



Supplement of

Source-explicit estimation of brown carbon in the polluted atmosphere over the North China Plain: implications for distribution, absorption, and the direct radiative effect

Jiamao Zhou et al.

Correspondence to: Guohui Li (ligh@ieecas.cn)

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1 S1 WRF-Chem model general description and configuration

2 In this study, a specific version of the WRF-Chem model (Grell et al., 2005) with modified by Li et 3 al. (2010; 2011a; 2011b; 2012) is used to quantitatively estimate the radiative effect of brown carbon in 4 the NCP. The model was run at a horizontal resolution of 6km with 35 vertical levels, and configured 5 with a single domain (no nesting) of 300×300 grid cells centered at grid point at latitude of 38.0 N and 6 longitude of 116.0 W as shown in Table S1. The model contains a new flexible gas phase chemical 7 module which utilized with SAPRC chemistry mechanism based on the available emission inventory in 8 the present study. The gas-phase chemistry is solved by an Eulerian backward Gauss-Seidel iterative 9 technique with a number of iterations, inherited from NCAR-HANK (Hess et al., 2000).

10 For the aerosol simulations, the CMAQ/models3 aerosol module (AERO5) developed by US EPA 11 has incorporated into the model (Binkowski and Roselle, 2003). The particle size distribution is 12 represented as the superposition of three lognormal modes. The processes of coagulation, particles 13 growth by the addition of mass, and new particle formation are included. The wet deposition follows the 14 method in the CMAQ module and the dry deposition of chemical species is parameterized following 15 Wesely (1989). The photolysis rates are calculated using the Fast Tropospheric Ultraviolet and Visible 16 (FTUV) Radiation Model ((Tie, 2003; Li et al., 2005), with the aerosol and cloud effects on the 17 photochemistry (Li et al., 2011a). The inorganic aerosols is predicted with ISORROPIA (version 1.7) 18 (Nenes et al., 1998) which calculates the thermodynamic equilibrium between the ammonia-sulfate-19 nitrate-chloride-water aerosols and their gas phase precursors of H₂SO₄-HNO₃-NH₃-HCl-water vapor.

20 The organic aerosol (OA) module is based on the volatility basis-set (VBS) approach with aging 21 (Li et al., 2011b). The primary organic aerosol (POA) are assumed semi-volatile and photochemically 22 reactive (Robinson et al., 2007) and distributed in logarithmically spaced volatility bins. Nine surrogate 23 species are used for POA components followed by Shrivastava et al. (2008) with saturation 24 concentrations (C*) ranging from 10⁻² to 10⁶ µg m⁻³ at room temperature. The secondary organic aerosol 25 (SOA) formation from each anthropogenic or biogenic precursor is calculated using four semi-volatile VOCs with effective saturation concentrations of 1, 10, 100, and 1000 µg m⁻³ at 298 K. The SOA 26 27 formation via the heterogeneous reaction of glyoxal and methylglyoxal is parameterized as a first-order irreversible uptake by aerosol particles with an uptake coefficient of 3.7×10^{-3} (Liggio, 2005; Zhao et al., 28

- 29 2006; Volkamer et al., 2007). The OA module has reasonably reproduced the POA and SOA
- 30 concentration against measurements, and detailed model performance can be found in Li et al. (2011b),
- 31 Feng et al. (2016), and Xing et al. (2019).

32 Table S1 WRF-Chem model configurations.

Parameter	Configuration
Regions	The North China Plain (NCP)
Simulation period	January 1 to 30, 2014
Domain size	300×300
Domain center	38.0°N, 116.0°E
Horizontal resolution	6km × 6 km
Vertical resolution	35 vertical levels with a stretched vertical grid with spacing ranging from 30m near the surface, to 500m at 2.5km and 1km above 14km
Microphysics scheme	WSM 6-class graupel scheme (Hong and Lim, 2006)
Boundary layer scheme	MYJ TKE scheme (Janjić, 2002)
Surface layer scheme	MYJ surface scheme (Janjić, 2002)
Land-surface scheme	Unified Noah land-surface model (Chen and Dudhia, 2001)
Long-wave radiation scheme	Goddard longwave scheme (Chou et al., 2001)
Short-wave radiation scheme	Goddard shortwave scheme (Chou and Suarez, 1999)
Meteorological boundary and initial conditions	NCEP $1^{\circ} \times 1^{\circ}$ reanalysis data
Chemical initial and boundary conditions	MOZART 6-hour output (Horowitz et al., 2003)
Anthropogenic emission inventory	SAPRC-99 chemical mechanism emissions developed by Zhang et al. (2009)and Li et al.(2017)
Biogenic emission inventory	MEGAN model developed by Guenther et al. (2006)
Four-dimension data assimilation	NCEP ADP Global Air Observational Weather Data
Model spin-up time	24 hours

