

In reply to referee comments on acp-2014-274 Wyche *et al.*, 2014

The authors are grateful to both referees for their excellent, thoughtful and insightful reviews, their comments were most welcome and indeed were very useful for manuscript improvements to be implemented. The following document addresses all of the points raised by both reviewers and explicitly details all of the correction made to the manuscript (revised manuscript attached separately with “track changes” documenting alterations) in accordance with both of the referee’s suggestions.

Author response key:

Indented blue, italicised text denotes direct author response to preceding reviewer comment.

Indented red text denotes author changes to the manuscript in response to requests or recommendations from the reviewer.

Indented plain, black text denotes unchanged areas of the manuscript.

Anonymous Referee #1

Emissions of biogenic volatile organic compounds and subsequent photochemical production of secondary organic aerosol in mesocosm studies of temperate and tropical plant species

The authors describe experiments with emissions from real trees in a photochemical reaction chamber. Focus was put on SOA formation and characterization of the gaseous precursors and oxidation products by GC-MS, PTR-MS and a CIR-TOF-MS. Experiments were conducted with and without ammonium sulfate seeds. The authors contrast a monoterpene (& other BVOC) emitter, birch, with tropic trees fig and palm, which are isoprene emitter. The birch serves also to test the current results against previous experiments. As expected SOA mass from isoprene emitters is small and was only achieved in seeded experiments. The isoprene SOA yields are a high while the MT yields are in the ballpark of previous experiments. The authors show indirectly that the relative large yields for the isoprene emitter likely arise from other BVOC, partly below the detection limits. The study was well conceptual well designed and the manuscript presents new material, especially about emissions from tropical tree species and their oxidation products as well as their SOA formation potential. The manuscript is well written.

It should be published in ACP after the authors addressed the following major and minor points. Major remarks:

Contaminations?

Section 2.2, p. 14296f

The description of the plant chamber unit cast some doubt about contaminations: “PVC foam stripes ensured an airtight seal between chamber”; “Heavy-duty double-sided tape (RS components, UK) was used to secure the Teflon bags to the frame”

Was the plant chamber tested for contaminations? Could intrusions and emanations from tape material have affected the SOA yields for the isoprene emitter?

All non-Teflon components were on the exterior of the chamber or frame in order to prevent any contamination from these components entering the chamber (Plant Chamber Design, Methods and Materials). This is stated in section 2.2, p 14297, lines 2 and 3: “The interior of the plant chamber was only exposed to Teflon surfaces.”

Irrespective of this, both the plant chamber and reaction chamber were tested for contaminants separately and when joined together by running the system with an empty plant chamber and by carrying out a blank run prior to each set of experiments. For clarity, this statement has been added to the end of section 2.3.

Section 2.5, p. 14302

“Both the pots and soil were isolated by enclosing them in PFTE sheeting; this acted to prevent VOC emissions from the plastic pots and soil NO_x emissions from entering the chamber air.” Was this explicitly checked?

NO_x emissions into both the plant and reaction chamber were explicitly checked and no significant NO_x emissions from the pots or substrate were detected.

Mass balance considerations (section 4.1, p. 14315, l.26ff and section 4.2, p. 14319, l.4ff):

The authors argue that, *cum grano salis*, the mass balance is closed in the experiments (Figures 3 and 6). However, the mass should increase by addition of oxygen. The conservative quantity is the amount of carbon. Insofar the mass balance discussion is inconclusive; it should be cancelled/or weakened and or the Carbon balance should be discussed.

We thank the reviewer for pointing this out; it was clearly an oversight on our part during compilation of the text. Consequently, the following changes have been made to the manuscript to correct for this and to account for the reviewers' suggestions:

Section 4.1, p. 14315, l.26ff, paragraph altered:

“As can be seen from inspection of Fig. 3, the transfer of mass through the *Betula pendula* experiment **was appeared** roughly conservative, with **a small and steady the primary point of note being a small** loss of measured mass from the **reaction matrix over time (after ~ 220 min)**. With the addition of oxygen to the starting body of hydrocarbon material during such an experiment, the total measured mass (i.e. Σ VOCs + SOA) within the system would be expected to **increase with time**. The **absence of such a total measured mass gain (and indeed the mass deficit observed towards the end of the experiment)**, can most likely be accounted for by considering the **various** measurement uncertainties involved **in producing these data** (e.g. assumptions in PTR sensitivity, uncharacterised fragmentation following ionisation, instrument detection limits, etc.) and

influences imposed by the chamber walls (including potential loss of more highly oxidised material from the gas phase and greater than expected loss of SOA). Indeed, there is potential for a system mass increase by the end of the experiment to lie within the uncertainty bounds of the CIR-TOF-MS/PTR-MS measurements alone, e.g. the average, single compound PTR measurement uncertainty is $\sim \pm 30\%$, allowing the final measured value of $130 \mu\text{g m}^{-3}$ to have an upper limit of $170 \mu\text{g m}^{-3}$, i.e. greater than the starting value. Considering these results it seems that the system studied is reasonably well characterised given the complications involved in such a task.”

Section 4.2, p. 14319, l.4ff, paragraphs altered:

“~~As with the *Betula pendula* experiments,~~ As can be seen in Fig. 6 mass transfer through the *Ficus benjamina* system was ~~relatively conservative, with only~~ characterised by a slight mass decrease measured during the central phase just after the start of the experiment followed by a gradual increase in mass with time. As was stated in section 4.1 a mass increase is expected with time during such an experiment, owing to the addition of oxygen to the precursor hydrocarbon material. Consequently, when considering the data presented in Fig. 6 in the context of potential uncertainties involved (including difficult to characterise influences imposed by the chamber walls), it appears ~~indicating again~~ that the system being studied is reasonably well characterised.

By comparing Figs. 3 and 6 we see that the monoterpene dominated *Betula pendula* system, which produces larger and lower vapour pressure oxidation products than the isoprene dominated *Ficus* system, as well as measureable SOA, is the case which exhibits measured mass loss. From this contrast it is reasonable to assume a significant fraction of any mass deficit observed during *Betula pendula* oxidation could result from the loss of the heavier, lower volatility compounds that are present in the *Betula pendula* oxidation system but not in the *Ficus* system.”

Moreover, the mass concentrations axis on the right side in Figures 3 and 6 seems strange. Did you really measured exactly 100 respectively 200 $\mu\text{g}/\text{m}^3$ in the experiments?

The right axis of both Figures 3 and 6 were manually scaled to the nearest $100 \mu\text{g m}^{-3}$ for neatness of appearance and to allow the reader to best resolve the black “total mass” line against the coloured bars in the background. Altering the right axis scale in either case to values other than 200 and $100 \mu\text{g m}^{-3}$, respectively makes the graph more difficult for the reader to view and interpret. We have however, included additional text in the legend of figures 3 and 6 to clarify the role of the axes:

“**Figure 3.** Evolution of mass through the *Betula pendula* system (7 July 2009), showing the relative contribution of precursor compounds, oxidation products and SOA mass to total measured mass, with time (coloured bars, left axis) and total measured mass (i.e. $\Sigma\text{VOCs} + \text{SOA}$) with time (black line, right axis). Note: ammonium sulphate seed mass removed from the SOA mass concentration.”

“**Figure 6.** Evolution of mass through the *Ficus benjamina* system (23 June 2009), showing the relative contribution of precursor compounds and oxidation products to total measured mass, with time (coloured bars, left axis) and total measured mass (i.e. Σ VOCs + SOA) with time (black line, right axis).”

Missing detection limits:

Section 2.4.1, p. 14299

Detection limits are not stated for PTR-MS and CIR-TOF-MS, however, these are used to argue later, e.g. sec 3.1.1, p. 14307, lines 5f, in the isoprene case (sec. 4.2, p. 14320, line 18ff).

As CIR-TOF-MS and PTR-MS detection limits are method, reagent, reaction, matrix and analyte specific, it is always difficult to give a simple statement detailing the “detection limit” of the technique, however a statement has been added to section 2.4.1, p. 14300 giving a range of typical values along with a reference for further information:

“CIR-TOF-MS and PTR-MS detection limits are reagent, reaction, analyte and sample matrix specific. However, typical CIR-TOF-MS detection limits, using PTR ionisation from hydronium, are of the order $0.4 \text{ ppbV (10 min)}^{-1}$ for more polar compounds, such as OVOCs (e.g. 2-hexanone) and as much as 10 ppbV min^{-1} for certain less polar compounds, such as smaller hydrocarbons (e.g. 1-pentene). For further details see Wyche et al., 2007.”

SOA yields for birch:

Values in abstract and section 3.1.2 do not match the values given in section 4.1 (p. 14315, l. 13). This questions the comparison with Mentel et al. 2009, and the qualifying as low end yields.

We thank the reviewer for point this out, the values in section 4.1 are simply “typos”, and the values given in the abstract and section 3.1.2 are indeed the correct values. The values of 16 and 11 % given in section 4.1 have been corrected to 39 and 26 %. Furthermore, the statement referring to Mentel et al. 2009 has been altered appropriately:

“~~Furthermore~~In a comparable study to ours, Mentel et al., (2009) reported a fractional mass yield of 11 % for their *Betula pendula* experiments, i.e. slightly lower than those given here, but within the bounds of quantified experimental errors. The yield values obtained here for the *Betula pendula* mesocosm system lie ~~at roughly in the middle lower end~~ of the single precursor yield range.”

Minor remarks

p. 14305, §1st : [OH] for birch and palm experiments are not given? How large were they? p. 14306, l. 25: “...only $2.0 (\pm 1.0) \text{ ppbV}$ isoprene was detected”. The fraction of isoprene amounts to 10-15%. This is a significant contribution! Please, comment in the manuscript.

[OH] could not be calculated for Birch experiments because the ozone reacted preferentially with the large concentrations of monoterpenes and sesquiterpenes produced. Therefore it was not possible to use the more stable isoprene (as in the ficus experiments) to calculate the production of [OH] from the reaction of isoprene with ozone. The palm trees were only used in a mixed canopy experiment but these results were not discussed in the manuscript.

p. 14306, l. 25f : Table 3 states the species found but not their relative abundance !

We are not quite sure what the reviewer is referring too here; Table 3 gives the yields of the species listed, in effect, this is the abundance.

p. 14308, l. 14ff: The authors discuss lifetimes but they quote rate coefficients. I suggest to use a mean [OH] from the experiments and explicitly state the lifetimes.

As requested, lifetimes have now been given (p. 14308, l. 14ff)

p. 14308, l. 24ff: α -hydroxy carbonyl compounds from OH addition to double bonds have also short lifetimes.

This has been noted in the text on p. 14308 by inclusion of the following additions to the text:

“This relatively short lifetime gives further insight into the potential identity of the m/z 111 and 93 signals, perhaps indicating the presence of multiple C=C bonds in the hydrocarbon structure, as would be found in the primary C₇ aldehydes obtained from the oxidation of ocimene or myrcene for example. Other short-lived biogenic oxidation products that could exist in such mesocosm systems, include α -hydroxy carbonyls, similarly formed following OH addition to a C=C bond.”

p. 14321, §2nd: This paragraph is difficult understand. A few more descriptive words about the concept and what the authors are aiming at will be helpful.

The paragraph has been altered as suggested:

“For 78 of the 120 measurement-and-parameter sets tested, the estimated residual SOA mass resulting solely from isoprene oxidation was negative – i.e., production of SOA from isoprene oxidation was not required to close the mass balance. Values were calculated based on the widest range of peak masses observed during the *ficus* experiments ($M_p = 1.3 \mu\text{g m}^{-3}$ and $5.5 \mu\text{g m}^{-3}$), and assume the lowest (29 %) and highest (100 %) VOC terpene yields and lowest (5 %) and highest (47 %) SOA yields from non-isoprene precursors, respectively, as observed in previous experiments conducted within this chamber. These ranges result in calculated residual SOA mass of -28.5 to $+5.0 \mu\text{g m}^{-3}$ produced

solely from isoprene oxidation. Hence, there are combinations of measurements, observations and oxidation/phase-change parameters — omitting isoprene and its oxidation products — that can account for ~20 times the observed aerosol mass production, and other combinations of measurements and parameters that leave up to ~ 90 % of the condensed mass to be explained by isoprene oxidation. If, instead of using the limiting cases, the closest approximation to the *ficus cyathistipula* system is used (i.e. $Y_{\text{VOC}} = 77\%$ and $Y_{\text{SOA}} = Y_{\text{SOA}} = \alpha\text{-pinene} = 15\%$), non-isoprene products could have accounted for around 145 % of the SOA mass that was produced. We have no way of assigning formal likelihoods to each set of measurements and parameters in this exercise, but we note that the great preponderance of parameter combinations do not require an isoprene contribution to the SOA mass (i.e. 78/120 measurement-and-parameter sets tested) under our experimental conditions. Moreover, our experiments produce much less SOA mass than would be expected from published experiments using individual mono- and sesquiterpenes.”

p. 14326/14327: Did the authors found indications for induced emissions in PTR-MS data etc. ?

No, we did not observe any evidence to suggest that exposure to ozone induced the emission of additional VOCs from the trees, over and above the ones emitted in the absence of ozone; i.e. there was no evidence of additional (i.e. induced) emissions after the onset of ozone exposure – only the formation of ozone-VOC reaction products.

The reference Hamilton et al. (2013) is missing in reference list.

The reference has been added:

Hamilton, J. F., Alfarra, M. R., Robinson, N., Ward, M. W., Lewis, A. C., McFiggans, G. B., Coe, H., and Allan, J. D.: Linking biogenic hydrocarbons to biogenic aerosol in the Borneo rainforest. *Atmos. Chem. Phys.*, 13(22), 11295-11305, doi:10.5194/acp-13-11295-2013, 2013.

Table 2: The authors should state relative abundances, at least for the main components.

A new table 2 has been created with abundances and inserted in the manuscript.

Figure 3 and Figure 6: There are already reaction products in the first bin. I suggest plotting the VOC mix just before the oxidation starts into the first bin.

Yes, the reviewer is correct there are, and there should be, reaction products in the first bin(s), as the first bin is the 10 minute data point from lights on to + 10 minutes after lights on. Hence, as the reaction has begun, we expect products to begin to evolve. If the reader requires information on pre-lights on data, they

can simply refer to Figs. 2 and 5, which provide this information in a much clearer manner. We feel that adding more data to Figs. 3 and 6 (that is already given in the previous figures), which are already somewhat congested with data, would reduce the clarity of the information we are trying to impart. As such and as the reviewer has no major concerns regarding these figures, we would like to leave them as they are.

Typos

p. 14295, l. 6: Jiang *Done*

p. 14309, l. 21: Hex *e* nal ? *Done*

p. 14310, l. 2: towards *the end of* the experiment ? *Done*

p. 14314, l. 11: calcu *l* ation *Done*

p. 14317, l. 6: Figure *10* ? *Done*

p. 14323, l. 1: (Kiendler-Scharr et al., 2009 *a*) ? *Done*

p. 14326, l. 6: (Kiendler-Scharr et al., 2009 *a*) ? *This line does not refer to this reference?*

Figure 9, captions, a blank is missing before Ficus *Done*

Anonymous Referee #2

Interactive comment on “Emissions of biogenic volatile organic compounds and subsequent photochemical production of secondary organic aerosol in mesocosm studies of temperate and tropical plant species” by K. P. Wyche et al.

Received and published: 7 August 2014

Overall Comment and Recommendation:

This manuscript examines the emissions of BVOCs from silver birch and three South-east Asian tropical plants grown in a whole-tree chamber. These BVOCs were then transferred to an irradiation chamber for subsequent production and characterization of gas- and aerosol-phase oxidation products. I have a number of reservations about the current version of this manuscript. I list these below and strongly encourage the authors to address these carefully before acceptance of this manuscript can be considered.

I was surprised how little particle-phase data was presented and discussed from off-line filter analyses.

