall, and winter,
614 respectively, compared with the NME between the Ω_{CMAQ} and Ω_{OMI} (without AKs
615 application). Overall, the $\Omega_{\text{CMAQ,AK}}$ were an annual average of ~28% (in terms of the NMB;
616 from 7% to 60% with seasonal variation) smaller in East Asia than the Ω_{OMI} , indicating
617 possible underestimations of the NO_x emissions used in this study.

618 To assess the seasonal and spatial discrepancies, several sensitivity studies, shown in
619 Table 1, were performed considering several uncertainty factors such as (i) monthly variation
620 of NO_x emission, (ii) influences of different NO_x emissions in East Asia, and (iii) reaction
621 probabilities of N_2O_5 . In Table 5, we summarize the relative changes in the NO_2 columns
622 from the sensitivity simulations with respect to those from the standard simulation (Case 1).
623 From the sensitivity simulations, we found that:

- 624 – Monthly variations in NO_x emissions have a strong impact on tropospheric NO_2
625 columns. The relative changes ranged from -31.16% to 65.37% over China, when the
626 monthly factors from Han et al. (2009) were used. However, Han et al.'s monthly
627 variations (2009) resulted in even larger discrepancies between $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI}
628 over several regions in China. The monthly variations of the INTEX-B NO_x inventory
629 had a tendency to result in better agreements between the $\Omega_{\text{CMAQ,AK}}$ and Ω_{OMI} over
630 China.
- 631 – As shown in Table 5, when REAS v1.11 inventory data over China were used in the
632 CMAQ model simulations, the $\Omega_{\text{CMAQ,AK}}$ become -31.45% to -58.44% lower over
633 China than those from the case with the INTEX-B inventory. Based on this, the NO_x
634 emissions from the REAS v1.11 NO_x emissions appeared to be more underestimated
635 over China than the INTEX-B NO_x emissions.

636 – In the sensitivity test of $\gamma_{\text{N}_2\text{O}_5}$, it appeared that the $\gamma_{\text{N}_2\text{O}_5}$ parameterization would not
637 be a negligible factor, particularly during the winter. The $\Omega_{\text{CMAQ,AK}}$ from Brown et al.
638 (2006)’s parameterization were ~19% larger over East Asia than the $\Omega_{\text{CMAQ,AK}}$ from
639 the combined parameterization of Riemer et al. (2003) and Evans and Jacob (2006).
640 In this study, the conventional $\gamma_{\text{N}_2\text{O}_5}$ parameterizations (Schemes II, III, and IV)
641 showed almost no practical differences in the $\Omega_{\text{CMAQ,AK}}$ and tended to produce better
642 $\Omega_{\text{CMAQ,AK}}$ data over East Asia than Schemes I and V.

643
644 One of the main driving forces of this study was to correct our previous conclusions
645 (Han et al., 2011), in which AKs were not employed for the comparison between the Ω_{OMI} and
646 Ω_{CMAQ} . Again, this study indicated that the bottom-up NO_x emissions of the INTEX-B,
647 CAPSS, and REAS v1.11 inventories used in the CMAQ model simulations would be rather
648 underestimated over East Asia. In the sensitivity studies, the influences of different NO_x
649 emissions and monthly variation in NO_x emissions can also significantly influence the levels
650 of the $\Omega_{\text{CMAQ,AK}}$ in East Asia. Moreover, we showed that the $\gamma_{\text{N}_2\text{O}_5}$ parameterization could be
651 another important factor in the winter. Because other possible uncertainty factors still exist, as
652 discussed in Sect. 3.2.4, further analyses are definitely necessary in future studies.

653 The estimation of “top-down” NO_x emissions has also been carried out in East Asia
654 (Stavrakou et al., 2008; Lin et al., 2010; Mijling et al., 2013) using satellite-derived NO_2
655 columns. However, in such top-down estimations, other uncertain (limiting) factors exist,
656 such as the lifetime of NO_x (i.e., τ_{NO_x}). The uncertainty in τ_{NO_x} is also linked with the factors
657 discussed herein in Sect. 3.2.4. In addition, even in the top-down NO_x emission, the random
658 and smoothing errors should be reduced/minimized via temporal and/or spatial averaging and
659 the application of AKs, respectively, as demonstrated herein.

660 Improvements in the NO_x emissions data or evaluation of the accuracy of bottom-up

661 NO_x emission fluxes in East Asia can improve air quality modeling and chemical weather
662 forecasting over East Asia. Thus, much effort should be focused on this issue in the future,
663 particularly on the circumstances over East Asia. In this context, efforts in inverse modeling
664 to improve the NO_x emissions data over East Asia, such as adjoint modeling with measured
665 data and top-down estimations of the NO_x emissions with satellite observations, could also
666 contribute to improving the performance of air quality modeling and the accuracy of chemical
667 weather forecasting over East Asia (Park et al., 2013).

668

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677

678

679 **Appendix A**

680 For statistical analyses between the CMAQ-calculated and OMI-retrieved tropospheric NO₂
 681 columns, several statistical parameters below are introduced in Table A1.

- 682 1. **Absolute errors and biases:** The Mean Error (ME) and Mean Bias (MB) are statistical
 683 parameters used to measure how close the estimated values ($\Omega_{CMAQ,AK}$ in this study) are to
 684 the observed values (Ω_{OMI} in this study). The distinction between the two parameters is that
 685 the MB provides information on overestimation (i.e. positive values) or underestimation
 686 (i.e. negative values) of the estimated values.
- 687 2. **Relative errors and biases:** The Mean Normalized Gross Error (MNGE) and Mean
 688 Normalized Bias (MNB) are statistical parameters used to measure the relative differences
 689 normalized by the observed values. The values of the MNGE and MNB can be
 690 significantly inflated (or overstated), when observations are sometimes close to zero. In
 691 this case, the Normalized Mean Error (NME) and Normalized Mean Bias (NMB) can be
 692 useful statistical parameters, because they avoid over-inflating the measured range.
 693 However, these bias parameters have an issue of asymmetry, meaning that overestimations
 694 (i.e., $+\infty$) are weighted more than the equivalent underestimations (i.e., -100), as shown in
 695 Table A1. The Mean Fractional Bias (MFB) provides equal weight to both sides, which
 696 range from -200 to $+200$, as shown in Table A1.
- 697 3. **Agreements:** The Pearson correlation coefficient (R) is a statistical parameter to measure
 698 the degree to which both the estimated and observed values are linearly related. The value
 699 of $R=1$ indicates perfect agreement between both values, whereas $R=0$ means no linear
 700 relationship. The Pearson correlation coefficient can sometimes be numerically unstable,
 701 depending on the sample size. The Index of Agreement (IOA) is a standardized measure of
 702 the degree of estimation error, ranging from 0 to 1 (Willmott, 1981). Unlike the Pearson
 703 correlation coefficient, the IOA can account for additive and proportional differences in the
 704 estimated and observed means and variances. The value of 0 indicates no agreement
 705 between the estimated and observed values, whereas the value of 1 indicates perfect
 706 agreement.

707
 708 **Table A1. Statistical parameters used in this study.**

| Parameters (unit) | Equations ¹⁾ | Range |
|---|---|---|
| Mean Error (molecules cm ⁻²) | $ME = \frac{1}{N} \sum_{i=1}^N \Omega_{CMAQ,AK} - \Omega_{OMI} $ | 0 to $+\infty$ |
| Mean Bias (molecules cm ⁻²) | $MB = \frac{1}{N} \sum_{i=1}^N (\Omega_{CMAQ,AK} - \Omega_{OMI}) = \overline{\Omega_{CMAQ,AK}} - \overline{\Omega_{OMI}}$ | $-\overline{\Omega_{NO2/OMI}}$ to $+\infty$ |
| Mean Normalized Gross Error (%) | $MNGE = \frac{1}{N} \sum_{i=1}^N \frac{ \Omega_{CMAQ,AK} - \Omega_{OMI} }{\Omega_{OMI}} \times 100$ | 0 to $+\infty$ |
| Mean Normalized Bias (%) | $MNB = \frac{1}{N} \sum_{i=1}^N \left(\frac{\Omega_{CMAQ,AK} - \Omega_{OMI}}{\Omega_{OMI}} \right) \times 100$ | -100 to $+\infty$ |
| Normalized Mean Error (%) | $NME = \frac{\sum_{i=1}^N \Omega_{CMAQ,AK} - \Omega_{OMI} }{\sum_{i=1}^N \Omega_{OMI}} \times 100$ | 0 to $+\infty$ |
| Normalized Mean Bias (%) | $NMB = \frac{\sum_{i=1}^N (\Omega_{CMAQ,AK} - \Omega_{OMI})}{\sum_{i=1}^N \Omega_{OMI}} \times 100$ | -100 to $+\infty$ |

| | | |
|--|--|--------------|
| Mean Fractional Error (%) | $MFE = \frac{1}{N} \sum_{i=1}^N \frac{ \Omega_{CMAQ,AK} - \Omega_{OMI} }{\left(\frac{\Omega_{CMAQ,AK} + \Omega_{OMI}}{2}\right)} \times 100$ | 0 to +200 |
| Mean Fractional Bias (%) | $MFB = \frac{1}{N} \sum_{i=1}^N \frac{(\Omega_{CMAQ,AK} - \Omega_{OMI})}{\left(\frac{\Omega_{CMAQ,AK} + \Omega_{OMI}}{2}\right)} \times 100$ | -200 to +200 |
| Pearson correlation coefficient (dimensionless) | $R = \frac{\sum_{i=1}^N (\Omega_{CMAQ,AK} - \overline{\Omega_{CMAQ,AK}})(\Omega_{OMI} - \overline{\Omega_{OMI}})}{\sqrt{\sum_{i=1}^N (\Omega_{CMAQ,AK} - \overline{\Omega_{CMAQ,AK}})^2 \sum_{i=1}^N (\Omega_{OMI} - \overline{\Omega_{OMI}})^2}}$ | -1 to +1 |
| Index of agreement (dimensionless) | $IOA = 1 - \frac{\sum_{i=1}^N (\Omega_{CMAQ,AK} - \Omega_{OMI})^2}{\sum_{i=1}^N (\Omega_{CMAQ,AK} - \overline{\Omega_{OMI}} + \Omega_{OMI} - \overline{\Omega_{OMI}})^2}$ | 0 to +1 |

709 ¹⁾ $\Omega_{CMAQ,AK}$ and Ω_{OMI} indicate the CMAQ-calculated NO₂ columns with the consideration of AKs and the OMI-
710 retrieved NO₂ columns, respectively. N represents the number of data samples.

References

- 711
712
713 Archibald, A. T., Cooke, M. C., Utembe, S. R., Shallcross, D. E., Derwent, R. G., and
714 Jenkin, M. E.: Impacts of mechanistic changes on HO_x formation and recycling in the
715 oxidation of isoprene, *Atmos. Chem. Phys.*, 10, 8097-8118, doi:10.5194/acp-10-8097-
716 2010, 2010.
- 717 Atkinson, R., Baulch, D. L., Cox, R. A., Crowley, J. N., Hampson, R. F., Hynes, R. G.,
718 Jenkin, M. E., Rossi, M. J., and Troe, J.: Evaluated kinetic and photochemical data for
719 atmospheric chemistry: Volume I - gas phase reactions of O_x, HO_x, NO_x and
720 SO_x species, *Atmos. Chem. Phys.*, 4, 1461-1738, doi:10.5194/acp-4-1461-2004, 2004.
- 721 Bertram, T. H., Heckel, A., Richter, A., Burrows, J. P., and Cohen, R. C.: Satellite
722 measurements of daily variations in soil NO_x emissions, *Geophys. Res. Lett.*, 32(24),
723 L24812, doi:10.1029/2005GL024640, 2005.
- 724 Bertram, T. H., Thornton, J. A., Riedel, T. P., Middlebrook, A. M., Bahreini, R., Bates, T. S.,
725 Quinn, P. K., and Coffman, D. J.: Direct observations of N₂O₅ reactivity on ambient
726 aerosol particles, *Geophys. Res. Lett.*, 36, L19803, doi:10.1029/2009GL040248, 2009.
- 727 Binkowski, F. S. and Roselle, S. J.: Models-3 Community Multi-scale Air Quality (CMAQ)
728 model aerosol components: 1. model description, *J. Geophys. Res.*, 108 (D6), 4183,
729 doi:10.1029/2001JD001409, 2003.
- 730 Boersma, K. F., Eskes, H. J., Veefkind, J. P., Brinksma, E. J., van der A, R. J., Sneep, M.,
731 van den Oord, G. H. J., Levelt, P. F., Stammes, P., Gleason, J. F., and Bucsela, E. J.:
732 Near-real time retrieval of tropospheric NO₂ from OMI, *Atmos. Chem. Phys.*, 7, 2103-
733 2118, doi:10.5194/acp-7-2103-2007, 2007.
- 734 Boersma, K. F., Jacob, D. J., Trainic, M., Rudich, Y., DeSmedt, I., Dirksen, R., and
735 Eskes, H. J.: Validation of urban NO₂ concentrations and their diurnal and seasonal
736 variations observed from the SCIAMACHY and OMI sensors using in situ surface
737 measurements in Israeli cities, *Atmos. Chem. Phys.*, 9, 3867-3879, doi:10.5194/acp-9-
738 3867-2009, 2009.
- 739 Boersma, K. F., Eskes, H. J., Dirksen, R. J., van der A, R. J., Veefkind, J. P., Stammes, P.,
740 Huijnen, V., Kleipool, Q. L., Sneep, M., Claas, J., Leitão, J., Richter, A., Zhou, Y., and
741 Brunner, D.: An improved tropospheric NO₂ column retrieval algorithm for the Ozone
742 Monitoring Instrument, *Atmos. Meas. Tech.*, 4, 1905-1928, doi:10.5194/amt-4-1905-
743 2011, 2011a.
- 744 Boersma, K. F., Braak, R., and van der A, R. J.: Dutch OMI NO₂ (DOMINO) data product
745 v2.0 HE5 data file user manual, TEMIS website, available
746 at:<http://www.temis.nl/airpollution/no2.html> (last access: 2 Oct 2014) 2011b.
- 747 Brown, S. S., Dibb, J. E., Stark, H., Aldener, M., Vozella, M., Whitlow, S., Williams, E. J.,
748 Lerner, B. M., Jakoubek, R., Middlebrook, A. M., DeGouw, J. A., Warneke, C., Goldan,
749 P. D., Kuster, W. C., Angevine, W. M., Sueper, D. T., Quinn P. K., Bates, T. S.,
750 Meagher, J. F., Fehsenfeld, F. C., and Ravishankara, A. R.: Nighttime removal of NO_x
751 in the summer marine boundary layer, *Geophys. Res. Lett.*, 31, L07108, doi:
752 10.1029/2004GL019412, 2004.
- 753 Brown, S. S., Ryerson, T. B., Wollny, A. G., Brock, C. A., Peltier, R., Sullivan, A. P., Weber,
754 R. J., Dube, W. P., Trainer, M., Meagher, J. F., Fehsenfeld, F. C., and Ravishankara, A.
755 R.: Variability in nocturnal nitrogen oxide processing and its role in regional air
756 quality, *Science*, 311, 67-70, 2006.
- 757 Brown, S. S., Dubé, W. P., Fuchs, H., Ryerson, T. B., Wollny, A. G., Brock, C. A., Bahreini, R.,
758 Middlebrook, A. M., Neuman, J. A., Atlas, E., Roberts, J. M., Osthoff, H. D., Trainer,
759 M., Fehsenfeld, F. C., and Ravishankara, A. R.: Reactive Uptake Coefficients for N₂O₅

760 Determined from Aircraft Measurements during the Second Texas Air Quality Study:
761 Comparison to Current Model Parameterizations, *J. Geophys. Res.* 114:D00F10.
762 doi:10.1029/2008JD011679, 2009.

