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Atmospheric data support a multi-decadal shift in the global methane budget towards natural tropical emissions

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Abstract. We use the GEOS-Chem global 3-D model and two inverse methods (the maximum a posteriori and ensemble Kalman filter) to infer regional methane (CH₄) emissions and the corresponding stable-carbonisotope source signatures from 2004-2020 across the globe using in situ and satellite remote sensing data. We use the Siegel estimator to determine linear trends from the in situ data. Over our 17-year study period, we estimate a linear increase of $3.6 \text{ Tg yr}^{-1} \text{ yr}^{-1}$ in CH₄ emissions from tropical continental regions, including North Africa, southern Africa, tropical South America, and tropical Asia. The second-largest increase in CH₄ emissions over this period $(1.6 \text{ Tg yr}^{-1} \text{ yr}^{-1})$ is from China. For boreal regions we estimate a negative emissions trend of $-0.2 \text{ Tg yr}^{-1} \text{ yr}^{-1}$, and for northern and southern temperate regions we estimate trends of $0.03 \text{ Tg yr}^{-1} \text{ yr}^{-1}$ and $0.2 \text{ Tg yr}^{-1} \text{ yr}^{-1}$, respectively. These increases in CH₄ emissions are accompanied by a progressively isotopically lighter atmospheric δ^{13} C signature over the tropics, particularly since 2012, which is consistent with an increased biogenic emissions source and/or a decrease in a thermogenic/pyrogenic emissions source with a heavier isotopic signature. Previous studies have linked increased tropical biogenic emissions to increased rainfall. Over China, we find a weaker trend towards isotopically lighter $\delta^{13}C$ sources, suggesting that heavier isotopic source signatures make a larger contribution to this region. Satellite remote sensing data provide additional evidence of emissions hotspots of CH₄ that are consistent with the location and seasonal timing of wetland emissions. The collective evidence suggests that increases in tropical CH₄ emissions are from biogenic sources, with a significant fraction from wetlands. To understand the influence of our results on changes in the hydroxyl radical (OH), we also report regional CH₄ emissions estimates using an alternative scenario of a $0.5 \% \text{ yr}^{-1}$ decrease in OH since 2004, followed by a larger 1.5 % drop in 2020 during the first COVID-19 lockdown. We find that our main findings are broadly insensitive to those idealised year-to-year changes in OH, although the corresponding change in atmospheric CH₄ in 2020 is inconsistent with independent global-scale constraints for the estimated annual-mean atmospheric growth rate.

1 Introduction

Changes in atmospheric methane (CH₄) over the last few decades have unfolded without clear explanation, exposing inadequacies in our measurement coverage and our ability to definitively attribute those changes to individual emissions and losses. The climatic importance of atmospheric CH₄ lies in its ability to absorb and emit infrared radiation at wavelengths that are relevant to outgoing terrestrial radiation and incoming shortwave radiation (Allen et al., 2023). Consequently, atmospheric CH₄ helps to maintain Earth's radiative balance and surface and atmospheric temperatures. Atmospheric CH₄ is derived from emissions due to thermogenically (organic matter broken down at high temperatures and pressures, mainly released during extraction and transport of fossil fuels), pyrogenically (through incomplete combustion of organic matter), and biogenically (microbial activity) based production pathways. The main loss process is from from the hydroxyl radical (OH), with minor losses from the reaction with chlorine, uptake from soils, and stratospheric loss. Methane is the second-most abundant anthropogenic greenhouse gas in terms of its anthropogenic radiative forcing. The global CH₄ growth rate was close to zero from 2000 to 2006 (Dlugokencky et al., 2020) but has since accelerated, with a global annual growth rate reported by the NOAA exceeding 15 ppb for the first time in 2020 and more than 18 ppb in 2021 (Feng et al., 2023). Concurrently, we are witnessing a progressively isotopically lighter signature of globally averaged CH₄ (more negative global-average atmospheric δ^{13} C value). Analysis of CH₄ mole fraction and δ^{13} C–CH₄ data suggests that thermogenic sources are unlikely to be the dominant driver of the post-2006 global-mean increase in atmospheric CH₄ (Lan et al., 2021). A growing body of work has proposed a range of hypotheses to explain short periods of observed global and regional variations in atmospheric CH₄ (Turner et al., 2019). In this study, we take a step back to look at observed CH₄ variations from 2004 to 2020 in order to capture some of the zero-growth-rate period and the subsequent increase in growth rate of CH₄ post-2007. We argue that monthly variations are part of a large-scale shift in predominately thermogenic energy emissions from high northern latitudes to biogenic emissions from the tropics, driven by larger emissions over tropical North Africa and tropical South America.

The post-2007 increase in atmospheric CH₄ has been the focus of many studies and has been attributed to different plausible hypotheses associated with changes in various emissions sources and the OH sink (Turner et al., 2019). These studies have reached their conclusions using in situ mole fraction observations alone or in combination with other observations, e.g. in situ δ^{13} C (Schaefer et al., 2016; Rice et al., 2016; Nisbet et al., 2016; Fujita et al., 2020; Lan et al., 2021; Basu et al., 2022; Oh et al., 2022), satellite observations (Worden et al., 2017; McNorton et al., 2018; Yin et al., 2021; Feng et al., 2022), or other trace gases, using a

variety of analysis methods and computational models. Typical emissions sizes and uncertainties are indicated in Table 1, adapted from Saunois et al. (2020). Our approach is unique in that, for our δ^{13} C inversion, we are solving for the δ^{13} C isotopic source signature of a geographical region. From the isotopic source signature of a region, we can determine how the source balance within a particular region has shifted over time, e.g. larger or smaller contributions from pyrogenic and biogenic sources, and consequently gain understanding of the geographical shifts in the CH₄ budget.

Methane oxidation by the OH radical in the troposphere is responsible for 80% of the total CH₄ sink globally. Changes in OH may have played a role in recent changes in atmospheric CH₄ (Rigby et al., 2017; Turner et al., 2017), but the magnitude of this influence is uncertain (its short atmospheric lifetime of < 1 s makes direct measurement of global variability very difficult). Reducing values of OH effectively increases atmospheric CH₄ and therefore has the same effect as increasing emissions of CH₄. Chemical reactions responsible for removing CH₄ from the atmosphere are faster for lighter isotopologues of CH₄. This isotopic fractionation therefore leads to an atmosphere enriched in heavier isotopes relative to the globally emitted CH₄. Lan et al. (2021) simulated CH₄ and δ^{13} C in a 3-D chemistry transport model covering the period 1984-2016 and found that changes in OH proposed by Turner et al. (2017) are not consistent with the trend of increasingly isotopically light δ^{13} C observed in the atmospheric record. We explore the impact of reducing OH in a sensitivity study, taking into account a larger OH decrease during 2020 (Peng et al., 2022; Feng et al., 2023) that was associated with widespread reductions in nitrogen oxide emissions (Cooper et al., 2022).

Here, we calculate trends in regional CH₄ emissions and isotopic δ^{13} C source signatures across the world from 2004– 2020 using in situ mole fraction and δ^{13} C data and satellite column-averaged dry-air mole fraction data. This is achieved by using three sets of inversions: two maximum a posteriori inversions using ground-based data (solving separately for regional emissions and isotopic sources signatures) and an ensemble Kalman filter inversion using Greenhouse gases Observing SATellite (GOSAT) data (solving for regional CH₄ emissions).

