

1 **Emissions of biogenic volatile organic compounds and subsequent photochemical**  
2 **production of secondary organic aerosol in mesocosm studies of temperate and**  
3 **tropical plant species**

4 Kevin P. Wyche<sup>1,2</sup>, Annette C. Ryan<sup>3\*</sup>, C. Nicholas Hewitt<sup>3</sup>, M. Rami Alfarra<sup>4,5</sup>, Gordon  
5 McFiggans<sup>4</sup>, Timo Carr<sup>1</sup>, Paul S. Monks<sup>1</sup>, Kirsty L. Smallbone<sup>2</sup>, Gerard Capes<sup>5</sup>,  
6 Jacqueline F. Hamilton<sup>6</sup>, Thomas A. M. Pugh<sup>7</sup>, and A. Robert MacKenzie<sup>8</sup>

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8 1. Department of Chemistry, University of Leicester, Leicester, LE1 7RH, UK

9 2. School of Environment and Technology, University of Brighton, Brighton, BN2 4GJ

10 3. Lancaster Environment Centre, Lancaster University, Lancaster, LA1 4YQ, UK

11 4. School of Earth, Atmospheric and Environmental Sciences, The University of  
12 Manchester, M13 9PL, UK

13 5. National Centre for Atmospheric Science, The University of Manchester, M13 9PL,  
14 UK

15 6. Dept. of Chemistry, University of York, York, YO10 5DD, UK

16 7. Karlsruhe Institute of Technology, IMK-IFU, Garmisch-Partenkirchen, Germany

17 8. Birmingham Institute of Forest Research, University of Birmingham, B15 2TT, UK

18

19

20 \*For correspondence:

21 A. C. Ryan,

22 Lancaster Environment Centre,

23 Lancaster University,

24 Lancaster,

25 LA1 4YQ,

26 UK.

27

28 Tel: 01524 594 534

29 Email: [a.ryan@lancaster.ac.uk](mailto:a.ryan@lancaster.ac.uk)

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31 **Abstract:** Silver birch (*Betula pendula*) and three Southeast Asian tropical plant  
32 species (*Ficus cyathistipula*, *Ficus benjamina* and *Caryota millis*) from the  
33 pantropical fig and palm genera were grown in a purpose-built and environment-  
34 controlled whole-tree chamber. The volatile organic compounds emitted from these  
35 trees were characterised and fed into a linked photochemical reaction chamber where  
36 they underwent photooxidation under a range of controlled conditions (RH ~ 65 – 89  
37 %, VOC/NO<sub>x</sub> ~ 3 – 9 and NO<sub>x</sub> ~ 2 ppbV). Both the gas phase and the aerosol phase of  
38 the reaction chamber were monitored in detail using a comprehensive suite of on-line  
39 and off-line, chemical and physical measurement techniques.

40 Silver birch was found to be a high monoterpene and sesquiterpene, but low isoprene  
41 emitter, and its emissions were observed to produce measureable amounts of SOA via  
42 both nucleation and condensation onto pre-existing seed aerosol (Y<sub>SOA</sub> 26 – 39 %). In  
43 contrast, all three tropical species were found to be high isoprene emitters with trace  
44 emissions of monoterpenes and sesquiterpenes. In tropical plant experiments without  
45 seed aerosol there was no measurable SOA nucleation, but aerosol mass was shown to  
46 increase when seed aerosol was present. Although principally isoprene emitting, the  
47 aerosol mass produced from tropical fig was mostly consistent (i.e., in 78 out of 120  
48 aerosol mass calculations using plausible parameter sets of various precursor specific  
49 yields) with condensation of photooxidation products of the minor VOCs co-emitted;  
50 no significant aerosol yield from condensation of isoprene oxidation products was  
51 required in the interpretations of the experimental results. This finding is in line with  
52 previous reports of organic aerosol loadings consistent with production from minor

53 biogenic VOCs co-emitted with isoprene in principally-isoprene emitting landscapes  
54 in Southeast Asia. Moreover, in general the amount of aerosol mass produced from  
55 the emissions of the principally-isoprene-emitting plants, was less than would be  
56 expected from published single-VOC experiments, if co-emitted species were solely  
57 responsible for the final SOA mass. Interpretation of the results obtained from the fig  
58 data sets, leaves room for a potential role for isoprene in inhibiting SOA formation  
59 under certain ambient atmospheric conditions, although instrumental and  
60 experimental constraints impose a level of caution in the interpretation of the results.

61 Concomitant gas and aerosol phase composition measurements also provide a detailed  
62 overview of numerous key oxidation mechanisms at work within the systems studied  
63 and their combined analysis provides insight into the nature of the SOA formed.

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65 *Keywords: Secondary organic aerosol, biogenic volatile organic compounds, BVOC,*  
66 *gas-aerosol partitioning, isoprene, monoterpenes, mesocosm*

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71 **Introduction**

72 Atmospheric aerosols change the radiative balance of the Earth through scattering and  
73 absorbing incident solar radiation (Kim and Ramanathan, 2008); they directly and  
74 indirectly affect the properties and formation of clouds, thus altering the hydrological  
75 cycle (Gunthe et al., 2009; Junkermann et al., 2009; Stevens and Feingold, 2009); and  
76 they may have an impact on the efficiency of plant photosynthesis (Mercado et al.,  
77 2009), thereby modifying the uptake of atmospheric carbon. Hence, aerosol particles  
78 affect the Earth's climate in several ways (as reviewed in Hallquist et al., 2009; IPCC,  
79 2007; Isaksen et al., 2009; Carslaw et al., 2010) as well as having a detrimental impact  
80 on human health (e.g., Baltensperger et al., 2008).

81 A large fraction of the observed atmospheric aerosol composition is organic (Zhang et  
82 al., 2007). A primary organic component is emitted directly into the atmosphere from  
83 anthropogenic activities, such as biomass burning and fossil fuel combustion, or is  
84 emitted from natural sources, such as plant abrasion and the sea surface. Secondary  
85 aerosol particles are formed within the atmosphere by gas-to-particle conversion;  
86 those formed from gas-phase organic precursors are known as secondary organic  
87 aerosol (SOA) (e.g., Riipinen et al., 2012). There is considerable uncertainty  
88 surrounding the chemical transformation of anthropogenic and biogenic volatile  
89 organic compounds (AVOC and BVOC, respectively) from the gas phase to the  
90 aerosol phase and hence, considerable uncertainty in the global source of SOA  
91 (Hallquist et al., 2009; Donahue et al., 2009; Ng et al., 2006; Virtanen et al., 2010).

92 On a global scale, approximately 90 % of all volatile organic compound emissions  
93 originate from biogenic sources (Guenther et al. 2012), with almost half of this being  
94 emitted from tropical and subtropical forests. The ability of biogenic VOC to form  
95 SOA is therefore of particular interest and potential importance. Globally, isoprene  
96 (2-methyl-1,3-butadiene, C<sub>5</sub>H<sub>8</sub>) is the biogenic VOC with the largest mass emission  
97 rate. It is estimated to account for about 50 % of BVOC emissions by mass (Guenther,  
98 et al. 2012), but it is still uncertain how much it contributes to SOA formation (Karl et  
99 al., 2009;Carlton et al., 2009).

100 Modelling, laboratory chamber experiments and field studies provide a range of  
101 possible yields of SOA from isoprene, typically of the order 0.1 – 3 % by mass, with  
102 some values reported as high as 5.5 % (van Donkelaar et al., 2007;Kleindienst et al.,  
103 2009, 2007;Kroll et al., 2005, 2006;Claeys et al., 2004a;Edney et al., 2005;  
104 Brégonzio-Rozier et al., 2014). SOA yields from the further oxidation of first and  
105 subsequent generation isoprene oxidation products, such as methacrolein, are  
106 estimated to be as much as 15 % (Rollins et al., 2009;Carlton et al., 2009;Claeys et  
107 al., 2004b;Robinson et al., 2010). Recent work has highlighted that under low NO<sub>x</sub>  
108 conditions, SOA mass formed from isoprene oxidation could be influenced by the  
109 acidity of pre-existing aerosol via the reactive uptake of certain key isoprene  
110 oxidation products, namely isoprene epoxydiols (IEPOX; Surratt et al., 2010; Lin et  
111 al., 2012). More recently, Nguyen et al. (2014) found that the “pH dependence for OA  
112 formation from IEPOX was weak for AS particles”. There is further evidence from  
113 chamber studies using temperate tree species such as birch, spruce and pine that

114 isoprene may in fact suppress SOA formation from other VOC precursors, when  
115 present (Kiendler-Scharr et al., 2009a; Kanawade et al., 2011). It should be noted at  
116 this point that it is unclear in most cases how wall effects have been considered in the  
117 production of such yield values and whether the treatments employed are adequate  
118 such that the yields are comparable between chambers, or indeed between  
119 experiments.

120 Here, we characterised the BVOC emissions from three south-east Asian tropical  
121 plant species (*Ficus cyathistipula*, *Ficus benjamina* and *Caryota millis*) and in a series  
122 of coupled plant growth chamber–atmospheric reaction chamber experiments, we  
123 examined the ability of their oxidation products to contribute to SOA formation under  
124 atmospherically relevant conditions. In order to provide a geographically and  
125 chemically contrasting study, we replicated these experiments using common silver  
126 birch (*Betula pendula*). Silver birch has previously been shown to contribute to the  
127 formation of secondary organic aerosol via the emissions of mono- and sesquiterpenes  
128 (e.g. Kiendler-Scharr et al., 2009a; 2009b; Mentel et al., 2009). Seeded (ammonium  
129 sulphate) and un-seeded experiments were carried out to allow studies of both fresh  
130 nucleation and condensation onto pre-existing aerosol.

## 131 **Methods and materials**

### 132 ***2.1 Plant selection and pre-screening***

133 Three non-clonal specimens of common silver birch (*Betula pendula*), a monoterpene  
134 and isoprene emitting tree species; two species of fig (*Ficus benjamina* and *Ficus*  
135 *cyathistipula*), and one species of palm (*Caryota millis*), each approximately 1.5 m in  
136 height were used. Figs and palms are abundant in all tropical rainforests. We chose  
137 three species found in abundance throughout south and southeast Asia to be consistent  
138 with our field work (Hewitt et al., 2010a; MacKenzie et al., 2011). *Ficus benjamina*  
139 (Moraceae) is native to Malaysia and has previously been found to be a high isoprene  
140 emitter ( $0.03 - 8.7 \mu\text{g C g}^{-1} \text{h}^{-1}$ , potted and in soil) with emissions of the  
141 monoterpenes, limonene ( $0.02 \mu\text{g C g}^{-1} \text{h}^{-1}$ ) and  $\beta$ -ocimene ( $1.8 - 2.5 \mu\text{g C g}^{-1} \text{h}^{-1}$ ),  
142 and the sesquiterpenes  $\beta$ -caryophyllene and  $\alpha$ -copaene (Carvalho et al., 2005; Geron  
143 et al., 2006). In addition, emissions of benzaldehyde ( $0.53 \mu\text{g C g}^{-1} \text{h}^{-1}$ ) and acetaldehyde  
144 ( $69 \mu\text{g C g}^{-1} \text{h}^{-1}$ ) from potted specimens have been detected (Carvalho et al., 2005).  
145 No previous data are available on the BVOC emissions from *Ficus cyathistipula* or  
146 *Caryota millis*. Proton transfer-reaction mass spectrometry (PTR-MS) and gas  
147 chromatography–mass spectrometry (GC-MS) screening, prior to the start of the  
148 coupled chamber experiments, confirmed that both species were high isoprene  
149 emitters with *Ficus cyathistipula* also emitting limonene,  $\beta$ -phellandrene,  $\alpha$ -  
150 damascone and acetaldehyde. Analytical methods are described in detail in section  
151 2.4.

## 152 **2.2 Plant chamber design**



153 A 4.7 m<sup>3</sup> plant chamber was constructed out of two rectangular Teflon bag sections  
154 and a Teflon lid (0.05 mm FEP) (Adtech Polymer Engineering, UK), which were each  
155 supported by frames built using 25 mm<sup>2</sup> box aluminium (Speed Frame, RS  
156 Components, UK). The framework stood on a raised foil and Teflon covered marine  
157 plywood base. PVC foam strips (RS Components, UK) ensured an airtight seal  
158 between chamber sections. Heavy-duty double-sided tape (RS components, UK) was  
159 used to secure the Teflon bags to the frame. The interior of the plant chamber was  
160 only exposed to Teflon surfaces.

161 Compressed air was constantly supplied to the plant chamber via a mass flow  
162 controller and regulator (ALICAT MCR-500 SLPM-D, Premier Control Technologies  
163 Ltd, UK) at 780 L min<sup>-1</sup> and 7.5 bar via a 12.7 mm (outer diameter- OD) reinforced  
164 tube. This was reduced to approx 1 bar and between 250 to 300 L min<sup>-1</sup> (+/- 0.8 %)  
165 dependent on the photosynthetic and transpiration rate of each plant species  
166 (equivalent to one complete air change every 15 - 20 mins). The air stream was  
167 passed through a 12.7 mm (OD) PTFE tube to three in-series filters to remove any  
168 pre-existing VOCs (activated carbon filter P3KFA14ASMN, Parker Pneumatic, UK),  
169 and submicrometer particles (HEPA CAP 75 filter capsule (FDP-780-050K, Fisher  
170 Scientific, UK)), and NO<sub>x</sub> (Purafil and activated charcoal, Purafil Inc. USA). Finally,  
171 the air was re-humidified by passing it through a 2 L Teflon barrel (Jencons, UK)  
172 filled with warmed distilled water. The plant chamber outlet air was either vented into  
173 the laboratory via a 50 mm (OD) stainless steel pipe and valve, or used to fill an 18  
174 m<sup>3</sup> Teflon reaction chamber.

175 To enhance mixing, air entered the plant chamber via a perforated 12.7 mm (OD)  
176 PTFE tube that circled the base of the chamber. One 12.7 mm stainless steel bulkhead  
177 fitting (Swagelok, UK) was inserted through the frame to secure the PTFE tube to the  
178 base of the plant chamber. A 50 mm (OD) stainless steel pipe was inserted into the  
179 upper corner of the chamber and supported by a Teflon (inner surface) and Nylon  
180 (outer surface) manifold (Plastics Direct, UK). The manifold also supported an EGM  
181 probe (EGM-4, PP Systems, UK), which recorded relative humidity (RH),  
182 temperature (T), CO<sub>2</sub> and photosynthetically active radiation (PAR).

183 Plants were kept in 255 – 330 mm (height) pots depending on species, watered to pot  
184 dripping point and sprayed twice weekly. Plant chamber conditions were maintained  
185 at 31 – 33.5 °C / 22 – 24 °C (day/night), 29 – 40 % / 33 – 44 % (day/night) RH, and  
186 335 – 385 ppmV / 390 – 404 ppmV (day/night) CO<sub>2</sub>. Owing to structural restrictions,  
187 PAR could not be measured directly under the growth lamps in the centre of the  
188 canopy. At the top edge of the canopy it was 500 μmol m<sup>-2</sup> s<sup>-1</sup> with a 12 hr day / night  
189 cycle.

### 190 ***2.3 Reaction chamber description***

191 The aerosol photochemical reaction chamber at the University of Manchester is  
192 composed of an 18 m<sup>3</sup> FEP Teflon bag mounted on three rectangular extruded  
193 aluminium frames (Alfarra et al., 2012). A bank of halogen lamps and a 6 kW Xenon  
194 arc lamp are mounted on the enclosure housing the bag, which is coated with  
195 reflective “space blanket” providing an integrating sphere, maximising the irradiance

196 in the bag and ensuring even illumination for the production of photochemical species  
197 such as the hydroxyl radical (OH). The air introduced to the bag is dried and filtered  
198 for gaseous impurities and particles, prior to humidification with high purity  
199 deionised water. A high capacity O<sub>3</sub> generator provides controlled ambient levels of  
200 O<sub>3</sub> (used as an oxidant) and high O<sub>3</sub> concentrations (serving as a cleaning agent  
201 between experiments).

202 Size-dependent (diffusional and gravitational) wall-loss rate constants were calculated  
203 based on particle mobility and the surface-to-volume ratio of the chamber (Verheggen  
204 and Mozurkewich, 2006). The diffusional loss rate uses a constant of proportionality,  
205 which can only be determined empirically. A time period was selected near the end of  
206 each experiment where the wall losses were deemed to be the dominant process  
207 affecting the size distribution. The volume size distribution at the beginning of this  
208 period had the calculated wall loss rate applied to simulate the evolution of the size  
209 distribution over the selected time period. If the calculated loss rate loss rate didn't  
210 reproduce the measured volume evolution within the specified tolerance (1 – 2 % in  
211 this work), the constant of proportionality for diffusional losses was adjusted such that  
212 the simulated volume at the end of the selected period matched the measured volume  
213 within the specified tolerance. The time-integrated gravitational and (optimised)  
214 diffusional loss rate constants were then applied to the volume size distribution  
215 throughout the experiment in order to reconstruct a wall loss corrected size  
216 distribution, which was then used to calculate the wall-loss-corrected particle mass.

217 Both the plant chamber and reaction chamber were tested for contaminants separately  
218 and when joined together by running the system with an empty plant chamber and by  
219 carrying out a “blank” run prior to each set of experiments.

## 220 ***2.4 Analytical techniques***

### 221 *2.4.1 Gas phase measurements*

222 The volatile and semi-volatile organic compounds and oxygenated volatile organic  
223 compounds in both the plant chamber and the reaction chamber were measured by  
224 soft-ionisation mass spectrometry (PTR-MS, CIR-TOF-MS, described below) and gas  
225 chromatography-mass spectrometry (GC-MS).

226 The proton-transfer-reaction mass spectrometry (PTR-MS) instrument employed  
227 (Ionicon, Austria) comprises two turbomolecular pumps, a heated silica steel inlet  
228 system and a 9.6 cm long stainless steel drift tube. The nominal response time is  
229 approximately 1 s. The operating parameters of the PTR-MS were held constant  
230 during measurements, except for the secondary electron multiplier voltage, which was  
231 optimised each day. The drift tube pressure, temperature and voltage were 2.2 hPa, 50  
232 °C, and 600 V, respectively. The central reaction chamber of the drift cell was  
233 operated at an  $E/N$  (*i.e.* electric field/gas number density) of 125 Td. The count rate of  
234  $\text{H}_3\text{O}^+\cdot\text{H}_2\text{O}$  ions was 1 – 2 % of the count rate of  $\text{H}_3\text{O}^+$  ions. The PTR-MS sampled  
235 continuously with a flow rate of 100 – 150 ml min<sup>-1</sup> through 3.2 mm PTFE tubing.

236 The chemical-ionisation-reaction time-of-flight mass spectrometer (CIR-TOF-MS)  
237 comprises a temperature controlled ( $40 (\pm 1) ^\circ\text{C}$ ) ion-source drift cell assembly  
238 coupled to an orthogonal time-of-flight mass spectrometer equipped with a reflectron  
239 array (Kore Technology Ltd, Ely, UK). The ion-source deployed was a hollow  
240 cathode discharge type (Blake et al., 2009) and the chemical ionization technique  
241 used was proton transfer reaction from hydrated hydronium ( $\text{H}_3\text{O}^+\cdot\text{H}_2\text{O}$ ) (Jenkin et  
242 al., 2012). Sample air was delivered in a continuous stream directly to the drift cell  
243 via a 0.5 m long, 6.35 mm (internal diameter) Teflon sample line, heated to  $40 (\pm 1)$   
244  $^\circ\text{C}$ , at a constant flow rate of  $80 \text{ ml min}^{-1}$ . The central reaction chamber of the drift  
245 cell was operated at an  $E/N$  ratio of  $\sim 90 - 100 \text{ Td}$ , with a tuned energy ramp at the  
246 base of the cell to remove potential water-cluster ions (e.g.  $\text{RH}^+\cdot\text{H}_2\text{O}$ ). Further  
247 information regarding the CIR-TOF-MS design and a detailed discussion regarding its  
248 operation can be found in Blake et al. (2004) and Wyche et al. (2007).

249 The PTR-MS and CIR-TOF-MS were calibrated using three different methods: (i)  
250 step-wise dilution of a gravimetrically prepared gas standard (BOC Special Gases,  
251 UK) containing a variety of VOCs and OVOCs; (ii) using calibration material  
252 produced in-house via the injection of liquid samples into 10 l Tedlar bags (SKC Inc.,  
253 USA) containing either humidified or dry, pure nitrogen; and (iii) using gas standards  
254 derived from permeation tubes (Vici Inc., US; Ecoscientific, UK), diluted, humidified  
255 and delivered by a commercial calibration unit (Kintec, model: 491). Where  
256 experimental calibration was not possible for a specific compound, either the  
257 calibration sensitivity for a structurally similar surrogate was used or calculated

258 concentrations were employed (Jenkin et al., 2012). For the quantification of isobaric  
259 signals, a single sensitivity value was used, e.g.  $\alpha$ -pinene sensitivity for  
260  $\Sigma$ (monoterpenes) and  $\beta$ -caryophyllene sensitivity for  $\Sigma$ (sesquiterpenes); again  
261 working on the principal that structurally similar compounds possess similar PTR and  
262 CIR sensitivities.

