

## Responses to the referee's comments

### Referee 2#

#### General Comment:

This paper presents the measurements of ozone and its precursors at four stations in Beijing area, which are in line along the regional transport route of air pollutants. The air quality impacts of ozone transported from urban to the downwind rural areas have been investigated in several megacities, e.g., Mexico City. Nevertheless, field measurement data from a network in Beijing is still rather limited. Given the unique features of ozone precursors in Beijing, this paper is still helpful to improving our understandings of the transport and chemistry of ozone in a megacity.

**Authors:** Thanks a lot for the referee's comments. We have made revisions according to the referee's specific comments, which are very helpful for us to improve the manuscript.

#### Specific Comment:

(1) **The referee:** It was indicated, and actually hypothesized, that the concentration and composition of ozone precursors were different in urban and downwind rural areas. However, the results (as shown in Figure 4) show that the compositions of NMHCs were rather similar at the BL and SDZ sites in spite of the substantially lower HC levels observed at SDZ. While some differences were actually indicated in the text and Table 3, it seems very confusing on the characteristics of O<sub>3</sub> precursors at the two sites.

**Authors:** It is true that the percentage contributions of aromatics, alkanes and alkenes to the total NMHCs are rather similar at the BL and SDZ sites. As suggested by the referee, we estimated the  $L_{OH}$  and OFP of CO. As shown in Fig. 6 of the revised manuscript, the percentage contributions from CO and biogenic NMHCs to the total OFP are about twice greater at SDZ than at BL. In addition, the contributions of individual NMHC species to the total  $L_{OH}$  and OFP are quite different at the two sites (Fig. 7). Moreover, the NO<sub>x</sub> level is rather lower at SDZ than at BL (see Table 2, Fig.8 and Fig. 9). These indicate different chemical characteristics of O<sub>3</sub> precursors at the two sites.

(2) **The referee:** The mixing ratio of isoprene at BL was significantly higher than at SDZ, 13.36 vs. 2.79 ppbv. The authors suggested that the anthropogenic sources of isoprene were negligible in Beijing. In this context, the lower isoprene level observed at the rural site, SDZ, needs to be explained.

**Authors:** In this study, isoprene shows negative and poor correlation with CO, however, it shows good correlation with temperature. It means isoprene is mostly biogenic rather than anthropogenic in summer in Beijing. Deciduous trees, which are often used to decorate the urban roads, and holly coniferous trees, which are common in most of the urban parks, can release substantial amounts of isoprene and monoterpenes into urban atmosphere in Beijing, respectively (Wang et al., 2003). The SDZ site is surrounded mainly by crops and fruit trees with much less emissions of isoprene and monoterpenes. Therefore, higher biogenic VOCs were observed at BL than at SDZ. We have added the discussions above into Section 3.2 in the revised manuscript.

(3) **The referee:** CO was used as an indicator of polluted air mass in this study. In addition to that, CO (and CH<sub>4</sub>) is also an important precursor of O<sub>3</sub>, in particular in a highly polluted area like Beijing. I suggest adding CO and CH<sub>4</sub> into the calculation of OFP (Table 3) and make some

discussion upon this point.

**Authors:** Thank the referee for constructive comments. As mentioned above, we have added CO into the calculation of  $L_{OH}$  and OFP and given related discussions in the revised manuscript. We have also added the CO emission strength for each county and distinct in Beijing into Fig. 1. The importance of CO in the ozone production at the rural site is also highlighted in the “summary and conclusions”. We have no measurement of CH<sub>4</sub> at our sites. The source profile of CH<sub>4</sub> is rather different from NO<sub>x</sub> and CO, the latter originating mainly from fossil fuel usage and biomass burning. Therefore, we do not consider CH<sub>4</sub> when calculating the  $L_{OH}$  and OFP.

(4) **The referee:** The ratios of HCs with different lifetimes were used to infer the photochemical ages of air masses. Such an analysis should be made along the transport of an air parcel or for a stagnant air mass. The current analysis shows different HC ratios at BL and SDZ, respectively, but seems ignoring the differences in the emission profiles at the two locations. Will this result in any bias in the analysis of O<sub>3</sub> formation/transport? Given that the samples were collected along the transport route, it could be useful to analyze the data in the context of upwind/downwind relationship.

**Authors:** The difference in HC ratios at BL and SDZ can be attributed to both the age of air masses and the differences in the emission profiles at the two locations. It should be noted that we focus on the changes in O<sub>3</sub> as a function of the HC ratio at the same site, either BL or SDZ, as shown in Fig. 12. In this case, the effect of emission profiles can be neglected. It is a good idea to analyze the data in the context of upwind/downwind relationship. However, at present we do not have sufficient data to perform such analysis. Maybe we can continue such investigation in the next field experiment.

#### **Technical Suggestion:**

(1) **The referee:** The unit of radiation in Figure 2 could be in error.

**Authors:** We have corrected this.

(2) **The referee:** The resolution in Figure 3 is not enough. It's rather hard to see the differences among the data from respective stations.

**Authors:** The resolution in Figure 3 and some other figures has been increased.