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Formation of chlorinated organic compounds from Cl atom-initiated reactions of aromatics and their detection in suburban Shanghai

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Abstract. Chlorine (Cl) atoms generated from the photolysis of atmospheric reactive chlorine species can rapidly react with various volatile organic compounds (VOCs), forming chlorine- and non-chlorine-containing low-volatile oxygenated organic molecules. Yet, the formation mechanisms of chlorine-containing oxygenated organic molecules (Cl-OOMs) from reactions of Cl atoms with aromatics in the presence and absence of NO_x are not fully understood. Here, we investigated Cl-OOMs formation from Cl-initiated reactions of three typical aromatics (i.e., toluene, m-xylene, and 1,2,4-trimethylbenzene (1,2,4-TMB)) in the laboratory and searched for ambient gaseous Cl-OOMs in suburban Shanghai. From our laboratory experiments, 19 Cl-containing peroxyl radicals and a series of Cl-OOMs originating from the Cl-addition-initiated reaction were detected, which provides direct evidence that the Cl-addition-initiated reaction is a non-negligible pathway. In addition, a total of 51 gaseous Cl-OOMs were identified during the winter in suburban Shanghai, 38 of which were also observed in laboratory experiments, hinting that Cl-initiated reaction of aromatics could serve as a source of Cl-OOMs in an anthropogenically influenced atmosphere. Toxicity evaluation of these Cl-OOMs shows potential adverse health effects. These findings demonstrate that Cl-OOMs can be efficiently formed via the Cl-addition pathway in the reactions between aromatics and Cl atoms and some of these Cl-OOMs could be toxic.

1 Introduction

Atmospheric chlorine atoms (Cl), together with hydroxyl radicals (OH), ozone (O₃), and nitrate radicals (NO₃), play vital roles in transforming volatile organic compounds (VOCs), leading to the formation of oxygenated organic molecules (OOMs) and secondary organic aerosol (SOA)(Priestley et al., 2018; Shang et al., 2021; Tham et al., 2016; Thornton et al., 2010). The involvement of

Cl atoms in atmospheric chemical processes was conventionally thought to be confined to the oceanic boundary layer (Keene et al., 1999; Knipping et al., 2000). Estimated Cl atom concentrations in the coastal region ranged from 10^2 to 10^5 molec. cm⁻³ (Thornton et al., 2010; Wingenter et al., 2005). Recently, a number of reactive chlorine species, such as nitryl chloride (ClNO₂), chlorine nitrate (ClONO₂), hypochlorous (HOCl), chlorine (Cl₂), bromine chloride (BrCl), and hydrochloric acid (HCl), were found

to lead to high concentrations of Cl atoms in the urban and suburban atmospheres (Le Breton et al., 2018; Peng et al., 2020; Priestley et al., 2018). During the daytime, peak Cl atom concentrations can reach 10⁶ molec. cm⁻³ (Wang et al., 2023). which is still less than the global average concentrations of OH radicals (Le Breton et al., 2018; Liu et al., 2017). Nonetheless, the reaction rate coefficients of VOCs with Cl atoms are generally 1–2 orders of magnitude larger than those of OH radicals, which can partially compensate for the lower concentration of Cl atoms when determining the relative importance of different reactive loss pathways of VOCs (Chen et al., 2023; Riva et al., 2015; Wang et al., 2005).

Apart from H-abstraction, Cl atoms can be added to VOCs forming chlorine-containing oxygenated organic molecules (Cl-OOMs) in reactions of VOCs with Cl atoms. For small alkenes (e.g., isoprene), the reaction mechanism is dominated by Cl-addition to the double bond, with some allylic hydrogen abstraction (approximately 15% for isoprene at 1 atm) (Finlayson-Pitts et al., 1999; Orlando et al., 2003; Ragains and Finlayson-Pitts, 1997). For larger biogenic VOCs (e.g., β -pinene), which contain a greater number of abstractable hydrogen atoms, H-abstraction becomes more significant; for instance, H-abstraction is thought to account for half of the overall initiation reactions (Finlayson-Pitts et al., 1999). Overall, both Cl-addition and H-abstraction pathways coexist for biogenic VOCs, with the Cl-addition pathway possibly being the more dominant pathway, such as isoprene (Wang et al., 2022) and d-limonene (Wang et al., 2019).

Given the ubiquitous existence of aromatics in the urban air and the recent detection of reactive chlorine species that are precursors of Cl atoms in the same atmosphere, the reaction mechanisms of aromatics and Cl atoms should be of concern. Although the reaction rate coefficient of benzene with Cl atoms under normal atmospheric conditions is relatively slow $(1.3 \times 10^{-15} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1})$ and this reaction is insignificant in the ambient atmosphere, its reaction mechanism provides valuable insights into reaction pathways of Cl atoms and aromatics (Shi and Bernhard, 1997). Sokolov et al. (1998) proposed that C₆H₆Cl radicals, formed through a Cl-addition pathway in reactions of Cl atoms and benzene, can either decompose back to benzene or further react in a non-aromatizing manner (Sokolov et al., 1998). In contrast, toluene, xylene, and trimethylbenzene react with Cl atoms at faster rates, urging a better understanding of their reaction mechanisms. Using the density functional theory and the conventional transition state theory, Huang et al. (2012) investigated the Cl-toluene reaction (Huang et al., 2012). Their findings indicate that the reaction rate coefficient for the H-abstraction pathway $(5.58 \times 10^{-11} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1})$ is substantially higher than that for the Cl- addition pathway $(0.91 \times 10^{-11} \,\mathrm{cm}^3 \,\mathrm{molec.}^{-1} \,\mathrm{s}^{-1})$, highlighting the significance of the H-abstraction pathway that accounts for approximately 86% of Cl-initiated reactions. Studies employing gas chromatography-mass spectrometry (GC-MS) to examine products of Cl-initiated reactions of toluene and iodidebased chemical ionization mass spectrometry (I-CIMS) to analyze products of Cl-initiated reactions of m-xylene, have both predominantly detected chlorine-free oxygenated organic compounds, which were supposed as main products (Cai et al., 2008; Wang et al., 2005). These results have guided atmospheric models to integrate the H-abstraction as the primary reaction pathway in their analytical frameworks (Ma et al., 2023; Peng et al., 2022). However, studies focusing on the Cl-addition pathway and its related products are sparse, and the significance of the Cl-addition pathway in the atmospheric reactions of Cl atoms and aromatics remains elusive. Over the past decade, thanks to the development of nitrate-based chemical-ionization atmospheric-pressureinterface long-time-of-flight mass spectrometers (nitrate-CI-APi-LToF), there have been significant advancements in the detection of highly oxygenated organic molecules and radicals, which is crucial for elucidating the mechanisms of Clinitiated reactions with VOCs (Bianchi et al., 2019; Ehn et al., 2010). The formation of highly oxygenated organic products, including Cl-containing ones from the reactions of biogenic VOCs with Cl atoms, were gradually revealed, whereas whether or not significant amounts of Cl-containing highly oxygenated organics and radicals can be formed from Cl atoms and aromatics remains unclear (Wang et al., 2020).