263 CIR-TOF-MS and PTR-MS detection limits are reagent, reaction, analyte and sample  
264 matrix specific. However, typical CIR-TOF-MS detection limits, using PTR  
265 ionisation from hydronium, are of the order  $0.4 \text{ ppbV (10 min)}^{-1}$  for more polar  
266 compounds, such as OVOCs (e.g. 2-hexanone) and as much as  $10 \text{ ppbV min}^{-1}$  for  
267 certain less polar compounds, such as smaller hydrocarbons (e.g. 1-pentene). For  
268 further details see Wyche et al., 2007.

269 The GC-MS system (GC-MS Turbomass Gold, Perkin Elmer, USA) comprised a  
270 thermal desorption autosampler (Perkin-Elmer ATD 400) connected via a heated (200  
271  $^{\circ}\text{C}$ ) transfer line to a Hewlett-Packard 5890 GC with a 5970 mass-selective detector.  
272 Compounds were desorbed at  $280 \text{ }^{\circ}\text{C}$  for 5 min at  $25 \text{ mL min}^{-1}$  onto a Tenax-TA cold  
273 trap maintained at  $-30 \text{ }^{\circ}\text{C}$ . The cold trap was then heated to  $300 \text{ }^{\circ}\text{C}$  for 6 min to  
274 desorb compounds onto the GC column. Chromatographic separation was achieved  
275 using an Ultra-2 column (Agilent Technologies:  $50 \text{ m} \times 0.2 \text{ mm ID} \times 0.11 \text{ } \mu\text{m}$  film, 5  
276 % phenylmethyl silica). An initial oven temperature of  $35 \text{ }^{\circ}\text{C}$  was maintained for 2  
277 min, and then increased at  $4 \text{ }^{\circ}\text{C min}^{-1}$  to  $160 \text{ }^{\circ}\text{C}$  followed by an increase of  $45 \text{ }^{\circ}\text{C}$   
278  $\text{min}^{-1}$  to  $300 \text{ }^{\circ}\text{C}$ , which was maintained for 10 min. The carrier gas was Helium

279 supplied at a rate of  $1 \text{ mL min}^{-1}$ , with an injector temperature of  $250 \text{ }^\circ\text{C}$ . The limit of  
280 detection for isoprene and monoterpenes was approximately  $0.25 \text{ ng on column}$  and  $2$   
281  $\text{ng on column}$  for sesquiterpenes, corresponding to  $100 \text{ pptV}$  of isoprene,  $50 \text{ pptV}$  of  
282 monoterpenes, and to  $400 \text{ pptV}$  of sesquiterpenes in a  $1 \text{ L}$  sample. Sampling was  
283 conducted by drawing  $8 \text{ L}$  of the analyte air through  $6.35 \text{ mm}$  PTFE tubing onto the  
284 GC-MS sample tubes using a handheld pocket pump (SKC Ltd, UK) at a flow rate of  
285  $150 \text{ ml min}^{-1}$  (total sample time  $\sim 43 \text{ mins}$ ). Sample tubes were stored at  $4 \text{ }^\circ\text{C}$  until  
286 analysed. VOC quantification was by comparison with commercially available liquid  
287 standards (Aldrich, Fluka and Sigma) diluted in methanol. Isoprene quantification  
288 was by comparison with a  $700 \text{ ppbV}$  in  $\text{N}_2$  certified gas standard (BOC, UK).

289  $\text{NO}$  and  $\text{NO}_2$  mixing ratios were measured using a chemiluminescence gas analyser  
290 (Model 42i, Thermo Scientific, MA, USA). Ozone was measured using a UV  
291 photometric gas detector (Model 49C, Thermo Scientific, MA, USA).

#### 292 *2.4.2 Particle phase measurements*

293 Within the main reaction chamber, a scanning mobility particle sizer (SMPS) system  
294 was used to measure the particle size distribution and total aerosol mass concentration  
295 (without sample drying). A particle density of  $1.3 \text{ g cm}^{-3}$  was assumed for calculating  
296 the mass of SOA particles in un-seeded experiments (Alfarra et al., 2006; Bahreini et  
297 al., 2005). For seeded experiments, a density of  $1.77 \text{ g cm}^{-3}$  was used to calculate the  
298 ammonium sulphate seed mass and  $1.3 \text{ g cm}^{-3}$  was assumed for calculating the  
299 additional SOA mass. A water-based condensation particle counter (wCPC, TSI

300 3786) was used to count the total particle number concentration between 2.5 nm and  
301 approximately 3  $\mu\text{m}$ . Further instrument details can be found in Alfarra et al., 2012,  
302 and references therein.

303 Real-time broad chemical characterisation of the SOA was made using a compact  
304 Time-of-Flight Aerosol Mass Spectrometer (cToF-AMS, Aerodyne Research Inc.,  
305 USA). A detailed description of the instrument, its operation and calibrations can be  
306 found elsewhere (Drewnick et al., 2005;Canagaratna et al., 2007). The instrument was  
307 operated in the standard configuration, taking both mass spectrum (MS) and particle-  
308 time-of-flight (PToF) data and was calibrated for ionisation efficiency using 350 nm  
309 monodisperse ammonium nitrate particles. The vapouriser was set at approximately  
310 600 °C and data were collected at a time resolution of 2 min. A collection efficiency  
311 value of unity was applied to these data, based on evidence from a previous chamber  
312 study (Alfarra et al., 2006).

313 Filter samples for offline analysis were collected (without denuders) in a specially  
314 constructed holder, positioned in the chamber vent line. Aerosol samples were  
315 collected onto 47 mm quartz fibre filters (Whatman) at a rapid flow rate of 3  $\text{m}^3 \text{min}^{-1}$   
316 (sample time ca. 6 mins.). After sampling, filters were immediately placed in pre-  
317 cleaned glass vials and stored below -20 °C until analysis. The filter collection  
318 procedure employed here is much faster than traditional filter collection methods,  
319 which should minimise any potential negative and positive artefacts.



320 The filters were extracted into high purity water, filtered, evaporated to dryness and  
321 redissolved in 1 ml 50 % MeOH : 50 % H<sub>2</sub>O. The water-soluble compounds were  
322 analysed using liquid chromatography-ion trap mass spectrometry (LC-MS/MS).  
323 Reverse phase LC separation was achieved using an HP 1100 LC system equipped  
324 with an Eclipse ODS-C<sub>18</sub> column with 5 µm particle size (Agilent, 4.6 mm x  
325 150 mm). Samples (60 µl) were injected then eluted by gradient elution with solvents  
326 A: 0.1% v/v formic acid water (Optima grade, Fisher) and B: methanol (Optima  
327 grade, Fisher) and a gradient program of 3 % B at time 0 min to 100 % B at 60  
328 minutes with a flow rate of 0.6 ml min<sup>-1</sup>. Mass spectrometry analysis was performed  
329 in negative ionisation mode using an HCT-Plus ion trap mass spectrometer with  
330 electrospray ionisation (Bruker Daltonics GmbH). Electrospray ionisation (ESI) was  
331 carried out at 350 °C with a nebuliser pressure of 4.82 bar and a nitrogen drying gas  
332 flow of 12 l min<sup>-1</sup>. Further details can be found in Hamilton et al., 2013.

### 333 ***2.5 Experimental protocol***

334 Three plants were placed in the plant chamber a minimum of 48 hours prior to the  
335 start of the experiment. Both the pots and soil were isolated by enclosing them in  
336 PFTE sheeting; this acted to prevent VOC emissions from the plastic pots and soil  
337 NO<sub>x</sub> emissions from entering the chamber air. Three experiments were carried out on  
338 each species over a one-week period, after which the plants were removed and  
339 replaced with three plants of the next species, and the experiment cycle repeated.

340 Prior to each experiment, ozone was added to the chamber to give a mixing ratio of  
341 approximately 2 ppmV and was left overnight. The chamber was then filled and  
342 flushed several times using clean air from the facility's main inlet system (including  
343 Purafil, charcoal and HEPA filters as described above), until the total particle count  
344 (as measured by a water based condensation particle counter) was below  $10\text{ cm}^{-3}$  and  
345 the  $\text{O}_3$  and  $\text{NO}_x$  levels were less than 1 and 2 ppbV, respectively. At this point, the  
346 reaction chamber was flushed and then connected to the plant chamber for filling with  
347 the plant VOC emissions. Aerosol and gas phase composition and concentrations  
348 were continuously monitored throughout. At the end of the filling process, the plant  
349 chamber was disconnected from the reaction chamber, and within the space of  
350 roughly one minute, both the chamber lights were turned on and pure  $\text{O}_3$  was injected  
351 to provide an initial concentration of around 20 or 70 ppbV (experiment dependent).  
352 The switching on of the chamber lights marked the start of each experiment, which  
353 typically lasted 6 hours from this point. For experiments using pre-existing seed,  
354 polydisperse ammonium sulphate particles (diameter between 40 – 60 nm) were  
355 generated from an aqueous solution using an aerosol nebuliser (Topas, ATM 230) and  
356 injected without drying into the reaction chamber at the end of the filling from the  
357 plant chamber.

358 In our experiments we chose to use ammonium sulphate for the aerosol seeds, rather  
359 than acidic particles that could otherwise promote isoprenoid particulate mass  
360 formation. Whilst it is recognised that isoprenoid SOA mass can be enhanced by the  
361 presence of acidic aerosol seed as originally reported by Jang et al. (2002) and

362 subsequently by Limbeck et al. (2003), Edney et al. (2005), Kleindienst et al. (2007),  
363 Limbeck et al. (2007) and Surratt et al. (2007), we have limited our study to SOA  
364 formation in the mixed precursor systems without deliberate enhancement of particle  
365 mass by condensed phase reaction. There is clear evidence that isoprene oxidation can  
366 contribute to atmospheric SOA formation (e.g. Claeys et al., 2004, Edney et al., 2005)  
367 and we have previously found that enhancement in SOA from isoprene oxidation  
368 above the Bornean rainforest compared with the Amazon may result from an  
369 enhanced marine acidic sulphate contribution to sub-micron aerosol (Robinson et al.,  
370 2011). Intermediates in SOA formation from isoprene have been identified (e.g. Lin  
371 et al., 2012, 2013) and mechanisms for the acid catalysed formation proposed (Surratt  
372 et al., 2010). Whilst out of the scope of the current study, this should be the focus of  
373 future work.

374 Air samples were taken from three separate locations: 1) immediately before the plant  
375 chamber (pre-PC) for blank subtraction, 2) immediately after the plant chamber (post-  
376 PC) during the reaction chamber filling period for directly emitted BVOC and 3) from  
377 the reaction chamber (RC) during the experiment. RC air was monitored continuously  
378 using PTR-MS and CIR-TOF-MS for VOC decay and formation of reaction products.  
379 Air samples from the pre-PC and post-PC position, as well as RC air samples  
380 immediately at the start of each experiment and 1, 2, 4 and 6 hours after the lights  
381 were switched on, were collected on Tenax TA and Carbotrap filled stainless steel  
382 tubes (Supelco Inc, PA, USA) for GC-MS analysis.

383 Relative humidity (%), CO<sub>2</sub> (ppmV), PAR (μmol m<sup>-2</sup> s<sup>-1</sup>), and temperature (°C) in the  
384 plant chamber were recorded every 5 or 10 mins during reaction chamber filling, and  
385 every 15 or 20 mins overnight. System blanks were taken at the start and end of the  
386 experimental period. The reaction chamber background was checked and  
387 characterised through the performance of regular blank experiments (one in every five  
388 experiments). NO<sub>x</sub> (NO, NO<sub>2</sub> and NO<sub>3</sub>) and O<sub>3</sub> were continuously monitored in the  
389 reaction chamber. A list of all experiments and their general parameters is given in  
390 Table 1.

## 391 **2.6 Calculated OH concentrations**

392 Since isoprene losses are controlled by reaction with ozone and the hydroxyl radical  
393 (OH), the concentration of OH available to react with isoprene in the reaction  
394 chamber for each experiment was calculated based on the measured concentrations of  
395 O<sub>3</sub> and isoprene in each experiment, the rate of change in isoprene concentration, and  
396 the rate constants for the reactions of isoprene with OH and O<sub>3</sub>, using equation (1):

$$397 \frac{\frac{d[Isoprene]}{dt} + k_{O_3}[O_3][Isoprene]}{-k_{OH}[Isoprene]} = [OH] \quad \text{Eq}^n (1)$$

398 Hourly averaged concentrations of O<sub>3</sub> and isoprene were calculated for five of the  
399 experiments using the tropical fig. Using these data along with equation (1) a range of  
400 OH concentrations were obtained. For the first hour after lights on, [OH] was  
401 estimated to be 1.9 x 10<sup>5</sup> – 9.5 x 10<sup>5</sup> molecules cm<sup>-3</sup>, whereas, towards the end of the

402 experiment after roughly five hours, values of  $8.1 \times 10^5 - 1.9 \times 10^6$  molecules  $\text{cm}^{-3}$   
403 were obtained. In general, during the tropical fig experiments,  $[\text{OH}]$  estimated from  
404 isoprene and ozone was observed to steadily increase over the duration of the  
405 experiment from 0 – 5 hours after lights on.

## 406 ***2.7 VOC/NO<sub>x</sub> Conditions***

407 Figure 1 shows the time-dependent mixing ratios of ozone and oxides of nitrogen for  
408 each experiment set. Although every effort was made to keep the concentrations of  
409 oxides of nitrogen low, measurable amounts were present, giving initial VOC/NO<sub>x</sub>  
410 ratios of the order 2 – 6 and 3 – 9 (see Table 1), for the birch and fig experiments,  
411 respectively (where here, the VOC concentration is equal to the sum of all potential  
412 precursor concentrations). In terms of a “Sillman plot” (Sillman, 1999), the  
413 experiments were carried out in the “VOC sensitive regime”.

414 The absolute concentration of VOCs in the reaction chamber was roughly ten times  
415 greater than those measured over the rainforest during our field experiments  
416 (Mackenzie et al., 2011) and the VOC/NO<sub>x</sub> ratios employed here were as much as ten  
417 times lower (i.e. typical ratio of 20:1, isoprene:NO<sub>x</sub> over the rainforest) (Hewitt et al.,  
418 2010b). The source of the NO<sub>x</sub> in the reaction chamber (initially  $\sim 2 - 6$  ppbV NO<sub>x</sub>,  
419 but increasing to  $\sim 5 - 9$  ppbV after  $\sim 5$  hours) is attributed to a small amount of  
420 diffusion of outside ambient air across the porous Teflon membrane into the reaction  
421 chamber. The production of certain reactive intermediates in the oxidation of VOCs

422 (e.g., hydroxyl hydroperoxides from isoprene oxidation) is very sensitive to NO<sub>x</sub>  
423 concentrations in the reaction mixture.

## 424 **Results**

### 425 ***3.1 Experiments with *Betula pendula****

#### 426 *3.1.1 Gas phase*

427 Continuous gas phase monitoring with the CIR-TOF-MS and PTR-MS throughout the  
428 experiments, indicated successful transfer of VOC precursor material from the plant  
429 chamber to the reaction chamber prior to lights on. The data indicated that there was  
430 negligible loss of precursor compounds during the chamber transfer process (Fig. 2a).

431 Immediately after initiation of the photochemistry, the VOC precursor concentrations  
432 were observed to decay and product ions began to appear in the CIR and PTR mass  
433 spectra. Approximately sixty product-ion peaks were observed by the CIR-TOF-MS  
434 and the PTR-MS in the organic gas phase during a typical *Betula pendula* experiment.  
435 The temporal profiles of a number of the most abundant (O)VOCs measured are  
436 shown in Fig. 2. From a combination of the CIR-TOF-MS, PTR-MS and GC-MS  
437 observations (and from those observations discussed below for the tropical plant  
438 experiments), over fifty different hemi-, mono- and sesqui-terpene oxidation products  
439 were tentatively identified (Fig. 3 and Tables S1 – S5 in the supplementary  
440 information).

441 From initial inspection of the data, it is clear that monoterpenes dominate during the  
442 *Betula pendula* experiments (Fig. 2a and b), with strong signals observed in the CIR-  
443 TOF-MS and PTR-MS mass spectra at  $m/z$  137 (protonated parent ion) and 81  
444 (hydrocarbon fragment). A small amount of isoprene was also detected during *Betula*  
445 *pendula* experiments; however this was always significantly lower in magnitude than  
446 that of the sum of monoterpenes; for example, during the experiment on 07/07/09,  
447 12.6 ( $\pm$  3.8) ppbV monoterpenes were measured in the reaction chamber prior to  
448 lights on (c.f. 11.4 ppbV total monoterpenes measured at the post-PC position by the  
449 GC-MS), whereas only 2.0 ( $\pm$  1.0) ppbV isoprene was detected. Speciation of the  
450 monoterpenes by GC-MS indicated that the most dominantly emitted C<sub>10</sub> compounds  
451 from *Betula pendula* were  $\alpha$ - and  $\beta$ -pinene (Table 2).

452 C<sub>15</sub> sesquiterpenes (parent ion  $m/z$  205) were detected in the plant and reaction  
453 chambers during each *Betula pendula* experiment, with the most abundant species  
454 identified by GC-MS being  $\beta$ -caryophyllene (Table 2). Sesquiterpenes were also  
455 measured in the reaction chamber by CIR-TOF-MS (Fig. 2c), however for the  
456 majority of the experiments they were present at concentrations either close to or  
457 below the detection limit, hence they could not always be monitored as a function of  
458 reaction time. For the experiment on 07/07/09, 1.7 ( $\pm$  0.9) ppbV sesquiterpenes were  
459 measured by the CIR-TOF-MS prior to lights on (c.f. 2.2 ppbV total sesquiterpenes  
460 measured at the post-PC position by the GC-MS). An ion of  $m/z$  153 was also  
461 observed in the PTR and CIR mass spectra of the plant chamber emissions and

462 subsequently in the reaction chamber air, tentatively assigned (and here after referred  
463 to) as camphore.

464 During the reaction phase of the *Betula pendula* experiments the CIR-TOF-MS mass  
465 spectra were dominated by ions of relatively high mass (i.e.  $m/z > 100$ ) pertaining to  
466 products of both monoterpene and sesquiterpene oxidation. The ions of highest mass  
467 (i.e.  $m/z$  170 – 290) are characteristic of sesquiterpene oxidation, and have been  
468 observed recently during a similar chamber study investigating  $\beta$ -caryophyllene  
469 photo-oxidation (Jenkin et al., 2012). Drawing a comparison between these data and  
470 the detailed  $\beta$ -caryophyllene study conducted by Jenkin et al. (2012), a number of  
471 tentative assignments have been made for  $\beta$ -caryophyllene products, with the  
472 assumption that other precursor specific structural isomers may also occupy the same  
473 mass channels. A full list of example tentative assignments is given in the  
474 supplementary information (Table S2). In total the sum of all sesquiterpene products  
475 measured in the chamber was estimated to be  $\sim 1.5$  ppbV (assuming an average PTR  
476 sensitivity for such high mass, oxygenated, compounds).

477 In contrast to the small amounts of sesquiterpene products observed in the reaction  
478 chamber, the products observed in greatest abundance were those derived from  
479 monoterpene decay. The largest (combined) product signal measured by the CIR-  
480 TOF-MS was that of  $\Sigma(I_{111}, I_{93})$ , where  $I_x$  is the intensity of the mass spectrum at  $m/z$   
481  $= x$  (Fig. 2d). Previously, the  $m/z$  111 and 93 signals have been shown to correspond  
482 to various primary  $C_7$  unsaturated aldehydes formed during the oxidation of



483 unsaturated acyclic monoterpenes, such as myrcene, ocimene and linalool (Lee et al.,  
484 2006a; Lee et al., 2006b; Ng et al., 2006; Wyche et al., In Preparation). In the case of  
485 myrcene and ocimene, the  $m/z$  111 and 93 signals correspond to the parent ion ( $MH^+$ )  
486 and the dehydrated daughter fragment, respectively ( $MH^+-H_2O$ ), and in the case of  
487 linalool  $m/z$  111 corresponds to the dehydrated daughter ion and  $m/z$  93 is a further  
488 fragment. The concomitant  $m/z$  111 and 93 signals have also been reported to result  
489 from a  $C_7$  cyclic ketone formed during the oxidation of terpinolene (not found in the  
490 *Ficus* emission profile and  $< 1$  ppbV found in the *Betula* profile). The  $m/z$  111 and 93  
491 ions have previously been observed to be significant contributors to total ion signal in  
492 the PTR mass spectra during single precursor chamber experiments with concomitant  
493 SOA formation (Lee et al., 2006a; Lee et al., 2006b; Ng et al., 2006; Wyche et al., In  
494 Preparation), and the  $m/z$  111 ion has also been observed in ambient air measurements  
495 over a forested region (Holtzinger et al., 2005).

