



*Supplement of*

## **Impact of HO<sub>2</sub> aerosol uptake on radical levels and O<sub>3</sub> production during summertime in Beijing**

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## S1.1 FAGE instrumentation description

The University of Leeds Fluorescence Assay by Gas Expansion (FAGE) instrument made measurements of OH, HO<sub>2</sub> and RO<sub>2</sub> radicals and OH reactivity ( $k_{OH}$ ). The FAGE instrument set up is described fully in Whalley et al., 2018 while the OH reactivity instrument set up is described fully in Whalley et al., 2016. Both instruments are also described fully in Slater et al., 2020 and so only a brief description is given here.

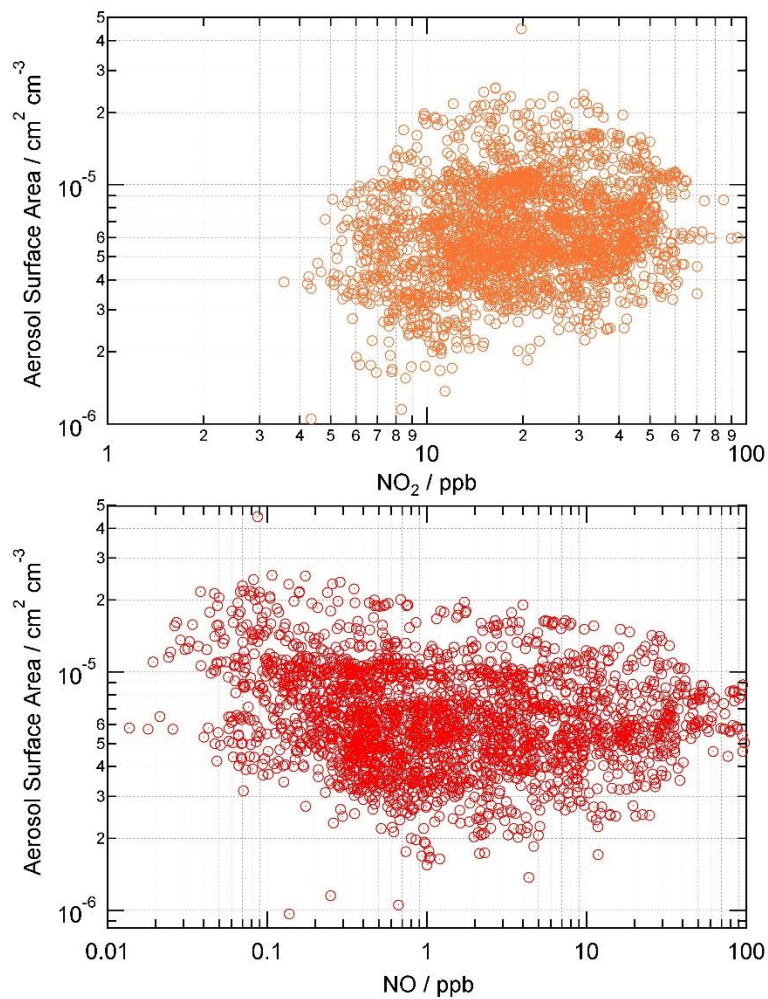
Two cells, a HO<sub>x</sub> cell and a RO<sub>x</sub> cell connected together with a side arm, were used to take radical measurements from the roof of the Leeds FAGE lab container. A RO<sub>x</sub>LIF flow reactor was also coupled to the RO<sub>x</sub> cell to allow for detection of RO<sub>2</sub> (total, complex and simple) as described by Fuchs et al., 2008. The HO<sub>x</sub> cell took sequential measurements of OH and the sum of OH and HO<sub>2</sub>, by the addition of NO (Messer, 99.5 %), which titrated HO<sub>2</sub> to OH for detection by Laser Induced Fluorescence (LIF) at 308 nm.

