



Modeling atmospheric sulfate oxidation chemistry via the oxygen isotope anomaly using the Community Multiscale Air Quality Model (CMAQ)

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Abstract. Atmospheric sulfate formation influences climate and air quality, yet its chemical pathways remain difficult to constrain. This study utilizes the oxygen isotope anomaly ($\Delta^{17}\text{O}$) of sulfate aerosol (ASO_4) as a tracer to distinguish formation processes. This work presents a simulation of $\Delta^{17}\text{O}(\text{ASO}_4)$ within the contiguous United States, conducted over full annual cycles, which enables the quantification of seasonal and spatial patterns of sulfate oxidation pathways and their response to major emission reductions, for the first time at this scale and temporal coverage. In 2019, $\Delta^{17}\text{O}(\text{ASO}_4)$ values were predicted to be below 1‰ in the Gulf Coast, indicating acidic, ASO_4 -rich conditions dominated by $\text{S(IV)} + \text{H}_2\text{O}_2$ oxidation, while values above 2‰ in the West suggested less acidic conditions, leading to enhanced ASO_4 production via $\text{S(IV)} + \text{O}_3$ oxidation. Peak $\Delta^{17}\text{O}(\text{ASO}_4)$ values of $\sim 4.5\text{‰}$ in April across the Western US reflected O_3 -driven ASO_4 formation during high ammonia (NH_3) emissions from fertilization. Between 2006 and 2019, mean $\Delta^{17}\text{O}(\text{ASO}_4)$ was predicted to increase by up to 2‰, driven by declining sulfur dioxide (SO_2) emissions from regulatory measures. Model comparisons with historical measurements show reasonable agreement in the acidic southeastern US (RMSE = 0.20‰, Baton Rouge, LA). However, the model overpredicts $\Delta^{17}\text{O}(\text{ASO}_4)$ in the Western US with RMSE values of 0.36‰ (La Jolla, CA) and 1.9‰ (White Mountain Research Center, CA). This overestimation suggests an excessive model response to aqueous $\text{S(IV)} + \text{O}_3$ reactions. These findings underscore the diagnostic potential of $\Delta^{17}\text{O}(\text{ASO}_4)$ for assessing sulfate formation mechanisms and pinpointing shortcomings in chemical transport models. However, $\Delta^{17}\text{O}(\text{ASO}_4)$ observations across the United States remain exceedingly limited, with most available data dating back to the late 1990s and early 2000s, highlighting the need for renewed measurement efforts.

1 Introduction

Atmospheric sulfate (SO_4^{2-}) plays a critical role in climate and air quality. As a major component of aerosols, SO_4^{2-} influences aerosol pH, atmospheric chemistry, and precipitation acidity (Calvo et al., 2013; Weber et al., 2016). SO_4^{2-} aerosols (ASO_4) significantly contribute to radiative forcing by scattering sunlight and serving as cloud condensation nuclei, which impacts cloud properties and the Earth's radiation balance (Lohmann and Feichter, 1997; Jones et al., 1994; Kaufman and Tanré, 1994). The anthropogenic influ-

ence on the ASO_4 budget, primarily from fossil fuel combustion, has been widely documented, contributing to regional and global climate effects (Langner et al., 1992; Smith et al., 2011). The presence of ASO_4 alters cloud albedo and lifetime, affecting regional and global climate patterns through indirect radiative forcing (Jones et al., 1994; Haywood and Boucher, 2000). Additionally, the health impacts of ASO_4 -containing particles underscore their importance in air quality management (Reiss et al., 2007). The formation of ASO_4 is influenced by complex interactions with secondary organic aerosols (SOA) and other atmospheric components.

Emerging research highlights the significant role of highly oxygenated organic molecules (HOMs) in enhancing ASO₄ formation under humid conditions (Hallquist et al., 2009; Bianchi et al., 2019). These interactions highlight the complex connections between ASO₄, atmospheric chemistry, and climate dynamics. Despite a 70 % reduction in ASO₄ concentrations over the past 15 years, aerosol acidity has remained high, primarily due to the buffering effect of ammonia partitioning between the gas and particle phases (Weber et al., 2016). This persistent acidity impacts both air quality and health, as it enhances the solubility of harmful metals and promotes acid-catalyzed chemical reactions in the atmosphere.

Despite their significance, atmospheric chemistry models often face significant challenges in accurately reproducing ASO₄ concentrations, potentially due to uncertainties surrounding ASO₄ formation mechanisms (Harris et al., 2013; Li et al., 2020; Vannucci et al., 2024). ASO₄ can originate from both primary emissions and secondary formation. Primary sources include natural emissions, such as sea salt, volcanic eruptions, and soil dust (Alexander et al., 2005; Arimoto et al., 2001; Savarino et al., 2003), as well as anthropogenic emissions from fossil fuel combustion (Langner et al., 1992; Smith et al., 2011; Sofen et al., 2011). Secondary ASO₄ formation involves complex oxidation processes that occur in both the gas phase and the aqueous phase. In the gas phase, sulfur dioxide (SO₂) is oxidized by hydroxyl radicals ($\cdot\text{OH}$), producing sulfuric acid (H₂SO₄). This sulfuric acid can either condense to form new particles or add mass to existing aerosols. The rate of this process is highly dependent on environmental conditions such as temperature and pH, which introduces significant uncertainties in predicting ASO₄ concentrations (Seigneur and Saxena, 1988). For instance, Vannucci et al. (2024) demonstrated that temperature plays a crucial role in modulating ASO₄ aerosol concentrations, particularly during summertime pollution episodes, where aerosol composition and temperature sensitivity can significantly impact model accuracy. Aqueous-phase ASO₄ formation occurs when dissolved sulfur species (S(IV) = SO₂·H₂O + HSO₃[−] + SO₃^{2−}) are oxidized by molecules including ozone (O₃), hydrogen peroxide (H₂O₂), and oxygen, catalyzed by transition metal ions (TMI) (e.g., Fe³⁺ and Mn²⁺). The role of other oxidants, such as hypohalous acids (HOX, X = Br and Cl), is increasingly recognized, particularly in marine boundary layers (Chen et al., 2016; Ishino et al., 2017). Chen et al. (2016) highlighted the significant contribution of HOX in ASO₄ formation in the remote marine boundary layer, estimating that 33 %–50 % of ASO₄ is produced via this pathway. This suggests that HOX may play a larger role in ASO₄ formation than previously recognized. Additionally, aqueous oxidation of S(IV) induced by nitrogen dioxide (NO₂) has also been proposed as a potential pathway, particularly under polluted and low-oxidant wintertime conditions (Sarwar et al., 2013). Although generally less important than H₂O₂ and O₃ ox-

idation, this pathway may contribute to ASO₄ formation in specific environments and conditions.

Sensitivity analyses have shown that the rate of aqueous-phase ASO₄ formation is particularly influenced by pH, oxidant availability, and environmental conditions, further complicating ASO₄ modeling (Pandis and Seinfeld, 1989). Harris et al. (2013) showed that TMI-catalyzed oxidation can dominate under specific conditions, particularly in the presence of coarse dust particles, significantly altering ASO₄ formation rates in cloud droplets. Similarly, Li et al. (2020) highlighted the critical role of TMI-driven SO₂ oxidation during haze episodes, where such pathways can account for up to 50 % of ASO₄ production under polluted conditions. Heterogeneous reactions on aerosol surfaces may also play a critical role in ASO₄ formation (Harris et al., 2013). These surface reactions involve the interaction of gaseous sulfur species with aerosols, significantly influencing ASO₄ formation and elevating the complexity of predicting ASO₄ concentrations. Meidan et al. (2019) emphasized the importance of Criegee intermediates (CIs) in ASO₄ formation, particularly in nocturnal power plant plumes, where SO₂ is oxidized under conditions with minimal photochemical activity. This study revealed that CIs could account for a significant portion of ASO₄ production in the absence of sunlight. Additionally, Liu et al. (2019) examined the role of stabilized Criegee intermediates (sCIs) in ASO₄ formation in the Beijing-Tianjin-Hebei region, showing that under certain atmospheric conditions, sCI-driven SO₂ oxidation can contribute substantially to secondary ASO₄ production, adding another layer of complexity to ASO₄ formation models. These interactions highlight the challenges in modeling ASO₄ concentrations, as heterogeneous reactions, TMIs, and Criegee intermediates all contribute to the uncertainty in atmospheric ASO₄ predictions.

The use of oxygen isotope mass-independent fractionation ($\Delta^{17}\text{O} = \delta^{17}\text{O} - 0.52 \times \delta^{18}\text{O}$) has emerged as a promising tool to explore atmospheric ASO₄ formation pathways (Alexander et al., 2004; Barkan and Luz, 2003; Kaiser et al., 2004; Michalski et al., 2003; Morin et al., 2007; Savarino et al., 2007; Walters et al., 2019; Weston, 2006). This isotopic indicator is crucial for tracking ASO₄ formation, providing a refined tool for model evaluation and prediction. This is because $\Delta^{17}\text{O}$ has distinct values associated with different oxidation processes, making it a powerful tool in understanding ASO₄ production mechanisms. The dominant source of $\Delta^{17}\text{O}$ in the lower atmosphere derives from O₃ formation. The average $\Delta^{17}\text{O}(\text{O}_3)$ near the surface is approximately 26 ‰ (Vicars and Savarino, 2014). This contrasts with other tropospheric oxidants, which have $\Delta^{17}\text{O}$ values near 0 ‰. Hydrogen peroxide (H₂O₂) has a $\Delta^{17}\text{O}$ value of about 1.6 ‰ due to the influence of O₃ involved in H₂O₂ formation (Savarino and Thiemens, 1999). Laboratory studies have shown that oxidants will proportionally transfer their $\Delta^{17}\text{O}$ values into the ASO₄ product. Table 1 summarizes the $\Delta^{17}\text{O}$ ranges associated with major tropospheric ASO₄ production

pathways based on oxygen isotopic mass balance (Alexander et al., 2005, 2009; Ishino et al., 2017; Savarino et al., 2000; Walters et al., 2019). The gas-phase oxidation of SO₂ by OH and metal-catalyzed O₂ oxidation yields $\Delta^{17}\text{O}(\text{ASO}_4)$ values near 0‰, indicating a negligible transfer of the $\Delta^{17}\text{O}$ signature. Similarly, aqueous-phase oxidation of SO₂ by hypohalous acids (HOX) results in $\Delta^{17}\text{O}(\text{ASO}_4)$ values around 0‰. In contrast, aqueous-phase oxidation involving H₂O₂ and O₃ exhibits significantly higher $\Delta^{17}\text{O}$ values. H₂O₂ oxidation produces $\Delta^{17}\text{O}(\text{ASO}_4)$ values around 0.8‰, while O₃ oxidation results in $\Delta^{17}\text{O}(\text{ASO}_4)$ values of about 6.5‰. These distinctions enable the ability to track ASO₄ formation.

