Response to Anonymous Referee #1

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We thank Referee #1 for their careful reading of the manuscript and comments; below we provide responses to each comment individually.

General Comment by Referee #1:

8 The paper presents the results of a detailed model analysis of nitryl chloride chemistry in polluted 9 continental outflow. This chemistry has been proposed as one of the main activation pathways for 10 chlorine in the troposphere and is certainly a subject that falls within the scope of Atmospheric Chemistry and Physics. The authors discuss in depth the impact of CINO2 chemistry on the oxidation 11 processes of VOC, on the formation of ozone and other pollutants (such as acyl peroxy nitrates); 12 interestingly, they also suggest that chlorinated VOC secondary products (such as acid chlorides) may 13 14 be important Cl sources in polluted regions. The results from this study provide many new insights into 15 our understanding of Cl chemistry and I recommend publication after the authors have addressed a few 16 minor questions. 17

18 Author responses follow each comment and are denoted with \*\*.19

## 21 GENERAL AND SPECIFIC COMMENTS

I have a few questions regarding the initialization of the model. Re the scaling of VOC measurements described in Sec 2: was the scaling applied only to ethanol and acetone? From line 11 it seems that all VOC were scaled. If so, it would probably be easier to just use the Atlantis dataset, I think.

\*\*A scaling factor was applied to all of the VOC which showed significant variation between the
Atlantis and ground site observations. Ethanol and acetone were simply given as examples of VOC
likely influenced by local emissions. Unfortunately, significantly fewer VOC were measured aboard the
Atlantis compared to the ground site – 13 vs. 44, respectively (see Supplemental Table S-1). We use the
ground site data set in order to constrain the model to the largest number of VOC possible.

33 Has the aircraft dataset been used at all? It is mentioned only in the introduction of the paper.

\*\*The aircraft data was not used in this study. Using the aircraft measurements to constrain the model in a diurnal sense would be difficult considering that 24-hour fixed type measurements are not feasible given aircraft measurement constraints. We chose to mention the aircraft measurements in order to direct readers to other relevant CINO2 studies considering that ambient CINO2 measurements are still relatively rare.

How were the data from the Atlantis selected? The text only says that the data in the LA region were
used, which is bit vague. Were the data filtered for distance from the coast and/or from the ground site?

\*\*We concentrate on the same time period described in Riedel et al. (2012) which we reference in the
manuscript. As discussed there Los Angeles outflow region was sampled during the period of May 16 –
31, 2010. We have added this date range to the main text.

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48 How accurate is the assumption of a constant 25C temperature?49

50 \*\*The mean temperature for the ground site was 17 °C with minimums near 13 °C and maximums near 51 24 °C during the day. While 25 °C is likely an overestimate during the night, it is important to note that

we are not trying to explicitly and completely represent the chemistry of the Los Angeles region with 1 2 the model, only to probe the effects of CINO2 formation in regions that might be similar to Los Angeles 3 during the CalNex study - polluted, coastal regions during late spring and early summer. For this purpose we feel that 25 °C represents an appropriate choice. That said we did perform a model run for a 4 5 10 °C case. The lower temperature enhances N2O5 formation given the temperature dependent 6 equilibrium between N2O5, NO3 and NO2. As a result, the maximum in ClNO2 and Cl-atoms increases by ~30% but because the modeled CINO2 in this case exceeds that observed, we would need to lower 7 8 the CINO2 yield and thus the actual impact on subsequent daytime chemistry would be negligible. We 9 have added the following statement to the main text to make this clear. "At lower model temperatures a 10 larger fraction of NOx will react as N2O5 with higher CINO2 levels and an increased morning Clburden relative to warmer cases. This result suggests that we are possibly overestimating the actual yield 11 12 of CINO2 per NO2 oxidized by ozone at night." 13 14 It would be useful to the community if the authors could make publicly available the expanded Cl+VOC 15 mechanism they have developed. Was the MCM protocol, as defined in the Jenkin/Saunders papers, 16 strictly followed (the protocol sets rules on how to exclude minor reaction channels and treat peroxy 17 radicals) or was the mechanism based on the expert judgment of the authors? 18 19 \*\*All Cl + VOC reactions incorporated into model mechanism and corresponding reaction rate 20 constants have been added to the supplemental information. The Matlab code containing the additional reactions and rate constants is also now freely available for download at: 21 22 ftp://ftp.atmos.washington.edu/thornton/UWCM/UWCM Riedel etal rxns.txt 23 We have added the following statement to the manuscript communicating this. "A complete list of the 24 added reactions and reaction rate constants is given in Supplemental Table S-2, and the MATLAB code 25 is freely available for download at ftp://ftp.atmos.washington.edu/thornton/UWCM/." For these reaction 26 mechanisms, we relied on our judgment and not the MCM protocol. In general we attempted to be 27 explicit and not simplify. For very fast reactions, such as those of alkoxy radicals we followed the 28 approach specified in Wolfe and Thornton (2011). The following statement clarifying this has been 29 added to the manuscript. "Similar to Wolfe and Thornton (2011), fast reactions, such as the reactions of alkoxy radicals, are treated as instantaneous in order to reduce model stiffness." 30 31 32 The authors tested the response of the model versus the reaction probabilities of ClNO3 and HOCl. 33 What about gamma(N2O5) and CINO2 yield? And the total aerosol surface area? 34 35 \*\*We chose to test the model sensitivity to the ClONO2 and HOCl reaction probabilities because those 36 quantities are more uncertain than gamma(N2O5) and the CINO2 yield at least for conditions 37 representative a polluted marine region like the Los Angeles outflow. Moreover, for gamma(N2O5), the 38 CINO2 yield, and total aerosol surface area, we have additional constraints provided by the CalNex field 39 measurements; N2O5 and CINO2 mixing ratios and aerosol surface area concentrations were measured 40 directly during the CalNex study. We focus on reproducing ClNO2 in excess of 1 ppbv. Therefore the N2O5-aerosol reaction probability must be large enough to allow for such CINO2 formation. Lowering 41 the N2O5-aerosol reaction probability by half, to 0.005, would require a CINO2 yield of nearly 100% in 42 43 order to produce >1 ppbv ClNO2. Lowering the N2O5-aerosol reaction probability to 0.001 or less 44 would result in insufficient CINO2 production. 45 For this reason gamma(N2O5) = 0.01 and a 50% ClNO2 yield represent a good choice. 46 47 It also seems to me that the estimated CINO2 photolysis rate as shown in Fig. S7 differs from the 48 observed rate in the period 6-11 am. How sensitive are the results to this parameter?

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- 50 \*\*The differences between the observed jClNO2 and that used by the model, which assumes clear sky 51 conditions and is generally larger than the observed jClNO2, would likely result in a slightly shorter

CINO2 lifetime compared to the observed conditions. While the small differences shown in
 Supplemental Figure S-7 would not significantly affect the presented results, if the observed jCINO2
 was used instead, we might expect a slightly slower decay in CINO2 throughout the morning and a
 smaller maximum in Cl atoms.

One of the largest uncertainties in this analysis seems to be HONO. Was it measured at any site during
CALNEX and how? How does modelled HONO compares with measured HONO? Heterogeneous
HONO formation in the model is mentioned in the summary but not really addressed earlier in the
discussion.

- \*\*HONO was measured at the CalNex Pasadena ground site via three different techniques. These are 11 12 described in detail in Young et al. (2012) which we reference frequently in the manuscript. We described the sources and sinks of HONO, in the model, with the statement: "Its abundance otherwise is 13 14 determined only by the reaction of OH + NO, HONO + OH, and the photolysis of HONO." As one 15 might expect, without a heterogeneous HONO source, the model under predicts HONO compared to the observations taken at the ground site. In order to investigate these effects, we constrained the modeled 16 17 HONO to the measured diurnal profile, and as we describe, we assume that by using these measurements as a constraint any heterogeneous HONO formation reactions are accounted for, and 18 19 perhaps overestimated because it is not clear that HONO sources in the MBL will be the same as over 20 land. We acknowledge the uncertainty in measured HONO especially regarding differences that might arise between parcels of air over land versus over water in the following statement. "However, as 21 22 discussed by Young et al. (2012), afternoon and daytime HONO concentrations are fairly uncertain, 23 especially when considering the extent to which HONO measurements over land represent HONO 24 concentrations in the marine boundary layer during morning hours within an air mass that was 25 transported offshore overnight."
  - TECHNICAL CORRECTIONS
- 30 page 28981, line 27: "surface area"
- 32 \*\*This change has been made.
- 34 page 28990, line 14: "dominates"
  - \*\*This change has been made.
- 38 page 28990, line 16: "sum of"
- 40 \*\*This change has been made.
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- Sources cited:
- 44 45
- 46 Riedel, T. P., Bertram, T. H., Crisp, T. A., Williams, E. J., Lerner, B. M., Vlasenko, A., Li, S.-M.,
- Gilman, J., de Gouw, J., Bon, D. M., Wagner, N. L., Brown, S. S., and Thornton, J. A.: Nitryl Chloride
   and Molecular Chlorine in the Coastal Marine Boundary Layer, Environmental Science & Technology,
   46, 10463, 10470, doi: 10.1001/oc204632r, 2012
- 49 46, 10463-10470, doi: 10.1021/es204632r, 2012.

- 1 Wolfe, G. M., and Thornton, J. A.: The Chemistry of Atmosphere-Forest Exchange (CAFE) Model -
- 2 Part 1: Model description and characterization, Atmos. Chem. Phys., 11, 77-101, doi: 10.5194/acp-11-3 77-2011, 2011.
- 4
- Young, C. J., Washenfelder, R. A., Roberts, J. M., Mielke, L. H., Osthoff, H. D., Tsai, C., Pikelnaya, O.,
- Stutz, J., Veres, P. R., Cochran, A. K., VandenBoer, T. C., Flynn, J., Grossberg, N., Haman, C. L.,
- 5 6 7 Lefer, B., Stark, H., Graus, M., de Gouw, J., Gilman, J. B., Kuster, W. C., and Brown, S. S.: Vertically
- 8 Resolved Measurements of Nighttime Radical Reservoirs; in Los Angeles and Their Contribution to the
- 9 Urban Radical Budget, Environmental Science & Technology, 46, 10965-10973, doi:
- 10 10.1021/es302206a, 2012.
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Response to Anonymous Referee #2

We thank Referee #2 for their careful reading of the manuscript and comments; below we provide responses to each comment individually.

6 General Comment by Referee #2:7

8 Riedel et al. report the addition of chlorine chemistry to the Master Chemical Mechanism and apply the 9 new mechanism to data collected during the Calnex-LA 2010 field campaign (focusing on the L.A. 10 urban outflow). The new mechanism allowed the identification and prediction of concentrations of several halogenated VOCs produced during the Cl initiated oxidation of alkenes and of the nature and 11 abundances of organic peroxy radicals produced. The authors confirm that the nocturnal conversion of 12 13 N2O5 to CINO2 and subsequent CINO2 photolysis increases O3 production on the following day in the study region. The authors also show that about 3/4 of the Cl produced by morning ClNO2 photolysis 14 15 converts to HCl, and that the remainder converts to ClO (via reaction of Cl with O3) or forms organochlorine molecules (via reaction of Cl with unsaturated hydrocarbons). Overall, the paper is 16 17 written well, is thorough, and presents important results that should be published after my minor concerns below have been addressed. 18 19

20 Author responses follow each comment and are denoted with \*\*.

23 General comments

1. The methodology used needs to be described in more detail. The additions are partly described on pg 28982 and in the supplement, but the paper is lacking a comprehensive table listing all of the reactions and rate coefficients that have been added to the model. Perhaps better still: Have the authors considered including the new MCM code as an appendix to this paper, or making it available for download on a web site or ftp server? After all, a considerable portion of this work is based on what was made freely available for download at the Leeds web site, and it would be a great service to the community if the authors were to follow the spirit of the MCM creators in this regard.

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\*\*We have added a table outlining all of the additional chlorine reactions and used rate constants not present in the native MCM to the supplemental information. The Matlab code containing the additional reactions and rate constants is also now freely available for download at:

ftp://ftp.atmos.washington.edu/thornton/UWCM\_Riedel\_etal\_rxns.txt. We have added the following statement to the manuscript communicating this. "A complete list of the added reactions and reaction rate constants is given in Supplemental Table S-2, and the MATLAB code is freely available for download at ftp://ftp.atmos.washington.edu/thornton/UWCM/."

