

31 **Abstract**

32 We used the nested grid version of the global three-dimensional Goddard
33 Earth Observing System chemical transport model (GEOS-Chem) to examine
34 the interannual variations (IAVs) of aerosols over heavily polluted regions in
35 China for years 2004–2012. The role of variations in meteorological
36 parameters was quantified by a simulation with fixed anthropogenic emissions
37 at year 2006 levels and changes in meteorological parameters over
38 2004–2012. Simulated PM_{2.5} (particles with a diameter of 2.5 μm or less)
39 aerosol concentrations exhibited large IAVs in North China (NC, 32–42°N,
40 110–120°E), with regionally averaged absolute percent departure from the
41 mean (APDM) values of 17%, 14%, 14%, and 11% in
42 December-January-February (DJF), March-April-May (MAM),
43 June-July-August (JJA), and September-October-November (SON),
44 respectively. Over South China (SC, 22–32°N, 110–120°E), the IAVs in PM_{2.5}
45 were found to be the largest in JJA, with the regional mean APDM values of 14%
46 in JJA and of about 9% in other seasons. Concentrations of PM_{2.5} over the
47 Sichuan Basin (SCB, 27–33°N, 102–110°E) were simulated to have the
48 smallest IAVs among the polluted regions examined in this work, with the
49 APDM values of 8–9% in all seasons. All aerosol species (sulfate, nitrate,
50 ammonium, black carbon, and organic carbon) were simulated to have the
51 largest IAVs over NC in DJF, corresponding to the large variations in
52 meteorological parameters over NC in this season. Process analyses were
53 performed to identify the key meteorological parameters that determined the
54 IAVs of different aerosol species in different regions. While the variations in
55 temperature and specific humidity, which influenced the gas-phase formation

56 of sulfate, jointly determined the IAVs of sulfate over NC in both DJF and JJA,
57 wind (or convergence of wind) in DJF and precipitation in JJA were the
58 dominant meteorological factors to influence IAVs of sulfate over SC and the
59 SCB. The IAVs in temperature and specific humidity influenced gas-to-aerosol
60 partitioning, which were the major factors that led to the IAVs of nitrate aerosol
61 in China. The IAVs in wind and precipitation were found to drive the IAVs of
62 organic carbon aerosol. We also compared the IAVs of aerosols simulated with
63 variations in meteorological parameters alone with those simulated with
64 variations in anthropogenic emissions alone; the variations in meteorological
65 fields were found to dominate the IAVs of aerosols in northern and southern
66 China over 2004–2012. Considering that the IAVs in meteorological fields are
67 mainly associated with natural variability in the climate system, the IAVs in
68 aerosol concentrations driven by meteorological parameters have important
69 implications for the effectiveness of short-term air quality control strategies in
70 China.

71

72 1 Introduction

73 Aerosols are major air pollutants that have adverse effects on human health,
74 reduce atmospheric visibility, and influence global climate change. With the
75 rapid economic development in China over the past decades, concentrations
76 of aerosols are now among the highest in the world (Fu et al., 2008; Cao et al.,
77 2012; Sun et al., 2013) driven mainly by the increases in direct and precursor
78 emissions (Streets et al., 2003). Aerosol concentrations in China have
79 variations on different time scales (Zhang et al., 2010a; Zhu et al., 2012); we
80 aim to understand interannual variations (IAVs) of aerosols in this study.
81 Understanding interannual variations in aerosols driven by variations in
82 meteorological parameters is especially important for air pollution control. For
83 example, the Action Plan for Air Pollution Prevention and Control released by
84 the State Council of China in year 2013 aims to reduce the annual mean PM_{2.5}
85 concentrations in the regions of Beijing-Tianjin-Hebei, Yangtze Delta, and
86 Pearl River Delta by 25%, 20%, and 15% respectively, as the concentrations in
87 year 2017 are compared with those in 2012. The role of interannual variations
88 in meteorological parameters needs to be separated from the impact of the
89 reductions in emissions in these targeted reductions.

90 The IAVs of aerosols were usually quantified in previous studies by
91 statistical variables such as standard deviation (SD), relative standard
92 deviation (RSD), mean absolute deviation (MAD), and absolute percent
93 departure from the mean (APDM), which are defined as

$$94 \quad SD = \sqrt{\frac{1}{n} \sum_{i=1}^n \left(C_i - \frac{1}{n} \sum_{i=1}^n C_i \right)^2} \quad (1)$$

$$95 \quad RSD = 100\% \times SD / \left(\frac{1}{n} \sum_{i=1}^n C_i \right) \quad (2)$$

96
$$\text{MAD} = \frac{1}{n} \sum_{i=1}^n \left| C_i - \frac{1}{n} \sum_{i=1}^n C_i \right| \quad (3)$$

97
$$\text{APDM} = 100\% \times \text{MAD} / \left(\frac{1}{n} \sum_{i=1}^n C_i \right) \quad (4)$$

98 where C_i is aerosol concentration of year i , and n is the number of years
99 examined. Therefore SD and MAD represent the absolute IAVs in aerosol
100 concentration, and RSD and APDM represent the IAVs relative to the average
101 concentration over the n years.

102 Large IAVs of aerosols have been reported in previous studies for different
103 aerosol species in different regions. Mahowald et al. (2003) showed that
104 annual mean mineral dust concentrations measured at 10 sites in the United
105 States over 1979–2000 had RSD values of 57–101%. Habib et al. (2006)
106 found by using the Total Ozone Mapping Spectrometer Absorbing Aerosol
107 Index datasets that the absorbing aerosol column burdens averaged over
108 April–May of 1981–1992 exhibited RSD of 16–30% in different regions of India.
109 Alston et al. (2012) showed by using ground-based measurements at 41 sites
110 in the southeastern United States (29 sites of $\text{PM}_{2.5}$ measurements provided
111 by Environmental Protection Agency and 12 sites by Georgia Dept. of Natural
112 Resources) that monthly-mean $\text{PM}_{2.5}$ concentrations in that region had APDM
113 values of 5–10% over 2000–2009. They also showed by using three sets of
114 aerosol optical depth (AOD) from Moderate Resolution Imaging
115 Spectroradiometer (MODIS) Terra, Multi-angle Imaging Spectroradiometer
116 Terra, and MODIS Aqua that monthly-mean AOD in the southeastern United
117 States had RSD values of 15–25% over 2000–2009.

118 Observations and modeling studies have also shown that aerosols in
119 China have large IAVs. Qu et al. (2010) reconstructed PM_{10} aerosol
120 concentrations at 86 Chinese cities using records of air pollution index from

121 summer 2000 to winter 2006, and reported that seasonal-median PM_{10} levels
122 exhibited APDM values of 15–35% in those cities. Yang et al. (2011) collected
123 weekly samples of carbonaceous aerosols over 2005–2008 at two sites in
124 Beijing and reported that year-to-year changes in emission and meteorology
125 altered annual-average fine organic carbon (OC) concentrations at the rural
126 site in Beijing by as much as 27% over the observational time period.

127 The IAVs of aerosols are influenced by both emissions and meteorology.
128 Meteorological parameters influence aerosol concentrations through altering
129 emissions, chemical reactions, transport, and deposition. For example,
130 increases in temperature enhance chemical production of sulfate in the
131 atmosphere (Aw and Kleeman, 2003; Dawson et al., 2007; Kleeman, 2008)
132 and decrease nitrate aerosol formation (Bellouin et al., 2011; Liao et al., 2006;
133 Pye et al., 2009; Racherla and Adams, 2006). Aerosol concentrations
134 decrease with increasing precipitation as wet deposition provides the main
135 aerosol sink (Balkanski et al., 1993; Dawson et al., 2007), and changes in
136 ventilation (wind speed, mixing depth) have large impacts on aerosols since
137 aerosols are mainly influenced by local meteorological conditions. Tai et al.
138 (2010) found that daily variation in meteorology as described by the multiple
139 linear regression could explain up to 50% of $PM_{2.5}$ variability based on an
140 11-year (1998–2008) observational record over the contiguous United States,
141 with temperature, relative humidity, precipitation, and circulation all being
142 important predictors.

143 Since concentrations of chemical species are the net results from
144 comprehensive physical and chemical processes, the Integrated Process
145 Rates (IPR) (Im et al., 2011) have been used to identify the dominant

146 processes (such as horizontal and vertical transport, emissions of primary
147 species, gas-phase chemistry, dry deposition, cloud processes, and aerosol
148 processes) that influence the concentrations of chemical species in the
149 Community Multi-scale Air Quality model (CMAQ) for episodic events (Jose et
150 al., 2002; Goncalves et al., 2009) as well as for yearly (Zhang et al., 2009) to
151 decadal simulations (Civerolo et al., 2010). The IPR analyses in these studies
152 ranked the roles of different processes in the formation and fate of a chemical
153 species. For example, the horizontal flows and gas-phase chemical reactions
154 in the morning and the vertical flows in the afternoon were found to be the main
155 factors in the formation of surface O₃ during a photochemical pollution episode
156 in the coastal area of South-Western Europe in summer (Goncalves et al.,
157 2009). Recently, a similar process analysis scheme was implemented within
158 the Weather Research and Forecasting model coupled with Chemistry (WRF
159 Chem) to understand the key photochemical and physical processes for the
160 formation of O₃ (Jiang et al., 2012) and PM₁₀ (Jiang et al., 2013).

161 The scientific goals of this work are: (1) to quantify the IAVs in
162 surface-layer aerosol concentrations in China resulted from the variations in
163 meteorological conditions during 2004–2012, using the global
164 three-dimensional chemical transport model GEOS-Chem, and (2) to identify
165 the key meteorological parameters that influenced the IAVs of aerosols in
166 different polluted regions of China by the IPR analyses. Section 2 describes
167 the model, emissions, and numerical experiments. Section 3 presents
168 simulated distributions of aerosols and IAVs in concentrations of different
169 aerosol species in China averaged over 2004–2012. The key meteorological
170 parameters that influenced IAVs of aerosols are examined by IPR in Section 4.

171 Section 5 discusses the impacts of anthropogenic and natural emissions on
172 IAVs of aerosols in China.

173

174 **2 Model description and numerical experiments**

175 **2.1 GEOS-Chem Model**

176 We simulated aerosols using the global chemical transport model
177 GEOS-Chem (version 9-01-02) driven by the GEOS-5 assimilated
178 meteorological fields from the Goddard Earth Observing System of the NASA
179 Global Modeling and Assimilation Office. We used the nested-grid capability of
180 the GEOS-Chem model over East Asia (11°S–55°N, 70°–150°E) with a
181 horizontal resolution of 0.5° latitude by 0.667° longitude and 47 vertical layers
182 up to 0.01hPa (Chen et al., 2009). Chemical boundary conditions were from
183 the global simulations performed at 4° x 5° horizontal resolution.

184 The GEOS-Chem model has fully coupled O₃-NO_x-hydrocarbon chemistry
185 and aerosols including sulfate (SO₄²⁻), nitrate (NO₃⁻), ammonium (NH₄⁺) (Park et
186 al., 2004; Pye et al., 2009), OC and BC (Park et al., 2003), sea salt (Alexander
187 et al., 2005), and mineral dust (Fairlie et al., 2007). The gas-aerosol
188 partitioning of nitric acid and ammonia is calculated using the ISORROPIA II
189 thermodynamic equilibrium model (Fountoukis and Nenes, 2007). Wet
190 deposition of soluble aerosols and gases follows the scheme of Liu et al. (2001)
191 and dry deposition follows a standard resistance-in-series model of Wesely
192 (1989). We do not examine IAVs of mineral dust and sea salt aerosols in this
193 study, because sea salt aerosol is not a major aerosol component in China
194 based on measurements (Ye et al., 2003; Duan et al., 2006) and mineral dust
195 aerosol simulation has very large uncertainties (Fairlie et al., 2007; Fairlie et al.,

196 2010).

197 Considering the large uncertainties in chemistry schemes of secondary
198 organic aerosol (SOA), SOA in our simulations was assumed to be the 10%
199 carbon yield of OC from biogenic terpenes (Park et al., 2003) and 2% carbon
200 yield of OC from biogenic isoprene (van Donkelaar et al., 2007).

201 **2.2 Emissions**

202 Global emissions of aerosol precursors and aerosols in the GEOS-Chem
203 model generally follow Park et al. (2003) and Park et al. (2004). Anthropogenic
204 emissions of NO_x, SO₂, BC, and OC (including emissions from power, industry,
205 residential, and transportation) in the Asian domain are overwritten by David
206 Streets' 2006 emission inventory (<http://mic.greenresource.cn/intex-b2006>)
207 (Zhang et al., 2009) in this work. Estimates of NH₃ emissions in China showed
208 large uncertainties in previous studies (Streets et al., 2003; Kim et al., 2006;
209 Zhang et al., 2010b; Huang et al., 2011; Huang et al., 2012). We used in our
210 simulations the most recent estimate of NH₃ emissions in China by Huang et al.
211 (2012), which was 9.8 Tg yr⁻¹, instead of 13.5 Tg yr⁻¹ from Streets et al. (2003).
212 Monthly variations in SO₂, NO_x, and NH₃ follow those in Wang et al. (2013).
213 Table 1 summarizes year 2006 annual emissions of NO_x, SO₂, NH₃, OC, and
214 BC in Asia (70–150°E, 11°S–55°N) and eastern China (98–130°E, 20–55°N).

215 Natural NO_x emissions from lightning were described by Sauvage et al.
216 (2007) and Murray et al. (2012), and those from soil were described by Yienger
217 and Levy (1995). Natural NH₃ emissions from soil, vegetation, and the oceans
218 were from the Global Emissions Inventory Activity inventory (Bouwman et al.,
219 1997). Biomass burning emissions were from the monthly Global Fire
220 Emissions Database-v2 inventory (van der Werf et al., 2006). Biogenic VOC

221 emissions were calculated from Model of Emissions of Gases and Aerosols
222 from Nature (Guenther et al., 2006).