In the next section, we describe the data and methods we use to quantify changes in regional CH₄ emissions and the corresponding regional stable isotope source signatures. In Sect. 3, we report our results of a posteriori regional CH₄ fluxes and regional δ^{13} C isotopic signatures, including analysis of sensitivity calculations that involve different assumptions about year-to-year changes in the OH sink. We conclude the paper in Sect. 4. **Table 1.** Global-mean emissions of different CH_4 source types from bottom-up inventories (Saunois et al., 2020) and our a posteriori emissions estimates and the corresponding conventional isotope ratios' signatures (Sherwood et al., 2017). Uncertainties are shown as max-min values in square brackets.

Source type	Annual-mean emission (Saunois et al., 2020) 2003–2012 (Tg/CH ₄)	Annual-mean emission (this study) 2004–2020 (Tg/CH ₄)	Isotopic ratio δ^{13} C (‰)
Gas and oil	80 [68–92]	82.0	-44.0 [±10.7]
Coal	42 [29–61]	53.7	-49.5 [±11.2]
Livestock	111 [106–116]	115.2	$-65.4[\pm 6.7]$
Waste	65 [60–69]	67.9	$-56.0[\pm 7.6]$
Biomass burning	17 [14–26]	14.3	$-26.2 [\pm 4.8]$
Termites	9 [3–15]	11.9	$-63.4[\pm 6.4]$
Wetlands	149 [102–182]	170.9	$-61.5 \ [\pm 5.4]$ (tropical)
			-71.5 [±5.4] (Arctic)
Rice	30 [25–38]	30.7	$-62.2 [\pm 3.9]$

2 Data and methods

2.1 In situ and satellite remote measurements of atmospheric methane

We use surface-level flask data as constraints on both regional CH₄ emissions and the corresponding δ^{13} methane isotopic source signatures. The CH₄ mole fraction data (version 2020-07; Dlugokencky et al., 2020) are taken from 31 National Oceanic and Atmosphere Administration -Global Monitoring Laboratory (NOAA-GML) sites around the world (Fig. 1). The data are monthly mean values that are averaged from discrete data as collected at each site; analysed at the NOAA - Earth System Research Laboratories (NOAA-ESRL) in Boulder, Colorado; and recorded according to the NOAA 2004A standard scale (Dlugokencky et al., 2005). Up to August 2019, the analysis was performed using gas chromatography (Steele et al., 1987; Dlugokencky et al., 1994, 2005), and since August 2019, cavity ring-down spectroscopy has been used (Dlugokencky et al., 2020). We also include data from a site in Siberia, Karasevoe (KRS), which is monitored by the National Institute for Environment Studies (NIES). This site was included to maximise geographical coverage of in situ data. The CH₄ mole fraction measurements from this site are continuous, covering the period 2004–2020, and were made from 65 m height (Sasakawa et al., 2010). A scale factor of 0.997 is applied to the NIES data in order to bring them into line with the NOAA 2004A scale (Zhou et al., 2009). The site constitutes part of the Japan-Russia Siberia Tall Tower Inland Observation Network (JR-STATION).

 δ^{13} C data are similarly monthly mean values that were calculated from discrete flask samples at NOAA network sites and reported on the international carbon isotope scale VPDB (Vienna Pee Dee Belemnite). Isotope ratio "delta" values represent the excess of a heavy, less abundant stable isotope (for δ^{13} C values, carbon-13) over the light, most abundant stable isotope (carbon-12) in a sample when compared to a standard. These measurements are useful as they are indicative of the source of the CH₄; biogenic sources are dominated by isotopically lighter signatures, and thermogenic sources are dominated by isotopically heavier signatures. For the NOAA network, isotopic analysis of δ^{13} C was performed at the University of Colorado Institute of Arctic and Alpine Research Stable Isotope Laboratory (CU-INSTAAR). They follow an isotope ratio mass spectrometry approach (Miller, 2002; Vaughn et al., 2004). The geographical locations of in situ measurement sites are shown in Fig. 1. These sites are a subset of the entire NOAA network's capacity for measuring CH₄ mole fractions. The sites included in the inversion (for both CH₄ and δ^{13} C) are those that cover the entire period of the inversion (2004–2020) without significant periods of measurement breaks to ensure a consistent interpretation of trends without consideration of possible biases introduced through the inclusion or exclusion of specific sites.

We also estimate CH₄ fluxes for 2010–2020 from the Japanese Greenhouse gases Observing SATellite (GOSAT) that was launched in 2009. GOSAT is in a sun-synchronous orbit with an equatorial local overpass time of 13:30. Since launch, it has provided continuous global observations of dry-air atmospheric-column-averaged CO2 (XCO2) and CH4 (XCH₄), retrieved from shortwave infrared wavelengths that are most sensitive to changes in CH₄ and CO₂ in the lower troposphere (Parker et al., 2020). We use the latest (v9) proxy XCH_4 : XCO_2 retrievals that use spectral absorption features around the wavelength of 1.6 µm (Parker et al., 2020; Palmer et al., 2021) because of the smaller bias and better global coverage than those provided by the full physics retrievals. Analyses show the precision of a single proxy retrieval is about 0.72%, with a global bias of 0.2% (Parker et al., 2011, 2015, 2020). In our calculations, we assume a higher observation uncertainty of 1.2% and deduct a globally uniform bias of 0.3 % to obtain better a posteriori agreement with the independent ground-based XCH₄ data by the Total Carbon Column Observing Network (TCCON). These uncertainties are detailed in Feng et al. (2022). To anchor



Figure 1. Map showing regions that are optimised in the CH₄ and δ^{13} C inversions in different colours. Black dots and labels show the location of ground-based measurement sites that measure CH₄ mole fraction. Red dots and labels indicate both CH₄ mole fraction and δ^{13} C measuring sites. Regions are named as follows: grey – North American boreal, yellow – North American temperate, light green – South American temperate, purple – Europe, blue – North Africa, light blue – southern Africa, pink – boreal Eurasia, orange – China, brown – India, peach – temperate Eurasia, red – tropical SE Asia, lilac – Oceania, white – oceans. Site identifiers are detailed in Table A2.

Table 2. Annual regional-mean a posteriori CH₄ emissions (Tg yr⁻¹) from 2004 to 2020. Regions include Australia (AUS), boreal Eurasia (BEUR), boreal North America (BNA), China (CHN), India (IND), Europe (EUR), temperate Eurasia (TEUR), temperate North America (TNA), temperate South America (TSA), southern Africa (SAf), North Africa (NAf), tropical Asia (TrAS), tropical South America (TrSA), oceans (OCN), tropics (Tr), boreal (BOR), northern temperate (TN), and southern temperate (TS). The last row marked as SLE denotes the Siegel linear estimate of the linear trend in CH₄ emissions (Tg yr⁻¹ yr⁻¹) over the 2004–2020 period.