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2. Calculated quantities were not compared with actual measurements. As such, the authors combined data from different measurement locations that are quite distant from each other. I agree that this was necessary to compensate for lack of certain measurements on the Atlantis. However, many of the estimated species used as model inputs and some of the species calculated (e.g., OH, HO2) were measured at the ground site. It would have made for a stronger and perhaps more interesting paper if the model presented here had been applied to and compared with measurements at the ground site only

rather than to a mixed ship/ground site data set. Perhaps something that could be considered for a futurepaper.

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50 \*\*It is true that the majority of modeling papers focus on comparing model outputs to field 51 measurements with an emphasis on reproducing the observations. However, in this study we did not

intend for the model to reproduce the CalNex observations explicitly but instead aimed to base the 1 2 model on a polluted coastal region similar to what was observed in the Los Angeles region during 3 CalNex in order to investigate the general effects of halogen chemistry in such a region. "The goal of these modeling studies is not to replicate the evolution of specific air masses in the LA Basin, but 4 5 instead to more generally probe the effect of multiphase reactive nitrogen and reactive halogen 6 chemistry on radical budgets, ozone production, and the fate of NOx in polluted coastal regions." We 7 expect that future studies will explicitly compare to the observations at the ground site, but this type of 8 box model is ill suited to that task given the need to faithfully represent hourly changes in transport and 9 boundary layer dynamics. 10

3. In the model, the inclusion of Cl production (from ClNO2 photolysis) has an effect that lasts
 throughout the entire simulated day. Is this because there is more total Cl available in the model when
 ClNO2 is included?

15 \*\*The referee is correct. The formation of CINO2 indeed provides an additional source of reactive chlorine by activating particulate chloride to Cl-atoms following CINO-2 photolysis. A portion of this 16 17 additional reactive Cl then proceeds through the more labile reservoirs HOCl and ClONO2 which allow it to have a more lasting impact over the course of the model day. We address this with the statement: 18 19 "To some extent these enhancements should be expected considering the larger Cl pool available for 20 recycling reactions when ClNO2 formation is allowed, but they give indication of the degree of indirect 21 coupling between CINO2 and Cl2 via the increased formation of reactive chlorine reservoirs like 22 ClONO2 and HOCl." 23

Specific comments

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pg 28976 lines 8-9. "... the fate of the Cl radicals and the overall impact of ClNO2 on regional photochemistry remain unclear" Unclear may be a bit too strong a word considering that we do have knowledge of some, if not most, of the chemistry, and we do have a good notion of what impact ClNO2 has on regional photochemistry in general. Consider rephrasing this sentence, for example by replacing the word "unclear" with "poorly constrained by measurements and models."

33 \*\*We have replaced "unclear" with "poorly constrained by measurements and models".

pg 28976, line 24 - pg 2897, line 10. Oum et al. [Science, 1998] reported the existence of a
 photochemical source of molecular chlorine from photolysis of ozone on sea salt aerosol. Please state
 whether the latter has been included in this paper, and if not, why not.

\*\*The model does not include a source of Cl2 from the photolysis of ozone in the presence of sea-salt particles as suggested by Oum et al. (1998). The Oum et al. (1998) study suggests a number of plausible mechanisms for Cl-atom production from O3 photolysis and subsequent heterogeneous reactions of H2O2 and/or OH with particle chloride to produce Cl2. Given the lack of a discrete mechanism to incorporate into the model we did not feel this potential source of Cl-atoms was complete enough to reliably include in the model.

- 46 pg 28979, lines 10/11 Please balance the chemical reactions (e.g., R9 and R10 are missing O2 as
   47 reactant).
- 49 \*\*We have added O2 over the reaction arrow.
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pg 28980, lines 18-20. Some of the data sets mentioned have been described in the literature - e.g., 1 2 Riedel et al., 2012a, Young et al., 2012, and Mielke et al., JGR, 2013. It would be appropriate to cite 3 those papers here. 4 5 \*\*We have added the suggested references. 6 7 pg 28981, lines 24-25. "Over the entirety of a model run temperature is held constant at 25 \_C" The 8 choice of temperature is critical as it affects reaction rates and model outcomes. A temperature of 25 \_C 9 seems too high for the nocturnal periods in this study region. Please include a sensitivity run at a lower 10 temperature (e.g., +10 \_C). 11 \*\*As suggested we performed a model run for a 10 °C case. The lower temperature enhances N2O5 12 13 formation given the temperature dependent equilibrium between N2O5, NO3 and NO2. As a result, the maximum in ClNO2 and Cl-atoms increases by ~30%. We have added the following statement to the 14 15 main text to make this clear. "At lower model temperatures a larger fraction of NOx will react as N2O5 with higher CINO2 levels and an increased morning Cl- burden relative to warmer cases. This result 16 17 suggests that we are possibly overestimating the actual yield of ClNO2 per NO2 oxidized by ozone at 18 night." 19 20 pg 28982, lines 26 and 27. The IUPAC database is continuously being updated. Please state the version 21 or year of the kinetics data used in this work. 22 23 \*\*We have added the following clarifying statement. "The reaction rate constants and product 24 branching for these reactions were taken from the IUPAC kinetics database as of May 11, 2012." 25 26 pg 28983, line 1 "CINO2 photolysis frequencies were estimated by scaling measured NO2 photolysis frequencies ... This approximation produces ... frequencies close to observations taken aboard the R/V 27 28 Atlantis" Please explain why the CINO2 photolysis frequencies were estimated even though they were 29 measured. Also, the CINO2 absorption cross-sections were recently revised by IUPAC (in June 2013). 30 Were the most recent values used in this work? 31 32 \*\*We chose to use the estimation as it allows the model to be more flexible for other investigations. Hardcoding the j-values measured by the R/V Atlantis into the model would have reduced the 33 functionality of the model to accurately represent latitudes and solar declination angles significantly 34 35 different from those of Los Angeles during May and June of 2010. These model results do not 36 incorporate the June 2013 IUPAC revisions to the CINO2 absorption cross-sections. These revisions are 37 not expected to significantly affect the results presented in our manuscript as the revisions agree well 38 with the previously used estimates. 39 40 pg 28983, line 5. ClONO2 and HOCl photolysis are mentioned here, but photolysis of Cl2 is not. Please 41 describe how its photolysis frequency was determined. 42 43 \*\*Cl2 photolysis frequencies were also determined using the TUV model. We have edited the 44 manuscript to reflect this. 45 46 pg 28984, line 5. "Gas-particle reaction probabilities in the model are set to 0.01 for N2O5 ... ... is within the typical range ... (< 0.001 - 0.03)" This range is quite large. Please consider sensitivity runs at 47 48 the extremes of this range. 49 50 \*\*We agree the listed range in N2O5-aerosol reaction probabilities is large. That said, the given range 51 represents N2O5 heterogeneous reaction probabilities over a large variety of different conditions 7

(particle sizes, particle compositions, relative humidities, temperatures, and locations). Given the high 1 2 levels of CINO2 (>1 ppbv) in these polluted coastal regions, the N2O5-aerosol reaction probability in 3 such regions must therefore be high enough to allow for such ClNO2 production. Please see our 4 responses to Referee #1 in this regard. 5 6 pg 28985 line 25. Please consider including a plot of the concentrations of CINO2, CIONO2, HOCl, 7 Cl2, OH, and CHOCl against hour of day from which the data in Fig 2 were derived. 8 9 \*\*The suggested plot has been added to the supplemental information and referenced in the main text. 10 pg 28987 line 16 "the reaction of OH with formyl chloride ... becomes a noticeable Cl source" The 11 authors speculate that this source may be higher in regions with alkene concentrations greater than Los 12 13 Angeles. I am not sure I would agree considering that alkenes would also react with NO3, slowing down 14 ClNO2 production. 15 16 \*\*We agree that enhanced concentrations of alkenes could potentially result in an enhancement in NO3 reactivity that might decrease CINO2 production. The effect will ultimately depend upon the abundance 17 of NO2 relative to the alkenes. At high NO2, the impact of higher alkenes on NO3 lifetime will be 18 19 somewhat buffered compared to the impact on Cl-atom reactivity. 20 21 pg 28988, paragraph starting on line 21 & Figure 4. It is difficult to follow this paragraph without 22 knowing the concentrations of ozone, NO, NO2, HO2, and the various VOCs that were present in the 23 model at 7 am and 3 pm. Consider calling out Figure S-9 earlier in the text and adding a table or graph 24 with key molecules (e.g., ozone, NO, NO2, HO2) to accompany Figures 4 and S-9. 25 26 \*\*As Figure S-9 is referenced in the same paragraph a few lines later we feel it is sufficient to direct the 27 reader to the relevant Cl-atom reactivity information. As requested, we have added a plot of NO, NO2, 28 and HO2 vs. model time to the supplemental information, and ozone mixing ratios are available in the 29 newly added Supplemental Figure S-12 (see next comment). 30 31 pg 28992, "3.3 Impact on ozone production rate" The model predicts \_10 ppbv of additional O3 as a 32 result of CINO2 chemistry (Figure 5C). To put this number in perspective, it would be useful to know how much total O3 the model produces in the absence of CINO2 and in its presence, rather than only 33 34 presenting the difference. Please consider adding this information to Figure 5, e.g., by modifying Figure 35 5C. 36 37 \*\*We have added a plot of total O3 mixing ratios predicted by the model for the with- and without-38 CINO2 cases to the supplemental information. 39 40 41 42 Sources cited: 43 44 Oum, K. W., Lakin, M. J., DeHaan, D. O., Brauers, T., and Finlayson-Pitts, B. J.: Formation of Molecular Chlorine from the Photolysis of Ozone and Aqueous Sea-Salt Particles, Science, 279, 74-76, 45 doi: 10.1126/science.279.5347.74, 1998. 46

# 1 An MCM modeling study of nitryl chloride (ClNO<sub>2</sub>) impacts on

## 2 oxidation, ozone production and nitrogen oxide partitioning in polluted

- 3 continental outflow
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- 5 Theran P. Riedel<sup>1,2</sup><sup>†</sup>, Glenn M. Wolfe<sup>3,4</sup>, Kenten T. Danas<sup>2</sup>, Jessica B. Gilman<sup>5,6</sup>, William C.
- 6 Kuster<sup>5,6</sup>, Daniel M. Bon<sup>5,6</sup>, Alexander Vlasenko<sup>7</sup>, Shao-Meng Li<sup>7</sup>, Eric J. Williams<sup>5,6</sup>, Brian M.
- 7 Lerner<sup>5,6</sup>, Patrick R. Veres<sup>5,6</sup>, James M. Roberts<sup>5</sup>, John S. Holloway<sup>5</sup>, Barry Lefer<sup>9</sup>, Steven S.
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#### 1 Abstract

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3 Nitryl chloride (ClNO<sub>2</sub>) is produced at night by reactions of dinitrogen pentoxide ( $N_2O_5$ ) on chloride 4 containing surfaces. ClNO<sub>2</sub> is photolyzed during the morning hours after sunrise to liberate highly 5 reactive chlorine atoms (Cl·). This chemistry takes place primarily in polluted environments where the 6 concentrations of N<sub>2</sub>O<sub>5</sub> precursors (nitrogen oxide radicals and ozone) are high, though it likely occurs 7 in remote regions at lower intensities. Recent field measurements have illustrated the potential 8 importance of CINO2 as a daytime Cl source and a nighttime NOx reservoir. However, the fate of the 9 Cl and the overall impact of ClNO<sub>2</sub> on regional photochemistry remain-unclear poorly constrained by 10 measurements and models. To this end, we have incorporated ClNO<sub>2</sub> production, photolysis, and 11 subsequent Cl reactions into an existing Master Chemical Mechanism (MCM version 3.2) box model 12 framework using observational constraints from the CalNex 2010 field study. Cl- reactions with a set of 13 alkenes and alcohols, and the simplified multiphase chemistry of N<sub>2</sub>O<sub>5</sub>, ClNO<sub>2</sub>, HOCl, ClONO<sub>2</sub>, and 14 Cl<sub>2</sub>, none of which are currently part of the MCM, have been added to the mechanism. The presence of 15 CINO<sub>2</sub> produces significant changes to oxidants, ozone, and nitrogen oxide partitioning, relative to model runs excluding ClNO<sub>2</sub> formation. From a nighttime maximum of 1.5 ppbv ClNO<sub>2</sub>, the daytime 16 17 maximum Cl concentration reaches 1x10<sup>5</sup> atoms cm<sup>-3</sup> at 7 AM, reacting mostly with a large suite of 18 volatile organic compounds (VOC) to produce 2.2 times more organic peroxy radicals in the morning 19 than in the absence of ClNO<sub>2</sub>. In the presence of several ppbv of nitrogen oxide radicals (NO<sub>x</sub> = NO + 20 NO<sub>2</sub>), these perturbations lead to similar enhancements in hydrogen oxide radicals (HO<sub>x</sub> = OH + HO<sub>2</sub>). 21 Neglecting contributions from HONO, the total integrated daytime radical source is 17% larger when 22 including CINO<sub>2</sub>, which leads to a similar enhancement in integrated ozone production of 15%. 23 Detectable levels (tens of pptv) of chlorine containing organic compounds are predicted to form as a 24 result of Cl addition to alkenes, which may be useful in identifying times of active Cl chemistry.