223 **2.3 Numerical Experiments**

224 To quantify IAVs of aerosols over 2004-2012, we performed the following
225 simulations of aerosols in China using the GEOS-Chem model driven by the
226 GEOS-5 meteorological fields:

227 (1) ANNmet: The simulation to examine how the IAVs of aerosols were
228 influenced by variations in meteorological parameters. Meteorological
229 fields, natural emissions, and biomass burning emissions were allowed to
230 vary from 2004 to 2012, while anthropogenic emissions were kept at the
231 year 2006 values.

232 (2) ANNmet_ATM: Sensitivity simulation for 2004–2012 to examine the
233 sensitivity of IAVs of aerosols to variations in atmospheric conditions alone.
234 All natural emissions (such as soil NO_x, lightning NO_x as well as biogenic
235 sources) that were sensitive to meteorological parameters were turned off.
236 Anthropogenic emissions were kept at the year 2006 values.
237 Meteorological fields and biomass burning emissions were allowed to vary
238 from 2004 to 2012.

239 (3) ANNemis: The simulation to examine how the IAVs of aerosols were
240 influenced by variations in anthropogenic emissions. Anthropogenic and
241 biomass burning emissions were allowed to vary from 2004 to 2012.
242 Meteorological parameters and hence natural emissions were kept at the
243 year 2006 values.

244 (4) ANNall: Simulation of aerosols for years 2004–2012 with yearly varying
245 meteorological parameters, biomass, natural and anthropogenic emissions.

246 The IAVs in anthropogenic emissions over 2004–2012 were obtained by
247 using scaling factors; the annual scaling factors for NO_x were taken from
248 Zhang et al. (2012) and those for SO₂, OC, and BC were taken from Lu et
249 al. (2011).

250 In ANNmet, the IAVs in meteorological fields influenced aerosol
251 concentrations in two ways. First, changes in meteorological parameters
252 influenced chemical reactions, transport, and deposition of aerosols. Second,
253 precursor emissions from natural sources varied with meteorological fields. We
254 performed one sensitivity simulation ANNmet_ATM with natural emissions
255 turned off. The differences between ANNmet and ANNmet_ATM represent the
256 differences in IAVs with and without natural emissions. Biomass burning
257 emissions were partly anthropogenic and partly natural, which were allowed to
258 vary over 2004–2012 in all the simulations. Comparison of ANNmet and
259 ANNemis tells us the relative importance of variations in meteorological
260 parameters and anthropogenic emissions in IAVs of aerosols.

261 The presentations of the IAVs of aerosols will be focused on three polluted
262 regions in China, North China (NC, 32–42°N, 110–120°E), South China (SC,
263 22–32°N, 110–120°E), and the Sichuan Basin (SCB, 27–33°N, 102–110°E), as
264 defined in Fig. 1.

265

266 **3 Simulated IAVs of aerosols resulted from IAVs of meteorological** 267 **parameters alone**

268 **3.1 Simulated distributions of aerosol concentrations**

269 Figure 2 shows simulated seasonal-mean surface-layer concentrations of
270 sulfate, nitrate, ammonium, OC, BC, and PM_{2.5} (sum of sulfate, nitrate,

271 ammonium, OC, and BC) averaged over 2004–2012 of simulation ANNmet.
272 Simulated aerosol concentrations were high over polluted eastern China
273 throughout the year. Sulfate concentrations in NC were 15–20 $\mu\text{g m}^{-3}$ in
274 June-July-August (JJA) with the strong photochemistry in that season, and the
275 concentrations over SC showed small values of 3–10 $\mu\text{g m}^{-3}$ in JJA as a result
276 of the large precipitation associated with the summer monsoon. The maximum
277 sulfate concentrations of 25–30 $\mu\text{g m}^{-3}$ were simulated over SCB in
278 December-January-February (DJF), as a result of the large SO_2 emissions
279 from winter heating. Our simulated seasonal variations in sulfate
280 concentrations agree well with those in Wang et al. (2013).

281 Simulated nitrate concentrations over NC were in the range of 15–40 μg
282 m^{-3} throughout the year, with maximum concentrations of about 40 $\mu\text{g m}^{-3}$ in
283 DJF. High NO_x emissions and low temperatures favored nitrate formation in
284 DJF. Nitrate concentrations showed values of 5–15 μg over SC and SCB m^{-3} in
285 JJA when temperatures were the highest. Simulated ammonium
286 concentrations were in the range of 5–20 $\mu\text{g m}^{-3}$ over NC, SC, and SCB, with
287 seasonal variations in these regions following those of nitrate.

288 The simulated distributions of OC and BC were similar to those of their
289 emissions, with the highest values in NC. Simulated OC and BC
290 concentrations were high in DJF and September-October-November (SON)
291 and low in March-April-May (MAM) and JJA, owing to the seasonal variations
292 in precipitation.

293 Simulated $\text{PM}_{2.5}$ concentrations were in the range of 70–90 $\mu\text{g m}^{-3}$ in NC
294 throughout the year, which were generally higher than the concentrations in
295 SC and SCB. In the surface layer, nitrate was predicted to be the most

296 abundant aerosol species over eastern China, followed by sulfate, ammonium,
297 OC, and BC. Wang et al. (2013) reported that high nitrate in the GEOS-Chem
298 model is likely caused by the overestimate of NH_3 emissions, and Kharol et al.
299 (2013) demonstrated that the persistent nitrate in GEOS-Chem in China is,
300 overall, as much linked to high NO_x emissions as it is to high NH_3 emissions.

301 **3.2 Simulated IAVs of aerosols**

302 Figures 3 and 4 show, respectively, the MAD and APDM values of seasonal
303 mean surface-layer concentrations of sulfate, nitrate, ammonium, OC, BC, and
304 $\text{PM}_{2.5}$ simulated in ANNmet. We also present the domain-averaged values of
305 MAD and APDM for NC, SC, and SCB in Tables 2 and 3.

306 The MAD and APDM values of sulfate, nitrate, and ammonium from
307 ANNmet indicated that concentrations of these species in NC had larger IAVs
308 than those in SC (Tables 2 and 3). Over NC, SC, and SCB, sulfate aerosol
309 showed regional mean APDM values of 10–24%. The largest IAVs of sulfate
310 were found over NC in DJF, with APDM values of exceeding 24%. These IAVs
311 were significant, considering that these were averages over 2004–2012; year
312 by year variations can be larger than the averages reported here. Our
313 simulated IAV of sulfate was close to the IAV of 14–20% reported by Gong et al.
314 (2010) for sulfate at the Canadian High Arctic.

315 The MAD and APDM values obtained in simulation ANNmet showed that
316 nitrate concentrations also had large IAVs (Tables 2 and 3). The APDM values
317 were 13–18% over NC where nitrate concentrations were the highest. The
318 APDM values of nitrate in NC did not show large variations with season, which
319 were larger than the APDM values of sulfate in JJA but smaller than those of
320 sulfate in DJF. The distribution and magnitude of the APDM values of

321 ammonium generally followed those of nitrate over polluted eastern China.

322 The spatial pattern of either MAD or APDM of OC was similar to that of BC
323 in all seasons. OC exhibited seasonal mean APDM values of 6–12% in
324 polluted NC, SC, and SCB throughout the year. Because BC is a chemically
325 inert tracer, the IAVs in BC obtained in ANNmet were caused by the variations
326 in transport and deposition. The APDM values of BC were about the same as
327 those of OC, except that the APDM values of BC were smaller in NC in MAM
328 and in SC in DJF, MAM, and JJA.

329 The IAVs of PM_{2.5} concentrations were the largest in NC; the regional
330 mean APDM values were 17%, 14%, 14%, and 11% in DJF, MAM, JJA, and
331 SON (Table 3), respectively. Over SC, the maximum APDM value of 14% was
332 found in JJA and the highest APDM values were about 9% in other seasons
333 (Table 3). Over SCB, PM_{2.5} showed the smallest IAVs among all regions, with
334 APDM values of 8–9% in all seasons.

335 **3.3 Comparisons of simulated IAVs of aerosols with measurements**

336 Simulated concentrations of sulfate, nitrate, and ammonium aerosols in China
337 have been evaluated in the study of Wang et al. (2013) and those of
338 carbonaceous aerosols in China have been evaluated by Fu et al. (2012), both
339 of which used the same one-way nested-grid capability of the GEOS-Chem.
340 Wang et al. (2013) found that simulated concentrations of sulfate, nitrate and
341 ammonium at 22 sites in East Asia exhibited annual mean biases of –10%,
342 +31%, and +35%, respectively, and Fu et al. (2012) showed that the simulated
343 annual mean concentrations of BC and OC averaged over rural and
344 background sites were underestimated by 56% and 76%, respectively.

345 For the purpose of this study, we evaluated the model's performance in

346 simulating the IAVs of aerosols by comparing simulated aerosol optical depths
347 (AODs) with satellite measurements, because of the lack of long-term
348 ground-based measurements in China (Chan and Yao, 2008). The Level 3
349 MODIS/Terra monthly products (MOD08_M3,
350 <http://ladsweb.nascom.nasa.gov/>) with $1^{\circ}\times 1^{\circ}$ equal-angle global grid were
351 obtained from NASA LAADS (Level 1 and Atmosphere Archive and Distribution
352 System). Collections 5 and 5.1 contain the time series of AODs from March
353 2000 to the present. We used AODs at the 550 nm wavelength, which
354 incorporated only the highest quality retrievals. Over East Asia, the MODIS
355 measurements have been well validated through many studies (Chin et al.,
356 2004; Park et al., 2011). In the GEOS-Chem model, the AODs of sulfate,
357 nitrate, ammonium, OC, BC, sea salt, and dust aerosols were calculated
358 based on aerosol mass concentration, extinction efficiency, effective radius,
359 particle mass density, and the assumed aerosol size distribution (Drury et al.,
360 2010). The hygroscopic growth of each aerosol species with relative humidity
361 was accounted for, using the hygroscopic growth factors listed in Martin et al.
362 (2003).

363 We compared simulated and observed IAVs in AODs for the cities of
364 Beijing (39.5°N , 116.2°E), Changsha (28.1°N , 112.6°E) and Chengdu (30.7°N ,
365 104.0°E) (Fig. 5), which were chosen to represent the model performance in
366 NC, SC and SCB regions, respectively. The correlation coefficients between
367 observed and modeled monthly mean AODs from ANNmet simulation were
368 0.83, 0.34, 0.13 for Beijing, Changsha, and Chengdu, respectively. The large
369 correlation coefficient for Beijing indicates that the model was able to capture
370 to some extent the observed IAVs in NC. The small correlation coefficients in

371 Chengdu can be explained in part by the complex topography and satellite
372 limitations such as cloud contamination (Xia et al., 2004). Note that in
373 simulation ANNmet, simulated AODs correlated well with the simulated column
374 burdens and surface-layer concentrations of $PM_{2.5}$; the correlation coefficients
375 between simulated monthly mean AODs and column burdens (surface-layer
376 concentrations) of $PM_{2.5}$ were 0.95 (0.81), 0.94 (0.81), and 0.91 (0.77) over
377 Beijing, Changsha, and Chengdu, respectively. The IAVs of observed AODs
378 agreed fairly well with the IAVs of surface-layer aerosol concentrations. For
379 example, the seasonal-mean APDM values of observed AODs were 18%
380 (DJF), 15% (MAM), 24% (JJA), and 16% (SON) for Beijing, close to the
381 seasonal-mean APDM values of surface-layer $PM_{2.5}$ shown in Fig. 4.

382

383 **4 Understanding the IAVs of aerosols by process analyses**

384 **4.1 IAVs of meteorological parameters**

385 Figure 6 shows seasonal mean temperature, specific humidity, precipitation,
386 850 hPa zonal and meridional winds for DJF and JJA. All these meteorological
387 fields were averaged over years 2004–2012. Over central and eastern China,
388 temperature, specific humidity and precipitation generally exhibited much
389 larger values in JJA than in DJF. At 850 hPa, strong westerlies were found in
390 NC in DJF, and prevailing southerlies occurred in JJA, reflecting the typical
391 features of winds in China.

392 Figures 7 and 8 show, respectively, the MAD and APDM values of
393 surface-layer temperature, specific humidity, and precipitation for DJF and JJA.
394 The APDM values of temperature in DJF were generally larger than those in
395 JJA. Piao et al. (2003) also showed that the largest IAV of temperature in

396 eastern China was in winter (December and February) based on the
397 reanalyzed temperatures over 1982–1999. Specific humidity showed APDM
398 values of 6–20% in DJF and of 2–8% in JJA over central and eastern China.
399 For precipitation, APDM values of 20–80% and 10–30% were calculated in
400 DJF and JJA, respectively, and the APDM values of precipitation were larger in
401 NC than in SC, which agreed with those reported in Qian and Lin (2005). The
402 variations in temperature and specific humidity can influence chemical
403 reactions of sulfate, nitrate and ammonium, while those in precipitation are
404 important for wet deposition of all aerosol species. The relatively large APDM
405 values of these meteorological parameters in DJF suggested large IAVs of
406 aerosols in this season.

407 **4.2 Process analyses**

408 The concentrations of aerosols are determined by emissions, chemical
409 reactions, transport, and deposition. Therefore, the IAVs of aerosols are
410 influenced by IAV of each of these processes. The weighted contribution of
411 each process to IAV of an aerosol is estimated here using

$$412 \quad \%PC_i = MAD_i / \sum_i^n MAD_i \quad (5)$$

413 where n is the number of processes considered, MAD_i is the MAD value of
414 process i and $\%PC_i$ is the relative contribution of process i to the sum of the
415 contributions from all processes (Im et al., 2011). Once the most important
416 processes are selected from this approach, meteorological variables to which
417 the processes are sensitive to are classified as the key meteorological
418 parameters that lead to IAV of the aerosol. Since we aimed to examine the
419 IAVs in surface-layer aerosol concentrations, our process analyses for an
420 aerosol species were performed for each region (NC, SC, or SCB) from the

421 surface to 1 km altitude. For an aerosol species, the budget (mass flux from
422 each process) was constructed for the selected region considering the mass
423 balance of this aerosol. Chemical production and removal, transport, as well as
424 wet and dry deposition of the aerosol were diagnosed at every time step and
425 summed over each season in simulation ANNmet.

