

**Biological residues
define the ice
nucleation properties
of soil dust**

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Biological residues define the ice nucleation properties of soil dust

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Abstract

Soil dust is a major driver of ice nucleation in clouds leading to precipitation. It consists largely of mineral particles with a small fraction of organic matter constituted mainly of remains of micro-organisms that participated in degrading plant debris before their own decay. Some micro-organisms have been shown to be much better ice nuclei than the most efficient soil mineral. Yet, current aerosol schemes in global climate models do not consider a difference between soil dust and mineral dust in terms of ice nucleation activity. Here, we show that particles from the clay and silt size fraction of four different soils naturally associated with 0.7 to 11.8 % organic carbon (w/w) can have up to four orders of magnitude more ice nuclei per unit mass active in the immersion freezing mode at -12°C than montmorillonite, the most efficient pure clay mineral. Most of this activity was lost after heat treatment. Removal of biological residues reduced ice nucleation activity to, or below that of montmorillonite. Desert soils, inherently low in organic content, are a large natural source of dust in the atmosphere. In contrast, agricultural land use is concentrated on fertile soils with much larger organic matter contents than found in deserts. It is currently estimated that the contribution of agricultural soils to the global dust burden is less than 20 %. Yet, these disturbed soils can contribute ice nuclei to the atmosphere of a very different and much more potent kind than mineral dusts.

1 Introduction

Cloud droplets do not freeze until temperatures drop below -38°C , unless they harbor heterogeneous ice nuclei (IN). IN are associated with particles, typically $> 0.5\ \mu\text{m}$ in diameter (DeMott et al., 2010), suspended in the atmosphere and they markedly affect the development of clouds and precipitation. The most efficient naturally occurring IN are proteins on the surface of biological particles. They have the capacity to catalyse freezing at temperatures near -2°C (Maki et al., 1974; Wilson et al., 2006). It has

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been suggested that all IN active at temperatures warmer than -10°C are of biological origin (Christner et al., 2008; DeMott et al., 2010). Non-biological IN are typically active only at colder temperatures (DeMott et al., 2010). The study of soil samples, and more precisely of decomposing leaves, first raised the issue of biological ice nucleation (Schnell and Vali, 1972, 1972, 1976) and its association with certain bacteria (Maki et al., 1974). However, in atmospheric modelling, soil dust is considered a purely mineral IN (Lohmann and Diehl, 2006; Hoose et al., 2010a; Diehl and Wurzler, 2010). Representation of biological IN in these models is limited to bacterial cells and fungal spores (Hoose et al., 2010a; Diehl and Wurzler, 2010). These are considered to come principally from plant canopies, and to a lesser degree soils (Lindemann et al., 1982). Although bacteria and other micro-organisms rarely constitute more than 2% (w/w) of all soil organic carbon (Kaiser et al., 1992), ice nucleation activity can be retained by dead cells (under certain conditions) and certain types of cell fragments. Furthermore, IN have also been found to be associated with certain plants and fungi (Lee et al., 1995), and these might also contribute to the soil organic matter. In the particle size ranges of clay and silt, the vast majority of biological material consists of more or less decomposed biological residues (Guggenberger et al., 1994) and is part of so-called organo-mineral complexes (Chenu and Plante, 2006). Even in soils with a small organic matter content (1–2% organic carbon (w/w)) up to 50% of the particle surface area can be covered with such residues (Kahle et al., 2002). Our focus is on the extent to which these small amounts of biological residues may be responsible for the IN activity within soil dust.

2 Materials and methods

Soil samples from the top 10 cm on patches of grassland (Table 1) were first sieved dry ($< 63\ \mu\text{m}$), then wet ($< 15\ \mu\text{m}$). We call the fraction $< 15\ \mu\text{m}$ “soil dust”. Particle size analysis by laser diffraction (Mastersizer X, Malvern Instruments, Worcestershire, UK) revealed that in all soil dusts, particles smaller than $3\ \mu\text{m}$ occupy a total volume fraction

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of 0.2 to 0.3 and particles $< 10 \mu\text{m}$ occupy a total volume fraction of about two-thirds. Assuming particles are spherical and have a density of 2.6 g cm^{-3} , we estimate from the detailed particle size analysis the total number of particles $> 0.5 \mu\text{m}$ diameter in $1 \mu\text{g}$ of soil dust to be around 8×10^4 .