Previous studies have utilized $\Delta^{17}\text{O}(\text{ASO}_4)$ observations to evaluate the impact of anthropogenic emissions on ASO₄ production routes. In polluted regions, anthropogenic emissions of metals such as Fe³⁺ and Mn²⁺ enhance O₂-catalyzed ASO₄ formation, particularly in the Northern Hemisphere during winter. This metal-catalyzed ASO₄ formation can suppress ASO₄ production via O₃ and H₂O₂ pathways, impacting $\Delta^{17}\text{O}(\text{ASO}_4)$ values and complicating model predictions (Savarino et al., 2000). Furthermore, ship emissions, which have been underrepresented in atmospheric models, significantly contribute to ASO₄ source in marine environments. Triple-oxygen isotope measurements suggest these emissions play a larger role in ASO₄ production than previously recognized, with implications for air quality and climate modeling (Dominguez et al., 2008). To fully utilize the diagnostic potential of $\Delta^{17}\text{O}(\text{ASO}_4)$, a comprehensive model framework is essential for interpreting ASO₄ formation. Previous models, such as GEOS-Chem, have incorporated $\Delta^{17}\text{O}$ tracking to investigate ASO₄ formation pathways, highlighting the growing importance of metal-catalyzed O₂ oxidation in polluted regions, which surpasses the traditional O₃ and H₂O₂ pathways (Sofen et al., 2011). Despite rising tropospheric O₃ levels since preindustrial times, $\Delta^{17}\text{O}(\text{ASO}_4)$ values in the Arctic have declined due to enhanced metal-catalyzed ASO₄ formation. Recent studies have applied $\Delta^{17}\text{O}$ of ASO₄ in chemical transport models to explore long-term changes and regional processes, including GEOS-Chem simulations coupled with ice core observations (Hattori et al., 2021; Peng et al., 2023) and CMAQ applications in East Asia (Itahashi et al., 2022; Lin et al., 2025). These works highlight the diagnostic potential of $\Delta^{17}\text{O}$ across diverse regions and timescales. Building upon these advances, our study presents the first CMAQ simulations of $\Delta^{17}\text{O}(\text{ASO}_4)$ within the contiguous United States over full annual cycles for 2006 and 2019, allowing for the assessment of seasonal and spatial patterns of ASO₄ oxidation pathways in response to emission reductions.

2 Methods

2.1 Model Description and EQUATES 2019 Dataset

This study utilizes the CMAQ (Community Multiscale Air Quality) version 5.4 model to simulate ASO₄ formation and its $\Delta^{17}\text{O}$ values across the contiguous United States (CONUS). The CMAQ model is configured with the cb6r5_ae7_aq chemical mechanism, which stands for Carbon Bond 6 revision 5, with aerosol 7 for standard cloud chemistry (Yarwood et al., 2010). This mechanism encompasses both gas-phase and aqueous-phase oxidation processes of SO₂, which are essential for accurately modeling ASO₄ formation. Specifically, it involves the oxidation of SO₂ by •OH in the gas phase and by H₂O₂ and O₃ in cloud droplets and aqueous environments. Cloud water pH in CMAQ is calculated dynamically within the default cloud chemistry module, which is based on the work of Walcek and Taylor (1986) and assumes instantaneous equilibrium among gas, aqueous, and ionic species. The pH is determined by the charge balance between dissolved acidic and basic ions. As S(IV) is oxidized to S(VI) and additional species are scavenged from interstitial aerosols, the pH evolves dynamically throughout cloud processing. The resulting pH fields respond to emissions and meteorological variability, directly governing the relative importance of the H₂O₂ and O₃ oxidation pathways for aqueous S(IV) oxidation. Previous evaluations have demonstrated that CMAQ accurately reproduces observed cloud droplet acidity, with differences generally within 0.5 pH units across multiple sites in the United States (Pye et al., 2020). The model's ability to capture these complex interactions facilitates a detailed assessment of ASO₄ dynamics under various atmospheric conditions (Appel et al., 2021).

The CMAQ simulations are based on the EQUATES (EPA's Air Quality Time Series Project) dataset, which provides a comprehensive and high-resolution emissions inventory derived from the 2017 National Emissions Inventory (NEI) (Benish et al., 2022; Foley et al., 2023). This dataset spans over two decades and provides detailed information on both natural and anthropogenic emissions, including those from industrial sources, vehicular traffic, power plants, and wildfires. It also accounts for seasonal and regional variations in emissions, enhancing the model's accuracy. The EQUATES 2019 dataset supplies critical inputs for CMAQ simulations, including emissions data, meteorological variables, as well as boundary and initial conditions, capturing pollutant variability across different seasons and regions.

Meteorological inputs for the CMAQ simulations were integrated from the Weather Research and Forecasting (WRF) model version 4.1.1. This integration provides detailed representations of temperature, wind speed, relative humidity, cloud cover, and precipitation rates. These meteorological factors influence cloud formation, pollutant dispersion, and oxidation processes. Boundary and initial conditions for the

CMAQ model were sourced from EQUATES to ensure accurate representation of pollutant inflows and outflows at the edges of the modeling domain. The initial conditions were established through a spin-up period starting on 15 December 2018, providing accurate starting concentrations for the 2019 simulation period. The CMAQ simulations were conducted at a resolution of 12×12 km over the CONUS domain using the Hyperion high-performance computing cluster at the University of South Carolina. This advanced computing infrastructure enabled the processing of large datasets and the execution of complex simulations necessary for this study.

2.2 Implementation of the Sulfur Tracking Mechanism (STM)

The Sulfur Tracking Mechanism (STM), utilized in the CMAQ model, provides a detailed analysis of ASO₄ formation pathways in the atmosphere (Appel et al., 2021). It distinguishes between various aqueous-phase and gas-phase formation processes and assesses contributions from emissions, initial conditions, and boundary conditions, offering valuable insights into the roles of these factors in overall ASO₄ production (Table 2). The sulfur budget comprises 14 ASO₄ species (AE) and 1 nonreactive ASO₄ species (NR), as documented in the CMAQ repository (<https://github.com/USEPA/CMAQ/blob/main/CCTM/src/MECHS/README.md>, last access: 1 March 2025). The STM output includes hourly simulations of the 15 tagged ASO₄ species across the model domain, which were then aggregated into monthly averages to analyze spatial and temporal variations in ASO₄ production. STM allows for an efficient way for the model to distinguish the contributions of different chemical pathways and emission contributions to ASO₄. This approach also enables a seamless calculation of $\Delta^{17}\text{O}$ of ASO₄.

A known bookkeeping bug in the STM implementation in CMAQ v5.4 resulted in the systematic underestimation of ASO₄ formed via the gas-phase SO₂ + OH pathway, despite the pathway being chemically active in the model. This issue has been documented by the CMAQ development team ([https://github.com/USEPA/CMAQ/wiki/CMAQ-Release-Notes:-Process-Analysis-&-Sulfur-Tracking-Model-\(STM\)](https://github.com/USEPA/CMAQ/wiki/CMAQ-Release-Notes:-Process-Analysis-&-Sulfur-Tracking-Model-(STM)), last access: 3 October 2025; U.S. Environmental Protection Agency, 2025) and has since been corrected in version 5.5. In our study, we fixed the issue by adjusting the call order of STM update routines in the sciproc.F module (matching the v5.5 fix) and ran the simulations using the corrected code. The updated STM diagnostic module used in this work is publicly available for reproducibility at: <https://doi.org/10.5281/zenodo.14954960> (Fang, 2025).

2.3 Calculation and Analysis of $\Delta^{17}\text{O}(\text{ASO}_4)$

The fractional contributions of each pathway, obtained from the STM, are used to calculate $\Delta^{17}\text{O}(\text{ASO}_4)$ across different grid cells.

$$f_i(\text{lat, lon, height, time}) = \frac{X_i}{\sum_{i=1}^n X_i} \quad (1)$$

where f_i represents the fractional contribution of pathway i ; X_i is the amount of ASO₄ produced by pathway i ; and $\sum_{i=1}^n X_i$ is the total Aitken mode and accumulation mode ASO₄ produced by all pathways except initial conditions and boundary conditions in each grid cell.

The gas-phase oxidation of SO₂ by $\cdot\text{OH}$ radical results in ASO₄ with no significant $\Delta^{17}\text{O}$ enrichment ($\sim 0\text{‰}$). ASO₄ formed via aqueous-phase oxidation by O₃ has a $\Delta^{17}\text{O}$ value of $\sim 6.5\text{‰}$, indicating significant cloud chemistry processes. ASO₄ formed via aqueous-phase oxidation by H₂O₂ has a $\Delta^{17}\text{O}$ value of $\sim 0.8\text{‰}$. Metal-catalyzed oxidation of SO₂ by O₂ in metal-rich environments results in a $\Delta^{17}\text{O}$ value of $\sim 0\text{‰}$ and does not exhibit a transfer of mass-independent fractionation signature. Although previous studies reported slightly negative $\Delta^{17}\text{O}$ values (-0.1‰ ; Hattori et al., 2021; Itahashi et al., 2022), this pathway contributes less than 10 % to total ASO₄ formation in our simulations, leading to a negligible ($< 0.01\text{‰}$) effect on the modeled $\Delta^{17}\text{O}(\text{ASO}_4)$. Therefore, it is approximated as 0‰ in this study. Heterogeneous reactions, such as those involving organic peroxides on aerosol surfaces, contribute to ASO₄ formation and are expected to have a $\Delta^{17}\text{O} \sim 0\text{‰}$. Although the fractional contributions (f_i) include all ASO₄ formation pathways diagnosed by the Sulfur Tracking Mechanism (STM), only H₂O₂ and O₃ carry non-zero $\Delta^{17}\text{O}$ signatures, all other pathways are assigned $\Delta^{17}\text{O} \approx 0\text{‰}$. Therefore, the $\Delta^{17}\text{O}(\text{ASO}_4)$ is calculated using the following equation:

$$\Delta^{17}\text{O}(\text{ASO}_4) = f_{\text{ASO}_4\text{AQH}_2\text{O}_2\text{J}} \times 0.8\text{‰} + f_{\text{ASO}_4\text{AQO}_3\text{J}} \times 6.5\text{‰} \quad (2)$$

where ASO₄AQH₂O₂J represents ASO₄ formed through the oxidation of SO₂ by H₂O₂; ASO₄AQO₃J represents ASO₄ formed through oxidation by O₃; the constants 0.8‰ and 6.5‰ correspond to the characteristic $\Delta^{17}\text{O}$ values for each pathway.