426 **4.2.1 Sulfate**

427 Processes that influence the IAVs of sulfate concentrations include
428 anthropogenic emissions, formation pathways (gas-phase oxidation of SO₂ by
429 OH, and in-cloud oxidation of SO₂ by ozone and hydrogen peroxide), mass
430 fluxes through the four lateral boundaries and the upper boundary at 1 km, wet
431 and dry deposition. On the basis of simulation ANNmet, Figure 9 shows the
432 mass fluxes and APDM values and Figure 10 shows the MAD values and
433 relative contributions of individual atmospheric processes to sulfate
434 concentrations in different regions for DJF and JJA. Transport fluxes were
435 calculated at the boundaries while other sources and sinks were the sums over
436 the grid cells of the region under 1 km.

437 With respect to sulfate budget over NC in DJF, vertical flux through the top
438 side and gas-phase sulfate formation by reaction of SO₂ with OH had the
439 largest values of 0.041 Tg S season⁻¹ and 0.019 Tg S season⁻¹, respectively
440 (Fig. 9a). Among the four horizontal fluxes through the lateral boundaries of
441 NC, the flux through the south boundary had the largest value of 0.017 Tg S
442 season⁻¹, while that through the west boundary had the smallest value of 0.003
443 Tg S season⁻¹. The in-cloud reactions of SO₂ with O₃ and H₂O₂ contributed
444 relatively small to sulfate formation, by 0.005 Tg S season⁻¹ and 0.004 Tg S
445 season⁻¹, respectively. Wet deposition was 0.006 Tg S season⁻¹ and dry

446 deposition was the smallest flux. Fig. 10 shows that gas-phase oxidation of
447 SO₂ by OH was the most important process that contributed to the IAV of
448 sulfate, with relative contribution of 30%, followed by the vertical flux through
449 the top side (18%). The relative contributions of horizontal fluxes through the
450 lateral boundaries were in the range of 3–11%. Each of the aqueous-phase
451 oxidations of SO₂ by O₃ and H₂O₂ accounted for 6% of the IAV of sulfate. The
452 relative contribution by wet deposition was 8% and that by dry deposition was
453 0.4% (Fig. 10a). In JJA, gas-phase formation of sulfate increased to 0.039 Tg
454 S season⁻¹ with relative contribution of 57%, making it the only process that
455 determined the IAV of sulfate concentration over NC in this season. Since
456 gas-phase oxidation of SO₂ is sensitive to temperature and humidity (Yao et al.,
457 2002; Zhang et al., 2012), we conclude that the variations in temperature and
458 specific humidity were the key factors that drove the IAV of sulfate over NC.
459 The relatively high MAD values of temperature in NC in both DJF and JJA (Fig.
460 7) as well as the large MAD values of specific humidity over NC in JJA (Fig. 7)
461 support the above conclusion.

462 Similar analyses were performed for sulfate in SC. In DJF, the gas-phase
463 formation of sulfate and wet deposition were the dominant source and sink of
464 sulfate in this region (Fig. 9b). The vertical flux, flux through the south
465 boundary, and wet deposition had the largest contributions to IAV of sulfate
466 with relative contributions of 24%, 18%, and 15%, respectively, indicating that
467 wind and precipitation were the main meteorological factors that determined
468 the IAV of sulfate concentration over SC in DJF. In JJA, sulfate formation from
469 the reaction of SO₂ with OH had a large value of 0.030 Tg S season⁻¹, but this
470 process had a very low APDM value of 6% (Fig. 9b). Fig. 10b showed that wet

471 deposition and in-cloud oxidation of SO_2 by H_2O_2 were the prevailing
472 processes that contributed, respectively, 20% and 18% to the IAV of sulfate
473 concentration in SC. These two processes corresponded well with the high
474 MAD values of precipitation in JJA as shown in Fig. 7.

475 With respect to sulfate budget over the SCB in DJF, vertical flux through
476 the top side, wet deposition, and gas-phase formation of sulfate had the
477 largest values of 0.050, 0.023, and 0.019 Tg S season^{-1} , respectively (Fig. 9c).
478 The vertical flux through the top side was the most important process that
479 contributed to the IAV of sulfate with a relative contribution of 33%, followed by
480 wet deposition (23%) and in-cloud oxidation of SO_2 by O_3 (12%) (Fig. 10c). We
481 can infer from these analyses that wind (or convergence of winds) was the
482 dominant meteorological factor to influence IAV of sulfate over SCB in DJF. In
483 JJA, wet deposition was the largest contributor to the IAV of sulfate
484 concentration with relative contribution of 35%, which corresponded well with
485 the high MAD values of precipitation in JJA (Fig. 7).

486 **4.2.2 Nitrate**

487 The processes that influence nitrate concentration include gas-to-aerosol
488 conversion of HNO_3 to form nitrate, mass fluxes through the four lateral
489 boundaries and the upper boundary at 1 km, wet and dry deposition. Among all
490 the processes, gas-to-aerosol conversion was found to be the key process that
491 drove the IAV of nitrate over the three studied regions. In DJF (JJA),
492 gas-to-aerosol conversion was calculated to have relative contributions of 56%
493 (71%), 45% (73%), and 62% (73%) in NC, SC, and SCB, respectively (Fig. 12),
494 indicating that the IAVs in temperature and specific humidity drove the IAVs of
495 nitrate in these polluted regions. As reported by Dawson et al. (2007),

496 temperature and humidity have the largest influences on gas-to-aerosol
497 partitioning of HNO₃.

498 **4.2.3 Organic carbon**

499 Processes that influence IAV of OC include mass fluxes through the four
500 lateral boundaries and the upper boundary, emissions from anthropogenic,
501 biomass, biofuel and biogenic sources, as well as wet and dry deposition. With
502 respect to OC budget over NC, emissions had the largest mass fluxes in OC
503 budget in both DJF and JJA (Fig. 13). However, Fig. 14 shows that transport
504 fluxes were the most important processes to drive the IAV of OC in both
505 seasons. The relative contributions of transport fluxes were in the range of
506 5–23% in DJF and 8.0–25% in JJA. Wet deposition made contributions of 9%
507 in DJF and 11% in JJA. Note that the relative contributions of biogenic and
508 biomass emissions were very small.

509 Similar analysis for SC showed that biofuel emission, fluxes through the
510 north and south boundaries had the largest values in OC budget in DJF (Fig.
511 13). Vertical transport, transport through the south boundary, and flux through
512 the north boundary had the largest contributions to the IAV of OC in DJF, with
513 relative contributions of 24%, 20%, and 15%, respectively (Fig. 14). In JJA,
514 transport through the north boundary had the largest relative contribution of
515 23%, followed by wet deposition of 16%.

516 Over SCB in DJF, vertical transport had the largest relative contribution of
517 31%, followed by mass flux through the south boundary of 24% (Fig. 14). In
518 JJA, wet deposition, vertical transport, and transport at the south boundary had
519 the largest contributions to the IAV of OC, with relative contributions of 30%,
520 15%, and 11%, respectively.

521 To conclude, wind was the most important meteorological parameter that
522 drove the IAV of OC in the three regions. Precipitation also played a crucial
523 role in JJA over SC and SCB.

524

525 **5 Impacts of anthropogenic emissions and meteorology-sensitive** 526 **natural emissions on IAVs of aerosols**

527 Table 2 and 3 show, respectively, the simulated MAD and APDM values of
528 SO_4^{2-} , NO_3^- , NH_4^+ , OC, BC, and $\text{PM}_{2.5}$ in ANNmet, ANNmet_ATM, ANNemis,
529 and ANNall experiments. As described in Section 2, the comparisons of the
530 APDM values from ANNemis with those obtained in ANNmet indicate the
531 relative importance of anthropogenic emissions and meteorological
532 parameters in the IAVs of aerosols. Based on the annual scaling factors taken
533 from Zhang et al. (2012) and Lu et al. (2011), the annual total emissions of NO_x ,
534 SO_2 , OC, and BC in China had APDM values of 7%, 5%, 3%, and 5%,
535 respectively, over years 2004–2012. In NC, the APDM values of
536 concentrations of all aerosol species obtained in ANNemis were much smaller
537 than those in ANNmet, indicating that the variations in meteorological
538 parameters played more important roles than variations in anthropogenic
539 emissions in driving the IAVs of aerosols. Similar results were found in SC,
540 except that the APDM values of nitrate aerosol driven by variations in
541 emissions alone became close to those driven by variations in meteorological
542 parameters alone. In SCB, the variations in emissions were as important as
543 those in meteorological fields for IAVs of nitrate, ammonium, and $\text{PM}_{2.5}$
544 aerosols in MAM, JJA, and SON. These results in SCB can be explained by
545 the nonlinear responses in nitrate and ammonium aerosols to variations in

546 precursor emissions. Wang et al. (2013) also showed that nonlinear responses
547 of sulfate, nitrate and ammonium to changes in precursor emissions differed
548 by region and season, as meteorological fields were kept at a specific year and
549 the same emission scaling factors were applied to all regions.

550 The differences in MAD and APDM values between ANNmet_ATM and
551 ANNmet represented the IAVs of aerosols caused by meteorology-sensitive
552 natural emissions. The roles of natural emissions were generally small, except
553 that the differences in APDM between ANNmet_ATM and ANNmet were large
554 in DJF and MAM for OC over SC (Table 3), which were caused by the high
555 biogenic emissions in the region (Fu and Liao, 2012).

556

557 **6 Conclusions**

558 We used the nested grid version of the GEOS-Chem model to estimate the
559 role of meteorology in the IAVs of aerosols over China for years 2004–2012.
560 We performed simulations, ANNmet (effects of variations in meteorology
561 alone), and ANNmet_ATM (same as ANNmet except that
562 meteorology-sensitive natural emissions were turned off), ANNemis (effects of
563 variations in anthropogenic emissions alone), and ANNall (combined effects of
564 variations in meteorology and anthropogenic emissions) to identify the key
565 parameters that influence the IAVs of aerosols.

566 We defined two parameters, mean absolute deviation (MAD) and absolute
567 percent departure from the mean (APDM), to quantify the IAVs in
568 concentrations of aerosols over 2004–2012. Results from simulation ANNmet
569 showed that, driven by changes in meteorological parameters alone, the
570 regional mean APDM values of sulfate, nitrate, and OC aerosols were in the

571 ranges of 10–24%, 9–18%, and 6–12%, respectively, over the studied regions
572 (NC, SC, and SCB) throughout the year. As a result of the IAVs of individual
573 aerosol species, simulated PM_{2.5} aerosol concentrations exhibited large IAVs
574 in NC, with regionally averaged APDM values of 17%, 14%, 14%, and 11% in
575 DJF, MAM, JJA, and SON, respectively. Over SC, the IAVs in PM_{2.5} were found
576 to be the largest in JJA; the regional mean APDM value was 14% in JJA and
577 about 9% in other seasons. Concentrations of PM_{2.5} over SCB were simulated
578 to have the smallest IAVs among the polluted regions examined in this work,
579 with the APDM values of 8–9% in all seasons. All aerosol species (sulfate,
580 nitrate, ammonium, black carbon, and organic carbon) were simulated to have
581 the largest IAVs over NC in DJF, corresponding to the large variations in
582 meteorological parameters over NC in DJF.

583 We applied process analyses for sulfate, nitrate and OC to identify key
584 meteorological parameters that led to IAVs of these aerosols over 2004–2012.
585 For sulfate in NC, gas-phase formation of sulfate was found to be the key
586 process that drove the IAV of sulfate in both DJF and JJA, with relative
587 contribution of 30% in DJF and of 57% in JJA, inferring that the variations in
588 temperature and specific humidity jointly determined the IAV of sulfate in NC.
589 Over SC and SCB, the most important process that dominated IAV of sulfate in
590 DJF was found to be the vertical flux through the top side with relative
591 contribution of 24% in SC and of 33% in SCB, and the key process in JJA was
592 found to be wet deposition with relative contributions of 20% and 35% in SC
593 and SCB, respectively. For nitrate, gas-to-aerosol conversion was found to be
594 the key process that dominated the IAVs of nitrate over the three regions, with
595 very high relative contributions of 45–62% in DJF and 71–73% in JJA,

596 indicating that temperature and specific humidity were the major factors that
597 drove the IAV of nitrate in China. For OC, transport was the most important
598 process that influenced the IAV of OC throughout the year, and precipitation
599 also played a crucial role in JJA over SC and SCB, associated with the East
600 Asian summer monsoon precipitation.

601 We also examined the relative importance of anthropogenic emissions and
602 meteorological parameters in the IAVs of aerosols. For all aerosol species
603 (sulfate, nitrate, ammonium, BC, and OC), the APDM values in ANNmet were
604 larger than those in ANNemis in NC and SC, indicating that the variations in
605 meteorological parameters played more important roles than variations in
606 anthropogenic emissions in driving the IAVs of aerosols in these two regions.
607 In SCB, the variations in emissions were found to be as important as those in
608 meteorological fields for IAVs of nitrate, ammonium, and PM_{2.5} aerosols in
609 MAM, JJA, and SON.