The RO<sub>x</sub>LIF reactor operated in 2 modes: a 'HO<sub>x</sub> mode' where a flow of CO (10 % in N<sub>2</sub>) was added to ambient sampled air close to the pin hole to convert all ambient OH to HO<sub>2</sub>; and a 'RO<sub>x</sub> mode' where NO (500 ppmv in N<sub>2</sub>) was added in addition to the CO flow to convert all RO<sub>2</sub> into OH before all OH was then rapidly converted by CO into HO<sub>2</sub>. The air from the RO<sub>x</sub>LIF reactor was then drawn into the FAGE low pressure fluorescence cell, whereupon pure NO (Messer, 99.5 %) was injected to convert HO<sub>2</sub> to OH. In HO<sub>x</sub> mode, the sum of OH, HO<sub>2</sub> and complex RO<sub>2</sub> was measured, while in RO<sub>x</sub> mode, the sum of OH, HO<sub>2</sub> and total RO<sub>2</sub> was measured. From this the concentration of complex RO<sub>2</sub> and HO<sub>2</sub>/OH from RO<sub>x</sub> can be determined.

An Inlet Pre-Injector was used attached to the HO<sub>x</sub> cell to remove ambient OH by injecting propane directly above the inlet of the cell. This leads to a background measurement while the laser is still online to the OH transition; this background is known as OH<sub>CHEM</sub>. OH<sub>CHEM</sub> includes signal from laser scatter and scattered solar radiation and any fluorescence signal from any OH generated inside the cell from an interference precursor. By comparing OH<sub>CHEM</sub> to the signal generated when the 308 nm laser tuned off the OH transition, OH<sub>WAVE</sub>, the contribution of any interference can be identified. While the laser is offline, OH<sub>WAVE</sub>, any signal seen is from laser scattered light and scattered solar radiation. Agreement between OH<sub>WAVE</sub> and OH<sub>CHEM</sub> was generally very good during the Summer AIRPRO campaign with an overall orthogonal distance regression slope of  $1.103 \pm 0.017$ , with the exception of an interference seen when O<sub>3</sub> levels were elevated (see Woodward-Massey et al., 2020 for details).

## S1.2 Correlation between aerosol surface area and $\text{NO}_x$ mixing ratio

During winter haze periods in Beijing, a strong correlation has been observed between  $\text{PM}_{2.5}$  and  $\text{NO}_x$  (Slater et al., 2020). During the Summer Beijing AIRPRO campaign in 2017, no strong correlation between  $\text{PM}_{2.5}$  and  $\text{NO}_x$  was seen. A correlation plot of  $\text{PM}_{2.5}$  aerosol surface area ( $\text{cm}^2 \text{cm}^{-3}$ ) versus  $\text{NO}$  and  $\text{NO}_2$  mixing ratio (ppb) is shown in Figure S1.

