1	Dicarboxylic Acids, Ketocarboxylic acids, α -Dicarbonyls, Fatty Acids
2	and Benzoic Acid in PM _{2.5} aerosol collected during CAREBeijing-2007:
3	an effect of traffic restriction on air quality
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34 Abstract

Thirty water-soluble organic species, including dicarboxylic acids, ketocarboxylic 35 36 acids, α -dicarbonyls, fatty acids, and benzoic acid were determined as well as organic carbon (OC), elemental carbon (EC) and water-soluble organic carbon (WSOC) in PM_{2.5} samples 37 38 collected during the Campaign of Air Quality Research in Beijing 2007 (CAREBeijing-2007) 39 in the urban and suburban areas of Beijing. The objective of this study is to identify the 40 influence of traffic emissions and regional transport to the atmosphere in Beijing during summer. PM_{2.5} samples collected with or without traffic restriction in Beijing are selected to 41 42 evaluate the effectiveness of local traffic restriction measure on air pollution reduction. The 43 average concentrations of the total quantified bifunctional organic compounds (TQBOC), total fatty acids and benzoic acid during the entire sampling period were 1184 ± 241 ng m⁻³, 44 597 ± 159 ng m⁻³ and 1496 ± 511 ng m⁻³ in PKU, and 1050 ± 303 ng m⁻³, 475 ± 114 ng m⁻³ 45 and 1278 ± 372 ng m⁻³ in Yufa. Oxalic acid (C₂) was found as the most abundant dicarboxylic 46 47 acid at PKU and Yufa, followed by phthalic acid (Ph). A strong even carbon number predominance with the highest level at palmitic acid ($C_{16:0}$), followed by stearic acid ($C_{18:0}$) 48 49 was found for fatty acids. According to the back trajectories modeling results, the air masses 50 were found to originate mainly from northeast, passing over southeast or south of Beijing 51 (heavily populated, urbanized and industrialized areas), during heavier pollution events, 52 whereas they are mainly from north or northwest sector (mountain areas without serious 53 anthropogenic pollution sources) during less pollution events. The data with wind only from 54 the same sector (minimizing the difference from regional contribution) but with and without 55 traffic restriction in Beijing were analyzed to evaluate the effectiveness of local traffic restriction measure on the reduction of local air pollution in Beijing. The results suggested 56 57 that the "traffic restriction" measure can reduce the air pollutants, but the decrease of 58 pollutants is generally smaller in Yufa compared to that in PKU. Moreover, an enhancement 59 of elemental carbon (EC) value indicates elevated primary emissions in Yufa during 60 restriction period than non-restriction period. This study demonstrates that even when primary exhaust was controlled by traffic restriction, the contribution of secondary organic species 61 62 formed from photochemical processes was critical with long-range atmospheric transport of pollutants. 63

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Keywords: dicarboxylic acids, ketocarboxylic acids, α-dicarbonyls, fatty acids, secondary
 organic carbon, Chinese aerosols

68 1. Introduction

69 Organic aerosol (OA) typically constitutes 20-90% of submicron aerosol (Jimenez et 70 al., 2009; Huang et al., 2014) and is influencing Earth's climate directly by absorbing and 71 scattering radiation and indirectly by acting as cloud condensation. OA also adversely affects 72 air quality and human health. However, uncertainties exist in the effect of OA on health and 73 climate, due in large part to the complexity of the OA composition (Hallquist et al., 2009; 74 Poschl, 2005; Hoffmann et al., 2011). OA can be primary emitted, or secondary produced by 75 photochemical reactions of gas-phase precursors. Due to polar functional groups formation 76 (e.g., carbonyl, carboxyl and hydroxyl), a major fraction of the SOA is thought to be water-77 soluble which, together with some water-soluble POA, accounts for about 40-80% of the OA 78 (Jaffrezo et al., 2005; Saxena and Hildemann, 1996).

79 Despite the dominant presence of WSOC in the atmosphere, there exist large 80 uncertainties associated with sources, the chemical composition, removal mechanisms and 81 atmospheric formation processing of aerosol WSOC. This is particularly evident in polluted 82 megacities where multiple sources of local and regional origins may significantly change the 83 chemical and physical properties of aerosol and therefore influence the air quality, climate 84 and human health. Dicarboxylic acids (diacids) are the most abundant organic compounds in 85 OA, which can be derived from primary emissions and/or secondary formation from different 86 precursor species via photochemical reactions (Glasius et al., 2000; Kawamura et al., 1996; 87 Kundu et al., 2010; Legrand et al., 2007). Fossil fuel combustion and biomass burning 88 (Falkovich et al., 2005; Ho et al., 2006; Kundu et al., 2010; Huang et al., 2014) are the major 89 primary sources whereas photochemical oxidation of volatile organic compounds (VOCs) 90 from biogenic and anthropogenic emissions (Kawamura et al., 1996; Mkoma and Kawamura, 91 2013) are the major secondary sources.

92 Beijing is one of the largest metropolitan cities in Asia and has become a heavily 93 polluted area due to the fast urbanization and industrialization over the past two decades. In 94 2009, more than 17.5 million residents and 4.0 million vehicles were reported in Beijing 95 (BMBS, 2010). Besides local emissions, the air flowed into Beijing from polluted 96 neighboring regions can have significant impact to the air quality in Beijing (Hatakeyama et 97 al., 2005; Luo et al., 2000; Mauzerall et al., 2000). Especially, the gas-to-particle partitioning 98 of semi-volatile organic compounds (SVOCs) and their subsequent aging via photochemical 99 processing during transport has been recognized to be a major air pollution source (Ding et 100 al., 2008; Guttikunda et al., 2005). Atmospheric aerosols have been investigated extensively 101 in China (An et al., 2007; Cao et al., 2003; Xu et al., 2008; Huang et al., 2014). However,

102 relevant studies on organic acids are still very scarce. With such limited information available 103 on organic acids despite the rapid urbanization and development (especially the increase in 104 traffic density), it is essential to seek a better understanding of organic acids in Beijing. For 105 the promised "Green Olympic Game" in 2008, many pollution control measures, such as 106 controlling traffic, halting industrial/construction activities, and sweeping roads, was taken to 107 improve the air quality. The 'traffic restriction' measure, which only allowed vehicles to be 108 on road in alternative business days according to their even and odd plate numbers, was 109 proposed to reduce air pollution.

110 To investigate the effects of the traffic restriction on the air quality of Beijing and to 111 accumulate experience and scientific evidence for the preparation of the 2008 Olympic game, 112 we conducted aerosol (PM_{2.5}) monitoring at two sites in Beijing during August 3–31, 2007. In 113 this study, PM_{2.5} samples collected were analyzed by gas chromatography flame ionization 114 detector (GC-FID) and gas chromatography mass spectrometry (GC-MS) to determine the 115 composition of low molecular weight (MW) diacids (C_2-C_{12}), ketocarboxylic acids ($\omega C_2-\omega C_9$, 116 pyruvic acid), α-dicarbonyls (C₂-C₃), benzoic acid and fatty acids (C₁₂-C₂₅). Moreover, OC, 117 EC, and WSOC were also analyzed. Through the intensive sampling campaign, the roles of 118 regional transport, local emissions and secondary formations of particulate matter in the 119 atmosphere of Beijing were investigated.

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121 **2.** Experiment

122 2.1 *PM*_{2.5} sampling

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Two sampling locations, Peking University (PKU) (39.98°N, 116.35°E) and Yufa 124 125 (39.51°N, 116.31°E) were selected in this study. The detailed descriptions of the sampling 126 locations were reported elsewhere (Ho et al., 2010). The air samplers were placed on the top 127 floor of the buildings (PKU: a 6-story building; Yufa: 4-story building). The meteorological 128 data such as wind speed, wind direction, relative humidity, and temperature were collected 129 during the sampling period. North and northwest of PKU are enclosed by mountains whereas 130 south and southeast of Yufa are surrounded by heavily industrialized and urbanized areas 131 such as Hebei province and Tianjin city.

Pre-heated (800 °C, 3h) quartz-fiber filters (47 mm QM-A Whatman quartz filters) were used to collect 24 h integrated $PM_{2.5}$ samples by Airmetrics mini-volume $PM_{2.5}$ samplers at a flow rate of 5 L min⁻¹. A DryCal[®] flow meter (BIOS International, Butler, NJ, USA) was used to calibrate the sampling flows before and after the sampling. Sampling was carried out simultaneously from 09:00 a.m. to 09:00 a.m. local time at the two sampling locations from 3^{rd} to 31^{st} August, 2007. The samples were properly kept in a freezer (-20°C) to prevent evaporation of semi volatile components and microbial degradation of organics.