610 The IAVs in meteorological fields are mainly associated with natural
611 variability in the climate system; hence the magnitudes of IAVs in aerosol
612 concentrations driven by meteorological parameters have important
613 implications for the effectiveness of short-term air quality control strategies in
614 China. We note that the changes in anthropogenic emissions on longer time
615 scales (for example, decades) may lead to linear trends in simulated aerosol
616 concentrations (Yang et al., 2014). For studies on longer time scales, the MAD
617 and APDM values need to be calculated after detrending the time series,
618 following the approach used in previous studies that examined interannual
619 variations in ozone concentrations (Camp et al., 2003), sea surface
620 temperature, partial pressure of CO₂ (Gruber et al., 2002), sea level pressure

621 (Thompson et al., 1998), and North Atlantic Oscillation index (Jung et al.,
622 2003).

623

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631

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883 eastern China due to the decadal-scale weakening of the East Asian
884 summer monsoon, Geophys. Res. Lett., 39, L09809, doi:
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886

887 Table 1. Summary of annual emissions of aerosols and aerosol precursors in
 888 Asia (70–150°E, 11°S–55°N) and eastern China (98–130°E, 20–55°N)

Species	Asia	Eastern China
NO_x (Tg N yr⁻¹)		
Aircraft	0.08	0.02
Anthropogenic	10.6	6.38
Biomass burning	1.09	0.14
Biofuel	0.02	<0.01
Fertilizer	0.35	0.12
Lightning	1.16	0.28
Soil	0.88	0.27
Total	14.18	7.21
SO₂ (Tg S yr⁻¹)		
Aircraft	0.01	<0.01
Anthropogenic	23.85	15.98
Biomass burning	0.36	0.04
Biofuel	<0.01	<0.01
Volcanoes	3.99	0.07
Ship	0.48	0.05
Total	28.69	16.14
NH₃ (Tg N yr⁻¹)		
Anthropogenic	14.18	7.26
Natural	2.33	0.49
Biomass burning	0.82	0.11
Biofuel	0.8	0.3
Total	18.13	8.16
OC (Tg C yr⁻¹)		
Anthropogenic	1.54	1.03
Biomass burning	4.62	0.67
Biofuel	3.47	1.44
Biogenic	2.5	0.55
Total	12.13	3.69
BC (Tg C yr⁻¹)		
Anthropogenic	1.52	0.96
Biomass burning	0.55	0.05
Biofuel	0.92	0.39
Total	2.99	1.4

Table 2. Simulated MAD values ($\mu\text{g m}^{-3}$) of aerosols in ANNmet, ANNmet_ATM, ANNemis, and ANNall.

Species	DIF				MAM				JJA				SON			
	ANNmet	ANNmet_ATM	ANNemis	ANNall												
NC																
SO ₄ ²⁻	2.0	2.1	0.5	2.0	1.3	1.3	0.5	1.3	1.4	1.4	0.4	1.5	1.5	1.5	0.5	1.6
NO ₃ ⁻	2.3	2.3	1.0	2.6	1.8	1.8	1.5	2.5	2.4	2.3	1.9	3.0	2.1	2.0	1.5	2.3
NH ₄ ⁺	1.2	1.2	0.2	1.2	1.0	0.9	0.4	1.1	1.1	1.1	0.5	1.3	1.0	1.0	0.4	1.0
OC	0.5	0.5	0.2	0.6	0.3	0.3	0.1	0.3	0.3	0.3	0.1	0.3	0.3	0.3	0.1	0.3
BC	0.3	0.3	0.2	0.3	0.1	0.1	0.1	0.2	0.1	0.1	0.1	0.2	0.2	0.2	0.1	0.2
PM _{2.5}	6.1	6.2	1.3	6.4	4.4	4.2	2.1	5.1	5.0	4.9	2.6	5.7	4.9	4.8	1.9	4.8
SC																
SO ₄ ²⁻	1.7	1.8	0.5	1.7	1.2	1.2	0.4	1.2	0.9	0.9	0.3	0.9	1.1	1.1	0.5	1.2
NO ₃ ⁻	1.4	1.4	0.8	1.7	1.2	1.2	1.4	2.0	1.2	1.2	1.3	1.8	1.3	1.3	1.4	1.9
NH ₄ ⁺	0.8	0.8	0.2	0.8	0.7	0.7	0.5	0.9	0.6	0.6	0.4	0.8	0.7	0.7	0.5	0.8
OC	0.5	0.3	0.1	0.5	0.3	0.2	0.1	0.3	0.2	0.2	0.1	0.2	0.3	0.2	0.1	0.3
BC	0.2	0.2	0.1	0.2	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
PM _{2.5}	3.7	3.6	1.4	4.1	3.1	3.0	2.1	4.2	2.8	2.8	2.0	3.6	3.1	3.1	2.3	3.8
SCB																
SO ₄ ²⁻	1.8	1.8	0.8	2.0	1.0	1.0	0.7	1.3	0.9	1.0	0.7	1.1	1.3	1.2	0.9	1.5
NO ₃ ⁻	1.2	1.2	0.7	1.5	0.7	0.7	0.9	1.1	0.8	0.8	0.8	1.3	0.8	0.8	0.9	1.3
NH ₄ ⁺	0.6	0.6	0.2	0.7	0.5	0.5	0.5	0.7	0.5	0.5	0.5	0.7	0.6	0.6	0.5	0.7
OC	0.3	0.3	0.1	0.4	0.2	0.2	0.1	0.3	0.2	0.2	0.1	0.2	0.2	0.2	0.1	0.2
BC	0.2	0.1	0.1	0.2	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
PM _{2.5}	3.1	3.1	1.7	3.7	2.0	2.0	2.1	3.1	2.3	2.4	2.0	3.2	2.5	2.5	2.3	3.4

891 Table 3. Simulated APDM values (%) of aerosols in ANNmet, ANNmet_ATM, ANNemis, and ANNall.

Species	DIF				MAM				JJA				SON			
	ANNmet	ANNmet_ATM	ANNemis	ANNall												
NC																
SO ₄ ²⁻	23.6	23.8	4.2	24.6	17.0	16.7	4.1	18.2	14.7	14.7	4.4	14.6	13.7	13.6	4.3	17.3
NO ₃ ⁻	18.2	18.3	5.0	20.0	15.6	15.7	8.2	19.2	17.0	17.4	11.1	19.6	13.0	13.2	8.0	15.1
NH ₄ ⁺	17.0	17.1	1.9	18.2	15.2	15.1	4.6	17.1	14.9	14.8	5.4	15.6	11.7	11.7	3.4	13.4
OC	12.1	12.1	5.1	13.1	10.7	8.7	5.0	11.3	8.3	8.2	5.1	8.5	7.7	7.6	5.1	8.5
BC	12.0	12.1	3.2	13.9	8.7	8.7	2.9	10.4	8.2	8.3	2.7	9.6	7.5	7.5	3.0	8.9
PM _{2.5}	16.9	16.9	2.5	17.9	13.8	13.8	4.7	15.8	13.7	13.7	5.3	14.5	10.7	10.8	3.6	12.5
SC																
SO ₄ ²⁻	13.9	14.1	4.4	15.5	11.8	11.6	4.6	14.1	15.7	15.1	6.2	17.6	10.6	10.5	4.5	12.1
NO ₃ ⁻	11.5	11.5	5.9	13.0	11.7	11.5	11.0	18.0	17.7	16.7	14.7	24.5	12.2	11.9	10.8	16.6
NH ₄ ⁺	9.3	9.2	2.9	10.9	10.5	10.3	6.4	14.5	16.1	15.4	9.6	19.8	9.5	9.4	5.9	13.6
OC	11.9	7.6	5.0	12.5	10.3	7.4	5.0	10.4	9.2	8.8	5.1	9.3	7.2	6.8	5.1	7.6
BC	8.6	7.8	2.8	10.5	7.7	7.2	2.5	9.8	8.3	8.2	2.1	10.2	6.9	6.9	2.5	7.9
PM _{2.5}	9.2	8.8	3.4	10.6	9.7	9.5	6.2	13.5	13.8	13.4	8.4	17.1	8.6	8.6	5.7	10.7
SCB																
SO ₄ ²⁻	14.4	14.2	5.4	16.6	10.1	9.6	7.4	13.8	10.8	10.2	8.9	15.0	10.7	10.6	7.3	13.7
NO ₃ ⁻	11.4	11.5	6.5	14.3	9.3	9.3	11.9	15.9	12.9	11.3	17.0	21.6	10.5	9.8	10.9	14.8
NH ₄ ⁺	9.0	9.0	3.1	10.3	8.5	8.2	7.9	12.7	10.8	10.0	10.8	16.2	8.7	8.5	6.9	11.1
OC	7.3	6.8	5.0	8.9	6.6	5.5	5.1	8.2	5.8	5.8	5.1	6.9	5.6	5.6	5.1	6.2
BC	7.4	7.4	3.0	9.5	5.3	5.4	2.8	8.1	5.4	5.4	2.5	8.0	5.8	5.8	2.8	7.7
PM _{2.5}	8.8	8.8	4.1	10.5	7.6	7.5	7.4	11.8	9.2	8.8	9.2	14.0	8.0	7.9	6.8	10.5

892 **Figure Captions**

893

894 **Fig. 1.** Polluted regions examined in this study, including North China (NC, 32°–42°N, 110°–120°E), South China (SC, 22°–32°N, 110°–120°E), and the
895 Sichuan Basin (SCB, 27°–33°N, 102°–110°E).
896

897

898 **Fig. 2.** Simulated seasonal-mean surface-layer concentrations ($\mu\text{g m}^{-3}$) of
899 sulfate, nitrate, ammonium, OC, BC, and $\text{PM}_{2.5}$ in ANNmet averaged over
900 2004–2012.

901

902 **Fig. 3.** Mean absolute deviation (MAD, $\mu\text{g m}^{-3}$) of surface-layer concentrations
903 of sulfate, nitrate, ammonium, OC, BC, and $\text{PM}_{2.5}$ simulated in ANNmet for
904 years 2004–2012.

905

906 **Fig. 4.** Absolute percent departure from the mean (APDM, %) of surface-layer
907 concentrations of sulfate, nitrate, ammonium, OC, BC, and $\text{PM}_{2.5}$ simulated in
908 ANNmet for years 2004–2012.

909

910 **Fig. 5.** MODIS AOD (black line, left axis), simulated AOD in ANNall (red dotted
911 line, left axis), simulated AOD in ANNmet (red line, left axis), simulated
912 surface-layer $\text{PM}_{2.5}$ concentrations in ANNmet (green line, $\mu\text{g m}^{-3}$, right axis),
913 and simulated column burden of $\text{PM}_{2.5}$ in ANNmet (blue line, mg m^{-2} , right axis)
914 over polluted cities of (a) Beijing, (b) Changsha, and (c) Chengdu.

915

916 **Fig. 6.** Seasonal mean surface air temperature (K), specific humidity (g kg^{-1}),
917 precipitation (mm d^{-1}), zonal and meridional wind at 850 hPa (m s^{-1}) in DJF and
918 JJA. Winds that were eastward or northward were positive, and those that
919 were westward and southward were negative. Meteorological fields were from
920 the GEOS-5 assimilated meteorological data and were averaged over
921 2004–2012.

922

923 **Fig. 7.** The MAD values of surface air temperature (K), specific humidity (g kg^{-1})
924 and precipitation (mm d^{-1}) in DJF and JJA based on the GEOS-5 assimilated
925 meteorological fields of 2004–2012.

926

927 **Fig. 8.** The APDM values (%) of surface air temperature, specific humidity, and
928 precipitation in DJF and JJA based on the GEOS-5 assimilated meteorological
929 fields of 2004–2012.

930

931 **Fig. 9.** Sulfate budget (mass flux from each process in Tg S season^{-1} , right)
932 and APDM of each flux (%), left) in (a) NC, (b) SC, and (c) SCB obtained from
933 simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport
934 fluxes were calculated at the boundaries while other sources and sinks were

935 calculated as the sums over the grid cells of the region under 1 km.

936

937 **Fig. 10.** The MAD (Tg S season^{-1} , right) and relative contribution (% , left) of
938 each sulfate process (flux) in (a) NC, (b) SC, and (c) SCB obtained from
939 simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport
940 fluxes were calculated at the boundaries while other sources and sinks were
941 calculated as the sums over the grid cells of the region under 1 km.

942

943 **Fig. 11.** Nitrate budget (mass flux from each process in Tg N season^{-1} , right)
944 and APDM of each flux (% , left) in (a) NC, (b) SC, and (c) SCB obtained from
945 simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport
946 fluxes were calculated at the boundaries while other sources and sinks were
947 calculated as the sums over the grid cells of the region under 1 km.

948

949 **Fig. 12.** The MAD (Tg N season^{-1} , right) and relative contribution (% , left) of
950 each nitrate process (flux) in (a) NC, (b) SC, and (c) SCB obtained from
951 simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport
952 fluxes were calculated at the boundaries while other sources and sinks were
953 calculated as the sums over the grid cells of the region under 1 km.

954

955 **Fig. 13.** Organic carbon budget (mass flux from each process in Tg C season^{-1} ,
956 right) and APDM of each flux (% , left) in (a) NC, (b) SC, and (c) SCB obtained
957 from simulation ANNmet. Blue lines are for DJF and red lines are for JJA.
958 Transport fluxes were calculated at the boundaries while other sources and
959 sinks were calculated as the sums over the grid cells of the region under 1 km.

960

961 **Fig. 14.** The MAD (Tg C season^{-1} , right) and relative contribution (% , left) of
962 each organic carbon process (flux) in (a) NC, (b) SC, and (c) SCB obtained
963 from simulation ANNmet. Blue lines are for DJF and red lines are for JJA.
964 Transport fluxes were calculated at the boundaries while other sources and
965 sinks were calculated as the sums over the grid cells of the region under 1 km.