763 Browne, E. C. and Cohen, R. C.: Effects of biogenic nitrate chemistry on the NO_x lifetime in
764 remote continental regions, *Atmos. Chem. Phys.*, 12, 11917-11932, doi:10.5194/acp-
765 12-11917-2012, 2012.

766 Browne, E. C., Min, K.-E., Wooldridge, P. J., Apel, E., Blake, D. R., Brune, W. H.,
767 Cantrell, C. A., Cubison, M. J., Diskin, G. S., Jimenez, J. L., Weinheimer, A. J.,
768 Wennberg, P. O., Wisthaler, A., and Cohen, R. C.: Observations of total RONO₂ over
769 the boreal forest: NO_x sinks and HNO₃ sources, *Atmos. Chem. Phys.*, 13, 4543-4562,
770 doi:10.5194/acp-13-4543-2013, 2013.

771 Butkovskaya, N. I., Kukui, A., Pouvesle, N., and Le Bras, G.: Formation of Nitric Acid in the
772 Gas-Phase HO₂ + NO Reaction: Effects of Temperature and Water Vapor, *J. Phys.*
773 *Chem. A*, 109(29), 6509–6520, doi:10.1021/jp051534v, 2005.

774 Butkovskaya, N., Rayez, M.-T., Rayez, J.-C., Kukui, A., and Le Bras, G.: Water vapor effect
775 on the HNO₃ yield in the HO₂ + NO reaction: Experimental and theoretical evidence, *J.*
776 *Phys. Chem. A*, 113(42), 11327–11342, doi: 10.1021/jp811428p, 2009.

777 Butler, T. M., Taraborrelli, D., Brühl, C., Fischer, H., Harder, H., Martinez, M., Williams, J.,
778 Lawrence, M. G., and Lelieveld, J.: Improved simulation of isoprene oxidation
779 chemistry with the ECHAM5/MESSy chemistry-climate model: lessons from the
780 GABRIEL airborne field campaign, *Atmos. Chem. Phys.*, 8, 4529-4546,
781 doi:10.5194/acp-8-4529-2008, 2008.

782 Byun, D. W. and Schere, K. L.: Review of the governing equations, computational algorithm,
783 and other components of the Models-3 Community Multi-scale Air Quality (CMAQ)
784 Modeling system, *Appl. Mech. Rev.*, 59(2), 51–77, 2006.

785 Carter, W. P. L.: Implementation of the SAPRC-99 Chemical Mechanism into the Models-3
786 Framework, United States Environmental Protection Agency, 2000.

787 Clarisse, L., Coheur, P. -F., Prata, F., Hadji-Lazaro, J., Hurtmans, D., and Clerbaux, C.: A
788 unified approach to infrared aerosol remote sensing and type specification, *Atmos.*
789 *Chem. Phys.*, 13, 2195-2221, 2013.

790 Cofala, J., Bertok, I., Borken-Kleefeld, J., Heyes, C., Klimont, Z., Rafaj, P., Sander, R.,
791 Schöpp, W., and Amann, M.: Emissions of Air Pollutants for the World Energy
792 Outlook 2012 Energy Scenarios, International Institute for Applied System Analysis
793 (IIASA), A-2361, Laxenburg, Austria, 2012.

794 Davis, J. M., Bhave, P. V., and Foley, K. M.: Parameterization of N₂O₅ reaction probabilities
795 on the surface of particles containing ammonium, sulfate, and nitrate, *Atmos. Chem.*
796 *Phys.*, 8, 5295-5311, 2008.

797 Dentener, F. J. and Crutzen, P. J.: Reaction of N₂O₅ on tropospheric aerosols: Impact on the
798 global distribution of NO_x, O₃, and OH levels, *J. Geophys. Res.*, 98, 7149-7163, 1993.

799 Emmons, L. K., Walters, S., Hess, P. G., Lamarque, J.-F., Pfister, G. G., Fillmore, D., Granier,
800 C., Guenther, A., Kinnison, D., Laepple, T., Orlando, J., Tie, X., Tyndall, G.,
801 Wiedinmyer, C., Baughcum, S. L., and Kloster, S.: Description and evaluation of the
802 Model for Ozone and Related chemical Tracers, version 4 (MOZART-4), *Geosci.*
803 *Model Dev.*, 3, 43-67, doi:10.5194/gmd-3-43-2010, 2010.

804 Eskes, H. J. and Boersma, K. F.: Averaging kernels for DOAS total-column satellite retrievals,
805 *Atmos. Chem. Phys.*, 3, 1285-1291, doi:10.5194/acp-3-1285-2003, 2003.

806 Evans, M. J. and Jacob, D. J.: Impact of new laboratory studies of N₂O₅ hydrolysis on global
807 model budgets of tropospheric nitrogen oxides, ozone, and OH, *Geophys. Res. Letts.*,
808 32, L09813, doi:10.1029/2005GL022469, 2005, 2005.

809 Fioletov, V. E., Bodeker, G. E., Miller, A. J., McPeters, R. D., and Stolarski, R.: Global and
810 zonal total ozone variations estimated from ground-based and satellite measurements:
811 1964-2000, *J. Geophys. Res.*, 107, D22, 4647, doi:10.1029/2001JD001350, 2002.

812 Fu, T., Jacob, D. J., Palmer, P. I., Chance, K., Wang, Y. X., Barletta, B., Blake, D. R., Staton, J.
813 C., and Pilling, M. J.: Space-based formaldehyde measurements as constrains on
814 volatile organic compound emissions in east and south Asia and implications for
815 ozone, *J. Geophys. Res.*, 112, D06312, doi:10.1029/2006JD007853, 2007.

816 Ghude, Sachin D., Pfister, Gabriele G., Jena, Chinmay, van der A, R. J., Emmons, Louisa K.,
817 and Kumar, Rajesh: Satellite constraints of nitrogen oxide (NO_x) emissions from India
818 based on OMI observations and WRF-Chem simulations, *Geophys. Res. Lett.*, 40, 1-6,
819 doi:10.1029/2012GL053926, 2013.

820 Han, K. M., Song, C. H., Ahn, H. J., Park, R. S., Woo, J. H., Lee, C. K., Richter, A.,
821 Burrows, J. P., Kim, J. Y., and Hong, J. H.: Investigation of NO_x emissions and NO_x-
822 related chemistry in East Asia using CMAQ-predicted and GOME-derived
823 NO₂ columns, *Atmos. Chem. Phys.*, 9, 1017-1036, doi:10.5194/acp-9-1017-2009,
824 2009.

825 Han, K. M., Lee, C. K., Lee, J., Kim, J. and Song, C. H.: A comparison study between model-
826 predicted and OMI-retrieved tropospheric NO₂ columns over the Korean peninsula,
827 *Atmos. Environ.*, 45, 2962-2971, 2011.

828 Han, K. M. and Song, C. H.: A budget analysis of NO_x column losses over the Korean
829 peninsula, *Asia-Pacific J. Atmos. Sci.*, 48(1), 55-65, 2012.

830 Han, K. M., Park, R. S., Kim, H. K., Woo, J. H., Kim, J., and Song, C. H.: Uncertainty in
831 biogenic isoprene emissions and its impacts on tropospheric chemistry in East Asia, *Sci.*
832 *Total Environ.*, vol. 463-464, 754-771, 2013.

833 Harris, G. W., Carter, W. P. L., Winer, A. M., Pitts, J. N., Platt, U., and Perner, D.:
834 Observations of nitrous acid in the Los Angeles atmosphere and implications for
835 predictions of ozone-precursor relationships, *Environ. Sci. Technol.*, 16(7), 414-419,
836 doi: 10.1021/es00101a009, 1982.

837 He, Y., Uno, I., Wang, Z., Ohara, T., Sugimoto, N., Shimizu, A., Richter, A., and Burrows, J.
838 P.: Variations of the increasing trend of tropospheric NO₂ over central east China
839 during the past decade, *Atmos. Environ.*, 41, 4865-4876, 2007.

840 Henderson, B. H., Pinder, R. W., Crooks, J., Cohen, R. C., Carlton, A. G., Pye, H. O. T., and
841 Vizuete, W.: Combining Bayesian methods and aircraft observations to constrain the
842 HO + NO₂ reaction rate, *Atmos. Chem. Phys.*, 12, 653-667, doi:10.5194/acp-12-653-
843 2012, 2012.

844 Herron-Thorpe, F. L., Lamb, B. K., Mount, G. H., and Vaughan, J. K.: Evaluation of a
845 regional air quality forecast model for tropospheric NO₂ columns using the OMI/Aura
846 satellite tropospheric NO₂ product, *Atmos. Chem. Phys.*, 10, 8839-8854,
847 doi:10.5194/acp-10-8839-2010, 2010.

848 Hilboll, A., Richter, A., and Burrows, J. P.: Long-term changes of tropospheric NO₂ over
849 megacities derived from multiple satellite instruments, *Atmos. Chem. Phys.*, 13, 4145-
850 4169, doi:10.5194/acp-13-4145-2013, 2013.

851 Hong, J. H., Lee, W. S., Kim, D. G., Lee, S. B., Kang, K. H.: 2006 Greenhouse gas and air
852 pollutants emissions in Korea, National Institute of Environmental Research (NIER),
853 Ministry of Environment of Korea, 2008.

854 Horowitz, L., Fiore, A. M., Milly, G. P., Cohen, R. C., Perring, A., Wooldridge, P. J., Hess, P.
855 G., Emmons, L. K., Lamarque, J. -F.: Observational constraints on the chemistry of
856 isoprene nitrates over the eastern United States, *J. Geophys. Res.*, 112, D12S08,
857 doi:10.1029/2006JD007747, 2007.

858 Hudman, R. C., Russell, A. R., Valin, L. C., and Cohen, R. C.: Interannual variability in soil
859 nitric oxide emissions over the United States as viewed from space, *Atmos. Chem.*
860 *Phys.*, 10, 9943-9952, doi:10.5194/acp-10-9943-2010, 2010.

861 Huijnen, V., Eskes, H. J., Poupkou, A., Elbern, H., Boersma, K. F., Foret, G., Sofiev, M.,
862 Valdebenito, A., Flemming, J., Stein, O., Gross, A., Robertson, L., D'Isidoro, M.,
863 Kioutsioukis, I., Friese, E., Amstrup, B., Bergstrom, R., Strunk, A., Vira, J.,
864 Zyryanov, D., Maurizi, A., Melas, D., Peuch, V.-H., and Zerefos, C.: Comparison of
865 OMI NO₂ tropospheric columns with an ensemble of global and European regional air
866 quality models, *Atmos. Chem. Phys.*, 10, 3273-3296, doi:10.5194/acp-10-3273-2010,
867 2010.

868 Itahashi, S., Uno, I., Irie, H., Kurokawa, J.-I., and Ohara, T.: Regional modeling of
869 tropospheric NO₂ vertical column density over East Asia during the period 2000–2010:
870 comparison with multisatellite observations, *Atmos. Chem. Phys.*, 14, 3623-3635,
871 doi:10.5194/acp-14-3623-2014, 2014.

872 Jacob, D. J.: Heterogeneous chemistry and tropospheric ozone, *Atmos. Environ.*, 34, 2131-
873 2159, 2000.

874 Jaeglé, L., Steinberger, L., Martin, R. V., and Chance, K.: Global partitioning of NO_x sources
875 using satellite observations: Relative roles of fossil fuel combustion, biomass burning
876 and soil emissions, *Faraday Discuss.*, 130, 407-423, doi:10.1039/b502128f, 2005.

877 Johnson, E. S., Bonjean, F., Lagerloef, G. S. E., and Gunn, J. T.: Validation and error analysis
878 of OSCAR sea surface currents, *J. Atmos. Oceanic Technol.*, 24, 688-701, 2007.

879 Karamchandani, P., Santos, L., Sykes, I., Zhang, Y., Tonne, C., and Seigneur, C.:
880 Development and evaluation of a state-of-the-science reactive plume model, *Environ.*
881 *Sci. Technol.*, 34, 870– 880, 2000.

882 Kim, H. S., Song, C. H., Park, R. S., Huey, G., and Ryu, J. Y.: Investigation of ship-plume
883 chemistry using a newly-developed photochemical/dynamic ship-plume model, *Atmos.*
884 *Chem. Phys.*, 9, 7531-7550, doi:10.5194/acp-9-7531-2009, 2009.

885 Kim, J.: GEMS (Geostationary Environment Monitoring Spectrometer) onboard the
886 GeoKOMPSAT to Monitor Air Quality in high Temporal and Spatial Resolution over
887 Asia-Pacific Region, EGU General Assembly 2012, 22-27 April 2012, Vienna, Austria,
888 p. 4051, 2012.

889 Kleipool, Q. L., Dobber, M. R., de Haan, J. F., and Levelt, P. E.: Earth surface reflectance
890 climatology from 3 years of OMI data, *J. Geophys. Res.*, 113, D18308,
891 doi:10.1029/2008JD010290, 2008.

892 Klimont, Z., Cofala, J., Xing, J., Wei, W., Zhang, C., Wang, S., Kejun, J., Bhandari, P., Mathur,
893 R., Purohit, P., Rafaj, P., Chambers, A., Amann, M., and Hao, J.: Projections of SO₂,
894 NO_x, and carbonaceous aerosols emissions in Asia, *Tellus*, 61B, 602-617, 2009.

895 Kubistin, D., Harder, H., Martinez, M., Rudolf, M., Sander, R., Bozem, H., Eerdeken, G.,
896 Fischer, H., Gurk, C., Klüpfel, T., Königstedt, R., Parchatka, U., Schiller, C. L.,
897 Stickler, A., Taraborrelli, D., Williams, J., and Lelieveld, J.: Hydroxyl radicals in the
898 tropical troposphere over the Suriname rainforest: comparison of measurements with
899 the box model MECCA, *Atmos. Chem. Phys.*, 10, 9705-9728, doi:10.5194/acp-10-
900 9705-2010, 2010.

901 Kurokawa, J., Ohara, T., Morikawa, T., Hanayama, S., Janssens-Maenhout, G., Fukui, T.,
902 Kawashima, K., and Akimoto, H.: Emissions of air pollutants and greenhouse gases
903 over Asian regions during 2000–2008: Regional Emission inventory in ASia (REAS)
904 version 2, *Atmos. Chem. Phys.*, 13, 11019-11058, doi:10.5194/acp-13-11019-2013,
905 2013.

906 Lamsal, L. N., Martin, R. V., van Donkelaar, A., Celarier, E. A., Bucsela, E. J., Boersma, K. F.,

907 Dirksen, R., Luo, C., and Wang, Y.: Indirect validation of tropospheric nitrogen
908 dioxide retrieved from the OMI satellite instrument: Insight into the seasonal variation
909 of nitrogen oxides at northern midlatitudes, *J. Geophys. Res.*, 115, D05302,
910 doi:10.1029/2009JD013351, 2010.