#### 26 1 Introduction

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Chlorine atoms (Cl·) are highly reactive, often having rate constants for reactions with volatile organic compounds (VOC) that are factors of 10 to 200 larger than the hydroxyl radical, OH, which is considered the atmosphere's primary initiator of oxidation. As a result, the presence of Cl· can lead to shorter lifetimes for VOC and an enhanced radical pool which can potentially enhance the production of ozone in polluted areas. The global tropospheric Cl· budget remains uncertain, with a large range in recent studies (~15 – 40 Tg Cl yr<sup>-1</sup>) developed from indirect means (Allan et al., 2007; Platt et al., 2004)

as tropospheric Cl- concentrations are not presently measurable by existing methods. There are a 1 2 number of potential Cl sources in the troposphere, the major sources are outlined in Reactions R1 - R5.

4	$HCl + OH \rightarrow Cl \bullet + H_2O$	(R1)
5	$Cl_2 + h\nu \rightarrow 2Cl \bullet$	(R2)
6	$BrCl + h\nu \rightarrow Cl \bullet + Br \bullet$	(R3)
7	$ClNO + h\nu \rightarrow Cl \bullet + NO$	(R4)
8	$ClNO_2 + h\nu \rightarrow Cl \bullet + NO_2$	(R5)

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10 The reaction of hydrochloric acid (HCl) with the hydroxyl radical (OH) is a daytime source of Cl. Typical HCl mixing ratios in the troposphere vary from 100 - 5000 pptv with the highest found in 11 12 polluted regions with direct HCl emissions from industrial processes and acid displacement of aqueous 13 chloride by HNO<sub>3</sub> and H<sub>2</sub>SO<sub>4</sub> (Keene et al., 2007). Cl<sup>.</sup> formed by HCl + OH tend to peak around midday with the peak in OH formed from  $O(^{1}D) + H_{2}O$ . Additionally, the oxidation of many VOC by 14 Cl· proceeds via a hydrogen abstraction to form HCl, thus recycling this Cl· source. 15

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17 Photolysis of molecular chlorine (Cl<sub>2</sub>) produces two Cl- and has been the focus of many Cl-18 investigations since it was first measured at elevated concentrations in ambient air (Finley and 19 Saltzman, 2006, 2008; Lawler et al., 2011; Riedel et al., 2012a; Spicer et al., 1998). Cl<sub>2</sub> mixing ratios were often on the order of tens of pptv with maximum reported mixing ratios near 100 - 200 pptv. 20 21 Direct Cl<sub>2</sub> emissions are related to power generation, water treatment, and oil refineries (Sarwar and 22 Bhave, 2007). Recently, a low pH Cl<sub>2</sub> production channel that may be atmospherically relevant has been identified in the reaction of N2O5 with chloride containing substrates, which involved CINO2 as an 23 24 intermediate (Roberts et al., 2008). In addition, Cl<sub>2</sub> can be formed in situ through multiphase chemistry 25 involving chlorine nitrate (ClONO<sub>2</sub>) and hypochlorous acid (HOCl). These species, in turn, can 26 photolyze to reform Cl· or ClO or react on acidic, chloride-containing particles to form Cl<sub>2</sub>. In polluted air, the reaction of ClO with NO, which completes a null cycle producing Cl- and NO<sub>2</sub>, limits the 27 28 potential for multiphase Cl<sub>2</sub> formation.

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30 BrCl photolysis to form Cl and atomic bromine is also thought to be an important Cl source, especially 31 in remote regions. In polar regions, BrCl mixing ratios on the order of tens of pptv have been measured 32 (Buys et al., 2013; Foster et al., 2001; Spicer et al., 2002). To our knowledge there have been no 33 reported observations of BrCl in ambient air outside of polar regions (Finley and Saltzman, 2008). BrCl can form through heterogeneous reactions of BrONO<sub>2</sub> and HOBr on acidic, chloride-containing
 particles in an analogous manner to the Cl<sub>2</sub> formation reactions described above or through reactions of
 ClONO<sub>2</sub> and HOCl on acidic, bromide-containing particles.

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5 Nitrosyl chloride (CINO) has also been proposed as a potential Cl· source (Raff et al., 2009). These 6 theoretical and laboratory studies have yet to be confirmed by field measurements of CINO in ambient 7 air. Using a regional 3-D chemical transport model, Raff et al. predict that CINO mixing ratios in 8 polluted marine areas could reach ppbv values. That said, the hydrolysis of CINO at moderate and high 9 relative humidity (RH>20%) will likely be sufficiently rapid to prevent the buildup of appreciable 10 atmospheric concentrations of CINO (Karlsson and Ljungström, 1996; Rubasinghege and Grassian, 2012; Scheer et al., 1997).

12

13 Since its proposed atmospheric formation by Finlayson-Pitts et al. (1989) and first observation in 14 ambient air by Osthoff et al. (2008), nitryl chloride (ClNO<sub>2</sub>) has been observed during a number of 15 different field studies worldwide with nighttime maximum mixing ratios ranging from tens of pptv to 16 over 2 ppbv (Kercher et al., 2009; Mielke et al., 2011; Mielke et al., 2013; Osthoff et al., 2008; Phillips et al., 2012; Riedel et al., 2012a; Thornton et al., 2010; Young et al., 2012). These observations have 17 18 occurred in both continental and marine locations illustrating the importance of ClNO<sub>2</sub> as a Cl· source in 19 a variety of different environments. CINO<sub>2</sub> represents a Cl source with clear anthropogenic origins as it 20 is formed at night by reactions involving  $NO_x$  ( $NO_2 + NO$ ), ozone, and chloride containing aerosols. 21 Anthropogenic activities associated with power generation, motor vehicle use, and agriculture now 22 dominate the global NO<sub>x</sub> source (Jaegle et al., 2005). Natural sources of NO<sub>x</sub>, such as microbial activity, 23 lightning, and wildfires, are also significant globally, but the impact of these  $NO_x$  sources on  $CINO_2$ 24 formation remain unknown. At night, a fraction of NOx is converted into CINO2 through Reactions R6-25 R8. The branching ratio between Reactions R8a and R8b, commonly referred to as the CINO<sub>2</sub> yield ( 26  $\phi_{CINO_2}$ ), is determined by the efficiency of CINO<sub>2</sub> formation from heterogeneous reactions of N<sub>2</sub>O<sub>5</sub>. The 27  $\phi_{ClNO_2}$  and the N<sub>2</sub>O<sub>5</sub>-particle reaction probability,  $\gamma$ (N<sub>2</sub>O<sub>5</sub>), are uncertain quantities that can vary 28 significantly depending on a number of factors such as particulate water, chloride, nitrate, and organic 29 content (Badger et al., 2006; Bertram and Thornton, 2009; McNeill et al., 2006; Mentel et al., 1999; 30 Thornton et al., 2003). After sunrise, the photolysis of ClNO<sub>2</sub> produces Cl<sup>.</sup> and NO<sub>2</sub>, thereby partially 31 circumventing the removal of  $NO_x$  through the formation and loss of 2HNO<sub>3</sub> (R8a).

(R6)

1 
$$NO_3 + NO_2 \leftrightarrow N_2O_5$$
 (R7)

$$2 \qquad N_2 O_5 + H_2 O \xrightarrow{\gamma} 2HNO_3 \tag{R8a}$$

3 
$$N_2O_5 + Cl^- \xrightarrow{\gamma} \phi_{CINO_2} ClNO_2 + (2 - \phi_{CINO_2})NO_3^-$$
 (R8b)

4

The Cl budget, hydrogen oxide and organic peroxy radical abundance ( $RO_x = OH + HO_2 + RO_2$ ),  $NO_x$ 5 lifetime and partitioning among other forms of reactive nitrogen, and the net ozone production rate are 6 7 linked through photochemical oxidation of VOC. As shown in Reactions R9 - R12 the oxidation of a 8 hydrocarbon (RH) is initiated by reaction with OH or Cl<sup>-</sup> to form the organic peroxy radical (RO<sub>2</sub>). In 9 polluted regions, the dominant fate of  $RO_2$  is to react with NO. The dominant channel of this reaction 10 eventually leads to a closed shell oxygenated hydrocarbon (OVOC - oxygenated volatile organic compound), hydroperoxyl radical (HO<sub>2</sub>) and NO<sub>2</sub>, while the minor channel leads to an alkyl nitrate 11 12 (RONO<sub>2</sub>). If the RO<sub>2</sub> is an acyl peroxy radical, then reaction with NO<sub>2</sub> produces acyl peroxy nitrates 13 (APN) such as acetyl peroxy nitrate (PAN). NO also reacts with HO<sub>2</sub> to form NO<sub>2</sub> and OH. Through 14 these reactions ozone is produced from the photolysis of NO<sub>2</sub>.

15

21

16	$RH + OH \xrightarrow{O_2} RO_2 + H_2O$	(R9)
17	$RH + Cl \bullet \xrightarrow{O_2} RO_2 + HCl$	(R10)
18	$RO_2 + NO \rightarrow OVOC + HO_2 + NO_2$	(R11a)
19	$RO_2 + NO \rightarrow RONO_2$	(R11b)
20	$RO_2 + NO_2 \leftrightarrow APN$	(R12)

Comment [TPR2]: Added O2 above reaction arrow.

22 From the above discussion, we expect that ClNO2 acts similarly, though not exactly the same, as an OH 23 source such as that from nitrous acid (HONO) photolysis or  $O(^{1}D) + H_{2}O$ . When the radical pool is 24 terminated via cross reactions between RO<sub>x</sub> and NO<sub>x</sub>, a higher production rate of HO<sub>x</sub> or Cl· will nearly 25 linearly increase the ozone production rate (Daum et al., 2000; Kleinman, 2005). Moreover, at high NO, 26 production of an RO<sub>2</sub> by Cl<sup>-</sup> attack directly increases the steady state concentration of OH and HO<sub>2</sub> due 27 to the rapid cycling between OH, HO<sub>2</sub> and RO<sub>2</sub>. However, the increased RO<sub>2</sub> due to Cl<sup>-</sup> arises from a 28 potentially different pool of hydrocarbons than that from OH, given the large differences in RH 29 abundance and relative reactivity towards OH and Cl. Moreover, ClNO<sub>2</sub> photolysis predominantly 30 occurs in the first few hours after sunrise, well before the maximum OH production rate from  $O(^{1}D)$  + 31 H<sub>2</sub>O and before the maximum in NO/NO<sub>2</sub>. Thus, the full impact of ClNO<sub>2</sub> on ozone production, VOC lifetime and NO<sub>x</sub> abundance and partitioning may not be the same as simply scaling the daytime average
 HO<sub>x</sub> production rate.

Here we examine the effects of  $CINO_2$  formation as predicted by a detailed box model that incorporates the Master Chemical Mechanism and is constrained by ground and ship-based ambient measurements taken during the CalNex 2010 field study. The aim is not to replicate specific observations or conditions, but rather to use the model to develop conceptual insights into the effect of a morning pulse of chlorine atoms in polluted air. We use the model to assess the impact of  $CINO_2$  on the Cl· budget,  $RO_x$  abundance,  $NO_x$  lifetime and partitioning, and the net ozone production rate.