The data from the filters was intended for use to highlight the formation of SOA and to highlight some major features in what is a complex sample. A full and rigorous analysis of the filter samples that we were able to record is outside of the scope of this (what is already a somewhat large) piece of work.

Further, the SOA yields presented and discussed have many hidden issues that have to be carefully addressed in a revised manuscript.

In the main, we agree with the reviewers' comments here regarding making yield estimates; they can be misleading and must be treated in context. Having said this, our methodology of yield calculation, taking into account the size dependent particle losses to the walls, is at least as rigorous as those from any other group or published work, more so than many in fact. We recognise that yields are completely dependent on the experimental conditions (e.g. T, RH, oxidant, VOC/NO_x ratio, VOC concentration, light intensity and spectrum); this clearly means that chamber SOA yields are not necessarily directly comparable between conditions, unless all others are held constant. Furthermore, we recognise that SOA yields are also dependent on the specific chamber in which the experiment is conducted (wall material, surface area to volume ratio, mixing timescale). This means that the yields are not necessarily comparable between chambers. Finally, and most importantly, the yield is dependent on the volatility of the products and the ratio of the mass transfer rate of the products to the particles and the loss rates to the walls.

However, owing to all of these issues, we have attempted to be deliberately transparent in our discussion of our SOA yield calculation methodology. We feel that owing to the extensive use of SOA yield values in the literature, it was necessary and in the context of all of the above, informative to the reader to do so.

Again, we fully recognise the views of the reviewer in this regard, however, we believe that as we have included all of our working methodology, and that the assumptions required are clear, any reader can understand what we have done in their own interpretation of the data/findings presented.

Overall, I recommend that this paper is not accepted until the Editor feels these specific comments below are adequately addressed.

Specific Comments:

1.) SOA yields and ELVOCs:

Based on the extra low-volatility VOCs (ELVOCs) recently discovered by Ehn et al. (Nature Letters, 2014) and isoprene-epoxydiols (IEPOX) (Paulot et al., 2009, Science), can the authors comment on the wall losses for these sticky low-volatile compounds? Specifically, how might these losses affect the interpretation of these results? This seems like a reasonable question to ask of the authors, especially considering that they report SOA yields throughout the entire text. These losses seem quite apparent, especially considering the effect of having pre-existing seed aerosol; that is, SOA was measurable and seed aerosol was typically more conducive to increasing SOA yields. My guess is as you guys increase the total surface area of your inorganic seed aerosol, you see more OA growth as a result of lower wall losses. Did the authors systematically test this? Since this is likely an issue, why focus the discussion on SOA yields?

As with every single other chamber experiment conducted so far, our yield calculations do not take into account the losses of these compounds. The text has been altered to state this explicitly:

“It should also be noted that along with previous caveats made regarding the role of the chamber walls and other measurement uncertainties, these yield values also do not take into account the potential loss of particularly “sticky” low volatility compounds (e.g. Ehn et al., 2014) to internal surfaces of the chamber.”

Losses to walls will be in competition with the losses that contribute to the condensational growth. Each of these losses results from the product of two terms - i) the mass transfer rate to the available surface and ii) the available effective absorptive surface / mass. Any ELVOCs that are produced, will be produced everywhere in the chamber and the seed particles will provide their effective mass uniformly throughout the chamber. It is likely that the ELVOC molecules formed away from the walls will collide with a particle before they collide with the wall and will stick irreversibly to it. ELVOC molecules formed nearer the walls will become increasingly likely to stick to the walls. The effective wall surface so far as the molecules in the middle of the bag are concerned is negligible (or it could alternatively be viewed that the ratio of likely number of collisions with the wall : likely number of collisions with particles, is very low), increasing towards the wall. The results should therefore be interpreted that the decrease in yields attributable to ELVOCs in a smaller

bag will be higher than in a larger bag. BUT ELVOCs have only been seen in terpene ozonolysis and hence this argument is not true for isoprene.

Furthermore, in addition to there being no evidence that isoprene actually yields ELVOCs, there is conflicting evidence about IEPOX. If it is assumed that reactive uptake requires acidity, then there is no reason to believe our walls are acidic and our particles are likely close to neutral (being originally ammonium sulphate). The neglect of wall effects with respect IEPOX uptake in such a case will likely be minimal. If we proceed with the findings of Nguyen et al., then the wall uptake will be in competition with the seed uptake and the same argument will apply as for ELVOCs above (and the decrease in yields attributable to IEPOX in a smaller bag will be higher than in a larger bag). But this is in apparent contradiction to the fact that the lowest yields have been observed in the largest bag (i.e. SAPHIR).

It is a naive simplification to state that more OA growth will occur with more seed. There are many other concomitant processes at play. We have not systematically tested this - the signal was very small and we were conducting very low realistic precursor concentration experiments.

2.) Page 14295, Lines 9-12:

The authors forget to mention the important effect of acidic aerosol on the reactive up- take of isoprene epoxydiols (IEPOX), which are major 2nd generation oxidation products from isoprene that yield most of the SOA from isoprene under low-NO conditions (Surratt et al., 2010, PNAS; Lin et al., 2012, ES&T).

On the contrary, Nguyen et al., (2014) found that “The results are consistent with weak correlations between IEPOX-derived OA and particle acidity or liquid water observed in field studies, as the chemical system is nucleophile-limited and not limited in water or catalyst activity”. The manuscript has been altered to reflect the work that supports the contention of the reviewer and that of Nguyen et al. to illustrate the conflicting literature evidence:

“Modelling, laboratory chamber experiments and field studies provide a range of possible yields of SOA from isoprene, typically of the order 0.4 – 3 % by mass, with some values reported as high as 5.5 % (van Donkelaar et al., 2007;Kleindienst et al., 2009, 2007;Kroll et al., 2005, 2006;Claeys et al., 2004a;Edney et al., 2005). SOA yields from the further oxidation of first and subsequent generation isoprene oxidation products, such as methacrolein, are is estimated to be as much as ~~up to~~ 15 % (Rollins et al., 2009;Carlton et al., 2009;Claeys et al., 2004b;Robinson et al., 2010). Recent work has highlighted that under low NO_x conditions, SOA mass formed from isoprene oxidation could be influenced by the acidity of pre-existing aerosol via the reactive uptake of certain key isoprene oxidation products, namely isoprene epoxydiols (IEPOX; Surratt et al., 2010; Lin et al., 2012). More recently, Nguyen et al. (2014) found that the “pH dependence for OA formation from IEPOX was weak for AS particles”. There is further evidence from chamber studies using temperate tree species such as birch, spruce and pine that isoprene may in fact suppress SOA formation from other VOC precursors, when present when present (Kiendler-

Scharr et al., 2009a; Kanawade et al., 2011).” It should be noted however; it is unclear in most cases how wall effects have been considered in the production of such yield values and whether the treatments employed are adequate such that the yields are comparable between chambers, or indeed between experiments.

3.) Page 14296, Line 25:

Did the plywood base emit BVOCs? Could these have leaked into chamber?

It is unlikely that the plywood base emitted VOC into the plant chamber as it was covered in its entirety in foil and then Teflon sheeting. To ensure any background VOC were accounted for, blank chamber samples were taken prior to each change of experiment and then background subtracted from the biogenic samples. Following comments from reviewer #1, this has been clarified in the manuscript (see above).

4.) Page 14301-14302, Line 28 and Lines 1-2:

Why wasn't the chemical composition dependent CE calculated as recently done by Middlebrook et al. (2012, AS&T)?

The AMS data reported in this manuscript was only limited to one SOA nucleation experiment where the particles were composed of only organic material, therefore the composition dependent CE suggestion is not relevant to the reported data. In addition, we only reported the fractional contribution of m/z 44 to total organic mass, which is, being a ratio, not dependent on the applied CE value.

5.) Page 14302, Lines 3-5:

Were denuders used in front of quartz filters? I worry that gaseous absorption of ELVOCs or IEPOX could have occurred on these filters, and thus, skewing the chemical composition results due to positive artifact formation. Have the authors confirmed there are no artifacts. I have to admit the data presented from these filters is very weak and limited, and doesn't really add much to the text in terms of sources of SOA.

Aerosols were collected without the use of denuders. However, they were collected using a flow rate of 3 m³/min over a period of around 6 minutes and were frozen within a further few minutes of collection. The procedure is much faster than traditional filter collection methods and should minimise the negative and positive artefacts. It is also important to emphasise that filters were collected at the end of the experiment, i.e. after several hours of photochemistry and therefore any ELVOCs, which were produced earlier should have had enough time to already condense onto particles. As the precursor VOCs concentrations had reached a minimum by this point, there was little source for ELVOC production when the filters were collected. It is also unlikely that any ELVOCs present would preferentially condense to the filters when the entire chamber contained more readily available particle surfaces on which to

condense to. As the filters were not acidic there is no evidence that IEPOX could condense effectively onto them.

Minor appropriate additions to the text have been applied:

“Filter samples for offline analysis were collected (without denuders) in a specially constructed holder, positioned in the chamber vent line. Aerosol samples were collected onto 47 mm quartz fibre filters (Whatman) at a rapid flow rate of 3 m³ min⁻¹ (sample time ca. 6 mins.). After sampling, filters were immediately placed in pre-cleaned glass vials and stored below -20 °C until analysis. The filter collection procedure employed here is much faster than traditional filter collection methods, which should minimise any potential negative and positive artefacts.”

6.) Page 14302, Line 7:

Why didn't the authors consider extracting the filters with a more flexible solvent (such as methanol) that can remove both polar and less polar SOA compounds? I worry that water extractions really removes only the most polar SOA constituents and from my experiences with monoterpenes (and especially sesquiterpenes). I have found methanol is a better solvent for these two classes of VOCs. Thus, do the authors worry they are missing important aerosol compounds in their off-line chemical analyses?

In contrast to the reviewer, in our experience we have found very little difference between using methanol and water as the extraction solvent (see Hamilton et al., Characterization of Polar Compounds and Oligomers in Secondary Organic Aerosol Using Liquid Chromatography Coupled to Mass Spectrometry, 2008, Analytical Chemistry), with slightly higher recoveries of most species in water. Therefore, we feel it is unlikely that methanol will improve the extraction.

7.) Page 14303, Line 12-15:

What is the concentration of your atomizing solution? This should be listed.

A typical ammonium sulphate solution concentration of 2 g/l was employed. However, as the number, mass and size distribution of the seed produced was always well characterised, the solution strength was not rigorously document.

How much volume of seed aerosol was injected into your experiments? This detail should be added to the experimental section or in the Table summarizing your experiments.

The injected seed volumes information is already included in Figure 8 in the form of seed mass at time zero, i.e. before the start of photochemistry.

Did you all calculate the aerosol pH of your atomized seed aerosol? Since BVOC-derived SOA (especially isoprene-derived SOA) is so sensitive on aerosol pH, this could be an important parameter to add as well.

No, we did not explicitly measure the solution pH and as the solution strength was not rigorously defined, it would be somewhat meaningless in this instance to attempt to calculate the aerosol pH. Aerosol acidity was not a key driver in this study, we did not have the time of resources to enter into this complex issue, which would comprise a whole separate study on its own. However, this is something that we are considering for the next phases of our work.

8.) Page 14303, Lines 16-28:

I would argue that for isoprene SOA formation, there is plenty of literature now that clearly shows that it forms due to REQUIRED acidity that allows for the reactive uptake and subsequent particle-phase chemistry of IEPOX (Surratt et al., 2010, PNAS; Lin et al., 2012; Nguyen et al., 2014, ACP) and MAE (Lin et al., 2013, PNAS). Without acidic particles, SOA formation from isoprene will be quite limited (e.g., Lin et al. 2012, ES&T). Thus, I wonder how relevant these experiments are for isoprene SOA formation?

Whilst it is recognised that there is near consensus in the literature that particle acidity is required for formation of SOA from isoprene (with the exception, for example, of Nguyen et al 2014 - who also investigated wet ammonium sulphate seed - and Brégonzio-Rozier et al., ACPD 2014), it is far from clear that ambient particles will always carry substantial acidity where isoprene photochemistry is active; for example the Amazon:

“Sulphate levels in Borneo are around four times greater than in the Amazon. Inspection of back trajectories suggests marine and anthropogenic sources of sulphate external to Borneo (Robinson et al., 2011a). A charge balance of sulphate and ammonium ions show excess sulphate over the oceans (Robinson et al., 2011b) compared to the ground site where charge is usually balanced. As acidic sulphate had been shown to play an important role in isoprene SOA formation in previous studies, its presence may contribute to the greater significance of MF in Borneo, although chamber studies have not shown sulphate isoprene SOA mechanisms likely to yield MF (Surratt et al., 2010)” - Robinson et al. (ACP 11, 1039-1050, 2011).

The presence of a more neutral or non-acidic, background aerosol in locations such as the Amazon (and also the lack of evidence for prolific distributions of acidic aerosol), lends support to the relevance of our work and lends a degree of justification to our use of an ammonium sulphate seed (which better approximates ambient conditions in such geographical locations). Indeed there perhaps exists a counter argument, which could debate the realism of conditions employed in some experiments that report the importance of acid-catalysed reactions, without reference to the real atmosphere.

In summary, one has to be cautious regarding the role of acidity of the aerosol. There is no doubt that Surratt et al. have shown the importance of acidity of the

aerosol, however in contrast Nguyen et al. (2014) and Brégonzio-Rozier et al., (2014) have shown SOA formation on neutral seed. As yet there appears to be no clear picture on the effect of acidity on isoprene SOA formation, we are not attempting here to make an argument for one case or another, we are simply reporting our observations and discussing them in open context.

The seed aerosol in our experiments were always “wet”,

9.) Page 14319, Lines 18-21:

I would argue that this is due to the lack of acidic sulfate aerosol. This has been repeatedly shown as a requirement to produce isoprene SOA (Edney et al., 2005; Surratt et al., 2007; Surratt et al., 2010; Lin et al., 2012; Nguyen et al. 2014).

We are slightly confused as to the referees’ contention here, owing to the contradiction they make in the next point, where the referee then states that they “would suspect a LOT of isoprene SOA (under near-neutral conditions) in the aerosol phase, especially based on recent work by Nguyen et al.”, i.e. in opposition to their statement here.

However, as we stated above in our reply to point (8), there does exist some lack of consensus in the literature regarding the role of pre-existing acidic aerosol in the production of SOA from isoprene oxidation (e.g. Nguyen showed OA formation on neutral ammonium sulphate aerosol). However, this section of the manuscript is discussing nucleation of gases in the absence of a pre-existing surface. Consequently, the argument regarding the presence of a certain type of seed is not relevant here. Moreover, in the same paragraph, just prior to this statement, we do indeed acknowledge the potential role played by a lack of seed in these particular experiments:

“A lack of SOA mass formation during our unseeded Ficus benjamina experiments could have resulted from a number of different factors, not least of which was simply the absence of a seed surface to help facilitate partitioning of the semi-volatile oxidation products to the aerosol phase and produce particles of sufficient size and measureable particle mass”.

In order to acknowledge the potential role played by a lack acid surface (specifically) here, this statement has been altered slightly in the revised manuscript:

“A lack of SOA mass formation during our unseeded Ficus benjamina experiments could have resulted from a number of different factors, not least of which was simply the absence of a seed surface (acidic or otherwise) to help facilitate partitioning of the semi-volatile oxidation products to the aerosol phase and produce particles of sufficient size and measureable particle mass”.

10.) Page 14321, Lines 27-28:

Why wasn’t off-line filter characterization data presented for the tropical plants to confirm that isoprene oxidation wasn’t making much SOA in the seeded experiments?

Since the experiments are very humid, I would suspect a LOT of isoprene SOA in the aerosol phase, especially based on recent work by Nguyen et al. (2014) that utilized only ammonium sulfate seed aerosol.