3 Results and Discussion

3.1 Predicted Fractional ASO₄ Formation and $\Delta^{17}\text{O}(\text{ASO}_4)$ in the Contiguous US in 2019

ASO₄ production in the contiguous United States arises from a combination of primary emissions and secondary formation pathways, the latter being dominated by H₂O₂- and

Table 1. Major ASO₄ formation pathways and their associated $\Delta^{17}\text{O}$ signatures. The pathways that are included in the CMAQ model using the cb6r5-ae7-aq mechanism are indicated.

Pathway	Reaction	$\Delta^{17}\text{O}$ (‰)	Notes	Included in CMAQ (cb6r5-ae7)
Gas-Phase	$\text{SO}_2 + \bullet\text{OH} \rightarrow \text{SO}_4^{2-} + \text{HO}_2$	~ 0	Dominant in photochemically active regions; negligible $\Delta^{17}\text{O}$ signature	Yes
Gas-Phase	$\text{SO}_2 + \text{sCI (Stabilized Criegee Intermediates)} \rightarrow \text{SO}_4^{2-}$	~ 0	Can enhance secondary ASO ₄ formation in regions influenced by biogenic VOC emissions; negligible $\Delta^{17}\text{O}$ signature	NO
Aqueous-Phase	$\text{HSO}_3^- + \text{H}_2\text{O}_2 \rightarrow \text{SO}_4^{2-} + \text{H}_2\text{O}$	0.8	Lower $\Delta^{17}\text{O}$ value, dominant under humid/cloudy conditions.	Yes
Aqueous-Phase	$\text{SO}_3^{2-} + \text{O}_3 \rightarrow \text{SO}_4^{2-} + \text{O}_2$	6.5	Higher $\Delta^{17}\text{O}$ value, significant for cloud chemistry.	Yes
Aqueous-Phase	$\text{SO}_3^{2-} + \text{O}_2$ (TMI = Transition Metal Ions, e.g., Fe^{3+} and Mn^{2+}) $\rightarrow \text{SO}_4^{2-}$	~ 0	Important in metal-rich aerosols; negligible $\Delta^{17}\text{O}$ signature	Yes
Aqueous-Phase	$\text{SO}_3^{2-} + \text{NO}_2 \rightarrow \text{SO}_4^{2-} + \text{NO}$	~ 0	Relevant under polluted and low-oxidant wintertime conditions; negligible $\Delta^{17}\text{O}$ signature	No
Aqueous-Phase	$\text{SO}_3^{2-} + \text{HOX (X = Br, Cl)} \rightarrow \text{SO}_4^{2-}$	~ 0	Dominant in marine environments with halogen chemistry; negligible $\Delta^{17}\text{O}$ signature	No
Heterogeneous	SO_2 (surface) + Organic peroxides $\rightarrow \text{SO}_4^{2-}$	~ 0	Significant role in submicron aerosol ASO ₄ formation; negligible $\Delta^{17}\text{O}$ signature.	No
Heterogeneous	SO_2 (surface) + H_2O_2 or O_3 on aerosols $\rightarrow \text{SO}_4^{2-}$	0.8–6.5	Highly variable; depends on aerosol composition and environmental conditions.	No

O₃-driven aqueous S(IV) oxidation and gas-phase oxidation of SO₂ via $\bullet\text{OH}$ (Fig. 1). These secondary reactions occur within cloud water, where SO₂ is oxidized by H₂O₂, O₃, and by O₂ (catalyzed by TMI). Compared to these dominant pathways, the TMI-catalyzed oxidation and reactions involving organic peroxides, such as methyl hydrogen peroxide (MHP) and peroxyacetic acid (PAA), have a minimal impact on ASO₄ production. While primary emissions contribute little overall, they exhibit localized hotspots in certain regions.

The fractional contributions of ASO₄ formation pathways demonstrate distinct spatial patterns that align with the predicted $\Delta^{17}\text{O}(\text{ASO}_4)$ variability. The H₂O₂ pathway ($f_{\text{S(IV)}+\text{H}_2\text{O}_2}$) is the most dominant, accounting for $35.4 \pm 14.0\%$ of the ASO₄ formation across the domain (Fig. 1).

This pathway is particularly influential in the Gulf Coast States, where abundant cloud cover and water vapor, acidic conditions (cloud pH < 6), and high concentrations of H₂O₂ (Figs. S1, S2 in the Supplement) support the oxidation of S(IV) in cloud droplets. The highest $f_{\text{S(IV)}+\text{H}_2\text{O}_2}$ in these regions contribute to the low $\Delta^{17}\text{O}(\text{ASO}_4)$ values below 1‰, due to the relatively lighter $\Delta^{17}\text{O}(\text{ASO}_4)$ resulted from the H₂O₂ pathway (0.8‰). Gas-phase oxidation of SO₂ by $\bullet\text{OH}$ ($f_{\text{SO}_2+\text{OH}}$) contributes to $34.4 \pm 9.5\%$ of the ASO₄ production across the domain (Fig. 1), exhibiting clear seasonal variability under photochemically active conditions, with the highest contributions occurring in summer (up to $\sim 75\%$) and lowest in winter (< 25%) (Fig. 5). The O₃ pathway ($f_{\text{S(IV)}+\text{O}_3}$) is the third most significant, contributing approximately $18.7 \pm 5.2\%$ to the ASO₄ formation across the do-

Table 2. Overview of the ASO₄ species in the Sulfur Tracking Mechanism (STM) incorporated into CMAQ.

Name	Group	Mode	Pathway
ASO4GASI	AE	Aitken	condensation of gas-phase reaction with OH
ASO4EMISI	AE	Aitken	source emission
ASO4ICBCI	AE	Aitken	initial conditions and boundary conditions
ASO4AQH2O2J	AE	Accumulation	H ₂ O ₂
ASO4AQO3J	AE	Accumulation	O ₃
ASO4AQFEMNJ	AE	Accumulation	O ₂ catalyzed by Fe ³⁺ and Mn ²⁺
ASO4AQMHPJ	AE	Accumulation	methyl hydrogen peroxide (MHP)
ASO4AQPAAJ	AE	Accumulation	peroxyacetic acid (PAA)
ASO4GASJ	AE	Accumulation	condensation of gas-phase reaction with OH
ASO4EMISJ	AE	Accumulation	source emission
ASO4ICBCJ	AE	Accumulation	initial conditions and boundary conditions
ASO4GASK	AE	Coarse	condensation of gas-phase reaction with OH
ASO4EMISK	AE	Coarse	source emission
ASO4ICBCK	AE	Coarse	initial conditions and boundary conditions
SULF_ICBC	NR	N/A	sulfuric acid vapor (SULF) from initial conditions and boundary conditions

N/A: Not applicable

main (Fig. 1). The highest $f_{S(IV)+O_3}$ occurs in the Western States, due to the high O₃ concentration and high cloud pH (Fig. S1), which facilitates the aqueous oxidation of S(IV) by O₃. With a higher $\Delta^{17}O(ASO_4)$ value resulting from the S(IV) + O₃ pathway of 6.5‰, the higher $f_{S(IV)+O_3}$ in these regions results in elevated $\Delta^{17}O(ASO_4)$ values, typically above 2‰. Minor pathways, such as those involving TMI, MHP, and PAA, contribute 2.3 ± 1.8 ‰, 0.25 ± 0.25 ‰, and 0.17 ± 0.12 ‰, respectively (Fig. 1), to ASO₄ formation across the US continuous domain. Primary ASO₄ emissions account for 8.7 ± 6.4 % of total ASO₄ (Fig. 1), with substantial contributions originating from urban and industrial regions. High SO₂ emissions from anthropogenic activities in these areas elevate the role of primary ASO₄, and their impact on $\Delta^{17}O(ASO_4)$ is correspondingly notable but limited to these localized hotspots.

Cloud pH is a critical determinant of ASO₄ formation pathways and $\Delta^{17}O(ASO_4)$ values, with lower cloud pH favoring the H₂O₂ pathway and higher cloud pH supporting the O₃ pathway (Seigneur and Saxena, 1988; Fahey and Pandis, 2001). The concentration of ASO₄ plays a dominant role in lowering cloud pH, primarily due to its origin from sulfuric acid (H₂SO₄). As a strong acid, H₂SO₄ dissociates completely, releasing significant amounts of hydrogen ions (H⁺) and causing substantial acidification of cloud water. In regions such as the Northeast, Southeast, and Midwest, relatively high SO₂ emissions result in elevated ASO₄ concentrations, which further favor the dominance of the H₂O₂ oxidation pathway over O₃, thereby sustaining low $\Delta^{17}O(ASO_4)$ values in the Northeast and Southeast (Fig. S1). This is due to the efficient conversion of dissolved S(IV) species to ASO₄, primarily through the aqueous S(IV) + H₂O₂ pathway under acidic cloud water. Frequent cloud occurrence and abundant oxidant availability accelerate SO₂ to ASO₄ production.