OOMs can lead to potential air quality, climate, and health effects. Due to their low volatility, OOMs have been identified as dominant precursors for the growth of newly formed particles and formation of SOAs, which are known for their negative effect on air quality and climate impacts (Ehn et al., 2014; Kulmala et al., 2013). Compared to non-chlorine-containing OOMs (non-Cl-OOMs), Cl-OOMs formed through the introduction of chlorine substituents make organic compounds more lipophilic, facilitating their interactions with hydrophobic sites and promoting enzymatic biotransformation in general, which can lead to adverse health effect in turn (Henschler, 1994; Zhang et al., 2019). Therefore, field and laboratory studies for characteristics and sources of Cl-OOMs from the reactions of Claromatics and their risk assessment upon human atmospheric exposure should be carried out.

In this study, we investigated non-Cl-OOMs and Cl-OOMs formation mechanisms from Cl-initiated reactions of toluene, m-xylene, and 1,2,4-trimethylbenzene (1,2,4-TMB) in the presence and absence of NO_x in a laboratory flow reactor, using a nitrate-CI-APi-LToF (Aerodyne Research, Inc. USA, and Tofwerk AG, Switzerland) and a Vocus proton-transferreaction long-time-of-flight mass spectrometer (Vocus-PTR-LToF) (Tofwerk AG, Switzerland). In addition, the nitrate CI-APi-LToF was also deployed in a field campaign in suburban Shanghai during winter to search for ambient gaseous Cl-OOMs. The toxicity of selected Cl-OOMs, which were simultaneously detected both in laboratory experiments and ambient observations, was evaluated by computational toxicity.

2 Materials and methods

2.1 Experimental Set-up in the Laboratory

A general scheme of the experimental setup is shown in Fig. S1 in the Supplement. Simulation experiments were conducted in a 6L quartz flow tube reactor with a total flow rate of $10L \, \mathrm{min}^{-1}$, resulting in a residence time of $\sim 36 \, \mathrm{s}$. This flow tube is covered by aluminum composite panels to avoid room light. Zero air with relative humidity (RH) less than 1 % generated from a Zero Air Generator (AADCO Instruments, Inc. USA) was used as carrier gas. The reaction temperature was maintained at around $20 \, ^{\circ}\mathrm{C}$.

Homemade aromatics– N_2 cylinders were prepared from their liquid standards (toluene, ≥ 99.0 %, Aladdin; m-xylene, ≥ 99.0 %, Aladdin; 1,2,4-TMB, ≥ 99.5 %, Aladdin) together with high-purity nitrogen gases. Cl atoms were produced by photolysis of chlorine (Cl₂, Shanghai Wetry Standard Reference Gas Analytical Technology Co., LTD) using 350 nm UV lights. In experiments with NO_x , NO (Air Liquid Co., LTD) was added into the flow tube to produce and sustain NO_x mixing ratios that were sufficiently high to be a competitive sink for RO_2 radicals. RH is controlled by changing zero air flowrates through the water bubbler. Before each experiment, the wall of the flow tube was cleaned with a water/alcohol solution and then purged with zero air for over 1 h. Details of the experimental setup are provided in Text S1.

The concentration of Cl atoms was controlled by adjusting the flow rate of Cl₂. The mean concentrations of Cl atoms were determined according to the decay of aromatic precursors (Fig. S2) and calculated using Eq. (1), with reaction rate coefficients k of 6.2×10^{-11} , 1.35×10^{-10} and 2.42×10^{-10} cm³ molec.⁻¹ s⁻¹ for reactions between Cl atoms and toluene, m-xylene, and 1,2,4-TMB, respectively (Wang et al., 2005), as follows:

$$[C1] = -1/kt \times \ln([Aromatics]_t/[Aromatics]_0)$$
 (1)

where [Aromatics]₀ and [Aromatics]_t are the initial concentration and the concentration after a reaction time t of aromatic precursors, respectively. [CI] is the estimated concentration of Cl atoms in the flow tube. In our flow tube experiments, the extent of oxidation is quantified using the parameter of Cl exposure, defined as [CI] multiplied by the reaction time t. Cl exposures in our experiments were in the range of $(1.2–2.0) \times 10^9$ molec. cm⁻³ s, equivalent to atmospheric oxidation times of roughly 0.6–1.3 h for aromatics at a daytime Cl atom concentration of 5×10^5 molec. cm⁻³ (Chang et al., 2004; Tham et al., 2016; Wang et al., 2023).

A Vocus-PTR-LToF and a nitrate-CI-APi-LToF (more details refer to Text S2 and Fig. S3) were simultaneously deployed to detect aromatic precursors and gaseous OOM products, respectively. In addition, an iodide-CIMS (I-CIMS, Tofwerk AG, Switzerland) was used in a separate experiment focusing on the m-xylene + Cl system to expand the detection of Cl-initiated reaction products. Their working prin-

ciples were described in details elsewhere (Eisele and Tanner, 1993; Krechmer et al., 2018). Signals of aromatic precursors and reaction products measured from the zero air were treated as their background. The resolving power of the nitrate CI-APi-LToF was up to around 8000 for ions with m/z larger than 200 Th. The ions of NO_3^- , $HNO_3 \cdot NO_3^-$, and $C_6H_5NO_3 \cdot NO_3^-$ were selected for mass calibration, and the calibration error is less than 1 ppm. When identifying the OOM signal peaks, the error is limited below 4 ppm.