911 Lelieveld, J., Butler, T. M., Crowley, J. N., Dillon, T. J., Fischer, H., Ganzeveld, L., Harder, H.,
912 Lawrence, M. G., Martinez, M., Taraborrelli, D., and Williams, J.: Atmospheric
913 oxidation capacity sustained by a tropical forest, *Nature*, 452, 737-740,
914 doi:10.1038/nature06870, 2008.

915 Levelt, P. F., van den Oord, G. H. J., Dobber, M. R., Mälkki, A., Visser, H., de Vries, J.,
916 Stammes, P., Lundell, J. O. V., and Saari, H.: The Ozone Monitoring Instrument, *IEEE*
917 *Trans. Geosci. Remote Sens.*, vol. 44(5), 1093-1101, 2006.

918 Lin, J.-T., McElroy, M. B., and Boersma, K. F.: Constraint of anthropogenic NO_x emissions in
919 China from different sectors: a new methodology using multiple satellite retrievals,
920 *Atmos. Chem. Phys.*, 10, 63-78, doi:10.5194/acp-10-63-2010, 2010.

921 Lin, J.-T.: Satellite constraint for emissions of nitrogen oxides from anthropogenic, lightning
922 and soil sources over East China on a high-resolution grid, *Atmos. Chem. Phys.*, 12,
923 2881-2898, doi:10.5194/acp-12-2881-2012, 2012.

924 Lin, J.-T., Liu, Z., Zhang, Q., Liu, H., Mao, J., and Zhuang, G.: Modeling uncertainties for
925 tropospheric nitrogen dioxide columns affecting satellite-based inverse modeling of
926 nitrogen oxides emissions, *Atmos. Chem. Phys.*, 12, 12255-12275, doi:10.5194/acp-
927 12-12255-2012, 2012.

928 Ma, J., Richter, A., Burrows, J. P., Nüß, H., and van Aardenne, J. A.: Comparison of model-
929 simulated tropospheric NO₂ over China with GOME-satellite data, *Atmos. Environ.*,
930 40, 593-604, 2006.

931 Macintyre, H. L. and Evans, M. J.: Sensitivity of a global model to the uptake of N₂O₅ by
932 tropospheric aerosol, *Atmos. Chem. Phys.*, 10, 7409-7414, doi:10.5194/acp-10-7409-
933 2010, 2010.

934 Martin, R. V., Sioris, C. E., Chance, K., Ryerson, T. B., Bertram, T. H., Wooldridge, P. J.,
935 Cohen, R. C., Neuman, J. A., Swanson, A., and Flocke, F. M.: Evaluation of space-
936 based constraints on global nitrogen oxide emissions with regional aircraft
937 measurements over and downwind of eastern North America, *J. Geophys. Res.*, 111,
938 D15308, doi:10.1029/2005JD006680, 2006.

939 McConnell, J. C., and McElroy, M. B.: Odd nitrogen in the atmosphere, *J. Atmos. Sci.*, 30(8),
940 1465-1480, 1973.

941 Mijling, B., van der A, R. J., and Zhang, Q.: Regional nitrogen oxides emission trends in East
942 Asia observed from space, *Atmos. Chem. Phys.*, 13, 12003-12012, doi:10.5194/acp-
943 13-12003-2013, 2013.

944 Mollner, A. K., Valluvadasan, S., Feng, L., Sprague, M. K., Okumura, M., Milligan, D. B.,
945 Bloss, W. J., Sander, S. P., Martien, P. T., Harley, R. A., McCoy, A. B., and Carter, W. P.
946 L.: Rate of Gas Phase Association of Hydroxyl Radical and Nitrogen Dioxide, *Science*,
947 330, 646-649, doi:10.1126/science.1193030, 2010.

948 Monaghan, A. J., Bromwich, D. H., and Wang, S. H.: Recent trends in Antarctic snow
949 accumulation from Polar MM5 simulations, *Phil. Trans. R. Soc. A*, 364, 1683-1708,
950 doi:10.1098/rsta.2006.1795, 2006.

951 Müller, J.-F., Stavrou, T., Wallens, S., De Smedt, I., Van Roozendaal, M., Potosnak, M. J.,
952 Rinne, J., Munger, B., Goldstein, A., and Guenther, A. B.: Global isoprene emissions
953 estimated using MEGAN, ECMWF analyses and a detailed canopy environment
954 model, *Atmos. Chem. Phys.*, 8, 1329-1341, doi:10.5194/acp-8-1329-2008, 2008.

955 Ohara, T., Akimoto, H., Kurokawa, J., Horii, N., Yamaji, K., Yan, X., and Hayasaka, T.: An

956 Asian emission inventory of anthropogenic emission sources for the period 1980-2020,
957 Atmos. Chem. Phys., 7, 4419-4444, 2007.

958 Pagsberg, P., Bjergbakke, E., Ratajczak, E., and Sillescu, A.: Kinetics of the gas phase
959 reaction $\text{OH} + \text{NO}(+\text{M}) \rightarrow \text{HONO}(+\text{M})$ and the determination of the UV absorption
960 cross sections of HONO, Chem. Phys. Lett., 272, 383-390, 1997.

961 Park, R. S., Han, K. M., Song, C. H., Park, M. E., Lee, S. J., Hong, S. Y., Kim, J., and Woo, J.
962 –H.: Current Status and Development of Modeling Techniques for Forecasting and
963 Monitoring of Air Quality over East Asia, J. KOSAE (in Korean), 29(4), 407-438,
964 2013.

965 Park, R.S., Lee, S., Shin, S.-K., and Song, C.H.: Contribution of ammonium nitrate to aerosol
966 optical depth and direct radiative forcing by aerosols over East Asia, Atmos. Chem.
967 Phys., 14, 2185-2201, 2014.

968 Platt, U. F., Winer, A. M., Biermann, H. W., Atkinson, R., and Pitts, J. N.: Measurement of
969 nitrate radical concentrations in continental air, Environ. Sci. Technol., 18(5), 365-369,
970 doi:10.1021/es00123a015, 1984.

971 Pleim, J. E.: A combined local and nonlocal closure model for the atmospheric boundary layer,
972 Part I: Model description and testing, J. Appl. Meteor. Climatol., 46, 1383-1395, 2007.

973 Pugh, T. A. M., MacKenzie, A. R., Hewitt, C. N., Langford, B., Edwards, P. M.,
974 Furneaux, K. L., Heard, D. E., Hopkins, J. R., Jones, C. E., Karunaharan, A., Lee, J.,
975 Mills, G., Misztal, P., Moller, S., Monks, P. S., and Whalley, L. K.: Simulating
976 atmospheric composition over a South-East Asian tropical rainforest: performance of a
977 chemistry box model, Atmos. Chem. Phys., 10, 279-298, doi:10.5194/acp-10-279-
978 2010, 2010.

979 Richter, A., Burrows, J. P., Nüß, H., Granier, C., and Niemeier, U.: Increase in tropospheric
980 nitrogen dioxide over China observed from space, Nature, 437, 129-132, 2005.

981 Richter, A., Begoin, M., Hilboll, A., and Burrows, J. P.: An improved NO_2 retrieval for the
982 GOME-2 satellite instrument, Atmos. Meas. Tech., 4, 1147-1159, 2011.

983 Riemer, N., Vogel, H., Vogel, B., Schell, B., Ackermann, I., Kessler, C., and Hass, H.: Impact
984 of the heterogeneous hydrolysis of N_2O_5 on chemistry and nitrate aerosol formation in
985 the lower troposphere under photosmog conditions, J. Geophys. Res., 108(D4), 4144,
986 doi:10.1029/2002JD002436, 2003.

987 Roberts, J. M., Flocke, F., Stroud, C. A., Hereid, D., Williams, E. J., Fehsenfeld, F. C., Brune,
988 W., Martinez, M., and Harder, H.: Ground-based measurements of peroxyacetylic
989 nitric anhydrides (PANs) during the 1999 Southern Oxidant Study Nashville Intensive,
990 J. Geophys. Res., 107(d21), 4554, doi:10.1029/2001JD000947, 2002.

991 Rodgers, C. D.: Inverse methods for atmospheric sounding: theory and practice, Series on
992 Atmospheric, Oceanic and Planetary Physics – Vol. 2, World Scientific Publishing,
993 Singapore, 43-63, 2000.

994 Rondon, A. and Sanhueza, E.: High HONO atmospheric concentrations during vegetation
995 burning in the tropical savannah, Tellus B, 41B(4), 474-477, doi: 10.1111/j.1600-
996 0889.1989.tb00323.x, 1989.

997 Sander, S. P., Abbatt, J., Barker, J. R., Burkholder, J. B., Friedl, R. R., Golden, D. M., Huie, R.
998 E., Kolb, C. E., Kurylo, M. J., Moortgat, G. K., Orkin, V. L., and Wine, P. H.:
999 Chemical Kinetics and Photochemical Data for Use in Atmospheric Studies,
1000 Evaluation number 17, NASA Panel for data evaluation, JPL Publication 10-6, Jet
1001 Propulsion Laboratory, Pasadena, <http://jpldataeval.jpl.nasa.gov> (last access: 27 June
1002 2014), 2011.

1003 Schneider, P. and van der A, R. J.: A global single-sensor analysis of 2002-2011 tropospheric
1004 nitrogen dioxide trends observed from space, J. Geophys. Res., 117, D16309,

doi:10.1029/2012JD017571, 2012.

1005 Shi, C., Fernando, Wang, Z., An, X., and Wu, Q.: Tropospheric NO₂ columns over East
 1006 Central China: Comparisons between SCIAMACHY measurements and nested
 1007 CMAQ simulations, *Atmos. Environ.*, 42, 7165-7173, 2008.

1008 Song, C. H., Chen, G., Hanna, S. R., Crawford, J., and Davis, D. D.: Dispersion and chemical
 1009 evolution of ship plumes in the marine boundary layer: Investigation of O₃/NO_y/HO_x
 1010 chemistry, *J. Geophys. Res.*, 108(D4), 4143, doi:10.1029/2002JD002216, 2003

1011 Song, C. H., Kim, H. S., von Glasow, R., Brimblecombe, P., Kim, J., Park, R. J., Woo, J. H.,
 1012 and Kim, Y. H.: Source identification and budget analysis on elevated levels of
 1013 formaldehyde within the ship plumes: a ship-plume photochemical/dynamic model
 1014 analysis, *Atmos. Chem. Phys.*, 10, 11969-11985, doi:10.5194/acp-10-11969-2010,
 1015 2010.

1016 Sörgel, M., Regelin, E., Bozem, H., Diesch, J.-M., Drewnick, F., Fischer, H., Harder, H.,
 1017 Held, A., Hosaynali-Beygi, Z., Martinez, M., and Zetzsch, C.: Quantification of the
 1018 unknown HONO daytime source and its relation to NO₂, *Atmos. Chem. Phys.*, 11,
 1019 10433-10447, doi:10.5194/acp-11-10433-2011, 2011.

1020 Stauffer, D. R. and Seaman, N. L.: Use of four-dimensional data assimilation in a limited-area
 1021 mesoscale model. Part I: experiments with synoptic-scale data, *Monthly Weather*
 1022 *Review*, 118 (6), 1250-1277, 1990.

1023 Stauffer, D. L. and Seaman, N. L., Multiscale four-dimensional data assimilation, *J. Applied*
 1024 *Meteorology*, 33 (3), 416-434, 1994.

1025 Stavrou, T., Müller, J. -F., Boersma, K. F., De Smedt, I., and van der A, R. J.: Assessing the
 1026 distribution and growth rates of NO_x emission sources by inverting a 10-year record of
 1027 NO₂ satellite columns, *Geophys. Res. Lett.*, 35, L10801, doi:10.1029/2008GL033521,
 1028 2008.

1029 Stavrou, T., Müller, J.-F., Boersma, K. F., van der A, R. J., Kurokawa, J., Ohara, T., and
 1030 Zhang, Q.: Key chemical NO_x sink uncertainties and how they influence top-down
 1031 emissions of nitrogen oxides, *Atmos. Chem. Phys.*, 13, 9057-9082, doi:10.5194/acp-
 1032 13-9057-2013, 2013.

1033 Stemmler, K., Ammann, M., Donders, C., Kleffmann, J., and George, C.: Photosensitized
 1034 reduction of nitrogen dioxide on humic acid as a source of nitrous acid, *Nature*, 440,
 1035 195-198, doi:10.1038/nature04603, 2006.

1036 Streets, D. G., Bond, T. C., Carmichael, G. R., Fernandes, S. D., Fu, Q., He, D., Klimont, Z.,
 1037 Nelson, S. M., Tsai, N. Y., Wang, M. Q., Woo, J. -H., and Yarber, K. F.: An inventory
 1038 of gaseous and primary aerosol emissions in Asia in the year 2000, *J. Geophys. Res.*,
 1039 108 (D21), 8809, doi:10.1029/2002JD003093, 2003.

1040 Svensson, R., Ljungström, E., and Lindqvist, O.: Kinetics of the reaction between nitrogen
 1041 dioxide and water vapour, *Atmos. Environ.*, 21(7), 1529-1539, 1987.

1042 Tie, X., Emmons, L., Horowitz, L., Brasseur, G., Ridley, B., Atlas, E., Stround, C., Hess, P.,
 1043 Klonecki, A., Madronich, S., Talbot, R., and Dibb, J.: Effect of sulfate aerosol on
 1044 tropospheric NO_x and ozone budgets: Model simulations and TOPSE evidence, *J.*
 1045 *Geophys. Res.*, 108(D4), 8364, doi: 10.1029/2001JD001508, 2003.

1046 Uno, I., He, Y., Ohara, T., Yamaji, K., Kurokawa, J.-I., Katayama, M., Wang, Z., Noguchi, K.,
 1047 Hayashida, S., Richter, A., and Burrows, J. P.: Systematic analysis of interannual and
 1048 seasonal variations of model-simulated tropospheric NO₂ in Asia and comparison with
 1049 GOME-satellite data, *Atmos. Chem. Phys.*, 7, 1671-1681, 2007.

