



Characteristics and processing of aqueous secondary organic aerosols during autumn in suburban Eastern China: role of aerosol liquid water, aerosol acidity, and photochemistry

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Abstract. Aqueous-phase secondary organic aerosols (aqSOA) constitute a large fraction of SOA, thereby exerting significant influence on air quality, climate, and human health. However, its formation pathways under real ambient conditions in Chinese urban regions remain insufficiently constrained. We conducted field measurements of particulate matters (PM) composition by deploying high-resolution aerosol mass spectrometry in a suburban environment during autumn in Nanjing, China. The characteristics and formation pathways of aqSOA are comprehensively investigated by using Positive Matrix Factorization (PMF) method. Our results show that aqSOA accounted for 27.6 % of oxidized organic aerosols, exhibiting elevated O : C ratios (0.78) and strong correlations with nitrate and aerosol liquid water (ALW). The important role of acid-catalyzed reactions is also revealed by the enhanced production of aqSOA at lower aerosol pH conditions. Under elevated nitrate and ALW levels, a pronounced morning aqSOA peak was frequently observed; whereas a noon-time aqSOA peak was also observed on several days, likely governed by photochemistry and aqueous-phase reactions. These findings highlight the critical roles of nitrate, ALW, acidity, and photochemistry in driving aqSOA production in polluted urban environments. This study advances the mechanistic understanding of aqSOA formation and provides insights into the mitigation of SOA in the Eastern China.

1 Introduction

Organic aerosols (OA) significantly affect air quality, human health, and climate by influencing radiative forcing and cloud formation (Kanakidou et al., 2005; Zhou et al., 2019). Secondary organic aerosols (SOA) can contribute 30%–70% of total organic aerosols (Huang et al., 2014; Xian et al., 2023). The formation of SOA has been attributed mainly to gas-phase oxidation of volatile organic compounds (VOCs), where the oxidized products subsequently partition into the aerosol phase (Hennigan et al., 2009; Seinfeld and Pankow, 2003; Ziemann and Atkinson, 2012). Recently, growing evidence highlights the important role of SOA generated through aqueous-phase processes in cloud droplets, fog, and aerosol liquid water (ALW) (Ervens et al., 2011; Kim et al., 2019; Sun et al., 2010). However, the formation mechanisms and sources of aqueous-phase SOA (aqSOA) in Chinese urban regions remain highly uncertain (Ervens et al., 2011; Huang et al., 2025; McNeill, 2015).

In recent years, concentrations of OA have gradually declined across many regions in China due to the implementation of air pollution control measures, but the relative contribution of SOA has increased markedly (Chen et al., 2024; Huang et al., 2025), with aqSOA constituting a substantial fraction. Numerous studies have demonstrated that elevated ALW can significantly promote aqSOA formation during winter haze episodes through multiphase reactions (Chen et al., 2021; Feng et al., 2022; Liu et al., 2019; Peng et al., 2021; Sun et al., 2016, 2019; Wang et al., 2023; Xiao et al., 2022; Xu et al., 2017; Zhao et al., 2019). Still, the exact processes driving aqSOA formation is not fully characterized. Field observations in urban Beijing have shown that ring-breaking oxidation and functionalization of polycyclic aromatic hydrocarbons of fossil-fuel-derived primary organic aerosols could lead to rapid aqSOA formation at high relative humidity (RH) during winter haze episodes (Wang et al., 2021). Another field study conducted in the North China Plain showed that the formation of aqSOA could be largely enhanced under favorable photochemical conditions with precursors originated from biomass burning activities (Kuang et al., 2020). Recently, field measurements in Hebei reported that high nitrate may support the potential formation/transformation from POA-related components to aqSOA (Gu et al., 2023). Laboratory studies reveal that accretion reactions, which play a crucial role in SOA formation, are highly sensitive to pH levels (Tilgner et al., 2021). Moreover, under the emerging dominance of nitrate in aerosol composition (Huang et al., 2025), the interplay among nitrate, ALW, and pH may complex the aqSOA formation and requires further investigation.

The Yangtze River Delta (YRD) region is one of the most densely populated and economically developed areas in China, characterized by intensive industrial activity, heavy traffic emissions, frequent regional pollution episodes, and high relative humidity (Liu et al., 2025). Several previous

studies have examined aqSOA processes in this region (Wang et al., 2016; Wu et al., 2018; Xian et al., 2023), but uncertainties regarding its sources, controlling factors, and formation mechanisms still remained. Consequently, it is challenging for current models to precisely simulate aqSOA and its contribution to the YRD region (Ervens et al., 2011; Rogers et al., 2025).

