



Chlorine enhances nocturnal heterogeneous uptake of NO_2 in coastal atmosphere under sea-land breeze circulation

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Abstract. Heterogeneous uptake of NO_2 serves as a significant source for reactive nitrogen species, playing an important role in atmospheric chemistry. Laboratory studies have demonstrated that chlorine can promote the heterogeneous NO_2 uptake, yet this effect under real ambient conditions remains poorly elucidated. Based on comprehensive field observations, a machine learning technique, and a multiphase chemical box model in the coastal city of Xiamen, China, this study reveals the enhancement effect of chlorine (Cl) on NO_2 uptake and quantifies the impact of this enhanced uptake on reactive nitrogen species during nocturnal sea-land breeze (SLB) periods. Compared with non-SLB days, nocturnal concentrations of nitrous acid (HONO) and particulate nitrate (NO_3^-) increased significantly during SLB days, with high mean value of NO_2 uptake rate constant (k_{NO_2}) reaching $9.70 \times 10^{-6} \text{ s}^{-1}$. Machine learning revealed that chlorine was the most important influencing factor for the enhanced k_{NO_2} . Incorporating this k_{NO_2} into the chemical box model substantially resolved the underestimation of HONO concentration and NO_3^- production under SLB conditions. Notably, nocturnal NO_2 uptake dominated HONO formation (83.9 %), while making a substantial contribution (46.4 %) to nitrate formation. This study highlights the critical role of chlorine-enhanced NO_2 uptake in atmospheric reactive nitrogen cycling and provides valuable insights for nocturnal chemistry in complex coastal environments.

1 Introduction

Nitrous acid (HONO) and nitrate (NO_3^-) are key reactive nitrogen species in the atmosphere. HONO, a major source of OH radicals, significantly contributes to the formation of O_3 and secondary aerosols by influencing radical budget (Acker et al., 2006; Alicke et al., 2003). NO_3^- is an important chemical component of particulate matter (PM), contributing to PM pollution and acid rain (Cao et al., 2022; Qu and Han, 2021). As a crucial source of both HONO and NO_3^- , the het-

erogeneous uptake of NO_2 has attracted considerable attention (Xuan et al., 2025; Carter et al., 1982).

Traditionally, the uptake of NO_2 in bulk solution was considered extremely slow, and thus this process was often negligible (Lee and Schwartz, 1981). However, numerous field observations have indicated that NO_2 uptake plays a significant role in the formation of HONO and NO_3^- . Previous studies have revealed a strong correlation between NO_2 and HONO after excluding homogeneous formation and primary emissions, suggesting efficient heterogeneous conversion of

NO₂ to HONO (Su et al., 2008; Zhang et al., 2024). Studies have also found that incorporating NO₂ uptake mechanisms into models can substantially improve the HONO simulation, especially at night (Zhang et al., 2023, 2024; Xuan et al., 2024; Shi et al., 2020). Using an aerosol-fog sampling system and a box model, Xu et al (2024), found that NO₂ uptake was the primary driver of the rapid increase in NO₃[−] during fog events. Currently, the reported reaction rate constant for NO₂ uptake vary widely across different types of surfaces, spanning several orders of magnitude from 10^{−7} s^{−1} to 10^{−5} s^{−1} (Zhang et al., 2025b). Multiple factors have been demonstrated to affect the NO₂ uptake rate constant (*k*NO₂) in laboratories, including NO₂ concentrations, aerosol particle size, ionic strength, pH, NH₃, and liquid water content (Liu and Abbatt, 2021; Gen et al., 2024; Zhang et al., 2025a). In the real atmosphere, research during haze events showed that parameterization schemes of *k*NO₂ considering relative humidity (RH) and NH₃ influences significantly improved the simulation results of heterogeneous HONO formation (Zhang et al., 2023). Long-term observations revealed that changes in aerosol chemical composition could alter aerosol pH, consequently affecting the uptake kinetics of NO₂ to HONO (Zhang et al., 2025b).

Investigations in coastal areas reported higher NO₂ uptake rates in marine air masses compared to continental ones (Zha et al., 2014; Yang et al., 2021), overall surpassing those measured at continental urban, suburban, and rural sites (Wentzell et al., 2010; Kleffmann et al., 2003; Zhang et al., 2020; Hao et al., 2020; Cui et al., 2018). These findings suggest that heterogeneous NO₂ uptake plays a particularly critical role in the coastal troposphere. Within the marine boundary layer, chloride ions are abundant due to sea spray. Chloride ions exhibit special surface propensity, making them tend to distribute at the air-water interface (Knipping et al., 2000; Piatkowski et al., 2014; Jungwirth and Tobias, 2006). Laboratory studies demonstrated that chloride ions can establish a surface-directed electric field, which attracts atmospheric NO₂ molecules and promotes their further uptake at air-water interface (Wang et al., 2025a; Zhang et al., 2025a). Nevertheless, the influence of chlorine on NO₂ uptake in realistic coastal atmosphere and the consequent impacts on reactive nitrogen cycling remain to be fully elucidated.

Sea-land breeze (SLB) events occur frequently in coastal urban areas (Chen et al., 2023; Liu et al., 2025; Xie et al., 2023). During SLB periods, the coastal boundary layer experiences intensive interactions between marine and continental air masses, providing a natural laboratory to investigate the effect of sea-salt chloride on NO₂ uptake. In this study, a multi-parameter field observation was conducted in Xiamen, a typical coastal city in China, during the autumn and winter seasons characterized by frequent SLB events (Chen et al., 2023). Based on observations, the characteristics of SLB days were revealed and *k*NO₂ under both SLB and non-SLB conditions were quantified. Using machine learning method,

the significant promoting effect of chlorine on NO₂ uptake during nocturnal SLB periods was demonstrated. Furthermore, the obtained *k*NO₂ was incorporated into an advanced multiphase box model to evaluate the impact of chlorine-enhanced NO₂ uptake on atmospheric reactive nitrogen cycling. This study highlights the role of previously overlooked chlorine on NO₂ uptake for nocturnal chemistry in coastal atmosphere.

2 Methods

2.1 Field measurements and sea-land breeze identification

Our observations were conducted at the Atmospheric Observation Supersite (24.61° N, 118.06° E) of the Institute of Urban Environment (IUE), Chinese Academy of Sciences in Xiamen, from 18 November 2023 to 6 January 2024. Xiamen is a coastal city in southeastern China. The observation site is located approximately 5 km from the shore and is surrounded by residential, commercial areas and major traffic arteries. Instruments were deployed on the roof of the IUE building at approximately 70 m above ground level. Virtually no high buildings obstructed the path between the site and the coast, making it a typical coastal urban site.

