Response to Anonymous Referee #1 re: Report #1 on Revised Manuscript

We are very thankful for the referee's continued assistance in improving our manuscript. We believe that an acceptable conclusion has been reached regarding the issue of air-surface exchange. Our comments are found embedded in the referee's text below in red.

Report #1
Submitted on 18 Nov 2013
Anonymous Referee #1

Summary: The authors have addressed some of the points raised, but not all of them. The model may be suitable to simulate exposure near the sources, but not in remote areas, which cover certainly more than half of the domain, at least not for semivolatiles, due to insufficient process resolution (air-surface exchange).
I recommend to reject the paper or insist that the results are limited to non-volatile PAHs (as suggested in the review to the submitted version).

Description of model used:
Sufficient in the revised version

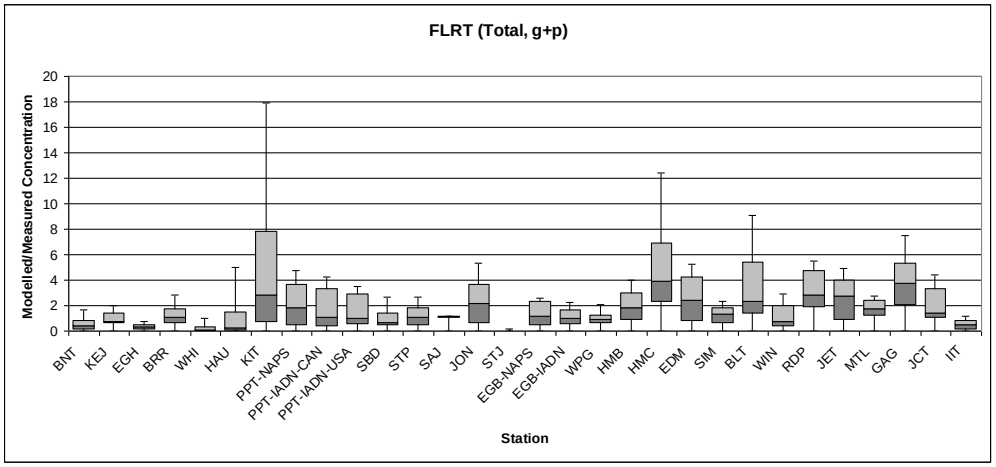
Gas-particle partitioning – discussion of unplausible model results:
The authors refuse to discuss, but plan to address in a subsequent publication.

Air-surface exchange and model evaluation by comparison with observed data / choice of monitoring sites:
The authors maintain that the neglect of air-surface exchange is insignificant as these "parametrizations likely have a minor effect in regions subject to ongoing PAH emissions" and "atmospheric PAH burdens in North America are likely not dominated by air-surface exchange".
However, in model study with air-surface exchange considered it was found for two semivolatile PAHs addressed in the Galarneau et al. study (anthracene and fluoranthene) that volatilization from soil exceeds depositions in summer not only in remote areas of North America (the US West, British Columbia, Ontario), but also in areas certainly influenced by sources (the US Midwest, southern Ontario, California) (Fig. S7b in Lammel et al., 2009). The reason is that the characteristic travel distance from the source in combination with lifetimes in soil lead to burdens in soil, decreasing with distance from source, which turn into secondary emissions and may seasonally exceed the deposition of advected primary and secondary emissions and model will underpredict concentrations at remote sites.

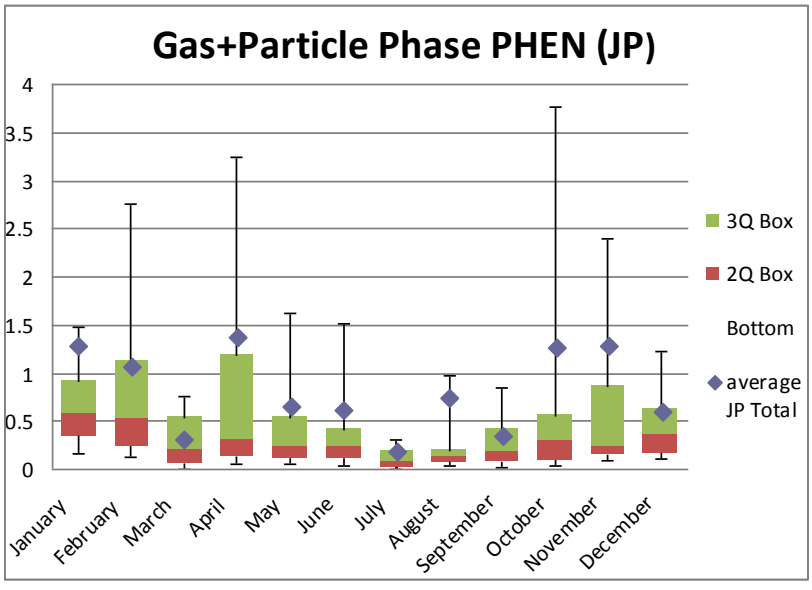
The atmospheric levels of fluoranthene and pyrene, respectively, are expected to drop to 37% and 14% (1/e and $1/e^2$, respectively) within characteristic travel distances from the sources of 160-450 km and 320-900 km, respectively. (These ranges conservatively represent upper estimates as based on the climatological 24-hour mean oxidant level, (0.38-1.24)x106 OH cm-3 within the model domain (Spivakovsky et al.. J. Geophys. Res. 2000), while more PAH is emitted during day-time.)

As also noted by a third referee, the potential impact of air-surface exchange needs to be acknowledged more fully in our manuscript and we have conducted the following analyses to support our new revision.

We have re-plotted the site-specific FLRT results with the x-axis now depicting measurement sites in increasing order of population within 25km. As can be seen in the figure, there is no clear relationship between population and underestimation by the model. In fact, the highest population of 3.86 million at IIT (Chicago) exhibits one of the lowest overall modelled-to-measured ratios.



However, as a result of referee comments, we have also replotted our data by month as shown below for PHEN. As can be seen in the figure, the model underprediction is greatest in summer which is consistent with the absence of volatilisation (or other potential missing sources or overestimated sinks). We observe a similar pattern for the other volatile PAHs (ANTH, FLRT and PYR) but not for the less volatile BaA, C+T and BaP. Since such behaviour is consistent with volatilisation in summer months, we have acknowledged this limitation on the volatile PAH results in our abstract and have included new figures and discussion (3.1.1) in the latest revision of the manuscript.



The authors insist that the spatial resolution (42km) is high enough to reproduce the gradients in urban areas. "As seen in Figure 4, the model performance does not differ much between sites." However, the sites indicating worst model performance (modelled/measured <0.1), WHI and STJ, are remote sites, unlike most other sites tested.

We apologize for the miscommunication. We did not mean to suggest that 42km resolution reproduces gradients **in** urban areas, but rather that this regional resolution is able to reproduce a substantial portion of the gradient between urban and rural areas.

Response to Anonymous Referee #3 re: Report #2 on Revised Manuscript

We are very thankful to the third referee for taking the time to review our revised manuscript. Responses to the report are embedded below in red.

Report #2
Submitted on 08 Dec 2013
Anonymous Referee #3

The authors need to re-write the latter part of Section 3.1.1 (Results & Discussion) and include a new figure (see referee report)

The authors present a revised manuscript detailing a modelling study that simulates PAH air concentrations across North America using a Eulerian CTM model. The simulation is based on the year 2002 and includes 7 PAHs that cover a relatively wide range in physical-chemical properties. This is one of only a handful of studies that has attempted to predict air concentrations at high spatial-resolution for these chemicals and considerable effort has been expended on gas-particle partitioning given the semi-volatile nature of these compounds.
The primary emissions of PAHs appear to be very well handled and detailed. These build on a methodology previously published by the lead author. Similarly a thorough model description is provided on the various fate/behaviour processes, including the use of two gas/particle partitioning approaches. The results are usefully presented including a thorough comparison with measurement data (modelled/measured ratios). I believe the paper is of publishable quality for ACP. However, the authors need to do more to address the concerns raised by a previous reviewer regarding the contribution of temperature-driven re-emission from secondary sources. Below are two
key points which the authors need to highlight/address:
(1) For the 7 PAHs studied here secondary sources (re: air-surface exchange) are most pertinent for phenanthrene (PHEN). Its isomer anthracene (ANTH) is probably too reactive for this processes to be fully significant (i.e. its half-life in air and soil is relatively short, plus its presence in air is erratic and can be driven in the main by notable point/primary sources).

The emissions inventory developed by the lead author was based on an extensive collection of published emissions data. No difference in emissions sources between PHEN and ANTH was observed; all sources had emissions of both isomers though the proportions varied among sources.

The relative importance of losses through OH reactivity will likely have an important effect on the total mass of the different volatile PAHs. Both emissions from air-surface exchange and losses due to OH reactivity would likely be at a maximum in the summer. Hence, while ANTH's increased OH reactivity might make it disappear from the air more quickly than PHEN, one might still observe patterns consistent with air-surface exchange. PAH emitting activities with a strong seasonal cycle (e.g. biomass burning) may also contribute to some of the missing summer mass. A test of the sensitivity of model results to the relative magnitudes of these processes would be interesting to pursue.

For the higher MW 4-ringed PAHs of fluoranthene and pyrene the magnitude of temperature—driven re-emission from secondary sources will be low relative to primary emissions (particularly over a single summer season re: 2002). Secondary emissions are negligible for the high MW PAHs (e.g. B[a]P). The authors need to make these points clear in the manuscript and cite studies that have observed this phenomenon and provide a few quantitative details (i.e. the observed seasonality in low MW PAH concentrations brought about by secondary reemission) (there are plenty of good studies in the literature for soil, water bodies, vegetation and urban surfaces).

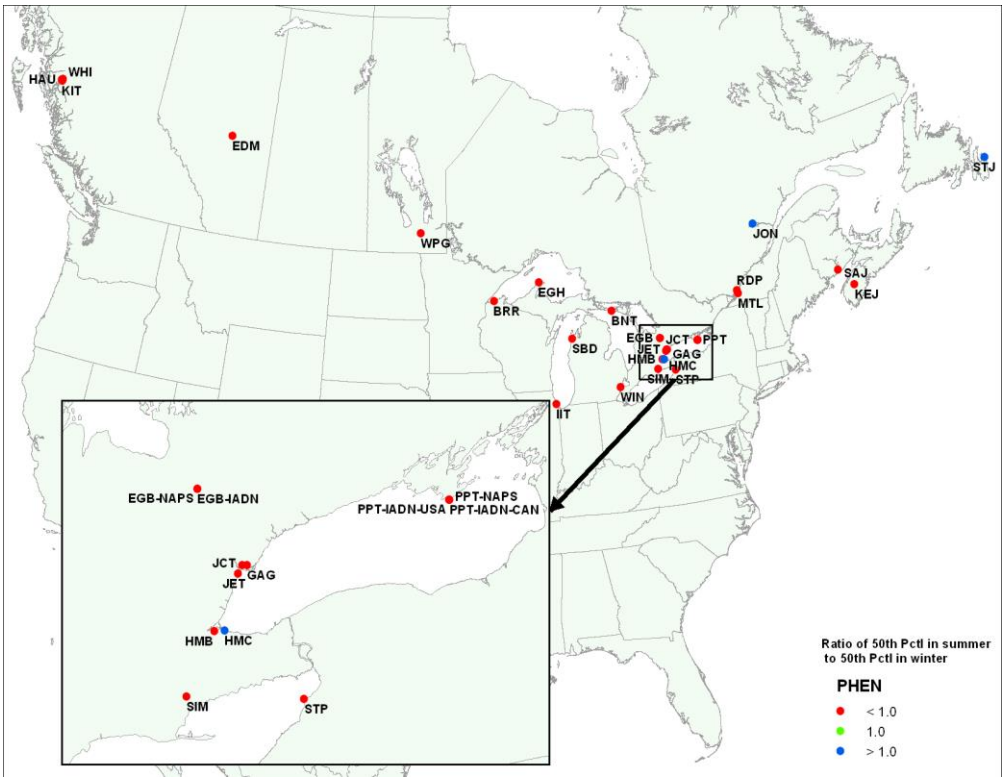
Please see our analysis below for information regarding the 4- and 5-ring species. We have added statements to Section 3.1.1 regarding the existing literature.