496 As can be seen from observation of Fig. 2(d), the  $\Sigma(I_{111}, I_{93})$  signal rises rapidly  
497 during the initial stages of the experiment, much more so than other monoterpene  
498 oxidation products (c.f. Fig. 2e), suggesting that the precursor has a much shorter  
499 lifetime with respect to OH and  $O_3$ . Of those monoterpenes speciated by the GC-MS,  
500 ocimene and linalool have the shortest lifetimes, with  $k_{OH} = 3.04$  and  $1.6 \times 10^{-10}$   $cm^3$   
501  $molecule^{-1} s^{-1}$  (average lifetimes with respect to OH  $\sim 44$  and  $55$  mins.), respectively,  
502 compared to  $k_{OH} = 7.4 \times 10^{-11}$   $cm^3 molecule^{-1} s^{-1}$  for  $\beta$ -pinene (average lifetime with  
503 respect to OH  $\sim 1458$  mins.) (Atkinson and Arey, 2003; Kim et al., 2011). The  $\Sigma(I_{111},$

504  $I_{93}$ ) signal peaks at around 60 – 100 mins at 3.0 ( $\pm$  0.7) ppbV (concentration estimated  
505 using pinonaldehyde sensitivity), before decaying at a greater rate than that of the  
506 precursor monoterpenes and the other monoterpene products. This relatively short  
507 lifetime gives further insight into the potential identity of the  $m/z$  111 and 93 signals,  
508 perhaps indicating the presence of multiple C=C bonds in the hydrocarbon structure,  
509 as would be found in the primary  $C_7$  aldehydes obtained from the oxidation of  
510 ocimene or myrcene for example. Other short-lived biogenic oxidation products that  
511 could exist in such mesocosm systems, include  $\alpha$ -hydroxy carbonyls, similarly formed  
512 following OH addition to a C=C bond.

513 Other dominant signals observed by the PTR-MS and CIR-TOF-MS during oxidation  
514 of the *Betula pendula* air matrix, include the sum of  $m/z$  169 + 151 + 107, which  
515 respectively correspond to the parent ion and two daughter fragments of a number of  
516 primary monoterpene keto-aldehydes (which, from the speciated monoterpene plant  
517 chamber data, are most likely to be pinonaldehyde, caronaldehyde and  $\alpha/\gamma$ -  
518 terpinaldehyde); and  $m/z$  139, corresponding to the parent ion of a number of primary  
519 monoterpene ketones (most likely to be nopinone and caronone, again when  
520 considering the monoterpenes speciated by the GC-MS). As shown in Fig. 2(e) the  
521 primary keto-aldehyde and ketone signals had similar temporal profiles to one  
522 another, growing at a slower rate than that of  $\Sigma(I_{111}, I_{93})$ , to peak concentrations of  
523 around 0.9 ( $\pm$  0.3) and 1.2 ( $\pm$  0.3) ppbV, respectively, as the monoterpene trace  
524 tended towards zero. The temporal profile for the sum of all other “monoterpene  
525 like” product ions (i.e. ions of  $m/z > 90$ ) was very similar to those of the primary keto-  
26

526 aldehyde(s) and ketone(s), peaking at a combined mixing ratio of approximately 3.5  
527 ppbV (assuming an average PTR sensitivity for such high mass, oxygenated,  
528 compounds).

529 During the oxidation of compounds emitted by *Betula pendula*, the primary isoprene  
530 products, methyl vinyl ketone (MVK) and methacrolein (MACR) (measured together  
531 at  $m/z$  71) were observed to evolve in the same manner as the primary monoterpene  
532 keto-aldehyde(s) and ketone(s), peaking at an approximate mixing ratio of  $0.4 (\pm 0.1)$   
533 ppbV (Fig. 2e). A series of lower  $m/z$  ions were also observed to evolve within the  
534 reaction chamber, including  $m/z$  61 (acetic acid), 59 (acetone), 47 (formic acid), 45  
535 (acetaldehyde), 33 (methanol) and 31 (formaldehyde). Each of these compounds has  
536 previously been associated with monoterpene oxidation and/or with off-gassing from  
537 illuminated chamber walls. Methanol, acetone and  $m/z$  99 (potentially *cis*-3-hexenal)  
538 were also observed within the reaction chamber prior to lights on, with a combined  
539 mixing ratio of approximately 20 ppbV.

### 540 3.1.2 Particle phase

541 From inspection of the CPC and SMPS data we see that SOA mass formed during  
542 oxidation of the *Betula pendula* air matrix. As can be seen from Fig. 4, during  
543 unseeded experiments nucleation occurred immediately after lights on, with no  
544 induction period prior to mass formation. After nucleation, SOA mass increased  
545 rapidly to  $\sim 11 \mu\text{g m}^{-3}$  by  $\sim 40$  minutes (experiment 06/07/09), followed by a relatively  
546 stable plateau (after the application of wall loss corrections) and a slight increase

547 towards the end of the experiment. In order to suppress nucleation, seed particles were  
548 introduced in some experiments, as has been used previously (Dommen et al.,  
549 2009;Meyer et al., 2009;Surratt et al., 2007;Kleindienst et al., 2006;Carlton et al.,  
550 2009). This more closely represents the conditions encountered in the ambient  
551 atmosphere where there is pre-existing aerosol. Consistent with the nucleation  
552 experiments described above, SOA mass was observed to increase as soon as the  
553 photochemistry was initiated when an ammonium sulphate seed was present (Fig. 4,  
554 experiment 07/07/09).

555 Using the wall-loss-corrected mass data, along with the corresponding quantity of the  
556 sum of precursor species reacted and equation (2), SOA yields were obtained for the  
557 *Betula pendula* oxidation system:

$$558 \quad Y_{SOA} = \frac{M_p}{\Delta(\Sigma VOC)} \quad \text{Eq}^n (2)$$

559 In this instance,  $Y_{SOA}$  = SOA mass yield,  $M_p$  = peak SOA mass ( $\mu\text{g m}^{-3}$ ) and  $\Delta(\Sigma VOC)$   
560 = the sum of gas phase precursors reacted by the time  $M_p$  is reached ( $\mu\text{g m}^{-3}$ ) (Odum  
561 et al., 1997). In order to determine  $\Delta(\Sigma VOC)$ , the time-dependent VOC mixing ratios  
562 for total sesquiterpenes, total monoterpenes, camphore and isoprene were  
563 independently converted to their corresponding mass concentrations ( $\mu\text{g m}^{-3}$ ) and the  
564 four data sets were combined to give a “total” VOC precursor decay profile. From  
565 the total VOC profile,  $\Delta(\Sigma VOC)$  was calculated, using the starting mass of  $\Sigma VOC$  at  
566 time = 0 and the mass of  $\Sigma VOC$  at the time of  $M_p$ . The uncertainty in  $\Delta(\Sigma VOC)$  is

567 estimated to be  $\pm 41$  %. Using equation (2) for the two *Betula pendula* experiments  
568 for which both gas phase mixing ratio and wall-loss-corrected aerosol data were  
569 available, SOA yield values of 39 and 26 % were obtained (Fig. 4). It should also be  
570 noted that along with previous caveats made regarding the role of the chamber walls  
571 and other measurement uncertainties, these yield values also do not take into account  
572 the potential loss of particularly “sticky” low volatility compounds (e.g. Ehn et al.,  
573 2014) to internal surfaces of the chamber.

574 It should be noted in the above yield calculations that the partitioning of material  
575 between the vapour phase and chamber walls has not been taken into account.  
576 Matsunaga and Ziemann (2010) showed that semi-volatile organic compounds move  
577 towards equilibrium between the walls and the vapour phase and that the equilibration  
578 timescale and equivalent absorptive mass of the walls was dependent on the  
579 molecular properties of the partitioning species. Kokkola et al. (2013) demonstrated in  
580 their model study that OVOC wall losses will have significant implications on their  
581 partitioning between the gas and particle phase, such that the mass components of  
582 very low volatility will almost completely be depleted to the chamber walls during the  
583 experiment while the depletion of OVOCs of higher volatilities is less efficient. The  
584 implications of such partitioning to chamber walls are such that comparison between  
585 any yields determined experimentally in different chambers should be conducted with  
586 caution. Even when calculated from experiments in the same chamber, yields should  
587 be interpreted qualitatively and relatively and not extrapolated to the atmosphere.

## 588 **3.2 Experiments with tropical species**

### 589 *3.2.1 Gas phase*

590 In order to study the contrast between species that primarily emit monoterpenes and  
591 those that primarily emit isoprene, and hence to better understand the isoprene-SOA  
592 system, the coupled plant chamber-reaction chamber system was employed to study  
593 several tropical species. Two species of fig and one species of palm were selected  
594 during the pre-experiment screening process. Those experiments using the figs, *Ficus*  
595 *cyathistipula* and *Ficus benjamina* gave the most complete data set; hence their  
596 results are used as a focus for discussion.

597 Fig. 5 shows the temporal evolution of a number isoprenoids detected in both the  
598 plant and reaction chambers (a and b) and the concomitant evolution of a selection of  
599 isoprenoid oxidation products (c and d), during a typical *Ficus benjamina* experiment  
600 (23/06/09). During a typical *Ficus benjamina* experiment, approximately thirty  
601 precursor and product-ion peaks were observed by the CIR-TOF-MS and PTR-MS in  
602 the gas phase. Tentative product identification is reported in the supplementary  
603 information (Fig. 6, Tables S1 – S5). A similar set of ions was observed during a  
604 typical *Ficus cyathistipula* experiment.

605 From inspection of Fig. 5(a) and (b), the dominance of isoprene in the *Ficus*  
606 *benjamina* system is clear, with 12.3 ( $\pm$  4.1) ppbV isoprene detected in the reaction  
607 chamber at lights on, compared to 0.8 ( $\pm$  0.4) ppbV monoterpenes, 0.5 ( $\pm$  0.9) ppbV

608 sesquiterpenes and an estimated 2.7 ( $\pm$  0.6) ppbV camphore. Speciation of the  
609 monoterpenes by GC-MS indicated that the most dominantly emitted C<sub>10</sub> compounds  
610 for *Ficus benjamina* were  $\alpha$ -pinene, limonene, sabinene and linalool and for *Ficus*  
611 *cyathistipula* were  $\alpha$ -pinene,  $\beta$ -pinene and limonene (Table 2). The sequiterpenes,  $\beta$ -  
612 caryophyllene and  $\alpha$ -cubebene were also identified.

613 Products of isoprene were observed to dominate the evolving *Ficus benjamina* and  
614 *Ficus cyathistipula* oxidation systems, with the isobaric primary species MACR and  
615 MVK comprising the strongest signals (measured together at  $m/z$  71). For example,  
616 during the *Ficus benjamina* experiment of 23/06/09, a combined peak MACR + MVK  
617 mixing ratio of 2.9 ( $\pm$  0.7) ppbV was observed (Fig. 5c).

618 Along with MACR and MVK, a series of other ions also associated with isoprene  
619 oxidation were detected during *Ficus benjamina* oxidation, including  $m/z$  117 and 99  
620 (4-hydroxy-2-methyl-but-2-enoic acid), 103 (C5-alkenediols, C4-hydroxydialdehydes  
621 and MPAN), 87 (C4-hydroxycarbonyls and methacrylic acid), 83 (3-methyl furan), 75  
622 (hydroxy acetone) and 31 (formaldehyde). Additionally, a signal of  $m/z$  101 was also  
623 measured, possibly corresponding to the sum of a series of C5-hydroxycarbonyls and  
624 C5-hydroxy hydroperoxides (Tuazon and Atkinson, 1990; Paulson and Seinfeld,  
625 1992; Jenkin *et al.*, 1997; Benkelberg *et al.*, 2000; Sprengnether *et al.*, 2002,  
626 Benkelberg *et al.*, 2000; Zhao *et al.*, 2004; Surratt *et al.*, 2006;  
627 <http://mcm.leeds.ac.uk/MCM>, v3.1). Of the signals observed, those of  $m/z$  83 and 87  
628 (tentatively assigned to be 3-methyl furan and C4-hydroxycarbonyls/methacrylic

629 acid), were the greatest in magnitude after MACR + MVK (Fig. 6). The temporal  
630 evolution of the sum of all of these products suggests that they are predominantly  
631 secondary in nature, forming in the chamber after MACR and MVK. They continued  
632 to increase in magnitude as the isoprene signal decreased and as the MACR + MVK  
633 signal began to fall (Figs. 5 and 6). During a typical *Ficus benjamina* experiment, the  
634 sum of these isoprene products was estimated to reach a peak mixing ratio of ~ 1.7  
635 ppbV.

636 A series of lower molecular weight ions were also observed to evolve within the  
637 reaction chamber, including  $m/z$  61 (acetic acid), 47 (formic acid), 45 (acetaldehyde),  
638 33 (methanol) and 31 (formaldehyde). Each of these compounds has previously been  
639 associated with isoprene oxidation and/or with off-gassing from illuminated chamber  
640 walls. The  $m/z$  43 and 46 signals, indicative of carbonyls and nitrates, respectively,  
641 were also observed to increase significantly during photo-oxidation, indicating the  
642 formation and evolution of such species with increasing experiment duration.

643 Besides ions pertaining to the oxidation products of isoprene, a number of spectral  
644 features typically derived from monoterpene oxidation products were also observed to  
645 form and evolve in the reaction chamber, including,  $m/z$  151, 125, 109, 107, 93 and  
646 91. To a first order approximation, the total peak quantity of oxidation products not  
647 believed to result from isoprene decay was estimated to be of the order 2 ppbV.  
648 However, it should be noted that the presence of isobaric interference in such a  
649 complex system, uncharacterised fragmentation, detection limits and the use of



650 pseudo and averaged calibration sensitivities, impose a certain level of unknown  
651 uncertainty upon this final value.

### 652 3.2.2 Particle phase

653 Contrary to the immediate and abundant formation of new particles in the un-seeded  
654 *Betula pendula* experiments, the total number of particles and total aerosol mass did  
655 not increase above background levels after lights were turned on in the un-seeded  
656 tropical *Ficus benjamina* experiments. Figure 7 shows the observed and wall-loss-  
657 corrected particle mass concentration during two typical *Ficus benjamina* experiments  
658 (22/06/09 and 23/06/09) along with a chamber background experiment. Owing to a  
659 lack of particle nucleation in those experiments, the total particle number  
660 concentration was too low for the wall loss correction (described in section 2.3) to be  
661 implemented. Instead, the average of the wall loss constants determined for the  
662 seeded experiments was used to calculate the wall loss corrected mass concentrations  
663 reported in Figure 7.

664 Figure 8 shows the observed and wall-loss-corrected particle mass concentration for  
665 ammonium sulphate seeded experiments using VOC emissions of *Ficus benjamina*  
666 and *Ficus cyathistipula*, as well as a seeded background experiment. The mass at the  
667 start of the experiment represents the initial ammonium sulphate mass. In order to  
668 quantify the formation of SOA mass during these experiments, the mass increase  
669 relative to the starting seed mass was determined in Figure 9 by subtracting the initial  
670 ammonium sulphate seed mass from the total wall loss corrected mass. The same

671 calculation was also performed for the *Betula pendula* seeded experiment (07/07/09).  
672 In contrast to the unseeded *Ficus benjamina* and *Ficus Cyathistipula* experiments,  
673 SOA mass was observed to form when a seed was present in the reaction chamber.  
674 The calculated SOA traces in Figure 9 illustrate a slower build-up of mass during the  
675 isoprene dominated *Ficus benjamina* (15/07/09) and *Ficus Cyathistipula* (30/06/09  
676 and 02/07/09) experiments compared to the much faster SOA mass formation in the  
677 monoterpene dominated *Betula pendula* experiment. Peak masses of the order 1.3 –  
678 5.5  $\mu\text{g m}^{-3}$  were observed, which when employed with the methodology described in  
679 section 3.1.2, produce SOA yields of 10 and 14 % for each of the two *Ficus*  
680 *Cyathistipula* experiments for which both gas and wall-loss-corrected aerosol data  
681 were available, i.e. 30/06/09 and 02/07/09, respectively (Fig. 9). Uncertainty in  
682  $\Delta(\Sigma\text{VOC})$  is estimated to be  $\pm 47\%$  and in the size distribution measurements used in  
683 the wall loss calculations, of the order of  $\pm 2\%$ . The uncertainties in the wall-loss  
684 correction will likely be substantially greater, but remain unquantified at present.

## 685 **Discussion and Conclusions**

### 686 **4.1 *Betula pendula***

687 In this study we coupled a plant chamber to a photochemical reaction chamber in  
688 order to investigate secondary organic aerosol production from a biogenically  
689 consistent mixture of biogenic volatile organic compounds. We studied silver birch

690 (*Betula pendula*), which emits predominantly monoterpenes, with some  
691 sesquiterpenes and oxygenated VOCs but only trace isoprene (Table 2, Fig 3).

692 Our *Betula pendula* experiments showed significant SOA formation (Fig. 4) both in  
693 the presence and absence of an ammonium sulphate seed, and reproduced the rate of  
694 production and growth of SOA observed in earlier published studies (Mentel et al.,  
695 2009;Carlton et al., 2009;e.g. VanReken et al., 2006;Hallquist et al., 2009;Kiendler-  
696 Scharr et al., 2009a;Kiendler-Scharr et al., 2009b).

697 The SOA yield values of 39 and 26 % obtained here for *Betula pendula* compare  
698 reasonably well with those reported within the literature for single precursor work  
699 conducted under similar conditions. For instance, for the two most abundant  
700 monoterpenes emitted by *Betula pendula*, i.e.  $\alpha$ -pinene and  $\beta$ -pinene, single  
701 precursors yields of the order 1 – 43 (16) and 3 – 30 %, respectively, have been  
702 observed (values given in parenthesis were obtained from the Manchester aerosol  
703 chamber). Similarly for other common monoterpenes such as limonene, myrcene,  $\Delta^3$ -  
704 carene and  $\alpha$ -terpenine, SOA yields of 9 – 34, 6 – 43 (15), 2 – 38 and 8 – 25 %,  
705 respectively, have been reported and for  $\beta$ -caryophyllene, 37 – 79 (50) % (Lee at al.,  
706 2006a and references therein; Alfarra at al., 2012). In a comparable study to ours,  
707 Mentel et al., (2009) reported a fractional mass yield of 11 % for their *Betula pendula*  
708 experiments, i.e. slightly lower than those given here, but within the bounds of  
709 quantified experimental errors. The yield values obtained here for the *Betula pendula*  
710 mesocosm system lie roughly in the middle of the single precursor yield range.

711

712 As can be seen from inspection of Fig. 3, the transfer of mass through the *Betula*  
713 *pendula* experiment appeared roughly conservative, with a small and steady loss of  
714 measured mass from the reaction matrix after  $\sim 220$  min. With the addition of  
715 oxygen to the starting body of hydrocarbon material during such an experiment, the  
716 total measured mass (i.e.  $\Sigma$ VOCs + SOA) within the system would be expected to  
717 increase with time. The absence of such a total measured mass gain (and indeed the  
718 mass deficit observed towards the end of the experiment), can most likely be  
719 accounted for by considering the various measurement uncertainties involved in  
720 producing these data (e.g. assumptions in PTR sensitivity, uncharacterised  
721 fragmentation following ionisation, instrument detection limits, etc.) and influences  
722 imposed by the chamber walls (including potential loss of more highly oxidised  
723 material from the gas phase and greater than expected loss of SOA). Indeed, there is  
724 potential for a system mass increase by the end of the experiment to lie within the  
725 uncertainty bounds of the CIR-TOF-MS/PTR-MS measurements alone, e.g. the  
726 average, single compound PTR measurement uncertainty is  $\sim \pm 30\%$ , allowing the  
727 final measured value of  $130 \mu\text{g m}^{-3}$  to have an upper limit of  $170 \mu\text{g m}^{-3}$ , i.e. greater  
728 than the starting value. Considering these results it seems that the system studied is  
729 reasonably well characterised given the complications involved in such a task.