A suite of atmospheric parameters was observed simultaneously. Ambient HONO was measured online by a water-based long-path absorption photometer (Zhichen Beijing, China) (Xuan et al., 2024). Sulfuric acid (H₂SO₄) and HCl were measured using a long time-of-flight chemical ionization mass spectrometer, equipped with a nitrate reagent ion source (Aerodyne Research Inc., USA) (Yang et al., 2023). The chemical composition of PM_{2.5} was analyzed using both a MARGA ADI 2080 (Metrohm Applikon, Switzerland) and an ACSM (Aerodyne Research Inc., USA) (Liu et al., 2022; Chen et al., 2022). Particle number size distribution (PNSD) in the range of 2–300 nm was measured by two scanning mobility particle sizers (TSI inc., USA) equipped with a long-DMA (2.5–62 nm) and a nano-DMA (7–300 nm), respectively (Li et al., 2025). Additionally, a range of commercial instruments were employed to monitor trace gases (NO, NO₂, SO₂, O₃, and VOCs), meteorological parameters (ambient temperature, RH, wind-speed, wind-direction and ultraviolet radiation), and photolysis rates (*J*O¹D, *J*NO₂, *J*HONO, *J*NO₃, *J*HCHO and *J*H₂O₂) (Liu et al., 2022). Detailed descriptions of instrument operation, species identification, and calibration are provided in Sect. S1 in the Supplement. Hourly BLH data were obtained from the ERA5 reanalysis dataset (Hersbach et al., 2020).

Sea-land breeze days were identified based on observed wind-speed and wind-direction according to the criteria established in our previous work (Chen et al., 2023). Briefly, the coastline in Xiamen is characterized by a northeast-southwest trend, the sea breeze is blowing from the east-south, and the land breeze is blowing from the west-north-

northeast. According to observed wind direction, a natural day could be divided into four periods. Specifically, 01:00–08:00 LT (land breeze period), 13:00–20:00 LT (sea breeze period), and 09:00–12:00 LT and 21:00–24:00 LT (mixed sea-land breeze period). A day was classified as an SLB day if the following three conditions were met: (1) the 24 h mean wind speed was $< 10 \text{ m s}^{-1}$; (2) during the land breeze period, the land breeze persisted for $\geq 4 \text{ h}$ and the sea breeze lasted $\leq 2 \text{ h}$; and (3) during the sea breeze period, sea breeze persisted for $\geq 4 \text{ h}$ and the land breeze lasted $\leq 2 \text{ h}$. As shown in Fig. S1 (light blue areas), a total of 23 SLB days were identified during the observation period. The remaining days were classified as non-SLB days. Figure S2 illustrates that the wind patterns on SLB days differed from those on non-SLB days, characterized by a higher prevalence of southeasterly sea breezes.

2.2 Calculation of HONO from heterogeneous formation processes and NO₂ uptake rate

Due to the absence of agricultural fields near the sampling site and its proximity to major traffic arteries, the primary emission of HONO was attributed to vehicles. Only nighttime data were used in our study, the uncertainties associated with photochemical process, such as NO₃[−] photolysis were excluded. Therefore, the nighttime corrected HONO (HONO_{corr}) is calculated as follows:

$$\text{HONO}_{\text{corr},t} = \text{HONO}_t - E_{\text{vehicle},t} - P_{\text{NO}+\text{OH},t} \quad (1)$$

where HONO_{*t*} is the observed HONO concentration, *E_{vehicle,t}* is the HONO emission from vehicles, and *P_{NO+OH,t}* is the HONO production from homogeneous reactions. *E_{vehicle,t}* is calculated as the NO_x concentration multiplied by the emission factor 0.005 (Zhang et al., 2025b). *P_{NO+OH,t}* is calculated as follows:

$$P_{\text{NO}+\text{OH},t} = k_{\text{NO}+\text{OH}} \times [\text{NO}] \times [\text{OH}] \times 3600 \text{ s} \quad (2)$$

where *k_{NO+OH}* (cm³ molec.^{−1} s^{−1}) is calculated in the Master Chemical Mechanism 3.3.1 (MCM v3.3.1 <https://mcm.york.ac.uk/MCM/>, last access: 4 September 2026), 3600 s represents the 1 h integration interval, and the concentration of OH radicals is estimated based on the steady-state balance of H₂SO₄ (Nie et al., 2022). Because H₂SO₄ is primarily produced by the reaction of SO₂ and OH and is mainly lost via condensation sink (CS) onto particle surfaces, the OH concentration can be estimated as below:

$$[\text{OH}] = \frac{[\text{H}_2\text{SO}_4] \times \text{CS}}{k_{\text{OH}+\text{SO}_2} \times [\text{SO}_2]} \quad (3)$$

where *k_{OH+SO₂}* is $5 \times 10^{-13} \text{ cm}^3 \text{ s}^{-1}$ (Seinfeld et al., 1998), and CS (s^{−1}) is calculated as follows (Kulmala et al., 2012):

$$\text{CS} = 2\pi D_v \int_0^{d_{p,\text{max}}} d_p \beta n(d_p) dd_p \quad (4)$$

where *D_v* is the diffusion coefficient of H₂SO₄, *d_p* is the particle diameter derived from the measured PNSD, *β* is the transitional correction factor, and *n(d_p)* is the particle number concentration of diameter *d_p*.

In line with previous studies (Zhang et al., 2025b, 2020, 2024), the heterogeneous conversion of NO₂ to form HONO_{corr} at night is considered a pseudo-first-order reaction. The reaction rate constant (*k_{NO₂}*, s^{−1}) is calculated based on the conversion of NO₂ during a given period (from *t*₁ to *t*₂) as follows:

$$k_{\text{NO}_2} = \frac{\text{HONO}_{\text{corr},t_2} - \text{HONO}_{\text{corr},t_1}}{\overline{\text{NO}_2}(t_2 - t_1)} \quad (5)$$

where $\overline{\text{NO}_2}$ is the mean concentration of NO₂ between *t*₁ and *t*₂.

To evaluate the sensitivity of the results to the assumed emission factor, the calculation of HONO_{corr} and *k_{NO₂}*

2.3 Machine learning technique

Random Forest (RF) algorithm, a machine learning technique extensively applied in atmospheric (Yang et al., 2024; Lin et al., 2026b; Zhang et al., 2025b), was employed to elucidate the primary drivers of *k_{NO₂}*

fects of photochemical processes on kNO₂ quantification, thereby enabling a more precise identification of key drivers. Detailed RF hyperparameter configurations and model evaluation are provided in Sect. S2. In this study, the dominant factors were identified based on RF factor importance, which quantifies the percentage contribution of each variable to kNO₂ (Breiman, 2001; Gregorutti et al., 2017). Furthermore, one-dimensional partial dependence plots (PDPs) were employed to delineate the main effects of individual variables, while bivariate PDPs were utilized to visualize and access the interaction effects between pairs of factors on kNO₂ (Goldstein et al., 2015).