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34 S2 Data and methodology

35 S2.1 Observation data description

36 The hourly near-surface measurements of O₃, NO₂, SO₂, CO and PM_{2.5} concentrations have been 37 released in public by the Ministry of Ecology and Environment of China since 2013. The submicron 38 sulfate, nitrate, ammonium, elemental carbon and organic aerosols obtained in two cities including 39 Beijing, Tianjin and the hourly observation of primary OA from, BB, RCC and motor vehicles emissions 40 and SOA in Beijing in January, 2014 are provided by Institute of Earth Environment, Chinese Academy 41 of Sciences. The organic carbon and elemental carbon concentrations are measured using a 42 thermal/optical reflectance carbon analyzer (Model 2001, DRI, USA) (Chow et al., 2004) and water-43 soluble ions are measured using a DX600 ion chromatograph (Dionex Inc., Sunnyvale, CA, USA)

44 (Zhang et al., 2011). The SWDOWN is measured by CM-11 pyranometers at five sites from Chinese 45 Ecosystem Research Network (CERN) in the NCP, including Beijing, Tianjin, Zhengzhou, Hefei, and 46 Ji'nan. The hourly measurement of OA in Beijing is measured by the Aerodyne high-resolution time-of-47 flight aerosol mass spectrometer (HR-ToF-AMS) with a PM_{2.5} lens from 9 to 25 January, 2014 at the 48 Institute of Remote Sensing and Digital Earth, Chinese Academy of Sciences (Li et al., 2018). The 49 positive matrix factorization (PMF) method is used to distinguish the sources of OA as hydrocarbon-like 50 OA, biomass burning OA, coal combustion OA (Elser et al., 2016), which are interpreted for surrogates 51 of primary OA (POA)-TRA, POA-BB, POA-COAL, and oxygenated OA is the surrogate of SOA in this 52 paper.

53 S2.2 Statistical metrics for simulation comparisons

54 In this study, the mean bias (MB), root mean square error (RMSE) and the index of agreement 55 (IOA) are used to evaluate the model performance in simulating air pollutants.

56
$$MB = \frac{1}{N} \sum_{i=1}^{N} (P_i - O_i)$$
 (S1)

57 RMSE =
$$\left[\frac{1}{N}\sum_{i=1}^{N}(P_i - O_i)^2\right]^{\frac{1}{2}}$$
 (S2)

58
$$IOA = 1 - \frac{\sum_{i=1}^{N} (P_i - O_i)^2}{\sum_{i=1}^{N} (|P_i - \overline{O}| + |O_i - \overline{O}|)^2}$$
 (S3)

Where P_i and O_i are the simulated and observed variables, respectively. *N* is the total number of the simulations for comparisons, and \overline{O} donates the average of the observations. The IOA ranges from 0 to 1, with 1 showing a perfect agreement of the simulations with the observations.

62 S3 Model performance

63 S3.1 Air pollutants simulations in the NCP

64 Comparison of observed (black dots) and simulated (solid dark blue lines) near-surface hourly mass 65 concentrations of (a) $PM_{2.5}$, (b) O_3 , (c) NO_2 , (d) SO_2 , and (d) CO averaged at available monitoring sites 66 in the NCP from January 1 to January 30, 2014 is shown in Fig. S1. The model successfully reproduces 67 the diurnal variation of near-surface $PM_{2.5}$ concentrations in the NCP with an IOA of 0.92 and a slightly 68 overestimation with a MB of 3.8 μ g m⁻³. The model generally captures well the temporal variations of 69 near-surface O_3 concentrations compared to observations in the NCP with an IOA of 0.90 while a 67 generally overestimates the O_3 concentrations a MB of 0.6 μ g m⁻³. The model also reasonably well yields 71 the temporal variation of NO₂, SO₂ and CO compared with observation, with IOA and MB of 0.82 and -

72 4.0 μ g m⁻³, 0.72 and -13.2 μ g m⁻³, 0.85 and 0.0 μ g m⁻³, respectively.