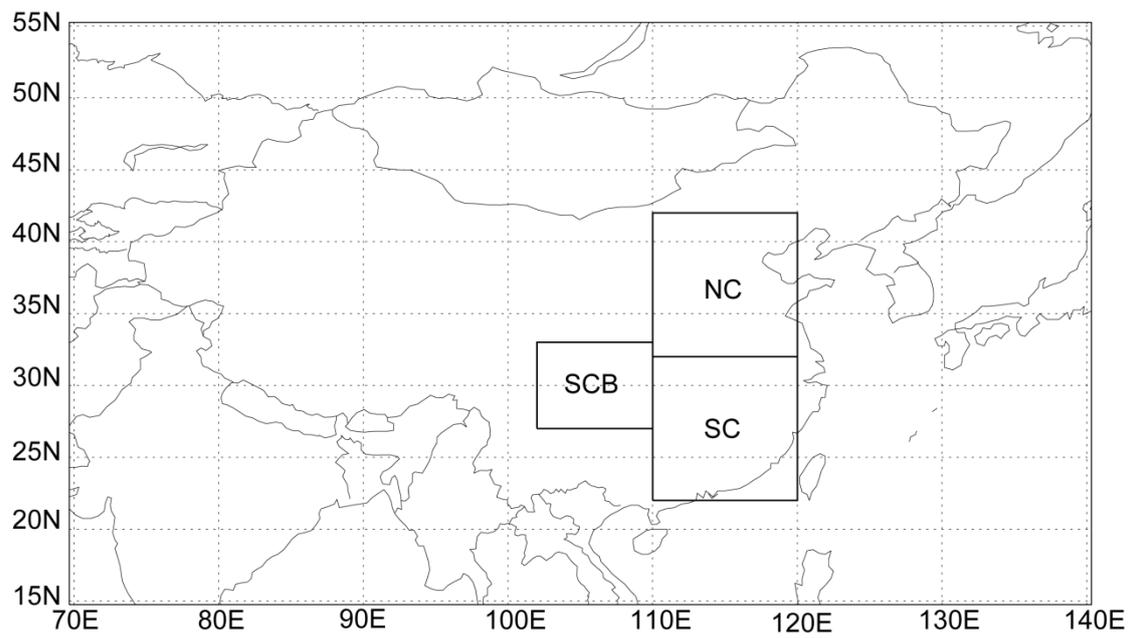


Fig. 1. Polluted regions examined in this study, including North China (NC, 32–42°N, 110–120°E), South China (SC, 22–32°N, 110–120°E), and the Sichuan Basin (SCB, 27–33°N, 102–110°E).

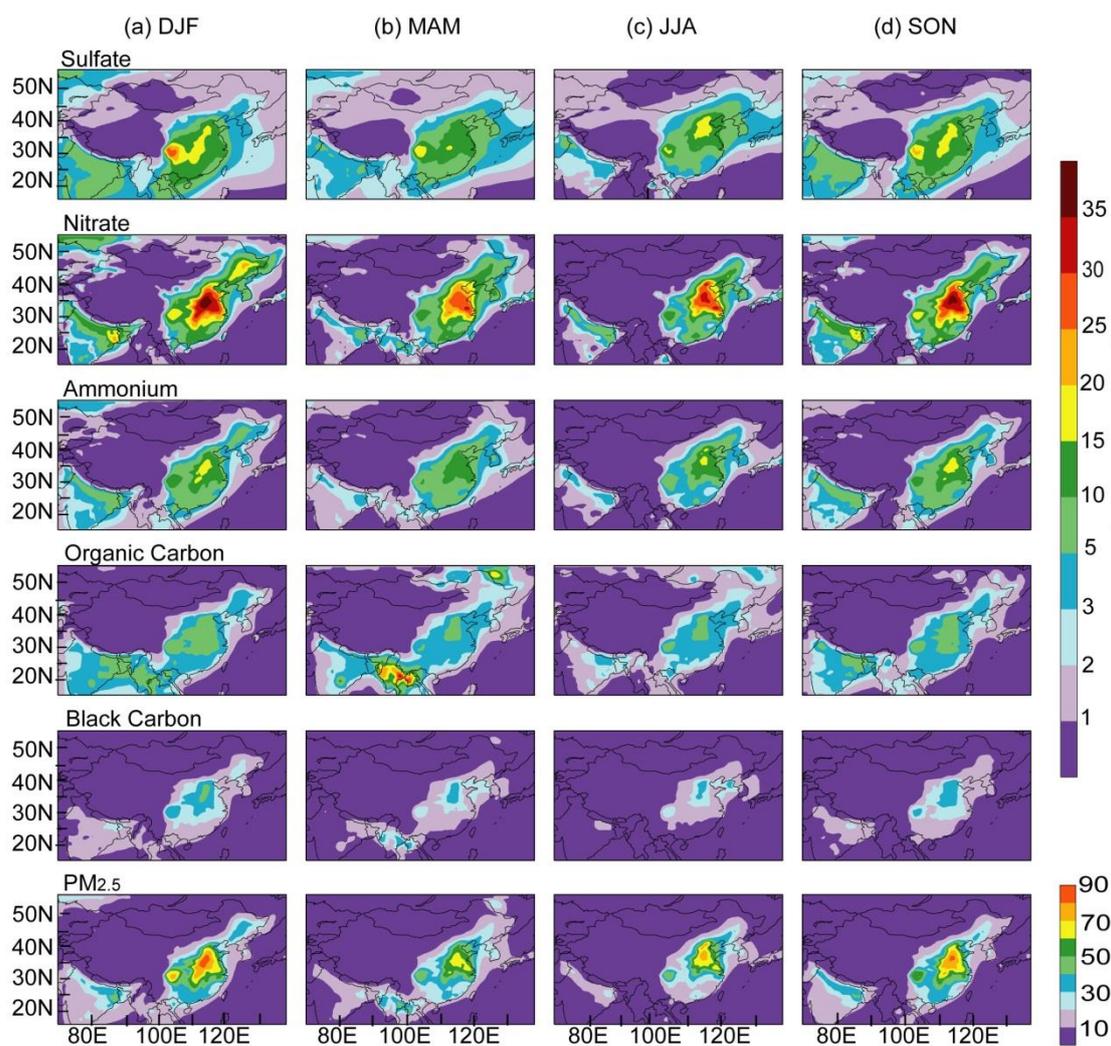


Fig. 2. Simulated seasonal-mean surface-layer concentrations ($\mu\text{g m}^{-3}$) of sulfate, nitrate, ammonium, OC, BC, and PM_{2.5} in ANNmet averaged over 2004–2012.

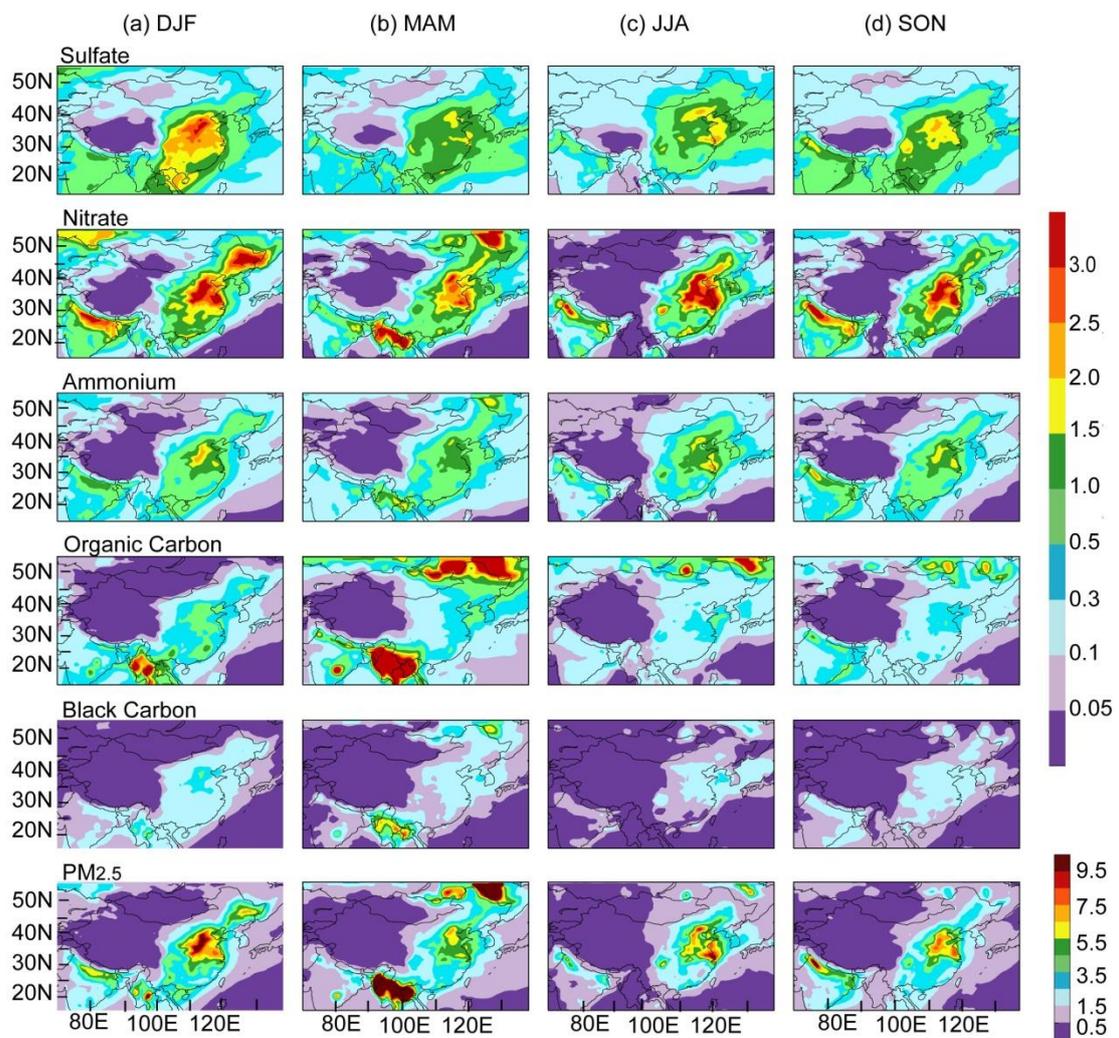


Fig. 3. Mean absolute deviation (MAD, $\mu\text{g m}^{-3}$) of surface-layer concentrations of sulfate, nitrate, ammonium, OC, BC, and PM_{2.5} simulated in ANNmet for years 2004–2012.

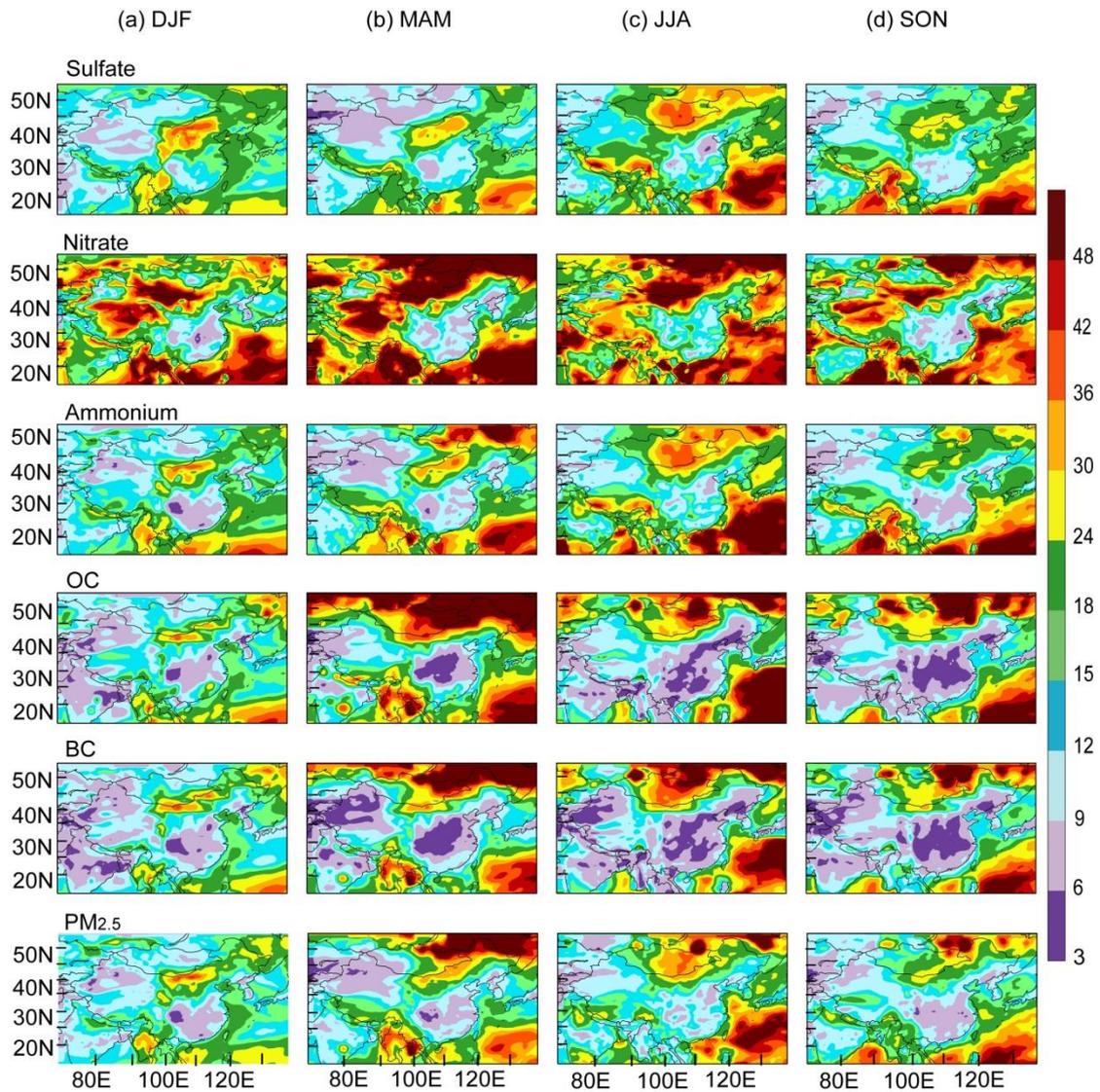


Fig. 4. Absolute percent departure from the mean (APDM, %) of surface-layer concentrations of sulfate, nitrate, ammonium, OC, BC, and PM_{2.5} simulated in ANNmet for years 2004–2012.