10

3

#### 11 2 Measurements and model description

12

13 We chose to constrain a box model using data taken during the CalNex field study, which occurred in 14 May and June of 2010 in the southern California region (Ryerson et al., 2013). The goal of these 15 modeling studies is not to replicate the evolution of specific air masses in the LA Basin, but instead to 16 more generally probe the effect of multiphase reactive nitrogen and reactive halogen chemistry on 17 radical budgets, ozone production, and the fate of  $NO_x$  in polluted coastal regions. There were multiple 18 measurement platforms involved in CalNex, three of which recorded both ClNO2 and extensive VOC 19 measurements: the Research Vessel Atlantis, a ground site located on the California Institute of 20 Technology campus in Pasadena, CA, and aircraft measurements taken on the NOAA WP-3D (Mielke 21 et al., 2013; Riedel et al., 2012a; Young et al., 2012). Though the R/V Atlantis sampled in many 22 locations along the southern California coast, we focus on the measurements made in and around Los 23 Angeles urban outflow due to the added constraints provided by the Pasadena ground site 24 measurements. CINO<sub>2</sub> mixing ratios in the nocturnal outflow from the Los Angeles region were 25 commonly over 500 pptv with maximums on the order of 2 ppbv (Riedel et al., 2012a; Wagner et al., 26 2012).

27

Cl· chemistry was incorporated into an existing model framework described in Wolfe and Thornton (2011) which is based on the Master Chemical Mechanism version 3.2 (MCM) developed at the University of Leeds (more information available at http://mcm.leeds.ac.uk/MCM) (Bloss et al., 2005; Jenkin et al., 1997; Jenkin et al., 2003; Saunders et al., 2003). Use of the MCM allows for explicit tracking of approximately 2800 chemical species and about 9000 different reactions with reaction rate constants derived from the International Union of Pure and Applied Chemistry (IUPAC) kinetics database (http://www.iupac-kinetic.ch.cam.ac.uk). Similar to Wolfe and Thornton (2011), fast reactions.

Comment [TPR3]: Added references.

#### such as reactions of alkoxy radicals, are treated as instantaneous in order to reduce model stiffness.

1 2

15

3 In total, 44 of the VOC measured at the Pasadena site are used to constrain the model. However, certain 4 VOC, for examplesuch as ethanol (median value = 8.2 ppbv) and acetone (median value = 3.8 ppbv), 5 measured at the Pasadena ground site appeared to often be dominated by highly localized emissions. To 6 more generally represent an urban air mass in the model, ground site VOC measurements were scaled 7 by those measured on the R/V Atlantis. A smaller number of VOC were measured aboard the R/V 8 Atlantis, so species not represented in the R/V Atlantis dataset were scaled by species of similar 9 structure (i.e., similar functional groups). For example, methanol was measured with median levels of 10 approximately 6 ppbv and 1 ppbv at the ground site and on the ship, respectively. Ethanol, however, 11 was only measured at the ground site. To estimate ethanol levels in the urban outflow and be more 12 representative of what the R/V Atlantis might have sampled, the ground site ethanol mixing ratios were simply scaled down by 1/6. For a complete list of the measured VOC used in the model see 13 14 Supplemental Information Table  $\mp$ S-1.

VOC and HCl mixing ratios are held to their ship-scaled hourly average diurnal profiles for a 69 hour 16 17 "spin-up" period. The diurnal HCl profile used is shown in Supplemental Figure S-1. NO<sub>2</sub>, O<sub>3</sub>, and CO 18 are held to mean values measured at the ground site during this spin-up period. In addition, we fix 19 methane at a mixing ratio of 1.8 ppmv. Over the entirety of a model run temperature is held constant at 25 °C and the aerosol surface area concentration is held constant at 350 µm<sup>2</sup> cm<sup>-3</sup>, which represents 20 some of the largest aerosol surface area concentrations encountered by the R/V Atlantis while sampling 21 22 Los Angeles outflow. At lower model temperatures a larger fraction of  $NO_x$  will react as  $N_2O_5$  with 23 higher ClNO<sub>2</sub> levels and an increased morning Cl- burden relative to warmer cases. This result suggests 24 that we are possibly overestimating the actual yield of CINO<sub>2</sub> per NO<sub>2</sub> oxidized by ozone at night. The 25 box model does not attempt to replicate the effects of meteorology and thus the processes of dilution 26 and deposition are not accurately incorporated. To maintain reasonable concentrations of the many modeled species which were not constrained by observations, we apply a continuous dilution rate of 27 28 1.5% per hour to all species. Formaldehyde and nitric acid have an additional deposition rate of 30% per 29 hour in order to keep mixing ratios at levels most similar to those sampled during the CalNex study (<6 30 ppbv for formaldehyde) (Warneke et al., 2011). The sensitivity of the results to this additional loss rate is minor (<20% adjustments to the Cl budget), and the need for this additional loss is likely related to 31 our neglect of deposition for intermediate organic oxidation products (in the case of formaldehyde) and 32 33 of HNO<sub>3</sub> itself. Isoprene, alpha-pinene, beta-pinene, and limonene are allowed to freely evolve at night 34 during the spin-up period to avoid unrealistic conditions whereby NO<sub>3</sub> reactions with these compounds

Formatted: Subscript Formatted: Subscript Formatted: Subscript Formatted: Subscript Formatted: Subscript Formatted: Subscript proceeded indefinitely throughout the night. That is, we assume that advection of air masses from the land out over the water (either at the surface or aloft of the continental nocturnal surface layer) would ultimately limit the source of reactive biogenic VOC. At hour 69, which represents sunset on the third model day, all species are released from observational constraints and the chemistry evolves freely for another 27 hours. We use the final 24-hours of a model run as the analysis period. All figures and calculations described here are performed on the model output from this period.

7

8 A number of reactions necessary for evaluating Cl production and reactivity are not included in the 9 MCM. Version 3.2 of the MCM only includes Cl- reactions with alkanes. In order to accurately 10 represent the chemistry, multiple mechanisms were added to the model framework. These include the Reactions R1 – R3, R8, R13 – R21, and a number of VOC + Cl reactions such as those for methanol, 11 12 ethanol, isopropanol, ethene, propene, formaldehyde, ethanal, propanal, acetone, benzene, styrene, o-13 xylene, toluene. Several of the added mechanisms are explicitly shown in the Supplemental Information 14 (Supplemental Figure S-2 – S-6). A complete list of the added reactions and reaction rate constants is 15 given in Supplemental Table S-2, and the MATLAB code is freely available for download at ftp://ftp.atmos.washington.edu/thornton/UWCM/. The reaction rate constants and product branching for 16 17 these reactions were taken from the IUPAC kinetics database as of May 11, 2012. Chlorinated products 18 not present in the MCM or available in the IUPAC database were assumed to react similarly and with 19 similar rate constants to non-chlorinated species of the same structure already in the MCM (see 20 Supplemental Table S-2). Additionally, our revised mechanism explicitly tracks gas-phase HCl 21 formation that results from hydrogen atom abstraction reactions by Cl. ClNO<sub>2</sub> photolysis frequencies 22 were estimated by scaling measured NO<sub>2</sub> photolysis frequencies down by a factor of 30 (i.e.,  $j_{NO_2}$ /30). 23 This approximation produces ClNO<sub>2</sub> photolysis frequencies close to observations taken aboard the R/V 24 Atlantis (Supplemental Figure S-7). Photolysis frequencies for Cl<sub>2</sub>, ClONO<sub>2</sub>, and HOCl were calculated 25 using the Tropospheric Ultraviolet and Visible (TUV) Radiation Model (available at 26 http://cprm.acd.ucar.edu/Models/TUV) and incorporated into the box model.

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28	$Cl \bullet + O_3 \rightarrow ClO + O_2$	(R13)
29	$ClO + NO_2 \rightarrow ClONO_2$	(R14)
30	$ClO + HO_2 \rightarrow HOCl + O_2$	(R15)
31	$ClONO_2 + h\nu \rightarrow Cl \cdot + NO_3$	(R16)

1	$HOCl + hv \rightarrow Cl \bullet + OH$	(R18)
2	$ClONO_2 + H^+ + Cl^- \xrightarrow{\gamma} Cl_2 + HNO_3$	(R19)
3	$HOCl + H^+ + Cl^- \xrightarrow{\gamma} Cl_2 + H_2O$	(R20)
4	$ClO + NO \rightarrow Cl \bullet + NO_2$	(R21)

5

6 For alkenes, the major pathway involves addition of chlorine to the double bond rather than the typical 7 hydrogen abstraction pathway (Atkinson et al., 2004). This pathway leads to chlorinated products which 8 might be detectable as tracers of Cl- chemistry in future studies. As we show below, such compounds 9 could be another avenue for Cl recycling. Reactions of Cl with isoprene, which also produces unique 10 chlorinated products, were not included in the model framework given its modest <1% contribution to 11 total Cl reactivity in the modeled Los Angeles outflow and the large increase in complexity when 12 incorporating the mechanism (Fan and Zhang, 2004; Tanaka et al., 2003). Instead, the products of Cl. + 13 isoprene were tracked as a single generic species with no chemical losses. However, in areas where 14 isoprene is a more significant contributor to Cl reactivity, it would be necessary to include a more 15 explicit isoprene oxidation mechanism to accurately capture the effects Cl., especially to assess any 16 chlorinated products that might form from these reactions (Riemer et al., 2008). In such locations the 17 products of chlorine-initiated isoprene oxidation are likely more pronounced than in the Los Angeles 18 region.

19

20 Gas-particle reaction probabilities in the model are set to 0.01 for N<sub>2</sub>O<sub>5</sub>, ClONO<sub>2</sub>, and HOCl (R8, R19, 21 R20). A  $\gamma = 0.01$  is within the typical range of  $\gamma(N_2O_5)$  measured on ambient aerosol (<0.001 – 0.03) at 22 elevated RH during various field studies (Bertram et al., 2009; Riedel et al., 2012b). Laboratory 23 measurements of ClONO<sub>2</sub> and HOCl uptake under stratospheric and tropospheric conditions on sulfuric 24 acid, sodium chloride, and sodium bromide particles and pure water droplets generally report  $\gamma$  values 25 <0.06 for CIONO<sub>2</sub> and HOCI (Deiber et al., 2004; Hanson and Ravishankara, 1994; Hanson et al., 1994). We make the upper-limit assumption that reactions of  $CIONO_2$  and HOCI on aerosol particles 26 27 produce only Cl<sub>2</sub> with unit efficiency. Given that Cl<sub>2</sub> production from heterogeneous reactions of CIONO<sub>2</sub> and HOCl is proportional to the product of  $\gamma$  and the yield, we use  $\gamma = 0.01$  and a 100% yield 28 29 on all particles in the model.

30

To examine the effects of ClNO<sub>2</sub> formation, we vary  $\phi_{ClNO_2}$  between 0% and 50% in successive model runs, which produce a without-ClNO<sub>2</sub> case and a with-ClNO<sub>2</sub> case, respectively. A 50% yield results in ~1.5 ppbv of ClNO<sub>2</sub> as shown in Figure 1, which is similar to levels in the Los Angeles outflow 1 conditions encountered during CalNex. We also performed a series of model runs where HONO was 2 constrained to observations made at the Pasadena ground site. Its abundance otherwise is determined 3 only by the reaction of OH + NO, HONO + OH, and the photolysis of HONO. Most of our main 4 conclusions reported here are relatively insensitive to HONO. Moreover, the vertical profiles of ClNO<sub>2</sub> 5 and HONO throughout the nocturnal and evolving daytime boundary layer are likely different (Young et 6 al., 2012), making our primary focus on ClNO<sub>2</sub> a reasonable simplification for a box model.