	AUS	BEUR	BNA	CHN	IND	EUR	TEUR	TNA	TSA	SAf	NAf	TrAS	TrSA	OCN	Tr	BOR	TN	TS
2004	9.3	13.6	15.7	62.2	37.8	51.0	61.7	46.5	28.0	43.0	28.8	47.9	70.9	11.0	190.6	29.3	159.1	28.0
2005	9.1	13.5	20.1	51.0	37.7	44.1	57.1	49.7	29.2	43.8	31.9	45.8	71.7	13.2	193.1	33.6	150.9	29.2
2006	9.2	13.7	20.6	53.8	31.7	48.4	60.0	52.2	29.8	44.2	32.1	47.2	72.6	12.9	196.1	34.3	160.6	29.8
2007	8.7	13.8	20.3	69.7	45.8	45.5	59.1	50.9	28.7	44.0	32.5	47.6	72.8	13.1	196.9	34.1	155.5	28.7
2008	8.4	12.0	18.1	60.7	38.4	42.5	53.7	56.8	28.7	42.5	31.4	43.6	69.4	12.2	186.8	30.2	153.0	28.7
2009	10.0	12.9	15.4	67.7	34.5	51.8	54.0	44.9	30.7	45.7	31.5	49.0	75.4	11.4	201.7	28.3	150.7	30.7
2010	10.7	10.9	24.9	64.5	38.2	48.3	58.8	51.5	33.9	51.5	39.5	53.7	85.5	13.8	230.2	35.8	158.7	33.9
2011	10.9	8.8	16.1	59.4	45.1	43.9	68.0	37.1	31.2	44.1	33.3	53.9	79.8	12.4	211.0	24.9	148.9	31.2
2012	13.0	9.0	19.1	62.8	37.9	45.2	77.0	42.8	28.0	44.2	40.8	60.9	75.2	13.1	221.1	28.1	165.0	28.0
2013	10.3	9.7	17.3	64.2	42.5	39.3	69.8	35.8	30.0	41.8	36.2	58.3	81.9	12.3	218.2	27.0	144.9	30.0
2014	10.6	9.4	21.4	68.1	31.5	42.9	69.6	47.3	29.9	42.9	36.4	61.0	94.0	12.7	234.2	30.8	159.7	29.9
2015	10.3	11.3	18.0	70.9	33.4	44.4	67.7	41.0	33.3	45.3	43.2	61.6	89.6	13.9	239.6	29.3	153.1	33.3
2016	10.6	10.0	18.2	73.6	39.9	41.9	73.5	35.7	30.7	45.8	39.3	65.0	95.7	14.2	245.7	28.2	151.1	30.7
2017	10.4	10.0	17.1	68.9	37.7	44.9	67.5	40.1	33.0	45.9	40.8	62.3	87.1	13.9	236.1	27.1	152.4	33.0
2018	11.0	11.1	16.7	74.1	37.3	46.2	73.1	38.3	32.8	46.7	43.1	65.7	82.5	14.2	237.9	27.7	157.6	32.8
2019	12.7	12.1	16.5	76.4	38.4	37.6	75.1	45.4	29.0	42.5	42.4	63.8	86.1	12.9	234.8	28.6	158.0	29.0
2020	11.2	10.6	19.7	80.2	43.5	49.3	72.7	42.9	29.6	45.8	53.9	62.5	84.1	14.2	246.2	30.3	164.9	29.6
SLE	0.1	-0.1	-0.2	1.6	0.0	-0.4	1.0	-0.7	0.2	0.1	1.0	1.5	1.2	0.1	3.6	-0.2	0.03	0.2

the constraints from the proxy XCH₄ : XCO₂ ratio (Fraser et al., 2014; Feng et al., 2017), we also assimilate the GLOB-ALVIEW CH₄ and CO₂ data (Schuldt et al., 2021), with assumed uncertainties of 0.5 ppm and 8 ppb for in situ measurements of CO₂ and CH₄, respectively. GLOBALVIEW constitutes a combination of CH₄ data from ground-based data (both flask and continuous) and aircraft data from 54 different laboratories, combined and published by the NOAA-GML (Schuldt et al., 2021). Locations of the assimilated GLOBALVIEW CH₄ (sub-)dataset are shown in Feng et al. (2022).

2.2 GEOS-Chem atmospheric chemistry and transport model

To relate CH_4 emissions to atmospheric CH_4 concentrations, we use v12.1 of the GEOS-Chem 3-D global chemical transport model (CTM) (Bey et al., 2001) at a horizontal resolution of 2° (latitude) by 2.5° (longitude) with 47 vertical levels from the surface to 80 km height using meteorological data from the MERRA-2 meteorological reanalyses (Gelaro et al., 2017) from the NASA Global Modeling and Assimilation Office (GMAO).

Our a priori emissions include (1) monthly EDGAR v6 anthropogenic emissions (Crippa et al., 2021) that account for emissions from oil and gas, coal, livestock, landfills, wastewater, rice, and other anthropogenic sources (including biofuel) from 2004 to 2018, after which we repeat 2018 emissions estimates; (2) monthly GFED-4 biomass burning emissions (version 4.1; Randerson et al., 2017); and (3) monthly v1.0 WetCHARTs wetland emissions (Bloom et al., 2017b). The Harvard–NASA Emissions COmponent (HEMCO) software within GEOS-Chem converts the emissions inventories at their native horizontal resolution to the GEOS-Chem $2^{\circ} \times 2.5^{\circ}$ resolution. Beyond the end of the emissions inventory, emissions are repeated yearly in a priori simulation.

Table 1 shows the δ^{13} C signatures for the source types included in our simulations. These are extracted as mean global values from Sherwood et al. (2017), which provide a database of global isotopic source signatures that are broken down into the same sectors as we employed in our simulations. However, individual source types show a wide range of source signatures, and this uncertainty is reflected in the assigned uncertainty given to the a priori source signatures in the inversion (Sect. 2.3). In the inversion, we differentiate between Arctic and tropical wetlands by applying a 10% isotopically lighter source signature to the Arctic source (Table 1) following Ganesan et al. (2018), who produced a global wetland source signature map based on published δ^{13} C data. Recent work showed that atmospheric simulations that included this isotopic distinction between Arctic and tropical wetlands provided clearer support for rising microbial emissions being responsible for a large fraction of the increase in atmospheric CH₄ since 2007 (Oh et al., 2022). In GEOS-Chem, we simulate isotopologues separately (i.e. for δ^{13} C, 12 CH₄, and 13 CH₄) and then calculate δ^{13} C values. The arithmetic underlying the conversion of isotope ratios to isotopologue emissions for input to the model is detailed in Appendix A.

We include the loss of atmospheric CH₄ from reaction with chlorine, soil uptake, and oxidation by OH. We use monthly 3-D fields of OH, calculated using the fullchemistry version of GEOS-Chem, and monthly 3-D fields of atomic chlorine (Sherwen et al., 2016). Stratospheric loss frequency fields are determined using the NASA Global Modeling Initiative (GMI) stratospheric model (Duncan et al., 2007). Estimates of the microbial consumption of CH₄ in soils are determined from Fung et al. (1991). The resulting atmospheric lifetime of CH₄ against OH is 9.77 years. The corresponding lifetime for methyl chloroform is 5.41 years, which is consistent with atmospheric observation of methyl chloroform. This lifetime also compares well with a multimodel study (Voulgarakis et al., 2013; Morgenstern et al., 2017) that reported global-mean lifetimes of CH_4 that range from 7.2–10.1 years. In our default model configuration, none of these loss processes include interannual variations.

To account for isotopic fractionation due to loss of CH₄ in the troposphere and stratosphere, we use published kinetic isotope effect (KIE) values. These values are employed to scale the reaction rate constants used in the simulations for 12 CH₄ and 13 CH₄ (Table A1). The OH and Cl sinks are handled in the hard coding of the model, whereas the soil sink is handled as a negative emission in the HEMCO file. Therefore, for the soil sink, the KIE is directly applied as a scale factor in the HEMCO configuration file (Snover and Quay, 2000; Burkholder et al., 2019).