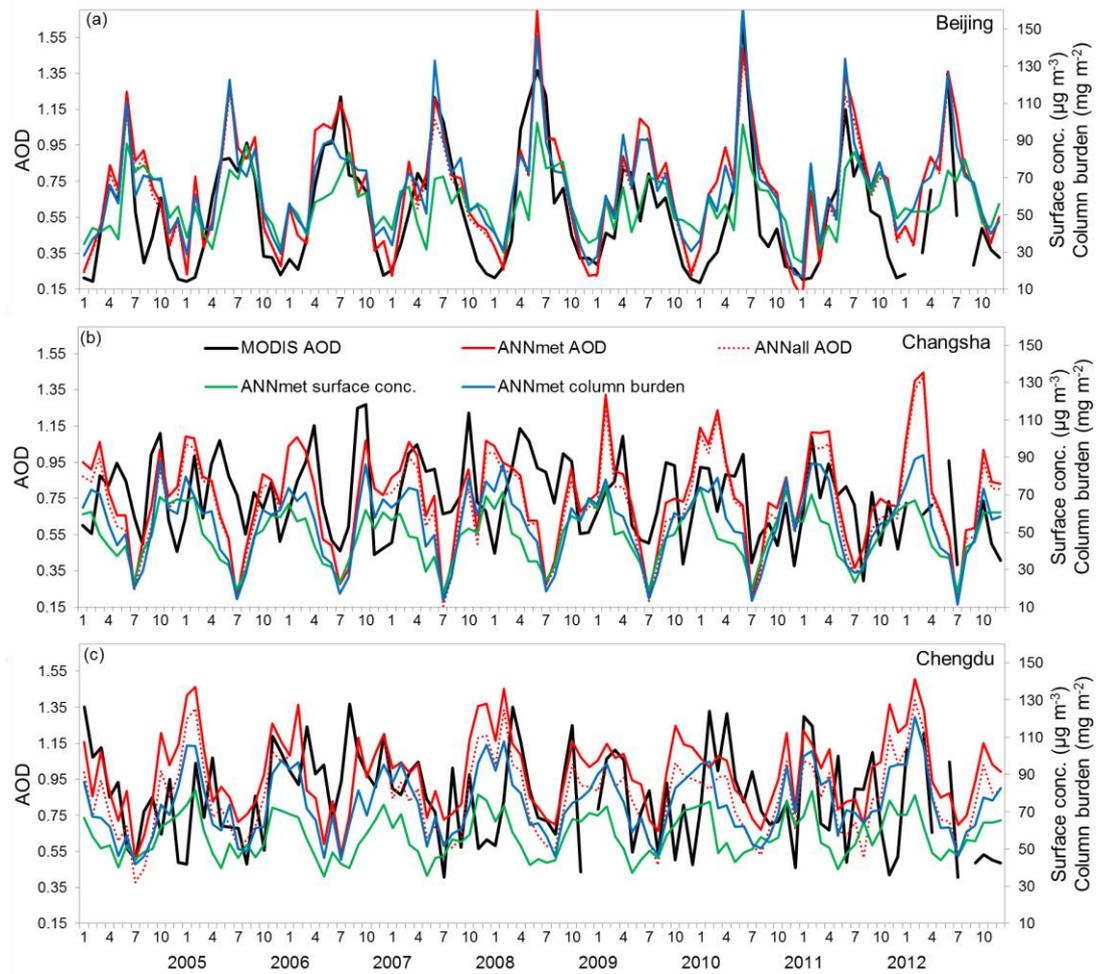


Fig. 5. MODIS AOD (black line, left axis), simulated AOD in ANNall (red dotted line, left axis), simulated AOD in ANNmet (red line, left axis), simulated surface-layer $\text{PM}_{2.5}$ concentrations in ANNmet (green line, $\mu\text{g m}^{-3}$, right axis), and simulated column burden of $\text{PM}_{2.5}$ in ANNmet (blue line, mg m^{-2} , right axis) over polluted cities of (a) Beijing, (b) Changsha, and (c) Chengdu.

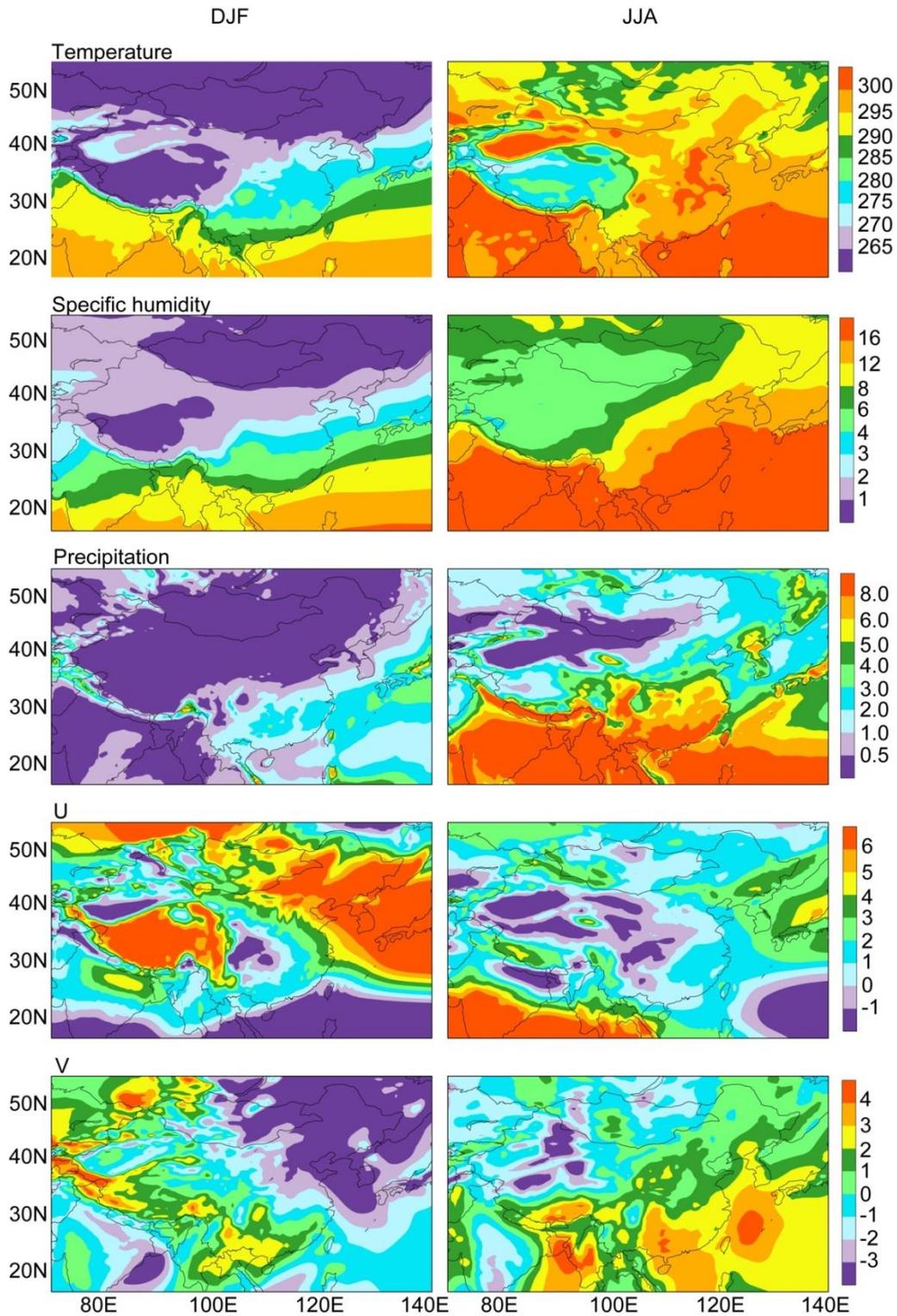


Fig. 6. Seasonal mean surface air temperature (K), specific humidity (g kg^{-1}), precipitation (mm d^{-1}), zonal and meridional wind at 850 hPa (m s^{-1}) in DJF and JJA. Winds that were eastward or northward were positive, and those that were westward and southward were negative. Meteorological fields were from the GEOS-5 assimilated meteorological data and were averaged over 2004–2012.

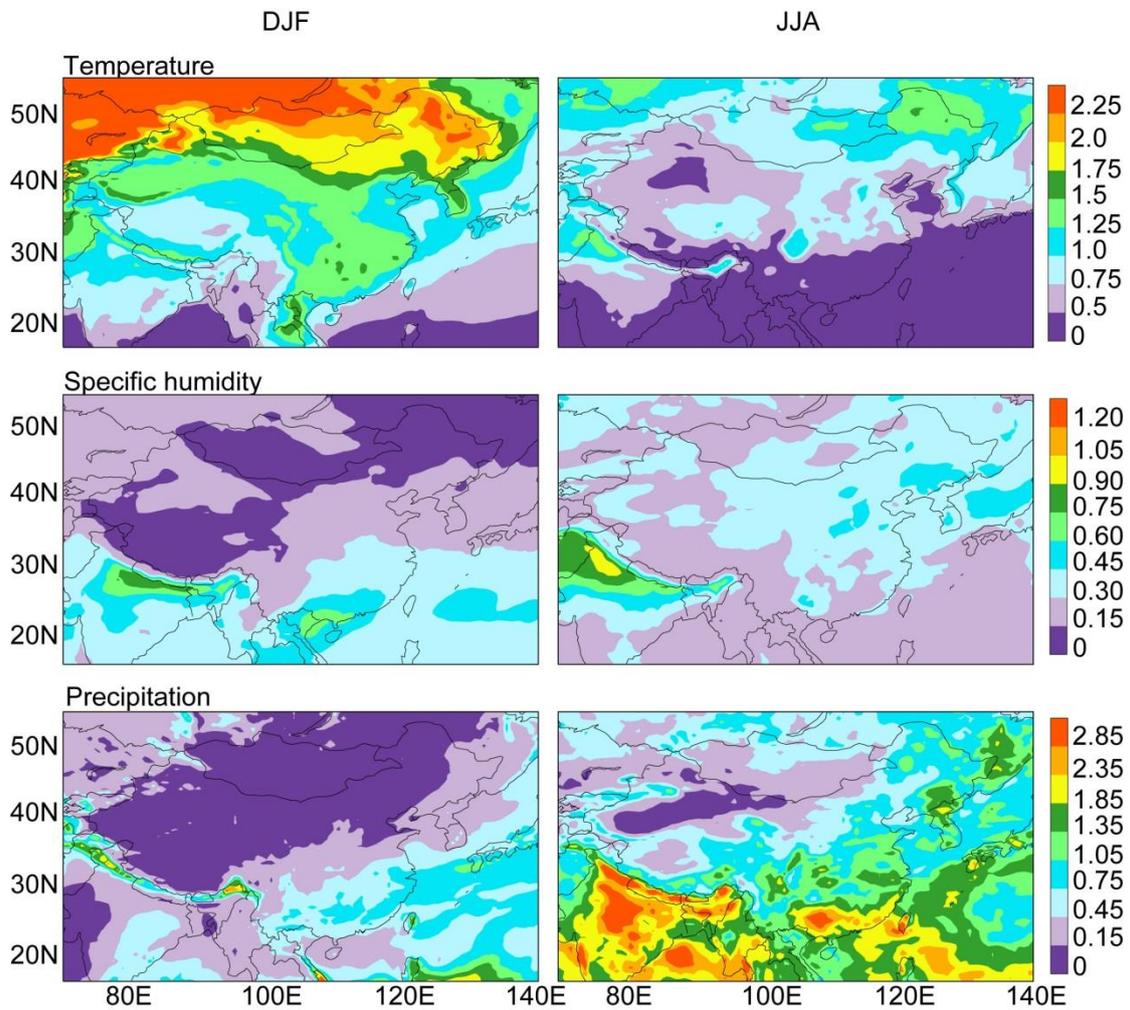


Fig. 7. The MAD values of surface air temperature (K), specific humidity (g kg^{-1}) and precipitation (mm d^{-1}) in DJF and JJA based on the GEOS-5 assimilated meteorological fields of 2004–2012.