1050 van der A, R. J., Peters, D. H. M. U., Eskes, H., Boersma, K. F., Van Roozendaal, M., De
 1051 Smedt, I., and Kelder, H. M., Detection of the trend and seasonal variation in
 1052 tropospheric NO₂ over China, *J. Geophys. Res.*, 111 (D12), 27, DOI:
 1053

1054 10.1029/2005JD006594, 2006.
1055 Wang, Y., McElory, M. B., Martin, R. V., Streets, D. G., Zhang, Q., and Fu, T.-M.: Seasonal
1056 variability of NO_x emissions over east China constrained by satellite observations:
1057 Implications for combustion and microbial sources, *J. Geophys. Res.*, 112, D06301,
1058 doi:10.1029/2006JD007538, 2007.
1059 Wild, O., Prather, M. J., and Akimoto, H.: Indirect long-term global radiative cooling from
1060 NO_x emissions, *Geophys. Res. Lett.*, 28(9), 1719-1722, 2001.
1061 Willmott, C. J.: On the validation of models, *Phys. Geogr.*, 2, 184-194, 1981.
1062 Xing, J., Wang, S. X., Chatani, S., Zhang, C. Y., Wei, W., Hao, J. M., Klimont, Z., Cofala, J.,
1063 and Amann, M.: Projections of air pollutant emissions and its impacts on regional air
1064 quality in China in 2020, *Atmos. Chem. Phys.*, 11, 3119-3136, doi:10.5194/acp-11-
1065 3119-2011, 2011.
1066 Yamartino, R. J.: Nonnegative, conserved scalar transport using grid-cell-centered, spectrally
1067 constrained Blackman cubics for applications on a variable-thickness mesh, *Mon. Wea.*
1068 *Rev.*, 121, 753-763, 1993.
1069 Zhang, Q., Streets, D. G., He, K., Wang, Y., Richter, A., Burrows, J. P., Uno, I., Jang, C. J.,
1070 Chen, D., Yao, Z., and Lei, Y.: NO_x emission trends for China, 1995-2004: The view
1071 from the ground and the view from space, *J. Geophys. Res.*, 112, D22306,
1072 doi:10.1029/2007JD008684, 2007.
1073 Zhang, Q., Streets, D. G., Carmichael, G. R., He, K. B., Huo, H., Kannari, A., Klimont, Z.,
1074 Park, I. S., Reddy, S., Fu, J. S., Chen, D., Duan, L., Lei, Y., Wang, L. T., and Yao, Z. L.:
1075 Asian emissions in 2006 for the NASA INTEX-B mission, *Atmos. Chem. Phys.*, 9,
1076 5131-5153, doi:10.5194/acp-9-5131-2009, 2009.
1077 Zhou, X., Zhang, N., TerAvest, M., Tang, D., Hou, J., Bertman, S., Alaghmand, M., Shepson,
1078 P. B., Carroll, M. A., Griffith, S., Dusanter, S., and Stevens, P. S.: Nitric acid
1079 photolysis on forest canopy surface as a source for tropospheric nitrous acid, *Nature*
1080 *Geoscience*, 4, 440-443, doi:10.1038/ngeo1164, 2011.
1081 Zyrichidou, I., Koukouli, M. E., Balis, D. S., Kioutsioukis, I., Poupkou, A., Katragkou, E.,
1082 Melas, D., Boersma, K. F., and van Roozendaal, M.: Evaluation of high resolution
1083 simulated and OMI retrieved tropospheric NO₂ column densities over Southeastern
1084 Europe, *Atmos. Res.*, 122, 55-65, 2013.
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Figure Captions

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Fig. 1. Monthly variation in NO_x emissions in China. Here, the ‘MEIC_2008’ and ‘MEIC_2010’ were obtained from the website, <http://www.meicmodel.org/>.

Fig. 2. Study domain and eight focus regions in this study: Central East China (CEC), Central East China 2 (CEC2), South China (SC), Sichuan Basin (SB), South Korea (SK), western part of Japan (JP1), eastern part of Japan (JP2), and entire domain (DM).

Fig. 3. Vertical distributions of averaging kernels (AKs) with error bars (one-sigma standard deviations from the mean) for four seasons over (a) CEC, (b) CEC2, (c) SC, (d) SB, (e) SK, (f) JP1, (g) JP2, and (h) DM regions (refer to Fig. 2 regarding the regions of analysis).

Fig. 4. Flow diagram for direct comparison between CMAQ-estimated and OMI-retrieved NO₂ columns.

Fig. 5. Spatial and seasonal distributions of CMAQ-calculated tropospheric NO₂ columns (a) without the applications of the AKs and (b) with the AKs and (c) OMI-retrieved NO₂ columns from the KNMI algorithm. Differences between OMI-retrieved and CMAQ-calculated NO₂ columns (d) before the applications of the AKs and (e) after the applications of the AKs.

Fig. 6. Seasonal scatter plots between CMAQ-calculated and OMI-retrieved NO₂ columns (Unit: $\times 10^{15}$ molecules cm⁻²) using seasonally averaged data sets over the CEC, CEC2, SC, SB, SK, JP1, JP2, and DM regions. Here, the AKs were applied to the CMAQ model simulations. R, S, Y-I, and N represent the correlation coefficient, linear regression slope, y-intercept, and the number of data points, respectively.

Fig. 7. Statistical analyses between CMAQ-calculated and OMI-retrieved NO₂ columns using the performance metrics defined in Table A1. Here, the color bars represent ME and MB at the top, MNGE, MNB, NME, NMB, MFE, and MFB in the middle, and IOA and R at the bottom. Here, light colors show good agreements while dark colors indicate poor agreements.

Fig. 8. Spatial distributions of (a) CMAQ-calculated NO₂ columns with the AKs and (b) OMI-retrieved NO₂ columns and (c) their differences for four seasonal episodes. Here, the monthly variations of NO_x emissions from Han et al. (2009) were applied to the CMAQ model simulations.

Fig. 9. CMAQ-calculated NO₂ columns using (a) INTEX-B inventory and (b) REAS inventory over China and (c) OMI-observed NO₂ columns for January.

Fig. 10. CMAQ-calculated NO₂ columns using five $\gamma_{\text{N}_2\text{O}_5}$ parameterizations from (a) Dentener and Crutzen (1993), (b) Riemer et al. (2003), (c) combination of Riemer et al. (2003) and Evans and Jacob (2005), (d) Davis et al. (2008), and (e) Brown et al., (2006) and (f) OMI-observed NO₂ columns for January.

Fig. S1. Scatter plots between daily $\Omega_{\text{CMAQ,AK}}$ and daily Ω_{OMI} over the DM regions for four

1135 **seasonal episodes.**

1136 **Fig. S2.** As in Fig. 6, except for the monthly variations of NO_x emissions from Han et al.
1137 (2009).

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1139 **Fig. S3.** As in Fig. 7, except for the monthly variations of NO_x emissions from Han et al.
1140 (2009).

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Table 1. Description of CMAQ model simulations conducted in this study.

| Cases | Sensitivity test | Month, year | Description | Section |
|-------|---|----------------------|--|-------------|
| 1 | Base-case simulation | Jan. – Dec., 2006 | - Seasonal variation of NO _x emission from INTEX-B inventory for China (Zhang et al., 2009) and from Han et al. (2009) for Korea and Japan. - NO _x emissions from INTEX-B, CAPSS, and REAS inventories for China, Korea, and Japan, respectively - Parameterization of $\gamma_{\text{N}_2\text{O}_5}$ from the combination of Riemer et al. (2003) and Evans and Jacob (2005) | Sect. 3.1 |
| 2 | Seasonal variation of NO _x emission | Jan. – Dec., 2006 | - As case 1 except for seasonal variation of NO _x emission from Han et al. (2009) for China (i.e. all the monthly factors from Han et al. (2009) for China, Korea, and Japan.) | Sect. 3.2.1 |
| 3 | Emission strength | Jan., 2006 | - As case 1 except for NO _x emissions from REAS inventory for China | Sect. 3.2.2 |
| 4 | Reaction probability of N ₂ O ₅ | Jan., 2006 | - As case 1 except for the $\gamma_{\text{N}_2\text{O}_5}$ parameterizations from: (i) Dentener and Crutzen (1993); (ii) Riemer et al. (2003); (iii) Davis et al. (2008); and (iv) Brown et al., (2006) | Sect. 3.2.3 |

Table 2. Average tropospheric NO₂ columns, standard deviations, and the normalized mean error (NME) with and without the application of AKs for four seasons.

| Region | Season | n ⁽¹⁾ | Ω_{CMAQ} (w/o AKs) ⁽²⁾ | $\Omega_{\text{CMAQ,AK}}$ (w/ AKs) ⁽²⁾ | Ω_{OMI} ⁽²⁾ | NME (w/o AKs) | NME (w/ AKs) |
|---------------|--------|------------------|---|---|--------------------------------------|---------------|--------------|
| CEC | Spring | 900 | 11.68 (6.19) ⁽³⁾ | 6.40 (3.95) ⁽³⁾ | 6.89 (4.07) | 74.48 | 29.48 |
| | Summer | 900 | 6.43 (4.09) | 2.60 (1.80) | 5.29 (3.02) | 45.55 | 53.06 |
| | Fall | 900 | 13.29 (7.71) | 7.18 (5.04) | 9.49 (5.89) | 52.08 | 32.79 |
| | Winter | 900 | 16.95 (9.52) | 11.08 (7.52) | 14.18 (8.05) | 37.52 | 31.77 |
| CEC2 | Spring | 820 | 10.49 (6.34) | 4.79 (4.12) | 4.45 (3.98) | 135.72 | 29.75 |
| | Summer | 820 | 6.01 (6.16) | 2.31 (2.92) | 3.02 (2.15) | 102.70 | 39.44 |
| | Fall | 820 | 12.36 (7.44) | 5.84 (4.39) | 4.97 (3.97) | 148.85 | 36.61 |
| | Winter | 820 | 20.07 (6.84) | 11.24 (5.54) | 8.49 (5.79) | 136.26 | 42.58 |
| SC | Spring | 1125 | 3.79 (2.87) | 1.16 (1.04) | 2.20 (2.03) | 81.80 | 50.26 |
| | Summer | 1124 | 2.65 (2.57) | 0.76 (0.85) | 1.77 (1.73) | 65.26 | 57.83 |
| | Fall | 1125 | 3.79 (2.79) | 1.27 (1.02) | 2.20 (2.31) | 79.89 | 44.80 |
| | Winter | 1125 | 8.98 (4.06) | 3.21 (1.88) | 3.24 (3.39) | 181.26 | 36.41 |
| SB | Spring | 408 | 4.25 (2.84) | 1.53 (1.09) | 2.56 (1.55) | 80.16 | 44.97 |
| | Summer | 420 | 2.34 (1.66) | 0.78 (0.59) | 2.14 (0.99) | 39.86 | 63.31 |
| | Fall | 418 | 6.37 (4.47) | 2.34 (1.76) | 2.71 (2.15) | 143.93 | 43.75 |
| | Winter | 403 | 11.55 (7.69) | 5.31 (4.14) | 3.43 (3.01) | 237.75 | 72.46 |
| SK | Spring | 260 | 9.14 (5.78) | 4.95 (3.50) | 5.24 (3.74) | 75.37 | 26.93 |
| | Summer | 260 | 7.52 (7.94) | 3.06 (3.60) | 3.41 (2.58) | 128.05 | 42.73 |
| | Fall | 260 | 8.85 (6.60) | 4.60 (3.71) | 4.81 (3.62) | 93.57 | 38.81 |
| | Winter | 260 | 12.30 (5.69) | 6.82 (3.37) | 6.68 (4.14) | 88.42 | 29.78 |
| JP1 | Spring | 204 | 4.61 (1.51) | 2.03 (0.73) | 3.58 (2.48) | 44.83 | 45.50 |
| | Summer | 204 | 2.47 (1.06) | 0.77 (0.33) | 2.91 (1.98) | 34.88 | 73.42 |
| | Fall | 204 | 4.62 (1.92) | 1.91 (0.90) | 3.57 (2.50) | 41.81 | 48.26 |
| | Winter | 204 | 7.63 (2.88) | 3.47 (1.41) | 4.48 (3.07) | 74.66 | 36.95 |
| JP2 | Spring | 285 | 3.90 (3.27) | 1.72 (1.75) | 3.09 (2.96) | 36.19 | 45.69 |
| | Summer | 286 | 2.41 (2.08) | 0.86 (0.81) | 2.64 (2.77) | 29.99 | 67.72 |
| | Fall | 286 | 3.96 (3.33) | 1.63 (1.66) | 3.12 (3.17) | 31.95 | 47.71 |
| | Winter | 279 | 5.84 (4.60) | 2.56 (2.45) | 3.92 (4.20) | 55.64 | 42.72 |
| Entire domain | Spring | 15175 | 3.02 (4.46) | 1.35 (2.39) | 1.97 (2.43) | 80.49 | 45.85 |
| | Summer | 15207 | 1.76 (3.09) | 0.64 (1.29) | 1.59 (1.72) | 59.27 | 63.15 |
| | Fall | 15224 | 3.31 (5.13) | 1.45 (2.72) | 2.06 (3.05) | 78.78 | 44.27 |
| | Winter | 14075 | 5.97 (7.31) | 2.96 (4.52) | 3.20 (4.79) | 98.13 | 40.31 |

⁽¹⁾ The number of data; ⁽²⁾ Unit, $\times 10^{15}$ molecules cm^{-2} ; ⁽³⁾ Standard deviations of the distributions of tropospheric NO₂ columns

Table 3. Reaction probabilities of N₂O₅ onto aerosol surfaces.

| References | Condensing medium | Reaction probability of N ₂ O ₅ ($\gamma_{N_2O_5}$) |
|--|---------------------|---|
| Dentener and Crutzen (1993) [†] (Scheme I Fig. 10 (a)) | Aqueous particles | $\gamma_{N_2O_5} = 0.1$ |
| Jacob (2000) ^{†,‡} | Aqueous particles | $\gamma_{N_2O_5} = 0.1$ (Range: 0.01-1) |
| Tie et al. (2003) [†] | Aqueous particles | $\gamma_{N_2O_5} = 0.04$ (Range: 0.0-0.10) |
| Rierner et al. (2003) [†] (Scheme II in Fig. 10 (b)) | Sulfate and Nitrate | $\gamma_{N_2O_5} = f \cdot \gamma_1 + (1-f) \cdot \gamma_2$ (Range: 0.02 - 0.002) $\gamma_1 = 0.02, \gamma_2 = 0.002; f = \frac{m_{SO_4^{2-}}}{m_{SO_4^{2-}} + m_{NO_3^-}}$ $m_{SO_4^{2-}}$ and $m_{NO_3^-}$: aerosol mass concentrations of sulfate and nitrate, respectively |
| Evans and Jacob (2005) [†] | Sulfate | $\gamma_{N_2O_5} = \alpha \times 10^\beta$ $\alpha = 2.79 \times 10^{-4} + 1.3 \times 10^{-4} \times RH - 3.43 \times 10^{-6} \times RH^2 + 7.52 \times 10^{-8} \times RH^3$ $\beta = 4 \times 10^{-2} \times (294 - T)$ (T ≥ 282K) $\beta = 0.48$ (T < 282K) |
| | OC | $\gamma_{N_2O_5} = RH \times 5.2 \times 10^{-4}$ (RH < 57%) $\gamma_{N_2O_5} = 0.03$ (RH ≥ 57%) |
| | BC | $\gamma_{N_2O_5} = 0.005$ |
| | Sea salt | $\gamma_{N_2O_5} = 0.005$ (RH < 62%) $\gamma_{N_2O_5} = 0.03$ (RH ≥ 62%) |
| | Dust | $\gamma_{N_2O_5} = 0.01$ |
| | | RH : fractional relative humidity; T : temperature (K) |
| Combination of parameterization by Evan and Jacob (2005) and Rierner et al. (2003) [†] (Scheme III in Fig. 10 (c)) | Sulfate and Nitrate | $\gamma_{N_2O_5} = f \cdot \gamma_1 + (1-f) \cdot \gamma_2$ $\alpha = 2.79 \times 10^{-4} + 1.3 \times 10^{-4} \times RH - 3.43 \times 10^{-6} \times RH^2 + 7.52 \times 10^{-8} \times RH^3$ $f = \frac{m_{SO_4^{2-}}}{m_{SO_4^{2-}} + m_{NO_3^-}}$ $\gamma_1 = \alpha \times 10^{0.48}; \gamma_2 = 0.1 \times \gamma_1$ (T < 282K) |