We created the initial conditions for atmospheric CH₄ by first scaling a standard CH₄ GEOS-Chem restart file (a file containing a default realistic distribution of CH₄ across the atmosphere) to conditions nearly representative of the start of our analysis in January 2004. We then ran the model 60 times, repeating 2004 MERRA-2 meteorology and emissions (corresponding to approximately 6 e-folding lifetimes for CH₄) to improve as far as possible the simulation of atmospheric gradients in CH₄ under the initial conditions. We then ran a single-year inversion for 2004 to optimise the isotope ratios and CH₄ concentrations relative to ground-based observations following the inverse method detailed below. The δ^{13} C inversion used the regional emissions estimate provided by the posteriori from the CH₄ inversion as a starting point, with sectoral emissions scaled as detailed in Appendix A. The output of this 2004 inversion is a final step in the creation of the initial conditions, which serve as a starting point for the longer inversion that we report here (2004–2020).

For all our calculations, we sample GEOS-Chem at the grid box and local time that correspond to the in situ and satellite remote sensing data. For the satellite data, we also apply scene-dependent averaging kernels to account for vertical structure. This approach allows us to directly compare the model with measurements. Regional trends are calculated by examining the grid boxes encompassed by a given region on the global grid.

2.3 Inverse methods

We use two inverse methods that reflect the volume and simplicity of the data being used. For in situ data we use the maximum a posteriori (MAP) inverse methods, and for the more voluminous satellite data we use an ensemble Kalman filter (EnKF). For brevity, we include only the essential details about either method and refer the reader to dedicated papers.

2.3.1 Maximum a posteriori

To infer regional a posteriori CH₄ fluxes and regional δ^{13} C emissions source signatures from the atmospheric measurements of CH₄, we use the maximum a posteriori solution

(MAP) inverse method (Rodgers, 2000). We solve for CH₄ fluxes and δ^{13} C emissions signatures from 14 geographical regions (Fig. 1). This method combines a priori knowledge and its uncertainty with the measurements and their uncertainties and has been used in a number of studies, e.g. Fraser et al. (2014) and McNorton et al. (2018).

The MAP solution and the associated a posteriori uncertainty are described as, respectively,

$$\boldsymbol{x}^{a} = \boldsymbol{x}^{b} + (\boldsymbol{H}^{T}\boldsymbol{B}^{-1}\boldsymbol{H} + \boldsymbol{R}^{-1})^{-1}\boldsymbol{H}^{T}\boldsymbol{B}^{-1}(\boldsymbol{y} - \boldsymbol{H}\boldsymbol{x}^{b}), \qquad (1)$$

$$\mathbf{A} = (\mathbf{H}^{\mathrm{T}}\mathbf{B}^{-1}\mathbf{H} + \mathbf{R}^{-1})^{-1}, \qquad (2)$$

using the convention that lower-case and upper-case variables denote vectors and matrices, where \boldsymbol{x} denotes the state vector that describes the estimated quantities, which in this study includes monthly CH₄ fluxes and δ^{13} C source signatures from regions across the world (Fig. 1). Subscripts "a" and "b" denote a posteriori and a priori CH₄ fluxes, respectively, and superscripts "-1" and "T" denote matrix inverse and transpose operations, respectively. The measurement vector y includes CH₄ mole fraction or δ^{13} C data. The matrices **B**, **A**, and **R** denote a priori, a posteriori, and measurement error covariance matrices, respectively. B and R are diagonal matrices. For B we assume uncertainties of 50 % of the regional CH₄ fluxes and 15 % for the δ^{13} C values, and for **R** we assume 10 ppb for the mole fraction data and 0.1% for the isotope data. These uncertainties were based on similar studies (Fraser et al., 2014; McNorton et al., 2016). We assume a model transport error of 12 ppb following Feng et al. (2022).

The Jacobian matrix **H** describes the sensitivity of the measurements to changes in the state vector, i.e. $\partial y/\partial x$. For the CH₄ mole fraction inversion, the Jacobian matrix describes the sensitivity of mole fractions in the model to changes in regional CH₄ emissions. We construct the matrix using a series of GEOS-Chem model runs. We systematically let each individual emitting region (described by the state vector) emit for 1 month, while all other regions emit as normal. The individual regional source is then switched off (emissions set to zero), and the effect of this on the 3-D atmospheric distribution of CH₄ mole fractions is recorded over the following 3 months. The result of this test is recorded at the grid boxes that correspond to the location of the measurement sites. The resulting mole fractions therefore describe the sensitivity of a particular measurement site to changes in a specific regional source up to 3 months after emission. This is repeated for every month within the inversion timescale and for every region described in the state vector.

For the δ^{13} C inversions, the Jacobian matrix describes the sensitivity of modelled δ^{13} C to changes in the regional isotopic source signatures. We construct the Jacobian as the difference between a control model calculation (using the CH₄ a posteriori regional emissions and mean source signature values from Sherwood et al., 2017) and perturbed-source-signature model calculation for the whole study pe-

riod (2004–2020). For the perturbed model calculation, we systematically perturb the isotopic source signature of each region (all of the sectors that are contained geographically within a region) isotopically heavier by 20% for the period 2004–2020. The difference in the δ^{13} C value between the control and perturbed run at the location of each measurement site is then divided by the value of δ^{13} C perturbation for the region's source signature to understand the effect of changing a region's source signature on the δ^{13} C value recorded at each measurement site location.

The output from the inversion is improved estimates of regional CH₄ fluxes and δ^{13} C source signatures. The model simulates the global atmosphere on a 2° × 2.5° horizontal grid. The a posteriori regional CH₄ fluxes and isotopic source signatures are applied to the grid boxes in the model which correspond to a given region in an a posteriori simulation.

2.3.2 Ensemble Kalman filter

We use an ensemble Kalman filter (EnKF) approach in performing the inversion using satellite data because we cannot easily evaluate the necessary matrix operations associated with an analytic inversion. Here we use an ensemble of flux perturbation pulses to represent uncertainty in our a priori estimate for regional monthly CH₄ fluxes. We subsequently use a global chemistry transport model (i.e. the GEOS-Chem v12) to track the transport and chemistry processes of the tagged emissions pulses in the atmosphere in order to project their spreads to the observation space. With the ensemble of a priori flux perturbations and the simulated observation impacts, we use the ensemble transform Kalman filter (ETKF) algorithm to numerically estimate the a posteriori CH₄ fluxes and the associated uncertainties by optimally comparing the model simulation with observations (see Feng et al., 2017, for more details). To reduce the computational costs, mainly from tracking tagged emissions pulses, we introduce a 4month moving lag window for each assimilation step because any observation has limited ability to distinguish between the signals emitted long (> 4 months) before from variations in the ambient background atmosphere (Feng et al., 2017). As a result, we are able to include a larger state vector, consisting of monthly scaling factors for 487 (476 land regions and 11 oceanic regions) regional CH₄ (and CO₂) pulse-like basis functions (Fig. S1 in Feng et al., 2022). We define these land sub-regions by dividing the 11 TransCom-3 (Gurney et al., 2002) land regions into 42 to 56 nearly equal sub-regions and use the 11 oceanic regions defined by the TransCom-3 experiment. Because of their smaller sizes, we have assumed a higher uncertainty percentage (60 %) for a priori emissions than the MAP approach described above. We also include spatial correlation with a correlation length of 500 km between the sub-regions.