Owing to the relatively low level of isoprene present in the chamber, insufficient aerosol mass was formed during these experiments to allow us to make compositional measurements using the techniques we had available e.g. see Figure 8. To clarify this point, the following statement has been added to the end of section 4.2 (page 14322), line 10:

“Unfortunately, insufficient SOA mass formed during *Ficus* experiments to allow us to conduct any form of compositional analysis.”

Here we are reporting that we did not see any mass formed above background; we are simply reporting what we saw in these instances. There is some degree of split in the literature regarding this and we hope that our findings can add to the debate.

11.) Page 14322, Lines 16-22:

Not EXACTLY. The methyl furan resulted from the decomposition of IEPOX-derived SOA, as recently shown by the Surratt group (Lin et al., 2012, ES&T; Budisulistiorini et al., 2013, ES&T).

This is correct. It was stated in Robinson et al. that the 3MF was a thermal decomposition product of isoprenoid SOA. However the following text and the suggested references have been added on P14322 to clarify:

“(i.e. thermal decomposition of isoprene derived SOA)”

12.) Page 14323, Lines 19-23:

Not only contrasting NO_x environments, but also contrasting aerosol acidity environments (as shown by Lin et al., 2012, ES&T; Lin et al., 2013, ACP; Pye et al., 2013, ES&T).

As was stated in the above replies to reviewer comments, there does exist some degree of disagreement in the literature (and in the reviewers commentary) regarding the need for/role of acidic seed (c.f. Nguyen et al., 2014; Brégonzio-Rozier et al., 2014; see previous replies). Owing to the volume of work required in this area the dependence on acidity should have to be the subject of further work and was never intended to be the primary driver of this study. However, as suggested by the reviewer, the additional potential importance of acidity of the environment has been noted with the following changes to the text, P14323, L 19 – 23:

“Understanding the exact role played by isoprene in air containing many different VOCs, and being able to account for the differing isoprene SOA yields under contrasting NO_x and acidity (Lin et al., 2012; Lin et al., 2013; Pye et al., 2013) environments, will undoubtedly help to significantly improve global

modelling estimates of total SOA loading even further (Couvidat and Seigneur, 2010)”

Minor Comments:

- 1.) Page 14294, Line 12: insert "have an" *Done*
- 2.) Page 14302, Line 9: Delete the "d" *Done*
- 3.) Page 14303, Line 4: insert comma between ppbv and respectively *Done*
- 4.) Page 14320, Line 10: Please provide acitations for this range of yields. *Done*
- 5.) Page 14324, Lines 1-2: Please provide supporting citations for this statement. *Done, Mentel et al., has been referenced.*

References added, following alterations as suggested by reviewer 2:

Budisulistiorini, S. H., Canagaratna, M. R., Croteau, P. L., Marth, W. J., Baumann, K., Edgerton, E. S., Shaw, S. L., Knipping, E. M., Worsnop, D. R., Jayne, J. T., Gold, A. and Surratt J. D.: Real-Time Continuous Characterization of Secondary Organic Aerosol Derived from Isoprene Epoxydiols in Downtown Atlanta, Georgia, Using the Aerodyne Aerosol Chemical Speciation Monitor, *Environ. Sci. Technol.*, 47 (11), 686–5694, doi: 10.1021/es400023n, 2013

Brégonzio-Rozier, L., Siekmann, F., Giorio, C., Pangui, E., Morales, S., B., Temime-Roussel, B., Gratien, A., Michoud, V., Ravier, S., Tapparo, A., Monod, A. and Doussin J.-F.: Gaseous products and Secondary Organic Aerosol formation during long term oxidation of isoprene and methacrolein, *Atmos. Chem. Phys. Discuss.*, 14, 22507-22545, 2014

Ehn, M., Thornton, J. A., Kleist, E., Sipila, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F., Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I. H., Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B., Jorgensen, S., Kjaergaard, H. G., Canagaratna, M., Dal Maso, M., Berndt, T., Petaja, T., Wahner, A., Kerminen, V. M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of low-volatility secondary organic aerosol, *Nature*, 506 (7489), 476-479, doi: 10.1038/nature13032, 2014

Lin, Y. H., Zhang, Z., Docherty, K. S., Zhang, H., Budisulistiorini, S. H., Rubitschun, C. L., Shaw, S. L., Knipping, E. M., Edgerton, E. S., Kleindienst, T. E., Gold, A., and Surratt, J. D.: Isoprene Epoxydiols as Precursors to Secondary Organic Aerosol Formation: Acid-Catalyzed Reactive Uptake Studies with Authentic Compounds, *Environ. Sci. Technol.*, 46 (1), 250–258, doi: 10.1021/es202554c, 2012

Lin, Y. H., Knipping, E. M., Edgerton, E. S., Shaw, S. L., and Surratt, J. D.: Investigating the influences of SO₂ and NH₃ levels on isoprene-derived secondary organic aerosol formation using conditional sampling approaches, *Atmos. Chem. Phys.*, 13, 8457-8470, doi: 10.5194/acp-13-8457-2013, 2013

Pye, H. O. T., Pinder, R. W., Piletic, I. R., Xie, Y., Capps, S. L., Lin, Y. H., Surratt, J. D., Zhang, Z., Gold, A., Luecken, D. J., Hutzell, W. T., Jaoui, J., Offenberg, J. H., Kleindienst, T. E., Lewandowski, M., and Edney E. O.: Epoxide Pathways Improve Model Predictions of Isoprene Markers and Reveal Key Role of Acidity in Aerosol Formation, *Environ. Sci. Technol.*, 47 (19), 11056–11064, doi: 10.1021/es402106h, 2013

1 **Emissions of biogenic volatile organic compounds and subsequent photochemical**
2 **production of secondary organic aerosol in mesocosm studies of temperate and**
3 **tropical plant species**

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30

31 **Abstract:** Silver birch (*Betula pendula*) and three Southeast Asian tropical plant
32 species (*Ficus cyathistipula*, *Ficus benjamina* and *Caryota millis*) from the
33 pantropical fig and palm genera were grown in a purpose-built and environment-
34 controlled whole-tree chamber. The volatile organic compounds emitted from these
35 trees were characterised and fed into a linked photochemical reaction chamber where
36 they underwent photooxidation under a range of controlled conditions (RH ~ 65 – 89
37 %, VOC/NO_x ~ 3 – 9 and NO_x ~ 2 ppbV). Both the gas phase and the aerosol phase of
38 the reaction chamber were monitored in detail using a comprehensive suite of on-line
39 and off-line, chemical and physical measurement techniques.

40 Silver birch was found to be a high monoterpene and sesquiterpene, but low isoprene
41 emitter, and its emissions were observed to produce measureable amounts of SOA via
42 both nucleation and condensation onto pre-existing seed aerosol (Y_{SOA} 26 – 39 %). In
43 contrast, all three tropical species were found to be high isoprene emitters with trace
44 emissions of monoterpenes and sesquiterpenes. In tropical plant experiments without
45 seed aerosol there was no measurable SOA nucleation, but aerosol mass was shown to
46 increase when seed aerosol was present. Although principally isoprene emitting, the
47 aerosol mass produced from tropical fig was mostly consistent (i.e., in 78 out of 120
48 aerosol mass calculations using plausible parameter sets of various precursor specific
49 yields) with condensation of photooxidation products of the minor VOCs co-emitted;
50 no significant aerosol yield from condensation of isoprene oxidation products was
51 required in the interpretations of the experimental results. This finding is in line with
52 previous reports of organic aerosol loadings consistent with production from minor

53 biogenic VOCs co-emitted with isoprene in principally-isoprene emitting landscapes
54 in Southeast Asia. Moreover, in general the amount of aerosol mass produced from
55 the emissions of the principally-isoprene-emitting plants, was less than would be
56 expected from published single-VOC experiments, if co-emitted species were solely
57 responsible for the final SOA mass. Interpretation of the results obtained from the fig
58 data sets, leaves room for a potential role for isoprene in inhibiting SOA formation
59 under certain ambient atmospheric conditions, although instrumental and
60 experimental constraints impose a level of caution in the interpretation of the results.

61 Concomitant gas and aerosol phase composition measurements also provide a detailed
62 overview of numerous key oxidation mechanisms at work within the systems studied
63 and their combined analysis provides insight into the nature of the SOA formed.

64

65 *Keywords: Secondary organic aerosol, biogenic volatile organic compounds, BVOC,*
66 *gas-aerosol partitioning, isoprene, monoterpenes, mesocosm*

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71 **Introduction**

72 Atmospheric aerosols change the radiative balance of the Earth through scattering and
73 absorbing incident solar radiation (Kim and Ramanathan, 2008); they directly and
74 indirectly affect the properties and formation of clouds, thus altering the hydrological
75 cycle (Gunthe et al., 2009; Junkermann et al., 2009; Stevens and Feingold, 2009); and
76 they may [have an](#) impact on the efficiency of plant photosynthesis (Mercado et al.,
77 2009), thereby modifying the uptake of atmospheric carbon. Hence, aerosol particles
78 affect the Earth's climate in several ways (as reviewed in Hallquist et al., 2009; IPCC,
79 2007; Isaksen et al., 2009; Carslaw et al., 2010) as well as having a detrimental impact
80 on human health (e.g., Baltensperger et al., 2008).

81 A large fraction of the observed atmospheric aerosol composition is organic (Zhang et
82 al., 2007). A primary organic component is emitted directly into the atmosphere from
83 anthropogenic activities, such as biomass burning and fossil fuel combustion, or is
84 emitted from natural sources, such as plant abrasion and the sea surface. Secondary
85 aerosol particles are formed within the atmosphere by gas-to-particle conversion;
86 those formed from gas-phase organic precursors are known as secondary organic
87 aerosol (SOA) (e.g., Riipinen et al., 2012). There is considerable uncertainty
88 surrounding the chemical transformation of anthropogenic and biogenic volatile
89 organic compounds (AVOC and BVOC, respectively) from the gas phase to the
90 aerosol phase and hence, considerable uncertainty in the global source of SOA
91 (Hallquist et al., 2009; Donahue et al., 2009; Ng et al., 2006; Virtanen et al., 2010).

92 On a global scale, approximately 90 % of all volatile organic compound emissions
93 originate from biogenic sources (Guenther et al. 2012), with almost half of this being
94 emitted from tropical and subtropical forests. The ability of biogenic VOC to form
95 SOA is therefore of particular interest and potential importance. Globally, isoprene
96 (2-methyl-1,3-butadiene, C₅H₈) is the biogenic VOC with the largest mass emission
97 rate. It is estimated to account for about 50 % of BVOC emissions by mass (Guenther,
98 et al. 2012), but it is still uncertain how much it contributes to SOA formation (Karl et
99 al., 2009;Carlton et al., 2009).

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100 Modelling, laboratory chamber experiments and field studies provide a range of
101 possible yields of SOA from isoprene, typically of the order 0.1–3 % by mass, with
102 some values reported as high as 5.5 % (van Donkelaar et al., 2007;Kleindienst et al.,
103 2009, 2007;Kroll et al., 2005, 2006;Claeys et al., 2004a;Edney et al., 2005;
104 Brégonzio-Rozier et al., 2014). SOA yields from the further oxidation of first and
105 subsequent generation isoprene oxidation products, such as methacrolein, are
106 estimated to be as much as 15 % (Rollins et al., 2009;Carlton et al., 2009;Claeys et
107 al., 2004b;Robinson et al., 2010). Recent work has highlighted that under low NO_x
108 conditions, SOA mass formed from isoprene oxidation could be influenced by the
109 acidity of pre-existing aerosol via the reactive uptake of certain key isoprene
110 oxidation products, namely isoprene epoxydiols (IEPOX; Surratt et al., 2010; Lin et
111 al., 2012). More recently, Nguyen et al. (2014) found that the “pH dependence for OA
112 formation from IEPOX was weak for AS particles”. There is further evidence from
113 chamber studies using temperate tree species such as birch, spruce and pine that

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114 isoprene may in fact suppress SOA formation from other VOC precursors, when
115 present when present (Kiendler-Scharr et al., 2009a; Kanawade et al., 2011). It should
116 be noted at this point that it is unclear in most cases how wall effects have been
117 considered in the production of such yield values and whether the treatments
118 employed are adequate such that the yields are comparable between chambers, or
119 indeed between experiments.

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120 Here, we characterised the BVOC emissions from three south-east Asian tropical
121 plant species (*Ficus cyathistipula*, *Ficus benjamina* and *Caryota millis*) and in a series
122 of coupled plant growth chamber–atmospheric reaction chamber experiments, we
123 examined the ability of their oxidation products to contribute to SOA formation under
124 atmospherically relevant conditions. In order to provide a geographically and
125 chemically contrasting study, we replicated these experiments using common silver
126 birch (*Betula pendula*). Silver birch has previously been shown to contribute to the
127 formation of secondary organic aerosol via the emissions of mono- and sesquiterpenes
128 (e.g. Kiendler-Scharr et al., 2009a; 2009b; Mentel et al., 2009). Seeded (ammonium
129 sulphate) and un-seeded experiments were carried out to allow studies of both fresh
130 nucleation and condensation onto pre-existing aerosol.

131 **Methods and materials**

132 ***2.1 Plant selection and pre-screening***

133 Three non-clonal specimens of common silver birch (*Betula pendula*), a monoterpene
134 and isoprene emitting tree species; two species of fig (*Ficus benjamina* and *Ficus*
135 *cyathistipula*), and one species of palm (*Caryota millis*), each approximately 1.5 m in
136 height were used. Figs and palms are abundant in all tropical rainforests. We chose
137 three species found in abundance throughout south and southeast Asia to be consistent
138 with our field work (Hewitt et al., 2010a; MacKenzie et al., 2011). *Ficus benjamina*
139 (Moraceae) is native to Malaysia and has previously been found to be a high isoprene
140 emitter (0.03 – 8.7 $\mu\text{g C g}^{-1} \text{h}^{-1}$, potted and in soil) with emissions of the
141 monoterpenes, limonene (0.02 $\mu\text{g C g}^{-1} \text{h}^{-1}$) and β -ocimene (1.8 – 2.5 $\mu\text{g C g}^{-1} \text{h}^{-1}$),
142 and the sesquiterpenes β -caryophyllene and α -copaene (Carvalho et al., 2005; Geron et
143 al., 2006). In addition, emissions of benzaldehyde (0.53 $\mu\text{g C g}^{-1} \text{h}^{-1}$) and acetaldehyde
144 (69 $\mu\text{g C g}^{-1} \text{h}^{-1}$) from potted specimens have been detected (Carvalho et al., 2005).
145 No previous data are available on the BVOC emissions from *Ficus cyathistipula* or
146 *Caryota millis*. Proton transfer-reaction mass spectrometry (PTR-MS) and gas
147 chromatography–mass spectrometry (GC-MS) screening, prior to the start of the
148 coupled chamber experiments, confirmed that both species were high isoprene
149 emitters with *Ficus cyathistipula* also emitting limonene, β -phellandrene, α -
150 damascone and acetaldehyde. Analytical methods are described in detail in section
151 2.4.

152 **2.2 Plant chamber design**

153 A 4.7 m³ plant chamber was constructed out of two rectangular Teflon bag sections
154 and a Teflon lid (0.05 mm FEP) (Adtech Polymer Engineering, UK), which were each
155 supported by frames built using 25 mm² box aluminium (Speed Frame, RS
156 Components, UK). The framework stood on a raised foil and Teflon covered marine
157 plywood base. PVC foam strips (RS Components, UK) ensured an airtight seal
158 between chamber sections. Heavy-duty double-sided tape (RS components, UK) was
159 used to secure the Teflon bags to the frame. The interior of the plant chamber was
160 only exposed to Teflon surfaces.