OOM concentrations are estimated by Eq. (2) (Kürten et al., 2016),

[OOMs] =

$$C \times \frac{\text{OOM} \cdot \text{NO}_3^-}{\text{NO}_3^- + \text{HNO}_3 \cdot \text{NO}_3^- + (\text{HNO}_3)_2 \cdot \text{NO}_3^-} \times T \quad (2)$$

 $OOM \cdot NO_3^-$, NO_3^- , $HNO_3 \cdot NO_3^-$, (HNO₃)₂ · NO₃[−] represent signals of corresponding ions in units of counts per second (cps). OOMs with an oxygen content of equal to or more than 6 (i.e., highly oxygenated organic molecules, HOMs) are assumed to cluster with NO₃ at the same rate coefficient as that of sulfuric acid (H₂SO₄), i.e., both at collision-limited rates (Bianchi et al., 2019; Ehn et al., 2010). Therefore, the calibration factor C for sulfuric acid is adopted as that of OOMs (Kürten et al., 2011, 2012). It should be noted we also used the same calibration factor C for quantification of OOMs with an oxygen number of less than 6, which may lead to relatively high uncertainties (Alage et al., 2024). A mass-dependent transmission correction factor T of our instrument is also taken into account in this study (Heinritzi et al., 2016). The mass-dependent transmission correction factor is instrument-specific and determined by depleting the primary ion with a series of perfluorinated acids and comparing the primary ion signal depletion with the product signal increase (which would match for equivalent transmission efficiency) (Lu et al., 2020).

In addition, a NO_x monitor (Thermo, 49i) was utilized to measure NO_x concentrations in laboratory experiments. A nano-SMPS (Scanning Mobility Particle Sizer with a nano Differential Mobility Analyzer, TSI, USA) together with a PSM (Particle Size Magnifier, Airmodus, Finland) were used to detect particles in the range of sub-3 to 60 nm, indicating the absence of newly formed particles during all experiments.

Table 1 summarizes experimental conditions including mixing ratios of aromatic precursors (i.e., toluene, m-xylene, and 1,2,4-TMB), NO_x , and RH.

2.2 Field measurements

A field campaign was conducted from 14 December 2022, to 2 February 2023, at the Dianshan Lake (DSL) Air Quality Monitoring Supersite in suburban Shanghai, China (31.10° N, 120.98° E). This monitoring site is frequently impacted by regional transport and experiences episodes of an-

Table 1. Summary of experimental conditions in the laboratory experiments.

Exp.	Precursor	Initial precursor concentration (ppb)	Initial NO ^a (ppb)	Estimated Cl exposure $(\times 10^9 \text{ molec. cm}^{-3} \text{ s})$	RH ^b (%)	Non-Cl-OOMs molar yield (%)	Cl-OOMs molar yield (%)	Ratio (Cl-OOMs/ Total OOMs, %)	Ratio (Dimer/ Monomer, %) ^c
1	Toluene	80	0	2.0	< 1	5.1	2.4	32	3.5
2	Toluene	80	0	1.2	60	0.7	0.3	31	2.0
3	Toluene	84	45	2.7	< 1	6.8	2.6	28	1.7
4	Toluene	80	40	1.6	60	1.8	0.8	29	1.0
5	m-Xylene	87	0	2.0	< 1	0.6	0.5	43	3.2
6	m-Xylene	87	0	1.6	68	0.3	0.2	44	2.6
7	m-Xylene	90	45	1.7	< 1	1.4	0.6	31	0.9
8	m-Xylene	87	50	2.3	35	1.1	0.5	32	1.2
9	1, 2, 4-TMB	98	0	1.4	< 1	0.5	0.3	34	2.1
10	1, 2, 4-TMB	103	0	1.4	30	0.3	0.2	35	1.4
11	1, 2, 4-TMB	93	40	2.2	< 1	1.5	0.7	31	0.8
12	1, 2, 4-TMB	109	55	2.2	30	1.0	0.6	37	0.6

^a In the presence of NO experiments, there is more OH chemistry was involved in Cl-aromatics reactions and its influence can lead to relatively high uncertainty in the molar yields reported in the table for pathways influenced by OH chemistry. Approximately 90 % of NO was converted into NO₂ after the UV light was turned on. ^b Relative humidity. ^c The molar yield ratios of total OOM dimers to monomers. Monomers are defined as molecules with a carbon number equal to the carbon number of aromatic precursor (nC), and dimers are defined as molecules with carbon numbers ranging from 2nC -1 to 2nC +1.

thropogenic pollution. A detailed description of this site can be found elsewhere (Wu et al., 2023; Yang et al., 2022, 2023). A Vocus-PTR-LToF and a nitrate-CI-APi-LToF were both deployed in this field campaign to detect aromatics and Cl-OOMs, respectively. The detailed description of nitrate-CI-APi-LToF and Vocus-PTR-LToF, and their calibration in the field measurement are shown in the Supplement (Text S2).

2.3 Heath effect estimation

A number of Cl-OOMs were assessed using the Estimation Program Interface Suite (EPI, V.4.11, US EPA [code]) and Toxicity Estimation Software Tool (T.E.S.T., V.5.1.2, US EPA [code]) software provided by the United States Environmental Protection Agency (EPA), to estimate their persistence, bioaccumulation, and toxicity through calculated half-life for reactions with OH, bioconcentration factors (BCF), oral rat pLD₅₀ ($-\log_{10}(\text{pred})$, mol kg $^{-1}$), developmental toxicity, and mutagenicity. The models utilized the SMILES (Simplified Molecular Input Line Entry System) notation of the target compounds as input for the prediction.