In this study, we conducted real-time measurements of OA at the National Observation and Research Station for Atmospheric Processes and Environmental Change in Yangtze River Delta (SORPES) located in suburban Nanjing in the western YRD region during the autumn of 2020. The chemical characteristics and formation mechanisms of aqSOA, as well as the roles of nitrate, ALW, and aerosol acidity are investigated. By classifying diurnal variation patterns of aqSOA, we assessed the relative roles of photochemical and aqueous-phase processes. The results provide new insights into regional aqSOA formation in the YRD and have implications for the development of effective air pollution control strategies.

2 Materials and Methods

2.1 Sampling Site

The field campaign was conducted from 13 October to 30 December in 2020 at SORPES station (118°57'E, 32°07'N) located in the Xianlin campus of Nanjing University in Nanjing, China. This is a representative station of western YRD, surrounded by high vegetation cover, and also subject to abundant anthropogenic emissions (Ding et al., 2016, 2019; Dou et al., 2025; Liu et al., 2025).

2.2 Instrumentation

Real-time non-refractory PM₁ composition was measured using a high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS; hereafter, AMS; Aerodyne Research Inc.). An aerodynamic PM₁ lens was used to focus the particle into a beam, which was then impacted on the heated tungsten surface (~600 °C) and flash-vaporized. In our study, ambient aerosols were passed through a ~2 m long stainless-steel sampling tube, dried by a Nafion drying tube, and then introduced into the AMS. In order to obtain highly sensitive data, AMS was operated in V mode with a time resolution of 2 min (DeCarlo et al., 2006).

Other instruments were also employed at SORPES in support of these measurements. Black carbon (BC) was measured by the photoacoustic extinctions (PAX, Droplet Measurement Technologies Inc., USA). The meteorological parameters and gaseous pollutants were also measured simultaneously. Ozone (O₃), carbon monoxide (CO), nitric oxide (NO), nitrogen oxides (NO_x) and sulfur dioxide (SO₂) were measured using online analyzers (Thermo Fisher Scientific, USA). Ammonia (NH₃) was measured by the Picarro

G2103 gas analyzer (Picarro Inc., USA) (Liu et al., 2024). Temperature, RH and other meteorological parameters were monitored by meteorological sensors (GRWS100, Campbell, USA).

2.3 Data Analysis

The AMS data were processed by SQUIRREL (version 1.60P) and PIKA (version 1.20P) from the ToF-AMS Software Downloads Web page (<http://cires.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/index.html>, last access: 11 May 2024). The ionization efficiency (IE) was calibrated using 300 nm pure ammonium nitrate before and after the campaign. The relative ionization efficiency (RIE) of ammonium was determined from pure ammonium nitrate, yielding a value of 3.52. RIE values for OA, nitrate, sulfate, and chloride were set to their default values of 1.4, 1.1, 1.2, and 1.3, respectively (Canagaratna et al., 2007). As well, the collection efficiency (CE) was assigned a typical value of 0.5 for common environments, although it is recognized that this constant value may introduce additional uncertainty during periods with high nitrate content. Element ratios, including H : C, O : C, N : C and OM : OC, are calculated using the Improved-Ambient method (Canagaratna et al., 2015).

In addition, ALW content and aerosol acidity that are associated with inorganic species were estimated by the Extended Aerosol Inorganics Model (E-AIM), which is a well-known inorganic thermodynamic model without simplifying assumptions (Clegg et al., 1998; Wexler and Clegg, 2002; Pye et al., 2020). The key inputs to the E-AIM model include temperature, relative humidity, and the concentrations of inorganic ions (sulfate, nitrate, ammonium, chloride) as well as gas-phase ammonia. The main sources of uncertainty stem from thermodynamic parameters and the uncertainties of measured inorganic ions. In this study, the ALW associated with the hygroscopicity of organics was estimated following the method of Nguyen et al. (2016) using the κ -Kohler theory and the Zdanovskii–Stokes–Robinson mixing rule. The estimated organic-associated ALW was only 10.7 % of that associated with the inorganic aerosols. Therefore, the organic-associated ALW was neglected in the subsequent analysis.