(2) Section 3.1.1. para (line 419-434). The authors discuss the model results with regards to the exclusion of air-surface exchange in their model. This paragraph is essential but I disagree with the flow of their discussion as they state that the model summary results provide indefinite evidence. Far from it(!) – the model appears to be working very well - and from the evidence provided both in Table 1 (summary comparison between model and measured) and Fig 3, the model appears to fit with PAH behaviour as described in point (1) above. For PHEN, the model clearly under predicts the air concentrations. There is nothing ambiguous about this and the underestimated PHEN concentrations in the model must presumably be due to the reemission component which is missing from the model. This is fine and not a problem for the manuscript, but the authors need to state this clearly and examine this issue more carefully and include this observation/finding in the abstract. For example, I would like to see a map of the spatial distribution of model/measurement ratios for PHEN, for both winter (i.e. Dec – Feb) and summer (i.e. June-Aug). The hypothesis here is that the ratios will be closer to 1 in the winter and deviate further from 1 in the summer (i.e. as the re-emission component becomes more relevant during the warmer part of the year). The authors should 'play up' this part of the paper and not try to discount discrepancies as ambiguous or uncertain. Another case in point is ANTH. As ANTH is relatively volatile it is not surprising that the model average is below the measurement average (again akin to PHEN re: re-emission component) but the data are variable and skewed (hence similar model/measured medians). This is just what I would expect for ANTH released from notable point sources but with a relatively short half-life. In short, the authors need to highlight the role played by air-surface exchange on PHEN and ANTH concentrations, and use the model results to demonstrate/quantify this process rather than discount it as model ambiguity. In fact, I believe this could be a real strength to this paper, more so than the gas-particle (g-p) partitioning story (especially as the two g-p approaches show little difference).

Thank you for your suggestions regarding our model's systematic underprediction of phenanthrene. Such underestimation might be caused by
(1) underestimated emissions
(2) overestimated removal processes (OH reaction, deposition)
(3) overestimated dispersion from source areas
Processes within each of these three categories can be identified by examining differences or similiarities between species and over spatiotemporal scales.

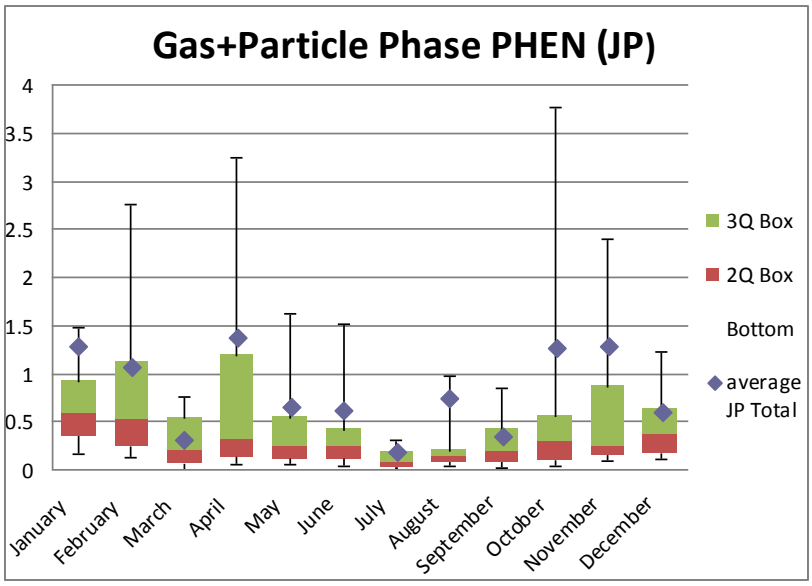
If case (1) is at play and emissions are underestimated due to missing air-surface exchange, we should observe modelled-to-measured concentration ratios that
(1) vary from lower in summer to higher in winter
(2) vary from lower in rural/background locations to higher in urban/industrial locations
(3) show a systematic decrease in underprediction from the most volatile species to the least

Attached is the figure requested by the reviewer. The modelled-to-measured concentration ratios were calculated for each site in summer and winter. The summer-to-winter ratio of the median values is plotted for each location at which measurement data were available. Modelled-to-measured concentration ratios for PHEN are indeed smaller in summer than in winter at all sites except Hamilton (Confederation Park), Jonquière, and St. John's. The first two of the latter locations are cities with substantial point source emissions from steel and aluminum processing, respectively.

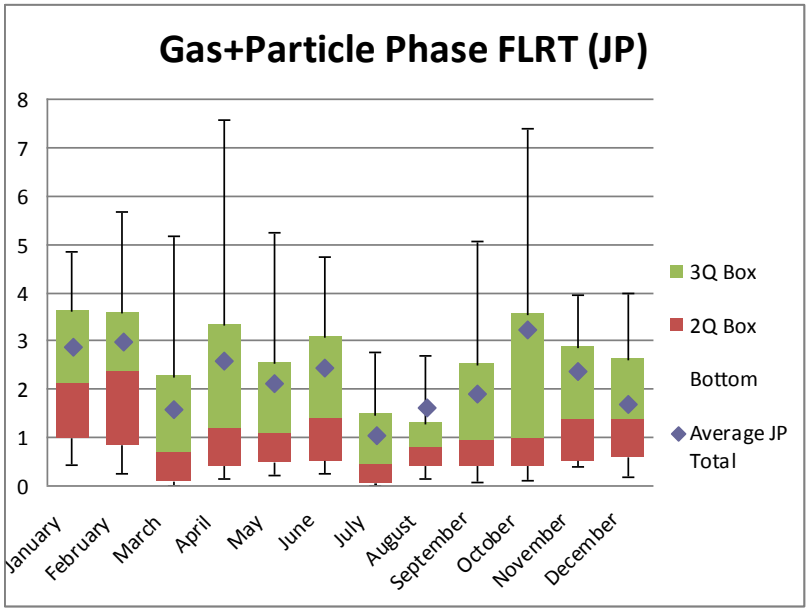


This result appears to support the argument that the underprediction of PHEN by our model is due to the missing air-surface exchange process. However, we observe similar patterns for ANTH, FLRT and PYR and not for BaA, C+T or BaP (not shown).

Since we find the map somewhat difficult to read, we have produced box-and-whisker plots of modelled-to-measured concentrations that are segregated by month. Therein we see a temporal pattern that shows greater underprediction in summer than in winter for PHEN.

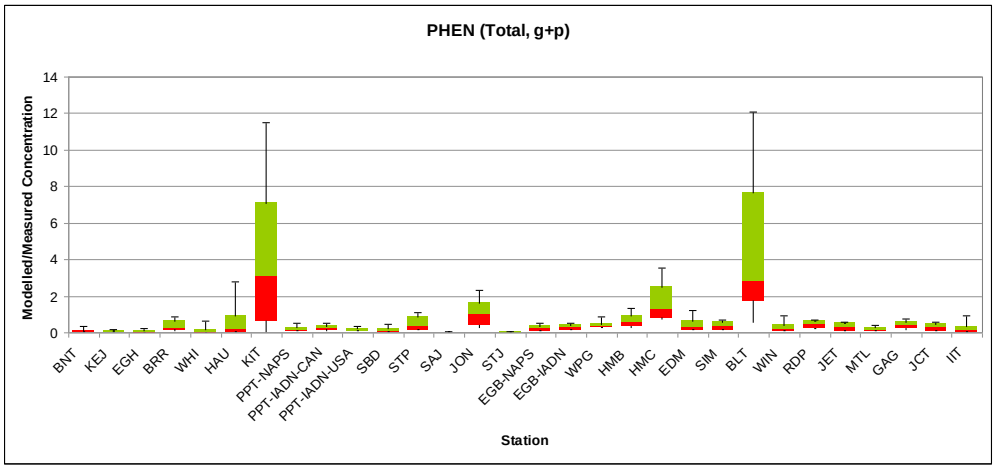


As with the maps, we see this again for the next three PAHs (increasing GC retention order/decreasing volatility). For example, FLRT exhibits a minimum in modelled-to-measured concentration ratio during the summer months. Yet this was the PAH species for which the model performed best and the referee did not expect air-surface exchange to be significant.



We have examined the issue in another way as well. We have plotted the box-and-whisker plots by location with the x-axis order indicating increasing population within 25km of the measurement site. If missing air-surface exchange is the cause of PHEN underprediction by the model, one

would expect to see the greatest underpredictions at the least populated sites and a steady increase toward equiprediction at the most populated or industrial locations.



Though the figure shows overpredictions at industrial sites (e.g., Jonquière JON and Hamilton HMC) and some of the most extreme underpredictions at the most rural sites (BNT, KEJ, and EGH), the pattern is not consistent. Underpredictions are also seen at SAJ and STJ and the site with the highest population (IIT Chicago) exhibits the 10th most extreme underprediction.

As a result of these augmented analyses, we respectfully suggest that the variation in model performance over the annual cycle is consistent with the absence of air-surface exchange. However, since ANTH, FLRT and PYR show similar temporal patterns to PHEN but are not similarly underpredicted, an additional factor must be contributing to the underprediction of PHEN. We suspect that the emissions for this species are erroneously low and we will be revisiting the inventory for future projects.

We hope that the addition of a new two-part figure (monthly PHEN and PYR box-and-whisker plots) and the following text to Section 3.1.1 will satisfy the request made by the referee. We have also added a statement to the Abstract.

"Though results are equivocal on an annual basis, monthly patterns observed in the model output are consistent with the absence of a seasonal source (e.g., air-surface exchange). Volatilisation from a variety of environmental compartments is typically stronger in warmer periods than in cooler ones (e.g., Nelson et al., 1998; Smith et al., 2001; Motelay-Massei et al., 2005; Bozlaker et al., 2008; Wang et al., 2011). Figure 4 shows the monthly distribution of modelled-to-measured concentration ratios for PHEN and PYR. Both exhibit higher values in winter than in summer as do ANTH and FLRT whereas this seasonality is not observed for the higher molecular BaA, C+T or BaP (not shown). These findings are consistent with a missing volatilisation source that emits during warmer weather. However, other factors could also be involved including overestimated loss terms (e.g., oxidation, deposition) or underestimated emissions (e.g., forest fires) during warmer periods. The investigation of the relevant causes is a priority for future model development. Regardless of the causes, the seasonal effect on model output appears to be compounded by further, as yet unidentified factors whereby PHEN is underpredicted throughout the year and ANTH, FLRT and PYR are overpredicted through some seasons, potentially due to air-surface exchange that leads to net deposition during cooler months."