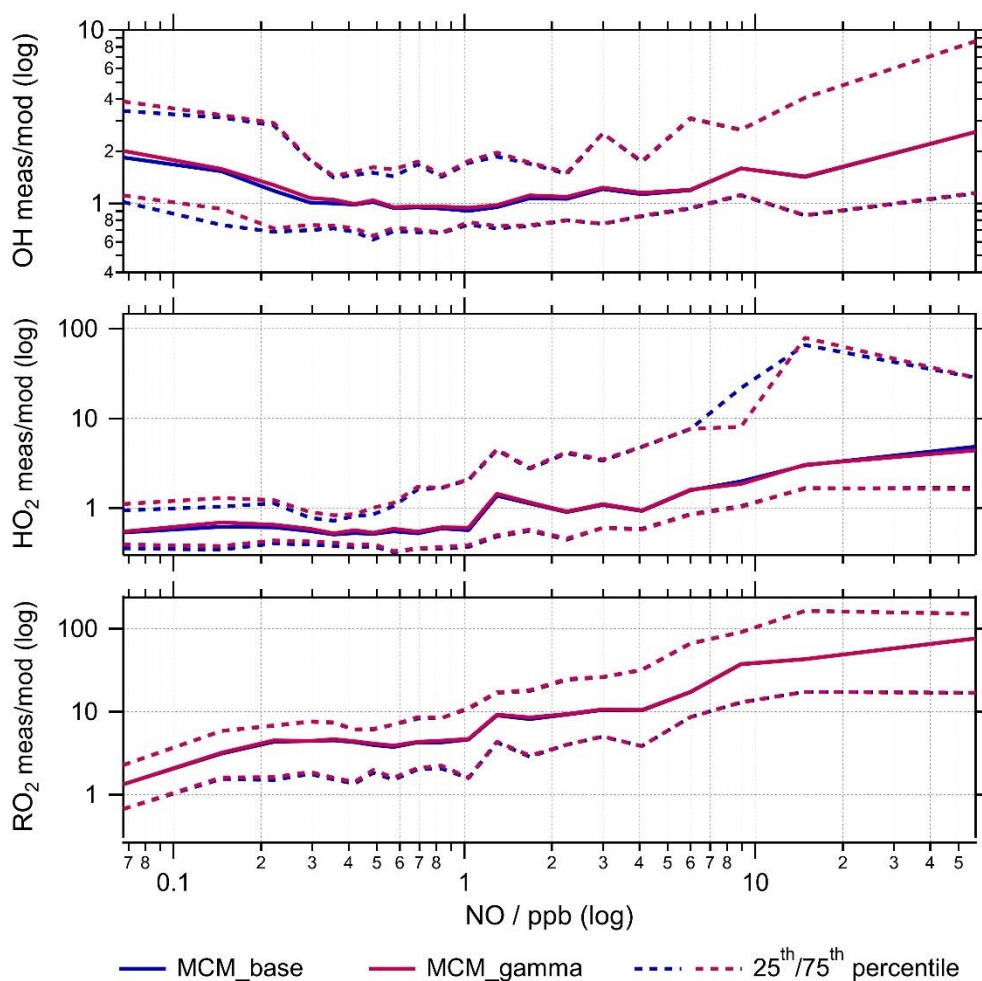


**Figure S1.** Correlation plot of  $\text{PM}_{2.5}$  aerosol surface area ( $\text{cm}^2 \text{cm}^{-3}$ ) versus  $\text{NO}_2$  (orange circles) and  $\text{NO}$  (red circles) mixing ratio (ppb) for Summer Beijing Campaign.