| | | |
|--|-------------------|---|
| | | $\gamma_1 = \alpha \times 10^\beta$; $\gamma_2 = 0.1 \times \gamma_1$; $\beta = 4 \times 10^{-2} \times (294 - T)$ (T \geq 282K) |
| Davis et al. (2008) [†] (Scheme IV in Fig. 10 (d)) | Aqueous particles | $\gamma_{N_2O_5, mix} = \sum_{i=1}^3 x_i \cdot \gamma_i$ $x_1 = 1 - (x_2 + x_3) \quad \text{for bisulfate}$ $x_2 = \max\left(0, \min\left(1 - x_3, \frac{c_{Ammo}}{c_{Nit} + c_{Sulf}} - 1\right)\right) \quad \text{for sulfate}$ $x_3 = \frac{c_{Nit}}{c_{Nit} + c_{Sulf}} \quad \text{for nitrate}$ |
| | Bisulfate (i=1) | $\lambda_1 = -4.559088 + 2.8593 \times RH - 0.111201 \times T_{287}$; $\gamma_1 = \min\left(\frac{1}{1 + e^{-\lambda_1}}, 0.08585\right)$ |
| | Sulfate (i=2) | $\lambda_2 = \lambda_1 - 0.369769$; $\gamma_2 = \min\left(\frac{1}{1 + e^{-\lambda_2}}, 0.053\right)$ |
| | Nitrate (i=3) | $\lambda_3 = -0.8107744 + 4.9017 \times RH$; $\gamma_3 = \min\left(\frac{1}{1 + e^{-\lambda_3}}, 0.0154\right)$ |
| | | c_{Ammo} , c_{Nit} , and c_{Sulf} : molar concentration of ammonium, nitrate, and sulfate, respectively |
| | Dry particles | $\gamma_{N_2O_5, mix} = (x_1 + x_2)\gamma_d + x_3 \times \min(\gamma_d, \gamma_3)$ $\lambda_d = -6.133764 + 3.5920 \times RH - 0.196879 \times T_{293}$; $\gamma_d = \min\left(\frac{1}{1 + e^{-\lambda_d}}, 0.0124\right)$ |
| Brown et al. (2006) [‡] (Scheme V in Fig. 10 (e)) | | $\gamma_{N_2O_5} = \frac{4k_{N_2O_5}}{c_{mean}A}$ <ul style="list-style-type: none"> i) 0.017 ± 0.004 (over Ohio and western Pennsylvania, US) ii) < 0.0010 (over eastern Pennsylvania and New Jersey, US) iii) < 0.0016 (over New York, US) $k_{N_2O_5}$: rate constant (s ⁻¹); c_{mean} : mean molecular speed of N ₂ O ₅ (cm s ⁻¹); A : aerosol surface density (μm ² cm ⁻³) |

[†] Modeling study; [‡] Measurement study.

Table 4. Average tropospheric NO₂ columns, standard deviations and the ratios of the $\Omega_{\text{CMAQ,AK}}$ to the Ω_{OMI} , when different γ_{N2O5} parameterizations were applied to the CMAQ model simulations for January.

| Region | Scheme ⁽¹⁾ | n ⁽²⁾ | $\Omega_{\text{CMAQ,AK}}$ ⁽³⁾ | Ω_{OMI} ⁽³⁾ | R= $\Omega_{\text{CMAQ,AK}} / \Omega_{\text{OMI}}$ |
|---------------|-----------------------|------------------|--|--------------------------------------|--|
| CEC | Scheme I | 896 | 11.11 (8.49) ⁽⁴⁾ | 13.32 (9.00) | 0.78 |
| | Scheme II | | 12.40 (9.42) | | 0.87 |
| | Scheme III | | 12.32 (9.35) | | 0.86 |
| | Scheme IV | | 12.21 (9.27) | | 0.85 |
| | Scheme V | | 14.23 (10.08) | | 0.99 |
| CEC2 | Scheme I | 820 | 9.78 (6.14) | 8.05 (6.34) | 1.21 |
| | Scheme II | | 11.37 (6.77) | | 1.41 |
| | Scheme III | | 11.43 (6.82) | | 1.42 |
| | Scheme IV | | 11.24 (6.74) | | 1.40 |
| | Scheme V | | 13.53 (7.70) | | 1.68 |
| SC | Scheme I | 1125 | 2.47 (1.75) | 2.98 (3.09) | 0.83 |
| | Scheme II | | 2.88 (1.88) | | 0.96 |
| | Scheme III | | 2.80 (1.83) | | 0.94 |
| | Scheme IV | | 2.77 (1.82) | | 0.93 |
| | Scheme V | | 3.44 (2.06) | | 1.15 |
| SB | Scheme I | 386 | 5.05 (4.43) | 3.34 (2.55) | 1.51 |
| | Scheme II | | 5.68 (4.83) | | 1.70 |
| | Scheme III | | 5.43 (4.63) | | 1.63 |
| | Scheme IV | | 5.44 (4.65) | | 1.63 |
| | Scheme V | | 6.78 (5.65) | | 2.03 |
| SK | Scheme I | 260 | 6.80 (3.71) | 6.70 (4.64) | 1.01 |
| | Scheme II | | 7.43 (3.83) | | 1.11 |
| | Scheme III | | 7.29 (3.79) | | 1.09 |
| | Scheme IV | | 7.26 (3.79) | | 1.08 |
| | Scheme V | | 8.42 (4.03) | | 1.26 |
| JP1 | Scheme I | 202 | 3.51 (1.75) | 4.35 (2.58) | 0.81 |
| | Scheme II | | 3.96 (1.92) | | 0.91 |
| | Scheme III | | 3.80 (1.86) | | 0.87 |
| | Scheme IV | | 3.81 (1.87) | | 0.88 |
| | Scheme V | | 4.34 (2.03) | | 1.00 |
| JP2 | Scheme I | 192 | 2.69 (2.60) | 4.68 (4.60) | 0.57 |
| | Scheme II | | 2.89 (2.73) | | 0.62 |
| | Scheme III | | 2.81 (2.67) | | 0.60 |
| | Scheme IV | | 2.82 (2.68) | | 0.60 |
| | Scheme V | | 3.18 (2.84) | | 0.68 |
| Entire domain | Scheme I | 12901 | 2.88 (4.82) | 3.23 (5.14) | 0.89 |
| | Scheme II | | 3.27 (5.40) | | 1.01 |
| | Scheme III | | 3.22 (5.38) | | 1.00 |
| | Scheme IV | | 3.20 (5.33) | | 0.99 |
| | Scheme V | | 3.82 (6.22) | | 1.18 |

⁽¹⁾ Scheme I (Dentener and Crutzen, 1993), Scheme II (Riemer et al., 2003), Scheme III (combination of Riemer et al., 2003 and Evans and Jacob, 2005), Scheme IV (Davis et al., 2007), Scheme V (Brown et al., 2006); ⁽²⁾ The number of data; ⁽³⁾ Unit, $\times 10^{15}$ molecules cm^{-2} ; ⁽⁴⁾ Standard deviations of the distributions of tropospheric NO₂ columns

Table 5. Relative changes in the CMAQ-calculated NO₂ columns for several case studies, compared to those from the standard case simulation (Case 1).

| Case | Sensitivity test | Season | Relative change ⁽¹⁾ (%) | | | | | | | | |
|------|---|--|------------------------------------|--------|--------|--------|------------------------|---------|---------|------------------------|--------|
| | | | CEC | CEC2 | SC | SB | SK | JP1 | JP2 | DM | |
| 2 | NO _x seasonal variation (Han et al., 2009) | Spring | 33.46 | 32.44 | 38.47 | 32.65 | (15.31) ⁽²⁾ | (10.94) | (6.68) | {30.67} ⁽³⁾ | |
| | | Summer | -31.16 | -28.99 | -26.37 | -26.42 | (-1.40) | (-1.44) | (-1.00) | {-21.96} | |
| | | Fall | -21.74 | -23.05 | -23.90 | -21.97 | (-2.12) | (-1.20) | (-0.84) | {-18.67} | |
| | | Winter | 21.25 | 22.34 | 22.99 | 65.37 | (12.95) | (8.04) | (7.36) | {23.30} | |
| 3 | Emission strength (REAS v1.11) | Jan. | -32.55 | -48.32 | -31.45 | -58.44 | (-0.72) | (27.04) | (-1.02) | {-30.49} | |
| 4 | γ _{N2O5} | (Scheme I: Dentener and Crutzen, 1993) | Jan. | -9.76 | -14.43 | -11.71 | -7.13 | -6.74 | -7.51 | -4.32 | -10.84 |
| | | (Scheme II: Riemer et al., 2003) | Jan. | 0.72 | -0.54 | 2.69 | 4.54 | 1.91 | 4.23 | 2.87 | 1.52 |
| | | (Scheme IV: Davis et al., 2008) | Jan. | -0.85 | -1.60 | -1.08 | 0.04 | -0.33 | 0.37 | 0.38 | -0.87 |
| | | (Scheme V: Brown et al., 2006) | Jan. | 15.59 | 18.44 | 22.72 | 24.76 | 15.57 | 14.17 | 13.02 | 18.52 |

$$^{(1)} \text{ Relative change (\%)} = \frac{\Omega_{CASE,i} - \Omega_{CASE,1}}{\Omega_{CASE,1}} \times 100$$

^{(2), (3)} Since the sensitivity parameters were applied only to China for the case 2 and 3 simulations, the relative changes in the parentheses over the SK, JP1, and JP2 regions indicate indirect impacts caused by long-range transports of the changes from China. The relative changes in the brackets in the entire domain (DM region) also include such indirect impacts from China.

Table S1. As Table 2, except for applying the seasonal variations of NO_x emission fluxes from Han et al. (2009) to the CMAQ model simulations.

| Region | Season | n ⁽¹⁾ | Ω_{CMAQ} (w/o AKs) ⁽²⁾ | $\Omega_{\text{CMAQ,AK}}$ (w/ AKs) ⁽²⁾ | Ω_{OMI} ⁽²⁾ | NME (w/o AKs) | NME (w/ AKs) |
|---------------|--------|------------------|---|---|--------------------------------------|---------------|--------------|
| CEC | Spring | 900 | 15.28 (7.98) ⁽³⁾ | 8.54 (5.20) ⁽³⁾ | 6.89 (4.07) ⁽³⁾ | 143.17 | 43.31 |
| | Summer | 900 | 4.44 (2.76) | 1.79 (1.21) | 5.29 (3.02) | 35.70 | 66.29 |
| | Fall | 900 | 10.41 (6.15) | 5.62 (4.00) | 9.49 (5.89) | 37.26 | 41.58 |
| | Winter | 900 | 20.22 (11.19) | 13.44 (9.08) | 14.18 (8.05) | 63.21 | 33.44 |
| CEC2 | Spring | 820 | 13.85 (7.47) | 6.35 (4.95) | 4.45 (3.98) | 211.06 | 49.74 |
| | Summer | 820 | 4.28 (4.52) | 1.64 (2.12) | 3.02 (2.15) | 52.54 | 49.80 |
| | Fall | 820 | 9.44 (6.05) | 4.49 (3.57) | 4.97 (3.97) | 91.89 | 30.49 |
| | Winter | 820 | 24.38 (7.59) | 13.75 (6.41) | 8.49 (5.79) | 187.02 | 64.48 |
| SC | Spring | 1125 | 5.27 (3.79) | 1.60 (1.38) | 2.20 (2.03) | 143.60 | 42.03 |
| | Summer | 1124 | 1.94 (1.88) | 0.56 (0.63) | 1.77 (1.73) | 40.26 | 68.33 |
| | Fall | 1125 | 2.84 (2.17) | 0.97 (0.78) | 2.20 (2.31) | 46.96 | 56.34 |
| | Winter | 1125 | 11.01 (4.71) | 3.95 (2.21) | 3.24 (3.39) | 241.40 | 48.83 |
| SB | Spring | 408 | 5.65 (3.75) | 2.04 (1.44) | 2.56 (1.55) | 129.04 | 36.98 |
| | Summer | 420 | 1.71 (1.20) | 0.58 (0.43) | 2.14 (0.99) | 32.91 | 73.01 |
| | Fall | 418 | 4.91 (3.45) | 1.83 (1.38) | 2.71 (2.15) | 96.49 | 45.96 |
| | Winter | 403 | 18.47 (12.17) | 8.78 (6.54) | 3.43 (3.01) | 438.87 | 160.21 |
| SK | Spring | 260 | 10.16 (6.06) | 5.70 (3.85) | 5.24 (3.74) | 94.08 | 28.25 |
| | Summer | 260 | 7.45 (7.90) | 3.02 (3.57) | 3.41 (2.58) | 126.90 | 43.06 |
| | Fall | 260 | 8.71 (6.59) | 4.51 (3.69) | 4.81 (3.62) | 91.61 | 39.04 |
| | Winter | 260 | 13.51 (5.77) | 7.70 (3.52) | 6.68 (4.14) | 105.18 | 36.27 |
| JP1 | Spring | 204 | 5.01 (1.47) | 2.25 (0.73) | 3.58 (2.48) | 53.34 | 42.50 |
| | Summer | 204 | 2.46 (1.06) | 0.76 (0.33) | 2.91 (1.98) | 34.93 | 73.80 |
| | Fall | 204 | 4.58 (1.93) | 1.89 (0.91) | 3.57 (2.50) | 41.01 | 48.68 |
| | Winter | 204 | 8.15 (3.04) | 3.75 (1.56) | 4.48 (3.07) | 85.57 | 36.53 |
| JP2 | Spring | 285 | 4.11 (3.32) | 1.84 (1.78) | 3.09 (2.96) | 40.45 | 42.56 |
| | Summer | 286 | 2.40 (2.10) | 0.85 (0.82) | 2.64 (2.77) | 30.22 | 68.04 |
| | Fall | 286 | 3.94 (3.34) | 1.62 (1.66) | 3.12 (3.17) | 31.45 | 48.14 |
| | Winter | 279 | 6.20 (4.58) | 2.74 (2.44) | 3.92 (4.20) | 63.79 | 41.80 |
| Entire domain | Spring | 15175 | 3.88 (5.70) | 1.77 (3.10) | 1.97 (2.43) | 119.82 | 49.39 |
| | Summer | 15207 | 1.38 (2.41) | 0.50 (1.00) | 1.59 (1.72) | 49.35 | 70.20 |
| | Fall | 15224 | 2.69 (4.12) | 1.18 (2.19) | 2.06 (3.05) | 55.88 | 48.79 |
| | Winter | 14075 | 7.28 (8.89) | 3.62 (5.54) | 3.20 (4.79) | 135.63 | 50.42 |

⁽¹⁾ The number of data; ⁽²⁾ Unit, $\times 10^{15}$ molecules cm^{-2} ; ⁽³⁾ Standard deviations of the distributions of tropospheric NO₂ columns