161 Compressed air was constantly supplied to the plant chamber via a mass flow
162 controller and regulator (ALICAT MCR-500 SLPM-D, Premier Control Technologies
163 Ltd, UK) at 780 L min⁻¹ and 7.5 bar via a 12.7 mm (outer diameter- OD) reinforced
164 tube. This was reduced to approx 1 bar and between 250 to 300 L min⁻¹ (+/- 0.8 %)
165 dependent on the photosynthetic and transpiration rate of each plant species
166 (equivalent to one complete air change every 15 - 20 mins). The air stream was
167 passed through a 12.7 mm (OD) PTFE tube to three in-series filters to remove any
168 pre-existing VOCs (activated carbon filter P3KFA14ASMN, Parker Pneumatic, UK),
169 and submicrometer particles (HEPA CAP 75 filter capsule (FDP-780-050K, Fisher
170 Scientific, UK)), and NO_x (Purafil and activated charcoal, Purafil Inc. USA). Finally,
171 the air was re-humidified by passing it through a 2 L Teflon barrel (Jencons, UK)
172 filled with warmed distilled water. The plant chamber outlet air was either vented into
173 the laboratory via a 50 mm (OD) stainless steel pipe and valve, or used to fill an 18
174 m³ Teflon reaction chamber.

175 To enhance mixing, air entered the plant chamber via a perforated 12.7 mm (OD)
176 PTFE tube that circled the base of the chamber. One 12.7 mm stainless steel bulkhead
177 fitting (Swagelok, UK) was inserted through the frame to secure the PTFE tube to the
178 base of the plant chamber. A 50 mm (OD) stainless steel pipe was inserted into the
179 upper corner of the chamber and supported by a Teflon (inner surface) and Nylon
180 (outer surface) manifold (Plastics Direct, UK). The manifold also supported an EGM
181 probe (EGM-4, PP Systems, UK), which recorded relative humidity (RH),
182 temperature (T), CO₂ and photosynthetically active radiation (PAR).

183 Plants were kept in 255 – 330 mm (height) pots depending on species, watered to pot
184 dripping point and sprayed twice weekly. Plant chamber conditions were maintained
185 at 31 – 33.5 °C / 22 – 24 °C (day/night), 29 – 40 % / 33 – 44 % (day/night) RH, and
186 335 – 385 ppmV / 390 – 404 ppmV (day/night) CO₂. Owing to structural restrictions,
187 PAR could not be measured directly under the growth lamps in the centre of the
188 canopy. At the top edge of the canopy it was 500 μmol m⁻² s⁻¹ with a 12 hr day / night
189 cycle.

190 ***2.3 Reaction chamber description***

191 The aerosol photochemical reaction chamber at the University of Manchester is
192 composed of an 18 m³ FEP Teflon bag mounted on three rectangular extruded
193 aluminium frames (Alfarra et al., 2012). A bank of halogen lamps and a 6 kW Xenon
194 arc lamp are mounted on the enclosure housing the bag, which is coated with
195 reflective “space blanket” providing an integrating sphere, maximising the irradiance

196 in the bag and ensuring even illumination for the production of photochemical species
197 such as the hydroxyl radical (OH). The air introduced to the bag is dried and filtered
198 for gaseous impurities and particles, prior to humidification with high purity
199 deionised water. A high capacity O₃ generator provides controlled ambient levels of
200 O₃ (used as an oxidant) and high O₃ concentrations (serving as a cleaning agent
201 between experiments).

202 Size-dependent (diffusional and gravitational) wall-loss rate constants were calculated
203 based on particle mobility and the surface-to-volume ratio of the chamber (Verheggen
204 and Mozurkewich, 2006). The diffusional loss rate uses a constant of proportionality,
205 which can only be determined empirically. A time period was selected near the end of
206 each experiment where the wall losses were deemed to be the dominant process
207 affecting the size distribution. The volume size distribution at the beginning of this
208 period had the calculated wall loss rate applied to simulate the evolution of the size
209 distribution over the selected time period. If the calculated loss rate loss rate didn't
210 reproduce the measured volume evolution within the specified tolerance (1 – 2 % in
211 this work), the constant of proportionality for diffusional losses was adjusted such that
212 the simulated volume at the end of the selected period matched the measured volume
213 within the specified tolerance. The time-integrated gravitational and (optimised)
214 diffusional loss rate constants were then applied to the volume size distribution
215 throughout the experiment in order to reconstruct a wall loss corrected size
216 | distribution, which was then used to calculate the wall-loss-corrected particle mass.

217 Both the plant chamber and reaction chamber were tested for contaminants separately
218 and when joined together by running the system with an empty plant chamber and by
219 carrying out a “blank” run prior to each set of experiments.

220 ***2.4 Analytical techniques***

221 *2.4.1 Gas phase measurements*

222 The volatile and semi-volatile organic compounds and oxygenated volatile organic
223 compounds in both the plant chamber and the reaction chamber were measured by
224 soft-ionisation mass spectrometry (PTR-MS, CIR-TOF-MS, described below) and gas
225 chromatography-mass spectrometry (GC-MS).

226 The proton-transfer-reaction mass spectrometry (PTR-MS) instrument employed
227 (Ionicon, Austria) comprises two turbomolecular pumps, a heated silica steel inlet
228 system and a 9.6 cm long stainless steel drift tube. The nominal response time is
229 approximately 1 s. The operating parameters of the PTR-MS were held constant
230 during measurements, except for the secondary electron multiplier voltage, which was
231 optimised each day. The drift tube pressure, temperature and voltage were 2.2 hPa, 50
232 °C, and 600 V, respectively. The central reaction chamber of the drift cell was
233 operated at an E/N (*i.e.* electric field/gas number density) of 125 Td. The count rate of
234 $\text{H}_3\text{O}^+\cdot\text{H}_2\text{O}$ ions was 1 – 2 % of the count rate of H_3O^+ ions. The PTR-MS sampled
235 continuously with a flow rate of 100 – 150 ml min⁻¹ through 3.2 mm PTFE tubing.

236 The chemical-ionisation-reaction time-of-flight mass spectrometer (CIR-TOF-MS)
237 comprises a temperature controlled ($40 (\pm 1) ^\circ\text{C}$) ion-source drift cell assembly
238 coupled to an orthogonal time-of-flight mass spectrometer equipped with a reflectron
239 array (Kore Technology Ltd, Ely, UK). The ion-source deployed was a hollow
240 cathode discharge type (Blake et al., 2009) and the chemical ionization technique
241 used was proton transfer reaction from hydrated hydronium ($\text{H}_3\text{O}^+\cdot\text{H}_2\text{O}$) (Jenkin et
242 al., 2012). Sample air was delivered in a continuous stream directly to the drift cell
243 via a 0.5 m long, 6.35 mm (internal diameter) Teflon sample line, heated to $40 (\pm 1)$
244 $^\circ\text{C}$, at a constant flow rate of 80 ml min^{-1} . The central reaction chamber of the drift
245 cell was operated at an E/N ratio of $\sim 90 - 100 \text{ Td}$, with a tuned energy ramp at the
246 base of the cell to remove potential water-cluster ions (e.g. $\text{RH}^+\cdot\text{H}_2\text{O}$). Further
247 information regarding the CIR-TOF-MS design and a detailed discussion regarding its
248 operation can be found in Blake et al. (2004) and Wyche et al. (2007).

249 The PTR-MS and CIR-TOF-MS were calibrated using three different methods: (i)
250 step-wise dilution of a gravimetrically prepared gas standard (BOC Special Gases,
251 UK) containing a variety of VOCs and OVOCs; (ii) using calibration material
252 produced in-house via the injection of liquid samples into 10 l Tedlar bags (SKC Inc.,
253 USA) containing either humidified or dry, pure nitrogen; and (iii) using gas standards
254 derived from permeation tubes (Vici Inc., US; Ecoscientific, UK), diluted, humidified
255 and delivered by a commercial calibration unit (Kintec, model: 491). Where
256 experimental calibration was not possible for a specific compound, either the
257 calibration sensitivity for a structurally similar surrogate was used or calculated

258 concentrations were employed (Jenkin et al., 2012). For the quantification of isobaric
259 signals, a single sensitivity value was used, e.g. α -pinene sensitivity for
260 Σ (monoterpenes) and β -caryophyllene sensitivity for Σ (sesquiterpenes); again
261 working on the principal that structurally similar compounds possess similar PTR and
262 CIR sensitivities.

263 CIR-TOF-MS and PTR-MS detection limits are reagent, reaction, analyte and sample
264 matrix specific. However, typical CIR-TOF-MS detection limits, using PTR
265 ionisation from hydronium, are of the order $0.4 \text{ ppbV} (10 \text{ min})^{-1}$ for more polar
266 compounds, such as OVOCs (e.g. 2-hexanone) and as much as 10 ppbV min^{-1} for
267 certain less polar compounds, such as smaller hydrocarbons (e.g. 1-pentene). For
268 further details see Wyche et al., 2007.

269 The GC-MS system (GC-MS Turbomass Gold, Perkin Elmer, USA) comprised a
270 thermal desorption autosampler (Perkin-Elmer ATD 400) connected via a heated (200
271 $^{\circ}\text{C}$) transfer line to a Hewlett-Packard 5890 GC with a 5970 mass-selective detector.
272 Compounds were desorbed at $280 \text{ }^{\circ}\text{C}$ for 5 min at 25 mL min^{-1} onto a Tenax-TA cold
273 trap maintained at $-30 \text{ }^{\circ}\text{C}$. The cold trap was then heated to $300 \text{ }^{\circ}\text{C}$ for 6 min to
274 desorb compounds onto the GC column. Chromatographic separation was achieved
275 using an Ultra-2 column (Agilent Technologies: $50 \text{ m} \times 0.2 \text{ mm ID} \times 0.11 \text{ } \mu\text{m}$ film, 5
276 % phenylmethyl silica). An initial oven temperature of $35 \text{ }^{\circ}\text{C}$ was maintained for 2
277 min, and then increased at $4 \text{ }^{\circ}\text{C min}^{-1}$ to $160 \text{ }^{\circ}\text{C}$ followed by an increase of $45 \text{ }^{\circ}\text{C}$
278 min^{-1} to $300 \text{ }^{\circ}\text{C}$, which was maintained for 10 min. The carrier gas was Helium

279 supplied at a rate of 1 mL min^{-1} , with an injector temperature of $250 \text{ }^\circ\text{C}$. The limit of
280 detection for isoprene and monoterpenes was approximately $0.25 \text{ ng on column}$ and 2
281 ng on column for sesquiterpenes, corresponding to 100 pptV of isoprene, 50 pptV of
282 monoterpenes, and to 400 pptV of sesquiterpenes in a 1 L sample. Sampling was
283 conducted by drawing 8 L of the analyte air through 6.35 mm PTFE tubing onto the
284 GC-MS sample tubes using a handheld pocket pump (SKC Ltd, UK) at a flow rate of
285 150 ml min^{-1} (total sample time $\sim 43 \text{ mins}$). Sample tubes were stored at $4 \text{ }^\circ\text{C}$ until
286 analysed. VOC quantification was by comparison with commercially available liquid
287 standards (Aldrich, Fluka and Sigma) diluted in methanol. Isoprene quantification
288 was by comparison with a 700 ppbV in N_2 certified gas standard (BOC, UK).

289 NO and NO_2 mixing ratios were measured using a chemiluminescence gas analyser
290 (Model 42i, Thermo Scientific, MA, USA). Ozone was measured using a UV
291 photometric gas detector (Model 49C, Thermo Scientific, MA, USA).

292 *2.4.2 Particle phase measurements*

293 Within the main reaction chamber, a scanning mobility particle sizer (SMPS) system
294 was used to measure the particle size distribution and total aerosol mass concentration
295 (without sample drying). A particle density of 1.3 g cm^{-3} was assumed for calculating
296 the mass of SOA particles in un-seeded experiments (Alfarra et al., 2006; Bahreini et
297 al., 2005). For seeded experiments, a density of 1.77 g cm^{-3} was used to calculate the
298 ammonium sulphate seed mass and 1.3 g cm^{-3} was assumed for calculating the
299 additional SOA mass. A water-based condensation particle counter (wCPC, TSI

300 3786) was used to count the total particle number concentration between 2.5 nm and
301 approximately 3 μm . Further instrument details can be found in Alfarra et al., 2012,
302 and references therein.

303 Real-time broad chemical characterisation of the SOA was made using a compact
304 Time-of-Flight Aerosol Mass Spectrometer (cToF-AMS, Aerodyne Research Inc.,
305 USA). A detailed description of the instrument, its operation and calibrations can be
306 found elsewhere (Drewnick et al., 2005;Canagaratna et al., 2007). The instrument was
307 operated in the standard configuration, taking both mass spectrum (MS) and particle-
308 time-of-flight (PToF) data and was calibrated for ionisation efficiency using 350 nm
309 monodisperse ammonium nitrate particles. The vapouriser was set at approximately
310 600 °C and data were collected at a time resolution of 2 min. A collection efficiency
311 value of unity was applied to these data, based on evidence from a previous chamber
312 study (Alfarra et al., 2006).

313 | Filter samples for offline analysis were collected (without denuders) in a specially
314 | constructed holder, positioned in the chamber vent line. Aerosol samples were
315 | collected onto 47 mm quartz fibre filters (Whatman) at a rapid flow rate of 3 m³ min⁻¹
316 | (sample time ca. 6 mins.). After sampling, filters were immediately placed in pre-
317 | cleaned glass vials and stored below -20 °C until analysis. The filter collection
318 | procedure employed here is much faster than traditional filter collection methods,
319 | which should minimise any potential negative and positive artefacts.

320 The filters were extracted into high purity water, filtered, evaporated to dryness and
321 redissolved in 1 ml 50 % MeOH : 50 % H₂O. The water-soluble compounds were
322 analysed using liquid chromatography-ion trap mass spectrometry (LC-MS/MS).
323 Reverse phase LC separation was achieved using an HP 1100 LC system equipped
324 with an Eclipse ODS-C₁₈ column with 5 µm particle size (Agilent, 4.6 mm x
325 150 mm). Samples (60 µl) were injected then eluted by gradient elution with solvents
326 A: 0.1% v/v formic acid water (Optima grade, Fisher) and B: methanol (Optima
327 grade, Fisher) and a gradient program of 3 % B at time 0 min to 100 % B at 60
328 minutes with a flow rate of 0.6 ml min⁻¹. Mass spectrometry analysis was performed
329 in negative ionisation mode using an HCT-Plus ion trap mass spectrometer with
330 electrospray ionisation (Bruker Daltonics GmbH). Electrospray ionisation (ESI) was
331 carried out at 350 °C with a nebuliser pressure of 4.82 bar and a nitrogen drying gas
332 flow of 12 l min⁻¹. Further details can be found in Hamilton et al., 2013.

333 *2.5 Experimental protocol*

334 Three plants were placed in the plant chamber a minimum of 48 hours prior to the
335 start of the experiment. Both the pots and soil were isolated by enclosing them in
336 PFTE sheeting; this acted to prevent VOC emissions from the plastic pots and soil
337 NO_x emissions from entering the chamber air. Three experiments were carried out on
338 each species over a one-week period, after which the plants were removed and
339 replaced with three plants of the next species, and the experiment cycle repeated.

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340 Prior to each experiment, ozone was added to the chamber to give a mixing ratio of
341 approximately 2 ppmV and was left overnight. The chamber was then filled and
342 flushed several times using clean air from the facility's main inlet system (including
343 Purafil, charcoal and HEPA filters as described above), until the total particle count
344 (as measured by a water based condensation particle counter) was below 10 cm^{-3} and
345 the O_3 and NO_x levels were less than 1 and 2 ppbV, respectively. At this point, the
346 reaction chamber was flushed and then connected to the plant chamber for filling with
347 the plant VOC emissions. Aerosol and gas phase composition and concentrations
348 were continuously monitored throughout. At the end of the filling process, the plant
349 chamber was disconnected from the reaction chamber, and within the space of
350 roughly one minute, both the chamber lights were turned on and pure O_3 was injected
351 to provide an initial concentration of around 20 or 70 ppbV (experiment dependent).
352 The switching on of the chamber lights marked the start of each experiment, which
353 typically lasted 6 hours from this point. For experiments using pre-existing seed,
354 polydisperse ammonium sulphate particles (diameter between 40 – 60 nm) were
355 generated from an aqueous solution using an aerosol nebuliser (Topas, ATM 230) and
356 injected without drying into the reaction chamber at the end of the filling from the
357 plant chamber.