2.4 Multiphase chemical box model

The multiphase chemical box model was established based on the Framework for 0-D Atmospheric Modeling (F0AM) (Wolfe et al., 2016). As shown in Tables S1 and S2, comprehensive HONO and NO₃[−] budgets were incorporated into the F0AM model to simulate the variations of HONO and NO₃[−]. The ISORROPIA II model (Seinfeld et al., 1998) was integrated into the model to determine aerosol pH, aerosol liquid water content and the partitioning ratio of HNO₃ and NO₃[−]. Six simulation scenarios were conducted: simulations without heterogeneous NO₂ uptake for (1) SLB and (2) non-SLB days; simulations with NO₂ uptake for (3) SLB and (4) non-SLB days, using the mean kNO₂ derived from SLB days for the SLB simulation and that derived from non-SLB days for the non-SLB simulation, respectively; and two additional sensitivity simulations for SLB days using the (5) 25th and (6) 75th percentiles of the kNO₂ distribution derived from SLB days. The ability of the model to reproduce real atmospheric processes was evaluated by comparing the simulated HONO and NO₃[−] with observations. Due to the underestimation of total particulate nitrate concentrations measured by instruments, the performance of NO₃[−] simulations was evaluated by comparing the trends of the simulated NO₃[−] production rate with the observed NO₃[−] concentration. In contrast, as HONO predominantly exists in the gas phase, the simulated HONO was evaluated directly against the observed HONO concentrations.

As for the model simulation, each simulation was preceded by a 3 d spin-up to allow intermediate species concentrations to stabilize. The mean diurnal pattern of trace gas concentrations, meteorological parameters, and the results from the ISORROPIA II model during SLB days and non-SLB days were used as input data to constrain the box model. Notably, the concentrations of HONO and NO₃[−] were not constrained by observed values. More details about the model incorporating mechanisms, modules, and simulation settings are provided in Sect. S3 and our previous study (Lin et al., 2026a).

3 Results and discussion

3.1 Characteristics of atmospheric species during SLB days

Figure S1 shows the time series of the observed parameters. Based on the identification criteria detailed in the Methods section, 23 d were classified as SLB days, accounting for 48 % of entire observation period. The diurnal patterns of major observed parameters are summarized in Fig. S4, revealing that the primary differences between the two types of days predominantly emerged at night. Compared to non-SLB days, SLB days exhibited rapid nighttime accumulation of HONO and NO₃[−], accompanied by elevated chlorine levels, NO_x concentrations, and aerosol surface area (SA) concentrations. Additionally, consistent with our previous study, SLB days were characterized by higher temperature and RH at night (Chen et al., 2022). Figure 1 and Table S3 further summarizes the nighttime observational results on SLB and non-SLB days. By definition, SLB days are influenced by persistent sea breezes, which likely enhances the impact of sea-salt aerosols. As shown in Fig. 1a, the Cl[−] concentrations were comparatively higher on SLB days (0.154 μg m^{−3}) than on non-SLB days (0.139 μg m^{−3}). Specifically, this overall increase (Fig. S5) was primarily driven by elevated levels during the sea breeze (18:00–20:00 LT) and mixed sea-land breeze (21:00–00:00 LT) periods, while concentrations declined during the land breeze (01:00–06:00 LT) period. The elevation pattern of Cl[−] indicates that the more abundant Cl[−] transported by stronger sea breezes on SLB days can persist into the subsequent mixed sea-land breeze periods. Additionally, Fig. S6 shows a moderate positive correlation between Na⁺ and Cl[−] on SLB days ($r = 0.67$), compared with a weaker relationship on non-SLB days ($r = 0.28$). Because sodium chloride (NaCl) is a major component of sea salt, this contrast is consistent with a greater sea-salt influence on Cl[−] on SLB days. As for gaseous chlorine, Fig. 1b shows that the concentration of HCl was significantly higher on SLB days (96.2 ppt) than on non-SLB days (67.5 ppt). This further confirms that, under the influence of sea-salt aerosols on SLB days, chlorine levels are distinctly elevated in the nighttime coastal troposphere.

Figure 1c indicates that the concentration of NO₂ on SLB days (mean: 28.3 ppb) was significantly higher than that on non-SLB days (23.2 ppb). Within our study area, ship emissions, cross-sea bridge traffic, and road traffic contribute substantial NO_x, and the enhanced local circulation on SLB days can promote NO₂ accumulation (Chen et al., 2023). The mean concentrations of HONO and NO₃[−] on SLB day were 0.836 ppb and 3.61 μg m^{−3}, respectively, which were significantly higher than the corresponding values of 0.497 ppb and 3.13 μg m^{−3} on non-SLB days (Fig. 1d, e). To better understand the elevated HONO levels, HONO_{corr} concentrations, defined as the observed HONO concentration subtracts contributions from primary emissions and gas-

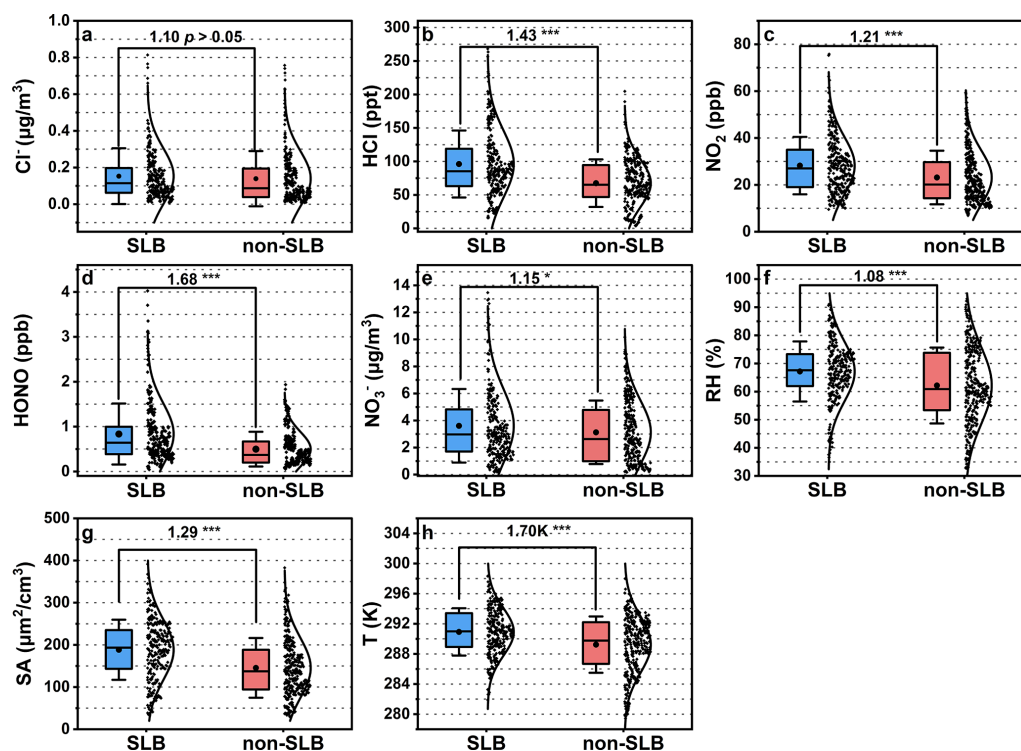
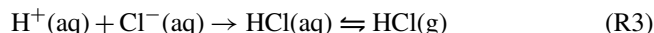
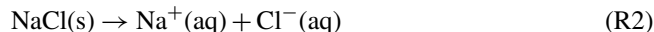
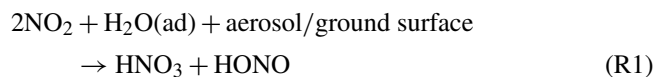