3 Results and discussion

3.1 OOM molar yields

The molar yields of OOMs are determined as OOMs formed $(\Delta M, \text{ molar cm}^{-3})$ divided by precursor reacted $(\Delta \text{Ar, molar cm}^{-3})$:

$$Molar yield = \frac{\Delta M}{\Delta Ar}$$
 (3)

Table 1 summarizes non-Cl-OOMs' and Cl-OOMs' molar yields and OOMs' dimer-to-monomer ratios from our laboratory experiments. The molar yields of non-Cl-OOMs and Cl-OOMs from reactions of three aromatic precursors with Cl

atoms in the absence of NO_x are within the ranges of 0.3 %-5.1 % and 0.2 %–2.4 %, respectively. These values are comparable with previously reported: HOM molar yields (0.8 %– 4.0 %) detected by nitrate-CI-APi-LToF for the reactions between α -pinene and Cl atoms (Wang et al., 2020) and non-Cl-OOMs molar yields (4.4 %–8.8 %) detected by H₃O⁺-Chemical Ionization Mass Spectrometry(H₃O⁺-CIMS) for the reactions between m-xylene and Cl atoms (Bhattacharyya et al., 2023). The low molar yields observed in both this study and previous studies may be attributed to the preference of different detection techniques. Nevertheless, the observed ratio of Cl-OOMs to non-Cl-OOMs, ranging from 29 % to 44 % (Table 1), further indicates the non-negligibility of the Cl-OOMs products among the total OOMs products. In the presence of $40-55 \text{ ppb NO}_x$, our flow tube experiments show that the molar yields of non-Cl-OOMs and Cl-OOMs from three aromatics are within the ranges of 1.0 %-6.8% and 0.5%–2.6%, respectively. The addition of NO_x can slightly increase the molar yields of both non-Cl-OOMs and Cl-OOMs.

Typically, the fate of peroxy radicals in flow tube experiments is largely influenced by their reactions with other RO₂, HO₂, or NO species, which are contingent upon the specific experimental conditions (Bianchi et al., 2019; DeMore et al., 1997). In experiments without NO_x, while the termination of RO₂ was primarily anticipated to be governed by RO₂-RO₂ and HO₂-RO₂ reactions. However, after the addition of NO_x, the reaction between RO₂ and NO tended to predominate, leading to a reduction in the dimer-to-monomer ratio by 50.5 % (Exp. 3 and Exp. 1 in Table 1).

In contrast to OH reaction, the presence of Cl, ClO, and Cl₂ in chlorine-involved reaction may introduce additional reaction pathways related to Cl-OOM formation. However, these pathways make only minor contributions to the formation of Cl-OOMs under our experimental conditions (see Text S4 for detailed analysis).

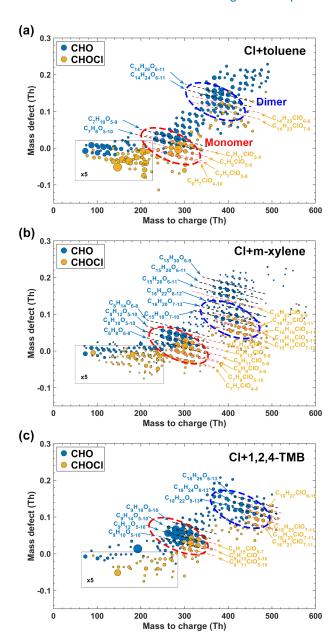


Figure 1. Mass defect plots of OOM products detected by a nitrate-CI-APi-LTOF from Cl-initiated reactions of (a) toluene, (b) m-xylene, and (c) 1,2,4-trimethylbenzene, respectively, in the absence of NO_X . The detected products are marked by their exact mass (with NO_3^- reagent ions) and mass defect (exact mass subtracted by its unit mass). The lines annotate the general chemical formulae. Chlorine-containing and non-chlorine-containing formulae are shown in different colors. The size of the circle corresponds to the concentration of products.

Although both the non-Cl-OOMs and Cl-OOMs molar yields increased in the presence of NO_x , the ratio of Cl-OOMs to the total OOMs decreased when NO_x was added. This phenomenon can be attributed to the significant contribution of OH chemistry, primarily resulting from the NO +

HO₂ reaction. The additional OH radicals can contribute to the formation of non-Cl-OOMs, as supported by recent experimental and modeling studies on the reaction dynamics between Cl atoms and isoprene (Wang et al., 2022). For our experiments, as NO_x was added into the flow tube, the concurrent presence of Cl atoms, OH radicals, and aromatics led to a notable increase in the total OOM molar yields, ranging from 25 % to 187 %, while reducing the ratios of Cl-OOMs to the total OOMs by approximately 9 %-28 % (Table 1). In addition, the increased Cl-OOM yields under NO_x conditions may result from the suppression of dimer or multimer formation (Table 1), which shifts the product distribution toward monomeric Cl-OOMs and thus leads to potential higher apparent yields. Moreover, NO_x-promoted chemistry facilitates the formation of nitrogen-containing OOMs with diverse structures and functionalities. The nitrate-CI-APi-LToF exhibits different detection sensitivities toward these species, which may also influence the estimated molar yields under high NO? conditions. Besides, high RH leads to low molar yields, which may be attributed to the depressed detection efficiency of OOMs and the elevated vapor wall loss under humid conditions (Huang et al., 2018). The product distribution remains unchanged under high humidity conditions (see Figs. 1b and S4), indicating that the presence of water does not significantly influence the reaction between Cl atoms and aromatics.

3.2 Characteristics of OOM products and peroxyl radicals

Mass defect plots of stabilized products from reactions between toluene (Exp. 1 in Table 1), m-xylene (Exp. 5 in Table 1), and 1,2,4-TMB (Exp. 9 in Table 1) and Cl atoms in the absence of NO_x are shown in Fig. 1. These products display similar distribution patterns, consisting of monomers (with carbon numbers equal to carbon numbers of the precursor, nC) and dimer products (with carbon numbers ranging from 2nC - 1 to 2nC + 1). The ratio of dimer products to monomer products are 3.5 %, 3.2 %, and 2.1 % for reactions of toluene, m-xylene, and 1,2,4-TMB with Cl atoms, respectively. Meanwhile, the products also can be classified into two groups: non-Cl-OOMs in blue and Cl-OOMs in orange in Fig. 1. In general, the total concentration of Cl-OOMs is lower than that of non-Cl-OOMs. Specifically, the concentrations of Cl-OOMs account for 47 %, 91 %, and 52% of the non-Cl-OOMs in the toluene, p-xylene, and 1,2,4-trimethylbenzene experiments, respectively. Both non-Cl-OOMs and Cl-OOMs products can be categorized into several bands, as indicated by the dashed lines in Fig. 1, each of which comprises compounds with varying numbers of oxygen atoms.