PAH Concentrations Simulated with the AURAMS-PAH Chemical Transport Model over Canada and the USA

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Abstract

The off-line Eulerian AURAMS (A Unified Regional Air quality Modelling System) chemical transport model was adapted to simulate airborne concentrations of seven PAHs: phenanthrene, anthracene, fluoranthene, pyrene, benz[a]anthracene, chrysene+triphenylene, and benzo[a]pyrene. The model was then run for the year 2002 with hourly output on a grid covering southern Canada and the continental USA with 42-km horizontal grid spacing. Model predictions were compared to ~5,000 24-hour-average PAH measurements from 45 sites, most of which were located in urban or industrial areas. Eight of the measurement sites also provided data on particle/gas partitioning which had been modelled using two alternative schemes. This is the first known regional modelling study for PAHs over a North American domain and the first modelling study at any scale to compare alternative particle/gas partitioning schemes against paired field measurements. The goal of the study was to provide output concentration maps of use to assessing human inhalation exposure to PAHs in ambient air. Annual average modelled total (gas + particle) concentrations were statistically indistinguishable from measured values for fluoranthene, pyrene and benz[a]anthracene whereas the model underestimated concentrations of phenanthrene, anthracene and chrysene+triphenylene. Significance for benzo[a]pyrene performance was close to the statistical threshold and depended on the particle/gas partitioning scheme employed. On a day-to-day basis, the model simulated total PAH concentrations to the correct order of magnitude the majority of the time. The model showed seasonal differences in prediction quality for volatile species which suggests that a missing emission source such as air-surface exchange should be included in future versions. Model performance differed substantially between measurement locations and the limited available evidence suggests that the model spatial resolution was too coarse to capture the distribution of concentrations in densely populated areas. A more detailed analysis of the factors influencing modelled particle/gas partitioning is warranted based on the findings in this study.

1. **Introduction**

Polycyclic aromatic hydrocarbons (PAHs) are ubiquitous air pollutants that tend to be most concentrated in areas of dense human population (Hafner et al., 2005) but are also detected at locations remote from local sources (Hung et al., 2005). Many PAH species have been classified as carcinogens (IARC, 2010) and they are implicated routinely as toxicants in airborne particulate matter (Kelly and Fussell, 2012). They are regulated under international agreements such as the Aarhus Protocol on Persistent Organic Pollutants. Benzo[a]pyrene, a commonly-reported PAH species, is subject to ambient air guidelines in many jurisdictions.

In Canada, PAHs meet the criteria for inclusion on the Toxic Substances List of the Canadian Environmental Protection Act (Environment Canada and Health Canada, 1994), and the resulting government obligation has been to reduce or minimise their release into the environment. Nationwide anthropogenic emissions of benzo[a]pyrene, a commonly-reported species, fell by 70% between 1990 and 2010 according to estimates made by the National Pollutant Release Inventory (Environment Canada, 2012). Though there are no federal guidelines for PAHs in Canadian air, a recent analysis of ambient monitoring data found that measured PAH concentrations regularly exceed the health-based guidelines set by the Canadian province of Ontario (Galarneau and Dann, 2011).

In the USA, PAHs are listed as Clean Air Act Hazardous Air Pollutants as part of the polycyclic organic matter (POM) class of compounds (US EPA, 2012) and have been identified as a regional cancer concern in the US National-Scale Air Toxics Assessment (US EPA, 2012). Industrial releases to air reported to the US Toxics Release Inventory (TRI) fell by 35% between 1995 and 2010 (US EPA, 2012). There is no federal US guideline for PAHs in ambient air.

PAH measurements are labour-intensive compared to those of criteria air contaminants such as ozone and particulate matter, and the processes governing their atmospheric fate are not yet well-understood. In an attempt to elucidate the spatiotemporal distributions of PAH sources and ambient concentrations, several numerical modelling studies have been published. Lagrangian frameworks have been used for Europe (Van Jaarsveld et al., 1997; Halsall et al., 2001) and China (Liu et al., 2007; Lang et al., 2007; Lang et al., 2008). Others studies have used box modelling (Prevedouros et al., 2004) and multimedia fate approaches (Yaffe et al., 2001; Prevedouros et al., 2008). Eulerian chemical transport models (CTMs) have been developed for Europe (Shatalov, 2005; Aulinger et al., 2007; Matthias et al., 2009; Gusev et al., 2011; Bieser et al., 2012) and east Asia (Zhang et al., 2009; 2011a; 2011b; Inomata et al., 2012), and three such studies on a global scale have also been published in recent years (Sehili and Lammel, 2007; Lammel et al., 2009; Friedman and Selin, 2012).

The aforementioned studies differ in many respects relating to the PAH species examined, the temporal variability of their emissions, and the spatial resolutions and process representations in the models. None has focussed exclusively on North America at the regional scale. As well, although several particle/gas partitioning mechanisms have been explored in other models, including Junge-Pankow adsorption (Junge, 1977;

Pankow, 1987), organic matter sorption (Finizio et al., 1997), and combined adsorption/absorption (Dachs and Eisenreich, 2000), no previous studies have evaluated model output against paired phase-distributed measurements for alternative partitioning expressions on the same domain.

This study presents the results of a chemical transport model, AURAMS-PAH, run over North America at 42-km horizontal grid spacing with hourly output for the year 2002. Seven PAH species were simulated with the model. Three isomer pairs of decreasing volatility and increasing particulate fraction comprise six of the species: phenanthrene (PHEN) and anthracene (ANTH) (178 g mol^{-1}), fluoranthene (FLRT) and pyrene (PYR) (202 g mol^{-1}), and benz[a]anthracene (BaA) and chrysene/triphenylene (C+T) (228 g mol^{-1}). The seventh PAH, benzo[a]pyrene (BaP) (252 g mol^{-1}), is not generally considered to be semivolatile but has been included due to its common use as a representative PAH species. Two particle/gas partitioning schemes, Junge-Pankow (JP: Junge, 1977; Pankow, 1987) and Dachs-Eisenreich (DE: Dachs and Eisenreich, 2000),were tested.

Model performance was evaluated against ~5,000 measurements from 45 stations in established networks in Canada and the USA. This is the first published model to be run and evaluated for PAH concentrations and their distributions between the particle and gas phases using two partitioning methods. It is also the first such model to be evaluated over a regional North American domain.

2. Methods

2.1 Model Description

AURAMS (A Unified Regional Air quality Modelling System) is an Eulerian CTM originally developed to simulate criteria air contaminants. The standard version of the model uses a sectional approach to represent the size distribution of airborne particles: 12 size bins from 0.01 to 40.96 μm in diameter and 9 particulate species (sulphate, nitrate, ammonium, elemental carbon, primary organic aerosol, secondary organic aerosol, crustal material, sea salt, and aerosol water) are usually considered. The model includes process representation for tropospheric gas-phase oxidative chemistry, the absorptive formation of secondary organic aerosols, inorganic heterogeneous chemistry, particle microphysics (nucleation, condensation, coagulation, etc.), cloud processing of aerosols, advective transport, vertical diffusion, and gas and particle emissions and deposition. A detailed overall description of AURAMS appears in Gong et al. (2006) while a description of the aerosol sectional approach and the microphysics modules of the model can be found in Gong et al. (2003a,b). Performance evaluation and model intercomparison results for AURAMS appear in McKeen et al. (2008), Smyth et al. (2009), Makar et al. (2010), Kelly et al. (2012) and Solazzo et al. (2012) among other publications.

A modified version of the AURAMS CTM known as AURAMS-PAH was developed to incorporate primary semivolatile organic compounds that are subject to sorptive partitioning. Starting from the standard AURAMS CTM had the advantage that a

number of required fields for modelling PAHs were already available. These included hydroxyl concentration, total particle surface area, and fractions of particle elemental carbon and organic carbon. The modifications made to AURAMS version 1.3.2 in order to simulate PAHs are described below. Physico-chemical property values used for each PAH in the modified code are found in Table S1.1 of the Supplementary Material.

2.1.1 Dry Deposition of Gases. Within AURAMS, gaseous dry deposition velocities are modelled using the inverse resistance analogy for several land-use categories (Zhang et al., 2002). Three resistances are assessed in AURAMS and only the first of these (aerodynamic resistance) is independent of the chemical species under consideration. The species-dependent resistances are the quasi-laminar sub-layer resistance and the surface or canopy resistance. The latter both depend on the gas-phase diffusivity of the compound in question, and this quantity was calculated in the model according to the Fuller et al. method described in Reid et al. (1987).

Surface or canopy resistance is the most complex of the three gaseous dry deposition component resistances and tends to dominate total dry deposition (Zhang et al., 2002). One of its sub-components, mesophyll resistance, was set to 100 s m^{-1} for species that are relatively insoluble in water and have small oxidizing capacities, as is the case for PAHs. The remaining sub-components (cuticle and ground resistances) are determined by scaling to O_3 and SO_2 settings based on physico-chemical qualifications. For the PAHs, scaling factors to O_3 and SO_2 for both acetaldehyde and C_3 carbonyls, the least soluble organic compounds considered in AURAMS aside from the PAHs, were used. Unsubstituted compounds such as PAHs are generally considered to have high resistances to deposition whereas carbonyl resistances are thought to be lower (Zhang et al., 2002). However, published observations of PAH deposition led us to assume that deposition velocities would be greater than zero (low resistances) and we therefore used the best-available homologues in AURAMS to represent PAHs. This is an uncertainty in the model that merits future attention.

Volatilisation of gaseous PAHs can occur from exposed water (Hoff et al., 1996), soil (Jones, 1994), and impervious urban surfaces (Diamond et al., 2000). Net gaseous deposition to the Great Lakes in 2002 was downward (Blanchard et al., 2005) suggesting that PAH fugacities in air exceeded those in surface compartments at the regional scale. Volatilisation was not included in this first-generation version of AURAMS-PAH and the effect of this omission is presented in Section 3.1.1.

2.1.2 Gas-Phase Reactions. Reactions of gas-phase PAHs with hydroxyl radicals are considered in this model. Since these reactions consume relatively little hydroxyl due to the trace concentrations of PAH, their reactions were simulated outside the AURAMS gas-phase chemistry solver. PAH oxidative loss was estimated as a first-order process using the model-predicted OH concentration immediately preceding particle-gas partitioning. Only seven new gas-phase concentration fields were added to the CTM; PAH reaction products were not tracked in the model, either as individual gas-phase species or as contributors to SOA.