Figure 1. Box plots of key observed parameters during nocturnal SLB and non-SLB periods. (a) Cl[−], (b) HCl, (c) NO₂, (d) HONO, (e) NO₃[−], (f) RH, (g) SA, and (h) ambient temperature (*T*). The box shows the 25th–75th percentiles with whiskers representing ± 1 standard deviation. The black dot and black line within the box represent the mean and the median. Differences between SLB and non-SLB days were tested using a two-sample *t*-test. Statistical significances are denoted as * ($p < 0.05$), ** ($p < 0.01$) and *** ($p < 0.001$). For all panels except (h), the number adjacent to the significance symbol denotes the ratio of the mean value on SLB days to that on non-SLB days; in panel (h), it denotes the absolute difference between the means.

phase production, were calculated to represent the nocturnal HONO produced from heterogeneous conversion. The ratio of HONO_{corr} / NO₂ was notably higher on SLB days (0.030) than on non-SLB days (0.021), indicating more active heterogeneous conversion of NO₂ to HONO on SLB days. Multiple studies have confirmed that NO₂ uptake is a primary source of nocturnal heterogeneous HONO formation (Zhang et al., 2023, 2022). Here, the higher SA concentrations on SLB days were also conducive to heterogeneous reactions (Fig. 1g). Furthermore, Fig. 2a–b indicate a moderate positive correlation between HONO_{corr} and NO₃[−], with the correlation coefficient (*r*) increasing from 0.56 on non-SLB days to 0.76 on SLB days. The HONO_{corr} concentrations under the two conditions overlapped mainly below 1.5 ppb. Within this overlapping range, the correlation on SLB days remained significant ($r = 0.63$, $p < 0.001$, Fig. S7a) but was close to that on non-SLB days ($r = 0.56$), indicating that the wider concentration range partly contributed to the higher full-range correlation on SLB days, while the positive relationship between HONO_{corr} and NO₃[−] persisted within the overlapping range. As shown in Reaction (R1), NO₂ uptake via disproportionation simultaneously produces HONO and HNO₃, the latter of which is further converted to NO₃[−] via

gas-particle partitioning. Thus, nocturnal NO₂ uptake is also likely an important contributor to the formations of NO₃[−] on SLB days. Figure 2c, d show that the full-range correlation between HONO_{corr} and gaseous HCl was stronger on SLB days than on non-SLB days ($r = 0.58$ versus 0.16). Within the overlapping HONO_{corr} range below 1.5 ppb, the correlation on SLB days decreased to ($r = 0.47$) but remained significant ($p < 0.001$, Fig. S7b) and appreciably higher than that on non-SLB days. This result indicates that the stronger HCl–HONO_{corr} relationship during SLB periods cannot be explained solely by the wider HONO_{corr} range. The H⁺ generated by NO₂ uptake can combine with Cl[−] (Reactions R2–R3). Simultaneously, the aerosol pH in the coastal region of southeastern China was relatively low (approximately 1.92 ± 0.34 during nocturnal SLB periods), chloride depletion could easily occur, leading to the observed positive correlation between HONO_{corr} and HCl. The enhanced chloride depletion can also explain our observations of a significant increase in gaseous HCl but not in Cl[−] concentrations on SLB days. Similarly, the rapid decline in Cl[−] concentration during land breeze periods (Fig. S4) could also be explained

by chloride depletion.



3.2 Enhanced rate of NO₂ uptake and its influencing factors

The values of kNO₂ on SLB days and non-SLB days were derived as Fig. S8. Overall, the average kNO₂ values on SLB days ($9.70 \times 10^{-6} \text{ s}^{-1}$) were significantly higher than those on non-SLB days ($6.72 \times 10^{-6} \text{ s}^{-1}$), by a factor of approximately 1.44. Across the tested emission factors, the mean HONO_{corr} decreased from 0.58 to 0.51 ppb on SLB days, whereas kNO₂ varied by no more than 6.2 %, indicating limited sensitivity of the derived rate constant to the assumed emission factor (Fig. S3). As presented in Table S4, a prolonged increase in the HONO_{corr}/NO₂ ratio was also observed during 17 of the 23 SLB days (see an example in Fig. S9). In contrast, this phenomenon was virtually absent on non-SLB days, indicating a sustained nocturnal enhancement of kNO₂ during SLB periods. Compared with existing field observations, the kNO₂ on SLB days was broadly consistent with the levels reported under the sea-case scenario in coastal cities of China such as Qingdao ($1.25 \times 10^{-5} \text{ s}^{-1}$) and Hong Kong ($8.81 \times 10^{-6} \text{ s}^{-1}$) (Yang et al., 2021; Zha et al., 2014), while it was markedly higher than those values observed in major non-coastal cities including Beijing (7.78×10^{-7} – $5.75 \times 10^{-6} \text{ s}^{-1}$) (Wang et al., 2017b; Zhang et al., 2020, 2022; Xuan et al., 2023; Jia et al., 2020), Nanjing (1.19×10^{-6} – $2.22 \times 10^{-6} \text{ s}^{-1}$) (Zheng et al., 2020; Liu et al., 2019), and Guangzhou (4.44×10^{-6} – $6.67 \times 10^{-6} \text{ s}^{-1}$) (Su et al., 2008; Li et al., 2012; Qin et al., 2009). The consistently higher kNO₂ in coastal areas implies that coastal atmospheric environments are likely more conducive to NO₂ uptake.