73 The spatial pattern of calculated and observed average near-surface concentrations of PM2.5, SO2, 74 NO₂ and O₃ along with simulated winds in January 2014 in the NCP is shown in Fig. S2. The simulations 75 of four air pollutants distributions are general in good agreement with the observations in the NCP, while 76 partly biases of modeling still exist. It shows that the air in the NCP in January 2014 is much polluted 77 with the monthly near-surface $PM_{2.5}$ concentrations over 150 µg m⁻³. The observed and simulated 78 highest average near-surface PM2.5 concentrations are found in Beijing, Hebei, Henan, Shandong, north 79 Anhui and north Jiangsu. Highest observed and simulated near-surface SO2 and NO2 concentrations 80 almost occurs in same areas in the NCP. But simulated highest SO₂ concentrations are mainly 81 concentrated around cities, while the distribution of NO2 shows more area uniformly which likely due to 82 their sources are different, the former mainly emits from point sources and the latter mainly comes from 83 more area sources. The simulated O₃ concentrations are rather low in the NCP which is consistent with 84 measurements.



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Figure S1. Comparison of observed (black dots) and simulated (dark blue lines) diurnal profiles of nearsurface hourly mass concentrations of (a) PM_{2.5}, (b) O₃, (c) NO₂, (d) SO₂, and (d) CO averaged at monitoring sites in the NCP from January 1 to January 30, 2014. The light blue wavy lines represent error bars plotted using standard deviation.



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Figure S2. Pattern comparisons of simulated (color counters) vs. observed (colored circles) near-surface mass
 concentrations of (a) PM_{2.5}, (b) SO₂, (c) NO₂, and (d) O₃ averaged in January 2014. The black arrows indicate
 simulated surface winds.

Figure S3 provides the time series variations of simulated and observed aerosol species including OA (1.6 times of measurement OC), EC, ammonium, sulfate, and nitrate at Beijing and Tianjin city from January 1 to January 30, 2014. It shows that the WRF-Chem model generally predicts the temporal variations of the aerosol species against the field measurements reasonably with relatively high IOA value. The model yields the main peaks of aerosol species but with some frequently underestimates or overestimates which is mostly linked to the uncertainty of emission inventory and meteorological variations.

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- 102





Figure S3. Comparison of measured (black dots) and simulated (blue lines) daily profiles of submicron aerosol
 species of (a) OA, (b) EC, (c) ammonium, (d) sulfate, and (e) nitrate at two sites (Beijing and Tianjin) in the
 NCP from January 1 to January 30, 2014.

108 S3.2 Downward shortwave flux comparison

Figure S4 shows the comparison of measured (black dots) and simulated (blue lines) diurnal profiles of the SWDOWN reaching the ground surface in (a) Beijing, (b) Tianjin, (c) Zhengzhou, (d) Hefei, and (e) Ji'nan from 01 January 2014 to 30 January 2014. Although the MB and RMSE values suggest bias in the model performance, but in overall, the model generally captures the diurnal patterns quite well, as reflected by the average IOA values up to 0.95 across all five cities. The biases of SWDOWN between model and field study may be caused by the cloud cover and optical thickness calculation in the model, which is due to the horizontal resolution of the model is insufficient to resolve the cumulus clouds.



Figure S4. Comparison of measured (black dots) and simulated (blue lines) diurnal profiles
of the SWDOWN reaching the ground surface in (a) Beijing, (b) Tianjin, (c) Zhengzhou, (d)
Hefei, and (e) Ji'nan from January 1 to January 30, 2014.