Table S2. Average tropospheric NO₂ columns, standard deviations and the ratios of the $\Omega_{\text{CMAQ,AK}}$ to the Ω_{OMI} , when the INTEX-B and REAS NO_x emissions were applied into China for January.

| Region | Inventory for China | n ⁽¹⁾ | $\Omega_{\text{CMAQ,AK}}$ ⁽²⁾ | Ω_{OMI} ⁽²⁾ | R= $\Omega_{\text{CMAQ,AK}} / \Omega_{\text{OMI}}$ |
|---------------|---------------------|------------------|--|--------------------------------------|--|
| CEC | INTEX-B | 896 | 12.32 (9.35) ⁽³⁾ | 14.32 (9.00) | 0.86 |
| | REAS | | 8.31 (6.42) | | 0.58 |
| CEC2 | INTEX-B | 820 | 11.43 (6.82) | 8.05 (6.34) | 1.42 |
| | REAS | | 5.91 (4.36) | | 0.73 |
| SC | INTEX-B | 1125 | 2.80 (1.83) | 2.98 (3.09) | 0.94 |
| | REAS | | 1.92 (1.51) | | 0.64 |
| SB | INTEX-B | 386 | 5.43 (4.63) | 3.34 (2.55) | 1.63 |
| | REAS | | 2.26 (1.69) | | 0.68 |
| SK | INTEX-B | 260 | 7.29 (3.79) | 6.70 (4.64) | 1.09 |
| | REAS | | 7.24 (4.47) | | 1.08 |
| JP1 | INTEX-B | 202 | 3.80 (1.86) | 4.35 (2.58) | 0.87 |
| | REAS | | 4.83 (2.92) | | 1.11 |
| JP2 | INTEX-B | 192 | 2.81 (2.67) | 4.68 (4.60) | 0.60 |
| | REAS | | 2.78 (2.49) | | 0.59 |
| Entire domain | INTEX-B | 12901 | 3.22 (5.38) | 3.23 (5.14) | 1.00 |
| | REAS | | 2.24 (3.56) | | 0.69 |

⁽¹⁾ Number of data; ⁽²⁾ Unit, $\times 10^{15}$ molecules cm^{-2} ; ⁽³⁾ Standard deviations of the distributions of tropospheric NO₂ columns