358 In our experiments we chose to use ammonium sulphate for the aerosol seeds, rather
359 than acidic particles that could otherwise promote isoprenoid particulate mass
360 formation. Whilst it is recognised that isoprenoid SOA mass can be enhanced by the
361 presence of acidic aerosol seed as originally reported by Jang et al. (2002) and

362 subsequently by Limbeck et al. (2003), Edney et al. (2005), Kleindienst et al. (2007),
363 Limbeck et al. (2007) and Surratt et al. (2007), we have limited our study to SOA
364 formation in the mixed precursor systems without deliberate enhancement of particle
365 mass by condensed phase reaction. There is clear evidence that isoprene oxidation can
366 contribute to atmospheric SOA formation (e.g. Claeys et al., 2004, Edney et al., 2005)
367 and we have previously found that enhancement in SOA from isoprene oxidation
368 above the Bornean rainforest compared with the Amazon may result from an
369 enhanced marine acidic sulphate contribution to sub-micron aerosol (Robinson et al.,
370 2011). Intermediates in SOA formation from isoprene have been identified (e.g. Lin
371 et al., 2012, 2013) and mechanisms for the acid catalysed formation proposed (Surratt
372 et al., 2010). Whilst out of the scope of the current study, this should be the focus of
373 future work.

374 Air samples were taken from three separate locations: 1) immediately before the plant
375 chamber (pre-PC) for blank subtraction, 2) immediately after the plant chamber (post-
376 PC) during the reaction chamber filling period for directly emitted BVOC and 3) from
377 the reaction chamber (RC) during the experiment. RC air was monitored continuously
378 using PTR-MS and CIR-TOF-MS for VOC decay and formation of reaction products.
379 Air samples from the pre-PC and post-PC position, as well as RC air samples
380 immediately at the start of each experiment and 1, 2, 4 and 6 hours after the lights
381 were switched on, were collected on Tenax TA and Carbotrap filled stainless steel
382 tubes (Supelco Inc, PA, USA) for GC-MS analysis.

383 Relative humidity (%), CO₂ (ppmV), PAR (μmol m⁻² s⁻¹), and temperature (°C) in the
384 plant chamber were recorded every 5 or 10 mins during reaction chamber filling, and
385 every 15 or 20 mins overnight. System blanks were taken at the start and end of the
386 experimental period. The reaction chamber background was checked and
387 characterised through the performance of regular blank experiments (one in every five
388 experiments). NO_x (NO, NO₂ and NO₃) and O₃ were continuously monitored in the
389 reaction chamber. A list of all experiments and their general parameters is given in
390 Table 1.

391 **2.6 Calculated OH concentrations**

392 Since isoprene losses are controlled by reaction with ozone and the hydroxyl radical
393 (OH), the concentration of OH available to react with isoprene in the reaction
394 chamber for each experiment was calculated based on the measured concentrations of
395 O₃ and isoprene in each experiment, the rate of change in isoprene concentration, and
396 the rate constants for the reactions of isoprene with OH and O₃, using equation (1):

$$397 \quad \frac{\frac{d[Isoprene]}{dt} + k_{O_3}[O_3][Isoprene]}{-k_{OH}[Isoprene]} = [OH] \quad \text{Eq}^n (1)$$

398 Hourly averaged concentrations of O₃ and isoprene were calculated for five of the
399 experiments using the tropical fig. Using these data along with equation (1) a range of
400 OH concentrations were obtained. For the first hour after lights on, [OH] was
401 estimated to be 1.9 x 10⁵ – 9.5 x 10⁵ molecules cm⁻³, whereas, towards the end of the

402 experiment after roughly five hours, values of $8.1 \times 10^5 - 1.9 \times 10^6$ molecules cm^{-3}
403 were obtained. In general, during the tropical fig experiments, [OH] estimated from
404 isoprene and ozone was observed to steadily increase over the duration of the
405 experiment from 0 – 5 hours after lights on.

406 **2.7 VOC/NO_x Conditions**

407 Figure 1 shows the time-dependent mixing ratios of ozone and oxides of nitrogen for
408 each experiment set. Although every effort was made to keep the concentrations of
409 oxides of nitrogen low, measurable amounts were present, giving initial VOC/NO_x
410 ratios of the order 2 – 6 and 3 – 9 (see Table 1), for the birch and fig experiments,
411 respectively (where here, the VOC concentration is equal to the sum of all potential
412 precursor concentrations). In terms of a “Sillman plot” (Sillman, 1999), the
413 experiments were carried out in the “VOC sensitive regime”.

414 The absolute concentration of VOCs in the reaction chamber was roughly ten times
415 greater than those measured over the rainforest during our field experiments
416 (Mackenzie et al., 2011) and the VOC/NO_x ratios employed here were as much as ten
417 times lower (i.e. typical ratio of 20:1, isoprene:NO_x over the rainforest) (Hewitt et al.,
418 2010b). The source of the NO_x in the reaction chamber (initially ~ 2 – 6 ppbV NO_x,
419 but increasing to ~ 5 – 9 ppbV after ~ 5 hours) is attributed to a small amount of
420 diffusion of outside ambient air across the porous Teflon membrane into the reaction
421 chamber. The production of certain reactive intermediates in the oxidation of VOCs

422 (e.g., hydroxyl hydroperoxides from isoprene oxidation) is very sensitive to NO_x
423 concentrations in the reaction mixture.

424 **Results**

425 ***3.1 Experiments with *Betula pendula****

426 *3.1.1 Gas phase*

427 Continuous gas phase monitoring with the CIR-TOF-MS and PTR-MS throughout the
428 experiments, indicated successful transfer of VOC precursor material from the plant
429 chamber to the reaction chamber prior to lights on. The data indicated that there was
430 negligible loss of precursor compounds during the chamber transfer process (Fig. 2a).

431 Immediately after initiation of the photochemistry, the VOC precursor concentrations
432 were observed to decay and product ions began to appear in the CIR and PTR mass
433 spectra. Approximately sixty product-ion peaks were observed by the CIR-TOF-MS
434 and the PTR-MS in the organic gas phase during a typical *Betula pendula* experiment.
435 The temporal profiles of a number of the most abundant (O)VOCs measured are
436 shown in Fig. 2. From a combination of the CIR-TOF-MS, PTR-MS and GC-MS
437 observations (and from those observations discussed below for the tropical plant
438 experiments), over fifty different hemi-, mono- and sesqui-terpene oxidation products
439 were tentatively identified (Fig. 3 and Tables S1 – S5 in the supplementary
440 information).

441 From initial inspection of the data, it is clear that monoterpenes dominate during the
442 *Betula pendula* experiments (Fig. 2a and b), with strong signals observed in the CIR-
443 TOF-MS and PTR-MS mass spectra at m/z 137 (protonated parent ion) and 81
444 (hydrocarbon fragment). A small amount of isoprene was also detected during *Betula*
445 *pendula* experiments; however this was always significantly lower in magnitude than
446 that of the sum of monoterpenes; for example, during the experiment on 07/07/09,
447 12.6 (\pm 3.8) ppbV monoterpenes were measured in the reaction chamber prior to
448 lights on (c.f. 11.4 ppbV total monoterpenes measured at the post-PC position by the
449 GC-MS), whereas only 2.0 (\pm 1.0) ppbV isoprene was detected. Speciation of the
450 monoterpenes by GC-MS indicated that the most dominantly emitted C_{10} compounds
451 from *Betula pendula* were α - and β -pinene (Table 2).

452 C_{15} sesquiterpenes (parent ion m/z 205) were detected in the plant and reaction
453 chambers during each *Betula pendula* experiment, with the most abundant species
454 identified by GC-MS being β -caryophyllene (Table 2). Sesquiterpenes were also
455 measured in the reaction chamber by CIR-TOF-MS (Fig. 2c), however for the
456 majority of the experiments they were present at concentrations either close to or
457 below the detection limit, hence they could not always be monitored as a function of
458 reaction time. For the experiment on 07/07/09, 1.7 (\pm 0.9) ppbV sesquiterpenes were
459 measured by the CIR-TOF-MS prior to lights on (c.f. 2.2 ppbV total sesquiterpenes
460 measured at the post-PC position by the GC-MS). An ion of m/z 153 was also
461 observed in the PTR and CIR mass spectra of the plant chamber emissions and

462 subsequently in the reaction chamber air, tentatively assigned (and here after referred
463 to) as camphore.

464 During the reaction phase of the *Betula pendula* experiments the CIR-TOF-MS mass
465 spectra were dominated by ions of relatively high mass (i.e. $m/z > 100$) pertaining to
466 products of both monoterpene and sesquiterpene oxidation. The ions of highest mass
467 (i.e. m/z 170 – 290) are characteristic of sesquiterpene oxidation, and have been
468 observed recently during a similar chamber study investigating β -caryophyllene
469 photo-oxidation (Jenkin et al., 2012). Drawing a comparison between these data and
470 the detailed β -caryophyllene study conducted by Jenkin et al. (2012), a number of
471 tentative assignments have been made for β -caryophyllene products, with the
472 assumption that other precursor specific structural isomers may also occupy the same
473 mass channels. A full list of example tentative assignments is given in the
474 supplementary information (Table S2). In total the sum of all sesquiterpene products
475 measured in the chamber was estimated to be ~ 1.5 ppbV (assuming an average PTR
476 sensitivity for such high mass, oxygenated, compounds).

477 In contrast to the small amounts of sesquiterpene products observed in the reaction
478 chamber, the products observed in greatest abundance were those derived from
479 monoterpene decay. The largest (combined) product signal measured by the CIR-
480 TOF-MS was that of $\Sigma(I_{111}, I_{93})$, where I_x is the intensity of the mass spectrum at m/z
481 = x (Fig. 2d). Previously, the m/z 111 and 93 signals have been shown to correspond
482 to various primary C_7 unsaturated aldehydes formed during the oxidation of

483 unsaturated acyclic monoterpenes, such as myrcene, ocimene and linalool (Lee et al.,
484 2006a; Lee et al., 2006b; Ng et al., 2006; Wyche et al., In Preparation). In the case of
485 myrcene and ocimene, the m/z 111 and 93 signals correspond to the parent ion (MH^+)
486 and the dehydrated daughter fragment, respectively ($MH^+ - H_2O$), and in the case of
487 linalool m/z 111 corresponds to the dehydrated daughter ion and m/z 93 is a further
488 fragment. The concomitant m/z 111 and 93 signals have also been reported to result
489 from a C_7 cyclic ketone formed during the oxidation of terpinolene (not found in the
490 *Ficus* emission profile and < 1 ppbV found in the *Betula* profile). The m/z 111 and 93
491 ions have previously been observed to be significant contributors to total ion signal in
492 the PTR mass spectra during single precursor chamber experiments with concomitant
493 SOA formation (Lee et al., 2006a; Lee et al., 2006b; Ng et al., 2006; Wyche et al., In
494 Preparation), and the m/z 111 ion has also been observed in ambient air measurements
495 over a forested region (Holtzinger et al., 2005).

496 As can be seen from observation of Fig. 2(d), the $\Sigma(I_{111}, I_{93})$ signal rises rapidly
497 during the initial stages of the experiment, much more so than other monoterpene
498 oxidation products (c.f. Fig. 2e), suggesting that the precursor has a much shorter
499 lifetime with respect to OH and O_3 . Of those monoterpenes speciated by the GC-MS,
500 ocimene and linalool have the shortest lifetimes, with $k_{OH} = 3.04$ and $1.6 \times 10^{-10} \text{ cm}^3$
501 $\text{molecule}^{-1} \text{ s}^{-1}$ (average lifetimes with respect to OH ~ 44 and 55 mins.), respectively,
502 compared to $k_{OH} = 7.4 \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$ for β -pinene (average lifetime with
503 respect to OH ~ 1458 mins.) (Atkinson and Arey, 2003; Kim et al., 2011). The $\Sigma(I_{111},$

504 I_{93}) signal peaks at around 60 – 100 mins at 3.0 (\pm 0.7) ppbV (concentration estimated
505 using pinonaldehyde sensitivity), before decaying at a greater rate than that of the
506 precursor monoterpenes and the other monoterpene products. This relatively short
507 lifetime gives further insight into the potential identity of the m/z 111 and 93 signals,
508 perhaps indicating the presence of multiple C=C bonds in the hydrocarbon structure,
509 as would be found in the primary C₇ aldehydes obtained from the oxidation of
510 ocimene or myrcene for example. Other short-lived biogenic oxidation products that
511 could exist in such mesocosm systems, include α -hydroxy carbonyls, similarly formed
512 following OH addition to a C=C bond.

513 Other dominant signals observed by the PTR-MS and CIR-TOF-MS during oxidation
514 of the *Betula pendula* air matrix, include the sum of m/z 169 + 151 + 107, which
515 respectively correspond to the parent ion and two daughter fragments of a number of
516 primary monoterpene keto-aldehydes (which, from the speciated monoterpene plant
517 chamber data, are most likely to be pinonaldehyde, caronaldehyde and α/γ -
518 terpinaldehyde); and m/z 139, corresponding to the parent ion of a number of primary
519 monoterpene ketones (most likely to be nopinone and caronone, again when
520 considering the monoterpenes speciated by the GC-MS). As shown in Fig. 2(e) the
521 primary keto-aldehyde and ketone signals had similar temporal profiles to one
522 another, growing at a slower rate than that of $\Sigma(I_{111}, I_{93})$, to peak concentrations of
523 around 0.9 (\pm 0.3) and 1.2 (\pm 0.3) ppbV, respectively, as the monoterpene trace
524 tended towards zero. The temporal profile for the sum of all other “monoterpene
525 like” product ions (i.e. ions of $m/z > 90$) was very similar to those of the primary keto-

526 aldehyde(s) and ketone(s), peaking at a combined mixing ratio of approximately 3.5
527 ppbV (assuming an average PTR sensitivity for such high mass, oxygenated,
528 compounds).

529 During the oxidation of compounds emitted by *Betula pendula*, the primary isoprene
530 products, methyl vinyl ketone (MVK) and methacrolein (MACR) (measured together
531 at m/z 71) were observed to evolve in the same manner as the primary monoterpene
532 keto-aldehyde(s) and ketone(s), peaking at an approximate mixing ratio of 0.4 (\pm 0.1)
533 ppbV (Fig. 2e). A series of lower m/z ions were also observed to evolve within the
534 reaction chamber, including m/z 61 (acetic acid), 59 (acetone), 47 (formic acid), 45
535 (acetaldehyde), 33 (methanol) and 31 (formaldehyde). Each of these compounds has
536 previously been associated with monoterpene oxidation and/or with off-gassing from
537 illuminated chamber walls. Methanol, acetone and m/z 99 (potentially cis-3-hexenal)
538 were also observed within the reaction chamber prior to lights on, with a combined
539 mixing ratio of approximately 20 ppbV.