Hydroxyl reaction rate constants were taken from the program AOPWIN which is part of the US EPA's EPI Suite (U.S. EPA, 2006). Measured constants are available for three low-molecular-weight PAHs considered here (phenanthrene, anthracene, fluoranthene) and these values were represented in AOPWIN. However, measurements for the remaining four PAHs are not available and the software predicted the same hydroxyl reaction rate constant of 50×10^{-12} cm^3 $molec^{-1}$ s^{-1} for these species.

2.1.3 Particle Representation of PAHs. Seven additional particle species, each with 12 size bins as in the original AURAMS configuration, were added to the model to represent the particle-bound PAH mass.

2.1.4 Particle/Gas Partitioning of PAHs. A new algorithm was developed for AURAMS-PAH to account for the sorptive particle/gas partitioning of PAHs. It is fully adaptable to other semivolatile species with similar atmospheric partitioning behaviour to PAHs such as dioxins and furans, PCBs, and organochlorine pesticides. The partitioning of PAHs to airborne particles was assumed to be fully reversible.

Two instantaneous equilibrium sorptive partitioning expressions were incorporated in the new partitioning subroutine. The first treated particle/gas partitioning as a Langmuirian adsorption process on a uniform particle surface (JP: Junge, 1977; Pankow, 1987). The model calculations began by adding the particulate PAH concentrations in all size bins (ΣC_p) and the gas-phase PAH concentration (C_g) to give a total PAH concentration (C_{TOT}) for each species. An updated bulk particulate fraction (ϕ) was then assigned according to the first part of Eq. (1):

$$\phi = \frac{c\Sigma\theta}{c\Sigma\theta + p_L^{\,0}} = \frac{\Sigma C_p}{C_{TOT}}, \qquad (1)$$

where c is a constant set at 0.173 J m^{-2} (estimated from Figure 3 in Junge, 1977), $\Sigma\theta$ is the total particle surface area concentration (m^2 m^{-3}) and $p_L^{\,0}$ is the saturated vapour pressure of the sub-cooled liquid (Pa) taken from the temperature-dependent values measured by Offenberg and Baker (1999; see Table S1.1). We have selected Junge's (1977) value of c over that estimated by Pankow (1987) since the latter was based on assumptions that have not been revisited in light of the numerous observations of PAH partitioning published since. The total particulate PAH concentrations dictated by ϕ were then redistributed among the particle size bins by prorating to the proportion of total aerosol surface area concentration within each size bin. The redistributed gas-phase PAH concentration was determined by difference between C_{TOT} and $\Sigma C_{p.}$

The second equilibrium partitioning expression available in the partitioning subroutine developed a partition coefficient (K_p, m^3 μg^{-1}) based on the contributions of two additive processes: absorption into particulate organic matter and adsorption onto particulate soot (DE: Dachs and Eisenreich, 2000)

$$K_p = 10^{-12}\left(1.5 f_{OC}/\rho_{oct} K_{OA} + f_{EC} K_{SA}\right) = \frac{\left(\Sigma C_p / C_{TSP}\right)}{C_g}, \quad (2)$$

where ρ_{oct} is the bulk density of octanol (0.82 kg L^{-1}), f_{OC} is the organic carbon fraction of the particulate matter (the 1.5 multiplier converts organic carbon to organic matter which is assumed to be well-represented by octanol), K_{OA} is the octanol-air partition coefficient (dimensionless), f_{EC} is the elemental carbon fraction of the particulate matter, K_{SA} is the soot-air partition coefficient (L kg^{-1}), ΣC_p is the particulate PAH concentration across all the size bins (ng m^{-3}), C_{TSP} is the total particulate matter concentration (μg m^{-3}), and C_g is the gas-phase concentration (ng m^{-3}).

Soot-air partition coefficients ($K_{SA,}$ L kg^{-1}) were estimated as the ratios of soot-water (K_{SW}) to air-water partition (K_{AW}) coefficients since direct K_{SA} measurements are not available for PAHs. K_{SW} values from Jonker and Koelmans (2002) were used in this model. These values vary substantially (up to a factor of 47) between relevant soots for each PAH considered here. Since a single K_{SW} was needed for each PAH in the model, representative values were determined by weighting the reported K_{SW} values by the contribution of their related combustion processes to the total emitted fine particulate matter ($PM_{2.5}$) used in the inventory of Galarneau et al. (2007). Temperature-dependent K_{AW} values were taken from Bamford et al. (1999). K_{OA} values were taken from the temperature-dependent expressions determined by Odabasi et al. (2006).

PAH partition coefficients were calculated according to the first part of Eq. (2). By determining the contribution of each size bin's organic matter and soot carbon to the totals across all size bins, the total particulate PAH was apportioned to each size bin. For example, if a total partition coefficient had contributions from the organic matter and soot carbon of 20% and 80%, respectively, and size bin 1 held 10% of the total particulate organic matter and 15% of the total soot carbon, the fraction of total particulate PAH assigned to size bin 1 would be 14% (viz., $0.2 \times 0.1 + 0.8 \times 0.15$). Gas-phase concentrations were then determined by difference between C_{TOT} and ΣC_p.

2.1.5 Below-cloud (Precipitation) Scavenging. Scavenging of gas and particle PAHs by liquid precipitation was calculated as per Gong et al. (2006). Particle scavenging assumed that particle-bound PAHs do not dissolve in falling rain; particle-bound PAHs were thus treated as passive aerosol tracers. Snow scavenging of gaseous PAHs was not considered in this version of AURAMS though particle-bound PAHs are scavenged by snow in the model as passive components of airborne particles.

2.1.6 Cloud Processing. Cloud processing in the model was treated in a similar manner to precipitation scavenging whereby gas-phase mass transfer to cloud water is species-dependent, whereas particulate interactions with cloud droplets are only affected by the presence of PAHs in terms of the size (mass and volume) that they represent as part of the overall aerosol. Solid-phase densities used to relate aerosol PAH mass to volume were taken from Mackay et al. (2006; see Table S1.1).

2.2 Model Domain, Emissions, and Boundary Conditions

The model domain included southern Canada and the continental USA (see Figure 1). It was run on a 42-km polar stereographic grid using off-line meteorology generated with the Global Environmental Multiscale numerical weather prediction model (GEM v 3.2.0: Côté et al., 1998a, b).

Emissions of PAHs were taken from the inventory of Galarneau et al. (2007) that had been updated from 2000 to 2002 and to which benzo[a]pyrene had been added using identical methods and data sources. As discussed in Galarneau et al. (2007), hourly PAH emissions fields were estimated with an emissions processing system using source-specific temporal profiles. The temporal profile library included 3020 month-of-year, 64 day-of-week, and 2672 hour-of-day temporal profiles for Canada and 1500, 49, and 680 analogous temporal profiles for the US. The overall temporal profile thus varies from grid cell to grid cell due to the different mixtures of source types found in each one.

All PAHs were emitted exclusively in the gas phase. Particle/gas partitioning took place at each 15-minute CTM time step according to the partitioning module described in Section 2.1.5. As mentioned in Section 2.1.2, no emissions of previously deposited PAHs were considered in this first-generation version of the model and the implications of this are discussed in Section 3.1.1. Emissions of SO_2, NO_x, NH_3, CO, volatile organic compounds (VOCs), and particulate matter were derived using Environment Canada and US EPA databases and methods for the year 2002.

Initial PAH concentrations at all lateral boundaries were set to zero in anticipation of pronounced spatial gradients away from localised source regions. As a result, modelled concentrations in Mexico and near its border with the US are not expected to be reliable, particularly since PAH emissions from Mexico have not been included in the model. Model output along the northern edge of the domain over western Canada is similarly expected to be unreliable since emission sources are located close to the model boundary in that region. The development of representative non-zero boundary concentrations is anticipated as part of future model development.

2.3 Evaluation Data

Observational PAH data used for comparison with model output were collected from four measurement networks: NAPS (Canada), IADN (Canada-US), CARB (California), and Rio Tinto Alcan (Kitimat, British Columbia, Canada). The measurement stations are depicted in Figure 2 and described in Section 2 of the Supplementary Material.

Measurement data were available from a total of 45 stations, 23 in Canada and 22 in the USA, all of which collected samples integrated over periods of 24 hours. Particle/gas partitioning was assessed at eight stations, three in Canada and five in the USA, all of which were operated by IADN.

The IADN phase-distributed data were also combined to yield total concentrations. These combined IADN data, along with NAPS and Rio Tinto data, yielded a total of 28 sites at which total PAH concentration for all the modelled PAHs could be assessed.

Particulate PAH measurements from the latter networks were determined from samples of total suspended particles (TSP). CARB provided data for benzo[a]pyrene in particles smaller than 2.5 μm in diameter ($PM_{2.5}$) at a further 17 locations.

Four model grid squares (Kitimat, Toronto, Hamilton, and Montreal) contained two or more measurement stations thus allowing for an assessment of the adequacy of modelling all seven PAHs at 42-km grid spacing.

3. Results

3.1 Total PAH Concentration

3.1.1 Overall Spatiotemporal Domain

Total PAH concentration refers to the sum of the gas and particulate concentrations whether these have been analysed together (e.g., NAPS) or separately (e.g., IADN). For stations at which the gas and particle phases were analysed separately, a valid total concentration was assumed to exist if at least one of the gas and particle phase concentrations was greater than the detection limit. Non-detectable values were assumed equal to zero for the calculation of total concentrations.

A representative plot of the spatial distribution of modelled annual average concentrations is presented in Figure 1 for fluoranthene. The remaining PAHs show similar spatial distributions and maps of their modelled concentrations are found in Section 3 of the Supplementary Material. All the PAHs show spatial distributions of their modelled concentrations that are consistent with regional dispersion of their emissions as depicted in Galarneau et al. (2007).

A summary of annual mean modelled and measured values over the entire spatiotemporal model domain is shown in Table 1. Only modelled values for which there was a corresponding measurement were included.

Table 1: Summary of 2002 Annual Modelled and Measured Total PAH Concentration Mean (Standard Deviation) Values (ng m^{-3})

PAH	Modelled – JP	Modelled - DE	Measured	n[1]
PHEN	12.75 (36.44)	12.76 (36.44)	36.06 (131.8)	790
ANTH	0.9123 (1.757)	0.9104 (1.759)	2.804 (11.56)	701
FLRT	6.781 (14.40)	6.888 (14.66)	9.179 (32.44)	789
PYR	5.727 (12.23)	6.009 (13.40)	5.733 (21.57)	785
BaA	1.227 (2.438)	1.328 (2.704)	1.326 (6.081)	610
C+T	1.511 (3.964)	1.473 (3.569)	3.303 (21.95)	721
BaP	1.173 (2.002)	1.424 (2.455)	0.9047 (3.238)	595

[1]n = number of modelled-measured data pairs

In comparing modelled results to measurements, the annual means were statistically indistinguishable at the 95% confidence level for FLRT, PYR, BaA, and BaP (JP) whereas they were statistically different for PHEN, ANTH, C+T, and BaP (DE). For

PHEN, ANTH, and C+T, modelled values were underestimated relative to measurements whereas they were overestimated for DE BaP.