5 We tested ice nucleation efficiency by immersion freezing, which represents the vast majority of nucleation events (Hoose et al., 2010b). Initial tests indicated that two soils were much more active than the others. So, the suspensions of the most active soils had to be diluted with distilled water before they could be characterized. Final concentrations were 0.1 (soil D), 2 (soil C) and 10 (soils A, B) $\mu\text{g ml}^{-1}$. To better characterize
10 the warmer end of the temperature range, all soil dusts, untreated and treated, were in addition tested at 5-times higher concentrations, respectively. To characterise the lower end of the temperature scale, soil D was also tested in a lower, third concentration ($0.02 \mu\text{g ml}^{-1}$). For some temperature steps, data from more than one concentration were available. This applies to all soils. The overlap of data between different
15 concentrations was in at least one, at most seven, temperature steps. Such duplicate (triplicate) data at one temperature step were log-averaged.

From each soil suspension 54 drops, consisting of $50 \mu\text{l}$ aliquots in 0.5 ml Eppendorf safe lock tubes, were exposed to decreasing temperatures (-4 to -15°C) in a water bath cooling at a rate of 1°C in 3 min. The number of frozen tubes was inspected
20 visually after each cooling step. Numbers of ice nuclei at each temperature step were calculated according to Vali (1971). Additional measurements were made with a suspension of montmorillonite ($10 \mu\text{g ml}^{-1}$; $50 \mu\text{g ml}^{-1}$) (Montmorillonite (aluminium pillared clay), Sigma Aldrich, CAS No. 139264-88-3) and with 5 sets of blanks, each set consisting of 54 drops ($50 \mu\text{l}$) distilled water in the same type of tube as for the
25 other samples. Only in one blank set did a tube freeze at $> -15^\circ\text{C}$. So, no background was subtracted from the results with soil suspensions. Numbers of frozen samples at a particular temperature were only considered when > 1 . To test the contribution of heat-sensitive components of organic matter to the nucleation activity, samples were then exposed for 10 min to temperatures $\sim 100^\circ\text{C}$ in a water bath and tested again

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(heat treated samples). Freezing nucleation properties of the mineral phase were determined for samples where the organic carbon had been removed by oxidation to CO₂ via exposure to H₂O₂ (35 %). Because these samples showed much lower activity in initial tests, they were all tested at concentrations of 10 and 50 µg ml⁻¹.

3 Results and discussion

To assess the contribution of biological residues to the IN efficiency of soil dust, we conducted immersion freezing tests with particles in the clay and finer silt fractions (< 15 µm) from soils in Western Mongolia (A), Germany (B), Hungary (C) and Eastern Siberia (D), harbouring 0.7, 5.1, 11.8 and 5.4 % organic carbon (w/w), respectively (Table 1). The three soil dusts with > 5 % organic carbon content were much more efficient IN than montmorillonite, currently considered to be the most efficient purely mineral IN (Fig. 1). At -12 °C, for example, the number of IN per µg of dust was 0.2 for montmorillonite and soil dust A, but 3, 24, and 1910 for soil dusts B, C, and D, respectively. At -8 °C, the number of IN per µg of montmorillonite was below the detection threshold, but was 0.03, 0.6, 5 and 206 for soil dusts A, B, C and D, respectively. At this temperature we assume that the active IN are of biological origin. Based on data from the literature for the total soil organic carbon content (1.5 × 10¹⁸ g) and the global bacterial population of soil (2.6 × 10²⁹ bacterial cells; Whitman et al., 1998) we estimated if intact bacteria could account for some of these IN. We assume that intact bacterial cell number density is proportional to organic carbon content, so there would be 1.7 × 10⁵ intact cells associated with 1 µg organic carbon in soil. For this calculation we also take into account that the known species of ice nucleation active bacteria, such as *Pseudomonas syringae* or *P. borealis* (Wilson et al., 2006), constitute a very small fraction of soil bacterial communities (< 0.1%). Furthermore, at -8 °C, the fraction of intact cells that produce ice nuclei does not exceed 0.01 for most strains. Given these assumptions, intact bacteria would account for 0.01, 0.1, 0.2 and 0.1 IN µg⁻¹ of soil dust A, B, C and D, respectively. Hence, the contribution of intact bacteria to the IN associated with these soils would seem to be negligible.