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140 2.2 *Chemical analysis*

141 OC and EC were analyzed (on a 0.526 cm^2 punch) by thermal analysis with optical 142 detection following the IMPROVE protocol on a Desert Research Institute (DRI) Model 2001 143 Thermal/Optical Carbon Analyzer (Atmoslytic Inc., Calabasas, CA, USA) (Cao et al., 2003; 144 Chow et al., 2005). The method detection limit (MDL) of OC and EC analysis is 0.8 and 0.4 μ gC cm⁻², respectively. Todetermine the WSOC, a total area of 2.63 cm² of the sample filter 145 was cut from each filter and 5 ml of Milli-Q water (18 MQ) was added into a 15 ml vial 146 147 where the sample was placed. Ultra-sonic water bath was used to extract the particles on the 148 filter for 1 hr. Syringe filters (0.2 µm PTFE membrane) were used to remove the insoluble 149 particles from the extracts. Filtered extract was then transferred into clean vials and analyzed total organic carbon (TOC) by using a Shimadzu TOC-V CPH Total Carbon Analyzer 150 (Columbia, MD, USA). The MDL is 0.01 μ gC m⁻³, with a precision of ±5%. The data 151 152 reported in this study were all corrected by the blanks.

153 The analytical procedures for water-soluble organic species were well reported 154 elsewhere (Kawamura and Yasui, 2005). Briefly, the sample was extracted with organic-free 155 water (10 ml \times 3) to isolate bifunctional organic compounds as well as fatty acids and benzoic 156 acid. After the extracts were concentrated using a vacuum rotary evaporator, 14% BF₃/n-157 butanol were added at 100 °C to convert the aldehyde groups to dibutoxy acetals and carboxyl groups to butyl esters. Homologous series of fatty acids were analyzed as butyl esters 158 159 (Mochida et al., 2007). No serious contamination (< 5% of real samples) was observed in our 160 analysis. The data reported in this study were corrected by the field blanks. The derivatized 161 samples were determined by a Agilent 6890GC/FID (Palo Alto, CA, USA) equipped with a 162 split/splitless injector, HP-5 fused silica capillary column (25 m \times 0.2 mm i.d. x 0.5 µm film 163 thickness) and an FID detector. Peak identifications were relied on the retention times of 164 authentic standards. ThermoQuest Trace MS (Austin, TX, USA) with a similar GC conditions 165 was used for mass spectral confirmation of the compounds. The reproducibility of the 166 methods was $< \pm 15\%$; recoveries of the bifunctional organic compounds fatty acids and 167 benzoic acid were > 70% (Kawamura and Yasui, 2005; Mochida et al., 2007). Field blanks 168 concentrations were < 15% of real samples, except for phthalic acid (up to 30%). The results 169 shown in this study were all corrected by the field blanks.

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171 **3. Results and discussion**

172 3.1 Overview of molecular compositions of bifunctional organic compounds in PKU and
173 Yufa

174 Average OC, EC, and WSOC concentrations in PKU and Yufa are illustrated in Table 175 1 and their levels during the entire sampling period were 14.9 ± 2.47 , 6.21 ± 1.90 and $5.59 \pm$ 1.49 μ gC m⁻³ in PKU, and 11.1 \pm 3.68, 5.6 \pm 1.83 and 4.55 \pm 1.79 μ gC m⁻³ in Yufa, 176 respectively. The WSOC accounted for $37 \pm 7\%$ and $40 \pm 7\%$ of OC in PKU and Yufa, 177 178 respectively. It was consistent with the WSOC/OC ratios (20-40%) at other metropolitan 179 cities (Ho et al., 2007; Yang et al., 2005), suggesting that WSOC is one of the main components in OA in China. Yufa is located at southern Beijing, which is close to the border 180 181 of Beijing Municipality and Hebei Province. Regional pollution from heavy industrialized and 182 urbanized areas, like Hebei province and Tianjin city, have a great impact to the air quality of 183 Yufa area.

184 The concentrations of bifunctional organic compounds measured in PKU and Yufa are 185 presented in Table 1. The concentrations of total quantified bifunctional organic compounds (TQBOC) varied from 730 to 1455 ng m⁻³ (average concentration: 1184 ± 241 ng m⁻³) in PKU. 186 and from 554 to 1621 ng m⁻³ (average concentration: 1050 ± 303 ng m⁻³) in Yufa. The results 187 are higher than measurements (average 813 ng m⁻³ in PKU; average 771 ng m⁻³ in Yufa) 188 reported in 2006 in same sampling locations (Ho et al., 2010), reflecting that there were 189 190 continuous increases of primary emissions and more aging of aerosols in Beijing. However, 191 the concentrations are close to those megacities studied recently (Ho et al., 2007).

Oxalic acid (C₂) was the most abundant diacid (435 ± 124 ng m⁻³ and 418 ± 130 ng m⁻³ 192 193 3 at PKU and Yufa, respectively) determined in this study, followed by phthalic acid (Ph) $(209 \pm 28.8 \text{ ng m}^{-3} \text{ and } 176.3 \pm 91.5 \text{ ng m}^{-3})$, and succinic acid (C₄) $(89.9 \pm 27.7 \text{ ng m}^{-3} \text{ and } 176.3 \pm 91.5 \text{ ng m}^{-3})$ 194 80.9 ± 26.9 ng m⁻³). These three species accounted for 65% of TQBOC in PKU and Yufa, 195 196 respectively. Oxalic acid was also recognized as predominant diacid in previous studies in 197 China (Ho et al. 2010, 2011). C₂ can be either released from combustion processes (e.g., fossil 198 fuel and biomass burning) (Kawamura and Kaplan, 1987; Narukawa et al., 1999) or 199 secondary produced by the oxidation of VOCs (Carlton et al., 2006; Warneck, 2005).

The average phthalic acid (Ph) concentrations measured in this study are higher than those reported by other studies (Ho et al., 2007; Wang and Kawamura, 2005). Three phthalic acids (phthalic acid (*o*-isomer), terephthalic acid (*p*-isomer) and isophthalic acid (*m*-ismoer)) were determined and these isomer species distribution was dominated by *o*-isomer, followed by *p*-isomer and *m*-isomer, which are consistent with studies measured in Mt. Tai, China and Pearl River Delta region (Fu et al., 2008; Ho et al., 2011). The abundant phthalic acid (Ph) can be released from incomplete combustion processes or secondary formed by oxidation of aromatic compounds (e.g., naphthalenes, NAP) (Kawamura and Kaplan, 1987; Kawamura and Yasui, 2005). In some previous studies, high levels of NAP were found in Beijing urban areas (Liu et al., 2007; Tao et al., 2007) and it can be thus one of the potential precursors to phthalic acid (Ph) formation (Ho et al., 2007).

Besides diacids (C_2-C_4) , azelaic acid (C_9) was the most abundant species among the 211 saturated diacids in both sampling locations (71.4 \pm 8.91 ng m⁻³ in PKU; 49.2 \pm 8.99 ng m⁻³ in 212 Yufa). C₉ is recognized as a photochemical reaction product of biogenic unsaturated fatty 213 214 acids, such as oleic $(C_{18:1})$ and linoleic $(C_{18:2})$ acids [the first number is carbon number and the 215 latter one refers to the amount of double bond] (Kawamura and Gagosian, 1987) and is 216 generally abundance in the high molecular weight homologues. The unsaturated fatty acids 217 are commonly determined in marine micro-organism or higher plant leaves. However, these 218 acids could be released by meat charbroiling also (Rogge et al., 1991). During long range 219 transport, photochemical oxidation of $C_{18:1}$ to C_9 via oxidants (e.g., ozone and/or OH radicals) 220 may occur in the air (Stephanou and Stratigakis, 1993). The C₉/C_{18:1} ratio determined was lower in Yufa (average value: 2.1) than that in PKU (average value: 5.1) suggesting that 221 222 significant secondary production of C₉ occurred in urban area of Beijing.

223 Diacids can be formed when ketocarboxylic acids, which are regarded as intermediate 224 product of mono-carboxylic acids oxidation, react with other pollutants in the air (He et al., 2013; Kawamura et al., 1996). The concentrations of total measured ketocarboxylic acids 225 varied from 87.4 to 169 ng m⁻³, (average value: 122 ± 28.8 ng m⁻³) in PKU and from 52.0 to 226 131 ng m⁻³ (average value: 97.0 \pm 22.9 ng m⁻³) in Yufa (Table 1). The concentrations in both 227 228 sampling sites are higher than those measured in rural site in Gosan. South Korea (53 ng m^{-3}) and megacities in China (summer: 37 ng m⁻³) (Ho et al., 2007; 2010; Kawamura et al., 2004). 229 230 These results reveal that the organic aerosols in PKU and Yufa were likely more 231 photochemically aged than that in other urban sites caused by photochemical reaction during 232 transportation (He et al., 2013). Glyoxylic acid (ωC_2) was found as the most abundant 233 ketocarboxylic acid, followed by pyruvic acid (Pyr). Their concentration levels are similar to 234 previous measurement in Tokyo, Japan (Kawamura and Yasui, 2005).