The mass spectrum of the detected OOMs monomer and dimer products from the reaction between m-xylene and Cl atoms with and without NO_x (Exp. 5 and 7 in Table 1) are shown in Fig. 2. Without NO_x , dominant monomer prod-

ucts for non-Cl-OOMs included $C_8H_{10}O_6$ and $C_8H_{12}O_{6-8}$, whereas for Cl-OOMs, $C_8H_{11}ClO_{6-8}$ and $C_8H_{13}ClO_{6-7}$ dominated. The most abundant dimer compounds for non-Cl-OOMs were $C_{16}H_{20}O_9$, $C_{16}H_{22}O_{10}$, and $C_{15}H_{18}O_8$, whereas for Cl-OOMs, $C_{16}H_{19}ClO_{8,10}$, $C_{16}H_{21}ClO_9$, and $C_{15}H_{17}ClO_8$ prevailed. Dimer products containing two Cl atoms were also observed, exemplified by $C_{16}H_{18,20}Cl_2O_{10}$. Under NO_x -present conditions, the main non-Cl-OOMs included $C_8H_{11}NO_{6-10}$ and $C_8H_{12}N_2O_{10}$, whereas Cl-OOMs were represented by $C_8H_{10}ClNO_{7-8}$ and $C_8H_{11}ClN_2O_{9-10}$.

Moreover, 29 peroxyl radicals in total were observed in our flow tube experiments with three precursors, as listed in Table S1 and illustrated in Figs. 3 and S5. It is crucial to note, however, that not every intermediate radical formed could be conclusively identified. This is primarily due to the inherent instability and the extremely brief life span of peroxyl radicals, which pose significant challenges for their detection. Currently, the absolute concentrations of radicals measured by I-CIMS cannot be reliably quantified. Therefore, for visualization and comparative purposes, the signal intensities of the radicals C₈H₁₀ClO₂₋₃ detected by I-CIMS were normalized to match the signal intensity of C₈H₁₀ClO₄ measured by nitrate-CI-APi-LToF, as presented in Fig. 3. It is essential to note that this normalization was solely used to facilitate consistent visual comparison across instruments. Consequently, the normalized I-CIMS signals are not utilized for quantitative comparisons of radical concentrations in the subsequent discussion.

Take the reaction between Cl atoms and m-xylene in the absence of NO_x for example, we detected peroxyl radicals including $C_8H_{10}ClO_{2-7}$, $C_7H_8ClO_{5-7}$, $C_8H_9O_6$, and C₇H₇O₆ (Figs. 3 and S5). High-resolution peak fittings of partial peroxy radicals from the raw mass spectrum obtained by nitrate-CI-APi-LToF are shown in Fig. S6. Notably, the dominant species among these radicals was C₈H₁₀ClO₆, constituting 27.4% of the total signal of identified peroxyl radicals in the reaction of Cl atoms and m-xylene and nearly three times of the C₈H₉O₆ radical signal. Also, similar ratio rules of chlorine-containing radicals (analogy of $C_8H_{10}ClO_6$) to non-chlorine-containing radicals (analogy of $C_8H_9O_6$) were observed for toluene (~ 4.2 times) and 1,2,4-TMB (~ 3.8 times). Although C₇H₈ClO₅₋₇ radicals were also discernible, their signal intensity was merely 22.3 % of the total signal of identified peroxyl radicals in the reaction of Cl atoms and m-xylene. Non-Cl-containing radicals were also detected, albeit with substantially lower signal values, accounting for only 15.5 % of the total identified radical signals.

3.3 Formation mechanisms of CI-OOMs

3.3.1 Cl-addition pathway

Given the similar product patterns for the three precursors (toluene, m-xylene, and 1,2,4-TMB), it is reasonable to in-

fer that the reaction mechanisms between different aromatics and Cl atoms are analogous. A generalized mechanism is thus proposed, elucidating the Cl-initiated reactions of mxylene in the absence of NO_x , as depicted in Scheme 1. This scheme serves as a representative example highlighting the potential pathways involved in the Cl-initiated reactions of aromatics.

Although the theoretical studies by Huang et al. (2012) show that the Cl-addition pathway accounts for only 14 % of Cl-initiated reactions of toluene (298 K), which is significantly lower than the 86% attributed to the H-abstraction pathway, the proportion of Cl-OOMs products from Claddition reaction should not be overlooked, as evidenced by the observation of peroxyl radicals and product distribution characteristics in this study (Huang et al., 2012). The initial reaction of m-xylene (C₈H₁₀) with Cl atoms can occur through two pathways: the Cl-addition pathway, leading to the formation of a C₈H₁₀Cl radical (C₈H₁₀Cl•, refer to Scheme 1) or the H-abstraction pathway, forming a C₈H₉ radical (C₈H₉•, refer to Scheme S1) and C₈H₉ radical (C₈H₉O₂• refer to Scheme S1). Then, both C₈H₉• and C₈H₁₀Cl· can in turn undergo autoxidation via the H-shift and addition of O2 to produce peroxy radicals of C8H10O5. or C₈H₁₀ClO₆ · (Vereecken and Nozière, 2020; Bianchi et al., 2019). The peroxyl radical $C_8H_{10}ClO_6$ was identified as the predominant species in terms of signal (Table S1). Although these Cl-OOMs-to-non-Cl-OOMs signal ratios may not accurately represent their relative concentrations due to sensitivity differences of these radicals towards the reagent ions $((HNO_3)_{0-1} \cdot NO_3^-)$, it is still noteworthy that these Cl-RO₂ overlooked in previous studies were directly observed in such a reaction system, thereby suggesting that Cl-addition pathway is indeed present in the initial reaction steps of reactions between Cl atoms and aromatics.