730 Certain insights into the mechanisms of SOA formation and growth during the *Betula*  
731 *pendula* experiments can be obtained through a combined examination of the VOC

732 data, the time-dependent growth curves (Fig. 10) and the aerosol composition data.  
733 The data in Fig. 10 demonstrates that during oxidation of the *Betula pendula*  
734 emissions and in absence of a seed, SOA mass evolution can be roughly split into two  
735 phases. In the early stages of the experiment after nucleation, SOA mass growth  
736 increased somewhat rapidly with respect to the amount of precursors reacted;  
737 however, after roughly 30 % of the initial precursor mass had been consumed, the rate  
738 of mass growth with respect to VOC precursor consumption was observed to reach an  
739 approximate steady state. When ammonium sulphate seed was present within the  
740 chamber, there was a similarly rapid initial growth with respect to VOC consumption,  
741 however this time subsequent aerosol evolution was characterised by a roughly linear  
742 mass increase to a much higher final mass by the end of the experiment. Considering  
743 the various species of precursor VOCs detected in the *Betula pendula* plant chamber  
744 air, and the relative lifetimes of these VOCs, it would seem possible that initially the  
745 shorter-lived sesquiterpenes react to form a significant proportion of the high mass,  
746 nucleating/condensing species, before being removed from the system (e.g. Jenkin et  
747 al., 2012). Subsequent aerosol mass formation as the air in the reaction chamber ages  
748 towards the central phase of the experiment, is then likely to result from the  
749 partitioning of relatively more volatile products formed from slower reacting  
750 monoterpenes, e.g. products such as pinic and pinonic acid from  $\alpha$ -pinene oxidation  
751 (e.g. Jenkin, 2004; Camredon et al., 2010), and products such as the primary acyclic  
752 unsaturated aldehydes, ( $m/z$  111 + 93); the temporal profile of which demonstrates

753 gas-phase loss concomitant with wall-loss-corrected aerosol growth reaching a steady  
754 state.

755 Further understanding of the composition and evolution of SOA typical of temperate  
756 plant environs comes from investigation of Figure 11, which shows the evolution of  
757 the fraction of the ratio of more/less oxygenated material present in the aerosol during  
758 the initial stages of a typical unseeded *Betula pendula* experiment (06/07/09). Figure  
759 11 was constructed using the ratio of  $m/z$  44 to 43 obtained from the AMS (*i.e.*  $f_{44/43}$ ,  
760 where  $m/z$  44 is derived from “more” oxidised material and  $m/z$  43 from “less”  
761 oxidised material, Ng et al., 2010). In this instance, the  $f_{44/43}$  ratio exhibits linear  
762 growth with time, from a value of  $\sim 0.8$  to  $\sim 1.1$ , suggesting an increase in the  
763 oxygenated content of the aerosol as the experiment ages. Such an increase in  
764 oxygenated content is generally observed when precursor species contain multiple  
765 C=C bonds (e.g. ocimene and myrcene), offering significant potential for higher  
766 aerosol O:C composition (e.g. perhaps species such as acyclic unsaturated aldehydes  
767 and their subsequent generations of products). Indeed, the evolution of the  $f_{44/43}$   
768 ratio observed here is consistent with those results obtained from single precursor  
769 experiments investigating the acyclic monoterpene, myrcene and the sesquiterpene,  $\beta$ -  
770 caryophyllene (Alfarra et al., 2012; Alfarra et al., 2013).

771 Off-line compositional analysis of the SOA collected at the end of the *Betula pendula*  
772 experiments supports the findings obtained from the on-line gas phase and bulk  
773 aerosol composition data. The LC-MS<sup>2</sup> analysis produced chromatograms with peaks

774 matching those seen in comparable single precursor BVOC experiments, with tracer  
775 compounds of both sesquiterpene and monoterpene SOA, detected (Fig. 12).  
776 Amongst the compounds observed were those of molecular weight (MW) 238, 242,  
777 254 and 256, corresponding to 4-(3,3-dimethyl-2-(3-oxopropyl)cyclobutyl)pent-4-  
778 enoic acid, 3-(3,3-dimethyl-2-(3-oxobutyl)cyclobutyl)-3-hydroxypropanoic acid,  $\beta$ -  
779 nocaryophyllonic acid/ $\beta$ -caryophyllinic acid and  $\beta$ -nocaryophyllinic acid,  
780 respectively, produced during  $\beta$ -caryophyllene oxidation (e.g. Alfarra et al., 2013)  
781 and MW 184, corresponding to cis-pinonic acid, produced during  $\alpha$ -pinene oxidation.  
782 Many other terpene oxidation tracers were detected, including compounds of  
783 molecular weight 118 and 200, which are also prevalent in SOA obtained from single  
784 precursor myrcene experiments. This observation is in line with findings obtained  
785 from the gas phase data, which suggest a relatively significant presence of gaseous  
786 acyclic unsaturated aldehydes that would originate from acyclic unsaturated terpenes,  
787 such as myrcene.

#### 788 **4.2 Tropical species**

789 In addition to *Betula pendula*, we studied three tropical plant species: two figs (*Ficus*  
790 *benjamina* and *Ficus cyathistipula*) and one palm (*Caryota millis*); in this work, we  
791 focus on results obtained from the fig plants. All three tropical species were found to  
792 be strong isoprene emitters, with very much smaller emissions of monoterpenes,  
793 sesquiterpenes, and oxygenated VOCs (Table 2, Fig. 6).

794 During the tropical plant experiments, the primary gas-phase isoprene oxidation  
795 products MACR, MVK, formaldehyde, isoprene hydroxy hydroperoxides and the  
796 secondary product hydroxyacetone were all observed (e.g. Figs. 5 and 6 and Tables  
797 S4 and S5 in the supplementary information). MACR, the isoprene hydroxy  
798 hydroperoxides (isoprene epoxide (IEPOX) and isoprene hydroperoxide (ISOPOOH))  
799 and hydroxyacetone are all believed to be precursors to SOA formation (Jaoui et al.,  
800 2010;Carlton et al., 2009;Kleindienst et al., 2009;Paulot et al., 2009;Kleindienst et al.,  
801 2007;Lee et al., 2006;Kroll et al., 2006;Surratt et al., 2006;Claeys et al.,  
802 2004b;Rollins et al., 2009;Robinson et al., 2010). In this study, with the exception of  
803 MACR and MVK, these products all formed at yields lower than those previously  
804 reported (Table 3), with MACR + MVK, hydroperoxides, hydroxy acetone and  
805 formaldehyde being observed to form in yields of 17 – 36, 1 – 3, 0 – 2 and 2 – 7 %,  
806 respectively during our work. This disagreement may result from differences in OH  
807 concentrations and NO<sub>x</sub> concentrations in each of the experimental studies. Other  
808 isoprene products tentatively identified from the CIR-TOF-MS and PTR-MS data  
809 include, C5-alkenediols, C4-hydroxycarbonyls/methacrylic acid and 3-methyl furan  
810 (Table S4), which have also previously been associated with SOA formation (e.g.  
811 Claeys et al., 2004; Surratt et al., 2006; Robinson et al., 2010). For a typical *Ficus*  
812 *benjamina* experiment (23/06/09) the sum of these and other potential isoprene  
813 products, excluding MACR + MVK, was estimated to have a combined gas phase  
814 yield of the order 18 % (Fig. 5).



815 As can be seen in Fig. 6 mass transfer through the *Ficus benjamina* system was  
816 characterised by a slight mass decrease just after the start of the experiment followed  
817 by a gradual increase in mass with time. As was stated in section 4.1 a mass increase  
818 is expected with time during such an experiment, owing to the addition of oxygen to  
819 the precursor hydrocarbon material. Consequently, when considering the data  
820 presented in Fig. 6 in the context of potential uncertainties involved (including  
821 difficult to characterise influences imposed by the chamber walls), it appears that the  
822 system being studied is reasonably well characterised.

823 By comparing Figs. 3 and 6 we see that the monoterpene dominated *Betula pendula*  
824 system, which produces larger and lower vapour pressure oxidation products than the  
825 isoprene dominated *Ficus* system, as well as measureable SOA, is the case which  
826 exhibits measured mass loss. From this contrast it is reasonable to assume a  
827 significant fraction of any mass deficit observed during *Betula pendula* oxidation  
828 could result from the loss of the heavier, lower volatility compounds that are present  
829 in the *Betula pendula* oxidation system but not in the *Ficus* system.

830 Despite the detection of a number of first- and second-generation gas phase products  
831 that have previously been directly linked with isoprene SOA composition (Claeys et  
832 al., 2004; Wang et al., 2004; Edney et al., 2005; Surratt et al., 2006; Healy et al.,  
833 2008), there was no accompanying evidence of SOA formation from the isoprene-  
834 emitting tropical plants during unseeded, nucleation style experiments (Fig. 7). A lack  
835 of SOA mass formation during our unseeded *Ficus benjamina* experiments could  
836 have resulted from a number of different factors, not least of which was simply the

837 absence of a seed surface (acidic or otherwise) to help facilitate partitioning of the  
838 semi-volatile oxidation products to the aerosol phase and produce particles of  
839 sufficient size and measureable particle mass (e.g. Kroll et al., 2006). Another  
840 potentially significant contributing factor in suppressing SOA formation during these  
841 experiments was our relatively low VOC/NO<sub>x</sub> ratio and the resultant gas phase  
842 chemistry. In the presence of high (i.e. ppbV-level) NO<sub>x</sub> mixing ratios, RO<sub>2</sub> radicals  
843 react with NO to produce mainly alkoxy (RO) radicals. For low molecular mass  
844 VOCs such as isoprene, these RO radicals generally fragment into smaller, more  
845 volatile products that do not easily partition from the gas phase to the aerosol phase,  
846 resulting in a low SOA yield (Surratt *et al.*, 2010). Conversely, under low NO<sub>x</sub>  
847 conditions, RO<sub>2</sub> radicals are known to undergo self- and cross-reactions to produce  
848 organic peroxides and hydroperoxides of relatively low volatility. For example,  
849 Surratt *et al.* (2010) showed that under high NO<sub>x</sub> conditions the yield of the  
850 potentially SOA forming gas phase IEPOX was reduced with respect to the equivalent  
851 value under low NO<sub>x</sub> conditions, where IEPOX formed in substantial yields (upward  
852 of 75 %) from the further oxidation of ISOPOOH by OH.

853 In contrast to our unseeded *Ficus* experiments, when an ammonium sulphate seed was  
854 present (and following wall loss correction), SOA mass was observed to form and  
855 evolve within the reaction chamber (Fig. 9). From estimates of the total  
856 concentrations of precursor VOCs within the reaction chamber matrix (primarily  
857 isoprene, e.g. Fig. 5), an SOA mass yield of the order 10 – 14 % was obtained for the  
858 *Ficus Cyathistpula* system. If it were to be assumed that the SOA were solely formed

859 from oxidation products of isoprene as the major emitted VOC, this yield would  
860 appear excessive in comparison with those obtained previously from single precursor  
861 isoprene studies, i.e.  $\sim 0.1 - 5.5 \%$  (van Donkelaar et al., 2007; Kleindienst et al.,  
862 2009, 2007; Kroll et al., 2005, 2006; Claeys et al., 2004a; Edney et al., 2005 (van  
863 Donkelaar et al., 2007; Kleindienst et al., 2009, 2007; Kroll et al., 2005, 2006; Claeys et  
864 al., 2004a; Edney et al., 2005; Brégonzio-Rozier et al., 2014). However, we must  
865 consider that the mesocosm system is in fact an ensemble of precursors, albeit an  
866 ensemble dominated by isoprene, analogous to ambient air above a tropical forested  
867 region (Hewitt et al., 2010; MacKenzie et al., 2011).

868 For the experiments of 30/06/09 and 02/07/09, for which SOA yields were obtained  
869 for the *Ficus Cyathistipula* system, a fraction of camphore was also observed in the air  
870 entering the reaction chamber (presumably for these two particular experiments,  
871 sesquiterpenes and monoterpenes were present at concentrations below the detection  
872 limits of the PTR-MS and CIR-TOF-MS). The concentration of camphore at lights on  
873 was estimated to be  $\sim 0.5 - 0.9$  ppbV for the *Ficus cyathistipula* system and  $\sim 1.4 -$   
874  $2.7$  ppbV for *Ficus benjamina*, and the sum of all *non-precursor* ions in the CIR-  
875 TOF-MS mass spectrum  $> m/z$  100 (indicative of *non-isoprene-like* oxidation  
876 products; excluding  $m/z$  103 and 117) was estimated to be of the order 2 ppbV by the  
877 end of the experiments. This calculation approximates the  $m/z > 100$  summation as  
878 one large, multifunctional analyte with a PTR sensitivity similar to pinonaldehyde (a  
879 typical, multifunctional, high MW molecule resulting from terpene oxidation).  
880 Continuing this assumption and taking a range of known VOC terpene product yields

881 ( $Y^P_{VOC}$ ) obtained from previous work at the Manchester chamber (i.e.  $Y^P_{VOC} = 100\%$   
882 the limiting case; 77 % from *Ficus Benjamina* oxidation; 55 % from *Betula Pendula*  
883 oxidation; and 29 % for pinonaldehyde and  $\Sigma(I_{111}, I_{93})$ ), a *non-isoprene*  $VOC_{precursor}$   
884 concentration may be estimated. Taking a range of known SOA yields obtained from  
885 the same reaction chamber (i.e.  $\alpha$ -pinene, myrcene, linalool and  $\beta$ -caryophyllene;  
886 Alfarra et al., 2013), the SOA yield obtained here for the *Betula pendula* system and  
887 the estimate of  $[VOC_{precursor}]$ , Eqn. (2) may be solved to provide a crude estimate of  
888 the mass of SOA formed from *non-isoprene* precursors. Consequently, an estimate of  
889 the residual SOA mass derived from isoprene oxidation within the *Ficus* system can  
890 be inferred for each of the experiments shown in Fig. 9.

891 For 78 of the 120 measurement-and-parameter sets tested, the estimated residual SOA  
892 mass resulting solely from isoprene oxidation was negative – i.e., production of SOA  
893 from isoprene oxidation was not required to close the mass balance. Values were  
894 calculated based on the widest range of peak masses observed during the *ficus*  
895 experiments ( $M_p = 1.3 \mu\text{g m}^{-3}$  and  $5.5 \mu\text{g m}^{-3}$ ), and assume the lowest (29 %) and  
896 highest (100 %) VOC terpene yields and lowest (5 %) and highest (47 %) SOA yields  
897 from non-isoprene precursors, respectively, as observed in previous experiments  
898 conducted within this chamber. These ranges result in calculated residual SOA mass  
899 of -28.5 to +5.0  $\mu\text{g m}^{-3}$  produced solely from isoprene oxidation. Hence, there are  
900 combinations of measurements, observations and oxidation/phase-change parameters  
901 — omitting isoprene and its oxidation products — that can account for ~20 times the  
902 observed aerosol mass production, and other combinations of measurements and

903 parameters that leave up to ~ 90 % of the condensed mass to be explained by isoprene  
904 oxidation. If, instead of using the limiting cases, the closest approximation to the *ficus*  
905 *cyathistipula* system is used (i.e.  $Y_{\text{VOC}} = 77\%$  and  $Y_{\text{SOA}} = Y_{\text{SOA}} = \alpha\text{-pinene} = 15\%$ ),  
906 non-isoprene products could have accounted for around 145 % of the SOA mass that  
907 was produced. We have no way of assigning formal likelihoods to each set of  
908 measurements and parameters in this exercise, but we note that the great  
909 preponderance of parameter combinations do not require an isoprene contribution to  
910 the SOA mass (i.e. 78/120 measurement-and-parameter sets tested) under our  
911 experimental conditions. Moreover, our experiments produce much less SOA mass  
912 than would be expected from published experiments using individual mono- and  
913 sesquiterpenes.

914 There are three principal reasons why the estimates of aerosol production from  
915 isoprene in the tropical plant experiments span such a large range. Firstly, the plants  
916 in the mesocosm emit a complicated mixture of biogenic VOCs, some of which are  
917 known to oxidise much more rapidly than isoprene and which will produce  
918 condensable compounds when oxidised. Secondly, these minor compounds co-  
919 emitted from principally-isoprene emitting tropical trees are imperfectly quantified  
920 because of the sensitivity of the chemical ionisation (PTR and CIR) instruments.  
921 Thirdly, these minor co-emissions are imperfectly characterised because many higher  
922 molecular weight compounds, such as the mono- and sesquiterpenes, are isobaric in  
923 the PTR and CIR instruments and so precise chemical structures cannot easily be  
924 assigned. Without better instrument detection sensitivity and high time resolution

925 chemical identification for the reactive compounds co-emitted with isoprene, it is not  
926 possible to constrain further the aerosol yield from the tropical plants. Unfortunately,  
927 insufficient SOA mass formed during *Ficus* experiments to allow us to conduct any  
928 form of compositional analysis.

### 929 ***4.3 Atmospheric significance***

930 Our results are specific to VOC/NO<sub>x</sub> ratios of 3 - 9 and NO<sub>x</sub> mixing ratios of ~  
931 2 ppbV. Note, however, that the three reasons given above for the uncertainty in the  
932 aerosol production ascribed to isoprene in our experiments will also pertain to field  
933 measurements, often being exacerbated by variability and the difficulties of operating  
934 in the field. A contribution of isoprene to SOA is *supported* by recent observations of  
935 isoprene related SOA formation above the tropical forest of Danum Valley, Borneo, a  
936 high isoprene, low NO<sub>x</sub> region (typical ratio of 20:1, isoprene/NO<sub>x</sub>) (Hewitt et al.,  
937 2010b). Robinson *et al* (2010) observed that up to 15 % by mass of atmospheric sub-  
938 micron organic aerosol above the tropical forest of Danum Valley was comprised of  
939 methyl furan, the most likely source of which is the oxidation of isoprene (i.e. thermal  
940 decomposition of isoprene derived SOA) (Ruppert and Becker, 2000; Robinson et al.,  
941 2010; Lin et al., 2012; Budisulistiorini et al., 2013). Although much smaller in  
942 magnitude, the monoterpene emissions measured at Danum Valley were more than  
943 adequate to account for the remaining sub-micron organic aerosol (MacKenzie et al.,  
944 2011), just as in the majority of aerosol mass calculations for principally-isoprene-  
945 emitting tropical trees, described above.

946 It has recently been proposed that isoprene can inhibit aerosol formation when present  
947 in air containing other potential SOA precursors, such as mono- and sesquiterpenes  
948 (Kiendler-Scharr et al., 2009a). Kiendler-Scharr et al. propose that isoprene could  
949 effectively act as an OH scavenger, suppressing new particle formation by slowing  
950 the oxidation of available monoterpenes (and presumably sesquiterpenes). In line  
951 with this thesis, interpretation of the results obtained from our seeded experiments  
952 with *Ficus* species leaves room for a potential role for isoprene in inhibiting SOA  
953 formation under certain atmospheric conditions, i.e. our results imply that isoprene  
954 may impact negatively on the overall SOA forming potential of air containing other  
955 biogenic SOA precursors. However, owing to the constraints laid upon our  
956 experiments by the instrumentation and apparatus employed, it is difficult to assign a  
957 given certainty level to the role played by isoprene in the ambient atmosphere and  
958 caution should be taken when interpreting such findings.

959 The fact that isoprene accounts for approximately 50 % of the total global burden of  
960 non-methane VOC, (Guenther et al., 2006), would make it a significant contributor to  
961 global SOA. It has been estimated that, even if the secondary organic aerosol yield  
962 from isoprene is small (e.g. 1 %), the overall contribution to total atmospheric aerosol  
963 could be up to 6 Tg yr<sup>-1</sup> (Carlton et al., 2009). Van Donkelaar *et al.* (2007) found that  
964 using an isoprene SOA yield of 2 % improved the relationship between model  
965 simulations and organic aerosol measurements, and contributed 10 – 50 % of the total  
966 organic aerosol loading over the United States during the summer. Understanding the  
967 exact role played by isoprene in air containing many different VOCs, and being able

968 to account for the differing isoprene SOA yields under contrasting NO<sub>x</sub> and acidity  
969 (Lin *et al.*, 2012; Lin *et al.*, 2013; Pye *et al.*, 2013) environments, will undoubtedly  
970 help to significantly improve global modelling estimates of total SOA loading even  
971 further (Couvidat and Seigneur, 2010).

972 Further to any such potential impacts imposed by isoprene, it has recently been shown  
973 that a range of other BVOC emissions, released in response to a range of  
974 environmental stress factors, can also have significant impacts on biogenic SOA  
975 formation and yield (Mentel *et al.*, 2013). It has been shown that the emissions of  
976 sesquiterpenes, methyl salicylate and C<sub>17</sub> BVOCs, released as a result of certain  
977 environmental stress factors have a net positive impact on SOA yield; whereas certain  
978 stress induced green leaf volatiles ((Z)-3-hexenol and (Z)-3-hexenylacetate) behave  
979 similarly to isoprene, suppressing SOA formation (Mentel *et al.*, 2013).