These high ASO₄ levels contribute significantly to lowering cloud pH in these areas, creating an acidic environment (Fig. S1). In contrast, in the Western States, SO₂ emissions and ASO₄ concentrations are comparatively lower (Fig. S1). This results in reduced acidification and a higher cloud pH, as the influence of ASO₄ on the acidity of cloud water is diminished. Ammonium in cloud water (NH₄), on the other hand, primarily acts as a buffering agent, mitigating the acidity caused by ASO₄ (Fig. S1). NH₃ reacts with H₂SO₄ to form (NH₄)₂SO₄ (ammonium sulfate), which reduces the availability of free H⁺ and partially neutralizes the acidification caused by ASO₄. However, the neutralizing capacity of NH₄ is limited. In regions with high ASO₄ concentrations, such as the Northeast and Southeast, the buffering effect of NH₄ is insufficient to fully counteract the strong acidity introduced by ASO₄. In the Midwest, where NH₃ emissions from agricultural activities, particularly fertilization, are significant, the resulting high concentrations of NH₄ partially neutralize the acidity from ASO₄ (Fig. S1). This interaction raises cloud pH slightly, preventing extreme acidification (Fig. S1). Nevertheless, even in regions with abundant NH₃ emissions, cloud water pH typically remains acidic because of the dominant influence of ASO₄ and other atmospheric acids.

3.2 Seasonal Variation in Fractional ASO₄ Formation and $\Delta^{17}O(ASO_4)$ in the Contiguous US in 2019

ASO₄ formation across the contiguous United States exhibits distinct seasonal patterns, shaped by varying contributions from the H₂O₂ and O₃ pathways, as well as shifts in cloud pH and precursor concentrations (Fig. S1). The isotopic composition of ASO₄, represented by $\Delta^{17}O(ASO_4)$, reflects the dominance of specific pathways under different meteorolog-

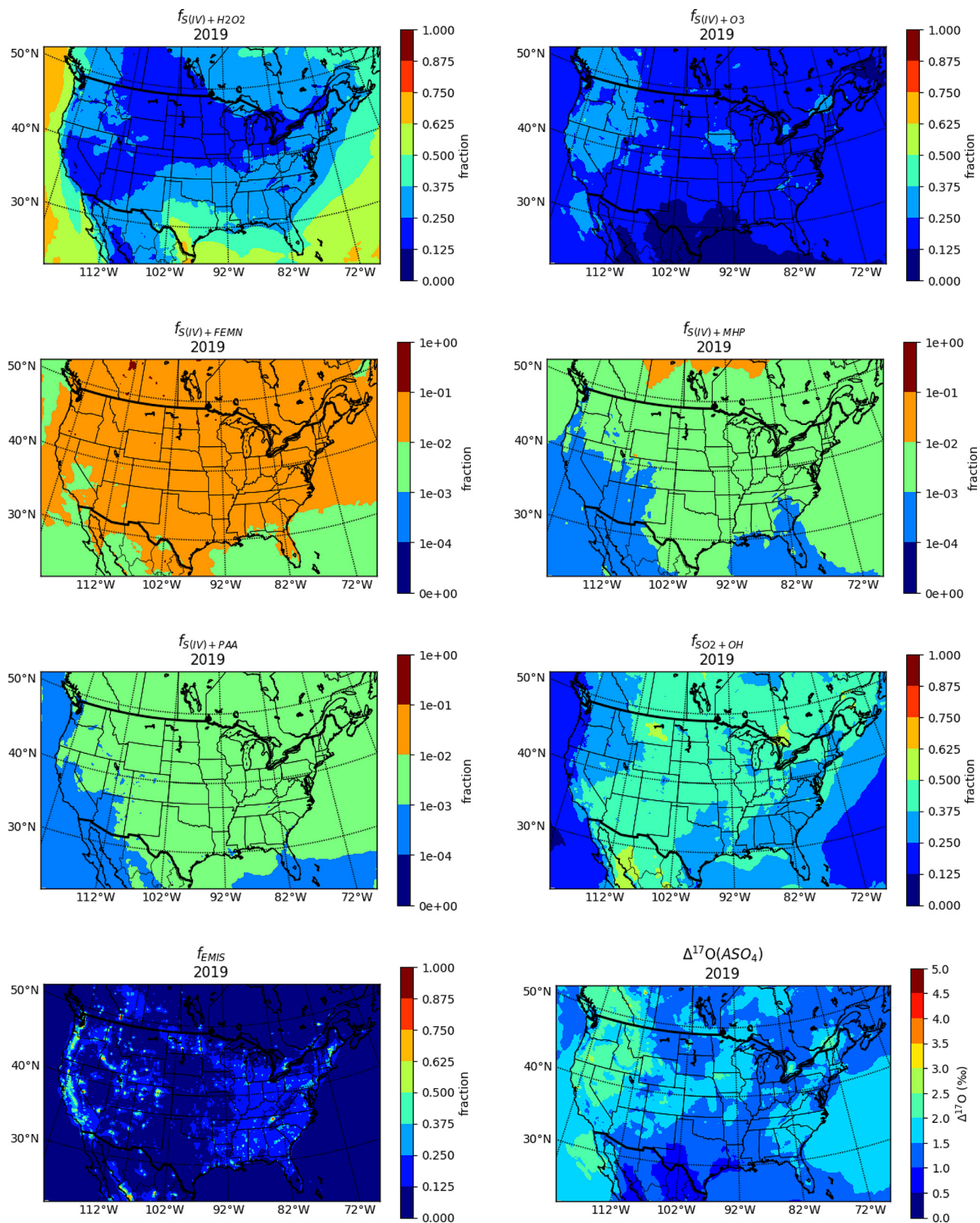


Figure 1. The annual fractional contribution from each ASO_4 formation pathway, along with $\Delta^{17}\text{O}(\text{ASO}_4)$ across the contiguous US in the year 2019, based on CMAQ simulation.

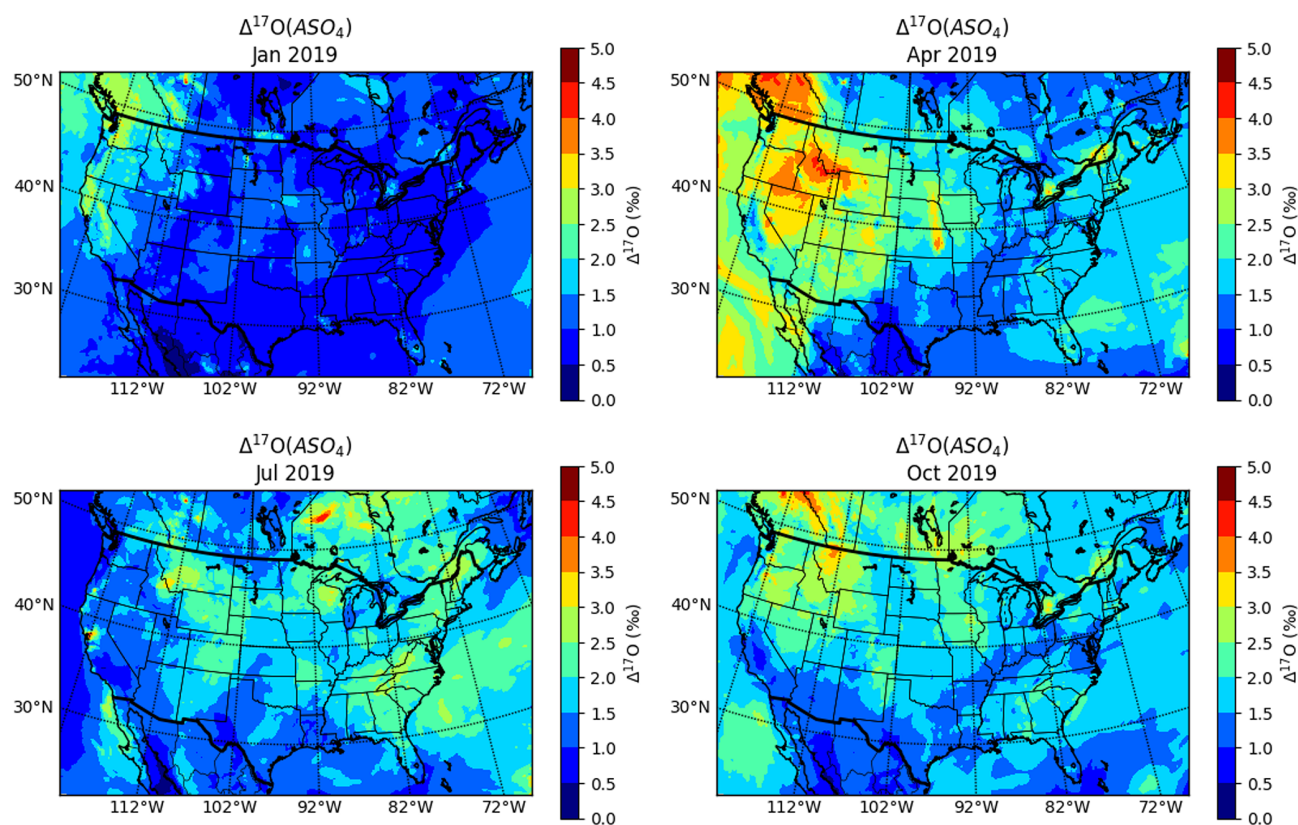


Figure 2. The simulated $\Delta^{17}\text{O}(\text{ASO}_4)$ values across the contiguous US for the year 2019 in each season (winter: January, spring: April, summer: July, fall: October), based on CMAQ simulation.

ical and chemical conditions. In regions with low cloud pH, the H_2O_2 pathway dominates, resulting in low $\Delta^{17}\text{O}$ values (Figs. S1, 2). Conversely, areas with high cloud pH favor the O_3 pathway, resulting in high $\Delta^{17}\text{O}$ values (Figs. S1, 2). Meanwhile, gas-phase $\text{SO}_2 + \cdot\text{OH}$ oxidation (Fig. 5) acts as a consistent background process throughout the year. Its near-zero $\Delta^{17}\text{O}$ signature moderates the contrast between H_2O_2 - and O_3 -dominated regimes, with a stronger influence during summer when photochemical activity peaks. Together with seasonal variations in cloud pH, ASO_4 , and ANH_4 (Figs. S3, S4, S6), this process shapes the overall spatiotemporal pattern of $\Delta^{17}\text{O}(\text{ASO}_4)$, reflecting the coupled effects of emissions, atmospheric chemistry, and meteorology.