In the reaction of Cl atoms and m-xylene, a certain fraction of the peroxyl radicals (C₈H₁₀ClO₆•) might also be derived from the secondary Cl-addition reaction between Cl atoms and a first-generation stabilized product C₈H₁₀O₆ (Scheme 1). Indeed, it is challenging to evaluate the exact contribution from secondary Cl-addition reaction to the formation of C₈H₁₀ClO₆•. However, it may be indirectly assessed via the potential secondary reactions between more dominant first-generation stabilized products ($C_8H_{12}O_x$) and Cl (Table S2), since $C_8H_{10}O_6$ and $C_8H_{12}O_x$ likely react with Cl at similar rates. If secondary Cl-addition reactions were significant in the reaction system of Cl atoms and m-xylene, $C_8H_{12}O_x$ should undergo secondary Cl-addition reactions to generate $C_8H_{12}ClO_x$ radicals. Yet, $C_8H_{12}ClO_x$ radicals were not detectable, which hints that secondary Cl-addition reactions could only play a minor role in our experiments. Therefore, it appears that the secondary Cl-addition reactions between stabilized products and Cl atoms are less significant compared with Cl-addition in the initial reaction steps.

Recently, Jahn et al. (2024) investigated the formation of secondary organic aerosols from the reaction of ethylbenzene

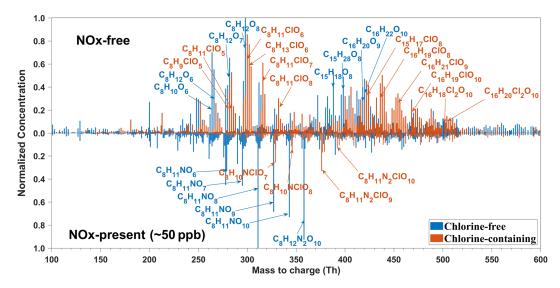


Figure 2. Mass spectra of OOM products detected by a nitrate-CI-APi-LToF from the reaction of m-xylene and Cl atoms under NO_x -free and NO_x -present conditions. The y axes in both figures are standardized by setting their maximum concentrations to 1.

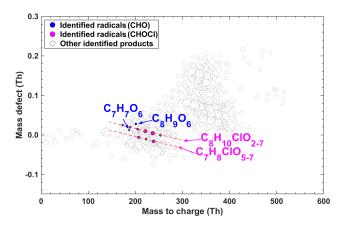


Figure 3. Mass defect plot of peroxyl radicals detected by a nitrate-CI-APi-LToF and an I-CIMS from Cl-initiated reactions of m-xylene without NO_x . $C_8H_{10}ClO_{2-3}$ were detected by I-CIMS while the remaining radicals were observed using nitrate-CI-APi-LToF. The detected products are marked by their exact mass (with NO_3^- reagent ions) and mass defect (exact mass subtracted by its unit mass). The lines annotate the general chemical formulas. Chlorine-containing and non-chlorine-containing formulas are shown in different colors. The size of the circle is proportional to the concentrations of peroxyl radicals.

and Cl atoms, and attributed the observed Cl-addition products to reactions involving non-aromatic C=C bonds (Jahn et al., 2024). They claimed that OH radical existed in their experiments, due to the presence of NO_x . As a result, approximately 40% of ethylbenzene reacted with Cl atoms, 30% with OH, and 30% remained unreacted. Thus, a secondary Cl addition to $C_8H_{11}NO_6$ for reaction of ethylbenzene and Cl atoms, forming $C_8H_{12}ClNO_8$, is possible (Scheme S1). In our experiments, although OH radicals can still be gen-

erated through H-abstraction from methyl groups by HO₂ (Bhattacharyya et al., 2023), the use of 20 ppm CO as an OH scavenger confirmed the presence of OH chemistry in our system. Nevertheless, the stable real-time signals of firstgeneration products and radicals from the Cl-addition reaction, observed both with and without OH chemistry, provide compelling evidence that the Cl-addition pathway remains the dominant mechanism for Cl-RO₂ formation under low-NO? conditions (see Text S5). Moreover, under high NO? conditions in our experiments, the dominant Cl-OOMs C₈H₁₀ClNO₇ (Scheme S2), contains two fewer H atoms than the C₈H₁₂ClNO₈ proposed by Jahn et al. (2024) as a secondary Cl-addition product (Scheme S3). While this difference might suggest the presence of an additional double bond or a ring in C₈H₁₀ClNO₇, it could also result from multigeneration oxidation chemistry. Specifically, H-abstraction from a carbon bearing an OH or OOH can form a carbonyl, thereby reducing the hydrogen count without requiring a new double bond.

3.3.2 Autoxidation and subsequent reactions of Cl-RO₂

Scheme 2 shows a proposed reaction mechanism for the autoxidation of the main Cl-RO₂ radical ($C_8H_{10}ClO_4$ •) generated from Cl atoms and m-xylene. Autooxidation of $C_8H_{10}ClO_4$ • leads to $C_8H_{10}ClO_6$ •. There are two distinct isomeric forms of $C_8H_{10}ClO_6$ •, denoted as $C_8H_{10}ClO_6$ • (I) and $C_8H_{10}ClO_6$ • (II), which is similar to results for aromatics + OH by Molteni et al. (2018). Notably, the formation of $C_8H_{10}ClO_6$ • (II) requires a second step of endocyclization, which is not competitive on account of its slow reaction rate, as inferred from several previous studies using both experimental and theoretical approaches of OH-initiated oxidation of aromatics (Wang et al., 2017; Xu

Scheme 1. Proposed reaction mechanisms of m-xylene with Cl atoms leading to the formation of Cl-OOMs. Radicals detected by nitrate-CI-APi-LToF are marked with black boxes.

et al., 2020). Therefore, the abundance of $C_8H_{10}ClO_6 \cdot$ (I) would likely be much higher than $C_8H_{10}ClO_6 \cdot$ (II), and our following discussion primarily focuses on the subsequent reaction processes involving $C_8H_{10}ClO_6 \cdot$ (I). Briefly, $C_8H_{10}ClO_6 \cdot$ reacts with $RO_2 \cdot$ to generate $C_8H_{9,11}ClO_5$ and $C_8H_{10}ClO_5 \cdot$ (Reaction pathway R2b) and with an HO₂ radical to produce $C_8H_{11}ClO_6$ (Reaction pathway R2c). Meanwhile, $C_8H_{10}ClO_6 \cdot$ has no more H atoms for another H-shift at an appreciable rate based on our current understanding. (Wang et al., 2024) In the presence of NO_x , $C_8H_{10}ClO_6 \cdot$ is terminated (Reaction pathway R2a), leading to the formation of $C_8H_{10}ClNO_7$ products.