980 Given the highly differing reported yields of isoprene SOA under various oxidant  
981 schemes, the uncertainty in the exact role played by isoprene and its oxidation  
982 products in realistic mixtures of VOCs (in particular in the context of SOA nucleation  
983 rates; Kiendler-Scharr *et al.*, 2009a) and the lack of knowledge regarding stress  
984 induced BVOCs, their atmospheric oxidation and their roles in biogenic SOA  
985 formation (and impact on chemical and physical properties), we suggest that there is a  
986 pressing requirement for additional, atmosphere-relevant laboratory and field studies  
987 to give us the necessary insight to successfully control biogenic SOA (Carlton *et al.*,  
988 2010).



989

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## 999 **5 References**

- 1000 Alfarrá, M. R., N. Good, K. P. Wyche, J. E. Hamilton, P. S. Monks, A. C. Lewis and  
1001 G. McFiggans (2013). "Water uptake is independent of the inferred composition of  
1002 secondary aerosols derived from multiple biogenic VOCs." Atmospheric Chemistry  
1003 and Physics **13**(23): 11769-11789. DOI: 10.5194/acp-13-11769-2013.  
1004
- 1005 Alfarrá, M. R., J. F. Hamilton, K. P. Wyche, N. Good, M. W. Ward, T. Carr, M. H.  
1006 Barley, P. S. Monks, M. E. Jenkin, A. C. Lewis and G. B. McFiggans (2012). "The  
1007 effect of photochemical ageing and initial precursor concentration on the composition  
1008 and hygroscopic properties of beta-caryophyllene secondary organic aerosol."  
1009 Atmospheric Chemistry and Physics **12**(14): 6417-6436. DOI: 10.5194/acp-12-6417-  
1010 2012.  
1011
- 1012 Alfarrá, M. R., D. Paulsen, M. Gysel, A. A. Garforth, J. Dommen, A. S. H. Prevot, D.  
1013 R. Worsnop, U. Baltensperger and H. Coe (2006). "A mass spectrometric study of  
1014 secondary organic aerosols formed from the photooxidation of anthropogenic and  
1015 biogenic precursors in a reaction chamber." Atmospheric Chemistry and Physics **6**:  
1016 5279-5293.  
1017

1018 Alfarra, M. R., D. Paulsen, M. Gysel, A. A. Garforth, J. Dommen, A. S. H. Prévôt, D.  
1019 R. Worsnop, U. Baltensperger and H. Coe (2006). "A mass spectrometric study of  
1020 secondary organic aerosols formed from the photooxidation of anthropogenic and  
1021 biogenic precursors in a reaction chamber." Atmospheric Chemistry and Physics **6**:  
1022 5279–5293.  
1023  
1024 Atkinson, R. and J. Arey (2003). "Gas-phase tropospheric chemistry of biogenic  
1025 volatile organic compounds: a review." Atmospheric Environment **37**: S197-S219.  
1026 DOI: 10.1016/s1352-2310(03)00391-1.  
1027  
1028 Bahreini, R., M. D. Keywood, N. L. Ng, V. Varutbangkul, S. Gao, R. C. Flagan, J. H.  
1029 Seinfeld, D. R. Worsnop and J. L. Jimenez (2005). "Measurements of secondary  
1030 organic aerosol from oxidation of cycloalkenes, terpenes, and m-xylene using an  
1031 Aerodyne aerosol mass spectrometer." Environmental Science and Technology, **39**:  
1032 15, 5674-5688  
1033  
1034 Baltensperger, U., J. Dommen, R. Alfarra, J. Duplissy, K. Gaeggeler, A. Metzger, M.  
1035 C. Facchini, S. Decesari, E. Finessi, C. Reinnig, M. Schott, J. Warnke, T. Hoffmann,  
1036 B. Klatzer, H. Puxbaum, M. Geiser, M. Savi, D. Lang, M. Kalberer and T. Geiser  
1037 (2008). "Combined determination of the chemical composition and of health effects  
1038 of secondary organic aerosols: The POLYSOA project." Journal of Aerosol Medicine  
1039 and Pulmonary Drug Delivery **21**(1): 145-154. DOI: 10.1089/jamp.2007.0655.  
1040  
1041 Benkelberg, H. J., O. Boge, R. Seuwen and P. Warneck (2000). "Product distributions  
1042 from the OH radical-induced oxidation of but-1-ene, methyl-substituted but-1-enes  
1043 and isoprene in NO<sub>x</sub>-free air." Physical Chemistry Chemical Physics **2**(18): 4029-  
1044 4039. DOI: 10.1039/b002053m.  
1045  
1046 Blake, R. S., P. S. Monks and A. M. Ellis (2009). "Proton-Transfer Reaction Mass  
1047 Spectrometry." Chemical Reviews **109**(3): 861-896. DOI: 10.1021/cr800364q.  
1048  
1049 Blake, R. S., C. Whyte, C. O. Hughes, A. M. Ellis and P. S. Monks (2004).  
1050 "Demonstration of proton-transfer reaction time-of-flight mass spectrometry for real-  
1051 time analysis of trace volatile organic compounds." Analytical Chemistry **76**(13):  
1052 3841-3845. DOI: 10.1021/ac0498260.  
1053  
1054 Budisulistiorini, S. H., Canagaratna, M. R., Croteau, P. L., Marth, W. J., Baumann, K.,  
1055 Edgerton, E. S., Shaw, S. L., Knipping, E. M., Worsnop, D. R., Jayne, J. T., Gold, A. and  
1056 Surratt J. D.: Real-Time Continuous Characterization of Secondary Organic Aerosol  
1057 Derived from Isoprene Epoxydiols in Downtown Atlanta, Georgia, Using the  
1058 Aerodyne Aerosol Chemical Speciation Monitor, *Environ. Sci. Technol.*, 47 (11), 686–  
1059 5694, doi: 10.1021/es400023n, 2013

1060 Brégonzio-Rozier, L., Siekmann, F., Giorio, C., Pangui, E., Morales, S., B., Temime-Roussel,  
1061 B., Gratien, A., Michoud, V., Ravier, S., Tapparo, A., Monod, A. and Doussin J.-F.: Gaseous  
1062 products and Secondary Organic Aerosol formation during long term oxidation of isoprene  
1063 and methacrolein, *Atmos. Chem. Phys. Discuss.*, 14, 22507-22545, 2014

1064  
1065 Camredon, M., J. F. Hamilton, M. S. Alam, K. P. Wyche, T. Carr, I. R. White, P. S.  
1066 Monks, A. R. Rickard and W. J. Bloss (2010). "Distribution of gaseous and  
1067 particulate organic composition during dark alpha-pinene ozonolysis." *Atmospheric*  
1068 *Chemistry and Physics* **10**(6): 2893-2917.

1069  
1070 Canagaratna, M. R., J. T. Jayne, J. L. Jimenez, J. D. Allan, M. R. Alfarra, Q. Zhang,  
1071 T. B. Onasch, F. Drewnick, H. Coe, A. Middlebrook, A. Delia, L. R. Williams, A. M.  
1072 Trimborn, M. J. Northway, P. F. DeCarlo, C. E. Kolb, P. Davidovits and D. R.  
1073 Worsnop (2007). "Chemical and microphysical characterization of ambient aerosols  
1074 with the aerodyne aerosol mass spectrometer." *Mass Spectrometry Reviews* **26**(2):  
1075 185-222. DOI: :10.1002/mas.20115.

1076  
1077 Carlton, A. G., R. W. Pinder, P. V. Bhave and G. A. Pouliot (2010). "To What Extent  
1078 Can Biogenic SOA be Controlled?" *Environmental Science & Technology* **44**(9):  
1079 3376-3380.

1080  
1081 Carlton, A. G., C. Wiedinmyer and J. H. Kroll (2009). "A review of Secondary  
1082 Organic Aerosol (SOA) formation from isoprene." *Atmospheric Chemistry and*  
1083 *Physics* **9**(14): 4987-5005.

1084  
1085 Carslaw, K. S., O. Boucher, D. V. Spracklen, G. W. Mann, J. G. L. Rae, S.  
1086 Woodward and M. Kulmala (2010). "A review of natural aerosol interactions and  
1087 feedbacks within the Earth system." *Atmos. Chem. Phys.* **10**(4): 1701-1737.

1088  
1089 Carvalho, L. R. F., P. C. Vasconcellos, W. Mantovani, C. S. Pool and S. O. Pisani  
1090 (2005). "Measurements of biogenic hydrocarbons and carbonyl compounds emitted  
1091 by trees from temperate warm Atlantic rainforest, Brazil." *Journal of Environmental*  
1092 *Monitoring* **7**(5): 493-499. DOI: 10.1039/b414881a.

1093  
1094 Claeys, M., B. Graham, G. Vas, W. Wang, R. Vermeylen, V. Pashynska, J. Cafmeyer,  
1095 P. Guyon, M. O. Andreae, P. Artaxo and W. Maenhaut (2004). "Formation of  
1096 secondary organic aerosols through photooxidation of isoprene." *Science* **303**(5661):  
1097 1173-1176.

1098  
1099 Claeys, M., W. Wang, A. C. Ion, I. Kourchev, A. Gelencser and W. Maenhaut  
1100 (2004). "Formation of secondary organic aerosols from isoprene and its gas-phase  
1101 oxidation products through reaction with hydrogen peroxide." *Atmospheric*  
1102 *Environment* **38**(25): 4093-4098. DOI: 10.1016/j.atmosenv.2004.06.001.

1103  
1104 Couvidat, F. and C. Seigneur (2010). "Modeling secondary organic aerosol formation  
1105 from isoprene oxidation under dry and humid conditions." Atmos. Chem. Phys.  
1106 Discuss. **10**(8): 20559-20605.  
1107  
1108 Dommen, J., H. Helleÿn, M. Saurer, M. Jaeggi, R. Siegwolf, A. Metzger, J. Duplissy,  
1109 M. Fierz and U. Baltensperger (2009). "Determination of the Aerosol Yield of  
1110 Isoprene in the Presence of an Organic Seed with Carbon Isotope Analysis."  
1111 Environmental Science & Technology **43**(17): 6697-6702.  
1112  
1113 Donahue, N. M., A. L. Robinson and S. N. Pandis (2009). "Atmospheric organic  
1114 particulate matter: From smoke to secondary organic aerosol." Atmospheric  
1115 Environment **43**(1): 94-106.  
1116  
1117 Drewnick, F., S. S. Hings, P. DeCarlo, J. T. Jayne, M. Gonin, K. Fuhrer, S. Weimer,  
1118 J. L. Jimenez, K. L. Demerjian, S. Borrmann and D. R. Worsnop (2005). "A new  
1119 time-of-flight aerosol mass spectrometer (TOF-AMS) - Instrument description and  
1120 first field deployment." Aerosol Science and Technology **39**(7): 637-658. DOI:  
1121 10.1080/02786820500182040.  
1122  
1123 Ehn, M., Thornton, J. A., Kleist, E., Sipila, M., Junninen, H., Pullinen, I., Springer,  
1124 M., Rubach, F., Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I. H.,  
1125 Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma, J., Kontkanen, J.,  
1126 Nieminen, T., Kurten, T., Nielsen, L. B., Jorgensen, S., Kjaergaard, H. G.,  
1127 Canagaratna, M., Dal Maso, M., Berndt, T., Petaja, T., Wahner, A., Kerminen, V. M.,  
1128 Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of low-  
1129 volatility secondary organic aerosol, *Nature*, 506 (7489), 476-479, doi:  
1130 10.1038/nature13032, 2014  
1131  
1132 Edney, E. O., T. E. Kleindienst, M. Jaoui, M. Lewandowski, J. H. Offenberg, W.  
1133 Wang and M. Claeys (2005). "Formation of 2-methyl tetrols and 2-methylglyceric  
1134 acid in secondary organic aerosol from laboratory irradiated isoprene/NOX/SO2/air  
1135 mixtures and their detection in ambient PM2.5 samples collected in the eastern United  
1136 States." Atmospheric Environment **39**(29): 5281-5289. DOI:  
1137 10.1016/j.atmosenv.2005.05.031.  
1138  
1139 Geron, C., S. Owen, A. Guenther, J. Greenberg, R. Rasmussen, J. Hui Bai, Q.-J. Li  
1140 and B. Baker (2006). "Volatile organic compounds from vegetation in southern  
1141 Yunnan Province, China: Emission rates and some potential regional implications."  
1142 Atmospheric Environment **40**(10): 1759-1773.  
1143

1144 Guenther, A., T. Karl, P. Harley, C. Wiedinmyer, P. I. Palmer and C. Geron (2006).  
1145 "Estimates of global terrestrial isoprene emissions using MEGAN (Model of  
1146 Emissions of Gases and Aerosols from Nature)." Atmospheric Chemistry and Physics  
1147 **6**: 3181-3210.

1148  
1149 Guenther, A. B., X. Jiang, C. L. Heald, T. Sakulyanontvittaya, T. Duhl, L. K.  
1150 Emmons and X. Wang (2012). "The Model of Emissions of Gases and Aerosols from  
1151 Nature version 2.1 (MEGAN2.1): an extended and updated framework for modeling  
1152 biogenic emissions." Geosci. Model Dev. **5**(6): 1471-1492. DOI: 10.5194/gmd-5-  
1153 1471-2012.

1154  
1155 Gunthe, S. S., S. M. King, D. Rose, Q. Chen, P. Roldin, D. K. Farmer, J. L. Jimenez,  
1156 P. Artaxo, M. O. Andreae, S. T. Martin and U. Poschl (2009). "Cloud condensation  
1157 nuclei in pristine tropical rainforest air of Amazonia: size-resolved measurements and  
1158 modeling of atmospheric aerosol composition and CCN activity." Atmospheric  
1159 Chemistry and Physics **9**(19): 7551-7575.

1160  
1161 Hallquist, M., J. C. Wenger, U. Baltensperger, Y. Rudich, D. Simpson, M. Claeys, J.  
1162 Dommen, N. M. Donahue, C. George, A. H. Goldstein, J. F. Hamilton, H. Herrmann,  
1163 T. Hoffmann, Y. Iinuma, M. Jang, M. E. Jenkin, J. L. Jimenez, A. Kiendler-Scharr,  
1164 W. Maenhaut, G. McFiggans, T. F. Mentel, A. Monod, A. S. H. Prevot, J. H. Seinfeld,  
1165 J. D. Surratt, R. Szmigielski and J. Wildt (2009). "The formation, properties and  
1166 impact of secondary organic aerosol: current and emerging issues." Atmospheric  
1167 Chemistry and Physics **9**(14): 5155-5236.

1168  
1169 Hamilton, J. F., Alfarra, M. R., Robinson, N., Ward, M. W., Lewis, A. C.,  
1170 McFiggans, G. B., Coe, H., and Allan, J. D.: Linking biogenic hydrocarbons to  
1171 biogenic aerosol in the Borneo rainforest. *Atmos. Chem. Phys.*, **13**(22), 11295-11305,  
1172 doi:10.5194/acp-13-11295-2013, 2013.

1173  
1174 Healy, R. M., J. C. Wenger, A. Metzger, J. Duplissy, M. Kalberer and J. Dommen  
1175 (2008). "Gas/particle partitioning of carbonyls in the photooxidation of isoprene and  
1176 1,3,5-trimethylbenzene." Atmospheric Chemistry and Physics **8**(12): 3215-3230.

1177  
1178 Hewitt, C. N., J. D. Lee, A. R. MacKenzie, M. P. Barkley, N. Carslaw, G. D. Carver,  
1179 N. A. Chappell, H. Coe, C. Collier, R. Commane, F. Davies, B. Davison, P. Di Carlo,  
1180 C. F. Di Marco, J. R. Dorsey, P. M. Edwards, M. J. Evans, D. Fowler, K. L.  
1181 Furneaux, M. Gallagher, A. Guenther, D. E. Heard, C. Helfter, J. Hopkins, T. Ingham,  
1182 M. Irwin, C. Jones, A. Karunaharan, B. Langford, A. C. Lewis, S. F. Lim, S. M.  
1183 MacDonald, A. S. Mahajan, S. Malpass, G. McFiggans, G. Mills, P. Misztal, S.  
1184 Moller, P. S. Monks, E. Nemitz, V. Nicolas-Perea, H. Oetjen, D. E. Oram, P. I.  
1185 Palmer, G. J. Phillips, R. Pike, J. M. C. Plane, T. Pugh, J. A. Pyle, C. E. Reeves, N. H.  
1186 Robinson, D. Stewart, D. Stone, L. K. Whalley and X. Yin (2010). "Overview:

1186 oxidant and particle photochemical processes above a south-east Asian tropical  
1187 rainforest (the OP3 project): introduction, rationale, location characteristics and  
1188 tools." Atmospheric Chemistry and Physics **10**(1): 169-199.  
1189  
1190 Hewitt, C. N., J. D. Lee, A. R. MacKenzie, M. P. Barkley, N. Carslaw, G. D. Carver,  
1191 N. A. Chappell, H. Coe, C. Collier, R. Commane, F. Davies, B. Davison, P. DiCarlo,  
1192 C. F. Di Marco, J. R. Dorsey, P. M. Edwards, M. J. Evans, D. Fowler, K. L.  
1193 Furneaux, M. Gallagher, A. Guenther, D. E. Heard, C. Helfter, J. Hopkins, T. Ingham,  
1194 M. Irwin, C. Jones, A. Karunaharan, B. Langford, A. C. Lewis, S. F. Lim, S. M.  
1195 MacDonald, A. S. Mahajan, S. Malpass, G. McFiggans, G. Mills, P. Misztal, S.  
1196 Moller, P. S. Monks, E. Nemitz, V. Nicolas-Perea, H. Oetjen, D. E. Oram, P. I.  
1197 Palmer, G. J. Phillips, R. Pike, J. M. C. Plane, T. Pugh, J. A. Pyle, C. E. Reeves, N. H.  
1198 Robinson, D. Stewart, D. Stone, L. K. Whalley and X. Yin (2010). "Overview:  
1199 oxidant and particle photochemical processes above a south-east Asian tropical  
1200 rainforest (the OP3 project): introduction, rationale, location characteristics and  
1201 tools." Atmos. Chem. Phys. **10**(1): 169-199.  
1202  
1203 Holzinger, R., A. Lee, K. T. Paw and A. H. Goldstein (2005). "Observations of  
1204 oxidation products above a forest imply biogenic emissions of very reactive  
1205 compounds." Atmospheric Chemistry and Physics **5**: 67-75.  
1206  
1207 Pye, H. O. T., Pinder, R. W., Piletic, I. R., Xie, Y., Capps, S. L., Lin, Y. H., Surratt, J.  
1208 D., Zhang, Z., Gold, A., Luecken, D. J., Hutzell, W. T., Jaoui, J., Offenberg, J. H.,  
1209 Kleindienst, T. E., Lewandowski, M., and Edney E. O.: Epoxide Pathways Improve  
1210 Model Predictions of Isoprene Markers and Reveal Key Role of Acidity in Aerosol  
1211 Formation, *Environ. Sci. Technol.*, **47** (19), 11056–11064, doi: 10.1021/es402106h,  
1212 2013  
1213  
1214 IPCC (2007). *Climate change 2007: The Physical Science Basis*, Cambridge  
1215 University Press, Cambridge.  
1216  
1217 Isaksen, I. S. A., C. Granier, G. Myhre, T. K. Berntsen, S. B. Dalsoren, M. Gauss, Z.  
1218 Klimont, R. Benestad, P. Bousquet, W. Collins, T. Cox, V. Eyring, D. Fowler, S.  
1219 Fuzzi, P. Jockel, P. Laj, U. Lohmann, M. Maione, P. Monks, A. S. H. Prevot, F. Raes,  
1220 A. Richter, B. Rognerud, M. Schulz, D. Shindell, D. S. Stevenson, T. Storelvmo, W.  
1221 C. Wang, M. van Weele, M. Wild and D. Wuebbles (2009). "Atmospheric  
1222 composition change: Climate-Chemistry interactions." Atmospheric Environment  
1223 **43**(33): 5138-5192. DOI: 10.1016/j.atmosenv.2009.08.003.  
1224  
1225 Jang, M. S., N. M. Czoschke, S. Lee and R. M. Kamens (2002). "Heterogeneous  
1226 atmospheric aerosol production by acid-catalyzed particle-phase reactions." Science,  
1227 **298**, 814-817.  
1228