126 S3.3 OA from different sources comparison in Beijing

Figure S5 presents a comparative analysis of temporal profiles of measured and simulated OA, POA from coal combustion (POA-COAL), biomass burning combustion (POA-BB), POA from vehicle exhaust (POA-TRA) and SOA in Beijing from January9 to 25, 2014. The model shows a good fit with

131despite some discrepancies in peak values and slightly overestimates as indicated by an RMSE of 33.1132 $\mu g/m^3$ and an MB of 5.0 $\mu g/m^3$, respectively. The model also generally tracks the measured diurnal133variations in POA-COAL mass concentrations, with an IOA of 0.81. The model frequently134underestimates or overestimates the POA-COAL mass concentrations and is also subject to missing the135observed POA-COAL peaks. The POA-COAL is mainly emitted from industries and residential coal136combustion. In general, the POA-COAL emissions from industries have clear diurnal variations but are137opposite for those from residential coal combustion, causing large model biases for the POA-COAL138simulation. The model performs well in capturing the general trend of POA-BB with an IOA of 0.86 and139a lower RMSE of 4.0 $\mu g/m^3$, while POA-Tra has a lower IOA of 0.56. Although the model captured the140major vehicle pollution events, some smaller peaks were not well reflected in the model. Modeled SOA141shows a fair correlation with observed data (IOA of 0.73) but also exhibits some of the higher variance142in peak concentrations, reflected in an RMSE of 11.3 $\mu g/m^3$. In general, the IOA values of all types of143OA suggest a reasonable model performance, particularly in capturing the temporal dynamics with some144quantitative in accuracies which largely associated with the influence of meteorological conditions and145emission sources uncertainties.	130	observed data with an IOA of 0.85, suggesting a reasonably accurate representation of OA variations,
variations in POA-COAL mass concentrations, with an IOA of 0.81. The model frequently underestimates or overestimates the POA-COAL mass concentrations and is also subject to missing the observed POA-COAL peaks. The POA-COAL is mainly emitted from industries and residential coal combustion. In general, the POA-COAL emissions from industries have clear diurnal variations but are opposite for those from residential coal combustion, causing large model biases for the POA-COAL simulation. The model performs well in capturing the general trend of POA-BB with an IOA of 0.86 and a lower RMSE of 4.0 μ g/m ³ , while POA-Tra has a lower IOA of 0.56. Although the model captured the major vehicle pollution events, some smaller peaks were not well reflected in the model. Modeled SOA shows a fair correlation with observed data (IOA of 0.73) but also exhibits some of the higher variance in peak concentrations, reflected in an RMSE of 11.3 μ g/m ³ . In general, the IOA values of all types of OA suggest a reasonable model performance, particularly in capturing the temporal dynamics with some quantitative in accuracies which largely associated with the influence of meteorological conditions and	131	despite some discrepancies in peak values and slightly overestimates as indicated by an RMSE of 33.1
134 underestimates or overestimates the POA-COAL mass concentrations and is also subject to missing the 135 observed POA-COAL peaks. The POA-COAL is mainly emitted from industries and residential coal 136 combustion. In general, the POA-COAL emissions from industries have clear diurnal variations but are 137 opposite for those from residential coal combustion, causing large model biases for the POA-COAL 138 simulation. The model performs well in capturing the general trend of POA-BB with an IOA of 0.86 and 139 a lower RMSE of 4.0 μ g/m ³ , while POA-Tra has a lower IOA of 0.56. Although the model captured the 140 major vehicle pollution events, some smaller peaks were not well reflected in the model. Modeled SOA 141 shows a fair correlation with observed data (IOA of 0.73) but also exhibits some of the higher variance 142 in peak concentrations, reflected in an RMSE of 11.3 μ g/m ³ . In general, the IOA values of all types of 143 OA suggest a reasonable model performance, particularly in capturing the temporal dynamics with some 144 quantitative in accuracies which largely associated with the influence of meteorological conditions and	132	μ g/m ³ and an MB of 5.0 μ g/m ³ , respectively. The model also generally tracks the measured diurnal
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137 opposite for those from residential coal combustion, causing large model biases for the POA-COAL 138 simulation. The model performs well in capturing the general trend of POA-BB with an IOA of 0.86 and 139 a lower RMSE of $4.0 \ \mu g/m^3$, while POA-Tra has a lower IOA of 0.56. Although the model captured the 140 major vehicle pollution events, some smaller peaks were not well reflected in the model. Modeled SOA 141 shows a fair correlation with observed data (IOA of 0.73) but also exhibits some of the higher variance 142 in peak concentrations, reflected in an RMSE of 11.3 $\mu g/m^3$. In general, the IOA values of all types of 143 OA suggest a reasonable model performance, particularly in capturing the temporal dynamics with some 144 quantitative in accuracies which largely associated with the influence of meteorological conditions and	135	observed POA-COAL peaks. The POA-COAL is mainly emitted from industries and residential coal
138 simulation. The model performs well in capturing the general trend of POA-BB with an IOA of 0.86 and 139 a lower RMSE of 4.0 μ g/m ³ , while POA-Tra has a lower IOA of 0.56. Although the model captured the 140 major vehicle pollution events, some smaller peaks were not well reflected in the model. Modeled SOA 141 shows a fair correlation with observed data (IOA of 0.73) but also exhibits some of the higher variance 142 in peak concentrations, reflected in an RMSE of 11.3 μ g/m ³ . In general, the IOA values of all types of 143 OA suggest a reasonable model performance, particularly in capturing the temporal dynamics with some 144 quantitative in accuracies which largely associated with the influence of meteorological conditions and	136	combustion. In general, the POA-COAL emissions from industries have clear diurnal variations but are
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144 quantitative in accuracies which largely associated with the influence of meteorological conditions and	142	in peak concentrations, reflected in an RMSE of 11.3 μ g/m ³ . In general, the IOA values of all types of
	143	OA suggest a reasonable model performance, particularly in capturing the temporal dynamics with some
145 emission sources uncertainties.	144	quantitative in accuracies which largely associated with the influence of meteorological conditions and
	145	emission sources uncertainties.