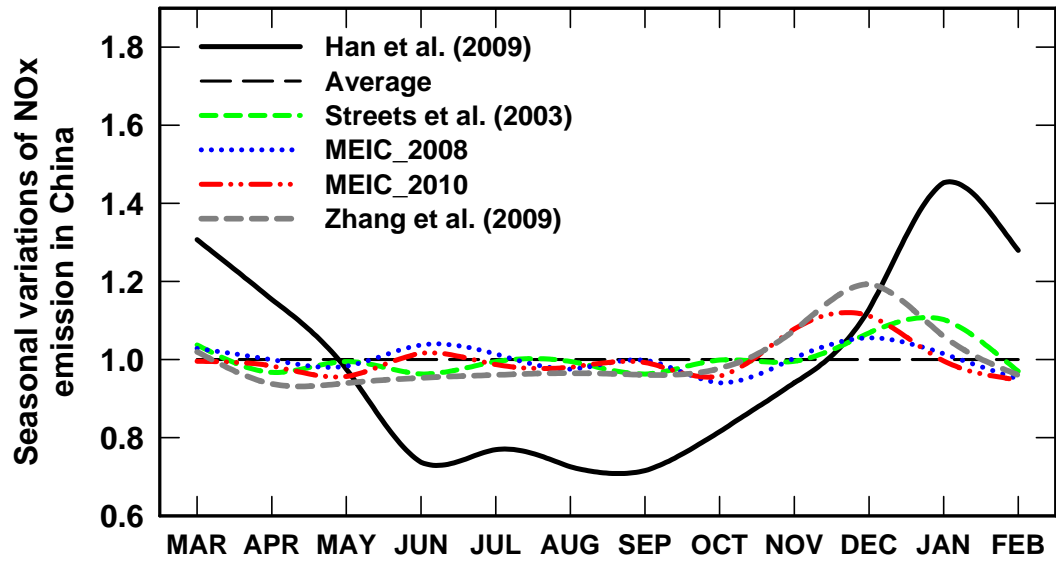


Fig. 1

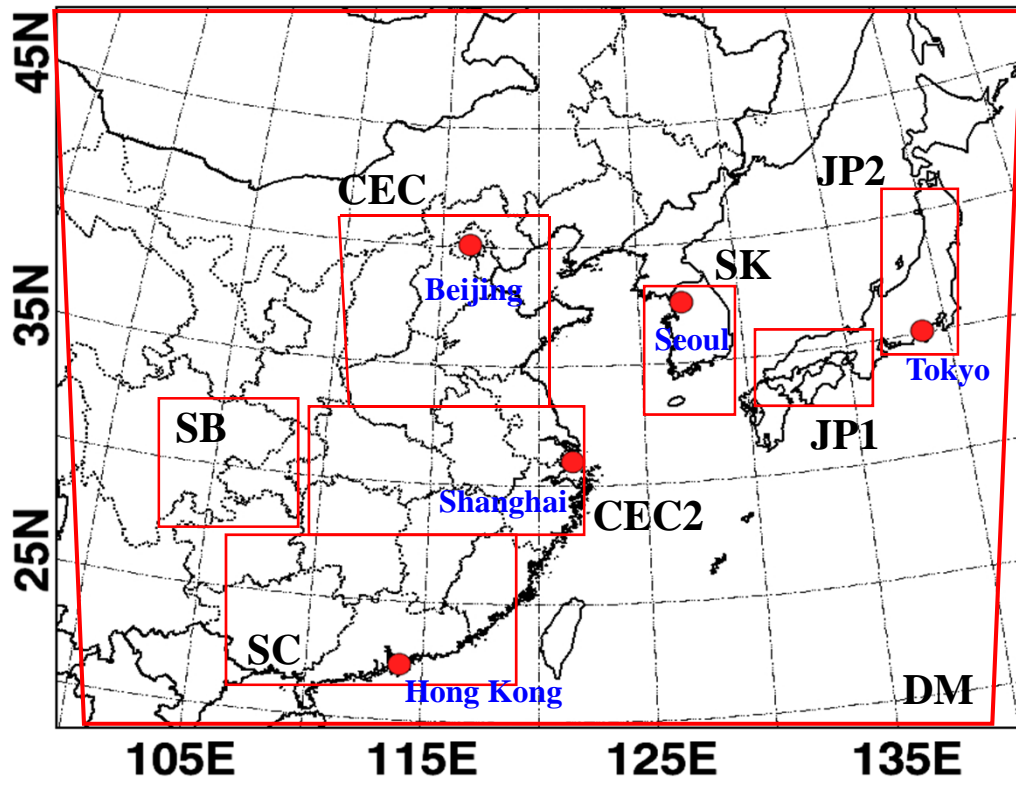


Fig. 2

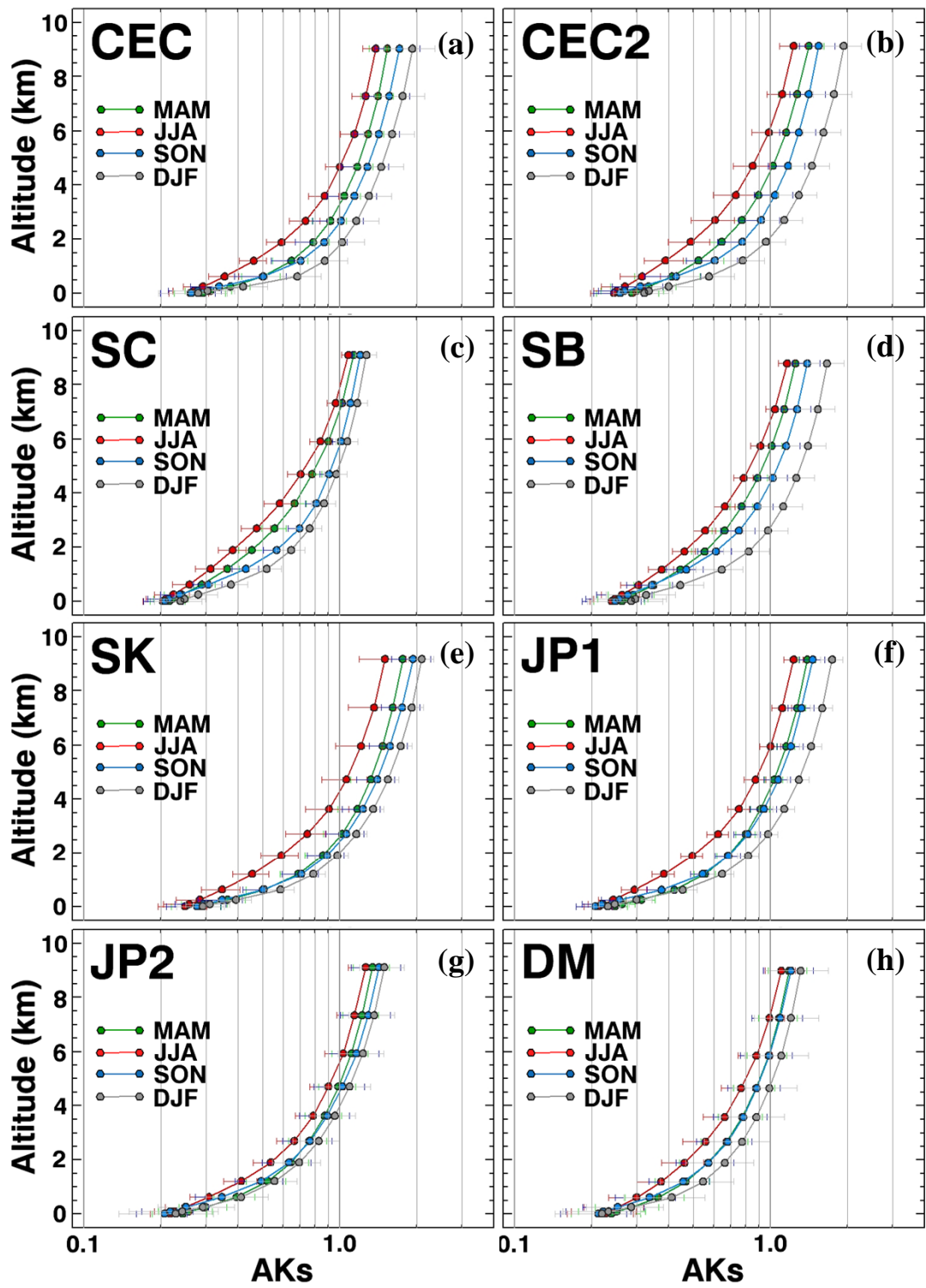


Fig. 3

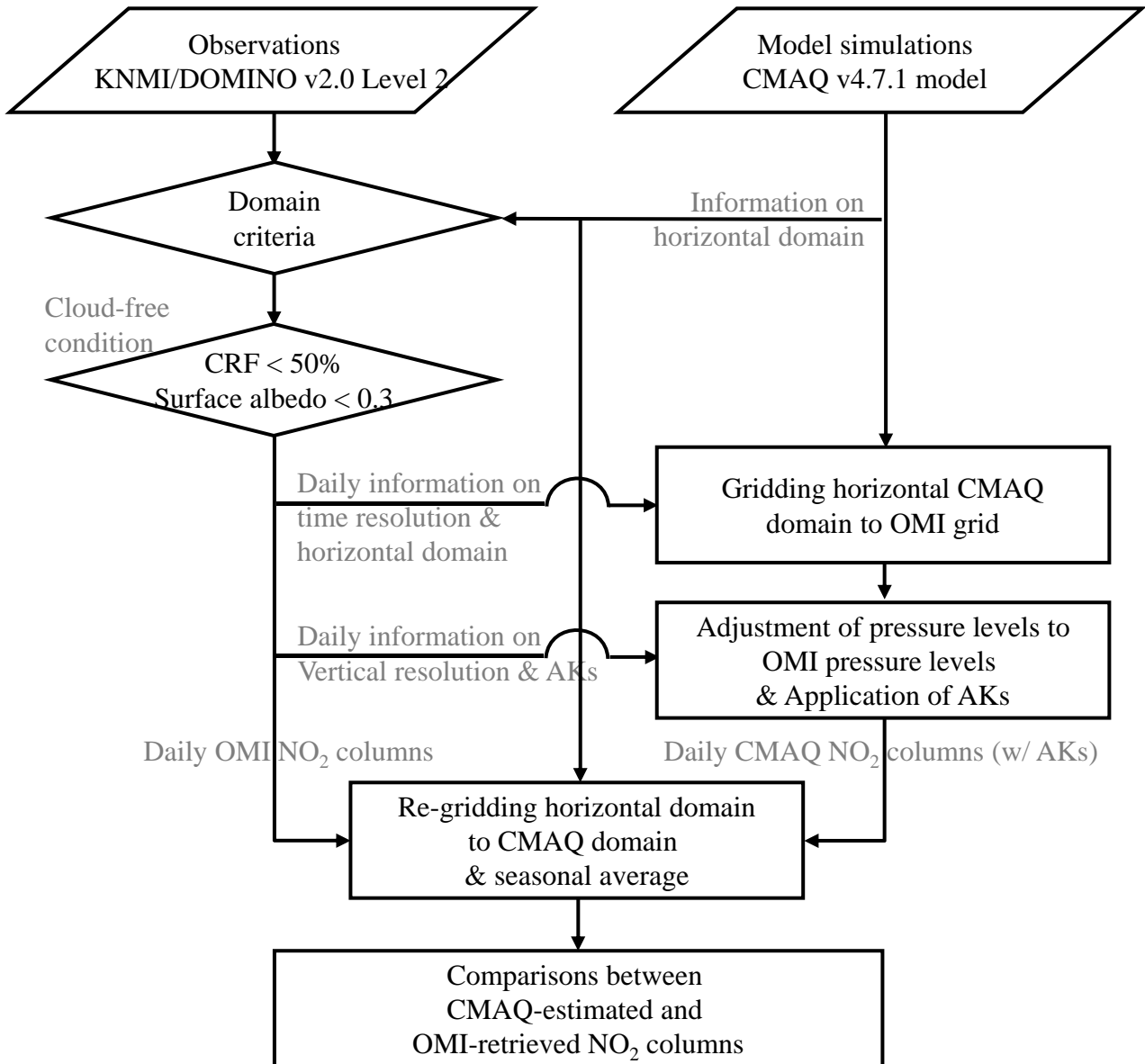


Fig. 4

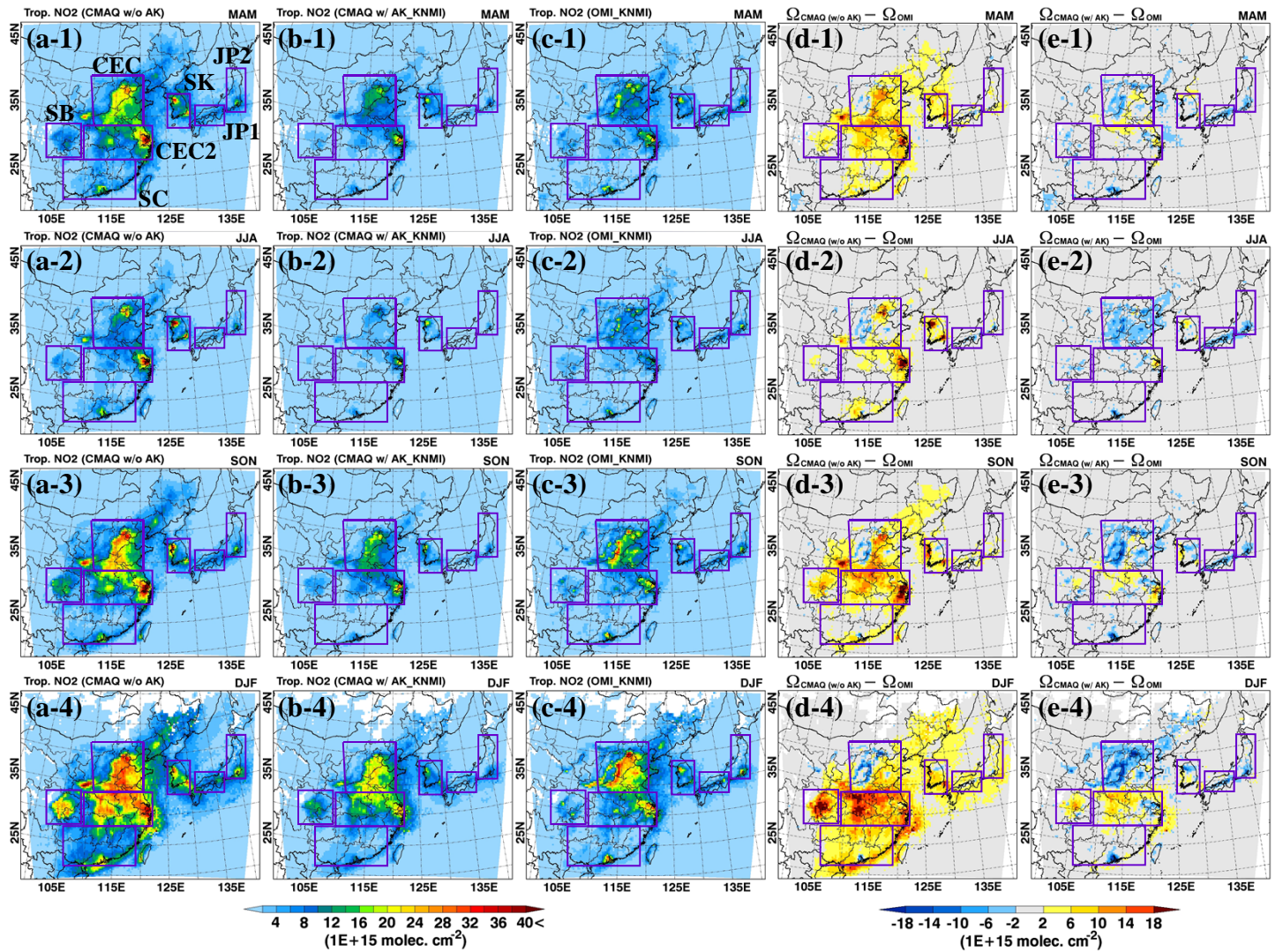


Fig. 5

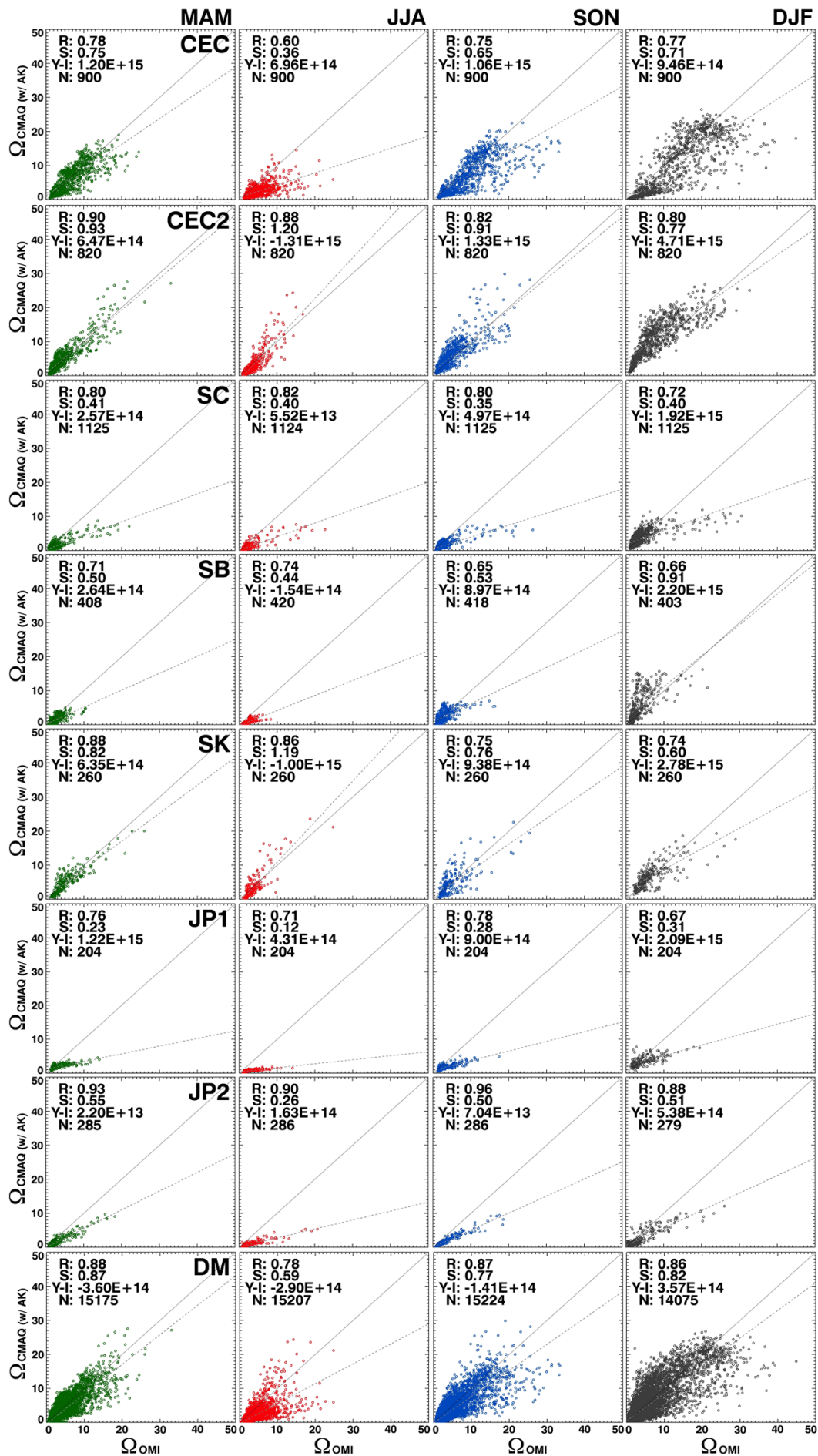


Fig. 6

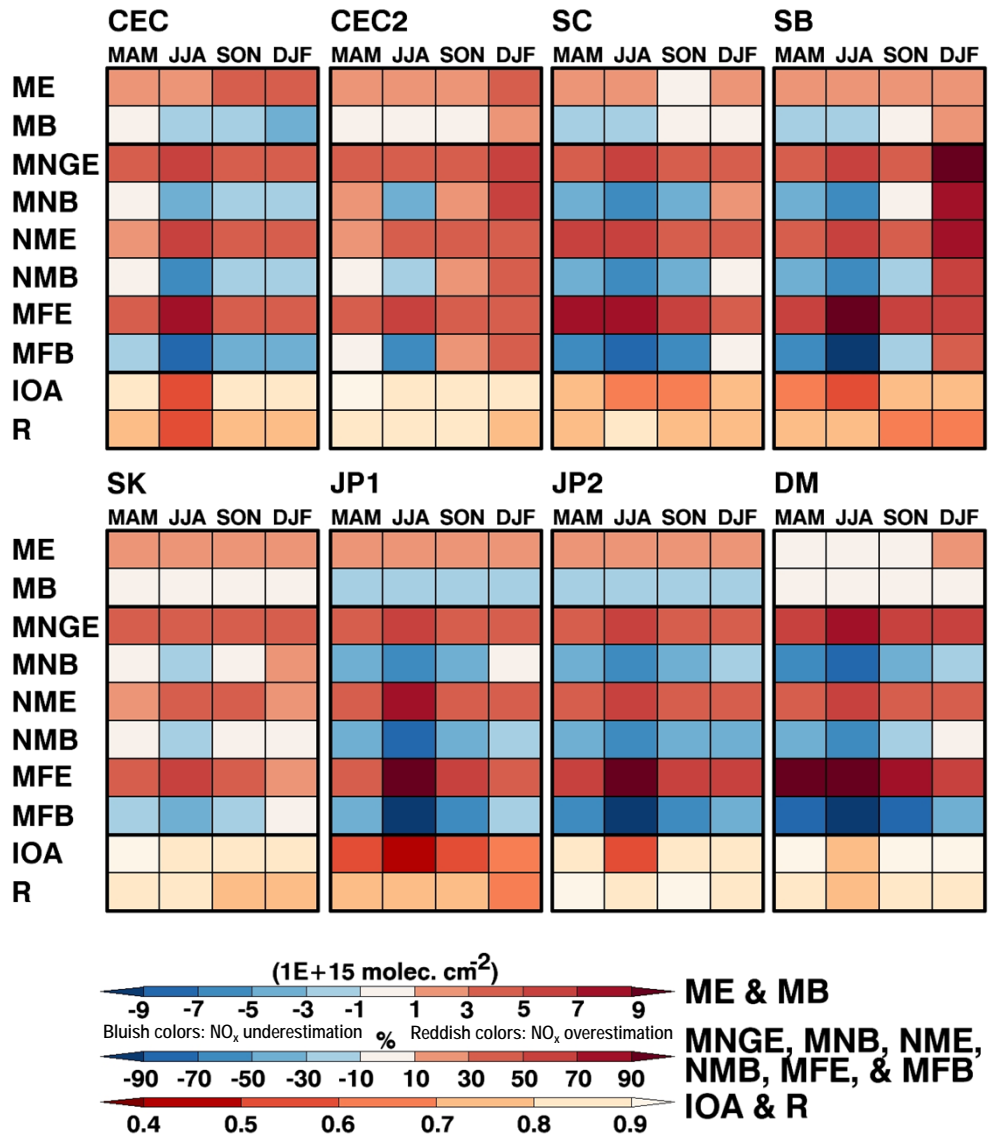