The model's temporal variability tended to be smaller than that of the corresponding measurements; the relative standard deviations of the measurements were 1.3 to 2.7 times greater than those of the modelled values. A similar observation has been made in the modelling of particulate matter with AURAMS and other regional air quality models (Solazzo et al., 2012) For PAHs, this effect was also seen by Matthias et al. (2009) who concluded that temporal variability in PAH emissions was not adequately represented by their inventory. This is a plausible contributing factor in the current study as well. Furthermore, meteorological parameters vary over a scale much finer than that used for regional air quality models. As a result, observed concentrations from point locations can be expected to exhibit greater variability than modelled concentrations determined for entire grid squares.

Differences in mean modelled total concentrations between the two partitioning versions (JP and DE) were statistically indistinguishable at 95% significance despite the finding that the two BaP model results differed in their comparison to measured values. The latter anomaly indicates that the BaP distributions were close to the 95% confidence threshold. As a result, no conclusion can be drawn about which partitioning mechanism was superior in simulating overall total PAH concentrations. Phase partitioning of semivolatile organic compounds (SVOCs) is a major determinant of their potential for long-range transport (Bidleman, 1988), yet it does not appear to have a large effect on the simulation of their total concentrations at the regional scale. Model performance in simulating phase partitioning is discussed in Section 3.2.

The model's performance was also more closely evaluated by examining the pertinent data distributions. Figure 3 depicts frequency distributions of the ratios of modelled-to-measured concentrations for all of the valid data pairs available for the model evaluation. Four PAH species (ANTH, FLRT, PYR, and C+T) yielded median values of the modelled-to-measured concentration ratio that were close to the ideal value of unity (1.1, 1.1, 1.5, and 1.4, respectively). PHEN showed an overall tendency toward underestimation by the model (0.2), whereas BaA and BaP tended toward overestimation (3.2/3.5 and 3.0/3.5 JP/DE, respectively).

BaA and BaP are reactive PAHs (e.g., Behymer and Hites, 1985; Pöschl et al., 2001; Kwamena et al., 2004; Esteve et al., 2006, Shiraiwa et al., 2009) and the exclusion of particle-bound reactions in this first-generation model may explain a portion of their overestimation in AURAMS-PAH as suggested in a comparable model for Europe (Matthias et al., 2009). However, BaP is subject to losses during sampling (Menichini, 2009) and some portion of the apparent model overestimation may in fact be due to measured concentrations that are biased low since the samplers used were not equipped with oxidant denuders. This presents a priority for future research since many jurisdictions use BaP as an indicator PAH when setting air quality standards.

As presented in Section 2.1.2, volatilisation of gaseous PAHs from surface compartments such as water and soil was not included in this first-generation version of AURAMS-

PAH. If such volatilisation were significant to the balance of PAHs in ambient air relative to the other processes simulated, one would expect an overall bias in model results whereby the most volatile PAHs, which are found predominantly in the gas phase, would be underestimated and the least volatile particulate species would be unaffected. Summary results provide indefinite evidence. Volatile PHEN~~phenanthrene~~ is systematically underestimated yet its similarly volatile isomer, ANTH~~anthracene~~, shows an ambiguous central tendency whereby its mean concentrations are underestimated by the model (Table 1) but its median concentrations are not (Figure 3). Less volatile but nonetheless predominantly gaseous FLRT~~fluoranthene~~ and PYR~~pyrene~~ show no tendency toward underestimation.

Though results are equivocal on an annual basis, monthly patterns observed in the model output are consistent with the absence of a seasonal source (e.g., air-surface exchange). Volatilisation from a variety of environmental compartments is typically stronger in warmer periods than in cooler ones (e.g., Nelson et al., 1998; Smith et al., 2001; Motelay-Massei et al., 2005; Bozlaker et al., 2008; Wang et al., 2011). Figure 4 shows the monthly distribution of modelled-to-measured concentration ratios for PHEN and PYR. Both exhibit higher values in winter than in summer as do ANTH and FLRT whereas this seasonality is not observed for the higher molecular BaA, C+T or BaP (not shown). These findings are consistent with a missing volatilisation source that emits during warmer weather. However, other factors could also be involved including overestimated loss terms (e.g., oxidation, deposition) or underestimated emissions (e.g., forest fires) during warmer periods. The investigation of the relevant causes is a priority for future model development. Regardless of the causes, the seasonal effect on model output appears to be compounded by further, as yet unidentified factors whereby PHEN is underpredicted throughout the year and ANTH, FLRT and PYR are overpredicted through some seasons, potentially due to air-surface exchange that leads to net deposition during cooler months.

~~These results warrant further investigation but suggest that volatilisation of gaseous PAHs from surface compartments may not be significant at the regional scale when model results are compared to 24-hour measurements from predominantly urban locations. This finding stands in contrast to the reported importance of air-surface exchange at remote locations determined through global-scale modelling (Lammel et al., 2009).~~

The range of modelled-to-measured concentration ratios shown in Figure 3 varied substantially by species. The ratios of 90th to 10th percentile values for PHEN, FLRT, PYR, BaA, and C+T spanned fewer than, or close to, two orders of magnitude (55, 59, 67/68, 67/63, and 100/93, respectively). The ratio for BaP was larger (180/270) and that for ANTH was very large (5900/7400), with extreme values tending toward underestimation for the latter species. As seen with the comparison of means, the two partitioning parametrizations used by AURAMS-PAH led to similar model performance overall when considering the distribution of total PAH concentrations.

Additional quantitative performance metrics are presented for the two particle/gas partitioning parametrizations in Tables S4.1 and S4.2 of the Supplementary Material.

Normalized mean bias and error have been included for completeness, but their utility in this evaluation is questionable given the large range of concentrations. Measured maximum to minimum concentration ratios range from 4.7E+06 (PYR) to 1.3E+09/1.4E+09 (C+T). Therefore, the mean measured concentrations used to normalize the bias and error do not represent the dataset well.

The correspondence between individual modelled-measured data pairs is weak as demonstrated by the low coefficients of determination, non-unity slopes, and high intercepts listed in Tables S4.1 and S4.2. However, the ability of the model to simulate observed concentrations within a certain tolerance is reasonable, especially when considering that PAHs are trace organic compounds subject to numerous sampling artefacts (McDow, 1999) and poor measurement precision (Galarneau, 2008). Depending on PAH species, 22-34% of modelled-measured data pairs fell within a factor of 2 of each other. This increased to 61-86% when considering a factor of 10. As a result, it can be stated with confidence that, on average, AURAMS-PAH was able to simulate atmospheric PAH concentrations in North America for rural to urban locations to the correct order of magnitude.

3.1.2 Site-Specific Performance

Model performance was not spatially uniform. Figure 5~~4~~ depicts the variation in distributions of individual modelled-to-measured concentration ratios across measurement sites for fluoranthene, the PAH species for which overall performance was best as determined by the median and spread in modelled-to-measured concentration ratios. Note that only JP partitioning values have been plotted since these are visually indistinguishable from those for DE partitioning.

Of the 30 sites depicted in Figure 5~~4~~ (CARB sites could not be considered since only benzo[a]pyrene was reported there), the median modelled-to-measured concentration ratio ranged from 0.061 (St. John's) to 4.0 (Hamilton – Confederation Park), whereas the median value for all sites was 1.1. The variability at individual sites is itself highly variable, with ratios of 90^{th} to 10^{th} percentile values of the modelled-to-measured concentration ratio ranging from 5.8 (Toronto – Junction Triangle) to 105,000 (Haul Road, near the Rio Tinto Alcan smelter in Kitimat, British Columbia). A low value of 1.1 was observed for Saint John, but this was based on only two modelled-measured data pairs. Sixteen of the 30 sites (53%) had median modelled-to-measured ratios that fell within a factor of two of the median value for all sites.

The other compounds varied spatially in a manner similar to fluoranthene with the following exceptions. ANTH exhibited atypically large underestimation at the three sites near the Rio Tinto Alcan smelter in Kitimat, suggesting that inaccurately low ANTH emissions are associated with the dominant source there. The reporting threshold for point-source ANTH emissions through the Canadian National Pollutant Release Inventory (NPRI) system is higher than the thresholds for other commonly-measured PAHs and no ANTH emissions were reported to the NPRI by Rio Tinto Alcan for 2002. The C+T performance at Jonquière, home to aluminum smelting facilities, suggests that reported emissions there are also too low. Emissions for other PAHs were reported from

this location for 2002 but not so for chrysene, which is called benzo[a]phenanthrene in the NPRI.

3.1.3 Model Grid Squares Containing Multiple Measurement Sites

The smoother the spatial distribution of a pollutant, the coarser the model resolution that can be used to simulate it. Four AURAMS-PAH model grid squares contain more than one measurement site, thus allowing for an assessment of the 42-km spatial resolution used for the evaluation runs. The multi-site grid squares are all located in Canada, and from west to east, they encompass sites in Kitimat (2 sites), Hamilton (2), Toronto (3), and Montreal (2) (see Tables S2.1 and S2.4).

Kitimat is a town 650 km northwest of Vancouver with approximately 9,000 residents whose largest employer is the aluminum smelter complex operated by Rio Tinto Alcan (District of Kitimat, 2009). Two measurement sites (Haul Road and Kitamaat Village) are located in the same model grid square and a third site (Whitesail) lies in an adjacent square even though it is only a few kilometers away. Hamilton is a city at the western end of Lake Ontario that is known colloquially as the Steel Capital of Canada and had a population of approximately 700,000 in 2010. It is part of the so-called "Golden Horseshoe" conurbation at the western end of Lake Ontario whose 2010 population, estimated as the sum of the populations of Oshawa, Toronto, Hamilton, and St. Catharines-Niagara, was over 7 million (Statistics Canada, 2011). Toronto and Montreal are the largest cities in Canada having 2010 populations of 5.7 and 3.9 million, respectively.

Table 2 lists the variability in contemporaneously measured concentrations at the four grid squares as represented by their coefficients of variation (COV). At any given site, the average COVs for the different PAH species tend to be similar to each other. Substantial differences exist between sites, however, particularly when grouping the urban sites (Hamilton, Toronto, and Montréal) against the industrial site at Kitimat. This is not unexpected. Urban areas include complex mixtures of point, area, and mobile sources that are distributed over distances similar to the scale of the model. Kitimat houses industrial operations within a relatively small area of otherwise rural land and wilderness. Steep spatial gradients in pollutant concentrations are expected there as a result.

Table 2: Average Coefficient of Variation (%) between Contemporaneous Measurements at Sites Falling Within the Same 42-km AURAMS-PAH Grid Square

Station	PHEN	ANTH	FLRT	PYR	BaA	C+T	BaP	O_3[2]	TSP
Kitimat	106	101	96.5	93.7	87.4	91.3	93.2	N/A	N/A
Hamilton	52.1	59.0	59.2	58.8	73.1	117	62.8	24.9	30.5
Toronto3[1]	36.2	48.7	42.5	39.6	42.2	32.9	43.0	N/A	22.1
Toronto2[1]	36.5	45.4	39.3	35.7	32.9	25.0	38.0	12.2	24.6
Montréal	49.2	52.9	45.8	44.6	55.2	55.3	51.6	35.3	29.0

[1] Toronto3 includes data from all three Toronto measurement sites. Toronto2 includes only data from the Gage Institute and Judson & Etona because O_3 data were not available from Junction Triangle.