1229 Jaoui, M., E. W. Corse, M. Lewandowski, J. H. Offenberg, T. E. Kleindienst and E.  
1230 O. Edney (2010). "Formation of organic tracers for isoprene SOA under acidic  
1231 conditions." Atmospheric Environment **In Press, Corrected Proof**.  
1232

1233 Jenkin, M. E., S. M. Saunders, and M. J. Pilling (1997). "The tropospheric  
1234 degradation of volatile organic compounds: a protocol for mechanism development."  
1235 Atmospheric Environment **31**(1): 81-104, 1997.  
1236

1237 Jenkin, M. E. (2004). "Modelling the formation and composition of secondary organic  
1238 aerosol from alpha- and beta-pinene ozonolysis using MCM v3." Atmospheric  
1239 Chemistry and Physics **4**: 1741-1757.  
1240

1241 Jenkin, M. E., K. P. Wyche, C. J. Evans, T. Carr, P. S. Monks, M. R. Alfarra, M. H.  
1242 Barley, G. B. McFiggans, J. C. Young and A. R. Rickard (2012). "Development and  
1243 chamber evaluation of the MCM v3.2 degradation scheme for beta-caryophyllene."  
1244 Atmospheric Chemistry and Physics **12**(11): 5275-5308. DOI: 10.5194/acp-12-5275-  
1245 2012.  
1246

1247 Junkermann, W., J. Hacker, T. Lyons and U. Nair (2009). "Land use change  
1248 suppresses precipitation." Atmospheric Chemistry and Physics **9**(17): 6531-6539.  
1249

1250 Kanawade, V. P., B. T. Jobson, A. B. Guenther, M. E. Erupe, S. N. Pressley, S. N.  
1251 Tripathi and S. H. Lee (2011). "Isoprene suppression of new particle formation in a  
1252 mixed deciduous forest." Atmospheric Chemistry and Physics **11**(12): 6013-6027.  
1253

1254 Karl, M., K. Tsigaridis, E. Vignati and F. Dentener (2009). "Formation of secondary  
1255 organic aerosol from isoprene oxidation over Europe." Atmospheric Chemistry and  
1256 Physics **9**(18): 7003-7030.  
1257

1258 Kiendler-Scharr, A., J. Wildt, M. Dal Maso, T. Hohaus, E. Kleist, T. F. Mentel, R.  
1259 Tillmann, R. Uerlings, U. Schurr and A. Wahner (2009a). "New particle formation in  
1260 forests inhibited by isoprene emissions." Nature **461**(7262): 381-384. DOI:  
1261 10.1038/nature08292.  
1262

1263 Kiendler-Scharr, A., Q. Zhang, T. Hohaus, E. Kleist, A. Mensah, T. F. Mentel, C.  
1264 Spindler, R. Uerlings, R. Tillmann and J. Wildt (2009b). "Aerosol Mass  
1265 Spectrometric Features of Biogenic SOA: Observations from a Plant Chamber and in  
1266 Rural Atmospheric Environments." Environmental Science & Technology **43**(21):  
1267 8166-8172.  
1268

1269 Kim, D. Y. and V. Ramanathan (2008). "Solar radiation budget and radiative forcing  
1270 due to aerosols and clouds." Journal of Geophysical Research-Atmospheres **113**(D2).  
1271

1272 Kim, S., A. Guenther, T. Karl and J. Greenberg (2011). "Contributions of primary and  
1273 secondary biogenic VOC to total OH reactivity during CABINEX (Community  
1274 Atmosphere-Biosphere INteractions Experiments)-09 field campaign." Atmospheric  
1275 Chemistry and Physics **11**: 8613-8623.  
1276  
1277 Kleindienst, T. E., E. O. Edney, M. Lewandowski, J. H. Offenberg and M. Jaoui  
1278 (2006). "Secondary organic carbon and aerosol yields from the irradiations of  
1279 isoprene and alpha-pinene in the presence of NO<sub>x</sub> and SO<sub>2</sub>." Environmental Science  
1280 & Technology **40**(12): 3807-3812.  
1281  
1282 Kleindienst, T. E., M. Lewandowski, J. H. Offenberg, M. Jaoui and E. O. Edney  
1283 (2007). "Ozone-isoprene reaction: Re-examination of the formation of secondary  
1284 organic aerosol." Geophysical Research Letters **34**(1).  
1285  
1286 Kleindienst, T. E., M. Lewandowski, J. H. Offenberg, M. Jaoui and E. O. Edney  
1287 (2009). "The formation of secondary organic aerosol from the isoprene plus OH  
1288 reaction in the absence of NO<sub>x</sub>." Atmospheric Chemistry and Physics **9**(17): 6541-  
1289 6558.  
1290  
1291 Kokkola, H., P. Yli-Pirila, M. Vesterinen, H. Korhonen, H. Keskinen, S.  
1292 Romakkaniemi, L. Hao, A. Kortelainen, J. Joutsensaari, D. R. Worsnop, A. Virtanen  
1293 and K. E. J. Lehtinen (2014). "The role of low volatile organics on secondary organic  
1294 aerosol formation." Atmospheric Chemistry and Physics **14**(3): 1689-1700. DOI:  
1295 10.5194/acp-14-1689-2014.  
1296  
1297 Kroll, J. H., N. L. Ng, S. M. Murphy, R. C. Flagan and J. H. Seinfeld (2005).  
1298 "Secondary organic aerosol formation from isoprene photooxidation under high-NO<sub>x</sub>  
1299 conditions." Geophysical Research Letters **32**(18). DOI: 10.1029/2005gl023637.  
1300  
1301 Kroll, J. H., N. L. Ng, S. M. Murphy, R. C. Flagan and J. H. Seinfeld (2006).  
1302 "Secondary organic aerosol formation from isoprene photooxidation." Environmental  
1303 Science & Technology **40**(6): 1869-1877. DOI: 10.1021/es0524301.  
1304  
1305 Lee, A., A. H. Goldstein, J. H. Kroll, N. L. Ng, V. Varutbangkul, R. C. Flagan and J.  
1306 H. Seinfeld (2006a). "Gas-phase products and secondary aerosol yields from the  
1307 photooxidation of 16 different terpenes." Journal of Geophysical Research-  
1308 Atmospheres **111**(D17). DOI: 10.1029/2006jd007050.  
1309  
1310 Lee, A., A. H. Goldstein, M. D. Keywood, S. Gao, V. Varutbangkul, R. Bahreini, N.  
1311 L. Ng, R. C. Flagan and J. H. Seinfeld (2006b). "Gas-phase products and secondary  
1312 aerosol yields from the ozonolysis of ten different terpenes." Journal of Geophysical  
1313 Research-Atmospheres **111**(D17). D07302, DOI:10.1029/2005JD006437.  
1314



- 1315 Liggio, J., S. M. Li, J. R. Brook and C. Mihele (2007). "Direct polymerization of  
1316 isoprene and  $\alpha$ -pinene on acidic aerosols." Geophysical Research Letters, 34, L05814,  
1317 10.1029/2006gl028468.
- 1318
- 1319 Limbeck, A., M. Kulmala and H. Puxbaum (2003). "Secondary organic aerosol  
1320 formation in the atmosphere via heterogeneous reaction of gaseous isoprene on acidic  
1321 particles." Geophysical Research Letters, 30, 1996, 10.1029/2003gl017738.
- 1322
- 1323 Lin, P., J. Z. Yu, G. Engling and M. Kalberer (2012). "Organosulfates in Humic-like  
1324 Substance Fraction Isolated from Aerosols at Seven Locations in East Asia: A Study  
1325 by Ultra-High-Resolution Mass Spectrometry." Environmental Science &  
1326 Technology, 46, 13118-13127, 10.1021/es303570v.
- 1327
- 1328 Lin, Y. H., Zhang, Z., Docherty, K. S., Zhang, H., Budisulistiorini, S. H., Rubitschun,  
1329 C. L., Shaw, S. L., Knipping, E. M., Edgerton, E. S., Kleindienst, T. E., Gold, A., and  
1330 Surratt, J. D.: Isoprene Epoxydiols as Precursors to Secondary Organic Aerosol  
1331 Formation: Acid-Catalyzed Reactive Uptake Studies with Authentic Compounds,  
1332 *Environ. Sci. Technol.*, 46 (1), 250–258, doi: 10.1021/es202554c, 2012
- 1333
- 1334 Lin, Y.-H., H. Zhang, H. O. T. Pye, Z. Zhang, W. J. Marth, S. Park, M. Arashiro, T.  
1335 Cui, S. H. Budisulistiorini, K. G. Sexton, W. Vizuete, Y. Xie, D. J. Luecken, I. R.  
1336 Piletic, E. O. Edney, L. J. Bartolotti, A. Gold and J. D. Surratt (2013). "Epoxide as a  
1337 precursor to secondary organic aerosol formation from isoprene photooxidation in the  
1338 presence of nitrogen oxides." Proceedings of the National Academy of Sciences, 110,  
1339 6718-6723, 10.1073/pnas.1221150110.
- 1340
- 1341 Lin, Y. H., Knipping, E. M., Edgerton, E. S., Shaw, S. L., and Surratt, J. D.:  
1342 Investigating the influences of SO<sub>2</sub> and NH<sub>3</sub> levels on isoprene-derived secondary  
1343 organic aerosol formation using conditional sampling approaches, *Atmos. Chem.*  
1344 *Phys.*, 13, 8457-8470, doi: 10.5194/acp-13-8457-2013, 2013
- 1345
- 1346 MacKenzie, A. R., B. Langford, T. A. M. Pugh, N. Robinson, P. K. Misztal, D. E.  
1347 Heard, J. D. Lee, A. C. Lewis, C. E. Jones, J. R. Hopkins, G. Phillips, P. S. Monks, A.  
1348 Karunaharan, K. E. Hornsby, V. Nicolas-Perea, H. Coe, A. M. Gabey, M. W.  
1349 Gallagher, L. K. Whalley, P. M. Edwards, M. J. Evans, D. Stone, T. Ingham, R.  
1350 Commane, K. L. Furneaux, J. B. McQuaid, E. Nemitz, Y. K. Seng, D. Fowler, J. A.  
1351 Pyle and C. N. Hewitt (2011). "The atmospheric chemistry of trace gases and  
1352 particulate matter emitted by different land uses in Borneo." Philosophical  
1353 Transactions of the Royal Society B-Biological Sciences 366(1582): 3177-3195. DOI:  
1354 10.1098/rstb.2011.0053.
- 1355
- 1356 Matsunaga, A. and P. J. Ziemann (2010). "Gas-Wall Partitioning of Organic  
Compounds in a Teflon Film Chamber and Potential Effects on Reaction Product and

1357 Aerosol Yield Measurements." Aerosol Science and Technology **44**(10): 881-892.  
1358 DOI: 10.1080/02786826.2010.501044.  
1359  
1360 Mentel, T. F., E. Kleist, S. Andres, M. Dal Maso, T. Hohaus, A. Kiendler-Scharr, Y.  
1361 Rudich, M. Springer, R. Tillmann, R. Uerlings, A. Wahner and J. Wildt (2013).  
1362 "Secondary aerosol formation from stress-induced biogenic emissions and possible  
1363 climate feedbacks." Atmospheric Chemistry and Physics **13**(17): 8755-8770. DOI:  
1364 10.5194/acp-13-8755-2013.  
1365  
1366 Mentel, T. F., J. Wildt, A. Kiendler-Scharr, E. Kleist, R. Tillmann, M. Dal Maso, R.  
1367 Fisseha, T. Hohaus, H. Spahn, R. Uerlings, R. Wegener, P. T. Griffiths, E. Dinar, Y.  
1368 Rudich and A. Wahner (2009). "Photochemical production of aerosols from real plant  
1369 emissions." Atmospheric Chemistry and Physics **9**(13): 4387-4406.  
1370  
1371 Mercado, L. M., N. Bellouin, S. Sitch, O. Boucher, C. Huntingford, M. Wild and P.  
1372 M. Cox (2009). "Impact of changes in diffuse radiation on the global land carbon  
1373 sink." Nature **458**(7241): 1014-U1087.  
1374  
1375 Meyer, N. K., J. Duplissy, M. Gysel, A. Metzger, J. Dommen, E. Weingartner, M. R.  
1376 Alfarra, A. S. H. Prevot, C. Fletcher, N. Good, G. McFiggans, A. M. Jonsson, M.  
1377 Hallquist, U. Baltensperger and Z. D. Ristovski (2009). "Analysis of the hygroscopic  
1378 and volatile properties of ammonium sulphate seeded and unseeded SOA particles."  
1379 Atmospheric Chemistry and Physics **9**(2): 721-732.  
1380  
1381 Ng, N. L., J. H. Kroll, M. D. Keywood, R. Bahreini, V. Varutbangkul, R. C. Flagan, J.  
1382 H. Seinfeld, A. Lee and A. H. Goldstein (2006). "Contribution of first- versus second-  
1383 generation products to secondary organic aerosols formed in the oxidation of biogenic  
1384 hydrocarbons." Environmental Science & Technology **40**(7): 2283-2297. DOI:  
1385 10.1021/es052269u.  
1386  
1387 Paulot, F., J. D. Crouse, H. G. Kjaergaard, A. Kurten, J. M. St Clair, J. H. Seinfeld  
1388 and P. O. Wennberg (2009). "Unexpected Epoxide Formation in the Gas-Phase  
1389 Photooxidation of Isoprene." Science **325**(5941): 730-733. DOI:  
1390 10.1126/science.1172910.  
1391  
1392 Paulson, S. E. and J. H. Seinfeld (1992). "DEVELOPMENT AND EVALUATION  
1393 OF A PHOTOOXIDATION MECHANISM FOR ISOPRENE." Journal of  
1394 Geophysical Research-Atmospheres **97**(D18): 20703-20715.  
1395  
1396 Riipinen, I., T. Yli-Juuti, J. R. Pierce, T. Petaja, D. R. Worsnop, M. Kulmala and N.  
1397 M. Donahue (2012). "The contribution of organics to atmospheric nanoparticle  
1398 growth." Nature Geosci **5**(7): 453-458.  
1399

1400 Robinson, N. H., J. F. Hamilton, J. D. Allan, B. Langford, D. E. Oram, Q. Chen, K.  
1401 Docherty, D. K. Farmer, J. L. Jimenez, M. W. Ward, C. N. Hewitt, M. H. Barley, M.  
1402 E. Jenkin, A. R. Rickard, S. T. Martin, G. McFiggans and H. Coe (2010). "Evidence  
1403 for a significant proportion of Secondary Organic Aerosol from isoprene above a  
1404 maritime tropical forest." Atmos. Chem. Phys. Discuss. **10**(11): 25545-25576.  
1405  
1406 Robinson, N. H., J. F. Hamilton, J. D. Allan, B. Langford, D. E. Oram, Q. Chen, K.  
1407 Docherty, D. K. Farmer, J. L. Jimenez, M. W. Ward, C. N. Hewitt, M. H. Barley, M.  
1408 E. Jenkin, A. R. Rickard, S. T. Martin, G. McFiggans and H. Coe (2011). "Evidence  
1409 for a significant proportion of Secondary Organic Aerosol from isoprene above a  
1410 maritime tropical forest." Atmospheric Chemistry and Physics **11**(3): 1039-1050.  
1411 DOI: 10.5194/acp-11-1039-2011.  
1412  
1413 Rollins, A. W., A. Kiendler-Scharr, J. L. Fry, T. Brauers, S. S. Brown, H. P. Dorn, W.  
1414 P. Dube, H. Fuchs, A. Mensah, T. F. Mentel, F. Rohrer, R. Tillmann, R. Wegener, P.  
1415 J. Wooldridge and R. C. Cohen (2009). "Isoprene oxidation by nitrate radical: alkyl  
1416 nitrate and secondary organic aerosol yields." Atmospheric Chemistry and Physics  
1417 **9**(18): 6685-6703.  
1418  
1419 Ruppert, L. and K. H. Becker (2000). "A product study of the OH radical-initiated  
1420 oxidation of isoprene: formation of C-5-unsaturated diols." Atmospheric Environment  
1421 **34**(10): 1529-1542.  
1422  
1423 Sillman, S. (1999). "The relation between ozone, NO<sub>x</sub> and hydrocarbons in urban and  
1424 polluted rural environments." Atmospheric Environment **33**(12): 1821-1845.  
1425  
1426 Sprengnether, M., K. L. Demerjian, N. M. Donahue and J. G. Anderson (2002).  
1427 "Product analysis of the OH oxidation of isoprene and 1,3-butadiene in the presence  
1428 of NO." Journal of Geophysical Research-Atmospheres **107**(D15). DOI:  
1429 10.1029/2001jd000716.  
1430  
1431 Stevens, B. and G. Feingold (2009). "Untangling aerosol effects on clouds and  
1432 precipitation in a buffered system." Nature **461**(7264): 607-613.  
1433  
1434 Surratt, J. D., A. W. H. Chan, N. C. Eddingsaas, M. Chan, C. L. Loza, A. J. Kwan, S.  
1435 P. Hersey, R. C. Flagan, P. O. Wennberg and J. H. Seinfeld (2010). "Reactive  
1436 intermediates revealed in secondary organic aerosol formation from isoprene."  
1437 Proceedings of the National Academy of Sciences **107**(15): 6640-6645. DOI:  
1438 10.1073/pnas.0911114107.  
1439  
1440 Surratt, J. D., M. Lewandowski, J. H. Offenberg, M. Jaoui, T. E. Kleindienst, E. O.  
1441 Edney and J. H. Seinfeld (2007). "Effect of Acidity on Secondary Organic Aerosol  
1442 Formation from Isoprene." Environmental Science & Technology **41**(15): 5363-5369.  
1443

1444 Surratt, J. D., S. M. Murphy, J. H. Kroll, N. L. Ng, L. Hildebrandt, A. Sorooshian, R.  
1445 Szmigielski, R. Vermeylen, W. Maenhaut, M. Claeys, R. C. Flagan and J. H. Seinfeld  
1446 (2006). "Chemical composition of secondary organic aerosol formed from the  
1447 photooxidation of isoprene." Journal of Physical Chemistry A **110**(31): 9665-9690.  
1448 DOI: 10.1021/jp061734m.  
1449  
1450 Surratt, J. D., S. M. Murphy, J. H. Kroll, N. L. Ng, L. Hildebrandt, A. Sorooshian, R.  
1451 Szmigielski, R. Vermeylen, W. Maenhaut, M. Claeys, R. C. Flagan and J. H. Seinfeld  
1452 (2006). "Chemical Composition of Secondary Organic Aerosol Formed from the  
1453 Photooxidation of Isoprene." The Journal of Physical Chemistry A **110**(31): 9665-  
1454 9690.  
1455  
1456 Surratt, J. D., A. W. H. Chan, N. C. Eddingsaas, M. Chan, C. L. Loza, A. J. Kwan, S.  
1457 P. Hersey, R. C. Flagan, P. O. Wennberg and J. H. Seinfeld (2010). "Reactive  
1458 intermediates revealed in secondary organic aerosol formation from isoprene."  
1459 Proceedings of the National Academy of Sciences, 107, 6640-6645,  
1460 10.1073/pnas.0911114107, 2010.  
1461  
1462 Tuazon, E. C. and R. Atkinson (1990). "A PRODUCT STUDY OF THE GAS-  
1463 PHASE REACTION OF ISOPRENE WITH THE OH RADICAL IN THE  
1464 PRESENCE OF NOX." International Journal of Chemical Kinetics **22**(12): 1221-  
1465 1236. DOI: 10.1002/kin.550221202.  
1466  
1467 van Donkelaar, A., R. V. Martin, R. J. Park, C. L. Heald, T. M. Fu, H. Liao and A.  
1468 Guenther (2007). "Model evidence for a significant source of secondary organic  
1469 aerosol from isoprene." Atmospheric Environment **41**(6): 1267-1274. DOI:  
1470 10.1016/j.atmosenv.2006.09.051.  
1471  
1472 VanReken, T. M., J. P. Greenberg, P. C. Harley, A. B. Guenther and J. N. Smith  
1473 (2006). "Direct measurement of particle formation and growth from the oxidation of  
1474 biogenic emissions." Atmospheric Chemistry and Physics **6**: 4403-4413.  
1475  
1476 Verheggen, B. and M. Mozurkewich (2006). "An inverse modeling procedure to  
1477 determine particle growth and nucleation rates from measured aerosol size  
1478 distributions." Atmos. Chem. Phys., 6, 2927-2942, doi:10.5194/acp-6-2927-2006.  
1479  
1480 Virtanen, A., J. Joutsensaari, T. Koop, J. Kannosto, P. Yli-Pirila, J. Leskinen, J. M.  
1481 Makela, J. K. Holopainen, U. Poschl, M. Kulmala, D. R. Worsnop and A. Laaksonen  
1482 (2010). "An amorphous solid state of biogenic secondary organic aerosol particles."  
1483 Nature **467**(7317): 824-827.  
1484  
1485 Wang, W., G. Vas, R. Dommissie, K. Loones and M. Claeys (2004). "Fragmentation  
1486 study of diastereoisomeric 2-methyltetrols, oxidation products of isoprene, as their

1487 trimethylsilyl ethers, using gas chromatography/ion trap mass spectrometry." Rapid  
1488 Communications in Mass Spectrometry **18**(16): 1787-1797. DOI: 10.1002/rcm.1553.  
1489  
1490 Wyche, K. P., R. S. Blake, A. M. Ellis, P. S. Monks, T. Brauers, R. Koppmann and E.  
1491 C. Apel (2007). "Technical note: Performance of Chemical Ionization Reaction Time-  
1492 of-Flight Mass Spectrometry (CIR-TOF-MS) for the measurement of atmospherically  
1493 significant oxygenated volatile organic compounds." Atmospheric Chemistry and  
1494 Physics **7**: 609-620.  
1495  
1496 Wyche, K. P., et al. (2014) "Mapping organic reactivity during secondary organic  
1497 aerosol formation". **In Preparation**.  
1498  
1499 Zhang, Q., J. L. Jimenez, M. R. Canagaratna, J. D. Allan, H. Coe, I. Ulbrich, M. R.  
1500 Alfarra, A. Takami, A. M. Middlebrook, Y. L. Sun, K. Dzepina, E. Dunlea, K.  
1501 Docherty, P. F. DeCarlo, D. Salcedo, T. Onasch, J. T. Jayne, T. Miyoshi, A. Shimono,  
1502 S. Hatakeyama, N. Takegawa, Y. Kondo, J. Schneider, F. Drewnick, S. Borrmann, S.  
1503 Weimer, K. Demerjian, P. Williams, K. Bower, R. Bahreini, L. Cottrell, R. J. Griffin,  
1504 J. Rautiainen, J. Y. Sun, Y. M. Zhang and D. R. Worsnop (2007). "Ubiquity and  
1505 dominance of oxygenated species in organic aerosols in anthropogenically-influenced  
1506 Northern Hemisphere midlatitudes." Geophysical Research Letters **34**(13). DOI:  
1507 10.1029/2007gl029979.  
  