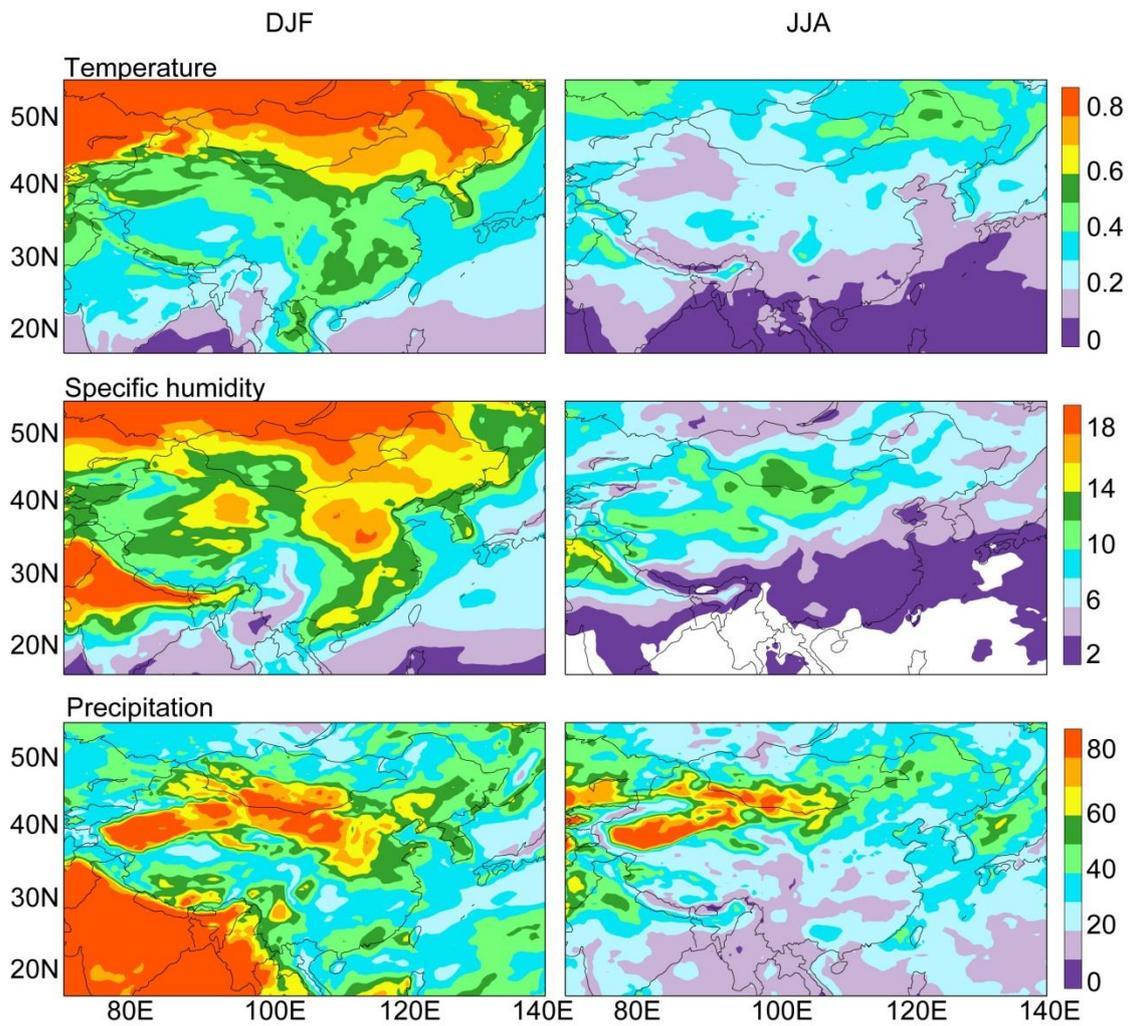


Fig. 8. The APDM values (%) of surface air temperature, specific humidity, and precipitation in DJF and JJA based on the GEOS-5 assimilated meteorological fields of 2004–2012.

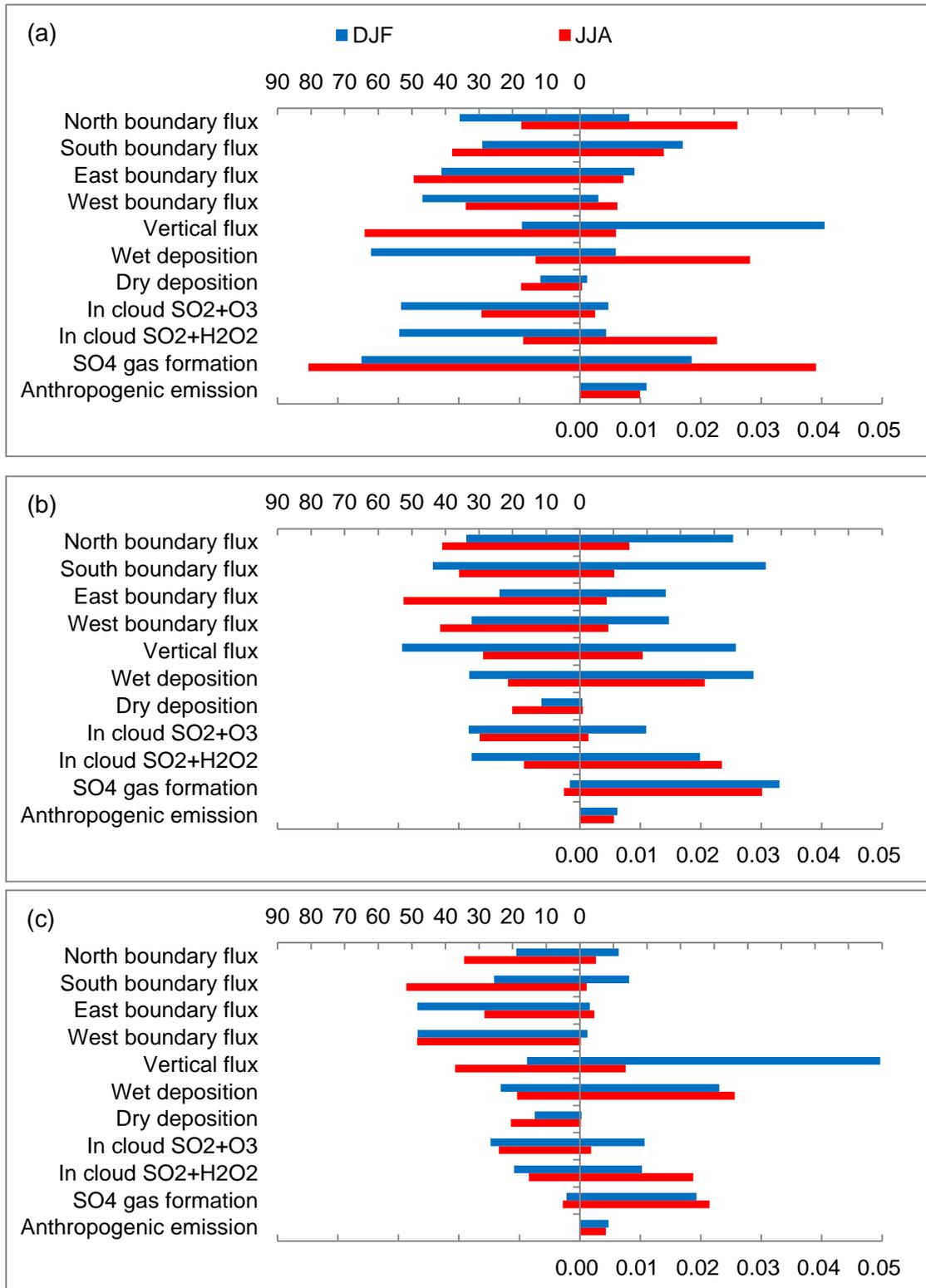


Fig. 9. Sulfate budget (mass flux from each process in Tg S season⁻¹, right) and APDM of each flux (% , left) in (a) NC, (b) SC, and (c) SCB obtained from simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport fluxes were calculated at the boundaries while other sources and sinks were calculated as the sums over the grid cells of the region under 1 km.

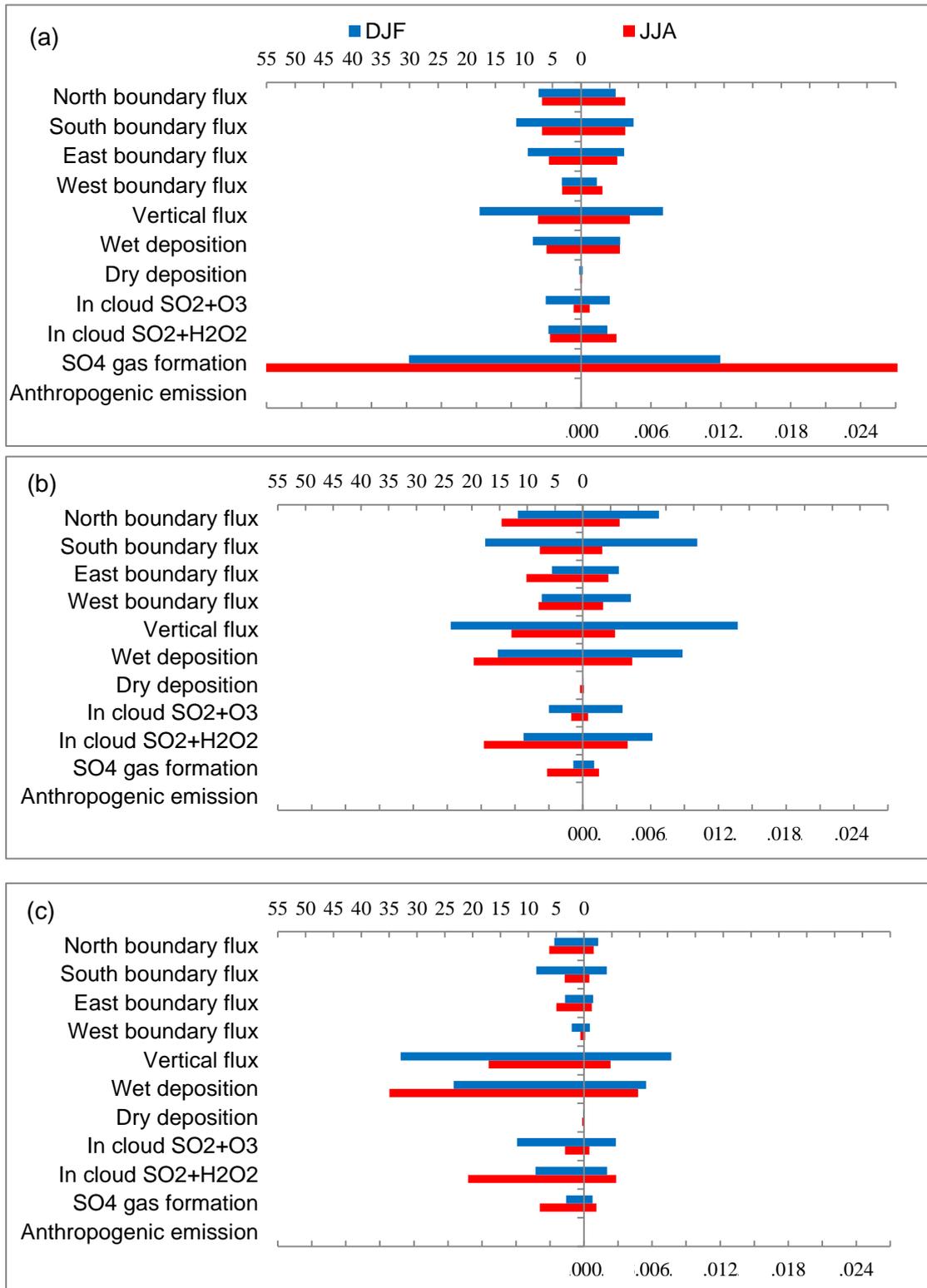


Fig. 10. The MAD (Tg S season⁻¹, right) and relative contribution (% , left) of each sulfate process (flux) in (a) NC, (b) SC, and (c) SCB obtained from simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport fluxes were calculated at the boundaries while other sources and sinks were calculated as the sums over the grid cells of the region under 1 km.

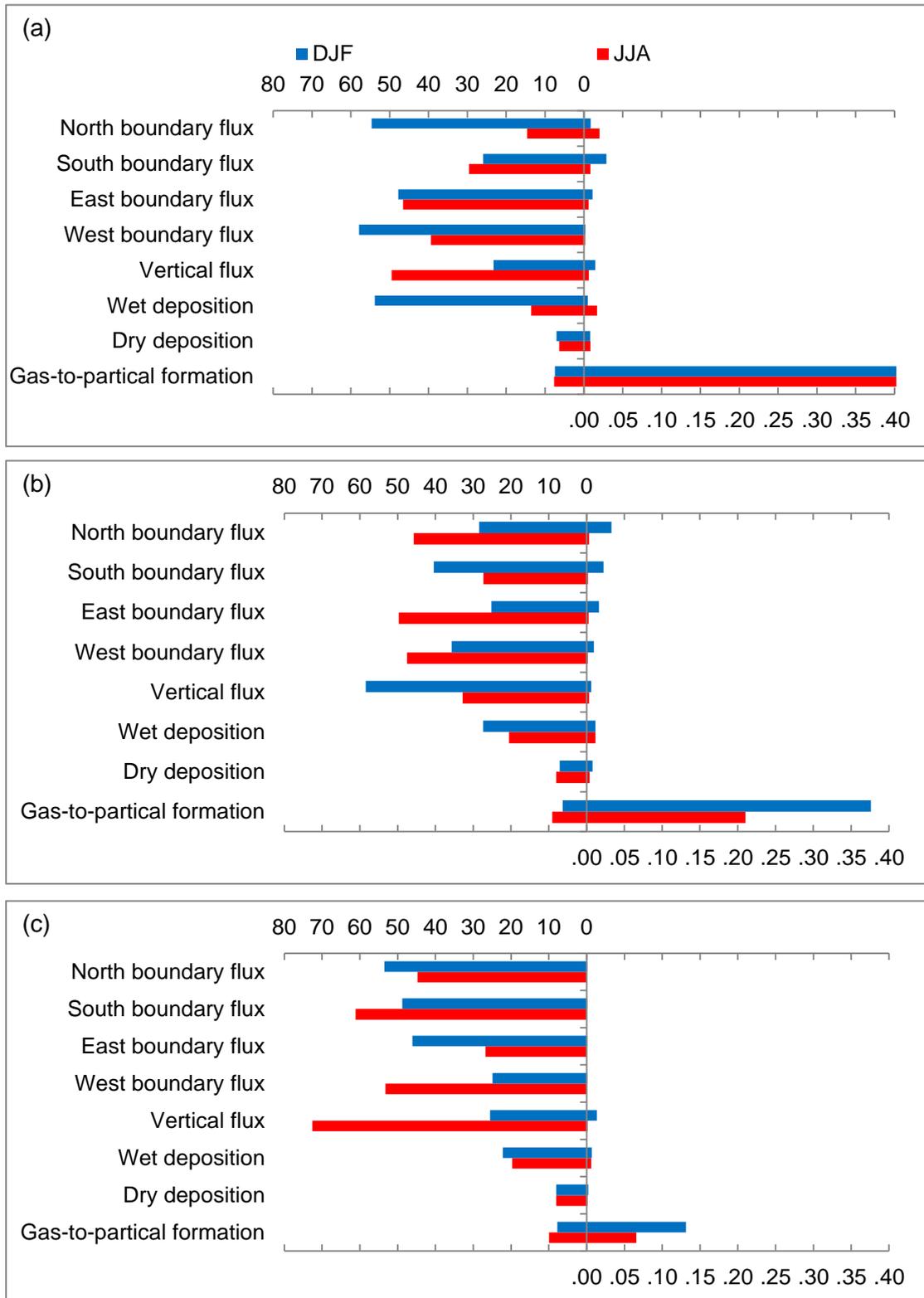


Fig. 11. Nitrate budget (mass flux from each process in Tg N season⁻¹, right) and APDM of each flux (% , left) in (a) NC, (b) SC, and (c) SCB obtained from simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport fluxes were calculated at the boundaries while other sources and sinks were calculated as the sums over the grid cells of the region under 1 km.