235 Concentrations of total measured α -dicarbonyls varied from 35.5 to 99.5 ng m⁻³ 236 (average value: 51.8 ± 17.9 ng m⁻³) in PKU and from 29.0 to 61.4 ng m⁻³ (average value: 44.2 237 ± 10.3 ng m⁻³) in Yufa. The two simplest α -dicarbonyl compounds (Glyoxal and 238 methylglyoxal) have recently attracted much attention as potential SOA precursors. These 239 compounds are formed by photochemical oxidation of both biogenic (e.g., isoprene and 240 terpenes) and anthropogenic VOCs (e.g., toluene, xylene) (Fick et al., 2004; Volkamer et al., 241 2001). They have been identified as the significant precursors in the heterogeneous processes 242 for SOA formation (Kroll et al., 2005). High concentrations of glyoxal and methylglyoxal 243 observed indicate the greater SOA formation potential in this region. α-Dicarbonyls levels 244 measured in PKU and Yufa were higher than previous results in other cities of China (average value: 12 ng m⁻³) (Ho et al., 2007). It indicates that the biogenic sources such as oxidation of 245 isoprene are more important than other urban cities in China. 246

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3.2 Overview of molecular compositions of fatty acids and benzoic acid in PKU and Yufa

249 Table 1 presents the average concentrations of straight chain saturated fatty acids 250 $(C_{12:0}-C_{25:0})$, the first number is carbon-number and the latter one refers to the amount of 251 double bonds), unsaturated fatty acid and benzoic acid. Total measured fatty acids concentrations varied from 459 to 1003 ng m⁻³ (average value: 597 ± 159 ng m⁻³) in PKU and 252 from 375 to 684 ng m⁻³ (average value: 475 ± 114 ng m⁻³) in Yufa. The distributions of fatty 253 254 acids were dominated by even carbon number with maximum at palmitic acid ($C_{16:0}$), 255 followed by stearic acid ($C_{18:0}$). This finding is consistent with previous measurements 256 reported in megacities of China (Fu et al., 2008; Ho et al., 2010). Both natural biogenic and 257 anthropogenic emissions represent the major sources of fatty acids, whereas, homologs $< C_{20}$ 258 are partially released from microbial sources (Simoneit and Mazurek, 1982). Additionally, 259 low MW fatty acids ($< C_{18}$) can be emitted by tire wear debris and traffic exhaust. Biomass 260 burning also produces high fractions of fatty acids which are the major components of plant 261 tissues and surface waxes. C_{16:0} and C_{18:0} were also the major organic compounds emitted from the meat cooking (Schauer et al., 1999, 2002; Zhao et al., 2007a, b). Higher 262 263 concentrations of fatty acids observed in PKU can be explained by the mixed contributions of 264 regional and local emissions in urban area. Interestingly the contributions of total quantified 265 fatty acids to OC are similar in both sites (3.1% in PKU and 3.2% in Yufa, respectively).

266 The even-over-odd carbon number preference in fatty acid ($C_{12:0}$ to $C_{25:0}$) is measured 267 by Carbon Preference Index (CPI) :

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$$CPI_{fatty acid} = \frac{\Sigma \text{ Even carbon number fatty acids}}{\Sigma \text{ Odd carbon number fatty acids}}$$

CPI is a measure to differentiate anthropogenic and biogenic sources and the values are 43.3
in PKU and 45.9 in Yufa, respectively. High CPI values observed in this study indicate that
biological sources such as vascular plant has significant influence in this region (Simoneit, 1984).

274 In this study, C_{18:1} was detected in all samples which can be directly emitted from 275 higher plants and soils. In urban areas, biomass burning and cooking are likely to be the main 276 anthropogenic sources for this acid (Rogge et al., 1993). Its concentrations varied from 2.9 to 33.0 ng m⁻³ (average value: 24.3 ± 8.9 ng m⁻³) and from 13.0 to 47.9 ng m⁻³ (average value: 277 24.6 ± 9.2 ng m⁻³) in PKU and Yufa, respectively. Oleic acid is a good tracer for unsaturated 278 279 organic aerosol and a representative compound for reactivity model (Rudich et al., 2007). The 280 diagnostic ratio of $C_{18:1}/C_{18:0}$ was used to determine the level of aerosol aging in this study. 281 Low values indicate that the air masses are more aged. The ratios in PKU and Yufa were 0.12 282 and 0.14, respectively, which suggests that unsaturated fatty acids are depleted by the 283 enhanced photochemical degradation in PKU (Wang et al., 2006). Moreover, the diagnostic 284 ratio of $C_{18:0}/C_{16:0}$ was applied as an indicator for source evaluation. Low ratios observed (< 285 0.25) in PM_{2.5} were likely originated from wood smoke, waxy leaf surface abrasions, and 286 foliar vegetation combustion; ratios that ranged between 0.25-0.5 were indicated for vehicle 287 exhausts; while ratios that ranged between 0.5-1 were obtained from hamburger charbroiling 288 and paved/unpaved road dust (Oliveira et al., 2007; Rogge et al., 2006). The C_{18:0}/C_{16:0} ratios 289 observed in this study had a range between 0.64 - 1.17 (average value: 0.85 in both locations) 290 in PKU and Yufa, indicating that the contribution of cooking emissions and paved/unpaved 291 road dust cannot be ruled out.

Almost all PM2.5 samples collected contained benzoic acid which has been identified 292 293 as a direct pollutant from the traffic emissions (Kawamura et al., 1985) and a indirect 294 pollutant produced from photo-degradation of aromatic compounds (e.g., toluene) released 295 from traffic exhausts (Suh et al., 2003). The average benzoic acid concentrations were $1496 \pm$ 511 ng m⁻³ in PKU and 1278 \pm 372 ng m⁻³ in Yufa, respectively. Although, benzoic acid is 296 297 semi-volatile organic species and is mainly found in gas phase (Fraser et al., 2003), it can be 298 formed in particulate phase via gas-to-particle partitioning. During ozone episode in August 2006, high concentration of toluene was determined in Beijing (11.4 μ g m⁻³) (Duan et al., 299 300 2008), which suggests that oxidation of toluene is one of the significant sources of benzoic 301 acid in the air.

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303 3.3 Less polluted air versus pollution events

304 Figure 1a and 1b show the temporal variation of mass concentrations of EC, OC and 305 WSOC in PKU and Yufa from 3 to 31 August 2007, respectively. Heavier air pollution events 306 were observed during 3, 5, 9, 15 and 31 August, as reflected by the elevated PM_{2.5} concentrations (i.e., range 96-191 μ g m⁻³, average 124 μ g m⁻³ in PKU and range 100-127 μ g 307 m^{-3} , average 110 µg m^{-3} in Yufa, respectively). The concentrations of OC, EC, and WSOC 308 309 significantly increased during these pollution events, but generally decreased for the less 310 polluted air mass events on 7, 13, 21 and 27 August, consistent with lower PM_{2.5} concentrations (i.e., a range of 65-77 μ g m⁻³, average 71 μ g m⁻³ in PKU and a range of 39-179 311 μ g m⁻³, average 62 μ g m⁻³ in Yufa, respectively). Similar temporal variations in total 312 313 quantified bifunctional organic compounds and fatty acids were observed in both PKU and 314 Yufa (Figure 1c and 1d). However, the temporal variation of benzoic acid is different from 315 the other compounds measured, indicating a different source or atmospheric processing for benzoic acid. 316

317 Ensemble 3 day air mass back trajectory analysis shows that the pollutants were 318 mainly from northeast, passing over southeast or south of Beijing, during heavier pollution 319 events, whereas they were mainly from north or northwest sector during less pollution events 320 (see Figure 2). South and southeast areas of Beijing are located close to the heavily 321 industrialized areas (e.g., Tianjin city, Shandong and Hebei province) whereas north and 322 northwest areas of Beijing are enclosed by the massive mountain ranges without impact of 323 anthropogenic pollution sources (Ho et al., 2010). As seen in Figures 3a and b, the 324 concentration levels of EC, OC, WSOC, diacids and ketocarboxylic acids in PKU and Yufa 325 are higher for heavier pollution episodes compared to the less polluted air events, suggesting 326 that high emission of carbonaceous aerosols and their precursor gases from neighboring 327 provinces and the subsequent transport to Beijing is one of the major sources responsible for 328 the elevated particulate pollutants in Beijing.