7

## 8 3 Results and discussion

9

10 3.1 Cl-atom budget

11

12 The model predicts that, integrated over a typical day in the Los Angeles outflow, ClNO<sub>2</sub> is the major driver of Cl- evolution. Neither HCl + OH nor multiphase chemistry involving ClONO<sub>2</sub> and HOCl to 13 14 produce  $Cl_2$  are competitive with the  $CINO_2$  source. Moreover, this picture is consistent with that 15 derived solely from observations in this region (Riedel et al., 2012a). Figure 1 shows the Cl-16 concentration predicted by the model during the 24-hour analysis period for both the with-ClNO<sub>2</sub> and 17 without-CINO<sub>2</sub> cases. When CINO<sub>2</sub> formation is included, the Cl<sup>.</sup> concentration reaches a maximum at 18  $\sim$ 7 AM (2 hours after model sunrise) with a value of 1.08x10<sup>5</sup> atoms cm<sup>-3</sup>. A substantially different 19 picture results from the without-CINO2 case where the maximum Cl- concentration occurs around noon and only reaches  $0.2 \times 10^5$  atoms cm<sup>-3</sup>. The assumptions made about the aerosol reaction probabilities of 20 CIONO<sub>2</sub> and HOCl partially drive the late afternoon Cl· profile, which, as a result, is more uncertain. 21 22 However, this afternoon Cl- concentration profile is not especially sensitive to the assumed reaction 23 probabilities. For example, increasing the reaction probabilities of ClONO<sub>2</sub> and HOCl from 0.01 to 0.1 24 does not substantially change the 24-hour profile. The maximum in Cl- concentration is increased by 25 ~10% still occurring in the early morning hours after sunrise (~7 AM), and the integrated Cl-26 concentration over the entire day is enhanced by only 20%. Additionally, the choice of 0.01 for a 27 ClONO2 and HOCl reaction probability and a 100% Cl2 yield is likely more realistic as the formation of 28 Cl<sub>2</sub> from these reactions is unlikely to be the sole product (Caloz et al., 1996; Santschi and Rossi, 2005). That said, to fully understand the impact of Cl chemistry in coastal urban areas, the fate of ClONO<sub>2</sub> 29 30 especially needs to be better constrained.

31

The evolution of Cl· largely follows that of the dominant source terms, as shown in Figure 2. In the absence of ClNO<sub>2</sub> formation (Fig. 2A), the bulk of Cl· production results from the HCl + OH production channel, and the maximum in Cl· production rate of  $0.5 \times 10^6$  atoms cm<sup>-3</sup> s<sup>-1</sup> coincides with the maximum

production from the HCl + OH channel. In the with-ClNO<sub>2</sub> case (Fig. 2B), the maximum Cl· production 1 rate occurs at 7 AM with a value of  $3.4 \times 10^6$  atoms cm<sup>-3</sup> s<sup>-1</sup> corresponding to the maximum contribution 2 from ClNO<sub>2</sub> photolysis. The ClNO<sub>2</sub> production channel represents 56% of Cl<sup>.</sup> production over the 3 course of the entire day, leading to nearly 3.7 times as many Cl produced relative to the without-ClNO2 4 5 case. By noon, CINO2 is largely depleted under the clear-sky model conditions, and other Cl production 6 channels like the reaction of HCl with OH and the photolysis of Cl<sub>2</sub> from HOCl and ClONO<sub>2</sub> 7 heterogeneous chemistry become more dominant. These production channels involving multi-phase Cl-8 recycling to form Cl<sub>2</sub> show significant enhancements when ClNO<sub>2</sub> formation is included. For example, 9 Cl production from ClONO<sub>2</sub> photolysis, HOCl photolysis, and Cl<sub>2</sub> photolysis are enhanced by factors 10 of 3.3, 2.2, and 3.3, respectively over the without-ClNO<sub>2</sub> case. To some extent these enhancements 11 should be expected considering the larger Cl pool available for recycling reactions when ClNO<sub>2</sub> 12 formation is allowed, but they give indication of the degree of indirect coupling between ClNO2 and Cl2 via the increased formation of reactive chlorine reservoirs like ClONO2 and HOCI. During CalNex, 13 14 molecular chlorine was also measured along with CINO<sub>2</sub> (Riedel et al., 2012a). Observations of 15 nighttime and early morning  $Cl_2$  were typically in the 5 – 50 pptv range. Modeled  $Cl_2$  levels are of similar magnitude to these observations, as well as previous observations of Cl<sub>2</sub> in this region (Finley 16 17 and Saltzman, 2006, 2008), and show a morning enhancement with slightly elevated levels throughout 18 the day but only with the inclusion of CINO<sub>2</sub> (see Supplemental Figure S-8). For additional information 19 on the model concentrations of the Cl<sup>.</sup> source species see Supplemental Figure S-9.

20

21 Given the lack of BrCl observations outside of polar regions, we do not include BrCl formation in the 22 model and therefore do not explicitly account for the potential Cl source, if any, represented by BrCl. 23 Considering Cl<sub>2</sub> represents 16% of the integrated Cl<sup>-</sup> source over the course of a model day, the typical 24 seawater ratio of chloride to bromide of ~650:1, and assuming that BrCl formation is not significantly 25 faster than 650 times Cl<sub>2</sub> formation, we estimate an upper limit Cl source from BrCl resulting from 26 ClONO<sub>2</sub> and HOCl reactions that is similar to that predicted from Cl<sub>2</sub>. That said, BrCl formed from 27 these reactions should not significantly bias our Cl estimates considering that we force these reactants 28 to produce exclusively Cl<sub>2</sub>, the photolysis of which forms 2Cl compared to only 1Cl from BrCl 29 photolysis. However, heterogeneous reactions of BrONO<sub>2</sub> and HOBr to form BrCl are not accounted for 30 at all. Using maximum HCl levels as a measure of particulate chloride displaced over Los Angeles and 31 the expected chloride to bromide ratio in seawater, 2.5 ppbv of chloride corresponds to  $\sim$ 4 pptv of 32 bromide available for BrCl formation. Incorporating this amount of total bromine into the model and 33 assuming  $\gamma(BrONO_2)$  and  $\gamma(HOBr) = 0.1$  with a unit yield of BrCl, we predict the model could be 34 neglecting a Cl- source from BrCl on the order of 5% of the Cl- concentration integrated over the model

day. Ambient measurements of BrCl in polluted coastal regions would be a particularly useful constraint
 on the extent of these Cl recycling reactions and the role of bromide.

3

The use of a comprehensive chemical mechanism such as the MCM also illustrates a potentially 4 5 important but heretofore overlooked source of Cl- in polluted regions. In the with-ClNO<sub>2</sub> case, the 6 reaction of OH with formyl chloride (CHOCl), produced from Cl- attack of alkenes, becomes a 7 noticeable Cl. source during the afternoon. Interestingly, CHOCl photolysis is predicted to be a Cl-8 source comparable in magnitude to that from HOCl photolysis (Figure 2B). In fact, because we possibly 9 overestimate the actual multi-phase recycling of ClONO2 and HOCl to form Cl2, Cl- release from such 10 acid chlorides may be more important than these multiphase processes in regions with significant alkene concentrations. This result suggests observations of acid chlorides would be as beneficial as Cl2 in 11 12 polluted regions.

13

14 HONO has a noticeable impact on the afternoon Cl budget via photolysis to form OH followed by the 15 reaction of OH + HCl. Constraining the model to the HONO diurnal profile measured at the Pasadena ground site leads to a 60% increase  $(1.4 \times 10^7 \text{ to } 2.26 \times 10^7 \text{ molecules cm}^{-3})$  in the daily maximum OH 16 17 concentration and a similar increase in the integrated Cl- formation rate from OH + HCl. Multiphase 18 recycling via ClONO<sub>2</sub> and HOCl are also increased as a result of the larger Cl<sup>.</sup> concentrations. 19 However, as discussed by Young et al. (2012), afternoon and daytime HONO concentrations are fairly 20 uncertain, especially when considering the extent to which HONO measurements over land represent 21 HONO concentrations in the marine boundary layer during morning hours within an air mass that was 22 transported offshore overnight.

23

24 We also investigate the formation potential of chlorinated products at potentially detectable quantities 25 resulting from the Cl oxidation of VOC. These species could represent "tracers" of Cl chemistry and 26 indicate the extent to which Cl- oxidation reactions are occurring. This issue has been investigated 27 previously in the Houston area using ground based measurements of potential products from Cl-28 oxidation of isoprene (Riemer et al., 2008). As stated above, chlorinated products of Cl reactions with 29 isoprene are not included our model and are likely of minor importance to total Cl- reactivity in the Los 30 Angeles region. Within our model framework, Cl- addition reactions with ethene and propene can nitrate 31 produce chlorinated products such as 2-chloroperoxypropionyl (2-chloroPPN), 32 chloroacetaldehyde, 2-chloroperoxyacetyl nitrate (2-chloroPAN), formyl chloride, 2-chloropropanal, 33 and chloroacetone - some of which have been previously investigated in polar regions (Keil and 34 Shepson, 2006). As with most of the previously mentioned effects, these enhancements are pronounced in the early part of the day when Cl· and VOC concentrations are highest in the model. As we show in
 Figure 3, morning enhancements are typically on the order of 5-30 pptv above the background generated
 during the spin-up period. Chloroacetaldehyde and formyl chloride exhibit the largest enhancements.

4

5 In a polluted region such as the Los Angeles basin and outflow, Cl- will be primarily lost through 6 reactions with VOC. Cl reactivity as determined by the model is illustrated in Figure 4, where it is 7 broken into Cl- reactions with methane, alkanes, alkenes, alcohols, carbonyls (aldehydes and ketones), 8 and ozone. Other Cl + VOC reactions, such as reactions with aromatics, are not shown as they were not 9 a significant sink of Cl relative to the species listed. In the morning (7 AM), the bulk of reactivity is due to reactions with a variety of alkanes, which comprise 42% of the 44 sec<sup>-1</sup> total reactivity. Later in the 10 day (3 PM), as the VOC are consumed in the model and  $O_3$  maximizes, the reaction with  $O_3$  to form 11 12 CIO is the dominant Cl· sink and represents about 60% of the 39 sec<sup>-1</sup> total reactivity. However, the majority of ClO (77% at 7 AM and 67% at 3 PM) will react with NO to give NO2 and reform Cl· that 13 14 can terminally react with VOC. Thus, we predict only 23 - 33% of the Cl·+O<sub>3</sub> reactions produce stable 15 reservoirs like ClONO<sub>2</sub> and HOCl; though other studies predict even less for the Los Angeles region (Young et al., 2013). Over the course of the day, no single VOC dominates the Cl- reactivity (see 16 17 Supplemental Figure S-119). Methane is consistently about 10% of the total. This implies that all VOC 18 reactive towards Cl are relevant when trying to estimate the total reactivity and that only using a small 19 subset may significantly underestimate the total. This behavior is different from that of OH, the 20 reactivity of which is often dominated by a few VOC or inorganic species such as CO and NO<sub>2</sub> in 21 highly polluted regions (Kato et al., 2011; Ren et al., 2003).

22

23 3.2 Impact on  $RO_x$  and  $NO_x$ 

24

25 The ClNO<sub>2</sub> formation described above leads to important and unique impacts on processes relevant to 26 tropospheric air quality. Cl· produced by ClNO<sub>2</sub> photolysis will react with VOC to produce RO<sub>2</sub> during 27 morning hours. The RO<sub>2</sub> will primarily react with NO under the polluted conditions to form HO<sub>2</sub> and 28 closed-shell oxygenated VOC or an alkyl nitrate. Alternatively, given the large NO<sub>2</sub>/NO ratio during the 29 morning, enhanced peroxy nitrate formation is possible via reaction of acyl peroxy radicals with NO<sub>2</sub>. 30 The HO<sub>2</sub> produced via  $RO_2$  + NO also reacts with NO to form OH, which in turn reacts with VOC to 31 form RO<sub>2</sub>. The chain length of this cycle tends to be somewhat short in the morning given higher 32 concentrations of  $NO_x$  leading to termination via  $OH + NO_2$  to form  $HNO_3$  which is efficiently removed 33 from the atmosphere through wet and dry deposition. Nonetheless, Cl- will enhance the abundance of 1 morning time  $RO_x$  (OH + HO<sub>2</sub> + RO<sub>2</sub>) as illustrated in Figure 5 which will lead to enhanced O<sub>3</sub> 2 production rates relative to a model run without ClNO<sub>2</sub>.