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540 3.1.2 Particle phase

541 From inspection of the CPC and SMPS data we see that SOA mass formed during
542 oxidation of the *Betula pendula* air matrix. As can be seen from Fig. 4, during
543 unseeded experiments nucleation occurred immediately after lights on, with no
544 induction period prior to mass formation. After nucleation, SOA mass increased
545 rapidly to $\sim 11 \mu\text{g m}^{-3}$ by ~ 40 minutes (experiment 06/07/09), followed by a relatively
546 stable plateau (after the application of wall loss corrections) and a slight increase

547 | towards the end of the experiment. In order to suppress nucleation, seed particles were
548 | introduced in some experiments, as has been used previously (Dommen et al.,
549 | 2009; Meyer et al., 2009; Surratt et al., 2007; Kleindienst et al., 2006; Carlton et al.,
550 | 2009). This more closely represents the conditions encountered in the ambient
551 | atmosphere where there is pre-existing aerosol. Consistent with the nucleation
552 | experiments described above, SOA mass was observed to increase as soon as the
553 | photochemistry was initiated when an ammonium sulphate seed was present (Fig. 4,
554 | experiment 07/07/09).

555 | Using the wall-loss-corrected mass data, along with the corresponding quantity of the
556 | sum of precursor species reacted and equation (2), SOA yields were obtained for the
557 | *Betula pendula* oxidation system:

558 |
$$Y_{SOA} = \frac{M_p}{\Delta(\Sigma VOC)} \quad \text{Eq}^n(2)$$

559 | In this instance, Y_{SOA} = SOA mass yield, M_p = peak SOA mass ($\mu\text{g m}^{-3}$) and $\Delta(\Sigma VOC)$
560 | = the sum of gas phase precursors reacted by the time M_p is reached ($\mu\text{g m}^{-3}$) (Odum
561 | et al., 1997). In order to determine $\Delta(\Sigma VOC)$, the time-dependent VOC mixing ratios
562 | for total sesquiterpenes, total monoterpenes, camphore and isoprene were
563 | independently converted to their corresponding mass concentrations ($\mu\text{g m}^{-3}$) and the
564 | four data sets were combined to give a “total” VOC precursor decay profile. From
565 | the total VOC profile, $\Delta(\Sigma VOC)$ was calculated, using the starting mass of ΣVOC at
566 | time = 0 and the mass of ΣVOC at the time of M_p . The uncertainty in $\Delta(\Sigma VOC)$ is

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567 estimated to be ± 41 %. Using equation (2) for the two *Betula pendula* experiments
568 for which both gas phase mixing ratio and wall-loss-corrected aerosol data were
569 available, SOA yield values of 39 and 26 % were obtained (Fig. 4). It should also be
570 noted that along with previous caveats made regarding the role of the chamber walls
571 and other measurement uncertainties, these yield values also do not take into account
572 the potential loss of particularly “sticky” low volatility compounds (e.g. Ehn et al.,
573 2014) to internal surfaces of the chamber.

574 It should be noted in the above yield calculations that the partitioning of material
575 between the vapour phase and chamber walls has not been taken into account.
576 Matsunaga and Ziemann (2010) showed that semi-volatile organic compounds move
577 towards equilibrium between the walls and the vapour phase and that the equilibration
578 timescale and equivalent absorptive mass of the walls was dependent on the
579 molecular properties of the partitioning species. Kokkola et al. (2013) demonstrated in
580 their model study that OVOC wall losses will have significant implications on their
581 partitioning between the gas and particle phase, such that the mass components of
582 very low volatility will almost completely be depleted to the chamber walls during the
583 experiment while the depletion of OVOCs of higher volatilities is less efficient. The
584 implications of such partitioning to chamber walls are such that comparison between
585 any yields determined experimentally in different chambers should be conducted with
586 caution. Even when calculated from experiments in the same chamber, yields should
587 be interpreted qualitatively and relatively and not extrapolated to the atmosphere.

588 **3.2 Experiments with tropical species**

589 **3.2.1 Gas phase**

590 In order to study the contrast between species that primarily emit monoterpenes and
591 those that primarily emit isoprene, and hence to better understand the isoprene-SOA
592 system, the coupled plant chamber-reaction chamber system was employed to study
593 several tropical species. Two species of fig and one species of palm were selected
594 during the pre-experiment screening process. Those experiments using the figs, *Ficus*
595 *cyathistipula* and *Ficus benjamina* gave the most complete data set; hence their
596 results are used as a focus for discussion.

597 Fig. 5 shows the temporal evolution of a number isoprenoids detected in both the
598 plant and reaction chambers (a and b) and the concomitant evolution of a selection of
599 isoprenoid oxidation products (c and d), during a typical *Ficus benjamina* experiment
600 (23/06/09). During a typical *Ficus benjamina* experiment, approximately thirty
601 precursor and product-ion peaks were observed by the CIR-TOF-MS and PTR-MS in
602 the gas phase. Tentative product identification is reported in the supplementary
603 information (Fig. 6, Tables S1 – S5). A similar set of ions was observed during a
604 typical *Ficus cyathistipula* experiment.

605 From inspection of Fig. 5(a) and (b), the dominance of isoprene in the *Ficus*
606 *benjamina* system is clear, with 12.3 (\pm 4.1) ppbV isoprene detected in the reaction
607 chamber at lights on, compared to 0.8 (\pm 0.4) ppbV monoterpenes, 0.5 (\pm 0.9) ppbV

608 sesquiterpenes and an estimated 2.7 (\pm 0.6) ppbV camphore. Speciation of the
609 monoterpenes by GC-MS indicated that the most dominantly emitted C₁₀ compounds
610 for *Ficus benjamina* were α -pinene, limonene, sabinene and linalool and for *Ficus*
611 *cyathistipula* were α -pinene, β -pinene and limonene (Table 2). The sesquiterpenes, β -
612 caryophyllene and α -cubebene were also identified.

613 Products of isoprene were observed to dominate the evolving *Ficus benjamina* and
614 *Ficus cyathistipula* oxidation systems, with the isobaric primary species MACR and
615 MVK comprising the strongest signals (measured together at m/z 71). For example,
616 during the *Ficus benjamina* experiment of 23/06/09, a combined peak MACR + MVK
617 mixing ratio of 2.9 (\pm 0.7) ppbV was observed (Fig. 5c).

618 Along with MACR and MVK, a series of other ions also associated with isoprene
619 oxidation were detected during *Ficus benjamina* oxidation, including m/z 117 and 99
620 (4-hydroxy-2-methyl-but-2-enoic acid), 103 (C5-alkenediols, C4-hydroxydialdehydes
621 and MPAN), 87 (C4-hydroxycarbonyls and methacrylic acid), 83 (3-methyl furan), 75
622 (hydroxy acetone) and 31 (formaldehyde). Additionally, a signal of m/z 101 was also
623 measured, possibly corresponding to the sum of a series of C5-hydroxycarbonyls and
624 C5-hydroxy hydroperoxides (Tuazon and Atkinson, 1990; Paulson and Seinfeld,
625 1992; Jenkin *et al.*, 1997; Benkelberg *et al.*, 2000; Sprengnether *et al.*, 2002,
626 Benkelberg *et al.*, 2000; Zhao *et al.*, 2004; Surratt *et al.*, 2006;
627 <http://mcm.leeds.ac.uk/MCM>, v3.1). Of the signals observed, those of m/z 83 and 87
628 (tentatively assigned to be 3-methyl furan and C4-hydroxycarbonyls/methacrylic

629 acid), were the greatest in magnitude after MACR + MVK (Fig. 6). The temporal
630 evolution of the sum of all of these products suggests that they are predominantly
631 secondary in nature, forming in the chamber after MACR and MVK. They continued
632 to increase in magnitude as the isoprene signal decreased and as the MACR + MVK
633 signal began to fall (Figs. 5 and 6). During a typical *Ficus benjamina* experiment, the
634 sum of these isoprene products was estimated to reach a peak mixing ratio of ~ 1.7
635 ppbV.

636 A series of lower molecular weight ions were also observed to evolve within the
637 reaction chamber, including m/z 61 (acetic acid), 47 (formic acid), 45 (acetaldehyde),
638 33 (methanol) and 31 (formaldehyde). Each of these compounds has previously been
639 associated with isoprene oxidation and/or with off-gassing from illuminated chamber
640 walls. The m/z 43 and 46 signals, indicative of carbonyls and nitrates, respectively,
641 were also observed to increase significantly during photo-oxidation, indicating the
642 formation and evolution of such species with increasing experiment duration.

643 Besides ions pertaining to the oxidation products of isoprene, a number of spectral
644 features typically derived from monoterpene oxidation products were also observed to
645 form and evolve in the reaction chamber, including, m/z 151, 125, 109, 107, 93 and
646 91. To a first order approximation, the total peak quantity of oxidation products not
647 believed to result from isoprene decay was estimated to be of the order 2 ppbV.
648 However, it should be noted that the presence of isobaric interference in such a
649 complex system, uncharacterised fragmentation, detection limits and the use of

650 pseudo and averaged calibration sensitivities, impose a certain level of unknown
651 uncertainty upon this final value.

652 3.2.2 Particle phase

653 Contrary to the immediate and abundant formation of new particles in the un-seeded
654 *Betula pendula* experiments, the total number of particles and total aerosol mass did
655 not increase above background levels after lights were turned on in the un-seeded
656 tropical *Ficus benjamina* experiments. Figure 7 shows the observed and wall-loss-
657 corrected particle mass concentration during two typical *Ficus benjamina* experiments
658 (22/06/09 and 23/06/09) along with a chamber background experiment. Owing to a
659 lack of particle nucleation in those experiments, the total particle number
660 concentration was too low for the wall loss correction (described in section 2.3) to be
661 implemented. Instead, the average of the wall loss constants determined for the
662 seeded experiments was used to calculate the wall loss corrected mass concentrations
663 reported in Figure 7.

664 Figure 8 shows the observed and wall-loss-corrected particle mass concentration for
665 ammonium sulphate seeded experiments using VOC emissions of *Ficus benjamina*
666 and *Ficus cyathistipula*, as well as a seeded background experiment. The mass at the
667 start of the experiment represents the initial ammonium sulphate mass. In order to
668 quantify the formation of SOA mass during these experiments, the mass increase
669 relative to the starting seed mass was determined in Figure 9 by subtracting the initial
670 ammonium sulphate seed mass from the total wall loss corrected mass. The same

671 | calculation was also performed for the *Betula pendula* seeded experiment (07/07/09).
672 | In contrast to the unseeded *Ficus benjamina* and *Ficus Cyathistipula* experiments,
673 | SOA mass was observed to form when a seed was present in the reaction chamber.
674 | The calculated SOA traces in Figure 9 illustrate a slower build-up of mass during the
675 | isoprene dominated *Ficus benjamina* (15/07/09) and *Ficus Cyathistipula* (30/06/09
676 | and 02/07/09) experiments compared to the much faster SOA mass formation in the
677 | monoterpene dominated *Betula pendula* experiment. Peak masses of the order 1.3 –
678 | 5.5 $\mu\text{g m}^{-3}$ were observed, which when employed with the methodology described in
679 | section 3.1.2, produce SOA yields of 10 and 14 % for each of the two *Ficus*
680 | *Cyathistipula* experiments for which both gas and wall-loss-corrected aerosol data
681 | were available, i.e. 30/06/09 and 02/07/09, respectively (Fig. 9). Uncertainty in
682 | $\Delta(\Sigma\text{VOC})$ is estimated to be $\pm 47\%$ and in the size distribution measurements used in
683 | the wall loss calculations, of the order of $\pm 2\%$. The uncertainties in the wall-loss
684 | correction will likely be substantially greater, but remain unquantified at present.

685 **Discussion and Conclusions**

686 **4.1 *Betula pendula***

687 | In this study we coupled a plant chamber to a photochemical reaction chamber in
688 | order to investigate secondary organic aerosol production from a biogenically
689 | consistent mixture of biogenic volatile organic compounds. We studied silver birch

690 (*Betula pendula*), which emits predominantly monoterpenes, with some
691 sesquiterpenes and oxygenated VOCs but only trace isoprene (Table 2, Fig 3).

692 Our *Betula pendula* experiments showed significant SOA formation (Fig. 4) both in
693 the presence and absence of an ammonium sulphate seed, and reproduced the rate of
694 production and growth of SOA observed in earlier published studies (Mentel et al.,
695 2009; Carlton et al., 2009; e.g. VanReken et al., 2006; Hallquist et al., 2009; Kiendler-
696 Scharr et al., 2009a; Kiendler-Scharr et al., 2009b).

697 The SOA yield values of 39 and 26 % obtained here for *Betula pendula* compare
698 reasonably well with those reported within the literature for single precursor work
699 conducted under similar conditions. For instance, for the two most abundant
700 monoterpenes emitted by *Betula pendula*, i.e. α -pinene and β -pinene, single
701 precursors yields of the order 1 – 43 (16) and 3 – 30 %, respectively, have been
702 observed (values given in parenthesis were obtained from the Manchester aerosol
703 chamber). Similarly for other common monoterpenes such as limonene, myrcene, Δ^3 -
704 carene and α -terpinene, SOA yields of 9 – 34, 6 – 43 (15), 2 – 38 and 8 – 25 %,
705 respectively, have been reported and for β -caryophyllene, 37 – 79 (50) % (Lee et al.,
706 2006a and references therein; Alfarra et al., 2012). In a comparable study to ours,

707 Mentel et al., (2009) reported a fractional mass yield of 11 % for their *Betula pendula*
708 experiments, i.e. slightly lower than those given here, but within the bounds of
709 quantified experimental errors. The yield values obtained here for the *Betula pendula*
710 mesocosm system lie roughly in the middle of the single precursor yield range.

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Deleted: Furthermore in a comparable study to ours, Mentel et al., (2009) reported a fractional mass yield of 11 % for their *Betula pendula* experiments. The yield values obtained here for the *Betula pendula* mesocosm system lie at the lower end of the single precursor yield range.

712 As can be seen from inspection of Fig. 3, the transfer of mass through the *Betula*
713 *pendula* experiment appeared roughly conservative, with a small and steady loss of
714 measured mass from the reaction matrix after ~ 220 min. With the addition of
715 oxygen to the starting body of hydrocarbon material during such an experiment, the
716 total measured mass (i.e. Σ VOCs + SOA) within the system would be expected to
717 increase with time. The absence of such a total measured mass gain (and indeed the
718 mass deficit observed towards the end of the experiment), can most likely be
719 accounted for by considering the various measurement uncertainties involved in
720 producing these data (e.g. assumptions in PTR sensitivity, uncharacterised
721 fragmentation following ionisation, instrument detection limits, etc.) and influences
722 imposed by the chamber walls (including potential loss of more highly oxidised
723 material from the gas phase and greater than expected loss of SOA). Indeed, there is
724 potential for a system mass increase by the end of the experiment to lie within the
725 uncertainty bounds of the CIR-TOF-MS/PTR-MS measurements alone, e.g. the
726 average, single compound PTR measurement uncertainty is $\sim \pm 30\%$, allowing the
727 final measured value of $130 \mu\text{g m}^{-3}$ to have an upper limit of $170 \mu\text{g m}^{-3}$, i.e. greater
728 than the starting value. Considering these results it seems that the system studied is
729 reasonably well characterised given the complications involved in such a task.

730 Certain insights into the mechanisms of SOA formation and growth during the *Betula*
731 *pendula* experiments can be obtained through a combined examination of the VOC

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Deleted: can be seen from inspection of Fig. 3, the transfer of mass through the *Betula pendula* experiment was roughly conservative, with the primary point of note being a small loss of material from the system over time (after ~ 220 minutes). The mass deficit observed towards the end of the experiment can most likely be accounted for by considering the measurement uncertainties involved (e.g. assumptions in PTR sensitivity, uncharacterised fragmentation following ionisation, instrument detection limits, etc) and influences imposed by the chamber walls (including potential loss of more highly oxidised material from the gas phase and greater than expected loss of SOA). However, in general, these results indicate that the system studied is reasonably well characterised. .