329 The OC to EC ratio (OC/EC) was used to estimate the transformation and emission 330 properties of carbonaceous aerosol. The average OC/EC ratios at less polluted air (PKU: 2.63; 331 Yufa: 2.19) events were slightly higher those found at the pollution episodes (PKU: 2.52; 332 Yufa: 2.05) at both sites. The slightly lower OC/EC ratio during pollution episodes is likely associated to high combustion emissions, especially from traffic exhaust. The slightly higher 333 334 OC/EC ratios observed during less polluted air events suggest that secondary formation of 335 OA was critical during less polluted air event. Bendle et al. (2007) reported that the 336 unsaturated-over-saturated C_{18} fatty acids ($C_{18:n}/C_{18:0}$) ratio could be used as a good indicator 337 to estimate the freshness of OM in marine samples. In this study, high ratios were recorded in 338 samples associated with pollution episode, whereas low ratios were observed in less polluted 339 air event with air masses originated from the north and northwest of Beijing. Low ratios 340 observed in less polluted air event represent an aged air mass, indicating longer residence 341 time for particle transformation and transportation (Alves et al., 2007).

342 Moreover, malonic acid (C_3) can be a byproduct of photochemical breakdown of 343 succinic acid (C₄) in the air. The C_3/C_4 ratio, which was used as a tracer of the enhanced 344 photochemical aging of OA (Kawamura and Ikushima, 1993), observed during less polluted 345 air event was higher than pollution episode in both sites (0.66 versus 0.58 in PKU and 0.57 346 versus 0.52 in Yufa). Higher C_3/C_4 ratios in less polluted air event suggest that secondary 347 formation of diacids are more significant in less polluted air event, which further indicates 348 secondary photochemical formation of particulate diacids is also critical during less polluted 349 air event.

350 It should be noted, however, that the concentrations of α -dicarbonyls and benzoic acid 351 in both PKU and Yufa are higher during less polluted air episodes compared to pollution 352 episodes. This indicates that local production or secondary formation could be important 353 source for these compounds. It is known that α -dicarbonyls are intermediate reaction products 354 (via photochemical oxidation) of a wide range of biogenic and anthropogenic VOCs 355 (Galloway et al., 2009). More distant sources lead to longer transport time and therefore to 356 increase chemical oxidation of glyoxal and methylglyoxal to their corresponding acids and further reaction products. This potentially reduces the local contribution of α -dicarbonyls in 357 358 Beijing. Positive correlation was observed between α -dicarbonyls and benzoic acid ($R^2 = 0.82$) in PKU and $R^2 = 0.65$ in Yufa) at both sites (Figure 4a and 4b), which further suggests that a 359 360 major fraction of α -dicarbonyls and benzoic acid are most likely produced in the local 361 atmosphere of Beijing through photochemical processing.

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3.4 Influence of local traffic on air quality between restriction and non-restriction periods

One goal of this sampling campaign is to study the traffic controls influence on the air quality in Beijing given the use of a large number of vehicles and the resulting high emission of particulate matter and precursor gases. As described above, the level of particulate pollutants in Beijing is significantly influenced by regional transport depending on the wind sector. Therefore, in the following discussion, only events with wind from the same sector (minimizing the difference from regional contribution) but with and without traffic restriction in Beijing are selected to evaluate the effectiveness of local traffic restriction measure on air pollution reduction. Measurements taken on 17 and 19 August represent the restriction events
and those taken on 3, 5, 9, 15 and 31 August represent the non-restriction events.

373 The concentration ratios of the restriction to the non-restriction periods (R/N) are 374 shown in Figure 5. A value of close to unity represents that the restriction does not have any 375 impact in the pollution controls. In PKU, the R/N ratios of EC, OC, WSOC, total diacids, total 376 ketocarboxylic acids and total α -dicarbonyls are much lower than 1, suggesting that these 377 pollutants or their precursors are closely related to the traffic emissions and that the "traffic 378 restriction" measure can reduce primary pollutants (e.g., EC) and the precursors of secondary 379 pollutants (e.g., diacids and α -dicarbonyls). A previous study (Zhang et al., 2011) also 380 indicated the reduction of anthropogenic elements in Beijing during the traffic restriction 381 period of August, 2007. The average OC/EC ratios observed at traffic restriction period (PKU: 382 2.69) was slightly higher than that found at non-restriction period (PKU: 2.52). The slightly 383 lower OC/EC ratio during non-restriction period was mainly due to the higher EC emissions 384 from traffic exhaust, while EC emissions were reduced during traffic restriction period. 385 However, the R/N ratios of benzoic acid and total fatty acids are higher than 1. A possible 386 explanation for this elevated R/N ratios is that these organics are mainly derived from 387 regional emissions. An alternative is that they are mainly produced from sources other than 388 vehicle emissions. For example, cooking emission that was not controlled under traffic 389 restriction period is a significant source of fatty acids in the air. More household cooking 390 activities can be found if the residents trended to stay home during the restriction period.

391 The profile of R/N ratio in Yufa is different from that in PKU. The concentrations of 392 OC, WSOC, total diacids and total fatty acids were lower during restriction period than those 393 during non-restriction period, suggesting that the "traffic restriction" measure indeed reduced 394 particulate pollutants. However, the decrease is generally smaller in Yufa compared to that in 395 PKU, indicating that the contribution of local traffic emission to air pollution in Yufa is 396 smaller. The R/N value >1 occurred to EC, total ketocarboxylic acids, total α -dicarbonyls and 397 benzoic acid. An enhanced EC value indicates elevated primary emissions in Yufa during 398 restriction period than non-restriction period. The potential contribution could be local rural 399 emissions (e.g., biomass burning and coal burning) and/or regional transport from polluted 400 neighboring provinces that are closer to Yufa. The average OC/EC ratios at traffic non-401 restriction period (Yufa: 2.05) events were slightly higher than those found at restriction 402 period (Yufa: 1.89). The lower OC/EC ratios during restriction period further suggest the 403 elevated emissions of EC from sources other than traffic at Yufa.

405 3.5 Ratios of selected species

406 The C_3/C_4 ratios measured in this study varied from 0.28 to 0.84 (average value: 0.59) 407 which are close to those measured in Northern China (0.61) (Ho et al., 2007), but higher than 408 that observed from traffic exhausts (0.3-0.5) (Kawamura and Kaplan, 1987). However, the 409 ratios determined in this study are much lower than the marine particles measured from 410 Pacific Ocean, where photochemical processing is commonly more intensive (Kawamura and 411 Sakaguchi, 1999). Higher C_3/C_4 ratios were observed in PKU (0.62) than in Yufa (0.56), 412 additionally, the ratios observed during traffic restriction period were higher than non-413 restriction periods in both sites (0.65 versus 0.58 in PKU and 0.61 versus 0.52 in Yufa). This 414 result suggests that C₃ is vigorously produced in traffic restriction period by photochemical 415 reaction of C₄ (Kawamura and Ikushima, 1993). Even though variations of the ratio were 416 small, these are sufficiently representatives to any minor rotations and vibrations of emission 417 sources. The results also suggested that secondary formation of diacids by photochemical 418 oxidation was critical during traffic restriction period despite primary exhaust was controlled.

419 Adipic acid (C_6) is considered as a reaction product of the photochemical oxidation of 420 cyclohexene, whereas C₉ is mainly emitted from unsaturated fatty acids (Hatakeyama et al., 421 1987; Kawamura and Gagosian, 1987). Therefore C_6/C_9 ratio has been applied to evaluate the 422 abundances of biogenic and anthropogenic sources to OA (Kawamura and Yasui, 2005). 423 C_6/C_9 ratios show higher values in non-restriction period (PKU: 0.40; Yufa: 0.61) than in 424 restriction period (PKU: 0.36; Yufa: 0.38) in this study. Higher C₆/C₉ ratios observed in non-425 restriction period support that anthropogenic organic compounds, especially from vehicles, 426 are the major source of OA during that period of time.

427 EC is a major component of vehicle exhaust, whereas C_2 is a major secondary organic 428 species in the air. Therefore, C_2/EC ratio can be used to assess the aging of the air mass. The 429 average C_2/EC ratios were 0.075 and 0.078 at PKU and Yufa, respectively, (which has a 430 range of 0.044 to 0.113), which are much higher than previously reported traffic exhaust ratio 431 (0.0022), but similar to those measured in the air over Shenzhen (0.063 in summer) (Huang 432 and Yu, 2007). The C_2/EC ratios generally showed higher values in restriction period (PKU: 433 0.081; Yufa: 0.077) than in non-restriction period (PKU: 0.067; Yufa: 0.074). The results are 434 consistent with the notion that the "traffic restriction" measure can reduce primary pollutants 435 (e.g., EC).