2.4 Source Apportionment of OA

Positive Matrix Factorization (PMF) analysis was applied to the high-resolution mass spectra of organic matrix for m/z 12–120 to resolve distinct OA factors from specific sources (Paatero and Tapper, 1994; Ulbrich et al., 2009). The data and error matrices were treated according to the procedures detailed in DeCarlo et al. (2010). By comparing the mass spectral profiles with previous studies and correlations with time series of tracers, five OA factors with $f_{\text{peak}} = 0$ were selected, including one primary organic aerosol (POA) factor, one nitrogenous OA (NOA) factor, and three SOA

factors, namely, less-oxidized oxygenated OA (LO-OOA), more-oxidized OOA (MO-OOA), and aqSOA. The detailed diagnostic plots are shown in Figs. S1 and S2 in the Supplement.

AqSOA exhibited strong correlations with unique fragment ions that are widely recognized as markers of aqueous-phase secondary products (Sun et al., 2016; Xu et al., 2019). For instance, significant correlations were observed with C_2O_2^+ (m/z 56, $r = 0.77$), a typical fragment of oxalate-related species, and with CH_3SO^+ (m/z 63, $r = 0.90$), which is indicative of organosulfur compounds (Fig. S3). These results provide chemical evidence supporting the aqueous-phase origin of aqSOA in Nanjing. Among these five OA factors, aqSOA exhibits typical characteristics of highly oxidized organic aerosols: fraction of m/z 44 (CO_2^+ , primarily from carboxylic acids and highly oxidized compounds) (Heald et al., 2010) in total signals exceeded that of m/z 43 (typically representing less oxidized compounds) (Fig. S2). Elemental analysis further indicated a strongly oxygenated character, with an average H : C ratio of 1.80 and an O : C ratio of 0.78. These values are comparable to those reported in other regions, such as northern Italy and Beijing (Gilardoni et al., 2016; Xu et al., 2019). The N : C ratio (0.05) was also elevated and similar to wintertime values observed in Beijing (N : C = 0.045; Xu et al., 2017), suggesting nitrogen-containing compound formation via aqueous-phase processing.

3 Results and Discussions

3.1 General Characteristics of aqSOA

The meteorological conditions in Nanjing were overall stable during the field campaign (October–December), with an average temperature and RH of 14.6 ± 4.6 °C and 65.1 ± 17.9 %, respectively (Fig. S6). The diurnal variation of RH ranged from 50 % to 85 %, resulting in relatively humid air conditions that favored aqueous-phase chemical reactions. The average wind speed of the prevailing northerly wind was relatively low (0.21 ± 0.16 m s⁻¹), resulting in the accumulation of local pollutants. In particular, the average NR-PM₁ concentration was 37.3 ± 20.6 μg m⁻³, indicating frequently occurred particulate matter pollution during this period.

Analysis of aerosol chemical composition revealed that organic aerosol was the dominant NR-PM₁ component in Nanjing, accounting for 40.9 % of the total PM₁ concentration, evidently higher than nitrate (30.7 %) and sulfate (13.9 %; Fig. S6). The time series of aqSOA showed significant variation, with peak concentrations up to 15.7 μg m⁻³ (Fig. 1a). The average concentration of aqSOA during the campaign was 3.1 μg m⁻³, accounting for 20.2 % (62.7 % in maximum) of the total OA and 27.6 % (78.2 % in maximum) of the total SOA (Fig. 1b). The average fraction of aqSOA in OA was much higher than Beijing (13 %–17 %) (Zhao et al.,