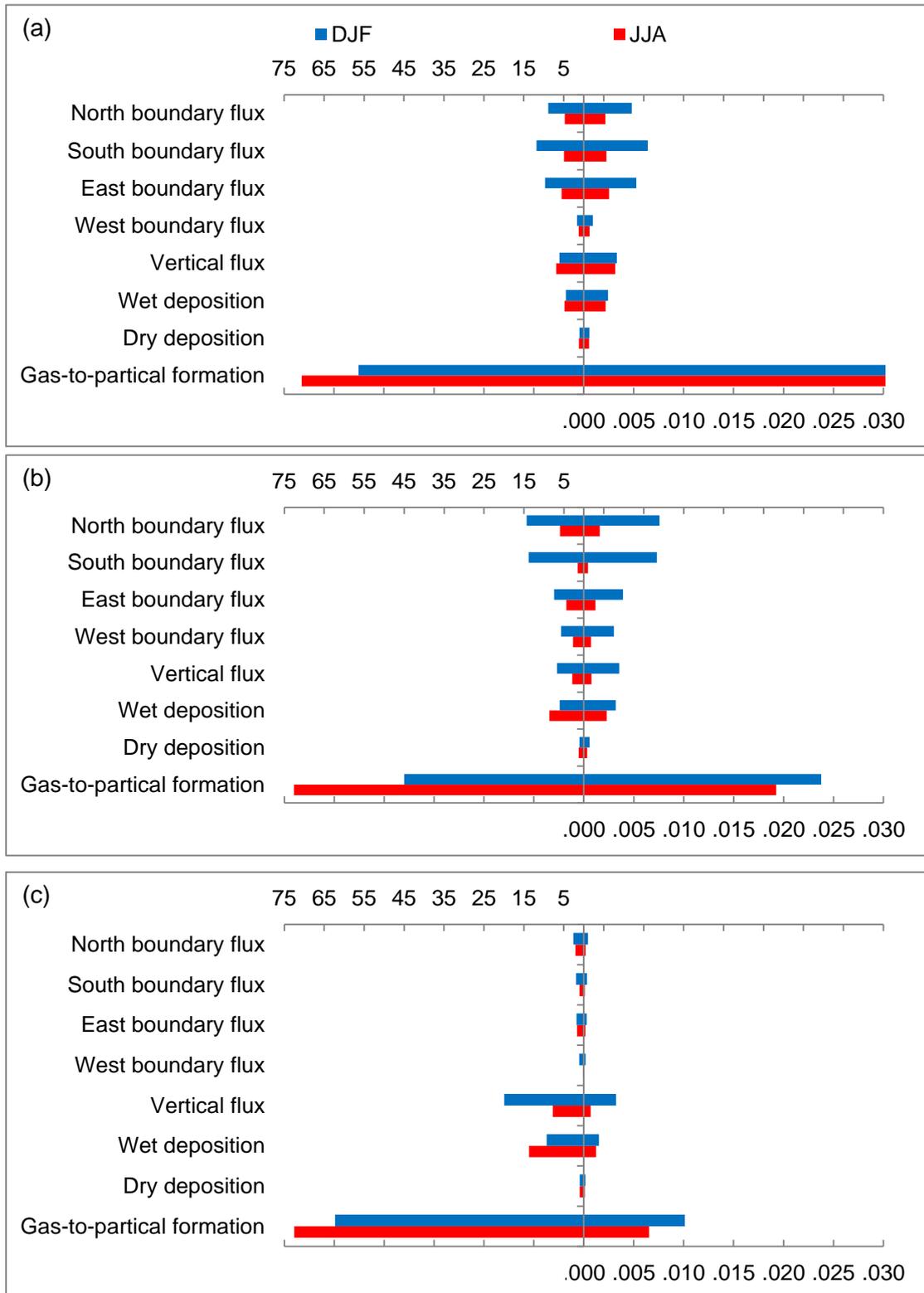


Fig. 12. The MAD (Tg N season^{-1} , right) and relative contribution (% , left) of each nitrate process (flux) in (a) NC, (b) SC, and (c) SCB obtained from simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport fluxes were calculated at the boundaries while other sources and sinks were calculated as the sums over the grid cells of the region under 1 km.

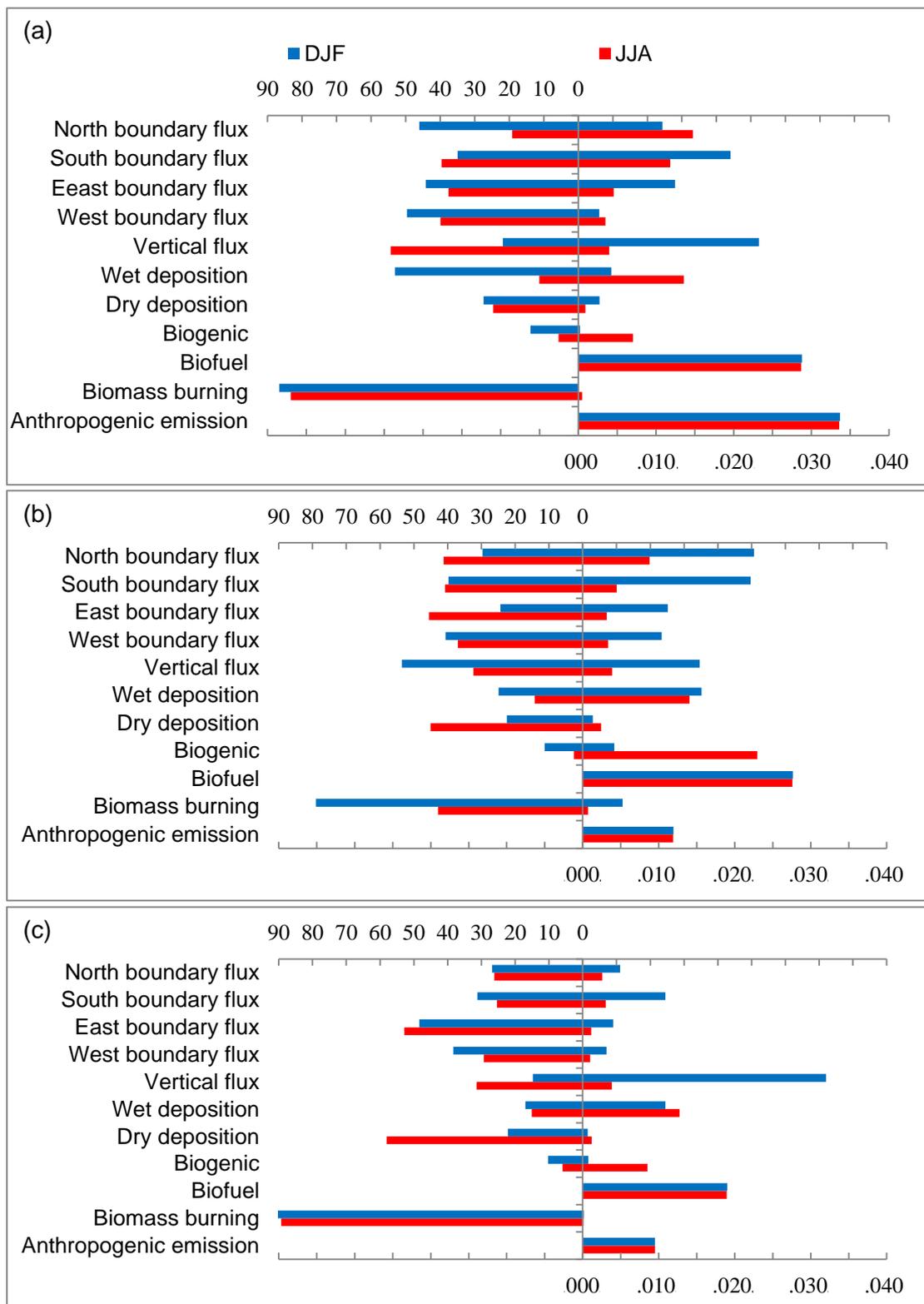


Fig. 13. Organic carbon budget (mass flux from each process in Tg C season⁻¹, right) and APDM of each flux (% , left) in (a) NC, (b) SC, and (c) SCB obtained from simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport fluxes were calculated at the boundaries while other sources and sinks were calculated as the sums over the grid cells of the region under 1 km.

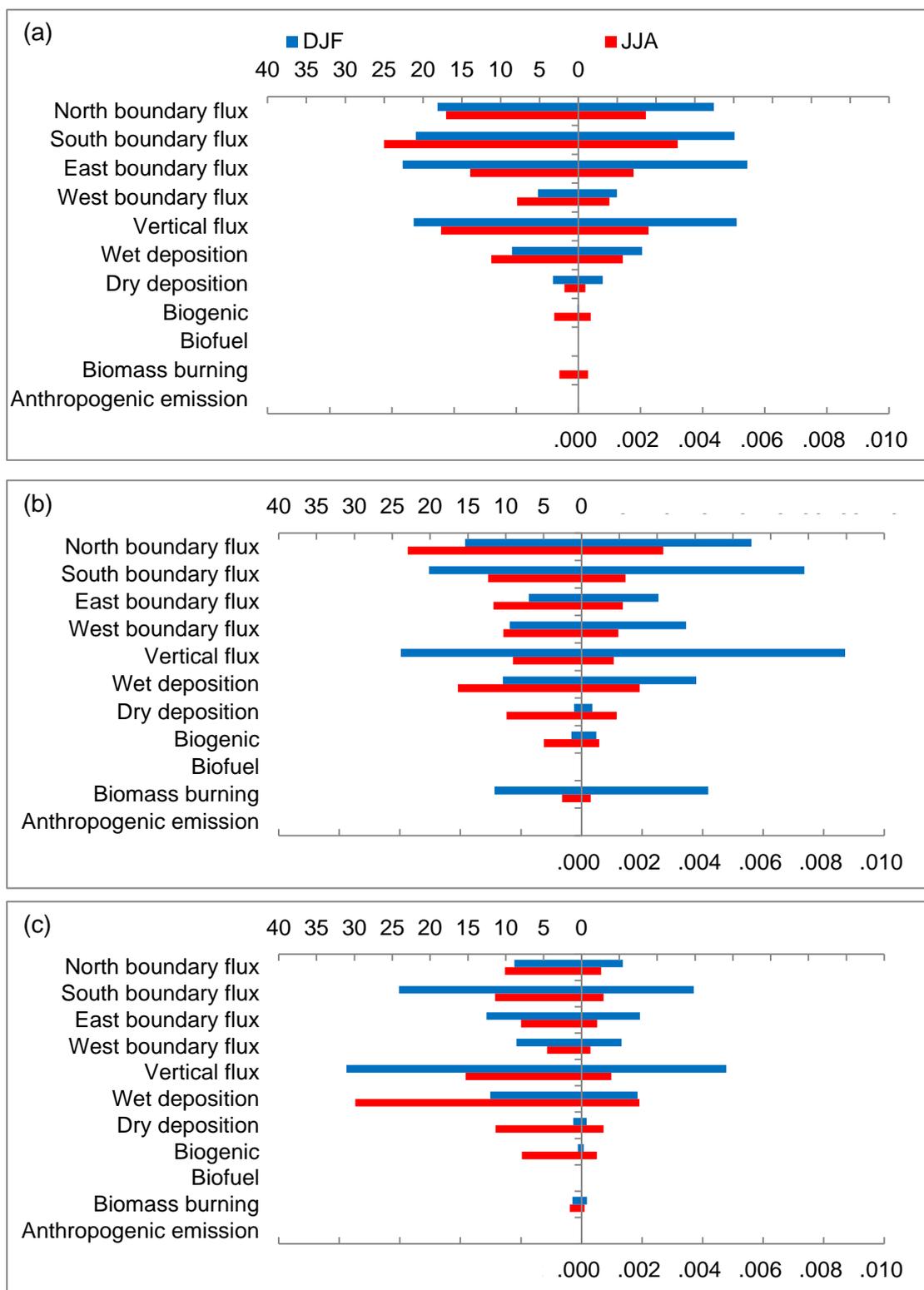


Fig. 14. The MAD (Tg C season⁻¹, right) and relative contribution (% , left) of each organic carbon process (flux) in (a) NC, (b) SC, and (c) SCB obtained from simulation ANNmet. Blue lines are for DJF and red lines are for JJA. Transport fluxes were calculated at the boundaries while other sources and sinks were calculated as the sums over the grid cells of the region under 1 km.