436 Moreover, C_2 /total diacids ratio can be applied as an indicator to assess the aging of 437 OA (Kawamura and Sakaguchi, 1999). In this study, the abundances of C_2 in total diacids 438 varied from less than 30% to 54%. Interestingly, the ratios of C_2 /total diacids generally 439 showed higher values in restriction period than in non-restriction period. The result indicates 440 that oxalic acid is preferentially formed in restriction period by the oxidation of its precursors 441 (other than anthropogenic VOCs, biogenic VOCs and their oxidation products may serve as 442 important precursors in restriction period) in the atmosphere. Further, ωC_9 is generated by 443 biogenic unsaturated fatty acids oxidation, revealing higher concentrations in restriction period (PKU: 3.47 ng m⁻³; Yufa: 2.49 ng m⁻³) than in non-restriction period (PKU: 1.82 ng m⁻³) 444 ³; Yufa: 2.12 ng m⁻³) (Yokouchi and Ambe, 1986). This result indicates that biogenic 445 emissions are important source for the formation of ωC_9 in restriction period, which can 446 447 further breakdown to produce lower molecular weight diacids including C_4 , C_3 , and C_2 . The 448 results further indicate that secondary formation of diacids by atmospheric oxidation was also 449 critical during traffic restriction period despite primary exhaust was controlled.

450

451 **4.** Summary and Conclusions

452 During the CAREBeijing-2007 in summer, molecular compositions of bifunctional 453 organic compounds, fatty acids and benzoic acid were studied in Beijing. Oxalic acid (C_2) 454 was detected as the most abundant diacid followed by phthalic (Ph) acid. Low MW 455 bifunctional organic compounds were found as the major water-soluble organic fraction, 456 accounting for more than 8.9% and 10.3% of WSOC in PKU and Yufa, respectively. 457 Additionally, total fatty acids and benzoic acid contributed 3.1% and 7.2% of OC in PKU and 458 3.2% and 9.3% of OC in Yufa, respectively. Bifunctional organic compounds can be released 459 from primary emissions (e.g., traffic exhaust and biomass burning) or formed by atmospheric 460 oxidation of VOCs in the Beijing atmosphere. Both natural biogenic (e.g., microbial) and 461 anthropogenic (e.g., traffic exhaust, cooking) sources provide the major inputs of fatty acids, 462 whereas benzoic acid was mainly formed by the photodegradation of aromatic compounds 463 such as toluene from traffic emission.

The concentrations of OC, EC and WSOC significantly increased during the heavy pollution events, but generally decreased during the less pollution events. Results of back trajectory analyses indicated that the air masses were originated mainly from northeast, passing over heavily populated, urbanized and industrialized areas during the heavy pollution events, whereas they were mainly from mountain clean areas during less pollution events.

In PKU, the restriction to non-restriction period (R/N) ratios of OC, EC, WSOC, total diacids, total ketocarboxylic acids and total α -dicarbonyls were much lower than 1, suggesting that the "traffic restriction" measure can reduce primary pollutants (e.g., EC) and the precursors of secondary pollutants (e.g., diacids and α -dicarbonyls). The R/N ratios of OC,

473 WSOC, total diacids and total fatty acids in Yufa were lower than 1, however, the values are 474 generally larger than those in PKU. Moreover, the R/N value >1 occurred to EC, total 475 ketocarboxylic acids, total α -dicarbonyls and benzoic acid, indicating that there are higher 476 contribution of local emissions (e.g., coal and biomass burning) and/or regional transport 477 from polluted neighboring provinces than local traffic emission in Yufa.

478 The C₃/C₄, C₂/EC and C₂/total diacids ratios observed during traffic restriction period 479 were higher than those of non-restriction periods at both sites. This result suggests that C₂ and 480 C₃ are secondarily more produced in traffic restriction period by the photochemical oxidation 481 of their precursors, indicating that even when primary exhaust was controlled, secondary 482 photochemical formation of particulate diacids was not controlled during traffic restriction 483 period. This study demonstrates that atmospheric oxidizing capability (photochemical aging) 484 is enhanced by the reduction of atmospheric loading of aerosol particles during the traffic 485 restriction period possibly due to the increased solar radiation reaching to the ground surface.

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493 **Reference**

- Alves, C., Oliveira, T., Pio, C., Silvestre, A. J. D., Fialho, P., Barata, F., and Legrand, M.:
 Characterisation of carbonaceous aerosols from the Azorean Island of Terceira, Atmos.
 Environ., 41, 1359-1373, doi:<u>http://dx.doi.org/10.1016/j.atmosenv.2006.10.022</u>, 2007.
- An, X., Zhu, T., Wang, Z., Li, C., and Wang, Y.: A modeling analysis of a heavy air pollution
 episode occurred in Beijing, Atmos. Chem. Phys., 7, 3103-3114, 2007.
- Bendle, J., Kawamura, K., Yamazaki, K., and Niwai, T.: Latitudinal distribution of terrestrial
 lipid biomarkers and n-alkane compound-specific stable carbon isotope ratios in the
 atmosphere over the western Pacific and Southern Ocean, Geochim. Cosmochim. Acta,
 71, 5934-5955, doi:10.1016/j.gca.2007.09.029, 2007.
- 503 Beijing statistical yearbook: <u>http://www.bjstats.gov.cn/</u>, access: Feb 2014, 2010.
- Cao, J. J., Lee, S. C., Ho, K. F., Zhang, X. Y., Zou, S. C., Fung, K., Chow, J. C., and Watson,
 J. G.: Characteristics of carbonaceous aerosol in Pearl River Delta Region, China
 during 2001 winter period, Atmos. Environ., 37, 1451-1460, doi:10.1016/s13522310(02)01002-6, 2003.
- Carlton, A. G., Turpin, B. J., Lim, H. J., Altieri, K. E., and Seitzinger, S.: Link between isoprene and secondary organic aerosol (SOA): Pyruvic acid oxidation yields low volatility organic acids in clouds, Geophys. Res. Lett., 33, L06822, doi:10.1029/2005gl025374, 2006.
- 512 Chow, J. C., Watson, J. G., Chen, L. W. A., Paredes-Miranda, G., Chang, M. C. O., Trimble,
 513 D., Fung, K. K., Zhang, H., and Yu, J. Z.: Refining temperature measures in
 514 thermal/optical carbon analysis, Atmos. Chem. Phys., 5, 2961-2972, 2005.
- 515 Ding, A. J., Wang, T., Thouret, V., Cammas, J. P., and Nedelec, P.: Tropospheric ozone
 516 climatology over Beijing: analysis of aircraft data from the MOZAIC program, Atmos.
 517 Chem. Phys., 8, 1-13, 2008.
- 518 Duan, J. C., Tan, J. H., Yang, L., Wu, S., and Hao, J. M.: Concentration, sources and ozone
 519 formation potential of volatile organic compounds (VOCs) during ozone episode in
 520 Beijing, Atmospheric Research, 88, 25-35, doi:10.1016/j.atmosres.2007.09.004, 2008.
- Falkovich, A. H., Graber, E. R., Schkolnik, G., Rudich, Y., Maenhaut, W., and Artaxo, P.:
 Low molecular weight organic acids in aerosol particles from Rondonia, Brazil,
 during the biomass-burning, transition and wet periods, Atmos. Chem. Phys., 5, 781797, 2005.
- Fick, J., Nilsson, C., and Andersson, B.: Formation of oxidation products in a ventilation
 system, Atmos. Environ., 38, 5895-5899, doi:10.1016/j.atmosenv.2004.08.020, 2004.
- Fraser, M. P., Cass, G. R., and Simoneit, B. R. T.: Air quality model evaluation data for
 organics. 6. C-3-C-24 organic acids, Environ. Sci. Technol., 37, 446-453,
 doi:10.1021/Es0209262, 2003.
- Fu, P. Q., Kawamura, K., Okuzawa, K., Aggarwal, S. G., Wang, G. H., Kanaya, Y., and
 Wang, Z. F.: Organic molecular compositions and temporal variations of summertime
 mountain aerosols over Mt. Tai, North China Plain, J. Geophys. Res-Atmos., 113,
 D19107, doi:10.1029/2008jd009900, 2008.
- Galloway, M. M., Chhabra, P. S., Chan, A. W. H., Surratt, J. D., Flagan, R. C., Seinfeld, J. H.,
 and Keutsch, F. N.: Glyoxal uptake on ammonium sulphate seed aerosol: reaction
 products and reversibility of uptake under dark and irradiated conditions, Atmos.
 Chem. Phys., 9, 3331-3345, 2009.
- Glasius, M., Lahaniati, M., Calogirou, A., Di Bella, D., Jensen, N. R., Hjorth, J., Kotzias, D.,
 and Larsen, B. R.: Carboxylic acids in secondary aerosols from oxidation of cyclic
 monoterpenes by ozone, Environ. Sci. Technol., 34, 1001-1010,
 doi:10.1021/Es990445r, 2000.