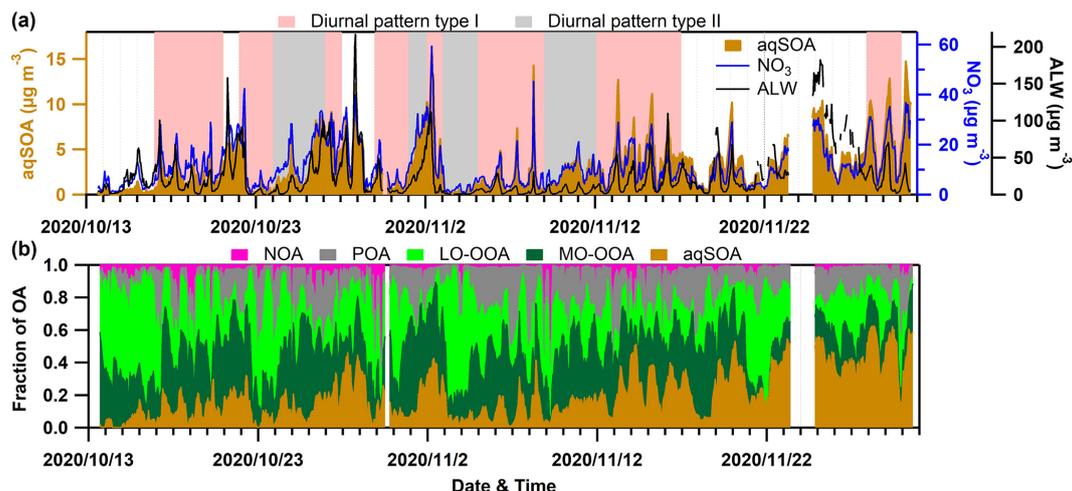


Figure 1. (a) Time series of aqSOA and nitrate concentrations. The pink area represents diurnal pattern type I, while the grey area represents diurnal pattern type II. (b) Time series of the fraction of five OA factors.

2019), indicating an important role of aqueous-phase processes in SOA formation in Nanjing. Moreover, both aqSOA concentrations and their relative contribution to total organic aerosol increased significantly with increasing RH (Fig. S7a), which was likely due to the increased availability of the aqueous reaction medium enhanced by water uptake by aerosols, which is crucial for the aqSOA formation. This is further confirmed with the sharply increased contribution of aqSOA to OA (up to 62.7 %) at RH levels above 80 %, suggesting that high-humidity environments in suburban Nanjing substantially promoted aqSOA formation through enhanced aqueous-phase chemistry.

The Van Krevelen diagram (Heald et al., 2010) provided further insights into the chemical aging of OA (Fig. S8). The H : C vs. O : C slope was near -0.5 , indicating OA oxidation primarily involving carboxylic acid and peroxide or alcohol functional groups addition without fragmentation and/or the addition of carboxylic acid functional groups with fragmentation (Ng et al., 2011). In particular, OA observed under high RH conditions ($> 80 %$) clustered in the upper plot region, suggesting distinct chemical evolution of OA when aqueous-phase reactions were involved. Specifically, the nearly zero slope of the relationship among POA, LO-OOA and aqSOA suggests that the observed increase in the O : C ratio of aqSOA may be related to oligomerization and hydroxyl formation through dark chemistry processes (Lim et al., 2010).

3.2 Enhanced aqSOA Formation Driven by Nitrate, ALW, and Acid Catalysis

As a strongly hygroscopic component, nitrate aerosol can enhance aerosol water uptake, thereby modifying the aqueous microenvironment for SOA production (Hodas et al., 2014; Sullivan et al., 2016). We observed strong correlation be-

tween aqSOA and nitrate concentrations ($R^2 = 0.83$; Fig. 2a) during the observation period, and the variation of aqSOA was highly consistent with that of nitrate aerosol (Fig. 1a). This suggests that nitrate aerosols may contribute importantly to aqSOA formation in Nanjing, potentially mediated by their influence on ALW and/or aqueous reactions. We estimated that the average contribution of organic nitrates to the total nitrate was approximately 9.5 % by assuming a NO^+ to NO_2^+ ratio of 10 for organic nitrates (Farmer et al., 2010). Thus, organic nitrates can be considered negligible in this study.

The correlation between aqSOA and aerosol liquid water (ALW; $r = 0.63$; Fig. 2b), which is mainly derived from the hygroscopic growth of inorganic salts such as nitrate, was moderate compared to nitrate. Still, their positive relationship indicates that higher ALW levels may enhance aqueous-phase chemical processes by providing the medium in which multiphase reactions can occur (Hodas et al., 2014). The presence of ALW promotes the partitioning of water-soluble organic precursors and allows for subsequent aqueous-phase reactions contributing to aqSOA formation. The results are consistent with previous studies in humid urban environments (Chen et al., 2021; Duan et al., 2022; Kuang et al., 2020).