732 data, the time-dependent growth curves (Fig. 10) and the aerosol composition data.
733 The data in Fig. 10 demonstrates that during oxidation of the *Betula pendula*
734 emissions and in absence of a seed, SOA mass evolution can be roughly split into two
735 phases. In the early stages of the experiment after nucleation, SOA mass growth
736 increased somewhat rapidly with respect to the amount of precursors reacted;
737 however, after roughly 30 % of the initial precursor mass had been consumed, the rate
738 of mass growth with respect to VOC precursor consumption was observed to reach an
739 approximate steady state. When ammonium sulphate seed was present within the
740 chamber, there was a similarly rapid initial growth with respect to VOC consumption,
741 however this time subsequent aerosol evolution was characterised by a roughly linear
742 mass increase to a much higher final mass by the end of the experiment. Considering
743 the various species of precursor VOCs detected in the *Betula pendula* plant chamber
744 air, and the relative lifetimes of these VOCs, it would seem possible that initially the
745 shorter-lived sesquiterpenes react to form a significant proportion of the high mass,
746 nucleating/condensing species, before being removed from the system (e.g. Jenkin et
747 al., 2012). Subsequent aerosol mass formation as the air in the reaction chamber ages
748 towards the central phase of the experiment, is then likely to result from the
749 partitioning of relatively more volatile products formed from slower reacting
750 monoterpenes, e.g. products such as pinic and pinonic acid from α -pinene oxidation
751 (e.g. Jenkin, 2004; Camredon et al., 2010), and products such as the primary acyclic
752 unsaturated aldehydes, (m/z 111 + 93); the temporal profile of which demonstrates

753 gas-phase loss concomitant with wall-loss-corrected aerosol growth reaching a steady
754 state.

755 Further understanding of the composition and evolution of SOA typical of temperate
756 plant environs comes from investigation of Figure 11, which shows the evolution of
757 the fraction of the ratio of more/less oxygenated material present in the aerosol during
758 the initial stages of a typical unseeded *Betula pendula* experiment (06/07/09). Figure
759 | 11 was constructed using the ratio of m/z 44 to 43 obtained from the AMS (*i.e.* $f_{44/43}$,
760 where m/z 44 is derived from “more” oxidised material and m/z 43 from “less”
761 oxidised material, Ng et al., 2010). In this instance, the $f_{44/43}$ ratio exhibits linear
762 growth with time, from a value of ~ 0.8 to ~ 1.1 , suggesting an increase in the
763 oxygenated content of the aerosol as the experiment ages. Such an increase in
764 oxygenated content is generally observed when precursor species contain multiple
765 C=C bonds (e.g. ocimene and myrcene), offering significant potential for higher
766 aerosol O:C composition (e.g. perhaps species such as acyclic unsaturated aldehydes
767 and their subsequent generations of products). Indeed, the evolution of the $f_{44/43}$
768 ratio observed here is consistent with those results obtained from single precursor
769 experiments investigating the acyclic monoterpene, myrcene and the sesquiterpene, β -
770 caryophyllene (Alfarra et al., 2012; Alfarra et al., 2013).

771 Off-line compositional analysis of the SOA collected at the end of the *Betula pendula*
772 experiments supports the findings obtained from the on-line gas phase and bulk
773 aerosol composition data. The LC-MS² analysis produced chromatograms with peaks

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774 matching those seen in comparable single precursor BVOC experiments, with tracer
775 compounds of both sesquiterpene and monoterpene SOA, detected (Fig. 12).
776 Amongst the compounds observed were those of molecular weight (MW) 238, 242,
777 254 and 256, corresponding to 4-(3,3-dimethyl-2-(3-oxopropyl)cyclobutyl)pent-4-
778 enoic acid, 3-(3,3-dimethyl-2-(3-oxobutyl)cyclobutyl)-3-hydroxypropanoic acid, β -
779 nocaryophyllonic acid/ β -caryophyllinic acid and β -nocaryophyllinic acid,
780 respectively, produced during β -caryophyllene oxidation (e.g. Alfara et al., 2013)
781 and MW 184, corresponding to cis-pinonic acid, produced during α -pinene oxidation.
782 Many other terpene oxidation tracers were detected, including compounds of
783 molecular weight 118 and 200, which are also prevalent in SOA obtained from single
784 precursor myrcene experiments. This observation is in line with findings obtained
785 from the gas phase data, which suggest a relatively significant presence of gaseous
786 acyclic unsaturated aldehydes that would originate from acyclic unsaturated terpenes,
787 such as myrcene.

788 **4.2 Tropical species**

789 In addition to *Betula pendula*, we studied three tropical plant species: two figs (*Ficus*
790 *benjamina* and *Ficus cyathistipula*) and one palm (*Caryota millis*); in this work, we
791 focus on results obtained from the fig plants. All three tropical species were found to
792 be strong isoprene emitters, with very much smaller emissions of monoterpenes,
793 sesquiterpenes, and oxygenated VOCs (Table 2, Fig. 6).

794 During the tropical plant experiments, the primary gas-phase isoprene oxidation
795 products MACR, MVK, formaldehyde, isoprene hydroxy hydroperoxides and the
796 secondary product hydroxyacetone were all observed (e.g. Figs. 5 and 6 and Tables
797 S4 and S5 in the supplementary information). MACR, the isoprene hydroxy
798 hydroperoxides (isoprene epoxide (IEPOX) and isoprene hydroperoxide (ISOPOOH))
799 and hydroxyacetone are all believed to be precursors to SOA formation (Jaoui et al.,
800 2010;Carlton et al., 2009;Kleindienst et al., 2009;Paulot et al., 2009;Kleindienst et al.,
801 2007;Lee et al., 2006;Kroll et al., 2006;Surratt et al., 2006;Claeys et al.,
802 2004b;Rollins et al., 2009;Robinson et al., 2010). In this study, with the exception of
803 MACR and MVK, these products all formed at yields lower than those previously
804 reported (Table 3), with MACR + MVK, hydroperoxides, hydroxy acetone and
805 formaldehyde being observed to form in yields of 17 – 36, 1 – 3, 0 – 2 and 2 – 7 %,
806 respectively during our work. This disagreement may result from differences in OH
807 concentrations and NO_x concentrations in each of the experimental studies. Other
808 isoprene products tentatively identified from the CIR-TOF-MS and PTR-MS data
809 include, C5-alkenediols, C4-hydroxycarbonyls/methacrylic acid and 3-methyl furan
810 (Table S4), which have also previously been associated with SOA formation (e.g.
811 Claeys et al., 2004; Surratt et al., 2006; Robinson et al., 2010). For a typical *Ficus*
812 *benjamina* experiment (23/06/09) the sum of these and other potential isoprene
813 products, excluding MACR + MVK, was estimated to have a combined gas phase
814 yield of the order 18 % (Fig. 5).

815 As can be seen in Fig. 6 mass transfer through the *Ficus benjamina* system was
816 characterised by a slight mass decrease just after the start of the experiment followed
817 by a gradual increase in mass with time. As was stated in section 4.1 a mass increase
818 is expected with time during such an experiment, owing to the addition of oxygen to
819 the precursor hydrocarbon material. Consequently, when considering the data
820 presented in Fig. 6 in the context of potential uncertainties involved (including
821 difficult to characterise influences imposed by the chamber walls), it appears that the
822 system being studied is reasonably well characterised.

823 By comparing Figs. 3 and 6 we see that the monoterpene dominated *Betula pendula*
824 system, which produces larger and lower vapour pressure oxidation products than the
825 isoprene dominated *Ficus* system, as well as measureable SOA, is the case which
826 exhibits measured mass loss. From this contrast it is reasonable to assume a
827 significant fraction of any mass deficit observed during *Betula pendula* oxidation
828 could result from the loss of the heavier, lower volatility compounds that are present
829 in the *Betula pendula* oxidation system but not in the *Ficus* system.

830 Despite the detection of a number of first- and second-generation gas phase products
831 that have previously been directly linked with isoprene SOA composition (Claeys et
832 al., 2004; Wang et al., 2004; Edney et al., 2005; Surratt et al., 2006; Healy et al.,
833 2008), there was no accompanying evidence of SOA formation from the isoprene-
834 emitting tropical plants during unseeded, nucleation style experiments (Fig. 7). A lack
835 of SOA mass formation during our unseeded *Ficus benjamina* experiments could
836 have resulted from a number of different factors, not least of which was simply the

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Deleted: As with the *Betula pendula* experiments, mass transfer through the *Ficus benjamina* system was relatively conservative, with only a slight mass deficit measured during the central phase of the experiment, indicating again that the system being studied is reasonably well characterised. Any deviations from mass conservation are assumed to result from measurement uncertainties and influences imposed by the chamber walls.

837 | absence of a seed surface (acidic or otherwise) to help facilitate partitioning of the
838 semi-volatile oxidation products to the aerosol phase and produce particles of
839 sufficient size and measureable particle mass (e.g. Kroll et al., 2006). Another
840 potentially significant contributing factor in suppressing SOA formation during these
841 experiments was our relatively low VOC/NO_x ratio and the resultant gas phase
842 chemistry. In the presence of high (i.e. ppbV-level) NO_x mixing ratios, RO₂ radicals
843 react with NO to produce mainly alkoxy (RO) radicals. For low molecular mass
844 VOCs such as isoprene, these RO radicals generally fragment into smaller, more
845 volatile products that do not easily partition from the gas phase to the aerosol phase,
846 resulting in a low SOA yield (Surratt *et al.*, 2010). Conversely, under low NO_x
847 conditions, RO₂ radicals are known to undergo self- and cross-reactions to produce
848 organic peroxides and hydroperoxides of relatively low volatility. For example,
849 Surratt *et al.* (2010) showed that under high NO_x conditions the yield of the
850 potentially SOA forming gas phase IEPOX was reduced with respect to the equivalent
851 value under low NO_x conditions, where IEPOX formed in substantial yields (upward
852 of 75 %) from the further oxidation of ISOPOOH by OH.

853 In contrast to our unseeded *Ficus* experiments, when an ammonium sulphate seed was
854 present (and following wall loss correction), SOA mass was observed to form and
855 evolve within the reaction chamber (Fig. 9). From estimates of the total
856 concentrations of precursor VOCs within the reaction chamber matrix (primarily
857 isoprene, e.g. Fig. 5), an SOA mass yield of the order 10 – 14 % was obtained for the
858 *Ficus Cyathistipula* system. If it were to be assumed that the SOA were solely formed

859 from oxidation products of isoprene as the major emitted VOC, this yield would
860 appear excessive in comparison with those obtained previously from single precursor
861 isoprene studies, i.e. $\sim 0.1 - 5.5\%$ ([van Donkelaar et al., 2007](#); [Kleindienst et al.,](#)
862 [2009, 2007](#); [Kroll et al., 2005, 2006](#); [Claeys et al., 2004a](#); [Edney et al., 2005](#) ([van](#)
863 [Donkelaar et al., 2007](#); [Kleindienst et al., 2009, 2007](#); [Kroll et al., 2005, 2006](#); [Claeys et](#)
864 [al., 2004a](#); [Edney et al., 2005](#); [Brégonzio-Rozier et al., 2014](#)). However, we must
865 consider that the mesocosm system is in fact an ensemble of precursors, albeit an
866 ensemble dominated by isoprene, analogous to ambient air above a tropical forested
867 region (Hewitt et al., 2010; MacKenzie et al., 2011).

868 For the experiments of 30/06/09 and 02/07/09, for which SOA yields were obtained
869 for the *Ficus Cyathistipula* system, a fraction of camphore was also observed in the air
870 entering the reaction chamber (presumably for these two particular experiments,
871 sesquiterpenes and monoterpenes were present at concentrations below the detection
872 limits of the PTR-MS and CIR-TOF-MS). The concentration of camphore at lights on
873 was estimated to be $\sim 0.5 - 0.9$ ppbV for the *Ficus cyathistipula* system and $\sim 1.4 -$
874 2.7 ppbV for *Ficus benamina*, and the sum of all *non-precursor* ions in the CIR-
875 TOF-MS mass spectrum $> m/z$ 100 (indicative of *non-isoprene-like* oxidation
876 products; excluding m/z 103 and 117) was estimated to be of the order 2 ppbV by the
877 end of the experiments. This calculation approximates the $m/z > 100$ summation as
878 one large, multifunctional analyte with a PTR sensitivity similar to pinonaldehyde (a
879 typical, multifunctional, high MW molecule resulting from terpene oxidation).
880 Continuing this assumption and taking a range of known VOC terpene product yields

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881 (Y^p_{VOC}) obtained from previous work at the Manchester chamber (i.e. $Y^p_{VOC} = 100\%$
882 the limiting case; 77 % from *Ficus Benjamina* oxidation; 55 % from *Betula Pendula*
883 oxidation; and 29 % for pinonaldehyde and $\Sigma(I_{111}, I_{93})$), a *non-isoprene VOC_{precursor}*
884 concentration may be estimated. Taking a range of known SOA yields obtained from
885 the same reaction chamber (i.e. α -pinene, myrcene, linalool and β -caryophyllene;
886 Alfara et al., 2013), the SOA yield obtained here for the *Betula pendula* system and
887 the estimate of $[VOC_{precursor}]$, Eqn. (2) may be solved to provide a crude estimate of
888 the mass of SOA formed from *non-isoprene* precursors. Consequently, an estimate of
889 the residual SOA mass derived from isoprene oxidation within the *Ficus* system can
890 be inferred for each of the experiments shown in Fig. 9.

891 For 78 of the 120 measurement-and-parameter sets tested, the estimated residual SOA
892 mass resulting solely from isoprene oxidation was negative – i.e., production of SOA
893 from isoprene oxidation was not required to close the mass balance. Values were
894 calculated based on the widest range of peak masses observed during the *ficus*
895 experiments ($M_p = 1.3 \mu\text{g m}^{-3}$ and $5.5 \mu\text{g m}^{-3}$), and assume the lowest (29 %) and
896 highest (100 %) VOC terpene yields and lowest (5 %) and highest (47 %) SOA yields
897 from non-isoprene precursors, respectively, as observed in previous experiments
898 conducted within this chamber. These ranges result in calculated residual SOA mass
899 of -28.5 to +5.0 $\mu\text{g m}^{-3}$ produced solely from isoprene oxidation. Hence, there are
900 combinations of measurements, observations and oxidation/phase-change parameters
901 — omitting isoprene and its oxidation products — that can account for ~20 times the
902 observed aerosol mass production, and other combinations of measurements and

903 parameters that leave up to ~ 90 % of the condensed mass to be explained by isoprene
904 oxidation. If, instead of using the limiting cases, the closest approximation to the *ficus*
905 *cyathistipula* system is used (i.e. $Y_{VOC} = 77\%$ and $Y_{SOA} = Y_{SOA} = \alpha$ -pinene = 15 %),
906 non-isoprene products could have accounted for around 145 % of the SOA mass that
907 was produced. We have no way of assigning formal likelihoods to each set of
908 measurements and parameters in this exercise, but we note that the great
909 preponderance of parameter combinations do not require an isoprene contribution to
910 the SOA mass (i.e. 78/120 measurement-and-parameter sets tested) under our
911 experimental conditions. Moreover, our experiments produce much less SOA mass
912 than would be expected from published experiments using individual mono- and
913 sesquiterpenes.

914 There are three principal reasons why the estimates of aerosol production from
915 isoprene in the tropical plant experiments span such a large range. Firstly, the plants
916 in the mesocosm emit a complicated mixture of biogenic VOCs, some of which are
917 known to oxidise much more rapidly than isoprene and which will produce
918 condensable compounds when oxidised. Secondly, these minor compounds co-
919 emitted from principally-isoprene emitting tropical trees are imperfectly quantified
920 because of the sensitivity of the chemical ionisation (PTR and CIR) instruments.
921 Thirdly, these minor co-emissions are imperfectly characterised because many higher
922 molecular weight compounds, such as the mono- and sesquiterpenes, are isobaric in
923 the PTR and CIR instruments and so precise chemical structures cannot easily be
924 assigned. Without better instrument detection sensitivity and high time resolution

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925 chemical identification for the reactive compounds co-emitted with isoprene, it is not
926 possible to constrain further the aerosol yield from the tropical plants. Unfortunately,
927 insufficient SOA mass formed during *Ficus* experiments to allow us to conduct any
928 form of compositional analysis.