1508 Zhao, J., R. Zhang, E. C. Fortner and S. W. North (2004). "Quantification of  
1509 hydroxycarbonyls from OH-isoprene reactions." Journal of the American Chemical  
1510 Society **126**: 2686-2687.  
  
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1512 **Tables**

1513 Table 1: List of experiments conducted and their general parameters

Date	Tree Species	Initial NO <sub>x</sub> / ppbV	VOC/NO <sub>x</sub>	Relative Humidity / %	Pre- existing Seed
22/06/09	<i>Ficus benjamina</i>	3	4.2	79	None
23/06/09	<i>Ficus benjamina</i>	6	2.7	75	None
25/06/09	<i>Ficus benjamina</i>	2	6.3	65	Sulphate
29/06/09	<i>Ficus</i> <i>cyathistipula</i>	2	9.4	71	None
30/06/09	<i>Ficus</i> <i>cyathistipula</i>	2	7.8	75	Sulphate
02/07/09	<i>Ficus</i> <i>cyathistipula</i>	3	5.6	78	Sulphate
06/07/09	<i>Betula pendula</i>	3	5.6	84	None
07/07/09	<i>Betula pendula</i>	3	5.5	73	Sulphate
09/07/09	<i>Betula pendula</i>	2	1.5	70	Sulphate
10/07/09	<i>Betula pendula</i> + 36 ppbV isoprene	2	5.5	70	Sulphate
13/07/09	<i>Ficus benjamina</i>	2	- <sup>1</sup>	87	Sulphate
15/07/09	<i>Ficus benjamina</i>	3	- <sup>1</sup>	89	Sulphate

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16/07/09	<i>Ficus benjamina</i> +	2	- <sup>1</sup>	85	None
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4.5 ppbV

limonene

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1514

1515 Notes:

1516 1. No quantified VOC data available

1517

1517

1518 Table 2: GC-MS identification of the biogenic VOC present in the plant chamber air  
 1519 immediately before RC filling began. Quantification of isoprene, total monoterpenes  
 1520 and total sesquiterpenes was carried out using PTR-MS and CIRMS (see Figures 2  
 1521 and 5).

Experiment	Compounds detected by GC-MS (abundance ppbV) (trace = <0.1 ppb)			
	Isoprene	Monoterpenes (and related)	Sesquiterpenes	Other
<i>B. Pendula</i>	Yes (0.25-1.19)	$\alpha$ -pinene (0.31 – 1.08)  $\beta$ -pinene (0.74 – 7.19)  ocimene (trace – 1.22)  $\Delta$ 3-carene (1.89 – 4.94)  $\gamma$ -terpinene (trace)  2, 4, 6-octatriene,2,6- dimethyl (trace)  4,7-methano-1H-indene, octahydro (trace)	$\beta$ -caryophyllene (0.15 – 0.22)  $\alpha$ -farnesene (0.14)  $\alpha$ -caryophyllene (0.59 – 0.92)  $\alpha$ -copaene (trace)  aromadendrene (0.45- 0.51)  $\alpha$ -cedrene (trace)  $\alpha$ -pyronene (trace)	acetaldehyde (trace)  caryophyllene- epoxide (trace)  nerolidol (trace)  linalool (trace)



<i>F. Benjamina</i>	Yes	$\alpha$ -pinene (trace)	$\alpha$ -cubebene (trace)	acetaldehyde (trace)
	(38.49)	limonene (trace)		benzoquinone
		sabinene (trace)		(trace)
		linalool (trace)		pyridine (trace)
				methyl salicate
				(trace)
				decanal (trace)
<i>F. Cyathistipula</i>	Yes (75.08)	$\alpha$ -pinene (trace)	$\beta$ -caryophyllene (trace)	acetic acid (trace)
		$\beta$ -pinene (trace)		
		limonene (trace)		
Mixed canopy	Yes	$\alpha$ -pinene (0.37)	None detected	<i>p</i> -dichlorobenzene
<i>F. Benjamina</i>		camphene (0.11)		methyl salicate
<i>F. Cyathistipula</i>		limonene (0.42)		
<i>C. Millis</i>		ocimene (trace)		

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1527 Table 3: Yields of isoprene oxidation products compared to literature values. Yields  
 1528 are an average from all *Ficus* experiments (seeded and unseeded) (n = 4) calculated at  
 1529 4 hours after lights on (HALO). Yield is based on the calculated relationship between  
 1530 the amount of isoprene reacted and the oxidation product in question.

1531

Isoprene	Hydro- peroxides <sup>1</sup>	MVK +MACR	Hydroxy- acetone	Form- aldehyde	Source
1	0.33	0.26	0.07		Williams et al 1999
1	0.05	0.36	0.05		Williams et al 2001 (NO <sub>x</sub> present)
1		0.33		0.67	Zang et al 2002
1		0.46 – 0.60		0.86 – 0.96	Niki et al 1983 Kamens et al 1982
1	0.18	0.25			Surratt et al 2010 NO <sub>x</sub> present NO <sub>x</sub> absent
1		0.27			Kleindienst et al 2009
1	0.01 –	0.17 – 0.36	0 – 0.02	0.02 – 0.07	This study

Isoprene	Hydro- peroxides <sup>1</sup>	MVK +MACR	Hydroxy- acetone	Form- aldehyde	Source
0.03					

1532

1533 Notes:

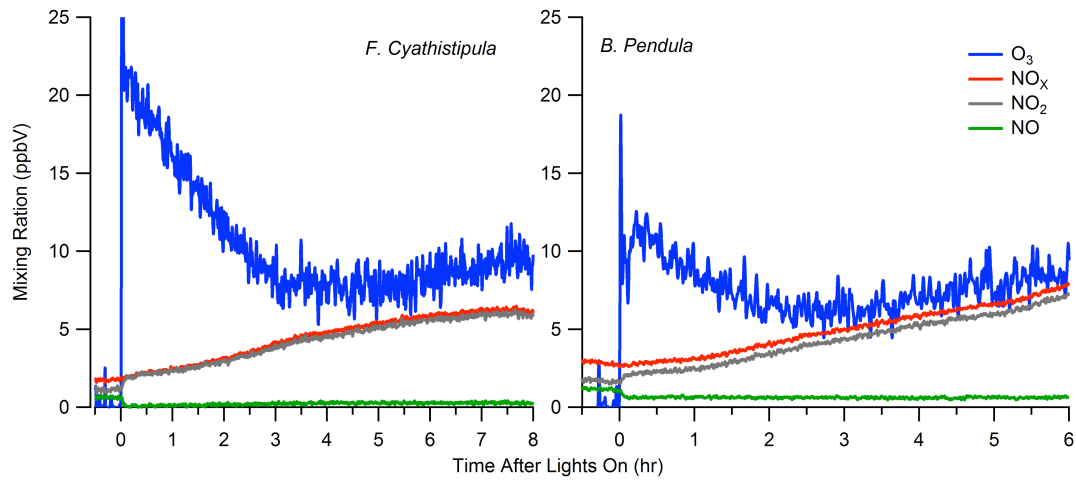
1534 1. Sum of isoprene epoxide (IEPOX) and isoprene hydroperoxide (ISOPOOH)

1535

1535

1536 **Figures**

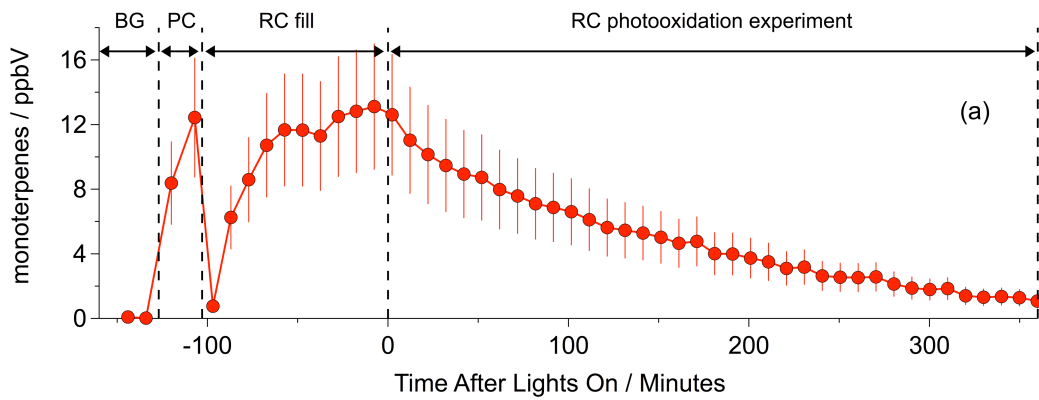
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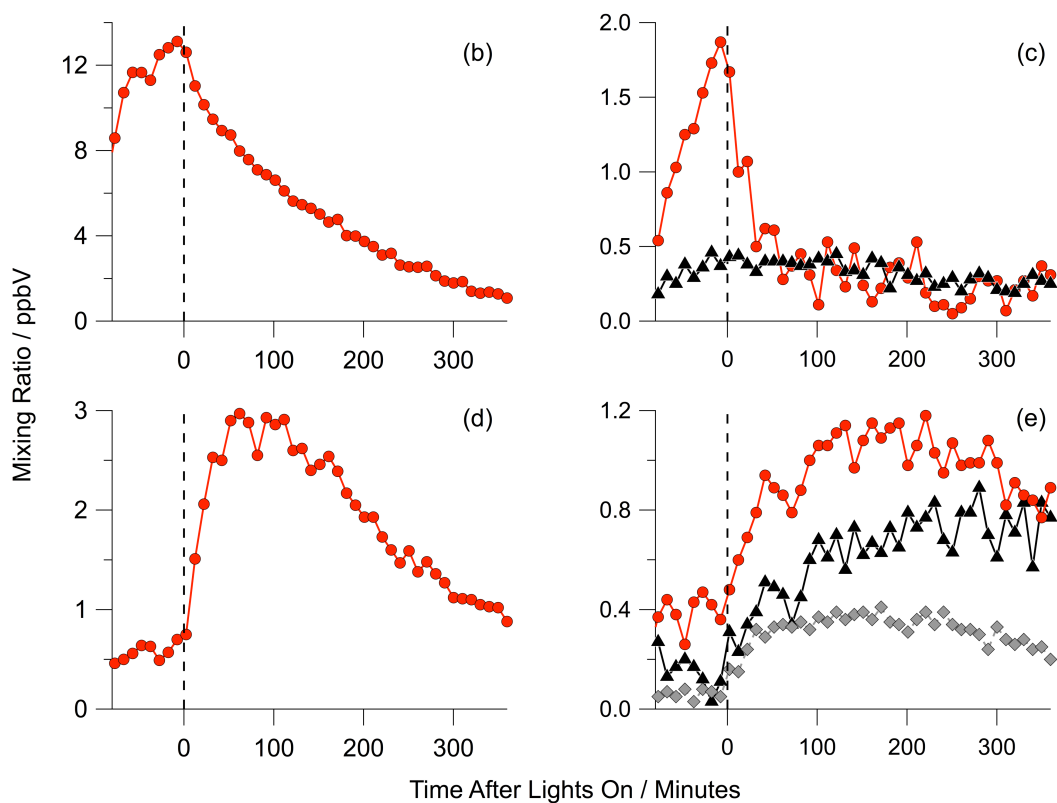
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1540 Figure 1: Temporal evolution of NO, NO<sub>2</sub>, NO<sub>x</sub> and O<sub>3</sub> during typical *Ficus*  
1541 *Cyathistipula* (a) and *Betula pendula* (b) experiments (25/06/09 and 07/07/09,  
1542 respectively).



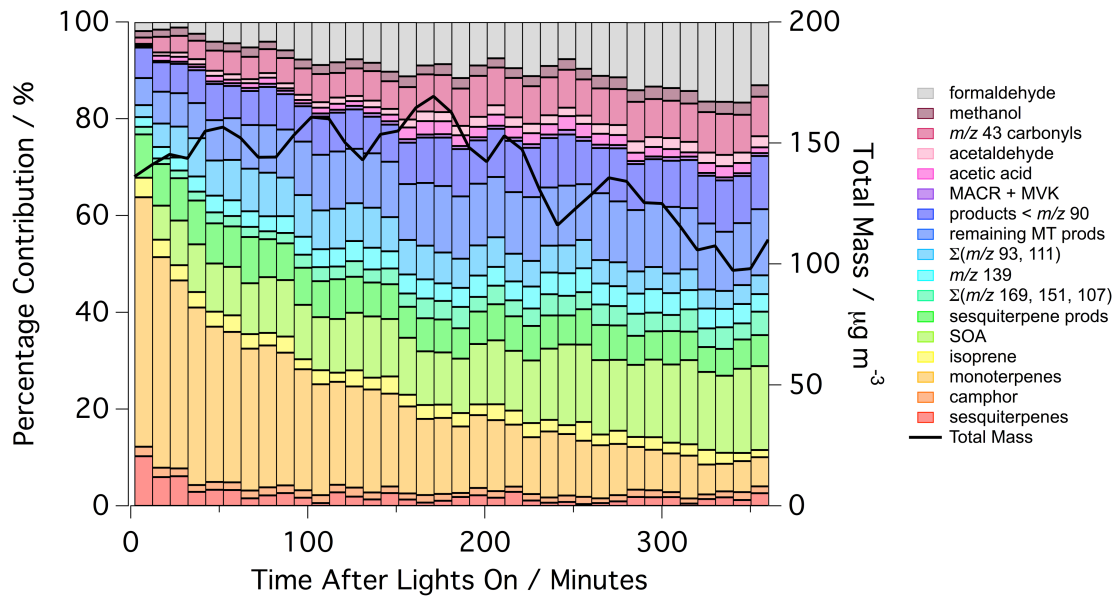
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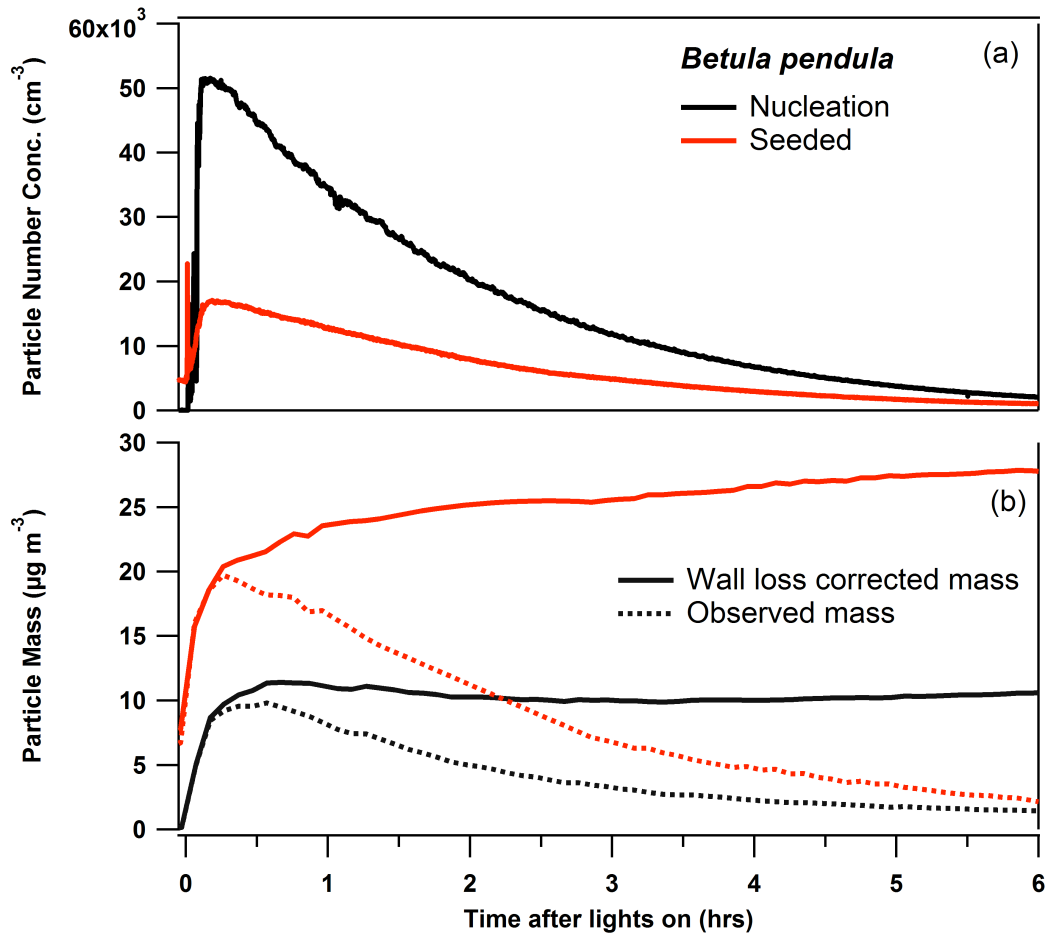


1544

1545 Figure 2: Temporal evolution of a series of isoprenoid “precursor” compounds and  
 1546 their oxidation products, as observed in the main reaction chamber during an example  
 1547 *Betula pendula* experiment (07/0709). The top panel (a) shows the entire experiment  
 1548 process in terms of monoterpene evolution, from background (BG) and plant chamber  
 1549 (PC) measurements, to reaction chamber (RC) fill and the main photooxidation  
 1550 experiment within the reaction chamber. Figure (b) shows monoterpene evolution, (c)  
 1551 shows sesquiterpenes (red circles and lines) and camphore (black triangles and lines),  
 1552 (d)  $\Sigma(I_{111}, I_{93})$  and (e) the primary ketone ( $m/z$  139) (red circles and lines), primary  
 1553 keto-aldehyde ( $m/z$  107 + 151 + 169) (black triangles and lines) and MVK + MACR  
 1554 ( $m/z$  71) (grey diamonds and dashed line).