150 Figure S5. Temporal profiles of measured (black dots) and simulated (blue lines) OA (a), POA-Coal (b), POA-

151 BB (c), POA-Tra (d) and SOA (e) in Beijing from January 9 to 25, 2014.



154 Figure S6. Vertical profile of secondary BrC to primary BrC ratio in NCP in January, 2014
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156 **References**

- Binkowski, F. S. and Roselle, S. J.: Models-3 Community Multiscale Air Quality (CMAQ) model
 aerosol component 1. Model description, J. Geophys. Res., 108, 479,
 https://doi.org/10.1029/2001JD001409, 2003.
- 160 Chen, F. and Dudhia, J.: Coupling an Advanced Land Surface-Hydrology Model with the Penn State-
- 161 NCAR MM5 Modeling System. Part I: Model Implementation and Sensitivity, Mon. Wea. Rev., 129,
- 162 569–585, https://doi.org/10.1175/1520-0493(2001)129<0569:CAALSH>2.0.CO;2, 2001.
- 163 Chou, M. D. and Suarez, M. J.: A solar radiation parameterization for atmospheric studies, NASA
- 164 Technique Report, Greenbelt, USA, NASA/TM-1999-104606/VOL15, 1999.
- Chou, M. D., Suarez, M. J., Liang, X. Z., Yan, M. H., and Cote, C.: A Thermal Infrared Radiation
 Parameterization for Atmospheric Studies, NASA Technique Report, Greenbelt, USA, NASA/TM-
- 167 2001-104606/VOL19, 2001.
- Chow, J. C., Watson, J. G., Chen, L. W. A., Arnott, W. P., Moosmüller, H., and Fung, K.: Equivalence
 of elemental carbon by thermal/optical reflectance and transmittance with different temperature
 protocols, Environmental science & technology, 38, 4414–4422, https://doi.org/10.1021/es034936u,
 2004.
- 172 Elser, M., Huang, R.-j., Wolf, R., Slowik, J. G., Wang, Q., Canonaco, F., Li, G., Bozzetti, C.,
- Daellenbach, K. R., Huang, Y., Zhang, R., Li, Z., Cao, J., Baltensperger, U., El-Haddad, I., and Prévôt,
 A. S. H.: New insights into PM_{2.5} chemical composition and sources in two major cities in China during
 extreme haze events using aerosol mass spectrometry, Atmos. Chem. Phys., 16, 3207–3225,
 https://doi.org/10.5194/acp-16-3207-2016, 2016.
- 177 Feng, T., Li, G., Cao, J., Bei, N., Shen, Z., Zhou, W., Liu, S., Zhang, T., Wang, Y., Huang, R.-j., Tie, X.,
- and Molina, L. T.: Simulations of organic aerosol concentrations during springtime in the Guanzhong
 Basin, China, Atmos. Chem. Phys., 16, 10045–10061, https://doi.org/10.5194/acp-16-10045-2016,
 2016.
- 181 Grell, G. A., Peckham, S. E., Schmitz, R., McKeen, S. A., Frost, G., Skamarock, W. C., and Eder, B.:
- 182 Fully coupled "online" chemistry within the WRF model, Atmospheric Environment, 39, 6957–6975,
- 183 https://doi.org/10.1016/j.atmosenv.2005.04.027, 2005.
- 184 Guenther, A., Karl, T., Harley, P., Wiedinmyer, C., Palmer, P. I., and Geron, C.: Estimates of global
- 185 terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from Nature),
- 186 Atmos. Chem. Phys., 6, 3181–3210, https://doi.org/10.5194/acp-6-3181-2006, 2006.
- 187 Hess, P. G., Flocke, S., Lamarque, J.-F., Barth, M. C., and Madronich, S.: Episodic modeling of the
- 188 chemical structure of the troposphere as revealed during the spring MLOPEX 2 intensive, J. Geophys.
- 189 Res., 105, 26809–26839, https://doi.org/10.1029/2000JD900253, 2000.
- Hong, S. Y. and Lim, J.O.J.: The WRF single-moment 6-class microphysics scheme (WSM6), Asia
 Pacific Journal of Atmospheric Sciences, 42, 129–151, 2006.
- 192 Horowitz, L. W., Walters, S., Mauzerall, D. L., Emmons, L. K., Rasch, P. J., Granier, C., Tie, X.,
- 193 Lamarque, J.-F., Schultz, M. G., Tyndall, G. S., Orlando, J. J., and Brasseur, G. P.: A global simulation
- 194 of tropospheric ozone and related tracers: Description and evaluation of MOZART, version 2, J.
- 195 Geophys. Res., 108, n/a-n/a, https://doi.org/10.1029/2002JD002853, 2003.