In addition to ALW, aerosol acidity also exerted a strong influence on aqSOA yields during the campaign (Lim et al., 2010; Tilgner et al., 2021). To illustrate, the dataset was separated by aerosol pH ($\text{pH} < 3$ vs. $\text{pH} > 3$) (Fig. 2b), the slopes of ALW-aqSOA correlations decreased with increasing pH, indicating that aqSOA production was higher under more acidic conditions at the same ALW level. This pattern demonstrates the importance of acid-catalyzed reactions in driving aqSOA formation. Previous studies have demonstrated that under acidic conditions (low pH, high H^+ concentration), non-oxidative aqueous organic chemi-

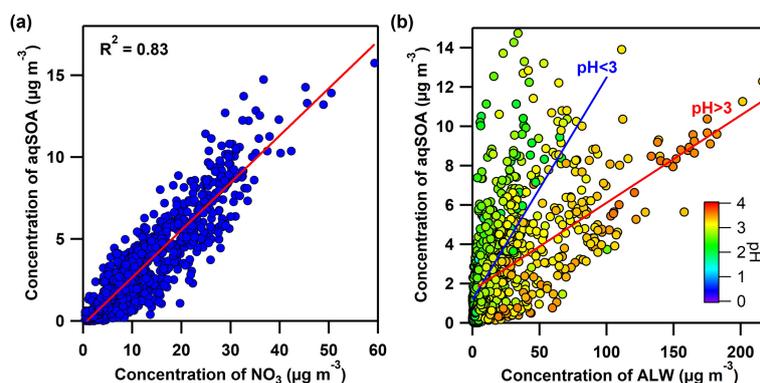


Figure 2. (a) Scatter plot of aqSOA and nitrate concentrations. (b) Scatter plot of aqSOA and ALW concentrations, colored by aerosol pH. The blue line represents the fitted line for data with aerosol pH < 3, while the red line represents the fitted line for data with aerosol pH > 3.

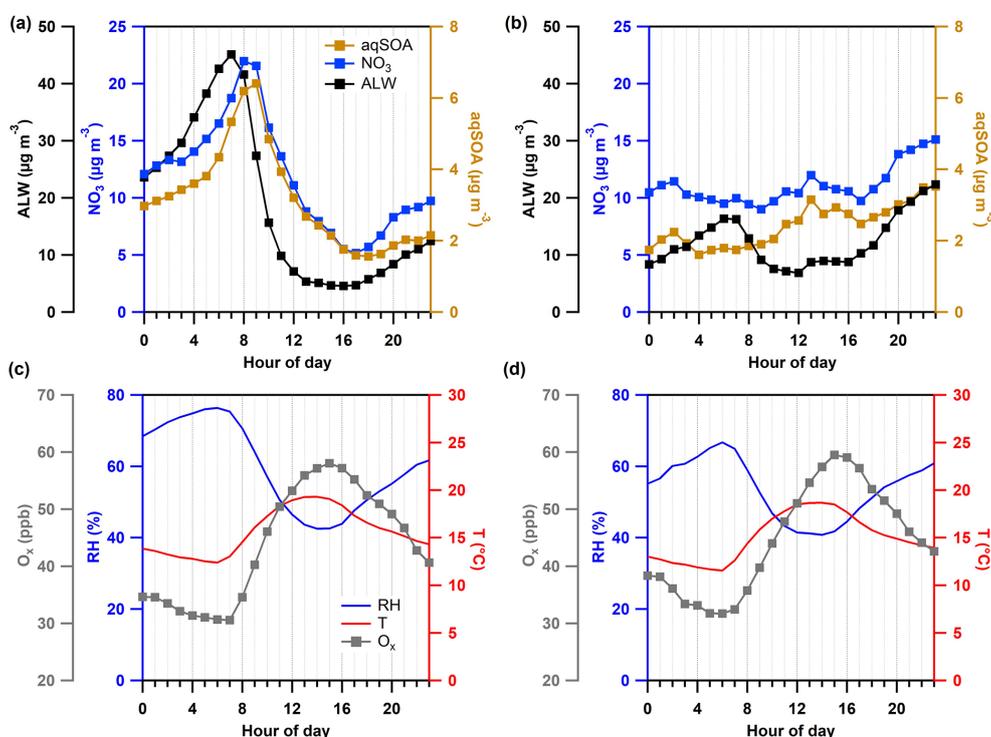


Figure 3. Diurnal variations for SOA, ALW, and nitrate in diurnal pattern types I (a) and II (b). Diurnal variations for temperature, humidity, and O_x concentration in diurnal pattern types I (c) and II (d).

cal processes, such as accretion reactions (aldol condensation, hemiacetal and acetal formation, and the esterification of carboxylic acids) are important formation pathways of aqSOA (Freedman et al., 2019; Tilgner et al., 2021). For example, the hydration of methylglyoxal and its subsequent acetal formation are highly pH-dependent, requiring a pH of less than 3.5 to occur (Yasmeen et al., 2010). As aerosol pH increases (lower H^+ availability), the catalytic efficiency of these acid-driven reactions diminishes, leading to a reduction in aqSOA production efficiency. Collectively, these findings suggest that aerosol acidity plays a pivotal regulatory role in

aqSOA formation, linking aqueous chemistry to the broader context of aerosol physicochemical properties in Nanjing.