[2] Ozone data have been aggregated to 24-hour concentrations contemporaneous with PAH measurements.

The COVs for ozone and total suspended particles (TSP) have also been included in Table 2 as comparative gaseous and particulate pollutants, respectively. Both vary less between sites in the same grid square than do PAHs. Ozone and a portion of TSP are secondary pollutants created by the mixing and reaction of precursor compounds. The atmospheric residence times required for their creation is consistent with a smoothing of the spatial variability in their concentrations though ozone variability is further complicated by reactions with NO_x near emissions from mobile sources. Conversely, unsubstituted PAHs are primary pollutants whose concentrations would be expected to vary in space over a finer resolution when multiple sources are found close by.

The results presented above suggest that a 42-km spatial resolution is not sufficiently fine to represent PAH concentrations in areas close to sources such as cities and industrial areas if an average model accuracy better than an order of magnitude is desired. AURAMS modelling of fine particulate matter has shown substantial improvement when grid spacing has been reduced to 2.5 km (Stroud et al., 2011), and similar results can be expected for the modelling of PAHs. No 42-km model grid squares in rural or background areas away from sources contain multiple measurement stations and, as a result, a comparison cannot be made for these areas. However, it is expected that spatial variation in PAH concentrations will be less in such areas and, as such, a 42-km resolution model may be sufficient there.

3.2 Particle/Gas Partitioning

As noted in Section 3.1, the choice of partitioning expression (JP or DE) had little effect on the simulation of total PAH concentrations. This implies that the partitioning from each approach is sufficiently similar that regional-scale differences in removal rates between gaseous and particulate PAHs have little effect. However, differences between the two expressions with respect to simulating phase-resolved concentrations were noted.

3.2.1 Overall Spatiotemporal Domain

Figure 6~~5~~ shows frequency distributions of the ratios of individual modelled-to-measured particulate fraction for all data pairs available to the model evaluation. Note that only the eight IADN stations are included since the gas and particle phases are analysed separately only at those sites.

Figure 6~~5~~ shows that PAH particulate fraction is underestimated for all species except BaP. The degree of underestimation decreases with increasing molecular weight. The particulate fractions of volatile PHEN and ANTH (178 g mol^{-1}) are underestimated by approximately two orders of magnitude whereas equipartitioning BaA and C+T (228 g mol^{-1}) have particulate fractions that are underestimated by only a factor of two. A similar pattern appears when examining the partition coefficient, K_p (not shown).

For all species other than BaP, Dachs-Eisenreich partitioning performs slightly better than Junge-Pankow partitioning in simulating measured particulate fractions. The all-site median particulate fraction simulated using DE is between 1.1 (PYR) and 2.9 (ANTH)

847 times higher than that using JP. However, the performance of the partitioning
848 expressions is highly dependent on the physico-chemical property values used. For
849 example, estimated soot-air partition coefficients vary by more than an order of
850 magnitude (Galarneau et al., 2006) and translate directly to variations in predicted
851 partitioning by the Dachs-Eisenreich expression. For Junge-Pankow partitioning, the
852 value of the constant, c, in Eq. 1 and the estimation of aerosol surface area also introduce
853 uncertainties. A full analysis of the sensitivity of modelled partitioning is beyond the
854 scope of this paper and is explored in a separate publication (Galarneau et al., in prep.).
855
856 **3.2.2 Site-Specific Performance**
857
858 As was the case for total concentration, there is substantial variability in the simulation of
859 partitioning between sites. Figure 7~~6~~ shows the variation in frequency distribution of
860 individual modelled-to-measured particulate fraction for fluoranthene using Dachs-
861 Eisenreich partitioning. Model performance for particulate fraction simulation is better at
862 urban (Chicago) or urban-influenced (Sturgeon Point, Egbert) sites than at those that are
863 remote (Eagle Harbor). An analysis of measured partitioning at IADN stations
864 (Galarneau et al., 2006) found that the proportionality between partitioning and volatility
865 varied between sites, and in some cases, over the annual cycle. Volatility is included in
866 both the JP (through p_L^0) and DE (through K_{OA} and K_{SA}) partitioning expressions and the
867 proportionality between it and partitioning magnitude is much smaller in model outputs
868 than in measurements. As noted earlier, factors involved in the performance of model
869 partitioning such as modelled particulate matter concentration and composition are
870 explored in a separate publication (Galarneau et al., in prep.)
871
872 **4. Conclusions**
873
874 This study described the first known modelling results for atmospheric PAHs at the
875 regional scale over North America. Predictions from the AURAMS-PAH model were
876 compared to roughly 5,000 24-hour average PAH measurements from 45 sites, eight of
877 which also provided data on particle/gas partitioning which had been modelled using two
878 different partitioning schemes.
879
880 The evaluation of the model is key to determining its potential utility as an input for
881 estimating the impacts of PAH inhalation exposure on human health. Annual average
882 modelled total (gas + particle) concentrations were statistically indistinguishable from
883 measured values for fluoranthene, pyrene and benz[a]anthracene, indicating the model's
884 potential utility for providing inputs to health impact estimation for these species. The
885 model annual average concentrations for phenanthrene, anthracene and
886 chrysene+triphenylene were biased low. For these species, the negative bias would have
887 to be considered if used as inputs to human health impact estimates as the model in its
888 present form underestimates long-term exposure.
889
890 The utility of the model for prediction purposes may also be considered on a day-to-day
891 basis though this is less relevant to the chronic health effects associated with carcinogenic
892 PAHs. The model simulated total PAH concentrations to the correct order of magnitude
893 64-86% of the time. That level of accuracy must be considered when assessing human

health impacts; annual exposure estimates are likely of more utility with the model in its current state.

The partitioning approach chosen did not have a significant impact on the model results for total concentrations though differences resulting form the choice of parametrization approached the 95% significance level for benzo[a]pyrene. At this time, neither of the two approaches used here provided a clear advantage for simulation accuracy of total concentrations.

As a first work of this nature, the analysis has suggested several avenues for further model development and improvement. Improved temporal emissions estimates for PAHs are key to improving model simulations of these species; simulated PAHs showed less temporal variability than the measurements. The reactions of particulate PAH species with atmospheric oxidants should be given further consideration since the more reactive species were overestimated in the current model. The addition of an air-surface exchange parametrisation should be evaluated as a potential response to the seasonally varying prediction capability of the model for the most volatile compounds. Model resolution has been shown to be a key factor in improving air pollution estimates in areas with high human exposures. While the 42-km horizontal grid spacing used in this study is finer than that used in global models, it was insufficient to capture the distribution of concentrations in densely populated areas. A more detailed analysis of the factors influencing modelled particle/gas partitioning is needed to improve the distribution of PAHs between the gas and particle phases in the atmosphere given that both partitioning schemes used here showed increasing negative biases for particle-bound PAH concentrations of increasing volatility.

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List of Figures

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Figure 1: Map of modelled (JP) annual average total (gas + particle) fluoranthene concentrations (ng m^{-3}).

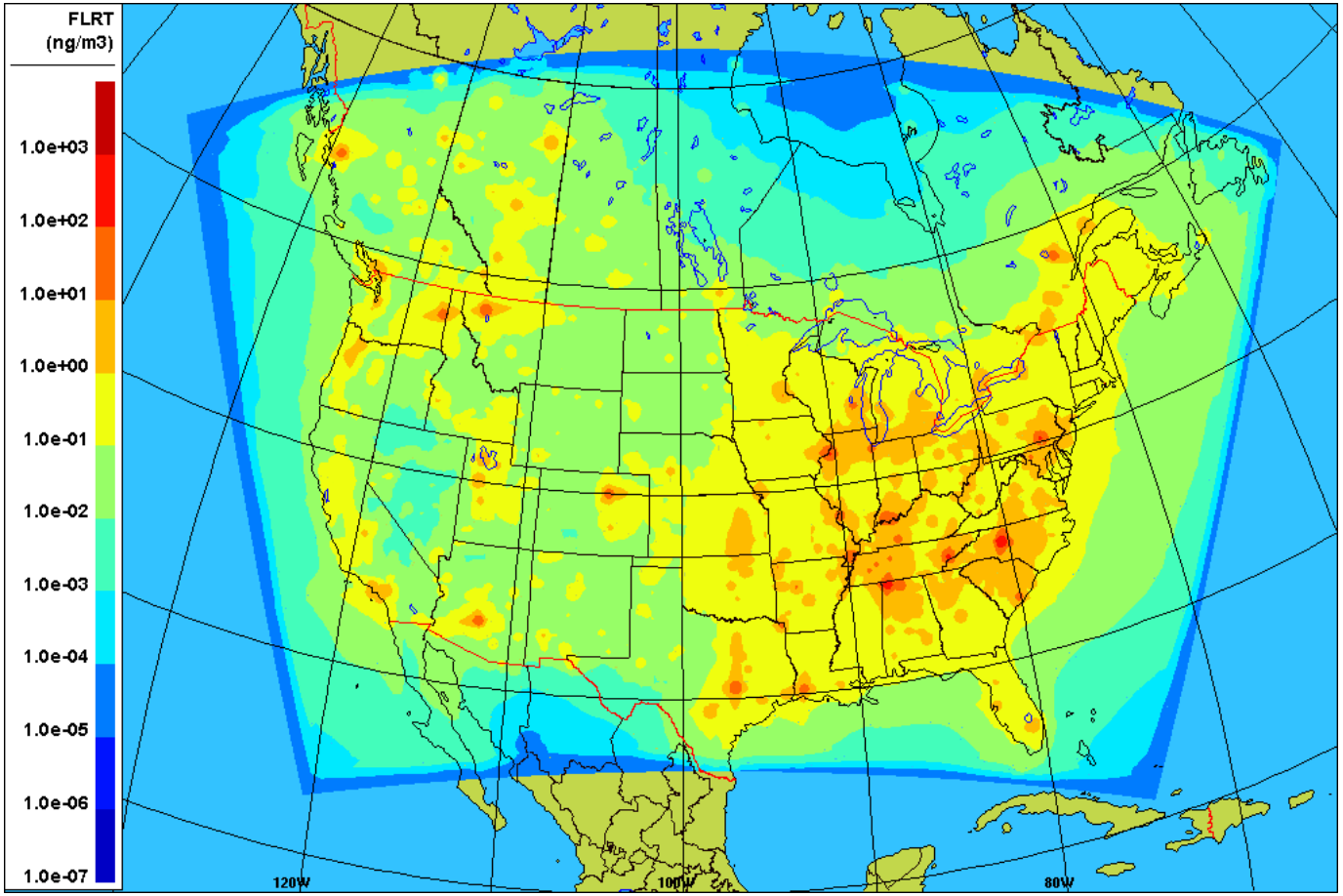


Figure 2: Map of measurement stations used in AURAMS-PAH evaluation.