- 196 Janjić, Z. I.: Nonsingular Implementation of the Mellor-Yamada Level 2.5 Scheme in the NCEP Meso
- 197 Model, 437, Ncep Office Note, Camp Springs, USA, 2002.
- 198 Li, G., Lei, W., Bei, N., and Molina, L. T.: Contribution of garbage burning to chloride and PM_{2.5} in
- 199 Mexico City, Atmos. Chem. Phys., 12, 8751–8761, https://doi.org/10.5194/acp-12-8751-2012, 2012.
- 200 Li, G., Bei, N., Tie, X., and Molina, L. T.: Aerosol effects on the photochemistry in Mexico City during
- 201 MCMA-2006/MILAGRO campaign, Atmos. Chem. Phys., 11, 5169–5182, https://doi.org/10.5194/acp-
- 202 11-5169-2011, 2011a.
- 203 Li, G., Zavala, M., Lei, W., Tsimpidi, A. P., Karydis, V. A., Pandis, S. N., Canagaratna, M. R., and
- 204 Molina, L. T.: Simulations of organic aerosol concentrations in Mexico City using the WRF-CHEM
- 205 model during the MCMA-2006/MILAGRO campaign, Atmos. Chem. Phys., 11, 3789–3809,
 206 https://doi.org/10.5194/acp-11-3789-2011, 2011b.
- 207 Li, G., Lei, W., Zavala, M., Volkamer, R., Dusanter, S., Stevens, P., and Molina, L. T.: Impacts of HONO
- 208 sources on the photochemistry in Mexico City during the MCMA-2006/MILAGO Campaign, Atmos.