A machine learning technique was used to further identify key drivers of the enhanced kNO₂ on SLB days. The RF models using Cl_{total} reproduced the calculated kNO₂ well during both SLB and non-SLB periods, with *R*² values of 0.91 and 0.87, respectively (Fig. S10c, d). Replacing Cl_{total} with either Cl[−] or HCl resulted in the respective chlorine variable ranking third after RH and SA (Figs. S10–S11). Particulate Cl[−] represents the residual particle-phase fraction, whereas HCl contains information on the fraction transferred to the gas phase through chloride depletion. Their respective importance therefore suggests that the Cl_{total} signal was not determined by either component alone, but reflected their complementary information on the Cl[−] burden before depletion. All VIFs were below 10, indicating no severe multicollinearity between the chlorine metrics and the other pre-

dictors (Table S5). Because all three metrics retain part of the chlorine-related information, the RF models showed similar *R*² values (0.89–0.91); nevertheless, the Cl_{total}-based model had the highest *R*² and provided a more physically representative description of the association between depletion-corrected particulate chloride and kNO₂. Notably, the role of Cl_{total} and RH diminished substantially on non-SLB days (Fig. 3b), underscoring the profound modulation of these two influencing factors by the SLB circulation. Figure 3c–e further elucidate the specific impacts of these primary drivers on nocturnal kNO₂ under SLB conditions. Specifically, Cl_{total} exhibits a positive correlation with kNO₂ (Fig. 3c). This phenomenon is likely attributable to the preferential partitioning of chloride ions at the gas–aqueous interface of aerosols (Zhang et al., 2025a; Shen et al., 2025; Wang et al., 2025a). The surface-bound Cl[−] can attract gaseous NO₂ molecules (Reaction R4) to form interfacial intermediates [Cl–NO₂][−], thereby facilitating the overall NO₂ uptake process. Subsequent interaction analysis was consistent with the mechanisms proposed in laboratory studies (Reactions R4–R5) and suggest that similar processes may operate in the coastal boundary layer. Figure 3f shows that at a fixed NO₂ concentration, elevated chloride levels amplify the positive effect of NO₂ molecules on kNO₂. This indicates that chloride ions actively promote the heterogeneous uptake of NO₂, consequently accelerating the kNO₂. Similar halogen-promoted effect has also been revealed in NO₂-initiated aqueous-phase oxidation of dissolved SO₂ at the air–aqueous interface and of nitrate photolysis at the aqueous-phase surface (Wang et al., 2025a; Shen et al., 2025). Figure 3d shows a significant positive relationship exists between RH and kNO₂, suggesting that moist sea breezes effectively supply the water molecules essential for the heterogeneous NO₂ uptake. This finding aligns with previously reported field observations highlighting RH-driven NO₂ uptake (Xuan et al., 2024). Analogous to Cl_{total}, the interaction between RH and NO₂ reveals that at a constant NO₂ level, the positive impact of NO₂ on kNO₂ is enhanced under elevated RH. It implies that RH could also facilitate NO₂ uptake by promoting the interaction between NO₂ molecules and heterogeneous interface. Furthermore, Fig. 3h shows a threshold-dependent joint enhancement: kNO₂ increased when RH exceeded approximately 80 % and Cl_{total} exceeded approximately $0.5 \mu\text{g m}^{-3}$, whereas the response was limited when either variable remained below these levels. Elevated RH can increase aerosol liquid water and maintain an aqueous particle phase, while Cl[−] in wetted chloride-containing particles can be preferentially accessible at the air–aqueous interface and enhance NO₂ adsorption and uptake (Zhang et al., 2025a; Wang et al., 2025a). Therefore, the bivariate PDPs response is consistent with high RH providing a favorable aqueous interfacial environment for chloride-associated NO₂ uptake. This interaction may partly explain the higher importance of RH and Cl_{total} on SLB days. As a primary factor on both SLB and non-SLB days, SA positively correlates with kNO₂ (Fig. 3e).

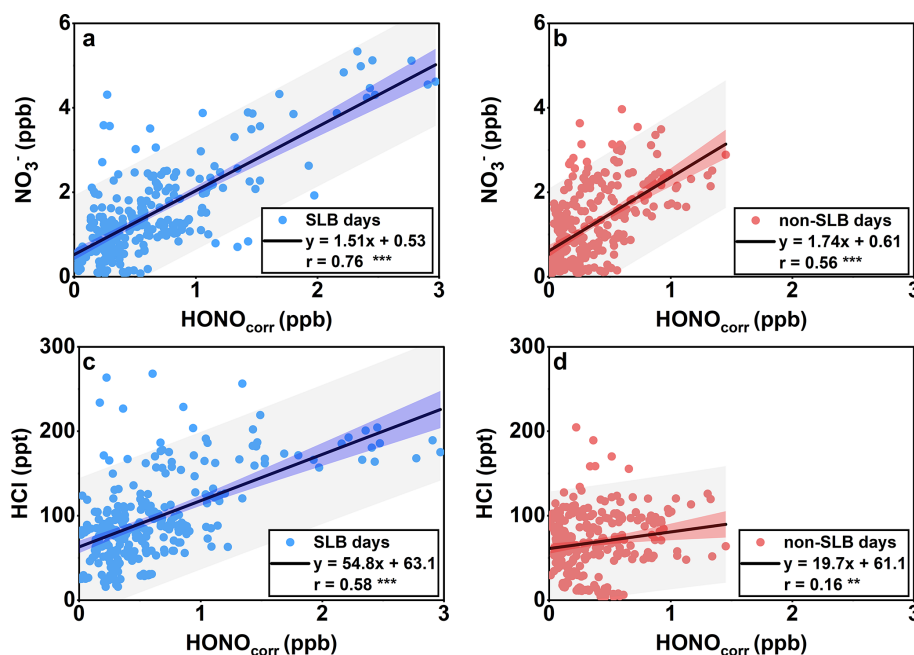
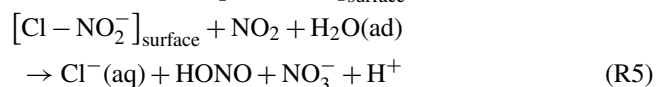
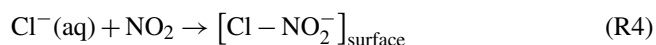


Figure 2. Relationship between HONO_{corr} and the important products of heterogeneous chemical processes during nocturnal SLB and non-SLB periods. HONO_{corr} denotes the corrected HONO concentration from secondary conversion. The products include (a, b) NO₃⁻ and (c, d) HCl. Blue and red dots represent observations during SLB days and non-SLB days, respectively. Each panel displays the linear regression fits and Pearson correlation coefficients (r) for the corresponding periods.