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Although dust D has just as much organic carbon as dust B, the number of IN per unit mass of this dust was substantially larger. So, probably the nature of the organic matter and processes contributing to its preservation play a role in this difference, not just the quantity. Soil D is from an environment where mean annual temperature is -9.5°C , whereas soil B experiences a mean annual temperature of $+9.8^{\circ}\text{C}$. Numbers of IN in leaf litter from cold environments are known to be larger, and onset of freezing is earlier, than in leaf litter from warm regions (Schnell and Vali, 1973). IN active sites in plants, bacteria and other micro-organisms are proteins (Wolber et al., 1986; Morris et al., 2004) and proteins are well preserved when sorbed to mineral surfaces (Kleber et al., 2007). Properties of leaf-derived and microbial IN in the respective climate zones may be transferred to organic coatings of mineral soil particles in these zones. This could explain the much larger number of IN per unit mass in soil D compared to soil C, although D has less organic carbon content than C in the analysed fine fraction (Table 1). Nevertheless, within a particular region, the number of IN may well correlate with the amount of biological residues associated with mineral dust. Further samples from Western Mongolia, but with different organic carbon contents than soil A, support such a presumption (Fig. 2). Within this region, the number of IN active at -8 and -12°C seems to increase by roughly one order of magnitude with each percent increase in organic carbon (w/w).

By heat treating (Christner et al., 2008) the particles of the soil dusts, we revealed that the agent responsible for the large number of IN active above -10°C , and in some cases even colder, was most likely proteinaceous rather than a mineral. Heating destroys the configuration, and thereby the function, of many biological molecules and in particular proteins which constitute the most active biological IN known. As expected, the ice nucleation activity of the purely mineral sample, montmorillonite, was little affected by heating. Soil dust A, which is associated with only 0.7% organic carbon (w/w) behaved similarly to montmorillonite. However, soil dusts B, C, and D with $> 5\%$ organic carbon (w/w) lost 70, 91 and 98% of their initial activity at -12°C after heating. Activity at -8°C was no longer detected after heating in soil dusts A, C and D. Soil

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dust B lost 82 % of its initial activity at -8°C by heating. When samples were treated with H_2O_2 to strip them of biological residues via oxidation, IN activity was further reduced either close to, or below, that of montmorillonite, which itself was little affected by the treatment. Treatment with H_2O_2 is a common procedure to oxidize organic matter from soil samples for subsequent analysis of the remaining minerals. These results illustrate that the IN activity of soils containing a certain level of biological residues is far greater than that of the basic mineral constituents composing the soil dust. They also suggest a direct role of the biological residues associated with mineral particles in this IN activity. A loss of IN activity in the mineral phase due to the treatments of the samples seems unlikely, although reactions of the mineral complexes in soil dusts that are not pure phases can not be ruled out. Perhaps more important than the increase in the number of IN per unit mass at any particular temperature is the upward shift in the onset of freezing caused by the biological residues. Observations over the Amazon seem to suggest that, despite small number concentrations, biological particles dominate the ice-nucleus population at warmer temperatures (Prenni et al., 2009). In our experiments it reaches into the range of temperatures where a process of multiplication of ice by shattering of crystals can be active in clouds (-8°C or warmer) (Hallett and Mossop, 1974).

4 Conclusions

While we can confidently state that small amounts of biological residues can greatly enhance the number of IN from soils active at a certain temperature and shift onset of freezing to warmer temperatures, there are still a number of questions to be answered before the atmospheric implications of these findings may be quantified. A critical point is to know how the IN are distributed among single particles in the atmosphere. From our particle size analysis we estimate that $1\ \mu\text{g}$ of analysed soil dust contains on average 8×10^4 particles $> 0.5\ \mu\text{m}$ in diameter. So, even in the most active soil D, no more than 1 in about 400 particles $> 0.5\ \mu\text{m}$ is likely to carry a site that is IN active

at -8°C and less than 1 in 40 carries a site active at -12°C . Particle size distribution of natural soil dust lifted to cloud height will not necessarily be the same as in our samples, where the largest particles were up to $15\ \mu\text{m}$ in diameter. From soil science we know that biological residue concentrations (w/w) in soil increase with decreasing particle size (Kahle et al., 2002). So, for the same mass, a collection of small particles will harbour a larger amount of biological residues than a collection of larger particles. The smaller a particle, the greater the likelihood it is transported to cloud heights. Consequently, air movements preferentially transport small soil particles, which carry above average loads of biological residues (Hoffmann et al., 2008), so probably also a greater number of IN per unit mass than is found in the bulk soil.