### **S1.3 HO<sub>2</sub> measured:modelled dependence on NO mixing ratio**

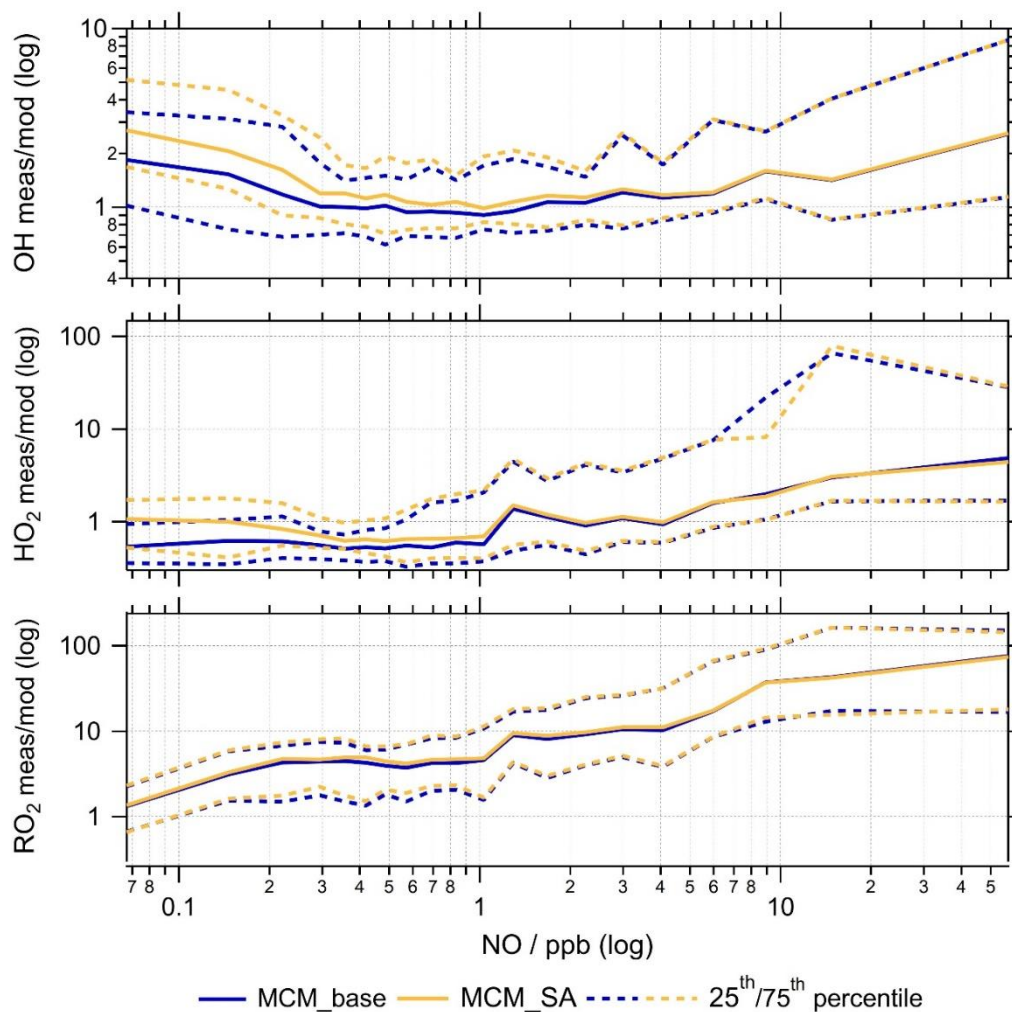
Figure S2 shows the ratio of measured to modelled OH, HO<sub>2</sub> and RO<sub>2</sub> radical concentrations binned against NO mixing ratio (ppb) for MCM\_gamma, compared to MCM\_base.

For the range of NO mixing ratios observed across the summer AIRPRO campaign, the OH measured to modelled ratio is close to 1 between ~0.3 and 2 ppb NO with the MCM\_base model beginning to under-predict OH slightly both below 0.3 ppb NO and above 2 ppb NO. Both HO<sub>2</sub> and RO<sub>2</sub> radical concentrations were strongly dependent on NO mixing ratio, with the model over-predicting HO<sub>2</sub> below ~ 1 ppb NO. For the entire campaign the average NO was 4.7 ppb with 45% of NO measurements taken across campaign being less than or equal to 1 ppb. Across all NO mixing ratios the measured to modelled ratio for RO<sub>2</sub> shows a large under-prediction, with the largest under-prediction at the highest NO mixing ratios. This is likely contributing to the underprediction of HO<sub>2</sub> at higher NO mixing ratios. From Figure S2 it can be seen that the addition of the calculated HO<sub>2</sub> uptake coefficient has had little effect across the range of NO mixing ratios measured during the summer AIRPRO campaign.



**Figure S2.** Ratio of measured to modelled OH, HO<sub>2</sub> and RO<sub>2</sub> radical concentrations using the MCM\_base (blue) and MCM\_gamma (dark pink) model binned over the range of NO mixing ratios (ppb) for the summer AIRPRO campaign. Solid lines show the median average measured to modelled radical concentration. Dashed lines show the 25<sup>th</sup>/75<sup>th</sup> percentiles.