This indicates that an increase in aerosol surface area concentrations provides effective reaction interfaces for NO₂ uptake process, consistent with most atmospheric heterogeneous processes (Wang et al., 2017a; Ma et al., 2026). Moreover, Fig. S12 shows that on SLB days, concentrations of the major NO₂ uptake products (HONO and NO₃⁻) increased with those key factors including Cl_{total}, RH, and SA, providing additional support for the above driving factor analysis. In general, the reasons underlying the enhanced kNO₂ on SLB days help to explain the multi-fold higher kNO₂ observed in coastal areas compared to non-coastal urban areas, highlighting ambient chlorine levels and RH as non-negligible contributors.



3.3 Enhanced NO₂ uptake promotes the atmospheric reactive nitrogen cycle

The above correlations analysis suggests that NO₂ uptake could play an important role in the production of HONO and NO₃⁻ on SLB days. Since the HONO and NO₃⁻ formation were influenced by multiple meteorological factors and many atmospheric reaction pathways, a multiphase chemical box model was employed to evaluate the role of enhanced NO₂ uptake. As described in Methods, six simula-

tion scenarios were set up. Incorporating NO₂ uptake mechanism substantially improved the simulated diurnal HONO profile during SLB periods: r increased from 0.57 to 0.94, while mean absolute error (MAE) and root mean square error (RMSE) decreased from 0.58 to 0.11 ppb and from 0.64 to 0.13 ppb, respectively (Fig. 4a). For NO₃⁻, incorporating NO₂ uptake increased the mean simulated net NO₃⁻ formation rate by 2.14 μg m⁻³ h⁻¹ and strengthened the temporal consistency between the simulated net formation rate and observed concentration, increasing r from 0.04 ($p > 0.05$) to 0.64 ($p < 0.001$, Fig. 4b). This improvement indicates that the added mechanism better represented the observed temporal variability in NO₃⁻ during SLB periods. Additionally, the NO₂ uptake could robustly explain the enhancements in both atmospheric reactive nitrogen species on SLB days across various kNO₂ quantiles (Fig. S13). In contrast, including NO₂ uptake under non-SLB conditions produced only limited improvements in the simulated HONO concentrations and the temporal covariation between NO₃⁻ production rates and observed NO₃⁻ concentrations (Fig. S14). This suggests that NO₂ uptake had less explanatory importance under non-SLB conditions than during SLB periods, consistent with the stronger observed HONO_{corr}–NO₃⁻ relationship on SLB days. Moreover, unlike on SLB days (Fig. S9), HONO_{corr}/NO₂ did not exhibit a sustained nocturnal increase on non-SLB days, which resulted in numerous negative estimates of kNO₂ estimated using Eq. (5). Excluding these negative values likely led to an overestimated mean

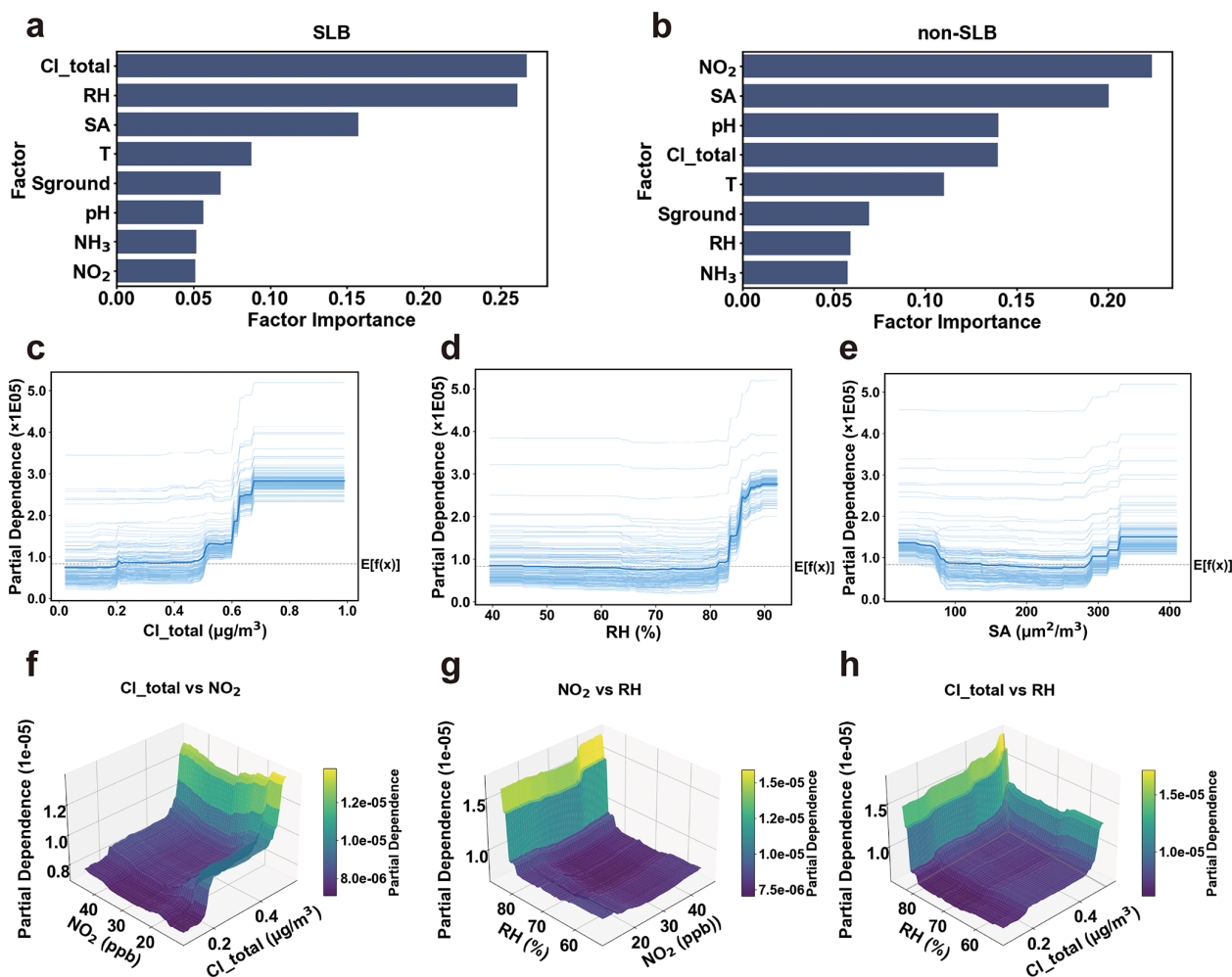


Figure 3. Main and interaction effects of influencing factors for k_{NO_2} . Factor importance determined by the random forest (a) SLB days and (b) non-SLB days. Partial dependence plots for the primary drivers on SLB days: (c) Cl_{total} , (d) RH, and (e) SA. Interactions between key factors on SLB days: (f) Cl_{total} and NO_2 , (g) RH and NO_2 , and (h) RH and Cl_{total} . In panels (a, b), S_{ground} denotes ground surface areas calculated by $1/\text{BLH}$ (boundary layer height). In panels (c–h), partial dependence reflects the extent to which the influencing factors affect k_{NO_2} .

k_{NO_2} and, consequently, to the overprediction of HONO and NO_3^- under non-SLB conditions.