2.4 Sensitivity of results to changes in assumed OH distributions

To examine the sensitivity of our results to changes in the magnitude of OH, we ran a single sensitivity run that is made up of two parts. First, we imposed a $0.5 \% \text{ yr}^{-1}$ uniform decrease on our 3-D OH field from 2004 to 2019, consistent with the 7 % reduction over 2003-2016 proposed by Turner et al. (2017). Second, we imposed a larger globalscale OH reduction of 1.5 % in 2020 based on recent studies (Miyazaki et al., 2021; Laughner et al., 2021) to describe a more abrupt change due to widespread reductions in nitrogen oxides (NO_x) associated with closing down manufacturing during the first Covid-19 lockdown (Cooper et al., 2022). Newer studies have suggested that the OH reduction in 2020 was closer to 1 % (Peng et al., 2022; Feng et al., 2023), but these estimates are also subject to uncertainties. The purpose of this numerical experiment is to determine the sensitivity of a posteriori CH₄ flux estimates to changes in assumed variations in OH and not to issue a proclamation about a time profile of OH that would simultaneously fit observed changes in CH₄ and δ^{13} C–CH₄.

We use the Siegel linear non-parametric estimates (Siegel, 1982) to fit a line to our a posteriori CH_4 emissions. This method is less sensitive to outliers, e.g. El Niño, that would otherwise compromise the linear-trend estimate (Palmer et al., 2021), and the resulting linear-trend estimate has lower variables than simpler methods. We find that Siegel trend estimates are similar to those estimated by the Theil–Sen estimator.

3 Results

Here, we report a posteriori estimates for total CH₄ emissions inferred from in situ and GOSAT data and then the corresponding a posteriori isotopic source signatures for δ^{13} C. We draw comparisons with previous studies throughout this section.

A posteriori emissions estimates of total CH₄

Figure 2 shows the annual-mean differences in regions between a priori emissions estimates and a posteriori emissions estimates for both ground-based (2004–2020) and GOSAT results (2009–2020). Absolute emissions values are plotted in Fig. 3 for completeness.

On a global scale, terrestrial a posteriori emissions inferred in situ and GOSAT data have progressively increased relative to a priori values since about 2014. The peak difference is in 2020, when we find increased emissions, relative to a priori emissions, of 68.5 ± 61.5 Tg yr⁻¹ for the in situ inversion and of 61.5 ± 37.3 Tg yr⁻¹ for the GOSAT inversion. Global ocean a posteriori CH₄ emissions inferred from in situ and GOSAT data support a negative bias in the a priori emissions, which we do not discuss further. As a zeroth-order check of our a posteriori emissions estimate of total CH₄ (Fig. 4) we compare the published NOAA atmospheric growth rate of CH₄ with our corresponding a posteriori atmospheric mole fractions. Generally, we find that the a posteriori values inferred from in situ and GOSAT data are consistent with the overall trend of the changes in the growth rate, with large year-to-year changes that we explain now in terms of regional emissions changes.

Changes in the global terrestrial emissions reflect changes from different geographical regions. Differences between a posteriori emissions estimates inferred from in situ and GOSAT data are partly due to differences in the geographic coverage of the datasets. Ground-based data have poorer geographic coverage, particularly over the tropics and the Southern Hemisphere, and satellite data are currently available at most once per day under cloud-free conditions. Using the in situ data, we find that the largest a posteriori emissions increases between 2004 and 2020, determined by the Siegel linear estimator, are over the tropics (3.6 Tg yr⁻¹ yr⁻¹; comprising North Africa, southern Africa, tropical South America, and tropical Asia), followed by China (1.6 Tg yr⁻¹ yr⁻¹). Smaller contributions (individually < 0.2 Tg yr⁻¹) were made elsewhere.

Table 1 provides an overview of our annual-mean sectorbased a posteriori emissions for 2004–2020. Generally, our values are close to the reported median values and within the range of values reported by Saunois et al. (2020).

Over the tropics, there is broad consistency between GOSAT and in situ data (Fig. 2) that highlights the negative bias in the a priori values over North Africa (bias of -8.6 Tg yr^{-1}), tropical Asia (bias of -7.2 Tg yr^{-1}), and tropical South America (bias of $-11.63 \text{ Tg yr}^{-1}$). The in situ and GOSAT data for China support a small, steady increase in emissions from 2009 to 2020 (1.0 Tg yr^{-1}) , with emissions inferred from the GOSAT data generally smaller than a priori values throughout the period (Fig. 2). Data over India have a small mean annual trend $(0.33 \text{ Tg yr}^{-1})$. In situ and GOSAT data are more consistent in sign (but not magnitude) at temperate latitudes (Fig. 2). A posteriori emissions from in situ and GOSAT data are generally lower by more than 12.0 and $5.6 \,\mathrm{Tg} \,\mathrm{yr}^{-1}$, respectively, over temperate North America and higher by more than 13.0 and 7.0 Tg yr⁻¹, respectively, over temperate Eurasia, with the smallest discrepancies relative to the a priori values before 2009. A posteriori emissions from boreal regions appear to be larger than a priori values by 4.2 Tg yr^{-1} before 2009 (Fig. 2). After 2009, in situ data become progressively more consistent with the a priori over North America and are typically smaller than a priori values over Eurasia by $\simeq 2.6 \text{ Tg yr}^{-1}$. GOSAT appears to show the converse situation; after 2009, data are lower than a priori values by 4.4 Tg yr⁻¹ over North America and comparable with a priori values over Eurasia. In the Southern Hemisphere, in situ data closely follow a priori values, as expected, since there are few places where data are collected. GOSAT data show a small but persistent increase in emissions with

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Figure 2. Annual-mean CH_4 a posteriori emissions estimates as a residual value relative to a priori values $(Tg yr^{-1})$ from each of the inversion regions in latitudinal order (geographic coverage indicated in Fig. 1) for both ground-based and GOSAT inversion results. Uncertainties are indicated, as calculated from inversion calculations, with an a priori uncertainty of 50% for the ground-based results and 60% for the GOSAT results. The ground-based a posteriori is in blue; the GOSAT a posteriori is in red.

time over southern Africa $(0.41 \text{ Tg yr}^{-1})$, highlighting the negative bias in a priori emissions over Australia and over temperate South America.

We use the a posteriori error covariance matrix from our MAP inversion (A; Eq. 2) to determine our ability to independently estimate CH₄ emissions from our geographical regions. Figure A1 shows no significant a posteriori correlations between neighbouring geographical regions in our state vector. This is consistent with the in situ data being able to independently estimate regional emissions estimates in our state vector.

Our a posteriori emissions estimates are broadly consistent with previous studies. For example, the increase in tropical emissions has been reported using GOSAT data or in situ data within a 3-D CTM inversion (McNorton et al., 2016; Fujita et al., 2020), which examined shorter time periods of 2003–2015 and 1995–2013, respectively. The increase over eastern Africa (which lies within our North Africa region) has been reported by several studies (Lunt et al., 2019, 2021; Pandey et al., 2021; Feng et al., 2022). Using GOSAT data, Sheng et al. (2021) reported that CH₄ emissions from China increased by 0.36 Tg yr⁻¹ from 2012 to 2017. Over the same time period, we estimate an increase of 0.64 and 0.50 Tg yr⁻¹ inferred from in situ and GOSAT data, respectively.