In January, the Western States exhibit the highest $\Delta^{17}\text{O}(\text{ASO}_4)$ values, exceeding 2 ‰ (Fig. 2). This is driven by the increased importance of the $\text{S(IV)} + \text{O}_3$ oxidation pathway (Fig. 3), supported by elevated cloud pH levels resulting from low ASO_4 concentrations (Figs. S3, S4). Conversely, the Gulf Coast States show the lowest $\Delta^{17}\text{O}(\text{ASO}_4)$ values, typically below 1 ‰ (Fig. S1), primarily due to the prevalence of the H_2O_2 pathway (Fig. 4). This pathway dominates under low cloud pH conditions caused by high ASO_4 concentrations and limited ANH_4 levels (Figs. S3, S4, S6). In the Midwest, moderate $\Delta^{17}\text{O}(\text{ASO}_4)$ values are shown,

reflecting a balance between pathways. Elevated ANH_4 levels partially neutralize the acidity from high ASO_4 concentrations (Figs. S4, S6). This neutralization raises cloud pH (Fig. S3), slightly lowering the fractional contribution of the $\text{S(IV)} + \text{H}_2\text{O}_2$ pathway (Fig. 4) and contributing to the intermediate $\Delta^{17}\text{O}(\text{ASO}_4)$ values. During this period, the $\text{SO}_2 + \cdot\text{OH}$ pathway remains weak due to limited photochemical activity but provides a modest background effect that slightly reduces the isotopic contrast between H_2O_2 - and O_3 -dominated regimes.

In April, $\Delta^{17}\text{O}$ values increase significantly, particularly in the Western States, rising above 3 ‰ (Fig. 2). This trend indicates an enhanced influence of the O_3 pathway, supported by elevated cloud pH and increased O_3 levels (Figs. 3, S3, S8). In contrast, the Gulf Coast States continue to exhibit low $\Delta^{17}\text{O}$ values (< 1.5 ‰) (Fig. 2), as the H_2O_2 pathway remains dominant due to persistently low cloud pH (Figs. 4, S3). This acidity is driven by high ASO_4 concentrations and low ANH_4 concentrations (Figs. S4, S6). Meanwhile, in the Midwest, cloud pH begins to rise as increased NH_3 levels, partially neutralizing the acidity from ASO_4 and shifting the balance of ASO_4 formation pathways (Figs. S3, S4, S7). The influence of the $\text{SO}_2 + \cdot\text{OH}$ pathway decreases relative to January (Fig. 5), as O_3 oxidation becomes more dominant,

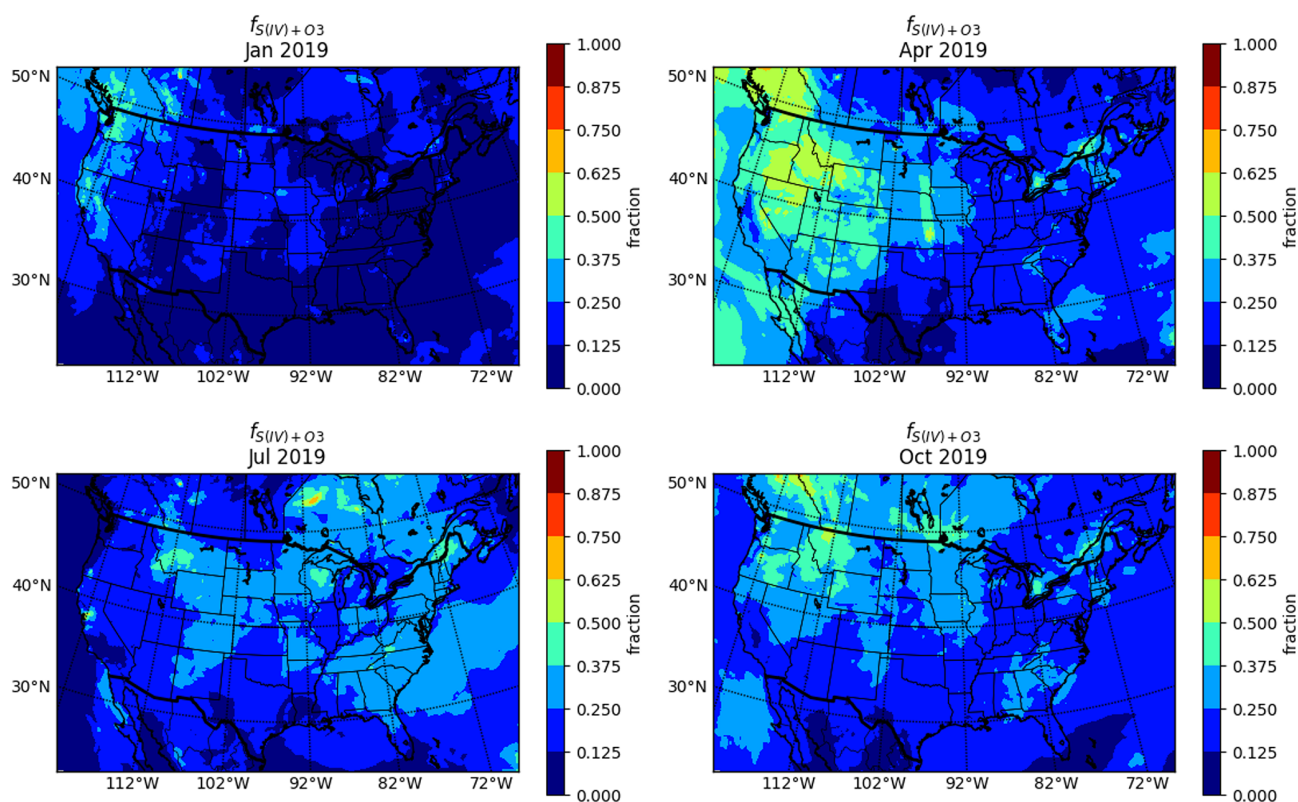


Figure 3. The fraction of ASO_4 formation from $\text{S(IV)} + \text{O}_3$ pathway across the contiguous US for the year 2019 in each season (winter: January, spring: April, summer: July, fall: October), based on CMAQ simulation.

thereby exerting a weaker moderating effect on the isotopic contrast across regions.

In July, $\Delta^{17}\text{O}$ values decrease in the Western States as the $f_{\text{S(IV)} + \text{H}_2\text{O}_2}$ increases compared to April (Fig. 2). This shift is driven by higher water vapor levels and increased cloud cover (Fig. S10), despite the regional consistently high cloud pH (Fig. S3). In the Gulf Coast States, $\Delta^{17}\text{O}$ values remain low, below 1.5‰ (Fig. S1), highlighting the continued dominance of the H_2O_2 pathway (Fig. 4) under conditions of abundant water vapor (Fig. S10), frequent cloud cover (Fig. S11), and persistently low cloud pH (Fig. S3). In the Midwest, cloud pH continues to rise from April (Fig. S3), driven by increasing NH_3 concentrations (Fig. S7), which partially neutralize the acidity caused by ASO_4 (Fig. S4). This elevation in cloud pH enhances the activity of the O_3 pathway (Fig. S2), leading to an increase in $\Delta^{17}\text{O}$ values compared to April. At the same time, the $\text{SO}_2 + \bullet\text{OH}$ pathway reaches its maximum importance (Fig. 5) under strong photochemical conditions, offsetting the isotopic enrichment from O_3 oxidation and contributing to the relative decrease in $\Delta^{17}\text{O}(\text{ASO}_4)$ compared to April.

In October, $\Delta^{17}\text{O}$ values in the Western States increase compared to July but remain slightly lower than in April (Fig. 2). This change is attributed to the enhanced $f_{\text{S(IV)} + \text{O}_3}$ (Fig. S2), supported by high cloud pH and low ASO_4 con-

centrations (Figs. S3, S4). In the Gulf Coast States, $\Delta^{17}\text{O}$ values remain low (Fig. 2), reflecting the continued dominance of the H_2O_2 pathway under acidic conditions sustained by high ASO_4 levels and low ANH_4 concentrations (Figs. S3, S4, S6). In the Midwest, decreasing NH_3 levels from July reduce the neutralization of acidity (Figs. S3, S7), making conditions less favorable for O_3 -driven ASO_4 formation (Fig. 3). This results in lower cloud pH (Fig. S3) and diminished $\Delta^{17}\text{O}$ values compared to earlier months. As photochemical activity weakens, the relative contribution of the $\text{SO}_2 + \bullet\text{OH}$ pathway declines (Fig. 5), reducing its moderating effect and allowing O_3 -driven isotopic enrichment to strengthen in high-pH regions.

Seasonal variations in ASO_4 formation and $\Delta^{17}\text{O}(\text{ASO}_4)$ highlight the interplay of chemical drivers and meteorological conditions. The dominance of the H_2O_2 pathway in acidic, ASO_4 -rich regions, such as the Gulf Coast States, leads to low $\Delta^{17}\text{O}$ values year-round. In contrast, the O_3 pathway prevails in higher pH regions such as the Western States, driving elevated $\Delta^{17}\text{O}$ values, particularly in April. The Midwest experiences transitional conditions, where cloud pH and NH_3 concentrations modulate the relative contributions of ASO_4 formation pathways. Alongside these seasonal and spatial contrasts, the gas-phase $\text{SO}_2 + \bullet\text{OH}$ pathway acts as a persistent, near-zero- $\Delta^{17}\text{O}$ background that