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Figure 5A shows the HO<sub>x</sub> production rate ( $P_{HOx}$ ) for both the with- and without-ClNO<sub>2</sub> cases. A factor 4 5 of 2.2 increase occurs in the early morning hours around 7 AM when Cl· production from ClNO2 6 photolysis is the major Cl source. P<sub>HOx</sub> remains elevated throughout the day relative to the without-7 CINO<sub>2</sub> case likely due to the larger ozone values in the with-CINO<sub>2</sub> case, thus illustrating that the 8  $CINO_2$  influence persists for more than just the early morning hours. 24-hour integrated HO<sub>x</sub> production 9 for the with- and without-ClNO<sub>2</sub> cases is 75 ppbv and 62 ppbv, respectively. Uncertainties in modeling 10 HONO have the largest impact on quantifying the perturbation of ClNO2 to P<sub>HOx</sub>. Constraining modeled 11 HONO to the diel average values measured at the ground site results in the same overall pattern of 12 CINO<sub>2</sub> effects on P<sub>HOx</sub> described above, just at a reduced relative magnitude - i.e., the inclusion of 13  $CINO_2$  formation increases  $P_{HOx}$  by ~35% in the morning, with moderate enhancements to  $P_{HOx}$ 14 sustained throughout the day resulting in an integrated P<sub>HOx</sub> of 116 ppbv and 105 ppbv for the with- and 15 with-out ClNO<sub>2</sub> cases, respectively. These two extremes in terms of HONO are likely both 16 representative - the latter HONO-rich case being representative of near surface chemistry while the 17 former HONO-poor case is more representative of the residual boundary layer aloft of the surface, 18 which dominates the column-average radical budget as described by Young et al. (2012).

20 Figure 5B shows the effects of ClNO<sub>2</sub> on OH, HO<sub>2</sub>, the sum over-of all 717 organic peroxy radicals 21 (RO<sub>2</sub>), and the sum of all 140 acyl peroxy nitrates (APN) predicted by the model. The ratio of the with-22 CINO<sub>2</sub> case relative to the without-CINO<sub>2</sub> case is shown. Inclusion of CINO<sub>2</sub> formation results in 23 significant changes in  $HO_x$  (OH and  $HO_2$ ) with 190% and 220% enhancements during the morning 24 hours in OH and both HO<sub>2</sub> and RO<sub>2</sub>, respectively. Such enhancements could be partly related to noted 25 discrepancies between measured and modeled morning HO<sub>x</sub> levels (Dusanter et al., 2009; Mao et al., 26 2010; Ren et al., 2003). Comparable enhancements in HO<sub>2</sub> and RO<sub>2</sub> were predicted by Osthoff et al. 27 (2008) using the MCM to assess measurements taken in and around the Houston ship channel. In remote 28 low-NO<sub>x</sub> regions, Cl and OH are largely uncoupled such that the presence of one does not largely 29 impact the abundance of the other. This condition then allows indirect quantification of Cl abundance 30 by comparing VOC which have different reaction rate constants for reaction with OH and Cl. (i.e., 31 Jobson et al. (1994), Platt et al. (2004), and Allan et al. (2007)). However, the presence of additional Cl-32 from ClNO<sub>2</sub> in a polluted region has the potential to significantly increase OH via the above 33 mechanism, especially in the morning hours thereby muting this effect. Constraining modeled HONO 34 again lowers the magnitude of these ClNO<sub>2</sub> induced morning perturbations to 25 and 50% increases in OH and the sum of HO<sub>2</sub> and RO<sub>2</sub>, respectively. Again, while even these HONO-rich perturbations are
 significant, these findings, together with the vertically resolved estimates of Young et al. (2012), further
 imply that strong vertical gradients in HONO will influence the ClNO<sub>2</sub> effects on morning oxidant
 evolution.

5

6 CINO<sub>2</sub> formation and photolysis has implications for the reactive nitrogen budget as well. CINO<sub>2</sub> is 7 relatively unreactive at night in these regions, thereby building up and allowing transport of NOx 8 downwind of the urban core with morning photolysis of CINO2 analogous to thermal decomposition of 9 acyl peroxy nitrates such as PAN. In addition to this direct impact on NO<sub>x</sub> transport, we find significant 10 perturbations to acyl peroxy nitrate formation to occur in the with-ClNO<sub>2</sub> case (see Figure 5B). A 50% 11 enhancement in total APN occurs before noon, and the enhancement remains elevated at around 10% 12 for the remainder of the day. This ClNO2 induced perturbation to APN formation largely persists even 13 when measured HONO values are incorporated. Additionally, the sum of alkyl nitrates is also enhanced 14 by 15% before noon with individual alkyl nitrates enhanced up to 60% in the with-ClNO<sub>2</sub> case. 15 Increasing the partitioning of NO<sub>x</sub> into reservoirs such as APN implies a greater potential for export of 16  $NO_x$  to remote regions. In fact, while CINO<sub>2</sub> formation extends the lifetime of NO<sub>x</sub> through the night, 17 our model suggests that faster OH + NO2, APN formation, and alkyl nitrate formation due to increased 18  $RO_x$  levels during the subsequent day largely offset this effect, such that  $NO_x$  is ~6% lower from sunrise 19 until noon in the with-ClNO<sub>2</sub> than in the without-ClNO<sub>2</sub> case.

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## 21 3.3 Impact on ozone production rate and odd-oxygen

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23 We illustrate the influence of  $CINO_2$  chemistry on ozone production in Figure 5C (also see 24 Supplemental Figure S-12). Over the entire model day, the difference between the integrated ozone 25 production rate with 1.5 ppbv CINO<sub>2</sub> and the integrated ozone production rate without CINO<sub>2</sub> is about 26 12 ppbv. The majority of this enhancement takes place over the first 5 hours after sunrise where at 6:30 27 AM the ozone production rate is enhanced by  $\sim 200\%$ . The pre-noon ozone mixing ratios relative to the 28 without-ClNO<sub>2</sub> case are increased by  $\sim 20\%$  with  $\sim 10\%$  increase over the remainder of the day including 29 peak ozone which occurs at about 5 PM. Such an influence is potentially large enough to affect 30 attainment of air quality standards in polluted coastal regions where exceedences are often only tens of ppbv over the current standard (Parrish et al., 2010; Qin et al., 2004; U.S. EPA, 2006). The 31 32 enhancement in ozone production scales nearly linearly with the ClNO<sub>2</sub> yield for this region, as 33 expected, given that the ozone production rate is approximately linear with the primary radical source in 34 a NOx-saturated environment like Los Angeles. Constraining modeled HONO to the observations

results in a slightly smaller 9 ppbv enhancement in the integrated ozone production rate due to ClNO<sub>2</sub>
 chemistry.

3

4 The above result is also interesting to consider in terms of the impact of nocturnal nitrogen oxide 5 chemistry on the odd-oxygen budget. Defining  $O_x$  as the sum of  $O_3$  and  $NO_2$ , our model predicts that 6 N<sub>2</sub>O<sub>5</sub> reactions on aerosol particles consume 9 ppbv O<sub>x</sub> at night. If we neglect ClNO<sub>2</sub> formation, this 9 7 ppbv  $O_x$  is permanently lost due to nitrate formation from  $N_2O_5$  hydrolysis. However, incorporating 8  $CINO_2$  formation, with a yield (branching ratio) of 50%, results in up to 12 ppbv O<sub>x</sub> produced the 9 subsequent day compared to the case where CINO<sub>2</sub> formation is neglected. Thus, due to CINO<sub>2</sub> 10 formation and its daytime impact on oxidants and ozone, nighttime N<sub>2</sub>O<sub>5</sub> chemistry does not net destroy 11 O<sub>x</sub> but is in fact potentially a net source, or at least a null cycle, for the Los Angeles region conditions 12 we simulate here.

13

14 While not directly comparable, our results appear generally consistent with a recent 3D CMAQ 15 modeling study of CINO<sub>2</sub> effects on ozone and particulate nitrate (Sarwar et al., 2012). In the Los Angeles region, the CMAQ modeling showed roughly a 2 – 4 ppbv increase in daytime ozone per ppbv 16 17 ClNO<sub>2</sub> photolyzed, with maxima approaching 8 ppbv/ppbv. Likely important in setting the actual ozone 18 enhancement caused by CINO2 is, among other possibilities, the extent to which a model mixes 19 background marine air with the polluted core during transport and the model predicted vertical 20 distribution of CINO2. These issues will be important to test with observations in order to validate 21 model representations of this process.

22

#### 23 4 Summary and conclusions

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25 These model results suggest that CINO<sub>2</sub> photolysis is likely a major Cl<sup>.</sup> source, if not the dominant 26 source, under conditions similar to those sampled in the Los Angeles region during CalNex 2010. The 27 impact of CINO<sub>2</sub> on potential daytime halogen atom recycling is substantial, with significant enhancements 28 predicted on other Cl reservoirs like ClONO2, HOCl, and Cl2. Relative to model runs without ClNO2 29 formation, the presence of ClNO<sub>2</sub> causes significant and non-negligible perturbations in HO<sub>x</sub>, RO<sub>2</sub>, APN, 30 and ozone production. Relative to a model without ClNO<sub>2</sub> formation and heterogeneous HONO 31 production, incorporating CINO<sub>2</sub> perturbed the integrated total radical and ozone production rates by 32 20%, with perturbations in  $RO_x$  and APN >100%. Moreover, we show that, given these effects, the 33 impact of N<sub>2</sub>O<sub>5</sub> reactions on aerosol particles is not a net sink of odd-oxygen but instead a net source for 34 the polluted coastal conditions we model here. The absolute magnitude of the perturbations in these quantities and processes relative to a model that does not include ClNO<sub>2</sub> will ultimately depend upon the
 presence of HONO and the abundance of ClNO<sub>2</sub> and HONO vertically as well as seasonally.

3

We conclude by noting that during winter, in locations such as the Northeastern U.S., the role of  $CINO_2$ may be substantially more important to the total radical budget given that  $O(^1D)$  production and  $H_2O$ vapor concentrations can both be factors of 5 lower than presented here, resulting in more than an order of magnitude reduction in primary OH abundances while  $CINO_2$  approaches similar concentrations (Kercher et al., 2009). This idea is consistent with the apparently important role of  $CINO_2$  at inland locations during wintertime as illustrated by recent studies at the Uintah Basin, Utah (Edwards et al., 2013).

11

## 12 Acknowledgements

13 This work was supported by a grant from the National Science Foundation (NSF CAREER ATM-14 0846183 to J.A.T). T. P. R. is grateful for an Earth System Science graduate fellowship from the 15 National Aeronautics and Space Administration (NASA NESSF NNX10AN48H). G. M. W. acknowledges support from a NOAA Climate and Global Change Postdoctoral Fellowship administered 16 17 by the University Corporation for Atmospheric Research. This research was support in part by the 18 NOAA Health of the Atmosphere Program. We also thank the crew of the R/V Atlantis and the 19 Pasadena ground site science team for their tireless efforts and continual support throughout the CalNex 20 study.

1	Figure Captions
2	
3	Figure 1. Model output for the analysis period of a model run showing ClNO <sub>2</sub> mixing ratios (heavy
4	green line, right y-axis) and Cl· concentrations for the case including ClNO <sub>2</sub> formation (heavy blue line,
5	left y-axis) and the case excluding ClNO <sub>2</sub> formation (dashed red line, left y-axis).
6	
7	Figure 2. Model calculated Cl <sup>.</sup> production channels (A) without ClNO <sub>2</sub> formation (top) and (B) with
8	CINO <sub>2</sub> formation (bottom).
9	
10	Figure 3. Mixing ratios of various chlorinated species tracked in the model for the with-ClNO <sub>2</sub> model
11	case.
12	
13	Figure 4. Modeled Cl reactivity at 7 AM (left) and 3 PM (right) grouped by reactant types.
14	
15	Figure 5. (A) The $HO_x$ production rate with $CINO_2$ formation (solid blue line) and without $CINO_2$
16	formation (dashed red line). (B) The ratio of the hydroxyl radical (solid black line), hydroperoxyl
17	radical (dashed green line), sum of organic peroxy radicals (solid pink line), and sum of acyl peroxy
18	nitrates (dotted cyan line) for the with-ClNO <sub>2</sub> case relative to the without-ClNO <sub>2</sub> case. (C) The
19	difference between the integrated ozone production rate with ClNO <sub>2</sub> formation and the integrated ozone