3.3.3 Dimer formation

The accretion reaction $(RO_2 + R'O_2 \rightarrow ROOR' + O_2)$ represents a pivotal source for dimer compounds, originating from highly oxidized and functionalized RO_2 radicals (Ehn et al., 2014; Zhao et al., 2018). As shown in Fig. 2, $C_{16}H_{19}ClO_8$ and $C_{16}H_{19}ClO_{10}$ are two typical accretion reaction products from reactions between Cl atoms and m-xylene without NO_x . In detail, $C_{16}H_{19}ClO_{10}$ can be formed through the accretion reaction between $C_8H_{10}ClO_6$ • and $C_8H_9O_6$ • (Scheme 3). The formation pathways for $C_{16}H_{19}ClO_8$ are more varied compared to that of $C_{16}H_{19}ClO_{10}$. It can be produced either via the accretion of a $C_8H_{10}ClO_6$ radical with a $C_8H_9O_4$ radical or through the reaction of a $C_8H_{10}ClO_4$ radical with a $C_8H_9O_6$ radical. Meanwhile, $C_{16}H_{20}Cl_2O_{10}$, which is also detected during the reaction, can be formed via the accretion reaction of two $C_8H_{10}ClO_6$ radicals.

Previous studies have reported dimer formation rates during the OH-initiated oxidation of 1,3,5-trimethylbenzene, ranging from 1.4×10^{-10} to 25×10^{-10} cm³ molec. $^{-1}$ s $^{-1}$ (Berndt et al., 2017). Compared to the OH reaction, Cl-RO₂ is produced in the Cl atom reaction, but its reaction rate in the accretion reaction remains unknown. If Cl-RO₂ have lower reactivity (i.e., slower rate coefficients) for accretion reactions compared to non-Cl-containing RO₂, the same generation rate of Cl-RO₂ and RO₂ would result in higher concentration of Cl-RO₂, resulting in the high Cl-RO₂ signal detected. However, the generation rate of both Cl-RO₂ and RO₂ are currently uncertain, further complicating the determination of the chemical mechanisms of Cl-initiated reactions even more challenging.

3.4 CI-OOMs detection in suburban Shanghai

As shown in Fig. 4a and Table S3, a total of 51 gaseous Cl-OOMs were identified during winter in suburban Shanghai, whose high-resolution peak fittings measured by nitrate CI-APi-LToF are shown in Fig. S7. Figure S8 further demonstrates the accuracy of our peak identification by showing the ratio of the fitted peak intensities versus the peak separation for identified Cl-OOM peaks and adjacent ions from our ambient nitrate-CI-APi-LToF measurements (Cubison and Jimenez, 2015). The peak separation, normalized to the half-width at half-maximum ($\chi = \Delta t/HWHM$), is greater than 1 for all Cl-OOMs. Notably, 22% of the peaks exhibit χ values between 1 and 2, indicating they are separable but closely spaced, while 78% of the Cl-OOMs ($\chi > 2$) are well-separated peaks.

Ambient Cl-OOMs consist of compounds mostly with a single Cl atom and only two with two Cl atoms. The carbon and oxygen numbers of these Cl-OOMs ranged from 2 to 14 and 1 to 11, respectively. Figure 4b shows the abundance distribution of gaseous Cl-OOMs on the basis of carbon and oxygen numbers in their molecular formulae. Approximately 80 % concentration of the identified Cl-containing molecules is C5-C9 Cl-OOMs, among which C6-C9 Cl-OOMs represent a large fraction of $\sim 79\%$. All assignments followed strict criteria for mass accuracy (<4 ppm) and isotopic pattern matching. Interference screening with adjacent formulas (Table S4) was performed to ensure unambiguous identification. Nonetheless, the detection of these low-oxygenated Cl-OOMs should be interpreted with caution, and we acknowledge that their measured concentrations may carry higher uncertainty compared to those of more oxygenated compounds.

38 Cl-OOMs observed in field measurements were also identified in our laboratory experiments, corresponding to reaction products of Cl atoms with toluene, m-xylene, or 1,2,4-TMB. These results indicate that ambient C6–C9 Cl-OOMs exemplified by C₇H₇N₂ClO₉, C₈H₁₀NClO₇, and C₉H₁₅ClO₈ are likely formed from Cl-initiated reactions with aromatic compounds. On the other hand, isoprene was considered as a precursor of C5 Cl-OOMs (Le Breton et al., 2018; Priestley et al., 2018; Wang and Ruiz, 2017). Several relatively abundant ambient Cl-OOMs, specifically C₂H₂Cl₂O₂, C₆H₄ClNO₃₋₄, C₈H₇ClO₆, and C₉H₆ClNO (Fig. 4a), although they were not observed in our flow tube experiments. The presence of C₂H₂Cl₂O₂ suggests possible contributions from the oxidation or atmospheric degrada-

Scheme 2. Reaction pathways of the bicyclic peroxyl radical $C_8H_{10}ClO_4$ in the Cl-initiated reaction of m-xylene. Blue, brown, and black formulae denote radicals, nitrogen-containing products, and Cl-OOMs products, respectively. Radicals and stable products detected by nitrate-CI-APi-LToF are marked with black boxes.

Scheme 3. Accretion reaction pathways of peroxyl radical $C_8H_{10}ClO_6$ • and $C_8H_9O_6$ • in the Cl-initiated reaction of m-xylene. Blue and black formulae denote radicals and stable products, respectively. Radicals and stable products detected by nitrate-CI-APi-LToF are marked with black boxes.

tion of chlorinated solvents, combustion emissions, or industrial processes (Wang et al., 2021). The $C_6H_4ClNO_{3-4}$ may originate from Cl-induced reaction of aromatics. Similarly, $C_8H_7ClO_6$ and C_9H_6ClNO could be derived from more oxidized aromatic compounds undergoing multi-step oxidation or fragmentation. The observation of these species highlights the complexity and diversity of Cl chemistry in suburban environments, underscoring the importance of further studies integrating broader precursor sets, longer aging times, and additional reaction pathways to better understand their sources and formation mechanisms.