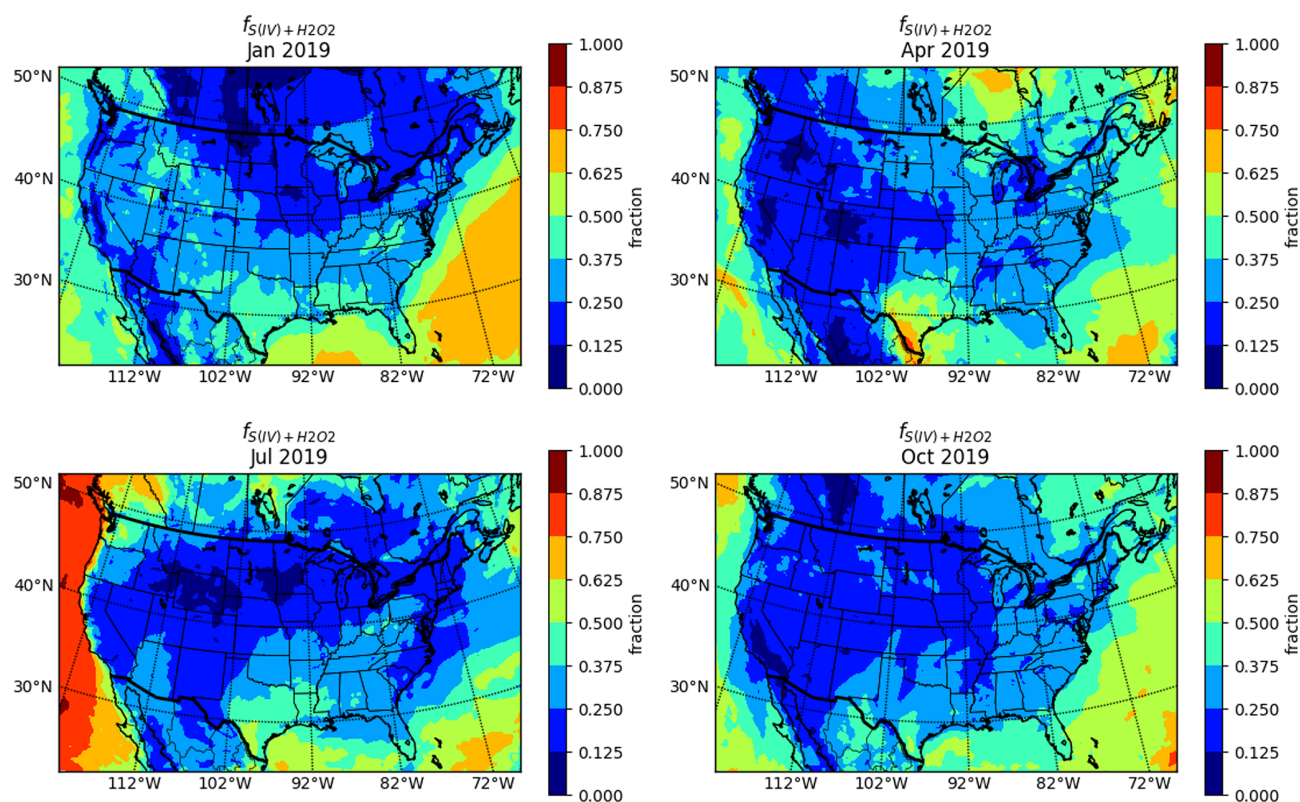


Figure 4. The fraction of ASO₄ formation from S(IV) + H₂O₂ pathway across the contiguous US for the year 2019 in each season (winter: January, spring: April, summer: July, fall: October), based on CMAQ simulation.

offsets isotopic enrichment from O₃ oxidation, particularly during summer when photochemical activity peaks. These findings underscore the dynamic nature of ASO₄ chemistry across seasons, emphasizing the importance of emissions, atmospheric composition, and cloud chemistry in shaping regional and seasonal patterns of ASO₄ formation.

3.3 Change in Fractional Annual ASO₄ Formation and $\Delta^{17}\text{O}(\text{ASO}_4)$ from 2006 to 2019

From 2006 to 2019, the annual $\Delta^{17}\text{O}(\text{ASO}_4)$ values across the contiguous US showed a consistent increase, highlighting the growing importance of the O₃ pathway in ASO₄ formation (Fig. 7). In the central and eastern US, $\Delta^{17}\text{O}(\text{ASO}_4)$ values increased by up to 2‰ (Fig. 7), primarily driven by significant reductions in SO₂ emissions, largely attributable to regulatory measures such as the Clean Air Act. These reductions led to lower ASO₄ concentrations, which elevated cloud pH and shifted the ASO₄ formation process toward the O₃ pathway (Fig. S12), resulting in ASO₄ with higher $\Delta^{17}\text{O}$ values. Conversely, the western US exhibited only modest increases in $\Delta^{17}\text{O}(\text{ASO}_4)$, typically less than 1‰ (Fig. 7). This is because the region historically favored O₃-dominated ASO₄ formation due to consistently high O₃ and cloud pH levels (Fig. S13), making the impacts of rising

cloud pH and reduced SO₂ emissions less obvious. Changes in H₂O₂ concentrations played a significant role in shaping these trends. In the central and eastern US, slight increases in H₂O₂ concentrations continued to support the H₂O₂ pathway, to a limited extent, even as the O₃ pathway became more dominant (Figs. 7, S12). In contrast, in the western US, H₂O₂ concentrations decreased slightly, resulting in a slight reduction in $f_{\text{S(IV)}+\text{H}_2\text{O}_2}$ (Fig. 7). A decrease in $f_{\text{SO}_2+\text{OH}}$ across the domain along with the negligible contributions from other pathways caused a relative increase in f_{EMIS} in these regions (Fig. 7). Between 2006 and 2019, the domain-averaged $f_{\text{SO}_2+\text{OH}}$ decreased by 12.5 %, primarily due to the enhanced contribution of the O₃ oxidation pathway ($f_{\text{S(IV)}+\text{O}_3}$) and lower SO₂ concentrations under reduced precursor emissions, which together shifted the overall oxidation balance toward aqueous-phase processes. Consistent with these trends, spatial patterns of concentration changes (Fig. S12) show substantial decreases in SO₂ and ASO₄, particularly in the eastern US, while NH₃ and ANH₄ increased, leading to higher cloud pH and favoring O₃ oxidation. Meanwhile, primary ASO₄ emissions, which do not carry a mass-independent signature and exhibit $\Delta^{17}\text{O}$ values close to 0‰, directly added to ASO₄ levels and tempered changes in $\Delta^{17}\text{O}(\text{ASO}_4)$ values (Fig. 7). This dynamic explains why the increases in $\Delta^{17}\text{O}(\text{ASO}_4)$ values from 2006

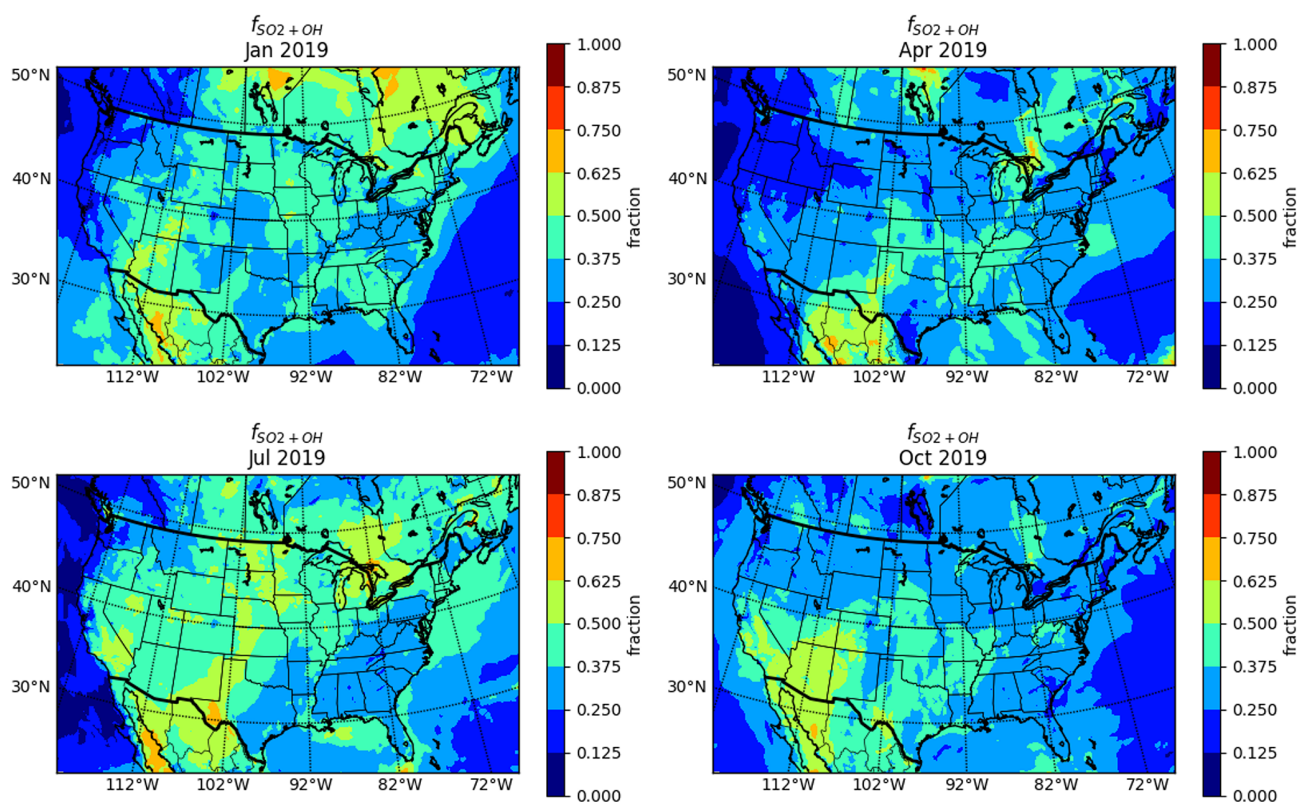


Figure 5. The fraction of ASO_4 formation from $\text{SO}_2 + \text{OH}$ pathway across the contiguous US for the year 2019 in each season (winter: January, spring: April, summer: July, fall: October), based on CMAQ simulation.

to 2019 were smaller in the western US compared to the central and eastern regions.

3.4 Comparison of Model $\Delta^{17}\text{O}(\text{ASO}_4)$ with Observations

The CMAQ simulations of $\Delta^{17}\text{O}(\text{ASO}_4)$ across the contiguous United States reveal significant insights into atmospheric ASO_4 formation over recent decades. However, observations of $\Delta^{17}\text{O}(\text{ASO}_4)$ in the contiguous US are very limited, with data primarily collected in the late 1990s at La Jolla, CA, and White Mountain Research Station, CA (Lee and Thiemens, 2001), and in the early 2000s at Baton Rouge, LA (Jenkins and Bao, 2006). These historical $\Delta^{17}\text{O}(\text{ASO}_4)$ data exhibit a range from 0.2‰ to 1.6‰ (Table S1). Due to the predicted change in ASO_4 chemistry from 2006 to 2019, the 2006 model simulation was chosen for evaluation against these observations (Figs. 8, 9). The comparison with historical $\Delta^{17}\text{O}(\text{ASO}_4)$ data is intended as a preliminary evaluation rather than a strict validation, given the temporal mismatch between the available observations (1990s–early 2000s) and the simulation years (2006, 2019). The observed range of 0.2‰ to 1.6‰ provides a useful benchmark for assessing whether the model produces realistic isotopic signatures. However, the limited number and dated nature of

these measurements preclude a comprehensive validation of ASO_4 chemistry. This further emphasizes the critical need for new $\Delta^{17}\text{O}(\text{ASO}_4)$ observations within the contiguous United States to enable robust model–observation comparisons.