209 Chem. Phys., 10, 6551–6567, https://doi.org/10.5194/acp-10-6551-2010, 2010.

- Li, G., Zhang, R., Fan, J., and Tie, X.: Impacts of black carbon aerosol on photolysis and ozone, J.
 Geophys. Res., 110, 1042, https://doi.org/10.1029/2005JD005898, 2005.
- Li, M., Liu, H., Geng, G., Hong, C., Liu, F., Song, Y., Tong, D., Zheng, B., Cui, H., Man, H., Zhang,
- Q., and He, K.: Anthropogenic emission inventories in China: a review, National Science Review, 4,
 834–866, https://doi.org/10.1093/nsr/nwx150, 2017.
- Li, X., Wu, J., Elser, M., Feng, T., Cao, J., El-Haddad, I., Huang, R., Tie, X., Prévôt, A. S. H., and Li,
- 216 G.: Contributions of residential coal combustion to the air quality in Beijing–Tianjin–Hebei (BTH),
- 217 China: a case study, Atmos. Chem. Phys., 18, 10675–10691, https://doi.org/10.5194/acp-18-10675218 2018, 2018.
- Liggio, J.: Reactive uptake of glyoxal by particulate matter, J. Geophys. Res., 110, 881,
 https://doi.org/10.1029/2004jd005113, 2005.
- 221 Nenes, A., Pandis, S. N., and Pilinis Christodoulos: ISORROPIA: A New Thermodynamic Equilibrium
- 222 Model for Multiphase Multicomponent Inorganic Aerosols, Aquatic Geochemistry, 4, 123–152, 1998.
- 223 Robinson, A. L., Donahue, N. M., Shrivastava, M. K., Weitkamp, E. A., Sage, A. M., Grieshop, A. P.,
- 224Lane, T. E., Pierce, J. R., and Pandis, S. N.: Rethinking organic aerosols: semivolatile emissions and225photochemical aging, Science (New York, N.Y.), 315, 1259–1262,
- 226 https://doi.org/10.1126/science.1133061, 2007.
- 227 Shrivastava, M. K., Lane, T. E., Donahue, N. M., Pandis, S. N., and Robinson, A. L.: Effects of gas
- 228 particle partitioning and aging of primary emissions on urban and regional organic aerosol
- 229 concentrations, J. Geophys. Res., 113, 2701, https://doi.org/10.1029/2007jd009735, 2008.
- Tie, X.: Effect of clouds on photolysis and oxidants in the troposphere, J. Geophys. Res., 108, 23,073,
 https://doi.org/10.1029/2003JD003659, 2003.
- 232 Volkamer, R., San Martini, F., Molina, L. T., Salcedo, D., Jimenez, J. L., and Molina, M. J.: A missing
- sink for gas-phase glyoxal in Mexico City: Formation of secondary organic aerosol, Geophys. Res. Lett.,
- 234 34, 641, https://doi.org/10.1029/2007GL030752, 2007.
- 235 Wesely, M. L.: Parameterization of surface resistances to gaseous dry deposition in regional-scale

- numerical models, Atmospheric Environment (1967), 23, 1293–1304, https://doi.org/10.1016/00046981(89)90153-4, 1989.
- 238 Xing, L., Wu, J., Elser, M., Tong, S., Liu, S., Li, X., Liu, L., Cao, J., Zhou, J., El-Haddad, I., Huang, R.,
- 239 Ge, M., Tie, X., Prévôt, A. S. H., and Li, G.: Wintertime secondary organic aerosol formation in Beijing-
- 240 Tianjin-Hebei (BTH): Contributions of HONO sources and heterogeneous reactions, Atmos. Chem.
- 241 Phys., 19, 2343–2359, https://doi.org/10.5194/acp-19-2343-2019, 2019.
- 242 Zhang, Q., Streets, D. G., Carmichael, G. R., He, K. B., Huo, H., Kannari, A., Klimont, Z., Park, I. S.,
- 243 Reddy, S., Fu, J. S., Chen, D., Duan, L., Lei, Y., Wang, L. T., and Yao, Z. L.: Asian emissions in 2006
- for the NASA INTEX-B mission, Atmos. Chem. Phys., 9, 5131–5153, https://doi.org/10.5194/acp-95131-2009, 2009.
- 246 Zhang, T., Cao, J. J., Tie, X. X., Shen, Z. X., Liu, S. X., Ding, H., Han, Y. M., Wang, G. H., Ho, K. F.,
- 247 Qiang, J., and Li, W. T.: Water-soluble ions in atmospheric aerosols measured in Xi'an, China: Seasonal
- 248 variations and sources, Atmospheric Research, 102, 110-119,
- 249 https://doi.org/10.1016/j.atmosres.2011.06.014, 2011.
- 250 Zhao, J., Levitt, N. P., Zhang, R., and Chen, J.: Heterogeneous reactions of methylglyoxal in acidic
- 251 media: Implications for secondary organic aerosol formation, Environmental science & technology, 40,
- 252 7682–7687, https://doi.org/10.1021/es060610k, 2006.