929 ***4.3 Atmospheric significance***

930 Our results are specific to VOC/NO_x ratios of 3 - 9 and NO_x mixing ratios of ~
931 2 ppbV. Note, however, that the three reasons given above for the uncertainty in the
932 aerosol production ascribed to isoprene in our experiments will also pertain to field
933 measurements, often being exacerbated by variability and the difficulties of operating
934 in the field. A contribution of isoprene to SOA is *supported* by recent observations of
935 isoprene related SOA formation above the tropical forest of Danum Valley, Borneo, a
936 high isoprene, low NO_x region (typical ratio of 20:1, isoprene/NO_x) (Hewitt et al.,
937 2010b). Robinson *et al* (2010) observed that up to 15 % by mass of atmospheric sub-
938 micron organic aerosol above the tropical forest of Danum Valley was comprised of
939 methyl furan, the most likely source of which is the oxidation of isoprene (*i.e. thermal*
940 *decomposition of isoprene derived SOA*) (Ruppert and Becker, 2000; Robinson et al.,
941 2010; Lin et al., 2012; Budisulistiorini et al., 2013). Although much smaller in
942 magnitude, the monoterpene emissions measured at Danum Valley were more than
943 adequate to account for the remaining sub-micron organic aerosol (MacKenzie et al.,
944 2011), just as in the majority of aerosol mass calculations for principally-isoprene-
945 emitting tropical trees, described above.

946 It has recently been proposed that isoprene can inhibit aerosol formation when present
947 in air containing other potential SOA precursors, such as mono- and sesquiterpenes
948 (Kiendler-Scharr et al., 2009a). Kiendler-Scharr et al. propose that isoprene could
949 effectively act as an OH scavenger, suppressing new particle formation by slowing
950 the oxidation of available monoterpenes (and presumably sesquiterpenes). In line
951 with this thesis, interpretation of the results obtained from our seeded experiments
952 with *Ficus* species leaves room for a potential role for isoprene in inhibiting SOA
953 formation under certain atmospheric conditions, i.e. our results imply that isoprene
954 may impact negatively on the overall SOA forming potential of air containing other
955 biogenic SOA precursors. However, owing to the constraints laid upon our
956 experiments by the instrumentation and apparatus employed, it is difficult to assign a
957 given certainty level to the role played by isoprene in the ambient atmosphere and
958 caution should be taken when interpreting such findings.

959 The fact that isoprene accounts for approximately 50 % of the total global burden of
960 non-methane VOC, (Guenther et al., 2006), would make it a significant contributor to
961 global SOA. It has been estimated that, even if the secondary organic aerosol yield
962 from isoprene is small (e.g. 1 %), the overall contribution to total atmospheric aerosol
963 could be up to 6 Tg yr⁻¹ (Carlton et al., 2009). Van Donkelaar *et al.* (2007) found that
964 using an isoprene SOA yield of 2 % improved the relationship between model
965 simulations and organic aerosol measurements, and contributed 10 – 50 % of the total
966 organic aerosol loading over the United States during the summer. Understanding the
967 exact role played by isoprene in air containing many different VOCs, and being able

968 | to account for the differing isoprene SOA yields under contrasting NO_x [and acidity](#)
969 | [\(Lin et al., 2012; Lin et al., 2013; Pye et al., 2013\)](#) environments, will undoubtedly
970 | help to significantly improve global modelling estimates of total SOA loading even
971 | further (Couvidat and Seigneur, 2010).

972 | Further to any such potential impacts imposed by isoprene, it has recently been shown
973 | that a range of other BVOC emissions, released in response to a range of
974 | environmental stress factors, can also have significant impacts on biogenic SOA
975 | formation and yield (Mentel et al., 2013). It has been shown that the emissions of
976 | sesquiterpenes, methyl salicylate and C₁₇ BVOCs, released as a result of certain
977 | environmental stress factors have a net positive impact on SOA yield; whereas certain
978 | stress induced green leaf volatiles ((Z)-3-hexenol and (Z)-3-hexenylacetate) behave
979 | similarly to isoprene, suppressing SOA formation [\(Mentel et al., 2013\)](#).

980 | Given the highly differing reported yields of isoprene SOA under various oxidant
981 | schemes, the uncertainty in the exact role played by isoprene and its oxidation
982 | products in realistic mixtures of VOCs (in particular in the context of SOA nucleation
983 | rates; Kiendler-Scharr et al., 2009) and the lack of knowledge regarding stress
984 | induced BVOCs, their atmospheric oxidation and their roles in biogenic SOA
985 | formation (and impact on chemical and physical properties), we suggest that there is a
986 | pressing requirement for additional, atmosphere-relevant laboratory and field studies
987 | to give us the necessary insight to successfully control biogenic SOA (Carlton et al.,
988 | 2010).

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1512 **Tables**

1513 Table 1: List of experiments conducted and their general parameters

Date	Tree Species	Initial NO _x / ppbV	VOC/NO _x	Relative Humidity / %	Pre- existing Seed
22/06/09	<i>Ficus benjamina</i>	3	4.2	79	None
23/06/09	<i>Ficus benjamina</i>	6	2.7	75	None
25/06/09	<i>Ficus benjamina</i>	2	6.3	65	Sulphate
29/06/09	<i>Ficus</i> <i>cyathistipula</i>	2	9.4	71	None
30/06/09	<i>Ficus</i> <i>cyathistipula</i>	2	7.8	75	Sulphate
02/07/09	<i>Ficus</i> <i>cyathistipula</i>	3	5.6	78	Sulphate
06/07/09	<i>Betula pendula</i>	3	5.6	84	None
07/07/09	<i>Betula pendula</i>	3	5.5	73	Sulphate
09/07/09	<i>Betula pendula</i>	2	1.5	70	Sulphate
10/07/09	<i>Betula pendula</i> + 36 ppbV isoprene	2	5.5	70	Sulphate
13/07/09	<i>Ficus benjamina</i>	2	- ¹	87	Sulphate
15/07/09	<i>Ficus benjamina</i>	3	- ¹	89	Sulphate

16/07/09	<i>Ficus benjamina</i> +	2	- ¹	85	None
	4.5 ppbV				
	limonene				

1514

1515 Notes:

1516 1. No quantified VOC data available

1517

1517

1518 Table 2: GC-MS identification of the biogenic VOC present in the plant chamber air
 1519 immediately before RC filling began. Quantification of isoprene, total monoterpenes
 1520 and total sesquiterpenes was carried out using PTR-MS and CIRMS (see Figures 2
 1521 and 5).

<u>Experiment</u>	<u>Compounds detected by GC-MS (abundance ppbV) (trace = <0.1 ppb)</u>			
	<u>Isoprene</u>	<u>Monoterpenes</u> <u>(and related)</u>	<u>Sesquiterpenes</u>	<u>Other</u>
<u>B. Pendula</u>	<u>Yes</u> <u>(0.25-1.19)</u>	<u>α-pinene (0.31 – 1.08)</u> <u>β-pinene (0.74 – 7.19)</u> <u>ocimene (trace – 1.22)</u> <u>Δ^3-carene (1.89 – 4.94)</u> <u>γ-terpinene (trace)</u> <u>2, 4, 6-octatriene,2,6-</u> <u>dimethyl (trace)</u> <u>4,7-methano-1H-indene,</u> <u>octahydro (trace)</u>	<u>β-caryophyllene (0.15</u> <u>– 0.22)</u> <u>α-farnesene (0.14)</u> <u>α-caryophyllene (0.59</u> <u>– 0.92)</u> <u>α-copaene (trace)</u> <u>aromadendrene (0.45-</u> <u>0.51)</u> <u>α-cedrene (trace)</u> <u>α-pyronene (trace)</u>	<u>acetaldehyde (trace)</u> <u>caryophyllene-</u> <u>epoxide (trace)</u> <u>nerolidol (trace)</u> <u>linalool (trace)</u>

<u>F. Benjamina</u>	<u>Yes</u>	<u>α-pinene (trace)</u>	<u>α-cubebene (trace)</u>	<u>acetaldehyde (trace)</u>
	<u>(38.49)</u>	<u>limonene (trace)</u>		<u>benzoquinone</u>
		<u>sabinene (trace)</u>		<u>(trace)</u>
		<u>linalool (trace)</u>		<u>pyridine (trace)</u>
				<u>methyl salicate</u>
				<u>(trace)</u>
				<u>decanal (trace)</u>
<u>F. Cyathistipula</u>	<u>Yes (75.08)</u>	<u>α-pinene (trace)</u>	<u>β-caryophyllene (trace)</u>	<u>acetic acid (trace)</u>
		<u>β-pinene (trace)</u>		
		<u>limonene (trace)</u>		
<u>Mixed canopy</u>	<u>Yes</u>	<u>α-pinene (0.37)</u>	<u>None detected</u>	<u>p-dichlorobenzene</u>
<u>F. Benjamina</u>		<u>camphene (0.11)</u>		<u>methyl salicate</u>
<u>F. Cyathistipula</u>		<u>limonene (0.42)</u>		
<u>C. Millis</u>		<u>ocimene (trace)</u>		

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1526

1527 Table 3: Yields of isoprene oxidation products compared to literature values. Yields
 1528 are an average from all *Ficus* experiments (seeded and unseeded) (n = 4) calculated at
 1529 4 hours after lights on (HALO). Yield is based on the calculated relationship between
 1530 the amount of isoprene reacted and the oxidation product in question.

1531

Isoprene	Hydro- peroxides ¹	MVK +MACR	Hydroxy- acetone	Form- aldehyde	Source
1	0.33	0.26	0.07		Williams et al 1999
1	0.05	0.36	0.05		Williams et al 2001 (NO _x present)
1		0.33		0.67	Zang et al 2002
1		0.46 – 0.60		0.86 – 0.96	Niki et al 1983 Kamens et al 1982
1	0.18	0.25			Surratt et al 2010 NO _x present NO _x absent
1		0.27			Kleindienst et al 2009
1	0.01 –	0.17 – 0.36	0 – 0.02	0.02 – 0.07	This study

Isoprene	Hydro- peroxides ¹	MVK +MACR	Hydroxy- acetone	Form- aldehyde	Source
0.03					

1532

1533 Notes:

1534 1. Sum of isoprene epoxide (IEPOX) and isoprene hydroperoxide (ISOPOOH)

1535

1535

1536 **Figure Captions**

1537

1538 Figure 1: Temporal evolution of NO, NO₂, NO_x and O₃ during typical *Ficus*
1539 *Cyathistipula* (a) and *Betula pendula* (b) experiments (25/06/09 and 07/07/09,
1540 respectively).

1541

1542 Figure 2: Temporal evolution of a series of isoprenoid “precursor” compounds and
1543 their oxidation products, as observed in the main reaction chamber during an example
1544 *Betula pendula* experiment (07/0709). The top panel (a) shows the entire experiment
1545 process in terms of monoterpene evolution, from background (BG) and plant chamber
1546 (PC) measurements, to reaction chamber (RC) fill and the main photooxidation
1547 experiment within the reaction chamber. Figure (b) shows monoterpene evolution, (c)
1548 shows sesquiterpenes (red circles and lines) and camphore (black triangles and lines),
1549 (d) $\Sigma(I_{111}, I_{93})$ and (e) the primary ketone (m/z 139) (red circles and lines), primary
1550 keto-aldehyde (m/z 107 + 151 + 169) (black triangles and lines) and MVK + MACR
1551 (m/z 71) (grey diamonds and dashed line).

1552

1553

1554 Figure 3: Evolution of measured mass through the *Betula pendula* system (7 July
1555 2009), showing the relative contribution of precursor compounds, oxidation products
1556 and SOA mass to total measured mass, with time (coloured bars, left axis) and total
1557 measured mass (i.e. Σ VOCs + SOA) with time (black line, right axis). Note:
1558 ammonium sulphate seed mass removed from the SOA mass concentration.

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Deleted: Figure 3: Evolution of mass through the *Betula pendula* system (07/07/09), showing precursor compounds, oxidation products and SOA mass with time. Note: ammonium sulphate seed mass removed from the SOA mass concentration. .

1559

1560 Figure 4: Particle number and mass concentrations measured during nucleation
1561 (06/07/09) and ammonium sulphate seeded (07/07/09) *Betula pendula* experiments
1562 (a). In the bottom panel (b), both the measured (dashed lines) and the wall loss
1563 corrected (solid lines) mass concentrations are shown.

1564

1565 Figure 5: Temporal evolution of a series of isoprenoid “precursor” compounds and
1566 their oxidation products, as observed in the main reaction chamber during an example
1567 *Ficus benjamina* experiment (23/06/09). Panel (a) shows isoprene (red circles and
1568 lines) and camphore (black triangles and lines) evolution, (b) shows monoterpenes
1569 (red circles and lines) and sesquiterpenes (black triangles and lines), (c) MVK +
1570 MACR (m/z 71) and (d) Σ (monoterpene products) (black triangles and lines) and
1571 Σ (non MVK+MACR isoprene products) (red circles and lines)

1572

1573 Figure 6: Evolution of measured mass through the *Ficus benjamina* system (23 June
1574 2009), showing the relative contribution of precursor compounds and oxidation
70

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1575 products to total measured mass, with time (coloured bars, left axis) and total
1576 measured mass (i.e. Σ VOCs + SOA) with time (black line, right axis).

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1577
1578 Figure 7: Observed and wall loss corrected particle mass concentrations during un-
1579 seeded *Ficus benjamina* (22/06/09, 23/06/09) and chamber background (26/06/09)
1580 experiments. The reaction chamber was filled with plant chamber air over a period of
1581 1 – 1.5 hours. Chamber filling was carried out in the dark. Ozone was added
1582 immediately prior to lights on. Time begins at the point at which the reaction chamber
1583 was illuminated, then increments in hours after lights on.

1584

1585 Figure 8: Observed and wall loss corrected particle mass concentrations during
1586 ammonium sulphate seeded *Ficus benjamina* (15/07/09), *Ficus cyathistipula*
1587 (30/06/09, 02/07/09) and chamber background (03/07/09) experiments. Ozone and
1588 ammonium sulphate seed were added immediately prior to lights on.

1589

1590 Figure 9: Calculated SOA mass concentrations during ammonium sulphate seeded
1591 experiments for *Betula pendula* (07/07/09), *Ficus benjamina* (15/07/09) and *Ficus*
1592 *cyathistipula* (30/06/09, 02/07/09). See text for details.

1593

1594 Figure 10: Time dependent growth curves for two typical *Betula pendula* experiments
1595 (red circles- nucleation experiment on 06/07/09 and black triangles- ammonium
1596 sulphate seeded experiment on 07/07/09), showing SOA growth behaviour with
1597 respect to consumption of the VOC precursors.

1598

1599 Figure 11: Temporal evolution of the m/z 44/43 ratio (red circles) during a typical
1600 *Betula pendula* experiment (06/07/09) and wall loss corrected SOA mass (black line);
1601 demonstrating the increase in oxygenated content of the SOA as the air matrix begins
1602 to age.

1603

1604 Figure 12: LC-MS² selected ion chromatograms derived from the off-line analysis of
1605 SOA collected on filters at the conclusion of a typical *Betula pendula* experiment
1606 (07/07/2009). Notes: Upper; m/z 183 = MW 184, 1 = cis-pinonic acid. Middle; m/z
1607 253 = MW 254, 2 = β -nocaryophyllonic acid, 3 = β -caryophyllinic acid, 4 = similar
1608 to sesquiterpene SOA. Lower; m/z 257 = MW 118 [$2*[M-H] + Na$], also seen in
1609 myrcene SOA, with same MS² spectra.