- Guttikunda, S. K., Tang, Y. H., Carmichael, G. R., Kurata, G., Pan, L., Streets, D. G., Woo, J.
 H., Thongboonchoo, N., and Fried, A.: Impacts of Asian megacity emissions on regional air quality during spring 2001, J. Geophys. Res-Atmos., 110, D20301, doi:10.1029/2004jd004921, 2005.
- Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M.,
 Dommen, J., Donahue, N. M., George, C., Goldstein, A. H., Hamilton, J. F.,
 Herrmann, H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L.,
 Kiendler-Scharr, A., Maenhaut, W., McFiggans, G., Mentel, T. F., Monod, A., Prevot,
 A. S. H., Seinfeld, J. H., Surratt, J. D., Szmigielski, R., and Wildt, J.: The formation,
 properties and impact of secondary organic aerosol: current and emerging issues,
 Atmos. Chem. Phys., 9, 5155-5236, 2009.
- Hatakeyama, S., Takami, A., Wang, W., and Tang, D. G.: Aerial observation of air pollutants
 and aerosols over Bo Hai, China, Atmos. Environ., 39, 5893-5898,
 doi:http://dx.doi.org/10.1016/j.atmosenv.2005.06.025, 2005.
- Hatakeyama, S., Ohno, M., Weng, J., Takagi, H., and Akimoto, H.: Mechanism for the
 formation of gaseous and particulate products from ozone-cycloalkene reactions in air,
 Environ. Sci. Technol., 21, 52-57, doi:10.1021/es00155a005, 1987.
- 559 He, N., Kawamura, K., Okuzawa, K., Kanaya, Y., and Wang, Z. F.: Diurnal variations of total carbon, dicarboxylic acids, ketoacids and α -dicarbonyls in aerosols in the northern 560 561 vicinity of Beijing, Atmos. Chem. Phys. Discuss., 13, 16699-16731, 562 doi:10.5194/acpd-13-16699-2013, 2013.
- Ho, K. F., Lee, S. C., Cao, J. J., Kawamura, K., Watanabe, T., Cheng, Y., and Chow, J. C.:
 Dicarboxylic acids, ketocarboxylic acids and dicarbonyls in the urban roadside area of
 Hong Kong, Atmos. Environ., 40, 3030-3040, 2006.
- Ho, K. F., Cao, J. J., Lee, S. C., Kawamura, K., Zhang, R. J., Chow, J. C., and Watson, J. G.:
 Dicarboxylic acids, ketocarboxylic acids, and dicarbonyls in the urban atmosphere of China, J. Geophys. Res-Atmos., 112, D22S27, doi:10.1029/2006jd008011, 2007.
- 569 Ho, K. F., Lee, S. C., Ho, S. S. H., Kawamura, K., Tachibana, E., Cheng, Y., and Zhu, T.: 570 Dicarboxylic acids, ketocarboxylic acids, α -dicarbonyls, fatty acids, and benzoic acid 571 in urban aerosols collected during the 2006 Campaign of Air Quality Research in 572 (CAREBeijing-2006), J. Geophys. Beijing Res., 115. D19312, doi:10.1029/2009jd013304, 2010. 573
- Ho, K. F., Ho, S. S. H., Lee, S. C., Kawamura, K., Zou, S. C., Cao, J. J., and Xu, H. M.:
 Summer and winter variations of dicarboxylic acids, fatty acids and benzoic acid in
 PM_{2.5} in Pearl Delta River Region, China, Atmos. Chem. Phys., 11, 2197-2208,
 doi:10.5194/acp-11-2197-2011, 2011.
- Hoffmann, T., Huang, R.-J., and Kalberer, M.: Atmospheric analytical chemistry, Anal.
 Chem., 83, 4649-4664, 2011.
- Huang, R.-J., Zhang, Y. L., Bozzetti, C., Ho, K. F., Cao, J. J., Han, Y. M., Dällenbach, K. R.,
 Slowik, J. G., Platt, S. M., Canonaco, F., Zotter, P., Wolf, R., Pieber, S. M., Bruns, E.
 A., Crippa, M., Ciarelli, G., Piazzalunga, A., Schwikowski, M., Abbaszade, G.,
 Schnelle-Kreis, J., Zimmermann, R., An, Z. S., Szidat, S., Baltensperger, U., El
 Haddad, I., Prévôt, A. S. H.: High secondary aerosol contribution to particulate
 pollution during haze events in China, Nature, 514, 218-222, 2014.
- Huang, X. F., and Yu, J. Z.: Is vehicle exhaust a significant primary source of oxalic acid in
 ambient aerosols?, Geophys. Res. Lett., 34, L02808, doi:10.1029/2006gl028457, 2007.
- Jaffrezo, J.-L., Aymoz, G., Delaval, C., and Cozic, J.: Seasonal variations of the water soluble
 organic carbon mass fraction of aerosol in two valleys of the French Alps, Atmos.
 Chem. Phys., 5, 2809-2821, 2005.

- 591 Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prevot, A. S. H., Zhang, Q., Kroll, J. H., 592 DeCarlo, P. F., Allan, J. D., Coe, H., Ng, N. L., Aiken, A. C., Docherty, K. S., Ulbrich, 593 I. M., Grieshop, A. P., Robinson, A. L., Duplissy, J., Smith, J. D., Wilson, K. R., Lanz, 594 V. A., Hueglin, C., Sun, Y. L., Tian, J., Laaksonen, A., Raatikainen, T., Rautiainen, J., 595 Vaattovaara, P., Ehn, M., Kulmala, M., Tomlinson, J. M., Collins, D. R., Cubison, M. 596 J., Dunlea, E. J., Huffman, J. A., Onasch, T. B., Alfarra, M. R., Williams, P. I., Bower, 597 K., Kondo, Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., 598 Salcedo, D., Cottrell, L., Griffin, R., Takami, A., Miyoshi, T., Hatakeyama, S., 599 Shimono, A., Sun, J. Y., Zhang, Y. M., Dzepina, K., Kimmel, J. R., Sueper, D., Jayne, 600 J. T., Herndon, S. C., Trimborn, A. M., Williams, L. R., Wood, E. C., Middlebrook, A. 601 M., Kolb, C. E., Baltensperger, U., and Worsnop, D. R.: Evolution of Organic 602 Aerosols in the Atmosphere, Science, 326, 1525-1529, doi:10.1126/science.1180353, 603 2009.
- Kawamura, K., and Kaplan, I. R.: Motor exhaust emissions as a primary source for
 dicarboxylic-acids in Los-Angeles ambient air, Environ. Sci. Technol., 21, 105-110,
 doi:10.1021/Es00155a014, 1987.
- Kawamura, K., and Gagosian, R. B.: Implications of [omega]-oxocarboxylic acids in the
 remote marine atmosphere for photo-oxidation of unsaturated fatty acids, Nature, 325,
 330-332, 1987.
- Kawamura, K., and Ikushima, K.: Seasonal-changes in the distribution of dicarboxylic-acids
 in the urban atmosphere, Environ. Sci. Technol., 27, 2227-2235,
 doi:10.1021/Es00047a033, 1993.
- Kawamura, K., and Sakaguchi, F.: Molecular distributions of water soluble dicarboxylic acids
 in marine aerosols over the Pacific Ocean including tropics, J. Geophys. Res-Atmos.,
 104, 3501-3509, doi:10.1029/1998jd100041, 1999.
- Kawamura, K., and Yasui, O.: Diurnal changes in the distribution of dicarboxylic acids,
 ketocarboxylic acids and dicarbonyls in the urban Tokyo atmosphere, Atmos. Environ.,
 39, 1945-1960, doi:10.1016/j.atmosenv.2004.12.014, 2005.
- Kawamura, K., Ng, L. L., and Kaplan, I. R.: Determination of organic-acids (C1-C10) in the
 atmosphere, motor exhausts, and engine oils, Environ. Sci. Technol., 19, 1082-1086,
 doi:10.1021/Es00141a010, 1985.
- Kawamura, K., Kasukabe, H., and Barrie, L. A.: Source and reaction pathways of
 dicarboxylic acids, ketoacids and dicarbonyls in arctic aerosols: One year of
 observations, Atmos. Environ., 30, 1709-1722, doi:10.1016/1352-2310(95)00395-9,
 1996.
- Kawamura, K., Kobayashi, M., Tsubonuma, N., Mochida, M., Watanabe, T., and Lee, M.:
 Organic and inorganic compositions of marine aerosols from East Asia: Seasonal
 variations of water-soluble dicarboxylic acids, major ions, total carbon and nitrogen,
 and stable C and N isotopic composition, Geo Soc S P, 9, 243-265,
 doi:http://dx.doi.org/10.1016/S1873-9881(04)80019-1, 2004.
- Kroll, J. H., Ng, N. L., Murphy, S. M., Varutbangkul, V., Flagan, R. C., and Seinfeld, J. H.:
 Chamber studies of secondary organic aerosol growth by reactive uptake of simple
 carbonyl compounds, J. Geophys. Res-Atmos., 110, D23207,
 doi:10.1029/2005jd006004, 2005.
- Kundu, S., Kawamura, K., Andreae, T. W., Hoffer, A., and Andreae, M. O.: Molecular
 distributions of dicarboxylic acids, ketocarboxylic acids and alpha-dicarbonyls in
 biomass burning aerosols: implications for photochemical production and degradation
 in smoke layers, Atmos. Chem. Phys., 10, 2209-2225, 2010.
- Legrand, M., Preunkert, S., Oliveira, T., Pio, C. A., Hammer, S., Gelencser, A., Kasper-Giebl,
 A., and Laj, P.: Origin of C-2-C-5 dicarboxylic acids in the European atmosphere