Generally, the CMAQ model reasonably reproduced $\Delta^{17}\text{O}(\text{ASO}_4)$ at the Baton Rouge, LA site, with a Root Mean Square Error (RMSE) of 0.20‰ ($n = 17$). This region is characterized by relatively low predicted $\Delta^{17}\text{O}(\text{ASO}_4)$ values, consistent with high regional SO_2 emissions and low cloud water pH that favor ASO_4 formation through aqueous $\text{S(IV)} + \text{H}_2\text{O}_2$ reactions. In contrast, the CMAQ-simulated $\Delta^{17}\text{O}(\text{ASO}_4)$ values tended to be overestimated at the California sites, suggesting possible inaccuracies in representing additional ASO_4 production pathways in this region. The La Jolla, CA site had an RMSE of 0.36‰ ($n = 31$), while the White Mountain, CA site had a notably higher RMSE of 1.9‰ ($n = 6$). Despite the limited number of $\Delta^{17}\text{O}(\text{ASO}_4)$ observations, a temporal analysis of model simulations versus observations indicates a consistent overprediction during the spring (Fig. 9). The $\Delta^{17}\text{O}(\text{ASO}_4)$ overestimation in spring could be associated with higher predicted cloud pH during this season, which promotes the $\text{S(IV)} + \text{O}_3$ oxidation pathway in the model (Fig. S26). The elevated cloud pH may result from increased NH_3 emissions, likely related

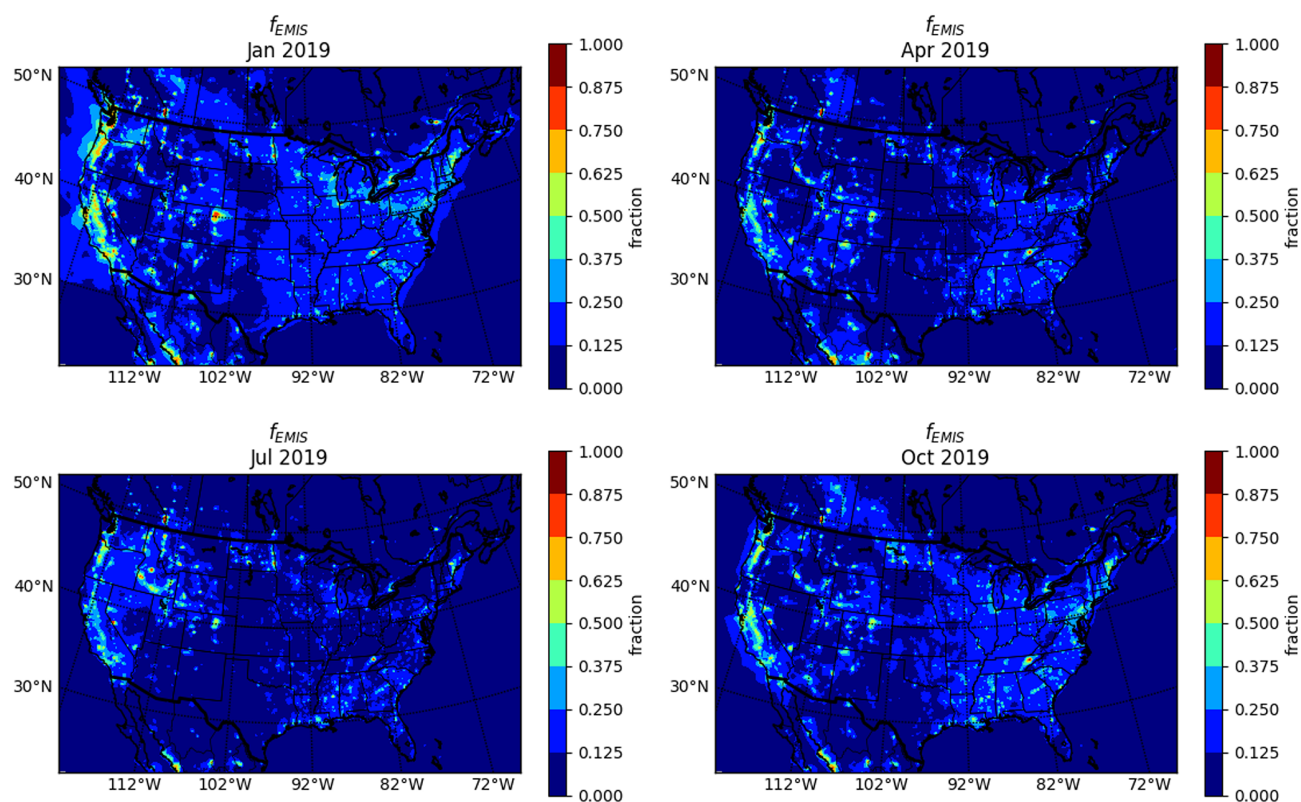


Figure 6. The fraction of ASO₄ from primary emission across the contiguous US for the year 2019 in each season (winter: January, spring: April, summer: July, fall: October), based on CMAQ simulation.

to fertilizer use in surrounding agricultural areas or to underrepresentation of marine boundary layer processes that could influence ASO₄ production (Guo et al., 2017; Lim et al., 2022; Zheng et al., 2024; Wang et al., 2025). Given the strong nonlinear pH dependence of ASO₄ formation, even moderate NH₃ emission biases can produce significant changes in isotopic composition. Future work should include explicit sensitivity simulations to better quantify the coupled effects of NH₃, cloud pH, and oxidant chemistry on modeled $\Delta^{17}\text{O}(\text{ASO}_4)$. While organic acids (e.g., formic, acetic) can locally influence cloud water acidity, their contribution to bulk pH is generally minor relative to the dominant SO₂-H₂SO₄-NH₃ system (Herrmann et al., 2015; Shah et al., 2020; Tsui et al., 2019). Still, future model developments should evaluate their role in regional cloud pH and isotopic composition. Additionally, certain ASO₄ formation pathways, such as marine boundary layer chemistry involving S(IV) oxidation by HOX, may not be fully captured, particularly at coastal sites like La Jolla, CA. These reactions can efficiently oxidize S(IV) even under moderately acidic conditions and produce ASO₄ with relatively low $\Delta^{17}\text{O}$ signatures (Chen et al., 2016; Ishino et al., 2017), which may partly explain the model overestimation at this site. Another possible factor is the omission of S(IV) oxidation induced by NO₂, an emerging multiphase pathway in polluted environ-

ments. Such reactions can proceed alongside metal-catalyzed and other aqueous pathways and are anticipated to result in low or near-zero $\Delta^{17}\text{O}$. Sensitivity simulations suggest that this mechanism can enhance ASO₄ concentrations by $\sim 0.4\%$ – 1.2% with a low rate constant and up to 4% – 20% with a higher rate constant, particularly under polluted, low-oxidant wintertime conditions, when the aqueous S(IV) oxidation by H₂O₂ and O₃ becomes less efficient (Sarwar et al., 2013), while its overall impact on $\Delta^{17}\text{O}(\text{ASO}_4)$ is expected to be minor.

Overall, the model-observation comparison of $\Delta^{17}\text{O}(\text{ASO}_4)$ suggests that CMAQ performs well in more acidic environments but struggles to simulate ASO₄ formation under less acidic conditions accurately. However, the limited availability of $\Delta^{17}\text{O}(\text{ASO}_4)$ observations constrains a more comprehensive evaluation of regional and temporal ASO₄ chemistry variations. This highlights the critical need for expanded observational datasets and model refinements to better represent the complex atmospheric ASO₄ dynamics. While this study highlights consistent patterns in ASO₄ oxidation pathways across the contiguous US, the evaluation of $\Delta^{17}\text{O}(\text{ASO}_4)$ remains constrained by the limited and dated nature of available measurements. Expanded and more recent datasets will be essential to validate and extend the findings presented here, particularly