Figure 4c, d illustrate the simulated HONO and NO_3^- budgets for the scenarios incorporating NO_2 uptake on SLB days. Despite potential overestimations of non-SLB k_{NO_2} , the mean NO_2 uptake rate remained significantly higher on SLB days than on non-SLB days, with an average rate of 1.03 ppb h^{-1} (Fig. S15). During nighttime, NO_2 uptake accounted for 83.8 % of HONO formation (Fig. 4e). The predominance of NO_2 uptake in nocturnal HONO formation has also been observed during PM pollution episodes in urban areas (Xuan et al., 2024; Jia et al., 2020). However, unlike the chlorine-driven reason identified here, those prior cases were primarily attributed to favorable heterogeneous reaction conditions, such as larger SA concentrations, lower BLH, and higher RH. In addition to NO_2 uptake, primary emis-

sions from vehicle exhaust, coal combustion, and biomass burning are also recognized as major contributors to nighttime HONO in urban areas (Wang et al., 2025b; Zhang et al., 2022). In this study, primary emissions contributed only 15.5 % to nocturnal HONO formation, substantially less than the contribution from NO_2 uptake. As for nighttime NO_3^- formation, many studies had indicated that N_2O_5 uptake was the dominant pathway in urban areas (Chen et al., 2020; Wang et al., 2018). However, the situation changed under the impact of the enhanced NO_2 uptake. Figure 4f shows that NO_2 uptake contributed 46.4 % to nocturnal NO_3^- formation, which was almost comparable to that of N_2O_5 uptake (51.6 %). These values were obtained using the BT09 parameterization of $\gamma_{\text{N}_2\text{O}_5}$ (Bertram and Thornton, 2009b). Additional simulations using the EJ05 and GRI09 schemes yielded contributions of 47.0 %–47.9 % for NO_2 uptake and

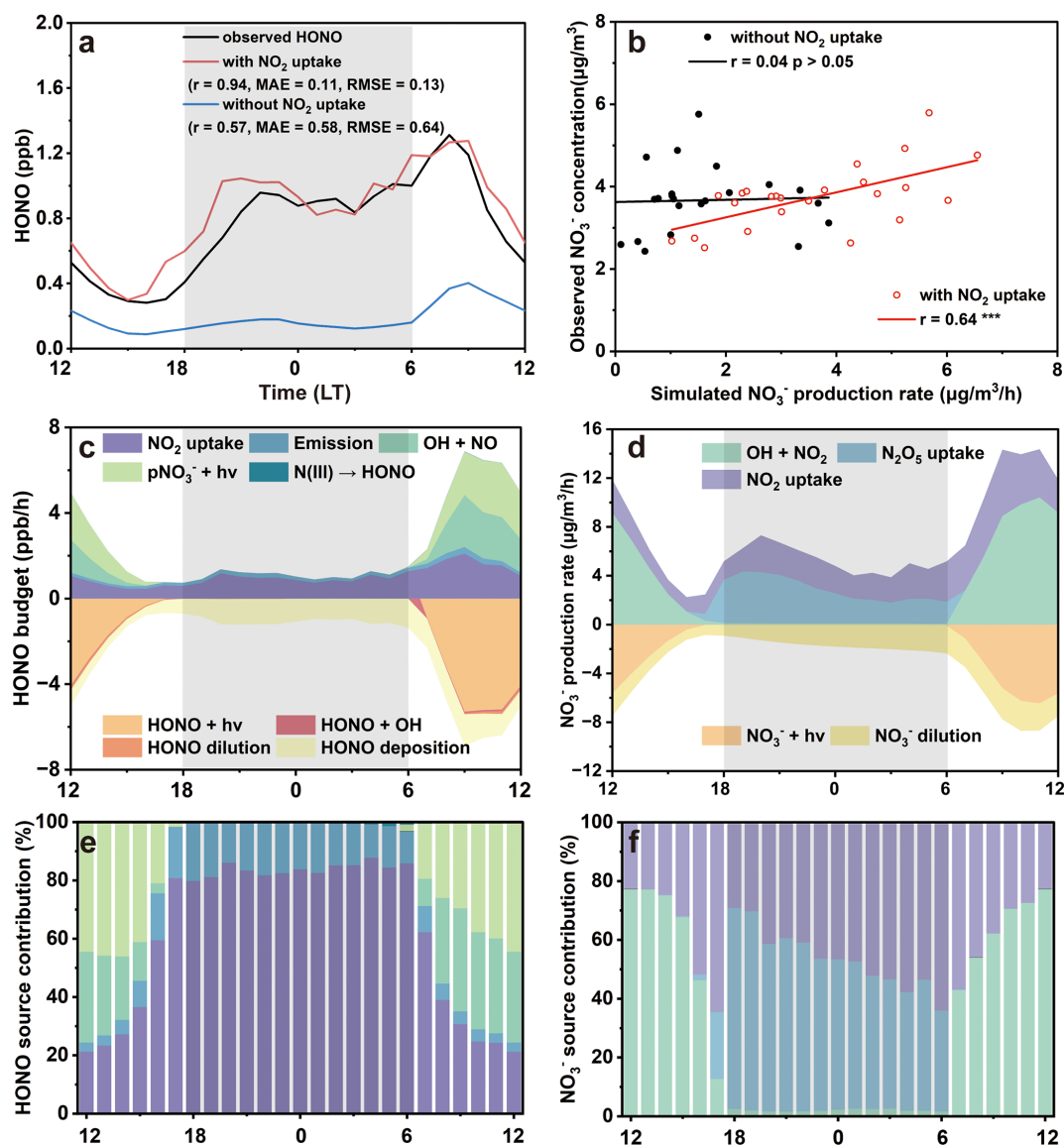


Figure 4. Quantified results from the multiphase chemical box model during SLB periods. (a) Diurnal variations in observed and simulated HONO concentrations with and without NO_2 uptake. (b) Relationship between the simulated NO_3^- production rate and observed concentration with and without NO_2 uptake. Diurnal variations of HONO (c) and NO_3^- budgets (d) during SLB days. Diurnal variations of source contributions to (e) HONO and (f) NO_3^- . The grey shaded areas represent the nighttime period.