- 20 production rate without  $ClNO_2$  formation.
- 21

#### 1 References

- Allan, W., Struthers, H., and Lowe, D. C.: Methane carbon isotope effects caused by atomic chlorine in
   the marine boundary layer: Global model results compared with Southern Hemisphere
   measurements, Journal of Geophysical Research: Atmospheres, 112, D04306, doi:
   10.1029/2006jd007369, 2007.
- Atkinson, R., Baulch, D. L., Cox, R. A., Crowley, J. N., Hampson, R. F., Hynes, R. G., Jenkin, M. E.,
  Rossi, M. J., and Troe, J.: Evaluated kinetic and photochemical data for atmospheric chemistry:
  Volume I gas phase reactions of Ox, HOx, NOx and SOx species, Atmos. Chem. Phys., 4,
  1461-1738, doi: 10.5194/acp-4-1461-2004, 2004.
- Badger, C. L., Griffiths, P. T., George, I., Abbatt, J. P. D., and Cox, R. A.: Reactive Uptake of N<sub>2</sub>O<sub>5</sub> by
   Aerosol Particles Containing Mixtures of Humic Acid and Ammonium Sulfate, The Journal of
   Physical Chemistry A, 110, 6986-6994, doi: 10.1021/jp0562678, 2006.
- Bertram, T. H., and Thornton, J. A.: Toward a general parameterization of N<sub>2</sub>O<sub>5</sub> reactivity on aqueous
   particles: the competing effects of particle liquid water, nitrate and chloride, Atmospheric
   Chemistry and Physics, 9, 8351-8363, doi: 10.5194/acp-9-8351-2009, 2009.
- Bertram, T. H., Thornton, J. A., Riedel, T. P., Middlebrook, A. M., Bahreini, R., Bates, T. S., Quinn, P.
   K., and Coffman, D. J.: Direct observations of N<sub>2</sub>O<sub>5</sub> reactivity on ambient aerosol particles, Geophysical Research Letters, 36, L19803, doi: 10.1029/2009gl040248, 2009.
- Bloss, C., Wagner, V., Jenkin, M. E., Volkamer, R., Bloss, W. J., Lee, J. D., Heard, D. E., Wirtz, K.,
  Martin-Reviejo, M., Rea, G., Wenger, J. C., and Pilling, M. J.: Development of a detailed
  chemical mechanism (MCMv3.1) for the atmospheric oxidation of aromatic hydrocarbons,
  Atmos. Chem. Phys., 5, 641-664, doi: 10.5194/acp-5-641-2005, 2005.
- Buys, Z., Brough, N., Huey, L. G., Tanner, D. J., von Glasow, R., and Jones, A. E.: High temporal
   resolution Br<sub>2</sub>, BrCl and BrO observations in coastal Antarctica, Atmospheric Chemistry and
   Physics, 13, 1329-1343, doi: 10.5194/acp-13-1329-2013, 2013.
- Caloz, F., Fenter, F. F., and Rossi, M. J.: Heterogeneous Kinetics of the Uptake of CIONO<sub>2</sub> on NaCl
   and KBr, The Journal of Physical Chemistry, 100, 7494-7501, doi: 10.1021/jp953099i, 1996.
- Daum, P. H., Kleinman, L., Imre, D. G., Nunnermacker, L. J., Lee, Y. N., Springston, S. R., Newman,
   L., and Weinstein-Lloyd, J.: Analysis of the processing of Nashville urban emissions on July 3
   and July 18, 1995, Journal of Geophysical Research: Atmospheres, 105, 9155-9164, doi:
   10.1029/1999jd900997, 2000.
- Deiber, G., George, C., Le Calve, S., Schweitzer, F., and Mirabel, P.: Uptake study of ClONO<sub>2</sub> and
   BrONO<sub>2</sub> by Halide containing droplets, Atmospheric Chemistry and Physics, 4, 1291-1299, doi:
   10.5194/acp-4-1291-2004, 2004.
- Dusanter, S., Vimal, D., Stevens, P. S., Volkamer, R., Molina, L. T., Baker, A., Meinardi, S., Blake, D.,
  Sheehy, P., Merten, A., Zhang, R., Zheng, J., Fortner, E. C., Junkermann, W., Dubey, M., Rahn,
  T., Eichinger, B., Lewandowski, P., Prueger, J., and Holder, H.: Measurements of OH and HO2
  concentrations during the MCMA-2006 field campaign Part 2: Model comparison and radical
  budget, Atmos. Chem. Phys., 9, 6655-6675, doi: 10.5194/acp-9-6655-2009, 2009.
- Edwards, P. M., Young, C. J., Aikin, K., deGouw, J., Dubé, W. P., Geiger, F., Gilman, J., Helmig, D.,
  Holloway, J. S., Kercher, J., Lerner, B., Martin, R., McLaren, R., Parrish, D. D., Peischl, J.,
  Roberts, J. M., Ryerson, T. B., Thornton, J., Warneke, C., Williams, E. J., and Brown, S. S.:
  Ozone photochemistry in an oil and natural gas extraction region during winter: simulations of a
  snow-free season in the Uintah Basin, Utah, Atmos. Chem. Phys., 13, 8955-8971, doi:
  10.5194/acp-13-8955-2013, 2013.
- Fan, J., and Zhang, R.: Atmospheric Oxidation Mechanism of Isoprene, Environmental Chemistry, 1,
   140-149, doi: 10.1071/en04045, 2004.
- Finlayson-Pitts, B. J., Ezell, M. J., and Pitts, J. N.: Formation of chemically active chlorine compounds
   by reactions of atmospheric NaCl particles with gaseous N<sub>2</sub>O<sub>5</sub> and ClONO<sub>2</sub>, Nature, 337, 241 244, doi: 10.1038/337241a0, 1989.

Finley, B. D., and Saltzman, E. S.: Measurement of Cl<sub>2</sub> in coastal urban air, Geophysical Research 1 2 Letters, 33, doi: 10.1029/2006gl025799, 2006. 3 Finley, B. D., and Saltzman, E. S.: Observations of Cl<sub>2</sub>, Br<sub>2</sub>, and I<sub>2</sub> in coastal marine air, Journal of 4 Geophysical Research-Atmospheres, 113, D21301, doi: 10.1029/2008jd010269, 2008. 5 Foster, K. L., Plastridge, R. A., Bottenheim, J. W., Shepson, P. B., Finlayson-Pitts, B. J., and Spicer, C. 6 W.: The role of Br<sub>2</sub> and BrCl in surface ozone destruction at polar sunrise, Science, 291, 471-7 474, doi: 10.1126/science.291.5503.471, 2001. 8 Hanson, D. R., and Ravishankara, A. R.: Reactive uptake of CIONO2 onto sulfuric-acid due to reaction 9 with HCl and H<sub>2</sub>O, Journal of Physical Chemistry, 98, 5728-5735, doi: 10.1021/j100073a026, 10 1994. 11 Hanson, D. R., Ravishankara, A. R., and Solomon, S.: Heterogeneous reactions in sulfuric-acid aerosols 12 - a framework for model-calculations, Journal of Geophysical Research-Atmospheres, 99, 3615-13 3629, doi: 10.1029/93jd02932, 1994. 14 Jaegle, L., Steinberger, L., Martin, R. V., and Chance, K.: Global partitioning of NO<sub>x</sub> sources using 15 satellite observations: Relative roles of fossil fuel combustion, biomass burning and soil 16 emissions, Faraday Discussions, 130, 407-423, doi: 10.1039/B502128F, 2005. 17 Jenkin, M. E., Saunders, S. M., and Pilling, M. J.: The tropospheric degradation of volatile organic compounds: A protocol for mechanism development, Atmospheric Environment, 31, 81-104, 18 19 doi: 10.1016/s1352-2310(96)00105-7, 1997. 20 Jenkin, M. E., Saunders, S. M., Wagner, V., and Pilling, M. J.: Protocol for the development of the 21 Master Chemical Mechanism, MCM v3 (Part B): tropospheric degradation of aromatic volatile 22 organic compounds, Atmospheric Chemistry and Physics, 3, 181-193, 2003. 23 Jobson, B. T., Niki, H., Yokouchi, Y., Bottenheim, J., Hopper, F., and Leaitch, R.: Measurements of 24 C2-C6 hydrocarbons during the Polar Sunrise1992 Experiment: Evidence for Cl atom and Br 25 atom chemistry, Journal of Geophysical Research: Atmospheres, 99, 25355-25368, doi: 26 10.1029/94JD01243, 1994. 27 Karlsson, R. S., and Ljungström, E. B.: Laboratory Study of CINO: Hydrolysis, Environmental Science 28 & Technology, 30, 2008-2013, doi: 10.1021/es950801f, 1996. 29 Kato, S., Sato, T., and Kajii, Y.: A method to estimate the contribution of unidentified VOCs to OH 30 reactivity, Atmospheric Environment, 45, 5531-5539, doi: 10.1016/j.atmosenv.2011.05.074, 31 2011. 32 Keene, W. C., Stutz, J., Pszenny, A. A. P., Maben, J. R., Fischer, E. V., Smith, A. M., von Glasow, R., 33 Pechtl, S., Sive, B. C., and Varner, R. K.: Inorganic chlorine and bromine in coastal New England air during summer, Journal of Geophysical Research-Atmospheres, 112, D10S12, doi: 34 35 10.1029/2006jd007689, 2007. Keil, A. D., and Shepson, P. B.: Chlorine and bromine atom ratios in the springtime Arctic troposphere 36 37 as determined from measurements of halogenated volatile organic compounds, Journal of 38 Geophysical Research: Atmospheres, 111, D17303, doi: 10.1029/2006JD007119, 2006. 39 Kercher, J. P., Riedel, T. P., and Thornton, J. A.: Chlorine activation by N<sub>2</sub>O<sub>5</sub>: simultaneous, in situ 40 detection of CINO2 and N2O5 by chemical ionization mass spectrometry, Atmospheric 41 Measurement Techniques, 2, 193-204, doi: 10.5194/amt-2-193-2009, 2009. 42 Kleinman, L. I.: The dependence of tropospheric ozone production rate on ozone precursors, 43 Atmospheric Environment, 39, 575-586, doi: 10.1016/j.atmosenv.2004.08.047, 2005. 44 Lawler, M. J., Sander, R., Carpenter, L. J., Lee, J. D., von Glasow, R., Sommariva, R., and Saltzman, E. 45 S.: HOCl and Cl<sub>2</sub> observations in marine air, Atmospheric Chemistry and Physics, 11, 7617-7628, doi: 10.5194/acp-11-7617-2011, 2011. 46 Mao, J., Ren, X., Chen, S., Brune, W. H., Chen, Z., Martinez, M., Harder, H., Lefer, B., Rappenglück, 47 48 B., Flynn, J., and Leuchner, M.: Atmospheric oxidation capacity in the summer of Houston 49 2006: Comparison with summer measurements in other metropolitan studies, Atmospheric 50 Environment, 44, 4107-4115, doi: http://dx.doi.org/10.1016/j.atmosenv.2009.01.013, 2010.

McNeill, V. F., Patterson, J., Wolfe, G. M., and Thornton, J. A.: The effect of varying levels of 1 2 surfactant on the reactive uptake of N2O5 to aqueous aerosol, Atmospheric Chemistry and 3 Physics, 6, 1635-1644, doi: 10.5194/acp-6-1635-2006, 2006.