The diurnal profiles of all 51 Cl-OOMs identified in the ambient were shown in Fig. S9. Similar to previous reports in northern Europe and Beijing (Le Breton et al., 2018; Priestley et al., 2018), Cl-OOMs increased with elevated solar radiation and their peaks appeared at around 12:00 p.m. (local time). This suggests that while ClNO₂ photolysis is a significant source of Cl atoms in the early morning, and other sources such as the photolysis of Cl₂, ClONO₂, HCl, ICl, and BrCl by sunlight can also contribute to Cl atom concentrations later in the day (Peng et al., 2020). Besides, some Cl-OOMs could be formed through secondary reactions involving OH radicals with Cl-VOCs or intermediates, rather than direct Cl atom-initiated reaction. While distinguishing

these pathways is beyond the scope of this study, the formation of Cl-OOMs is likely influenced by both Cl- and OH-initiated mechanisms under ambient conditions, especially in the presence of NO_x .

3.5 Health effect of ambient CI-OOMs

Table S5 presents a detailed toxicity estimation of various Cl-OOMs, including C₇H₇N₂ClO₉, C₈H₁₀NClO₇, and C₉H₁₅ClO₈, which were identified in the suburban Shanghai atmosphere and laboratory experiments. These estimations are based on the potential chemical structures of these Cl-OOMs. We characterized the toxicity of these Cl-OOMs using persistence, bioaccumulative and toxic (PBT) criteria, as outlined in Table S5. Additionally, we conducted a comparative analysis of their toxicity relative to that of naphthalene, a compound whose atmospheric toxicity has been extensively studied.

Compounds with short OH radical reaction half-lives and bioconcentration factors (BCF) are unlikely to be persistent and bioaccumulative. Among these Cl-OOMs, $C_8H_{10}NClO_7$ (II) exhibits the slowest degradation rate with OH radicals with a half-life of 5 d, which suggests that it is a potential

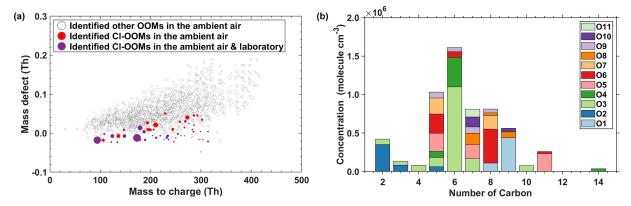


Figure 4. (a) Mass defect plot of detected OOMs by a nitrate-CI-APi-LToF in ambient air in suburban Shanghai. Cl-OOMs only observed in the ambient atmosphere were marked in red, and Cl-OOMs observed in both ambient and lab were marked in purple. The size of the circle is proportional to the concentration of compounds in the ambient. (b) Distribution of carbon and oxygen number of gas-phase Cl-OOMs identified in suburban Shanghai. The color codes correspond to the number of oxygen atoms.

new persistent pollutant (half-life > 2 d) in the atmosphere (Europe, 2023).

As indicated by the pLD $_{50}$ values, toxicity levels of most Cl-OOMs are lower than that of naphthalene, except for C $_8H_{10}NClO_8$, C $_8H_{10}NClO_7$ (II), and C $_7H_7N_2ClO_8$ (Table S5). Among these compounds, C $_8H_{10}NClO_7$ (I) is identified with the highest predicted toxicity, followed by compound C $_8H_{10}NClO_8$. The pLD $_{50}$ of these compounds is categorized within level 3, signifying their potential for a considerable acute toxicity.

Notably, the pLD $_{50}$ values of Cl-OOMs are found to be akin to those of naphthalene. However, given that their concentration in suburban Shanghai (0.55 ppt) is merely one percent of that of naphthalene (50 ppt) as measured in this study, the total probabilistic hazard quotient (PrHQ, defined as the product of estimated human exposure (ambient concentration) and pLD $_{50}$ in this study) of Cl-OOMs is lower. Despite this lower PrHQ, it is important to recognize that each of the evaluated Cl-OOMs may pose risks of developmental toxicity and mutagenicity, which underscores the need for a thorough understanding of the toxicological implications of Cl-OOMs in the atmosphere.

4 Conclusions

In summary, this study highlights the Cl-addition-initiated reaction as a non-negligible pathway in the reaction of Cl atoms with aromatics. Field measurements in suburban Shanghai revealed 51 gaseous Cl-OOM species, with 38 of these Cl-OOMs also detected in our laboratory experiments. This suggests that these Cl-OOMs likely derive from reactions between Cl atoms and aromatics. Considering the significant role of Cl atoms in daily atmospheric oxidation processes, overlooking the Cl-addition pathway could lead to ignoring the formation of Cl-OOMs from aromatic compounds in the atmosphere. This study, therefore, highlights the neces-

sity of incorporating both pathways in the models for a more accurate assessment of the atmospheric fate of Cl atoms and aromatics in urban settings. In addition, health effect evaluation indicates that all assessed Cl-OOMs may possess developmental toxicity, and nearly half of the compounds may exhibit carcinogenic effects. Considering the critical role of aromatics in the urban air and recent observations reporting increased levels of reactive chlorine species in polluted atmospheres, our study offers timely insights into the chemical processes between Cl atoms and aromatics occurring in anthropogenically influenced atmospheres and the adverse health effects of these Cl-containing reaction products.

Data availability. The data of this study are available at https://doi.org/10.5281/zenodo.16787937 (Li, 2025) and other data related to this study are available upon request from Lei Yao (lei_yao@fudan.edu.cn) and Lin Wang (lin_wang@fudan.edu.cn).

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Author contributions. LY and LW conceived and designed this study and revised the manuscript. CL analyzed and interpreted data, drafted and revised the manuscript. CL and YW contributed to the modeling of the data. MF and XC contributed to the health effect analysis.

Competing interests. The contact author has declared that none of the authors has any competing interests.

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