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1565

1566 Figure 4: Particle number and mass concentrations measured during nucleation

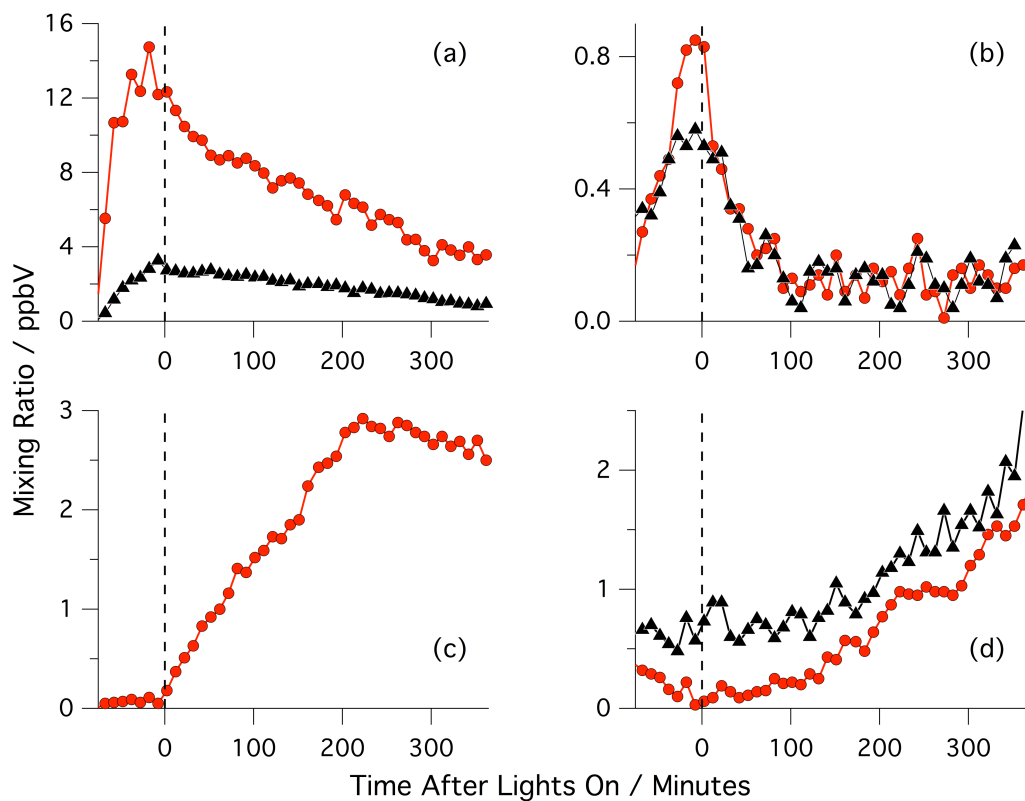
1567 (06/07/09) and ammonium sulphate seeded (07/07/09) *Betula pendula* experiments

1568 (a). In the bottom panel (b), both the measured (dashed lines) and the wall loss

1569 corrected (solid lines) mass concentrations are shown.

1570



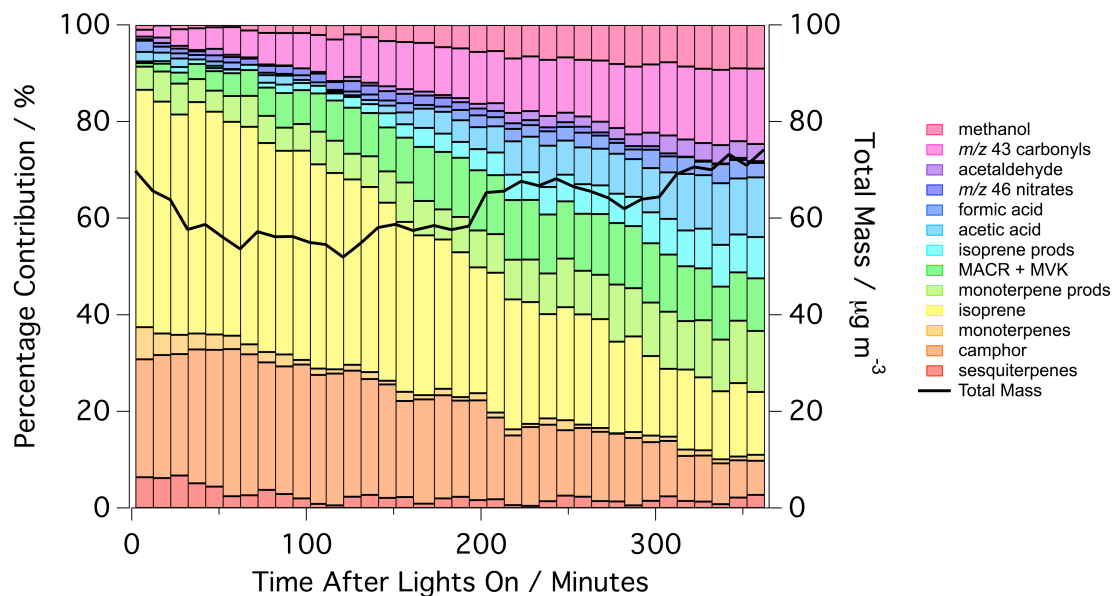


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1572

1573 Figure 5: Temporal evolution of a series of isoprenoid “precursor” compounds and  
 1574 their oxidation products, as observed in the main reaction chamber during an example  
 1575 *Ficus benjamina* experiment (23/06/09). Panel (a) shows isoprene (red circles and  
 1576 lines) and camphore (black triangles and lines) evolution, (b) shows monoterpenes  
 1577 (red circles and lines) and sesquiterpenes (black triangles and lines), (c) MVK +  
 1578 MACR ( $m/z$  71) and (d)  $\Sigma$ (monoterpene products) (black triangles and lines) and  
 1579  $\Sigma$ (non MVK+MACR isoprene products) (red circles and lines)

1580

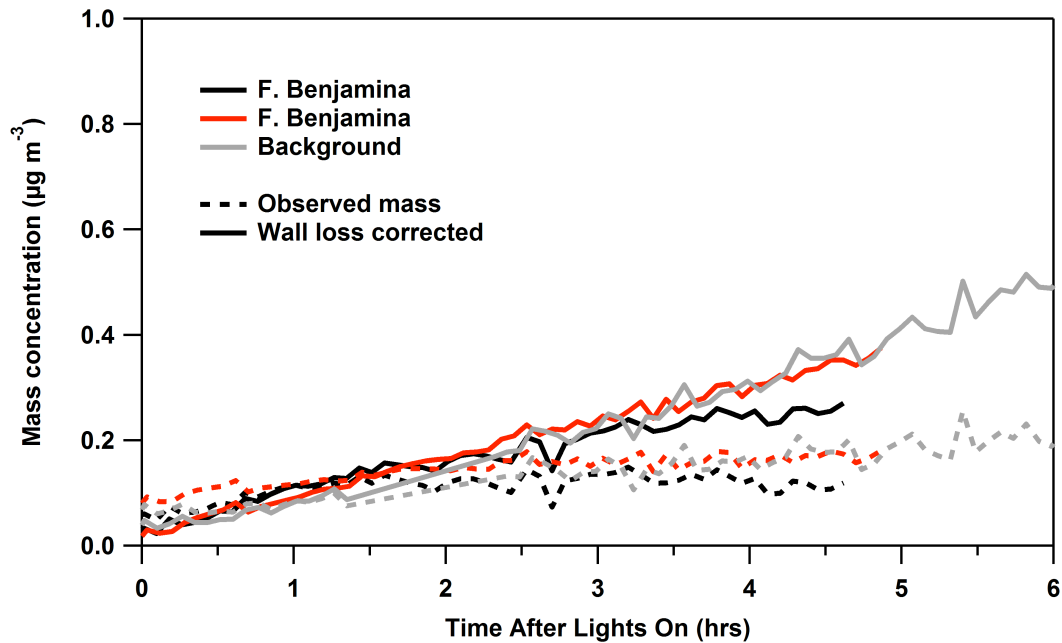


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1582

1583 Figure 6: Evolution of measured mass through the *Ficus benjamina* system (23 June  
 1584 2009), showing the relative contribution of precursor compounds and oxidation  
 1585 products to total measured mass, with time (coloured bars, left axis) and total  
 1586 measured mass (i.e.  $\Sigma$ VOCs + SOA) with time (black line, right axis).

1587

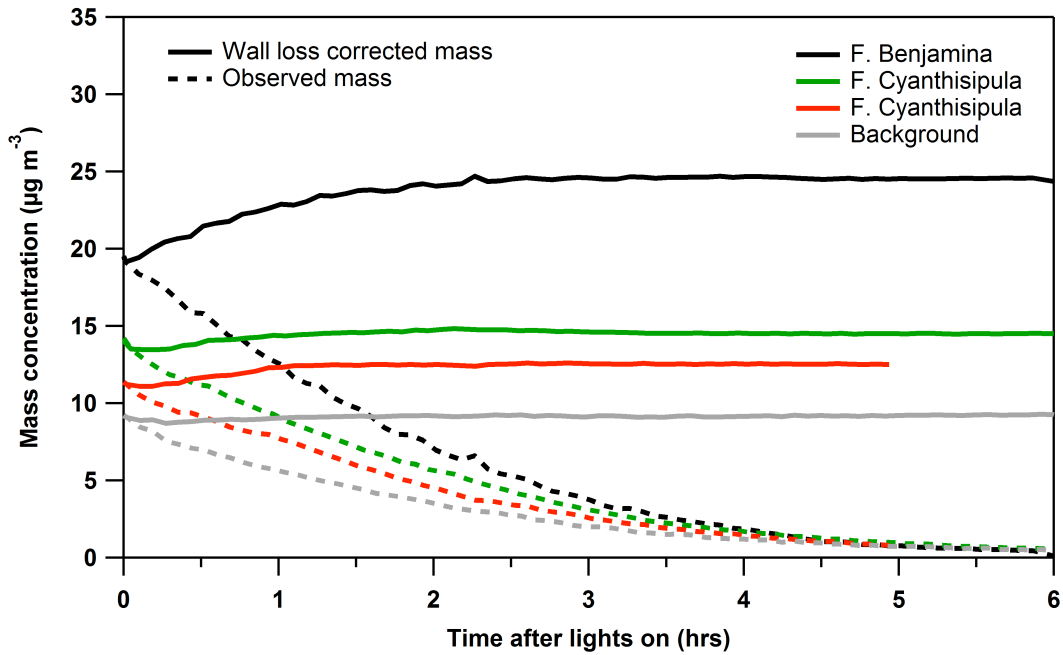


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1590 Figure 7: Observed and wall loss corrected particle mass concentrations during un-  
 1591 seeded *Ficus benjamina* (22/06/09, 23/06/09) and chamber background (26/06/09)  
 1592 experiments. The reaction chamber was filled with plant chamber air over a period of  
 1593 1 – 1.5 hours. Chamber filling was carried out in the dark. Ozone was added  
 1594 immediately prior to lights on. Time begins at the point at which the reaction chamber  
 1595 was illuminated, then increments in hours after lights on.

1596

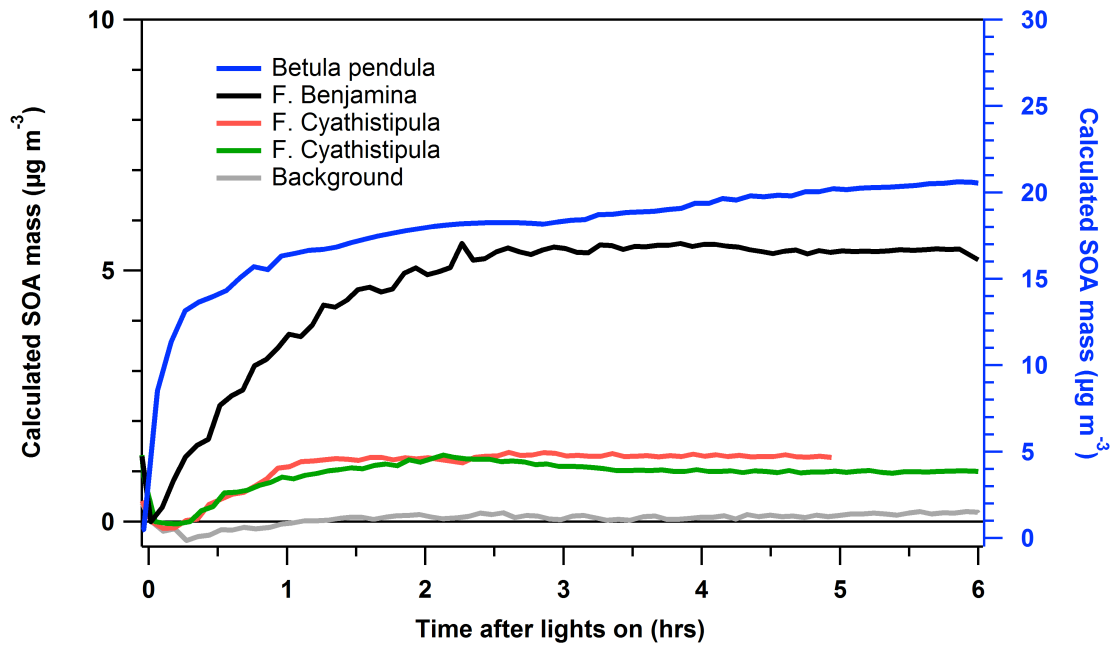


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1598

1599 Figure 8: Observed and wall loss corrected particle mass concentrations during  
 1600 ammonium sulphate seeded *Ficus benjamina* (15/07/09), *Ficus cyathistipula*  
 1601 (30/06/09, 02/07/09) and chamber background (03/07/09) experiments. Ozone and  
 1602 ammonium sulphate seed were added immediately prior to lights on.

1603

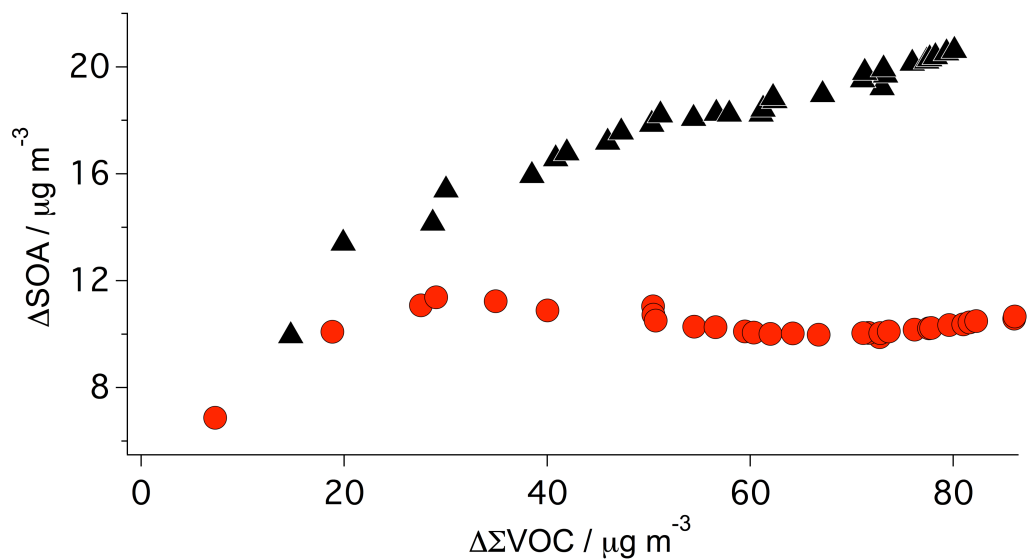


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1606 Figure 9: Calculated SOA mass concentrations during ammonium sulphate seeded  
 1607 experiments for *Betula pendula* (07/07/09), *Ficus benjamina* (15/07/09) and *Ficus*  
 1608 *cyathistipula* (30/06/09, 02/07/09). See text for details.

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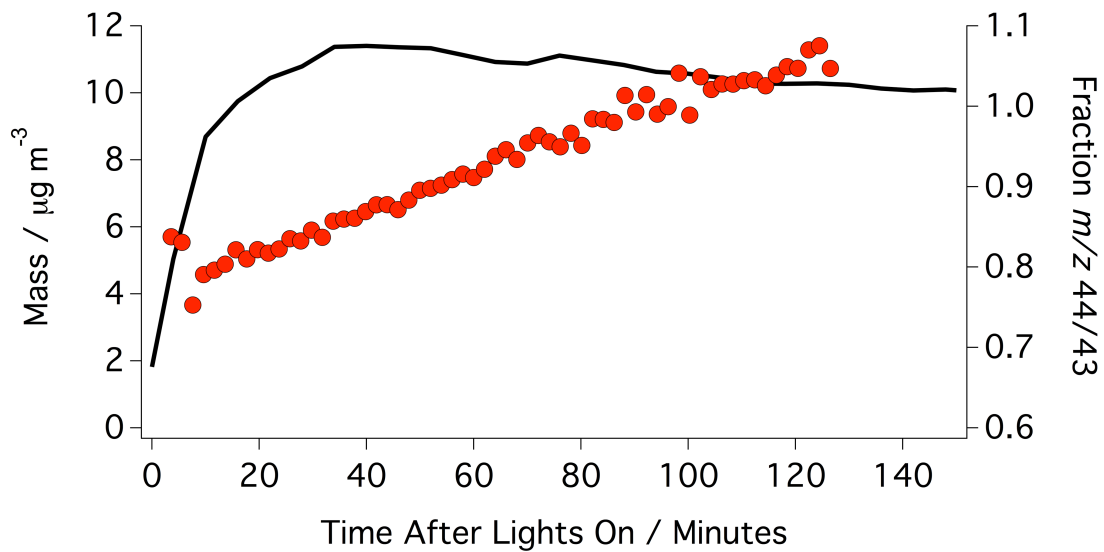
1612 Figure 10: Time dependent growth curves for two typical *Betula pendula* experiments

1613 (red circles- nucleation experiment on 06/07/09 and black triangles- ammonium

1614 sulphate seeded experiment on 07/07/09), showing SOA growth behaviour with

1615 respect to consumption of the VOC precursors.

1616

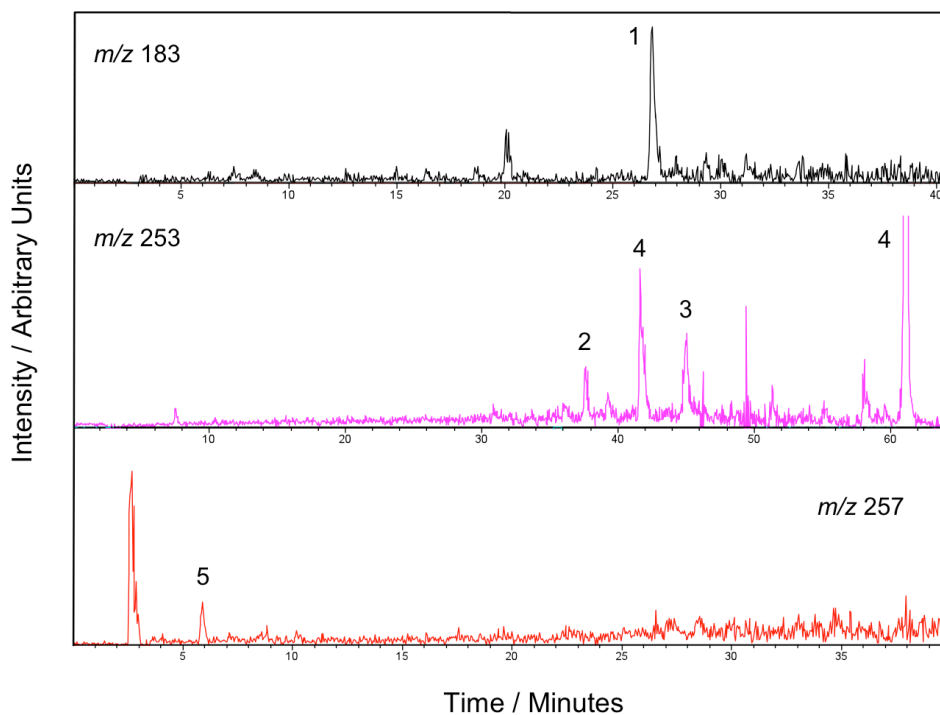


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1619 Figure 11: Temporal evolution of the  $m/z$  44/43 ratio (red circles) during a typical  
 1620 *Betula pendula* experiment (06/07/09) and wall loss corrected SOA mass (black line);  
 1621 demonstrating the increase in oxygenated content of the SOA as the air matrix begins  
 1622 to age.

1623



1624

1625 Figure 12: LC-MS<sup>2</sup> selected ion chromatograms derived from the off-line analysis of  
 1626 SOA collected on filters at the conclusion of a typical *Betula pendula* experiment  
 1627 (07/07/2009). Notes: Upper;  $m/z$  183 = MW 184, 1 = cis-pinonic acid. Middle;  $m/z$   
 1628 253 = MW 254, 2 =  $\beta$ -nocaryophyllonic acid, 3 =  $\beta$ -caryophyllinic acid, 4 = similar  
 1629 to sesquiterpene SOA. Lower;  $m/z$  257 = MW 118 [ $2*[M-H] + Na$ ]<sup>-</sup>, also seen in  
 1630 myrcene SOA, with same MS<sup>2</sup> spectra.