50.0 %–50.9 % for N_2O_5 uptake (Fig. S16 and Table S6), indicating that their comparable importance was insensitive to the choice of $\gamma\text{N}_2\text{O}_5$ parameterization (Evans and Jacob, 2005; Hallquist et al., 2003; Griffiths et al., 2009). While chlorine can also promote the uptake of N_2O_5 (Bertram and Thornton, 2009a), the relative enhancement in the contribution of NO_2 uptake to nocturnal NO_3^- formation was more pronounced. This might be related to the suppression of N_2O_5 uptake caused by increased NO_3^- generating directly from NO_2 uptake. The increase in NO_3^- component of aerosol can shift the equilibrium between N_2O_5 and NO_3^- toward N_2O_5 , thereby reducing the contribution of N_2O_5 to

NO_3^- formation (Bertram and Thornton, 2009a; Wagner et al., 2013). Additionally, the enhanced NO_2 uptake process also competed for NO_2 , an important precursor of N_2O_5 formation, especially in the NO_2 -limited regime for N_2O_5 formation (Lin et al., 2025), ultimately reducing the overall enhancement effect on N_2O_5 generation. Moreover, this comparable contribution of NO_2 and N_2O_5 uptake NO_3^- formation was also found during fog events (Xu et al., 2024). Although the observed phenomenon was similar, the dominant drivers differed between fog conditions and coastal SLB periods. The former was primarily attributed to favorable uptake conditions such as abundant aerosol water content and high

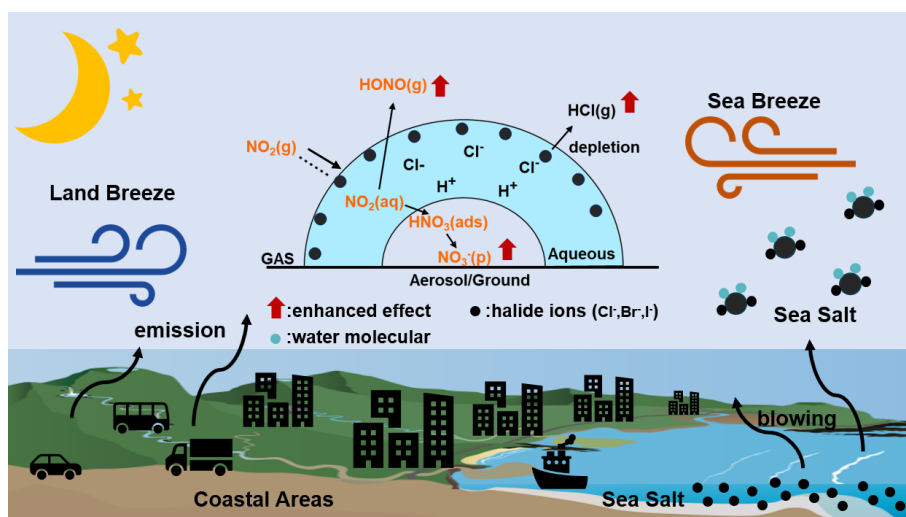


Figure 5. Schematic of the chlorine-enhanced heterogeneous uptake of NO_2 during sea-land breeze days in coastal areas. Red arrows, black solid circles, and blue solid circles represent the enhancement of the uptake process, halide ions and water molecular, respectively.

surface area concentration, whereas the latter was more due to the enhancement effect of chlorine levels and RH on the NO_2 uptake process. Overall, these results reveal the high contribution of NO_2 uptake to HONO and NO_3^- formation during SLB circulation periods, demonstrating that enhanced NO_2 uptake can significantly influence the atmospheric reactive nitrogen cycle in coastal urban areas.

4 Conclusions

Overall, our observations show that nighttime sea-land breeze period in this coastal urban site were characterized by stronger sea-salt aerosol influence, higher relative humidity, and elevated HONO and nitrate concentrations. Nighttime conditions led to a significant enhancement in the NO_2 uptake rate constant, mainly driven by increased chlorine levels and relative humidity. Multiphase chemical box model simulations further confirmed that this enhanced NO_2 uptake promoted HONO and nitrate formation. Figure 5 illustrates a schematic diagram of chlorine-enhanced heterogeneous uptake of NO_2 during sea-land breeze days. Under the influence of strong sea-land breeze circulation, the continuous influx of sea-salt aerosols transported by moist sea breezes leads to high chlorine levels during the night. Meanwhile, emission from vehicles and ships contribute substantial NO_2 , which accumulates in the coastal atmospheric boundary layer (Wen et al., 2023; Sun et al., 2025). Under these conditions, abundant chloride can effectively increase k_{NO_2} and promote the NO_2 uptake process, thereby influencing the formation of key atmospheric reactive nitrogen species, HONO and NO_3^- , as well as broader reactive nitrogen cycling. Notably, other sea-salt halides (Br^- , I^-) exhibit similar chemical properties to chlorine and may also contribute to the elevated k_{NO_2} (Zhang et al., 2025a).

Although the observations were limited to one coastal urban site in autumn and winter, chlorine-enhanced NO_2 uptake may also occur in other coastal environments where chloride-rich sea-salt aerosols, high RH, and sufficient NO_2 coexist. Its magnitude is expected to vary seasonally and regionally with sea-salt loading, aerosol water content, aerosol composition, and NO_2 abundance. Therefore, the quantitative contribution derived here should be considered site- and period-specific, and multi-season observations at other coastal sites are needed to assess its broader relevance. Additionally, our study focuses specifically on nocturnal NO_2 uptake to form HONO , given that daytime HONO formation involves more complex factors such as photo-enhanced heterogeneous reactions of NO_2 and nitrate photolysis (Stemmler et al., 2006; Han et al., 2016; Ye et al., 2017; Zhou et al., 2011). Future investigations are warranted to explicitly evaluate the impact of chlorine on daytime NO_2 uptake. Moreover, current air quality models often exhibit considerable biases in simulating secondary pollutants such as O_3 and $\text{PM}_{2.5}$ over coastal areas (Huang et al., 2025; Ma et al., 2025; Mao et al., 2022). One possible reason is that the influence of chlorine on chemical processes like NO_2 uptake has not yet been adequately considered. Given the dense traffic and substantial NO_2 emission in coastal cities of China, greater attention should be paid to the chlorine-enhanced NO_2 uptake process and its subsequent impact on nocturnal chemistry in coastal urban atmosphere.

Code and data availability. Data analysis methods are available on request from Jinsheng Chen (jschen@iue.ac.cn). The dataset can be accessed at Zenodo: <https://doi.org/10.5281/zenodo.21037074> (Lin et al., 2026c).

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