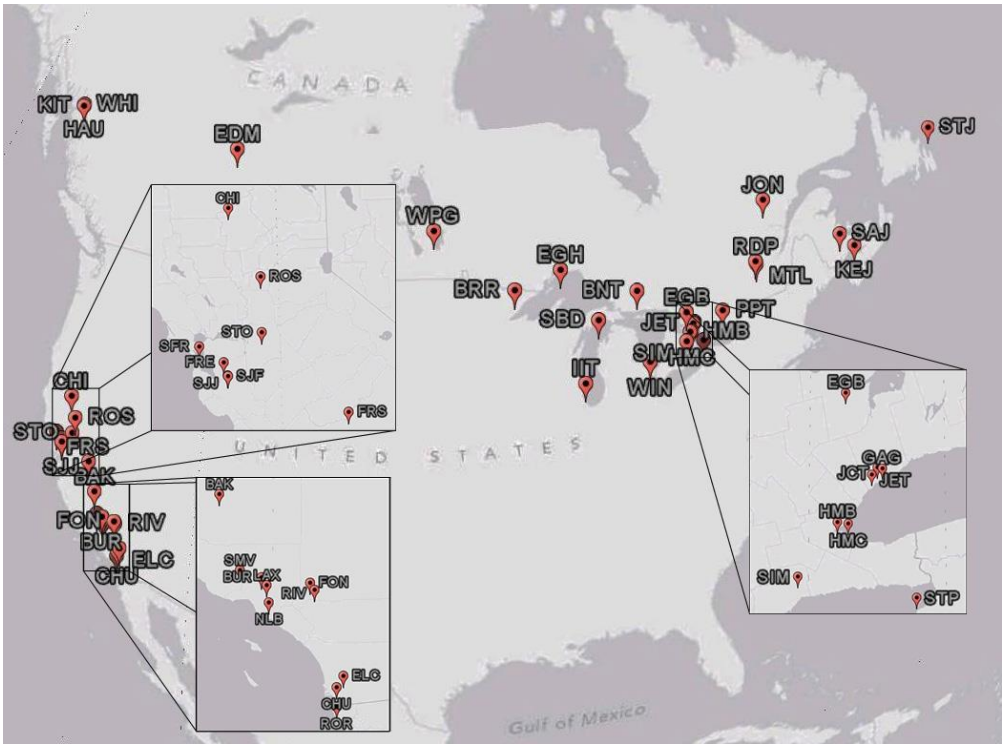
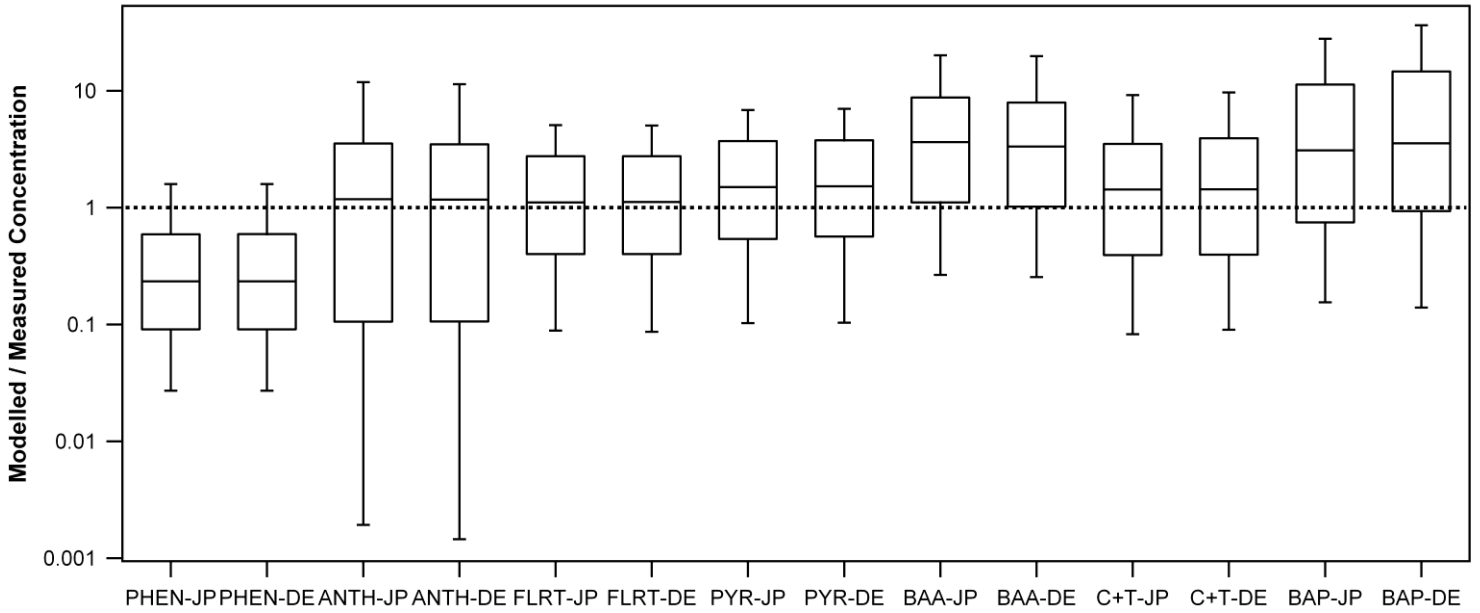


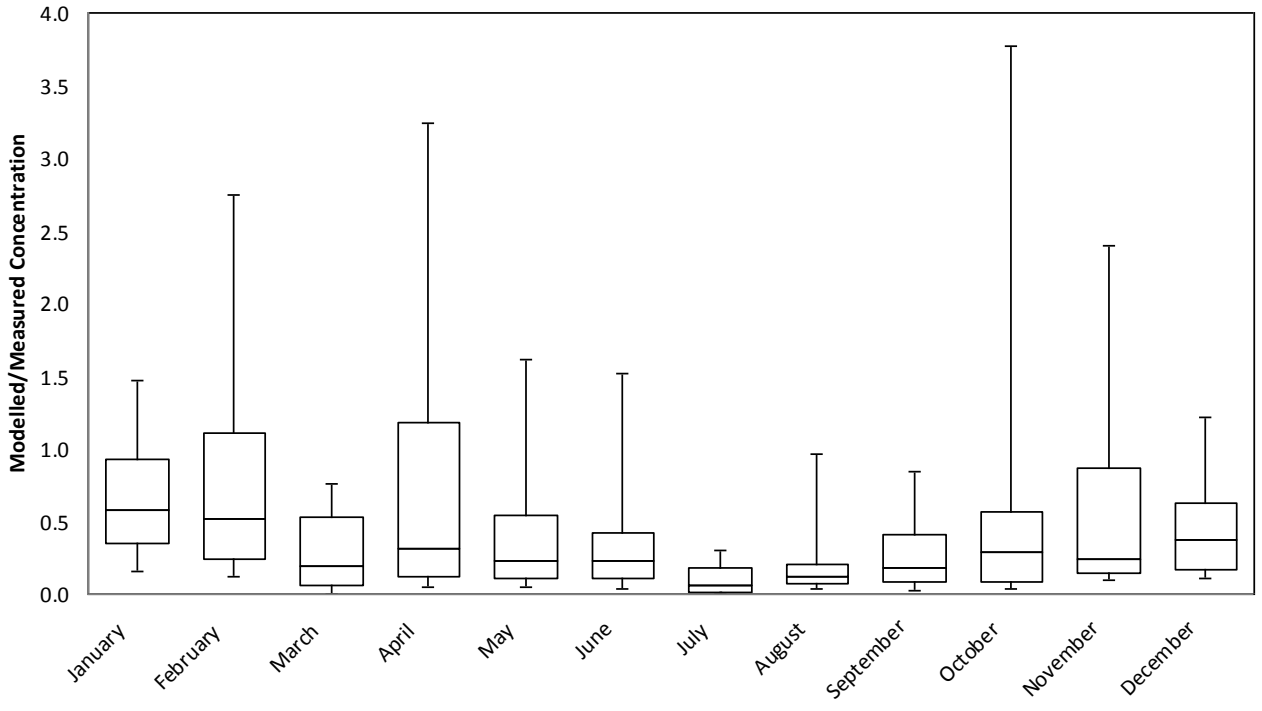
Figure 3: All-site ensemble of modelled-to-measured concentration ratios for total (gas + particle) PAHs using JP and DE partitioning expressions.



N.B. Box boundaries are 25th, 50th and 75th percentile values; whiskers are 10th and 90th percentile values. JP = Junge-Pankow partitioning; DE = Dachs-Eisenreich partitioning.

Figure 4: All-site ensemble of modelled-to-measured concentration ratios for total (gas + particle) PHEN (a) and PYR (b) using JP partitioning expression plotted by month.