Figure A2 shows observed CH_4 time series at groundbased sites that we use to determine the corresponding GEOS-Chem a priori and a posteriori mole fractions. A priori values already show excellent agreement with observations (mean residual of 14.1 ppb and root-mean-square error, RMSE, of 18.3 ppb), but this is generally improved after the model is fitted to the in situ data, with smaller mean residuals (12.5 ppb) and RMSE (17.0 ppb). This is consistent



Figure 3. A posteriori emissions estimates $(Tg yr^{-1})$ inferred from ground-based in situ data (blue) and GOSAT data (red, with record starting in 2010) for the geographical regions shown in Fig. 1. A priori emissions estimates are denoted by black dots, and a posteriori uncertainties are denoted by whisker bars.

with previous studies such as McNorton et al. (2018) that reported a posteriori RMSE values of 12.3 ppb. Figure A3 shows a posteriori CH₄ mole fractions at NOAA sites that we do not include in the inversion. This provides an additional and independent test of our ability to describe atmospheric CH₄ using a subset of NOAA data that we use in our inversion (Table A2). Generally, our a posteriori estimates agree with these independent data, but for some sites the model has difficulty reproducing the data, e.g. AMY (western South Korea), KZD (Kazakhstan), and SDZ (mainland China). This is because some sites are influenced by local sources that are not representative of the spatial scale of our transport model (\simeq 50 000 km²). Similarly, we find agreement using a posteriori mole fractions using GOSAT data (Fig. A4; mean residual of 29.1 ppb and RMSE 35.1 ppb).

A posteriori source signatures of δ^{13} C

Figure 5 shows source signatures of a posteriori regional δ^{13} C emissions inferred from ground-based in situ data. We group our results into approximately 3-year bands, as a residual from the 2004–2007 mean value, to show how the regional isotopic source signatures change across the time series.

Relative to a priori emissions (Fig. A5), a posteriori values from northern boreal regions (boreal North America and Eurasia) have isotopically lighter signatures (-62%), consistent with a larger contribution from isotopically lighter biogenic emissions and/or a smaller contribution from isotopically heavier thermogenic or pyrogenic emissions (Fig. A5). Conversely, a posteriori values from regions such as temperate Eurasia, Australia, and southern Africa have isotopically heavier source signatures (approximately -40%), suggesting a larger proportion of thermogenic or



Figure 4. A posteriori annual-mean atmospheric CH₄ growth rate inferred from in situ (black line) and GOSAT data (blue line) compared with the equivalent data as published by the NOAA (red line, with uncertainty as surrounding blue field; Dlugokencky et al., 2020). The green line denotes the annual atmospheric growth rate determined using the in situ mole fraction data from the sites included in the inversion ("Sites-Post"). To calculate the atmospheric growth rates from model calculations (NOAA-Post and GOSAT-post), we compare the average global CH₄ mole fraction in 1 year (the mean mole fraction of every grid box in every month of a year) with the mean value from the following year. The calculation is January–January in order to remove the effects of the seasonal cycle following the approach by the NOAA (Dlugokencky et al., 2020).

pyrogenic emissions and/or a smaller contribution from isotopically light biogenic emissions.

Figure 5 shows a general trend towards isotopically lighter regional source signatures of δ^{13} C across the time series. Our analysis suggests that this trend has been ongoing since 2012 and is observed in all regions worldwide but is strongest (compared with a priori estimates) over tropical and Southern Hemisphere regions. For example, tropical South America and southern Africa are 1.2% and 0.9% isotopically lighter than a priori values for 2019 and 2020, respectively.

Our analysis also highlights the period 2007–2012, when regional source signatures, particularly in Northern Hemisphere regions, become isotopically heavier compared with a priori source signatures (by 1.0 % between 2007–2009, by 0.8% between 2010-2012, and by 0.3% between 2013-2015). After 2012, regional source signatures of δ^{13} C generally become isotopically lighter. This result suggests that 2012 was a period when there was a change in the balance of global sources that determine changes in atmospheric CH₄. These isotopic shifts in 2008 and 2012 are noted by Nisbet et al. (2016), who used a box model and examine data from sites measured by the NOAA and Royal Holloway, University of London (RHUL). They found that changes in removal rates could not explain these anomalies so that these events were attributed to changing emissions. We find that China experiences a weaker shift in 2012 to an isotopically lighter $(\sim 0.1 \%) \delta^{13}$ C source signature compared to a priori values (Fig. 5) and compared to other temperate regions. This suggests that heavier isotopic source signatures (such as coal mines) make a larger contribution to this region.

Unlike the a posteriori total CH₄ emissions estimates, we find significant a posteriori correlations between neighbouring regions for δ^{13} C source signatures (Fig. A6). For example, there is a correlation of 0.95 between estimates for southern Africa and temperate South America, so these cannot be considered to be independent estimates. This result aligns with Basu et al. (2022), who used CH₄ mole fraction and δ^{13} C measurements to determine that tropical biogenic sources are driving CH₄ growth. They acknowledged that measurement coverage limited conclusions based exclusively on isotope ratio measurements. Nevertheless, they found a clear trend of stronger emissions of isotopically lighter CH₄, indicative of an increased role for biogenic emissions in the global source makeup.

We find that a posteriori regional δ^{13} C source signatures result in a time series of δ^{13} C that is more consistent with observations than a priori values (Fig. A7), as expected. This particularly affects the period 2008–2018, when a priori emissions source signatures are significantly isotopically lighter. Our a posteriori source signatures result in a mean observation—model residual and RMSE of 0.11 ‰ and 0.14 ‰, respectively. These are smaller than those corresponding to a priori values for the observation—model residual (0.37 ‰) and RMSE (0.41 ‰). Our comparison is consistent with Mc-Norton et al. (2018) (RMSE of 0.1 ‰) and Fujita et al. (2020) (RMSE of 0.08–0.25 ‰).

Sensitivity to assumptions about OH

Figure 6 shows the result of our sensitivity test, which assumes a $0.5 \% \text{ yr}^{-1}$ uniform decrease in our 3-D OH field from 2004 to 2019, followed by a more abrupt decrease of -1.5% in 2020, to describe the widespread reduction in nitrogen oxide emissions. This is an idealised sensitivity test that is inconsistent with global-scale constraints on estimates of the global-mean atmospheric growth of atmospheric CH₄; i.e. most of the observed global growth in atmospheric CH₄ can be explained by the changes in OH. Nevertheless this test provides us with some idea of the robustness of our results against changes in OH.

We find that this alternative assumption about OH does not significantly affect our results until much later in the time series (2017–2019), reflecting our large a posteriori uncertainties. We find a similar quality of fit of the a posteriori model to the data with or without considering the OH trend (not shown). This does not preclude a role for changes in OH, but the concurrent a posteriori shifts in CH₄ emissions and regional isotopic source signatures of δ^{13} C are consistent with decreasing OH playing a smaller role than increasing emissions with isotopically light δ^{13} C source signatures in determining observed changes in atmospheric CH₄ (Lan et al., 2021).

The larger, abrupt change in 2020 results in a marked reduction (approximately 6%, $\sim 40 \text{ Tg yr}^{-1}$) in the emissions necessary to explain the increase in atmospheric CH₄. There

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Figure 5. Regional and global a posteriori δ^{13} C emissions source signatures (‰) in 3-year groups (2004-06, 2007-09, 2010-12, 2013-15, 2016-18, 2019-20) as a residual from the 2004–2006 a posteriori regional emissions source signature value. The a priori equivalent is represented by black dots. The regions are those solved for in the CH₄ and δ^{13} C inversions and are indicated in Fig. 1.

is still debate about the impact of a posteriori CH_4 methane emissions. Peng et al. (2022) used in situ data and concluded that the increase in atmospheric CH_4 in 2020 could be attributed approximately equally to a decrease in OH and an increase in OH. Analysis of GOSAT suggests that increased emissions play a larger role (Qu et al., 2022; Feng et al., 2023).