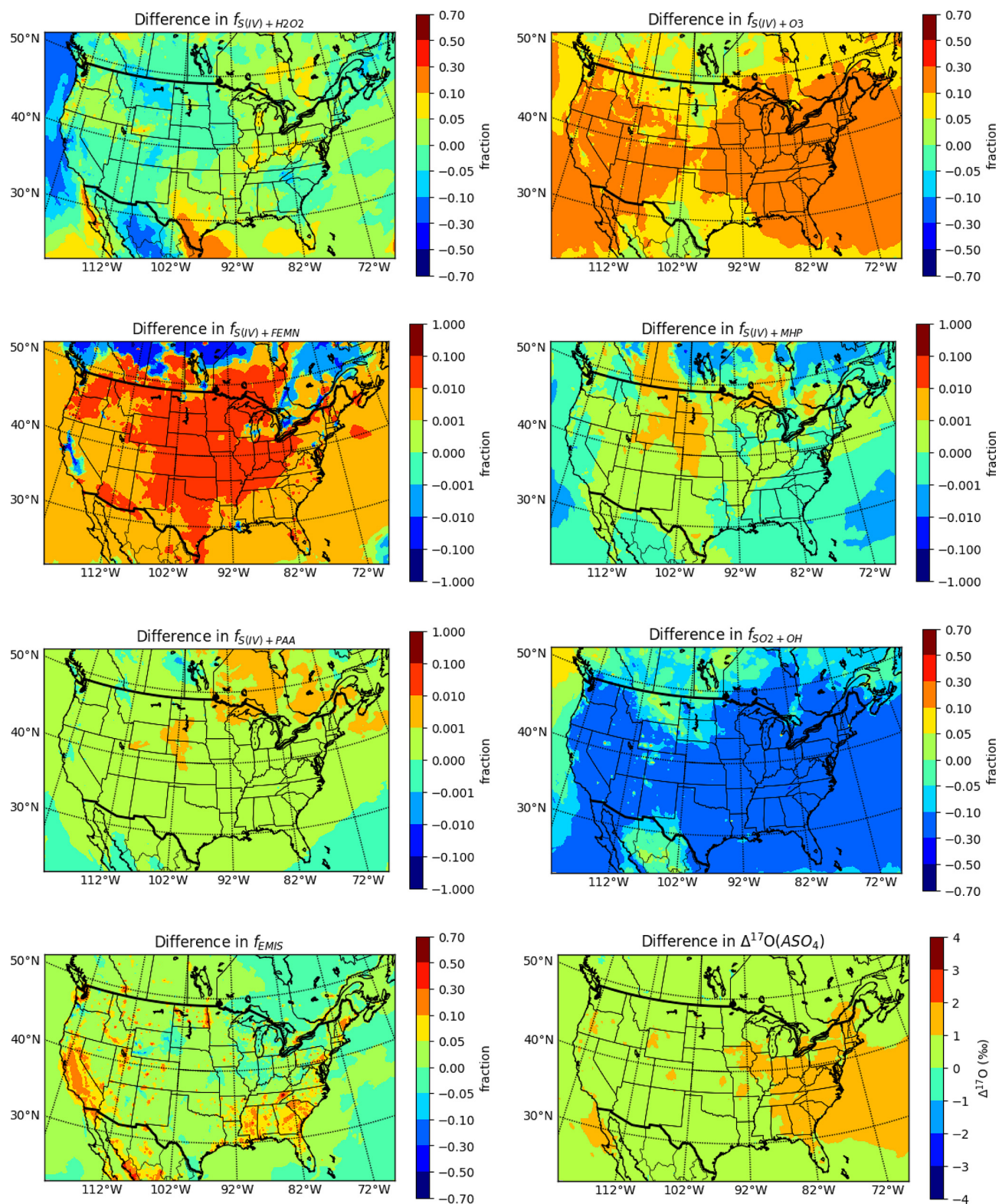


Figure 7. The change in the fraction from each ASO_4 formation pathway and $\Delta^{17}\text{O}$ values across the contiguous US, from 2006 to 2019, based on CMAQ simulation.

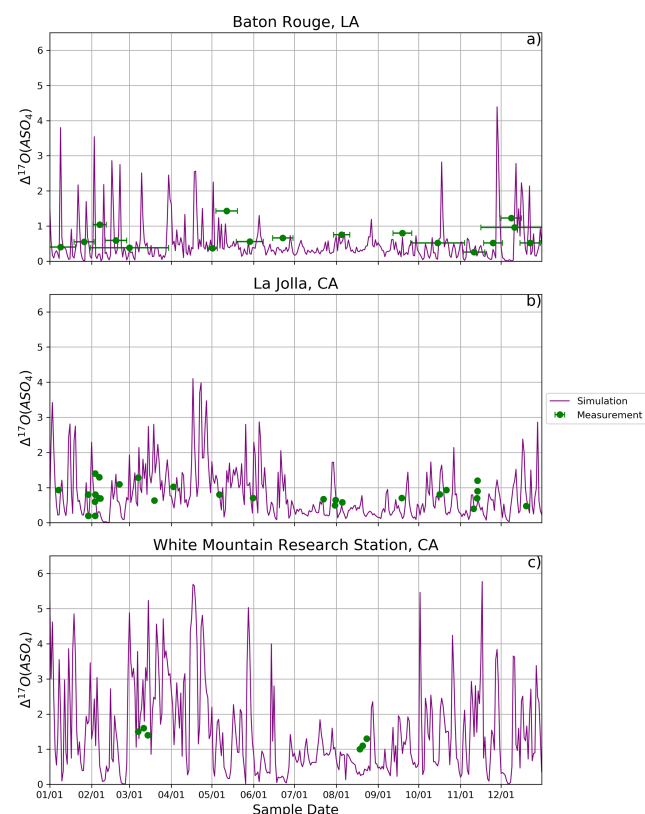


Figure 8. Temporal variations in $\Delta^{17}\text{O}(\text{ASO}_4)$ measurements and model simulations at (a) Baton Rouge, LA (top); (b) La Jolla, CA (middle); and (c) White Mountain Research Station, CA (bottom). The x axis error bars correspond to collection times.

to quantify seasonal and regional variability in ASO_4 formation.

4 Conclusions

This study modeled ASO_4 formation pathways and the $\Delta^{17}\text{O}(\text{ASO}_4)$ for the contiguous United States using the CMAQ model for 2006 and 2019. The results reveal distinct seasonal and regional patterns in ASO_4 chemistry, strongly influenced by photochemical conditions, emissions of SO_2 and NH_3 , and variations in cloud pH. From 2006 to 2019, significant changes in ASO_4 formation dynamics were observed, driven primarily by regulatory-driven reductions in SO_2 emissions. These shifts highlight the evolving balance between aqueous-phase oxidation pathways, particularly those driven by H_2O_2 and O_3 .

The reductions in SO_2 emissions due to the Clean Air Act resulted in lower cloud water ASO_4 , which subsequently increased cloud pH. This change shifted ASO_4 production toward the O_3 pathway, particularly in the eastern US, where the O_3 pathway was once limited by lower pH levels in 2006. By 2019, ASO_4 formation via O_3 oxidation had increased significantly, indicating a more efficient production mecha-

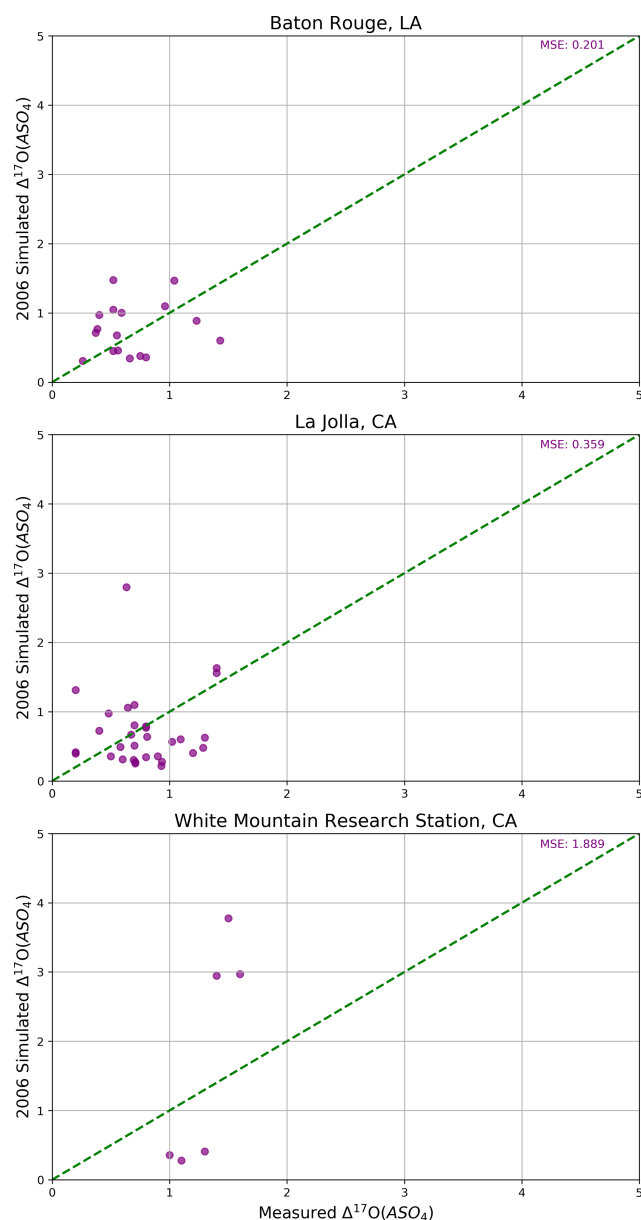


Figure 9. Comparison of $\Delta^{17}\text{O}(\text{ASO}_4)$ measurements and model simulations at La Jolla, CA, White Mountain Research Station, CA, and Baton Rouge, LA from 1996 to 2005.

nism under elevated pH conditions. The sub-linear response of ASO_4 concentrations to SO_2 emission reductions highlights the complexity of ASO_4 formation chemistry and the role of co-emitted species, such as NH_3 , in modifying pH levels and influencing pathway dominance.

The isotopic signature $\Delta^{17}\text{O}(\text{ASO}_4)$ serves as a powerful tracer for tracking shifts in ASO_4 formation pathways. In regions with limited photochemical activity, such as during winter or in areas with high primary ASO_4 emissions, lower $\Delta^{17}\text{O}(\text{ASO}_4)$ values were associated with greater contributions from primary ASO_4 emissions. Conversely, higher

$\Delta^{17}\text{O}(\text{ASO}_4)$ values reflected an increased role of the O_3 pathway, particularly in regions with elevated cloud pH, reduced SO_2 emissions, and higher ozone concentrations.

This work demonstrates a significant and predictable shift in ASO_4 chemistry over the study period. The introduction of $\Delta^{17}\text{O}(\text{ASO}_4)$ as a diagnostic tool for probing ASO_4 formation mechanisms provides a novel approach for investigating these changes. Expanding the measurement of $\Delta^{17}\text{O}(\text{ASO}_4)$ across diverse regions and time periods will be critical for validating and extending these findings. Future studies should prioritize exploring how changes in atmospheric composition and regulatory measures continue to influence ASO_4 chemistry, with a particular focus on understanding the increasing prominence of O_3 -driven chemistry. This effort will be crucial for enhancing atmospheric models and understanding the implications of ASO_4 chemistry on air quality, human health, and climate.

Code availability. The source code for CMAQ version 5.4 is available at <https://github.com/USEPA/CMAQ/tree/5.4> (last access: 1 March 2025).

Data availability. The input datasets for CMAQ simulation are available at https://cmas-equates.s3.amazonaws.com/index.html#CMAQ_12US1/INPUT/ (last access: 1 March 2025). The in-detail simulation results for $\Delta^{17}\text{O}(\text{ASO}_4)$ are archived on Zenodo.org (<https://doi.org/10.5281/zenodo.14954960>, Fang, 2025).

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Author contributions. HF and WW designed the study. HF conducted the model simulations and analysis with input from WW. HF wrote the manuscript with input from all authors. WW secured funding.

Competing interests. The contact author has declared that neither of the authors has any competing interests.

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