- inferred from year-round aerosol study conducted at a west10.1029/2006jd008019east transect, J. Geophys. Res-Atmos., 112, D23S07,
 doi:10.1029/2006jd008019, 2007.
- 644 Liu, Y. N., Tao, S., Yang, Y. F., Dou, H., Yang, Y., and Coveney, R. M.: Inhalation exposure 645 of traffic police officers to polycyclic aromatic hydrocarbons (PAHs) during the 646 winter in Beijing, China, Sci. Total Environ., 383, 98-105. 647 doi:10.1016/j.scitotenv.2007.05.008, 2007.
- Luo, C., John, J. C. S., Zhou, X. J., Lam, K. S., Wang, T., and Chameides, W. L.: A nonurban
 ozone air pollution episode over eastern China: Observations and model simulations, J.
 Geophys. Res-Atmos., 105, 1889-1908, doi:10.1029/1999jd900970, 2000.
- Mauzerall, D. L., Narita, D., Akimoto, H., Horowitz, L., Walters, S., Hauglustaine, D. A., and
 Brasseur, G.: Seasonal characteristics of tropospheric ozone production and mixing
 ratios over East Asia: A global three-dimensional chemical transport model analysis, J.
 Geophys. Res-Atmos., 105, 17895-17910, doi:10.1029/2000jd900087, 2000.
- Mkoma, S. L., and Kawamura, K.: Molecular composition of dicarboxylic acids, ketocarboxylic acids, alpha-dicarbonyls and fatty acids in atmospheric aerosols from Tanzania, East Africa during wet and dry seasons, Atmos. Chem. Phys., 13, 2235-2251, doi:10.5194/acp-13-2235-2013, 2013.
- Mochida, M., Umemoto, N., Kawamura, K., Lim, H. J., and Turpin, B. J.: Bimodal size distributions of various organic acids and fatty acids in the marine atmosphere:
 Influence of anthropogenic aerosols, Asian dusts, and sea spray off the coast of East Asia, J. Geophys. Res-Atmos., 112, D15209, doi:10.1029/2006jd007773, 2007.
- Narukawa, M., Kawamura, K., Takeuchi, N., and Nakajima, T.: Distribution of dicarboxylic
 acids and carbon isotopic compositions in aerosols from 1997 Indonesian forest fires,
 Geophys. Res. Lett., 26, 3101-3104, doi:10.1029/1999gl010810, 1999.
- Oliveira, C., Pio, C., Alves, C., Evtyugina, M., Santos, P., Goncalves, V., Nunes, T., Silvestre, 666 667 A. J. D., Palmgren, F., Wahlin, P., and Harrad, S.: Seasonal distribution of polar 668 organic compounds in the urban atmosphere of two large cities from the North and 669 South of Europe, Atmos. Environ., 41, 5555-5570, 670 doi:10.1016/j.atmosenv.2007.03.001, 2007.
- Poschl, U.: Atmospheric aerosols: Composition, transformation, climate and health effects,
 Angew Chem Int Edit, 44, 7520-7540, doi:10.1002/anie.200501122, 2005.
- Rogge, W. F., Medeiros, P. M., and Simoneit, B. R. T.: Organic marker compounds for
 surface soil and fugitive dust from open lot dairies and cattle feedlots, Atmos.
 Environ., 40, 27-49, doi:10.1016/j.atmosenv.2005.07.076, 2006.
- Rogge, W. F., Hildemann, L. M., Mazurek, M. A., Cass, G. R., and Simonelt, B. R. T.:
 Sources of fine organic aerosol .1. Charbroilers and meat cooking operations, Environ.
 Sci. Technol., 25, 1112-1125, doi:10.1021/Es00018a015, 1991.
- Rogge, W. F., Hildemann, L. M., Mazurek, M. A., Cass, G. R., and Simoneit, B. R. T.:
 Sources of Fine Organic Aerosol .2. Noncatalyst and Catalyst-Equipped Automobiles
 and Heavy-Duty Diesel Trucks, Environ. Sci. Technol., 27, 636-651, doi:Doi
 10.1021/Es00041a007, 1993.
- Rudich, Y., Donahue, N. M., and Mentel, T. F.: Aging of organic aerosol: Bridging the gap
 between laboratory and field studies, Annu. Rev. Phys. Chem., 58, 321-352,
 doi:10.1146/annurev.physchem.58.032806.104432, 2007.
- Saxena, P., and Hildemann, L. M.: Water-soluble organics in atmospheric particles: A critical
 review of the literature and application of thermodynamics to identify candidate
 compounds, J. Atmos. Chem., 24, 57-109, doi:10.1007/Bf00053823, 1996.

- Schauer, J. J., Kleeman, M. J., Cass, G. R., and Simoneit, B. R. T.: Measurement of emissions
 from air pollution sources. 1. C1 through C29 organic compounds from meat
 charbroiling, Environ. Sci. Technol., 33, 1566-1577, doi:10.1021/Es980076j, 1999.
- Schauer, J. J., Kleeman, M. J., Cass, G. R., and Simoneit, B. R. T.: Measurement of emissions
 from air pollution sources. 4. C1-C27 organic compounds from cooking with seed oils,
 Environ. Sci. Technol., 36, 567-575, doi:10.1021/Es002053m, 2002.
- 695 Simoneit, B. R. T.: Organic-matter of the troposphere .3. Characterization and sources of
 696 petroleum and pyrogenic residues in aerosols over the western United-States, Atmos.
 697 Environ., 18, 51-67, doi:10.1016/0004-6981(84)90228-2, 1984.
- Simoneit, B. R. T., and Mazurek, M. A.: Organic-matter of the troposphere .2. Natural
 background of biogenic lipid matter in aerosols over the rural western United-States,
 Atmos. Environ., 16, 2139-2159, doi:10.1016/0004-6981(82)90284-0, 1982.
- Stephanou, E. G., and Stratigakis, N.: Oxocarboxylic and alpha,omega-dicarboxylic acids Photooxidation products of biogenic unsaturated fatty-acids present in urban aerosols,
 Environ. Sci. Technol., 27, 1403-1407, doi:10.1021/Es00044a016, 1993.
- Suh, I., Zhang, R., Molina, L. T., and Molina, M. J.: Oxidation mechanism of aromatic peroxy
 and bicyclic radicals from OH-toluene reactions, J. Am. Chem. Soc., 125, 1265512665, doi:10.1021/ja0350280, 2003.
- Tao, S., Wang, Y., Wu, S. M., Liu, S. Z., Dou, H., Liu, Y. N., Lang, C., Hu, F., and Xing, B. 707 708 S.: Vertical distribution of polycyclic aromatic hydrocarbons in atmospheric boundary 709 winter, Atmos. Environ., 9594-9602, layer of Beijing in 41, 710 doi:10.1016/j.atmosenv.2007.08.026, 2007.
- Volkamer, R., Platt, U., and Wirtz, K.: Primary and secondary glyoxal formation from aromatics: Experimental evidence for the bicycloalkyl-radical pathway from benzene, toluene, and p-xylene, J. Phys. Chem. A, 105, 7865-7874, doi:10.1021/Jp010152w, 2001.
- Wang, G. H., and Kawamura, K.: Molecular characteristics of urban organic aerosols from Nanjing: A case study of a mega-city in China, Environ. Sci. Technol., 39, 7430-7438, doi:10.1021/Es051055+, 2005.
- Wang, G. H., Kawamura, K., Lee, S., Ho, K., and Cao, J.: Molecular, seasonal, and spatial distributions of organic aerosols from fourteen Chinese cities, Environ. Sci. Technol., 40, 4619-4625, 2006.
- Warneck, P.: Multi-phase chemistry of C2 and C3 organic compounds in the marine atmosphere, J. Atmos. Chem., 51, 119-159, doi:10.1007/s10874-005-5984-7, 2005.
- Xu, J., Zhang, Y. H., Fu, J. S., Zheng, S. Q., and Wang, W.: Process analysis of typical
 summertime ozone episodes over the Beijing area, Sci. Total Environ., 399, 147-157,
 doi:10.1016/j.scitotenv.2008.02.013, 2008.
- Yang, H., Yu, J. Z., Ho, S. S. H., Xu, J. H., Wu, W. S., Wan, C. H., Wang, X. D., Wang, X.
 R., and Wang, L. S.: The chemical composition of inorganic and carbonaceous materials in PM_{2.5} in Nanjing, China, Atmos. Environ., 39, 3735-3749, doi:<u>http://dx.doi.org/10.1016/j.atmosenv.2005.03.010</u>, 2005.
- Yokouchi, Y., and Ambe, Y.: Characterization of polar organics in airborne particulate matter,
 Atmospheric Environment (1967), 20, 1727-1734, doi:<u>http://dx.doi.org/10.1016/0004-6981(86)90121-6</u>, 1986.
- Zhang, R. J., Shen, Z. X., Zhang, L. M., Zhang, M. G., Wang, X., and Zhang, K.: Elemental
 Composition of Atmospheric Particles during Periods with and without Traffic
 Restriction in Beijing: The Effectiveness of Traffic Restriction Measure, Sola, 7, 061064, doi:10.2151/sola.2011-016, 2011.