6

7

8

- Mentel, T. F., Sohn, M., and Wahner, A.: Nitrate effect in the heterogeneous hydrolysis of dinitrogen 4 5 pentoxide on aqueous aerosols, Physical Chemistry Chemical Physics, 1, 5451-5457, doi: 10.1039/a905338g, 1999.
  - Mielke, L. H., Furgeson, A., and Osthoff, H. D.: Observation of CINO2 in a Mid-Continental Urban Environment, Environmental Science & Technology, 45, 8889-8896, doi: 10.1021/es201955u, 2011.
- 10 Mielke, L. H., Stutz, J., Tsai, C., Hurlock, S. C., Roberts, J. M., Veres, P. R., Froyd, K. D., Hayes, P. L., Cubison, M. J., Jimenez, J. L., Washenfelder, R. A., Young, C. J., Gilman, J. B., de Gouw, J. A., 11 Flynn, J. H., Grossberg, N., Lefer, B. L., Liu, J., Weber, R. J., and Osthoff, H. D.: 12 13 Heterogeneous formation of nitryl chloride and its role as a nocturnal NO<sub>x</sub> reservoir species 14 during CalNex-LA 2010, Journal of Geophysical Research: Atmospheres, 118, 10,638-610,652, 15 doi: 10.1002/jgrd.50783, 2013.
- Osthoff, H. D., Roberts, J. M., Ravishankara, A. R., Williams, E. J., Lerner, B. M., Sommariva, R., 16 17 Bates, T. S., Coffman, D., Quinn, P. K., Dibb, J. E., Stark, H., Burkholder, J. B., Talukdar, R. K., Meagher, J., Fehsenfeld, F. C., and Brown, S. S.: High levels of nitryl chloride in the 18 19 polluted subtropical marine boundary layer, Nature Geoscience, 1, 324-328, doi: 20 10.1038/ngeo177, 2008.
- 21 Parrish, D. D., Aikin, K. C., Oltmans, S. J., Johnson, B. J., Ives, M., and Sweeny, C.: Impact of 22 transported background ozone inflow on summertime air quality in a California ozone 23 exceedance area, Atmos. Chem. Phys., 10, 10093-10109, doi: 10.5194/acp-10-10093-2010, 24 2010.
- 25 Phillips, G. J., Tang, M. J., Thieser, J., Brickwedde, B., Schuster, G., Bohn, B., Lelieveld, J., and 26 Crowley, J. N.: Significant concentrations of nitryl chloride observed in rural continental Europe 27 associated with the influence of sea salt chloride and anthropogenic emissions, Geophysical 28 Research Letters, 39, L10811, doi: 10.1029/2012g1051912, 2012.
- Platt, U., Allan, W., and Lowe, D.: Hemispheric average Cl atom concentration from <sup>13</sup>C/<sup>12</sup>C ratios in 29 atmospheric methane, Atmos. Chem. Phys., 4, 2393-2399, doi: 10.5194/acp-4-2393-2004, 2004. 30
- 31 Qin, Y., Tonnesen, G. S., and Wang, Z.: One-hour and eight-hour average ozone in the California South 32 Coast air quality management district: trends in peak values and sensitivity to precursors, Atmospheric Environment, 38, 2197-2207, doi: 33 http://dx.doi.org/10.1016/j.atmosenv.2004.01.010, 2004. 34
- Raff, J. D., Njegic, B., Chang, W. L., Gordon, M. S., Dabdub, D., Gerber, R. B., and Finlayson-Pitts, B. 35 36 J.: Chlorine activation indoors and outdoors via surface-mediated reactions of nitrogen oxides 37 with hydrogen chloride, Proceedings of the National Academy of Sciences of the United States 38 of America, 106, 13647-13654, doi: 10.1073/pnas.0904195106, 2009.
- 39 Ren, X. R., Harder, H., Martinez, M., Lesher, R. L., Oliger, A., Simpas, J. B., Brune, W. H., Schwab, J. J., Demerjian, K. L., He, Y., Zhou, X. L., and Gao, H. G.: OH and HO<sub>2</sub> chemistry in the urban 40 atmosphere of New York City, Atmospheric Environment, 37, 3639-3651, doi: 10.1016/s1352-41 2310(03)00459-x, 2003. 42
- 43 Riedel, T. P., Bertram, T. H., Crisp, T. A., Williams, E. J., Lerner, B. M., Vlasenko, A., Li, S.-M., 44 Gilman, J., de Gouw, J., Bon, D. M., Wagner, N. L., Brown, S. S., and Thornton, J. A.: Nitryl Chloride and Molecular Chlorine in the Coastal Marine Boundary Layer, Environmental Science 45 & Technology, 46, 10463-10470, doi: 10.1021/es204632r, 2012a. 46
- Riedel, T. P., Bertram, T. H., Ryder, O. S., Liu, S., Day, D. A., Russell, L. M., Gaston, C. J., Prather, K. 47 A., and Thornton, J. A.: Direct N<sub>2</sub>O<sub>5</sub> reactivity measurements at a polluted coastal site, 48 49 Atmospheric Chemistry and Physics, 12, 2959-2968, doi: 10.5194/acp-12-2959-2012, 2012b.
- 50 Riemer, D. D., Apel, E. C., Orlando, J. J., Tyndall, G. S., Brune, W. H., Williams, E. J., Lonneman, W. 51 A., and Neece, J. D.: Unique isoprene oxidation products demonstrate chlorine atom chemistry

occurs in the Houston, Texas urban area, Journal of Atmospheric Chemistry, 61, 227-242, doi: 1 2 10.1007/s10874-009-9134-5, 2008. 3 Roberts, J. M., Osthoff, H. D., Brown, S. S., and Ravishankara, A. R.: N<sub>2</sub>O<sub>5</sub> oxidizes chloride to Cl<sub>2</sub> in acidic atmospheric aerosol, Science, 321, 1059-1059, doi: 10.1126/science.1158777, 2008. 4 5 Rubasinghege, G., and Grassian, V. H.: Surface-Catalyzed Chlorine and Nitrogen Activation: 6 Mechanisms for the Heterogeneous Formation of CINO, NO, NO2, HONO, and N2O from 7 HNO3 and HCl on Aluminum Oxide Particle Surfaces, The Journal of Physical Chemistry A, 8 116, 5180-5192, doi: 10.1021/jp301488b, 2012. 9 Ryerson, T. B., Andrews, A. E., Angevine, W. M., Bates, T. S., Brock, C. A., Cairns, B., Cohen, R. C., 10 Cooper, O. R., de Gouw, J. A., Fehsenfeld, F. C., Ferrare, R. A., Fischer, M. L., Flagan, R. C., Goldstein, A. H., Hair, J. W., Hardesty, R. M., Hostetler, C. A., Jimenez, J. L., Langford, A. O., 11 McCauley, E., McKeen, S. A., Molina, L. T., Nenes, A., Oltmans, S. J., Parrish, D. D., Pederson, 12 13 J. R., Pierce, R. B., Prather, K., Quinn, P. K., Seinfeld, J. H., Senff, C. J., Sorooshian, A., Stutz, 14 J., Surratt, J. D., Trainer, M., Volkamer, R., Williams, E. J., and Wofsy, S. C.: The 2010 15 California research at the Nexus of air quality and climate change (CalNex) field study, Journal of Geophysical Research: Atmospheres, n/a-n/a, doi: 10.1002/jgrd.50331, 2013. 16 17 Santschi, C., and Rossi, M. J.: The heterogeneous interaction of HOCl with solid KBr substrates: The catalytic role of adsorbed halogens, Physical Chemistry Chemical Physics, 7, 2599-2609, doi: 18 19 10.1039/b503071d, 2005. 20 Sarwar, G., and Bhave, P. V.: Modeling the effect of chlorine emissions on ozone levels over the eastern United States, Journal of Applied Meteorology and Climatology, 46, 1009-1019, doi: 21 22 10.1175/jam2519.1, 2007. 23 Sarwar, G., Simon, H., Bhave, P., and Yarwood, G.: Examining the impact of heterogeneous nitryl 24 chloride production on air quality across the United States, Atmospheric Chemistry and Physics, 25 12, 6455-6473, doi: 10.5194/acp-12-6455-2012, 2012. 26 Saunders, S. M., Jenkin, M. E., Derwent, R. G., and Pilling, M. J.: Protocol for the development of the 27 Master Chemical Mechanism, MCM v3 (Part A): tropospheric degradation of non-aromatic 28 volatile organic compounds, Atmospheric Chemistry and Physics, 3, 161-180, 2003. 29 Scheer, V., Frenzel, A., Behnke, W., Zetzsch, C., Magi, L., George, C., and Mirabel, P.: Uptake of 30 nitrosyl chloride (NOCl) by aqueous solutions, Journal of Physical Chemistry A, 101, 9359-9366, doi: 10.1021/jp972143m, 1997. 31 32 Spicer, C. W., Chapman, E. G., Finlayson-Pitts, B. J., Plastridge, R. A., Hubbe, J. M., Fast, J. D., and 33 Berkowitz, C. M.: Unexpectedly high concentrations of molecular chlorine in coastal air, Nature, 34 394, 353-356, doi: 10.1038/28584, 1998. 35 Spicer, C. W., Plastridge, R. A., Foster, K. L., Finlayson-Pitts, B. J., Bottenheim, J. W., Grannas, A. M., and Shepson, P. B.: Molecular halogens before and during ozone depletion events in the Arctic 36 37 at polar sunrise: concentrations and sources, Atmospheric Environment, 36, 2721-2731, doi: 38 10.1016/s1352-2310(02)00125-5, 2002. 39 Tanaka, P. L., Allen, D. T., McDonald-Buller, E. C., Chang, S., Kimura, Y., Mullins, C. B., Yarwood, 40 G., and Neece, J. D.: Development of a chlorine mechanism for use in the carbon bond IV 41 chemistry model, Journal of Geophysical Research: Atmospheres, 108, 4145, doi: 10.1029/2002jd002432, 2003. 42 43 Thornton, J. A., Braban, C. F., and Abbatt, J. P. D.: N<sub>2</sub>O<sub>5</sub> hydrolysis on sub-micron organic aerosols: 44 the effect of relative humidity, particle phase, and particle size, Physical Chemistry Chemical Physics, 5, 4593-4603, doi: 10.1039/b307498f, 2003. 45 Thornton, J. A., Kercher, J. P., Riedel, T. P., Wagner, N. L., Cozic, J., Holloway, J. S., Dube, W. P., 46 Wolfe, G. M., Quinn, P. K., Middlebrook, A. M., Alexander, B., and Brown, S. S.: A large 47 atomic chlorine source inferred from mid-continental reactive nitrogen chemistry, Nature, 464, 48 49 271-274, doi: 10.1038/nature08905, 2010. 50 U.S. EPA: Air Quality Criteria for Ozone and Related Photochemical Oxidants (2006 Final), U.S. 51 Environmental Protection Agency, Washington, DC, EPA/600/R-05/004aF-cF, 2006. 30

- Wagner, N. L., Riedel, T. P., Roberts, J. M., Thornton, J. A., Angevine, W. M., Williams, E. J., Lerner,
   B. M., Vlasenko, A., Li, S. M., Dube, W. P., Coffman, D. J., Bon, D. M., de Gouw, J. A.,
   Kuster, W. C., Gilman, J. B., and Brown, S. S.: The sea breeze/land breeze circulation in Los
   Angeles and its influence on nitryl chloride production in this region, Journal of Geophysical
   Research-Atmospheres, 117, D00V24, doi: 10.1029/2012jd017810, 2012.
- Warneke, C., Veres, P., Holloway, J. S., Stutz, J., Tsai, C., Alvarez, S., Rappenglueck, B., Fehsenfeld,
  F. C., Graus, M., Gilman, J. B., and de Gouw, J. A.: Airborne formaldehyde measurements using
  PTR-MS: calibration, humidity dependence, inter-comparison and initial results, Atmos. Meas.
  Tech., 4, 2345-2358, doi: 10.5194/amt-4-2345-2011, 2011.
- Wolfe, G. M., and Thornton, J. A.: The Chemistry of Atmosphere-Forest Exchange (CAFE) Model –
   Part 1: Model description and characterization, Atmos. Chem. Phys., 11, 77-101, doi:
   10.5194/acp-11-77-2011, 2011.
- Young, C. J., Washenfelder, R. A., Roberts, J. M., Mielke, L. H., Osthoff, H. D., Tsai, C., Pikelnaya, O.,
  Stutz, J., Veres, P. R., Cochran, A. K., VandenBoer, T. C., Flynn, J., Grossberg, N., Haman, C.
  L., Lefer, B., Stark, H., Graus, M., de Gouw, J., Gilman, J. B., Kuster, W. C., and Brown, S. S.:
  Vertically Resolved Measurements of Nighttime Radical Reservoirs; in Los Angeles and Their
  Contribution to the Urban Radical Budget, Environmental Science & Technology, 46, 1096510973, doi: 10.1021/es302206a, 2012.
- Young, C. J., Washenfelder, R. A., Edwards, P. M., Parrish, D. D., Gilman, J. B., Kuster, W. C., Mielke,
   L. H., Osthoff, H. D., Tsai, C., Pikelnaya, O., Stutz, J., Veres, P. R., Roberts, J. M., Griffith, S.,
   Dusanter, S., Stevens, P. S., Flynn, J., Grossberg, N., Lefer, B., Holloway, J. S., Peischl, J.,
   Ryerson, T. B., Atlas, E. L., Blake, D. R., and Brown, S. S.: Evaluating evidence for Cl sources
   and oxidation chemistry in a coastal, urban environment, Atmos. Chem. Phys. Discuss., 13,
- 24 13685-13720, doi: 10.5194/acpd-13-13685-2013, 2013.
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1 FIGURES





