4 Conclusions

We estimated regional CH₄ emissions and δ^{13} C source signatures for the period 2004–2020, inclusively, by fitting the GEOS-Chem 3-D atmospheric chemistry transport model to surface mole fraction data (Fig. 1) and GOSAT atmospheric column data (2010–2020) using Bayesian inverse methods. We used surface sites for which we had complete monthly

coverage over most of the study period (Table A2). Collectively, our results indicate that the post-2007 increases in CH₄ emissions are best explained by a progressive latitudinal shift in emissions from the northern mid-latitudes to tropical latitudes. A posteriori CH4 emissions estimates inferred from the ground-based and GOSAT data show larger tropical emissions, particularly over North Africa, tropical Asia, and tropical South America as well as over China and at the same time as mid-latitudinal emissions proportions decrease. Source signature estimates inferred from the δ^{13} C measurements (Fig. 1) over the same time period indicate that the latitudinal shift in CH₄ emissions is due to a larger proportion of sources with a lighter atmospheric δ^{13} C signature (e.g. biogenic source such as wetlands) and/or a smaller proportion of sources with a heavier atmospheric δ^{13} C signature (e.g. thermogenic or pyrogenic sources). Our results are broadly

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Figure 6. Annual-mean CH_4 emissions $(Tg yr^{-1})$ for each region of the inversion (indicated in Fig. 1) inferred from the ground-based data (dark blue) and the emissions estimates determined by reduced OH values (described in the text, shown in red). A priori regional emissions estimates are indicated by black dots. Regional uncertainties for the a posteriori emissions are indicated.

consistent with previous studies that focus on shorter contributing periods (McNorton et al., 2018; Nisbet et al., 2019; Fujita et al., 2020; Yin et al., 2021; Lan et al., 2021; Basu et al., 2022), providing confidence in our model assumptions and data selection. We find that our main results are robust against assuming a $0.5 \% \text{ yr}^{-1}$ OH decrease from 2004 to 2019, consistent with Turner et al. (2017), followed by an abrupt 1.5 % OH drop in 2020 that reflects the widespread decrease in nitrogen oxide emissions from shutting down manufacturing during the first Covid-19 lockdown. This is an idealised sensitivity test but nevertheless provides us with some idea of the robustness of our results against changes in OH. A more detailed discussion of the role of OH in 2020 is discussed elsewhere (Qu et al., 2022; Peng et al., 2022; Feng et al., 2023).

Sparse geographic coverage of ground-based data results in larger uncertainties for regional emissions estimates that are informed by fewer data, i.e. high and low latitudes in both hemispheres. For CH₄, this deficiency can be partly addressed using the satellite data, but isotope ratios cannot usefully be retrieved from Earth observation satellite instruments. We use only three measurement sites for δ^{13} C in the Southern Hemisphere, which have a continuous record over the period of study. A consequence of this data sparsity is strong correlations between source signatures from neighbouring regions (Fig. A6). We assume mean sectoral $\delta^{13}C$ source signatures from Sherwood et al. (2017). These values are highly uncertain, as different sectors produce a range of possible δ^{13} C values, and there are significant overlaps between recorded source signatures (Douglas et al., 2017), but the values chosen represent our current best knowledge of mean values. These data have greater value when they are used in a broader context with other data, as we describe in this study. We have used satellite observations to help identify that large-scale emissions change over regions that coincide with wetlands.

Collectively, empirical evidence, including in situ and GOSAT observation of CH₄ and in situ δ^{13} C data, points to an increasing biogenic source originating from the tropics. While we cannot definitively attribute these changes to increasing wetland emissions, there is sufficient contextual evidence, building on previous studies, to suggest that wetlands have played a significant role in recent growth of atmospheric CH₄. First, large changes in OH that would be needed to explain this atmospheric growth are inconsistent with increasingly isotopically light $\delta^{13}C$ observations in the atmospheric record (Lan et al., 2021). Second, we know from in situ data the broad geographical regions responsible for increasing CH₄ emissions and isotopically lighter δ^{13} C source signature, where the seasonal cycles are consistent with biogenic emissions peaking outside the burning season. Third, GOSAT provides us with additional information about the geographical distribution of CH₄ emissions; tropical emissions hotspots are colocated with known wetland regions (Lunt et al., 2019, 2021; Pandey et al., 2021; Wilson et al., 2021; Feng et al., 2022, 2023). Finally, we also have evidence from other satellite data, e.g. hydrology, that helps explain the growth of wetland emissions in the last decade (Lunt et al., 2019; Feng et al., 2022). Greater confidence in source attribution of changes in atmospheric CH₄ may come from collecting and interpreting δD and multiply substituted "clumped" isotopes (Douglas et al., 2017; Chung and Arnold, 2021), alongside δ^{13} C. This needs to be accompanied by field measurements of these isotope ratios to improve delineation between different sectors.

The evidence presented here is consistent with a growing body of work that points to a substantial increase in biogenic CH_4 emissions from the tropics. This increase will likely have major implications for achieving the goals of the Paris Agreement (Nisbet et al., 2019). Nature does not care about the origin of atmospheric CH_4 so that increasing biogenic emissions will require larger emissions reductions from anthropogenic sectors, placing additional pressure on citizens to reduce their carbon footprints.

Appendix A: Isotopologue emissions

To simulate the atmospheric isotope ratio δ^{13} C, the isotopologues 12 CH₄ and 13 CH₄ are considered separately in the model. To calculate the specific sectoral isotopologue emissions we use the emissions calculated from the CH₄ mole fraction simulation and the isotope ratios defined in Table 2. We consider the isotope 13 C relative to all isotopes in the sample (designated hereafter as 13x) using

$$13x = \frac{{}^{13}\text{C}}{{}^{12}\text{C} + {}^{13}\text{C}} = \frac{{}^{13}\text{C}/{}^{12}\text{C}}{1 + ({}^{13}\text{C}/{}^{12}\text{C})},$$
(A1)

where ${}^{13}C/{}^{12}C$ is calculated from the $\delta^{13}C$ reported on the international carbon isotope scale VPDB (Vienna Pee Dee Belemnite). This is the proportional molar abundance of the isotopologues containing ${}^{13}C$ (dominated by ${}^{13}CH_4$) relative to the isotopologues containing ${}^{12}C$ (dominated by ${}^{12}CH_4$). This value has to be adjusted before being applied in GEOS-Chem to convert from isotope ratio values to kilogram values used by emissions inventories:

$$SF13 = 13x \times \frac{M_{13}}{M_{\text{tot}}},\tag{A2}$$

where SF13 is the scale factor applied to each emissions type for the ¹³CH₄ simulation, M_{13} is the molecular weight of ¹³CH₄ (17.035 g mol⁻¹), and M_{tot} is the molecular weight of CH₄ (16.04 g mol⁻¹).

For the ¹²CH₄ counterpart to ¹³CH₄, we use a similar approach. The ratio of ¹²C to all isotopes in the sample (designated as 12x) is given by

$$12x = \frac{{}^{12}\text{C}}{{}^{13}\text{C} + {}^{12}\text{C}}.$$
(A3)

This is similarly adjusted from molar to mass ratio; SF12 is the scale factor for each emissions type in the ${}^{12}CH_4$ simulations:

$$SF12 = 12x \times \frac{M_{12}}{M_{\text{tot}}},\tag{A4}$$

where M_{12} is the molecular weight of 12 CH₄ (16.03 g mol⁻¹). Since 13 C and 12 C are the only stable carbon isotopes of CH₄, 13*x* and 12*x* should sum to 1.