Recent in situ detection of cloud ice crystal residues at high altitudes revealed that 60% of the dust particles contained biological material, which may have led to increased ice-nucleation efficiency (Pratt et al., 2009). Significant effects on stratiform mixed-phase clouds have been observed in simulations when parameters for kaolinite (a clay that is less IN active than montmorillonite) are used (Lohmann and Diehl, 2006). The same simulation comparing montmorillonite with biologically IN-activated soil particles would be a worthwhile exercise to re-evaluate the anthropogenic indirect aerosol effect caused by landuse or landuse change. Over-grazing of natural steppes in Inner Mongolia leads to wind erosion releasing soil particles to the atmosphere with 3–4% organic carbon (w/w) (Hoffmann et al., 2008). Tillage operations on arable soils can multiply the amount of dust emissions caused by wind erosion in temperate regions (Goossens et al., 2001), and high anthropogenically induced wind erosion rates often affect soils rich in organic matter (Moon, 2005; Galic et al., 2009). The effect of IN associated with these dust emissions on cloud processes might be considerably larger than expected from their currently-estimated contribution of less than 20% to the global dust burden (Forster et al., 2007).

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Table 1. Origin of samples (MAT = mean annual temperature) and their carbon content (in % w/w).

Sample	Region	Coordinates	Altitude	MAT (°C)	Organic C content (%)	
					Bulk soil	< 15 µm
A	Western Mongolia	46° 19' N; 95° 04' E	976 m	−0.1	0.2	0.7
B	Southern Germany	47° 38' N; 07° 40' E	300 m	9.8	2.9	5.1
C	Central Hungary	46° 58' N; 19° 33' E	125 m	10.4	2.7	11.8
D	Central Yakutia	62° 28' N; 126° 01' N	240 m	−9.5	3.3	5.4

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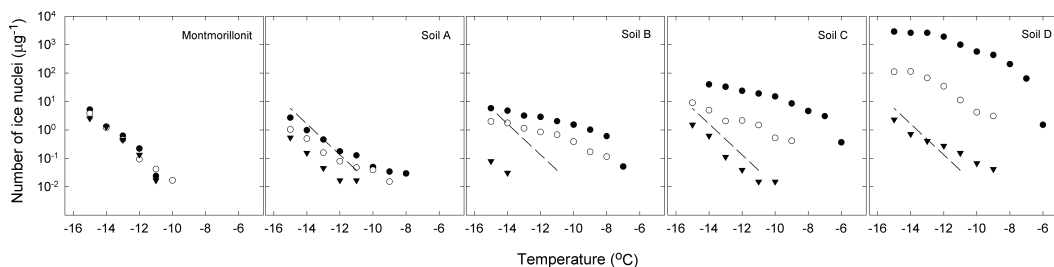


Fig. 1. Freezing nucleation properties of montmorillonite and dust particles ($< 15 \mu\text{m}$) from different soils (A, B, C, D) immersed in distilled H_2O . Initial numbers of ice nuclei (\bullet), numbers of ice nuclei still active after heating (\circ), and numbers of ice nuclei remaining after removal of organic matter (\blacktriangledown) are shown. Dashed lines indicate the curve for untreated montmorillonite, for comparison.

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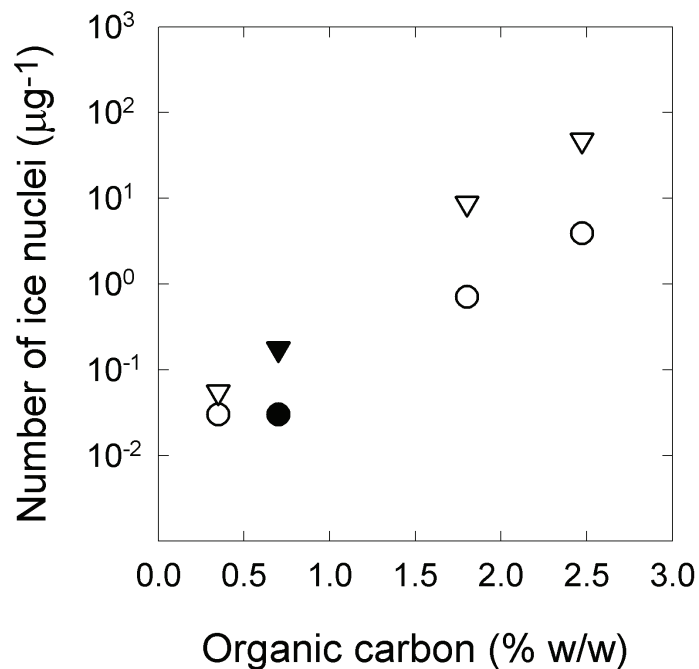


Fig. 2. Relationship between the number of ice nuclei and soil organic carbon (w/w) in soil dusts ($< 15 \mu\text{m}$) from Western Mongolia. Closed symbols are for data from soil A in Fig. 1. Open symbols are for soil dusts collected between 9 and 313 km from soil A. Circles and triangles show the number of ice nuclei active at -8°C and -12°C , respectively.

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