(a) PHEN



(b) PYR

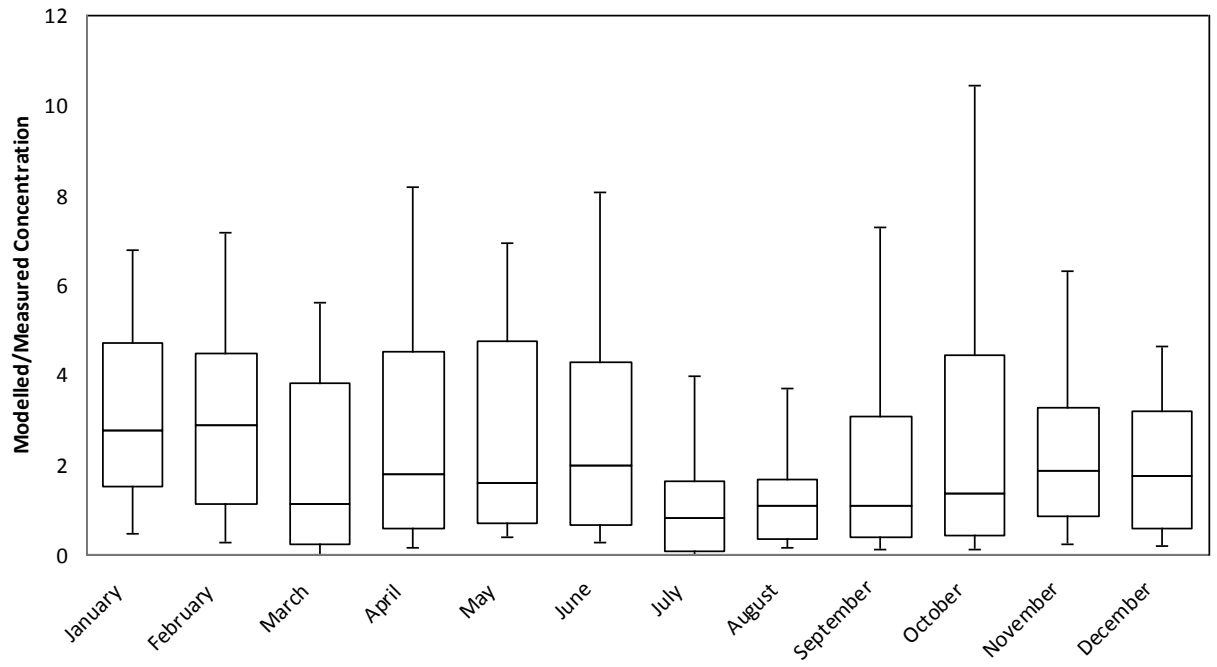
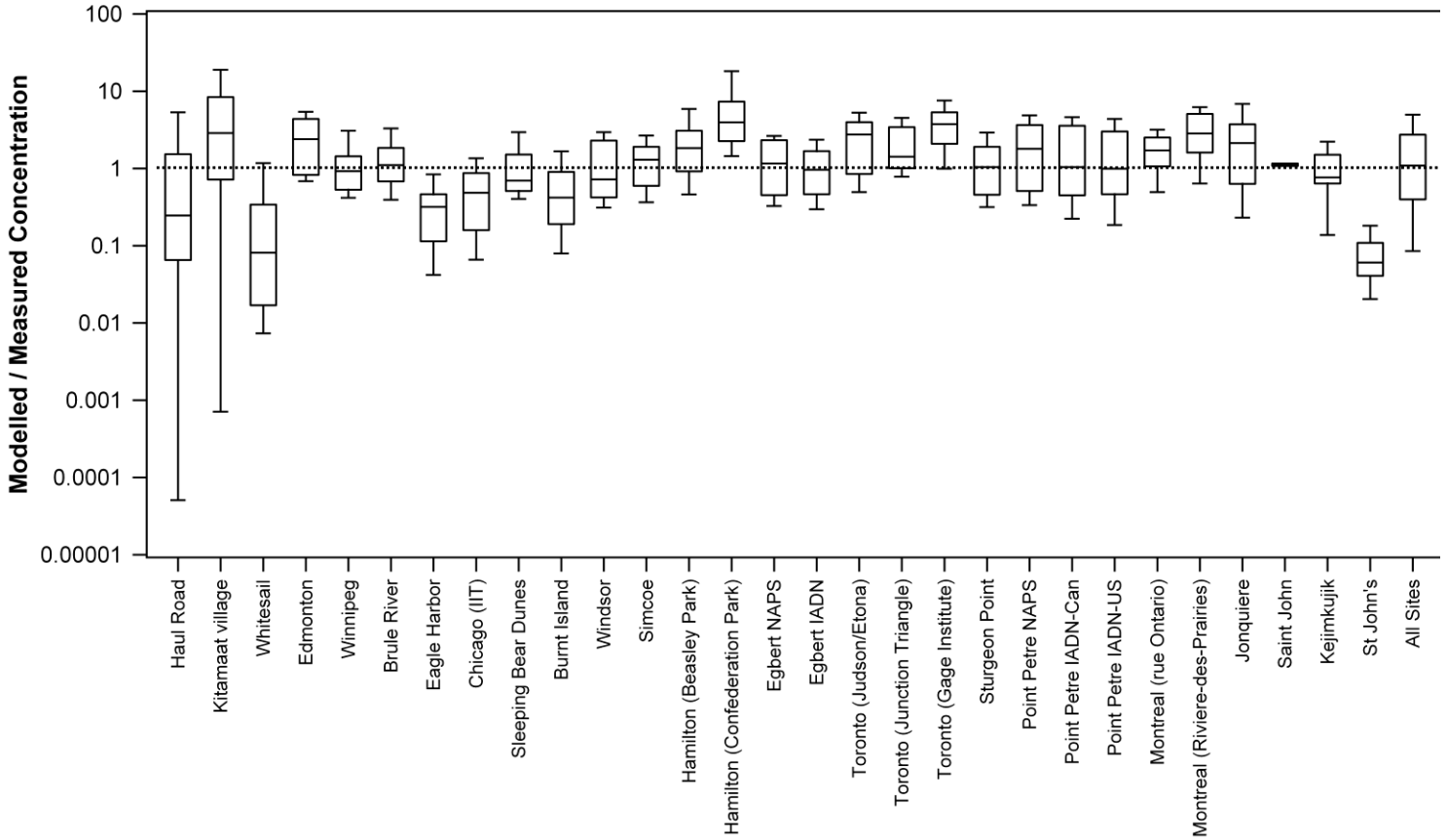
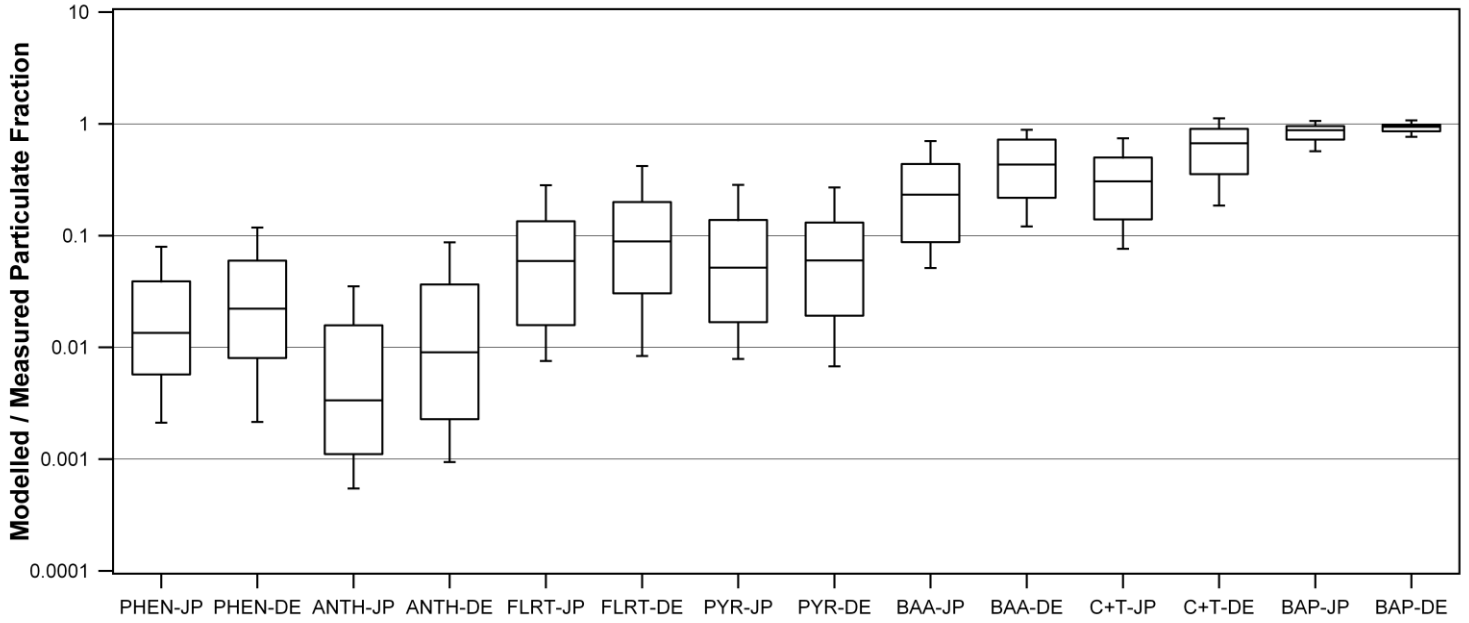

Modelled/Measured Concentration
12
10
8
6
4
2
0
January
February
March
April
May
June
July
August
September
October
November
December

Figure 45: Site-specific modelled-to-measured concentration ratios for total (gas + particle) fluoranthene for JP partitioning



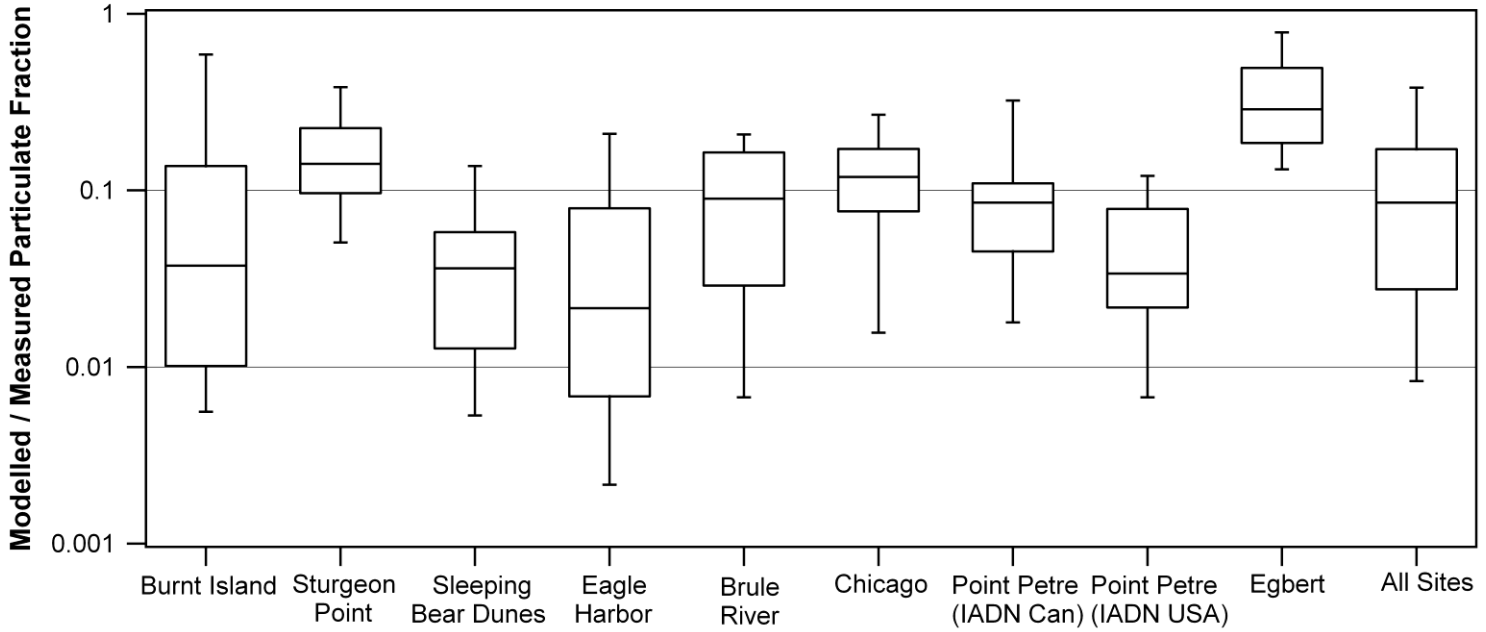
N.B. Box boundaries are 25^{th}, 50^{th} and 75^{th} percentile values; whiskers are 10^{th} and 90^{th} percentile values.

Figure 65: All-site ensemble of modelled-to-measured PAH particulate fraction ratios for JP and DE partitioning expressions.



N.B. Box boundaries are 25th, 50th and 75th percentile values; whiskers are 10th and 90th percentile values. JP = Junge-Pankow partitioning; DE = Dachs-Eisenreich partitioning.

Figure 7~~6~~: Site-specific modelled-to-measured partition coefficients for fluoranthene for DE partitioning for eight IADN sites.



N.B. Box boundaries are 25th, 50th and 75th percentile values; whiskers are 10th and 90th percentile values.

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