Thus, our measurements revealed a synergistic interplay among nitrate, ALW, and aerosol acidity in regulating aqSOA formation in Nanjing. Nitrate enhances the ALW content, which in turn promotes aqueous-phase reactions, while aerosol acidity governs the efficiency of these chemical processes. These processes together constituted the formation mechanism of aqSOA during the observation period.

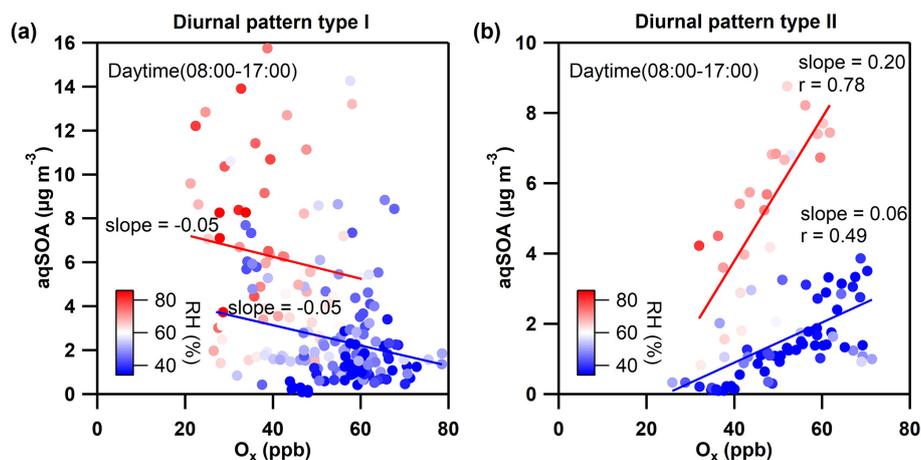


Figure 4. Scatter plots of aqSOA and O_x for Diurnal pattern types I (a) and II (b). Data points are colored by RH, with red and blue lines representing fits for $\text{RH} > 60\%$ and $\text{RH} < 60\%$ respectively. For diurnal pattern type I, the correlation coefficient was -0.22 ($N = 151$, $p < 0.01$) during daytime with $\text{RH} < 60\%$, and $r = -0.14$ ($N = 59$, $p = 0.3$) when $\text{RH} > 60\%$. For diurnal pattern type II, the correlation coefficient was 0.49 ($N = 67$, $p < 0.01$) under $\text{RH} < 60\%$, and $r = 0.78$ ($N = 23$, $p < 0.01$) under $\text{RH} > 60\%$.

3.3 Synergistic Role of Aqueous and Photochemical Processes

During the campaign, the average diurnal variation of aqSOA in Nanjing exhibited a distinct morning peak at approximately 09:00 LT (Fig. S2d). This pattern is different from most of the previous urban observations, where aqSOA concentrations typically peak during nighttime periods (Sun et al., 2016; Xu et al., 2019; Gu et al., 2023). To address this, the time series of aqSOA was analyzed on a daily basis, with rainy days excluded. Based on this analysis, two distinct diurnal variation patterns were identified: type I (22 d) and type II (9 d) (Fig. 1a). For diurnal pattern type I, the daily variation of aqSOA exhibits a pronounced morning peak, whereas the daytime concentration of aqSOA shows a noon peak for diurnal pattern type II (Fig. 3). These two distinct diurnal patterns of aqSOA could be attributed to multiple formation mechanisms of aqSOA or different meteorological conditions. It should be noted that the observed morning peak around 09:00–10:00 LT may be modulated not only by chemical production but also by concurrent physical processes, such as morning boundary-layer evolution and mixing/dilution effects. Nevertheless, the diurnal patterns and classification discussed herein are interpreted primarily under chemically constrained conditions, and the main conclusions about aqSOA formation pathways rely chiefly on chemical relationships and correlations.