- Zhao, Y. L., Hu, M., Slanina, S., and Zhang, Y. H.: Chemical compositions of fine particulate
 organic matter emitted from Chinese cooking, Environ. Sci. Technol., 41, 99-105,
 doi:10.1021/Es0614518, 2007a.
- Zhao, Y. L., Hu, M., Slanina, S., and Zhang, Y. H.: The molecular distribution of fine particulate organic matter emitted from Western-style fast food cooking, Atmos.
 Environ., 41, 8163-8171, doi:10.1016/j.atmosenv.2007.06.029, 2007b.
- 743

Compounds	PKU (n=10)	j8 _ ·		Yufa (n=10))	
$(ng m^{-3})$	Range	Average	S.D.	Range	Average	S.D.
	Dic	arboxylic a	acids			
Oxalic, C2	212-586	435	124	226-632	418	130
Malonic, C3	30.0-73.5	54.9	14.0	17.1-68.6	43.5	15.0
Succinic, C4	52.8-147	89.9	27.7	44.8-129	80.9	26.9
Glutaric, C5	13.7-59.2	36.0	14.4	16.0-168	41.3	46.2
Adipic, C6	15.1-35.1	26.7	6.03	10.8-73.1	24.5	18.1
Pimeric, C7	MDLs-6.44	2.79	2.56	MDLs-3.38	0.34	1.07
Suberic, C8	MDLs	MDLs		MDLs	MDLs	
Azelaic, C9	58.8-85.8	71.4	8.91	37.5-64.6	49.2	8.99
Sebacic, C10	MDLs-3.91	0.69	1.47	MDLs-2.78	0.28	0.88
Undecanedioic, C11	MDLs	MDLs		MDLs	MDLs	
Dodecanedioic, C12	MDLs	MDLs		MDLs	MDLs	
Methylmalonic, iC4	MDLs	MDLs		MDLs	MDLs	
Methylsuccinic, iC5	MDLs-10.6	3.82	2.75	MDLs-5.21	3.05	2.22
2-Methylglutaric, iC6	MDLs-10.5	6.09	2.79	MDLs-8.92	4.80	2.40
Maleic, M	9.18-20.5	15.6	3.39	9.69-17.2	13.5	2.03
Fumaric, F	MDLs	MDLs		MDLs	MDLs	
Methylmaleic, mM	7.52-12.2	9.60	1.40	5.75-8.83	7.21	0.92
Phthalic. Ph	171-250	209	28.9	80.6-415	176	91.5
Isophthalic, iPh	MDLs-15.1	7.27	5.76	MDLs-12.8	4.65	5.22
Terephthalic, tPh	12.2-51.8	30.4	10.1	6.83-53.2	30.9	13.8
Malic. hC4	MDLs-3.21	0.32	1.02	MDLs	MDLs	10.0
Ketomalonic, kC3	2.79-11.8	6.56	2.79	MDLs-10.4	6.67	3.07
4-Ketopimelic, kC7	MDLs-13.0	3.71	5.06	MDLs-11.0	3.21	4.33
Total diacids	599-1287	1010	220	473-1429	909	278
	Keto	carboxvlic	acids		,	270
Pvruvic	17.9-70.2	30.3	15.9	12.7-27.3	20.1	4.18
Glyoxylic. ω C2	49.2-105	72.9	16.1	32.8-81.7	59.7	14.5
3 -Oxopropanoic. ω C3	MDLs-3.29	1.53	1.23	MDLs-1.31	0.59	0.62
4-Oxobutanoic. ω C4	5.91-25.0	15.8	5.55	6.59-23.3	15.0	4.83
9-Oxononaoic. ω C9	MDLs-7.28	1.70	2.50	MDLs-5.03	1.61	2.19
Total ketoacids	87.4-169	122	28.8	52 0-131	97.0	22.9
Total Retouclus	07.1102	-Dicarbony	vls	52.0 151	71.0	22.9
Glyoxal, Gly	1.40-21.4	13.1	5.72	2.93-17.3	11.5	4.04
Methylglyoxal MeGly	23 3-81 3	38.7	16.2	21 3-51 3	32.7	8 76
Total dicarbonyls	35 5-99 5	51.8	17.9	29.0-61.4	44.2	10.3
Sum of bifunctional species	730-1455	1184	241	554-1621	1050	303
Sum of offunctional species	750 1155	Fatty acid	<u> </u>	551 1021	1050	505
Tridecanoic acid C120	5 08-16 4	10.8	3.92	MDLs-118	7 56	3 31
Tetradecanoic acid C _{14.0}	54 5-97 9	68.7	13.6	36 5-74 0	50.0	13.6
Hexadecanoic acid Cico	199-393	249	57.0	149-310	201	48.9
Heptadecanoic acid C _{16:0}	MDI s-13 3	4 32	4 53	MDI s-12.8	3 90	5 06
Octadecanoic acid C _{10.0}	134-462	219	94.0	122-237	170	36.6
Octadecenoic acid $C_{18:0}$	2.91-33.0	24.3	8 93	13 0-47 9	24.6	9.23
Eicosanoic acid C_{200}	MDL s-7 84	4 01	3 51	MDL s-7 71	3 28	3 50
Docosanoic acid Cara	5 69-13 6	9.24	2.43	$MDL_{s-15.4}$	7.01	4 95
Tetracosanoic acid $C_{22:0}$	MDI s-10.5	6.51	3 59	MDI s-12.4	7.11	3.09
	1010-10.0	507	150	375_68/	475	11/
Sum of fatty acids	459_1003		1.57	575-004	-TIJ	114
Sum of fatty acids	459-1003 933-2754	1/06	511	724,1030	1278	372
Sum of fatty acids Benzoic acid PM_{rec} (ug m ⁻³)	459-1003 933-2754 64 9 191	<u>1496</u> 103	511	724-1930	1278	372
Sum of fatty acids Benzoic acid $PM_{2.5} (\mu g m^{-3})$	459-1003 933-2754 64.9-191 12 1 10 3	1496 103	511 35.1 2.47	724-1930 38.6-160 4.63.17.0	1278 98.3	372 34.1 3.68
Sum of fatty acids Benzoic acid $PM_{2.5} (\mu g m^{-3})$ OC ($\mu g m^{-3}$) FC ($\mu g m^{-3}$)	459-1003 933-2754 64.9-191 12.1-19.3 4.48-11.0	1496 103 14.9	511 35.1 2.47	724-1930 38.6-160 4.63-17.0 2.96.8.62	1278 98.3 11.1 5.57	372 34.1 3.68

Table 1. Concentrations of dicarboxylic acids, ketocarboxylic acids and α-dicarbonyls, fatty acids and745benzoic acid in $PM_{2.5}$ samples during CAREBeijing 2007.

746 MDL: Method Detection Limit

- 750 Figure Captions
- 751

Figure 1. Temporal variations of EC, OC, WSOC and various organic compound classes at
PKU and Yufa during CAREBeijing-2007 [Note the difference in start time at PKU and Yufa
(i.e., the first three days)].

- Figure 2. 3-day air mass back trajectories on a) 05 August b) 13 August and c) 17 and 19
 August.
- 758
- Figure 3. Pollution events versus less polluted air at PKU (a and b) and Yufa (c and d),showing the variation of particulate pollutants.
- 761
- Figure 4. Positive correlation between α-dicarbonyls and benzoic acid observed at PKU (a) and Yufa (b).
- 764
- Figure 5. The R/N ratio of particulate compounds observed at PKU (a) and Yufa (b)
- 766



Figure 1









Figure 4