Fig. 7

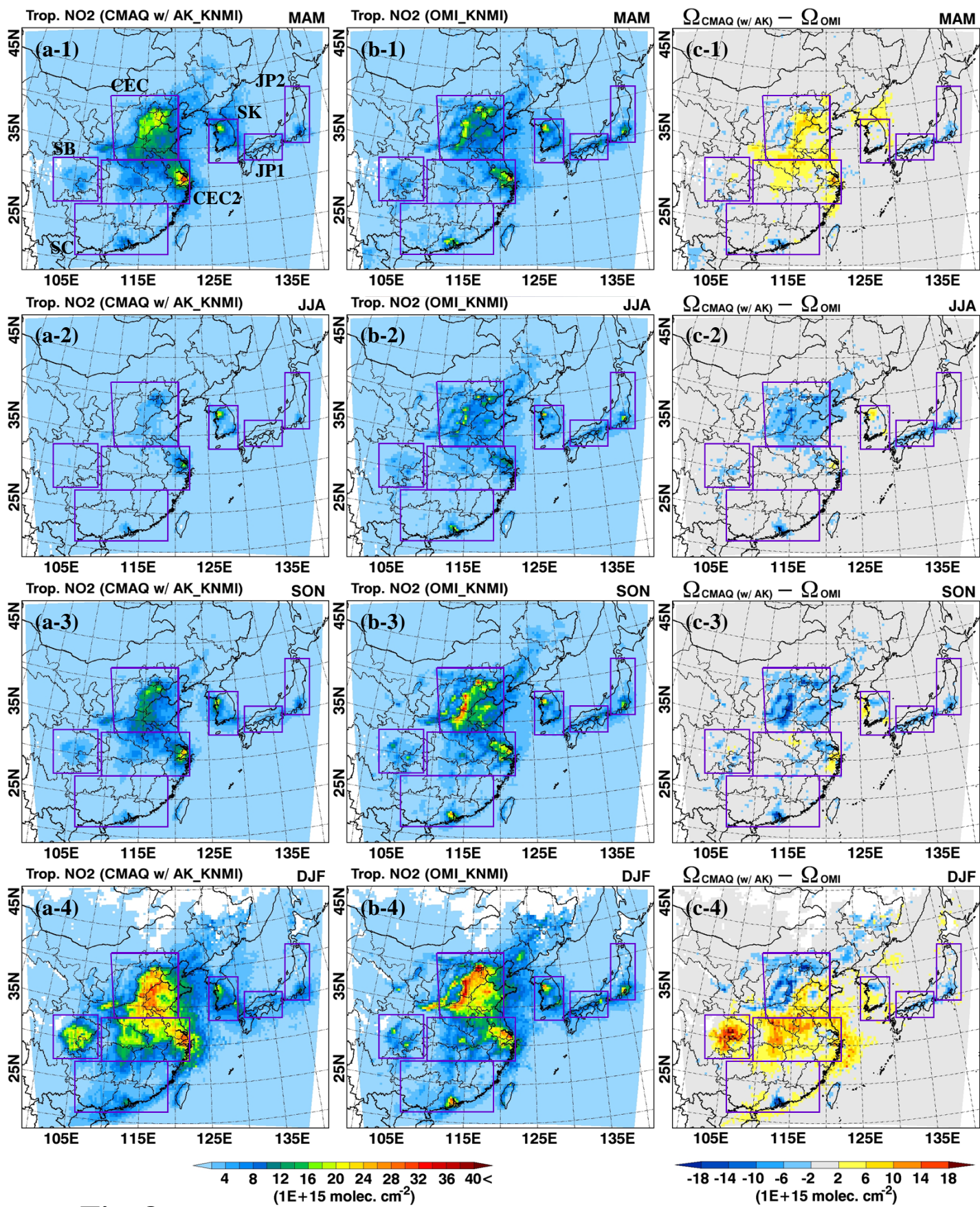


Fig. 8

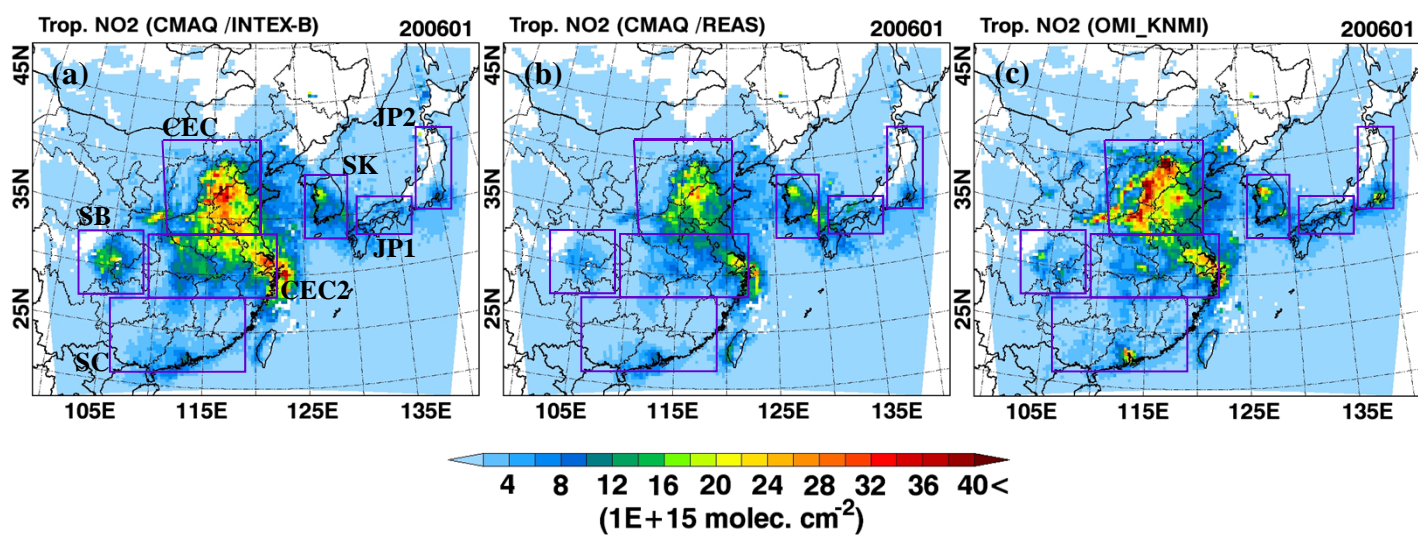


Fig. 9

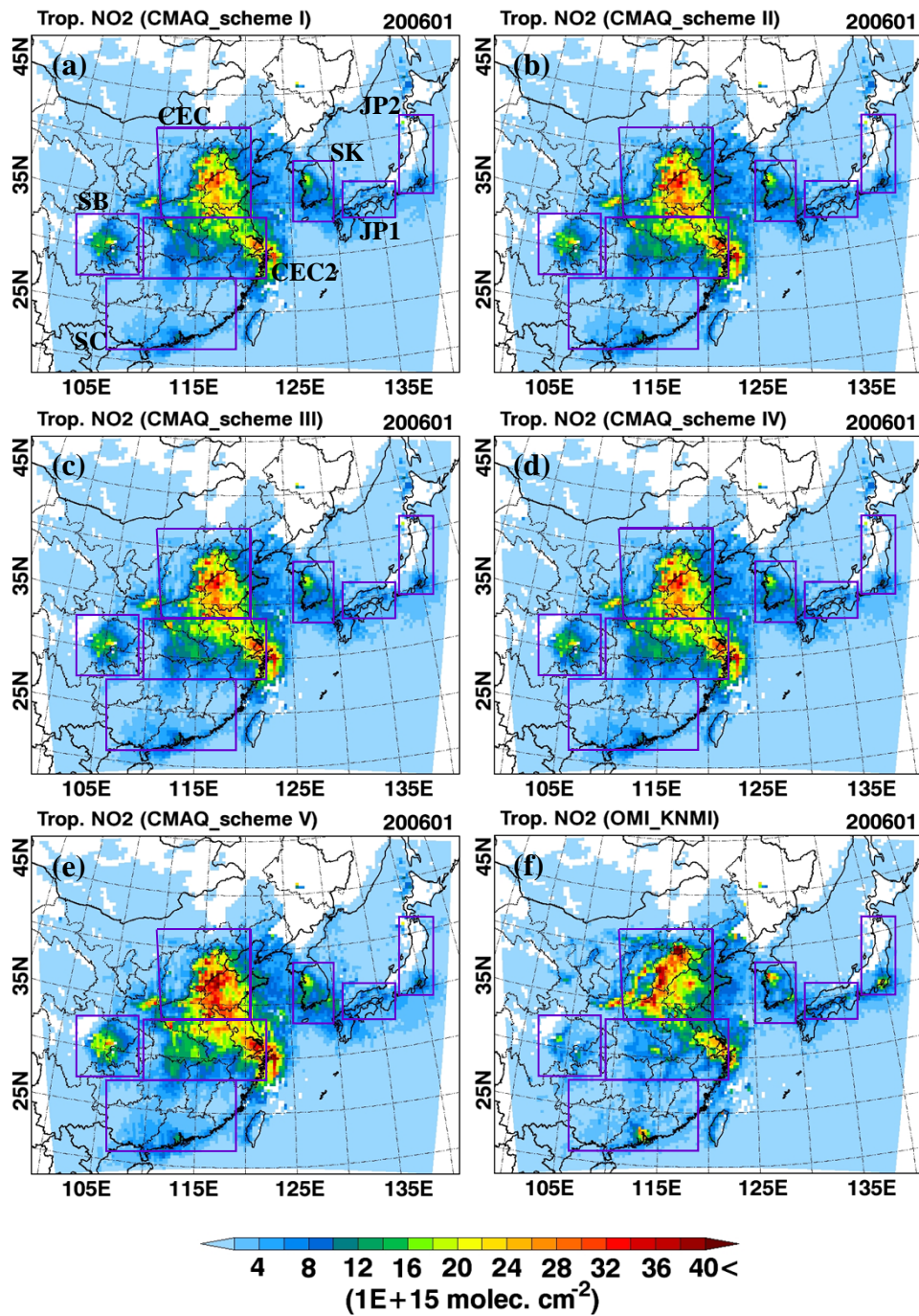


Fig. 10

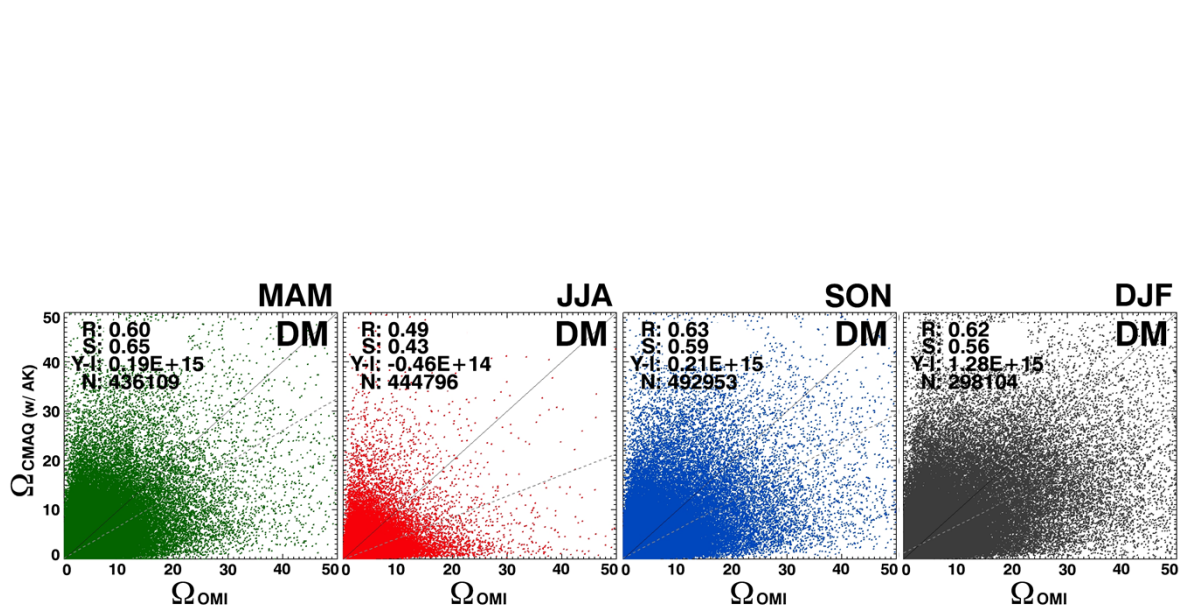


Fig. S1

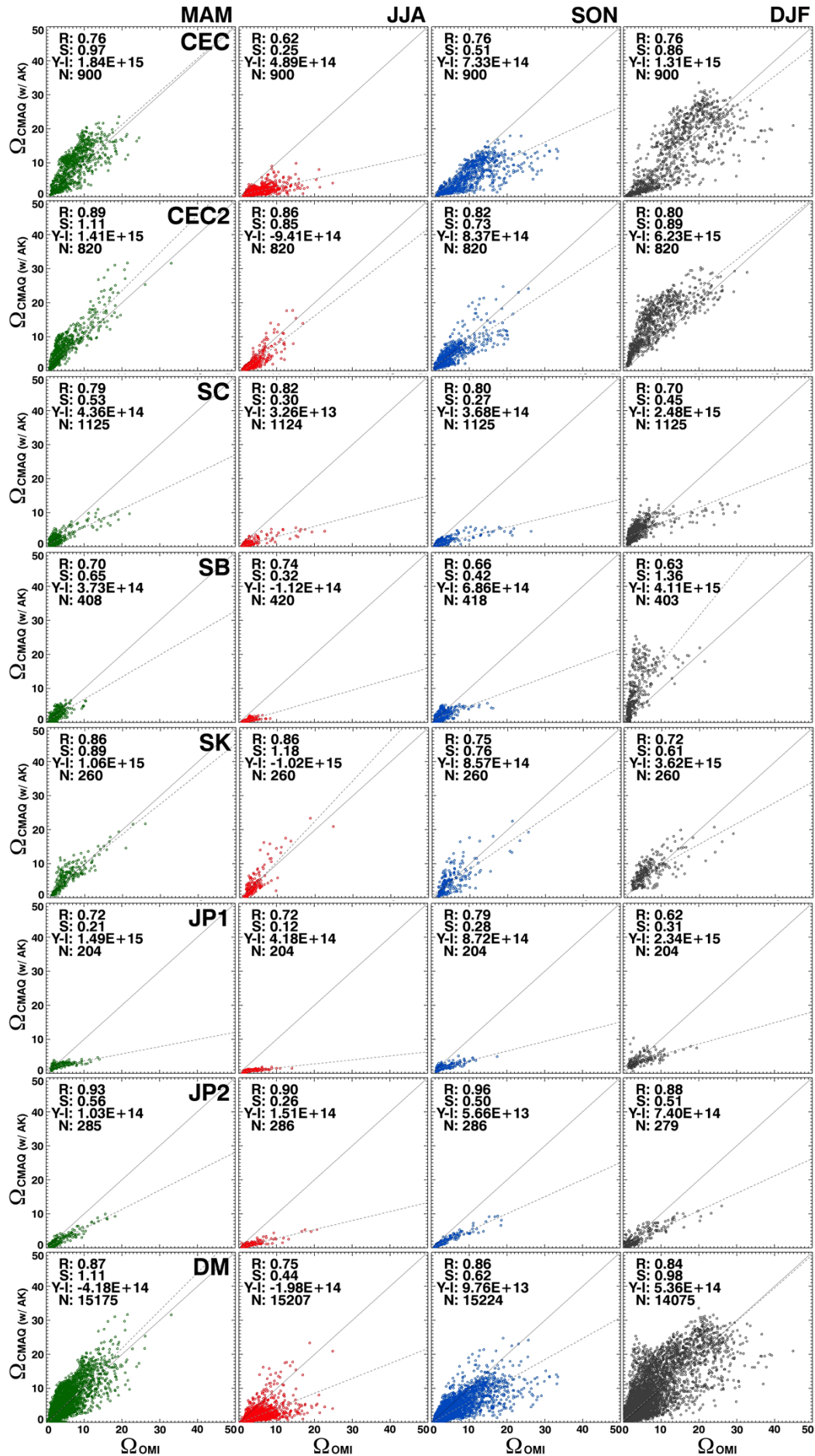


Fig. S2

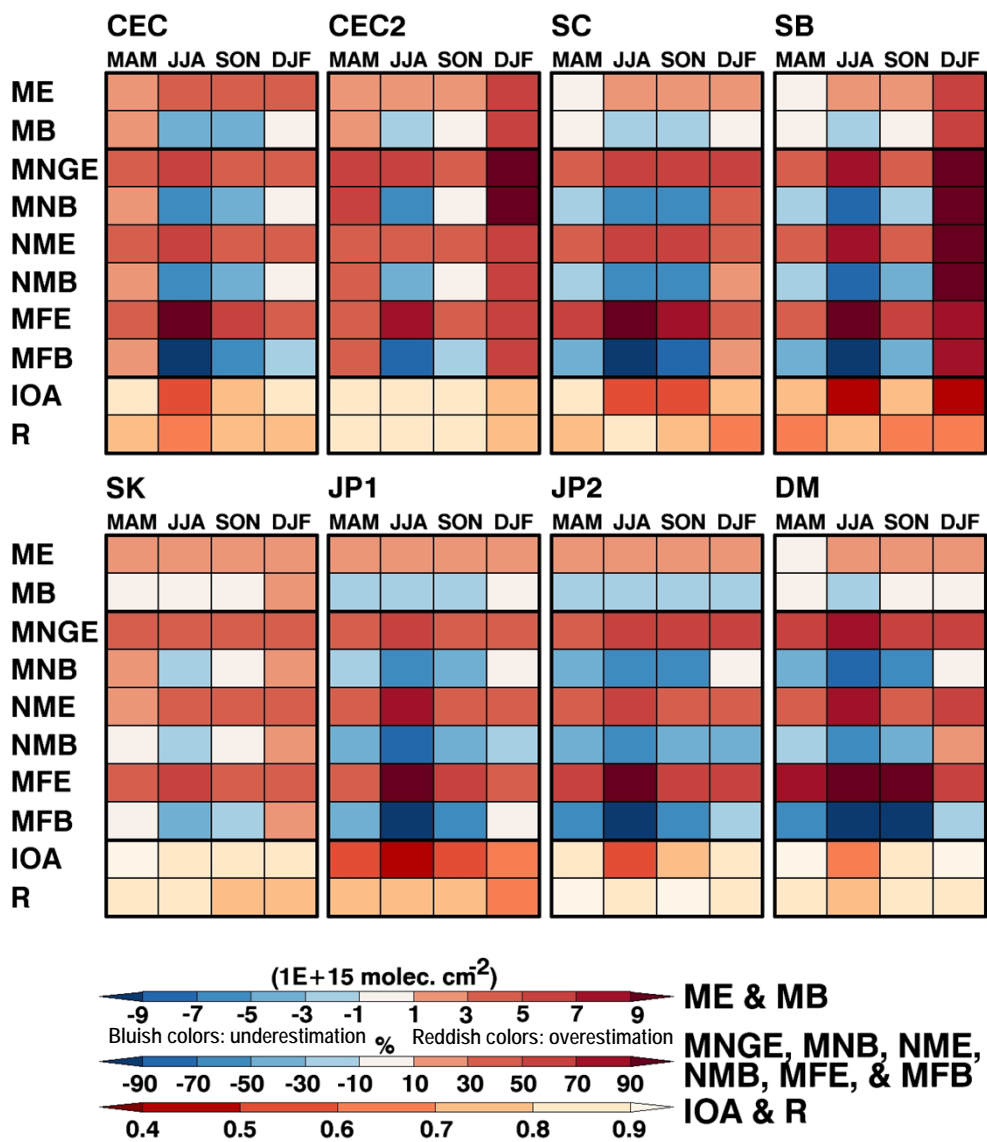


Fig. S3