During diurnal pattern type I, aqSOA concentrations exhibited a pronounced single peak between 09:00 and 10:00 LT, following a period of continuous nighttime accumulation from approximately 20:00 to 07:00 LT (Fig. 3a). Notably, the aqSOA peak lagged behind the nitrate peak by about one hour (based on the average peak-time difference), and nocturnal increases in nitrate, ALW, and aqSOA were

highly synchronized. This ~ 1 h lag may be attributed to the kinetic limitations of the nitrate-mediated aqSOA formation processes. Compared with type II, nighttime meteorological conditions during type I were characterized by higher RH (73.1 % vs. 61.2 %) and ALW concentrations ($34.2 \mu\text{g m}^{-3}$ vs. $12.6 \mu\text{g m}^{-3}$). Such humid conditions favor the formation of both nitrate and ALW, thereby enhancing aqSOA production through aqueous-phase reactions. Moreover, the higher RH conditions observed during type I events could further promote aqueous reaction pathways, as the nitrate formation from N_2O_5 hydrolysis on aqueous aerosol surfaces is strongly facilitated under such conditions (Sun et al., 2018). Overall, these observations indicate that aqSOA production under type I conditions is predominantly controlled by aerosol aqueous-phase processes, highlighting the important role of nocturnal multiphase chemistry in driving morning aqSOA peaks.

In contrast, aqSOA concentrations gradually increased during the daytime and peaked around 13:00 LT (Fig. 3b) during diurnal pattern type II, indicating an influence of photochemical processes. The diurnal variation of aqSOA is consistent with that of nitrate aerosols. This suggests that enhanced photochemical activity is associated with nitrate aerosol formation, which may subsequently mediate aqSOA production via water uptake. O_x is commonly regarded as a tracer for photochemical processes. The relationship between daytime (08:00–17:00 LT) aqSOA concentrations and O_x levels was examined for both diurnal patterns (Fig. 4). For diurnal type I, no significant correlation was observed, suggesting that daytime photochemistry played a minimal role in aqSOA production under these conditions. In contrast, diurnal type II displayed a positive correlation, implying that aqSOA formation was partially related to the photochemical processes.

Further analysis revealed that the slope of the aqSOA- O_x relationship under high RH conditions in diurnal type II was approximately three times higher than that under low RH conditions. This is because higher RH likely enhanced the partitioning of semi-volatile photochemical oxidation products into the aqueous phase and promoted aqueous-phase photochemical reactions, thereby facilitating aqSOA formation. These results indicate that daytime aqSOA production during type II is governed by the combined effects of photochemical oxidation and humidity-dependent aqueous processing.

4 Conclusions

We analyzed the characteristics and formation mechanisms of aqSOA in the autumn season in Nanjing, where aqSOA was found to constitute a significant fraction of organic aerosols (20.2 %). The results demonstrate that nitrate, ALW, and aerosol acidity act in concert to regulate aqSOA formation, with nitrate serving as a critical driver through its strong hygroscopicity. We observed the occurrence of a distinct morning peak of aqSOA in the YRD region, a feature that differs from most of the previously reported studies in urban environments. Based on diurnal variation analysis, two distinct diurnal patterns were identified: type I, dominated by nighttime aqueous-phase chemistry linked to nitrate and ALW accumulation; and Type II, shaped primarily by daytime photochemical oxidation. These findings highlight that aqSOA formation in this megacity is governed by the synergistic effects of both nocturnal aqueous reactions and daytime photochemical processes.

China is strictly controlling SO_2 emissions, which has led to a decrease in sulfate content in atmospheric particulate matter. Nitrate is becoming a more important inorganic component of $PM_{2.5}$, and its role in promoting aqSOA formation will become even more significant. Although this study did not directly resolve the specific precursor compounds of aqSOA, our results still highlight the importance of paying greater attention to nitrate-driven aqueous processes in future air quality assessments. In particular, coordinated management of nitrogen oxides (NO_x) alongside traditional particulate matter controls may provide an effective pathway for mitigating aqSOA burdens in a post-sulfate-dominated atmosphere.

Data availability. All data have been deposited in the Zenodo (Liu and Wang, 2026).

Supplement. The supplement related to this article is available online at <https://doi.org/10.5194/acp-26-3185-2026-supplement>.

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Competing interests. The contact author has declared that none of the authors has any competing interests.

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