Figure A1. A posteriori correlations between CH_4 emissions from geographical regions inferred from ground-based CH_4 mole fraction data. These correlations are determined by normalising the diagonal elements of the a posteriori error covariance matrix (Eq. 2).



Figure A2. Observed (red), a priori (grey), and a posteriori (black) model atmospheric mole fractions at a series of NOAA sites (subplot titles denote site codes; Table A2) covering a range of latitudes.



Figure A3. A posteriori (black) monthly estimates of atmospheric CH₄ simulated at NOAA sites across latitudes. Red indicates monthly mean CH₄ data from the NOAA network sites indicated. These sites were not included in the CH₄ inversion but are shown here to provide independent validation of a posteriori emissions. The sites included are Baltic Sea, Poland (55.35° N, 17.22° E); Cold Bay, Alaska (55.21° N, 162.72° W); Sary Taukum, Kazahkstan (44.08° N, 76.87° E); Shangdianzi, China (44.65° N, 117.12° E); Point Arena, USA (38.95° N, 123.74° W); Anmyeon-do, Republic of Korea (36.54° N, 126.38° E); Terceira Island, Azores (38.77° N, 27.37° W); Dongsha Island, Taiwan (20.70° N, 116.73° E); High Altitude Global Climate Observation Center, Mexico (18.98° N, 97.31° W); Mt. Kenya, Kenya (0.06° S, 37.29° E); Bukit Kototabang, Indonesia (0.20° S, 100.31° E); Arembepe, Brazil (12.77° S, 38.17° W); Gobabeb, Namibia (23.58° S, 15.03° E); Cape Point, South Africa (34.35° S, 18.49° E); and Drake Passage (59.00° S, 64.69° W).



Figure A4. Observed (red) and 3-hourly surface a posteriori CH_4 values inferred from GOSAT data (black) at the location of a number of NOAA sites (Table A2) from 2010–2020.



Figure A5. Monthly a priori (grey) and a posteriori (blue) regional δ^{13} C source signatures (%). Values are produced using ground-based in situ δ^{13} C data. Uncertainties in source signatures are indicated as shaded envelopes, with a priori uncertainties of 15%.



Figure A6. A posteriori correlations between δ^{13} C source signatures from geographical regions inferred from ground-based δ^{13} C data. These correlations are determined by normalising the diagonal elements of the a posteriori error covariance matrix (Eq. 2).



Figure A7. A priori (grey) and a posteriori (black) monthly estimates of atmospheric δ^{13} C simulated at NOAA sites across latitudes (site codes listed in Table A2). Red indicates monthly mean δ^{13} C data from CU-INSTAAR for the respective sites.

Table A1. Kinetic isotope effects (KIEs) for different isotopologues reacting with the three main sinks of CH ₄ (OH, Cl, soil) at 298 K. A
KIE indicates relative reaction rate compared with ¹² CH ₄ , the reaction rate constant is applied to the OH and Cl sinks and is dependent on
temperature (T) , and the scaling factor is applied to the soil sink at each time step (handled as a negative emission).

Isotopologue	Sink	KIF	Reaction rate constant	Scaling factor	Literature source
isotopologue	SIIIK	MIL	Reaction face constant	Sealing factor	
¹² CH ₄	OH	1	$2.45 \times 10^{-12} \times e^{\frac{-1775}{T}}$	n/a	Burkholder et al. (2019)
¹² CH ₄	Cl	1	$9.600 \times 10^{-12} \times e^{\frac{-1360}{T}}$	n/a	Kirschke et al. (2013)
¹² CH ₄	Soil	n/a	n/a	1	Snover and Quay (2000)
¹³ CH ₄	OH	1.0039	$2.44 \times 10^{-12} \times e^{\frac{-17/5}{T}}$	n/a	Burkholder et al. (2019)
¹³ CH ₄	Cl	1.06	$9.057 \times 10^{-12} \times e^{\frac{-1360}{T}}$	n/a	Feilberg et al. (2005)
¹³ CH ₄	Soil	n/a	n/a	1.0670	Snover and Quay (2000)

n/a – not applicable

Table A2. Sites that are included in the in situ inversions. All sites are part of the NOAA network except for KRS, which is part of the JR-STATION network, monitored by NIES Japan.

Code	Full Name	Latitude	Longitude
ALT	Alert Station	82.28	-62.30
ZEP	Ny-Ålesund, Svalbard	78.90	11.89
SUM	Summit, Greenland	72.60	-38.42
BRW	Barrow Station	71.32	156.61
ICE	Stórhöfði,Iceland	63.40	-20.29
KRS	Karasevoe, Siberia	58.14	82.25
MHD	Mace Head, Ireland	53.33	-9.90
SHM	Shemya Island, Alaska	52.71	174.12
UUM	Ulaan-Uul, Mongolia	44.45	111.09
NWR	Niwot Ridge, Colorado	40.05	-105.59
UTA	Wendover, Utah	39.90	-113.72
WLG	Mt. Waliguan, China	36.29	100.90
BMW	Bermuda	32.26	-64.88
WIS	Ketura, Israel	29.96	35.06
IZO	Izana, Tenerife	28.31	-16.50
MID	Midway Islands	28.22	-177.37
KEY	Key Biscayne, Florida	25.67	-80.16
ASK	Assekrem, Algeria	23.26	5.63
KUM	Cape Kumukahi, Hawaii	19.56	-154.89
MLO	Mauna Loa, Hawaii	19.54	-155.58
RPB	Ragged Point, Barbados	13.17	-59.43
SEY	Mahé, Seychelles	-4.68	55.53
ASC	Ascension Island	-7.97	-14.40
SMO	American Samoa	-14.25	-170.56
CGO	Cape Grim	-40.68	144.69
BHD	Baring Head	-41.40	174.87
CRZ	Crozet Islands	-46.43	51.85
USH	Ushuaia, Argentina	-54.84	-68.31
PSA	Palmer Station, Antarctica	-64.77	-64.05
SYO	Syowa Station, Antarctica	-69.01	39.59
SPO	South Pole, Antarctica	-89.98	-24.8

Code availability. The community-led GEOS-Chem model of atmospheric chemistry is maintained centrally by Harvard University (https://geoschem.github.io/, Bey et al., 2001) and is available on request. The ensemble Kalman filter code is publicly available as PyOSSE (https://www.nceo.ac.uk/data-tools/atmospheric-tools/, Feng et al., 2009).

Data availability. All the data and materials used in this study are freely available. The NOAA-GML and CU-INSTAAR groundbased CH₄ and δ^{13} C data are available from the NOAA GML FTP server (https://gml.noaa.gov/dv/data, last access: 24 July 2023, https://doi.org/10.15138/VNCZ-M766, Dlugokencky et al., 2020), subject to their fair-use policies. Data from the JR-STATION network were provided with the cooperation of NIES Japan. The University of Leicester GOSAT Proxy v9.0 XCH4 data are available from the Centre for Environmental Data Analysis data repository https://doi.org/10.5285/18ef8247f52a4cb6a14013f8235cc1eb at (Parker and Boesch, 2020) and from the Copernicus Climate Data Store. EDGAR data are available at https://edgar.jrc.ec.europa.eu/ (Crippa et al., 2021), GFED-4 data are available at https://www.globalfiredata.org/data.html (last access: 24 July 2023, https://doi.org/10.3334/ORNLDAAC/1293, Randerson et al., 2017), and WetCHARTs data are available at https://doi.org/10.3334/ORNLDAAC/1502 (Bloom et al., 2017b).

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Competing interests. The contact author has declared that none of the authors has any competing interests.

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