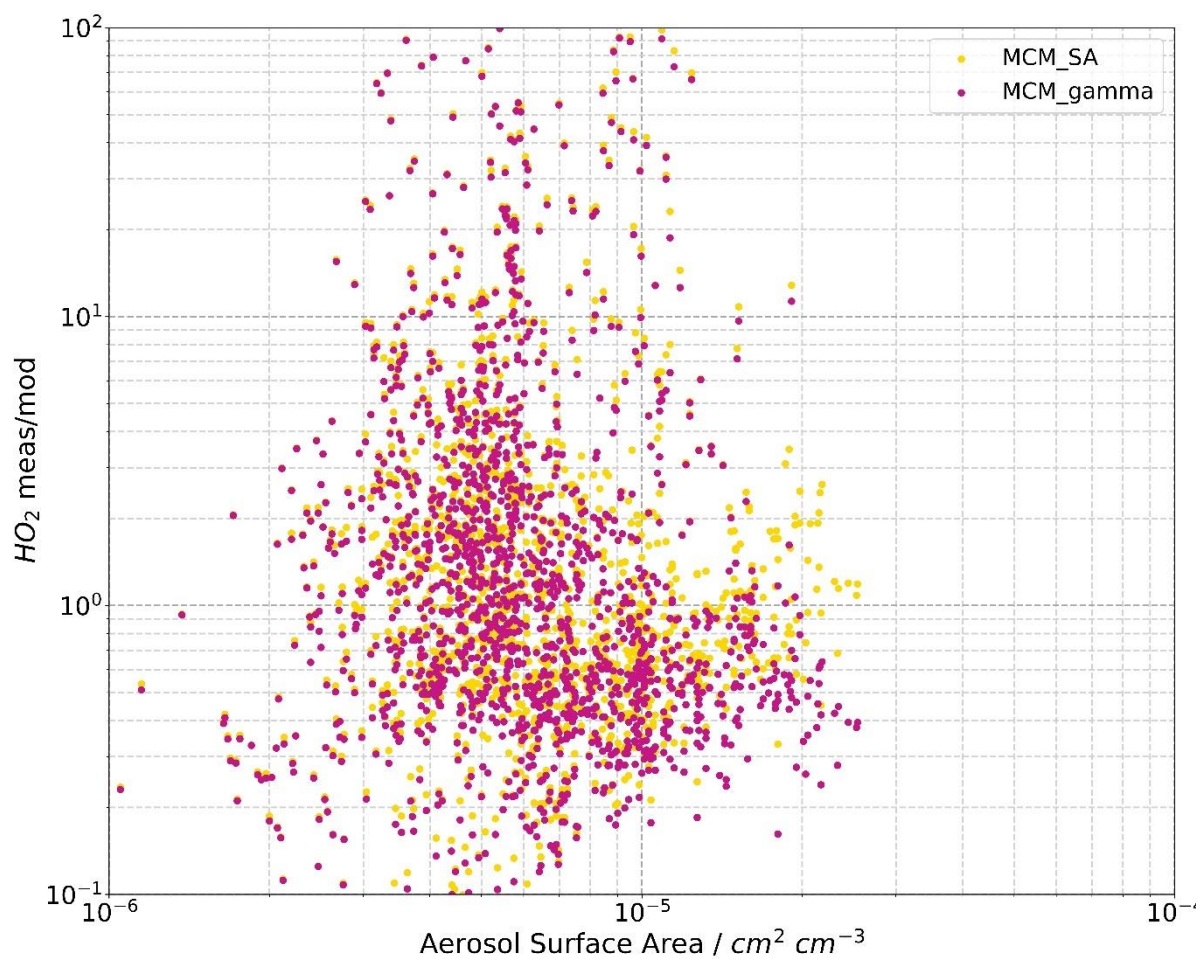
As seen in Figure S3, the addition of  $\gamma_{HO_2} = 0.2$  affected the ability of the model to reproduce the NO dependence of radical concentrations. While MCM\_base over-predicts HO<sub>2</sub> below  $\sim 1$  ppb NO, the over-prediction of HO<sub>2</sub> decreases below 1 ppb NO for MCM\_SA with HO<sub>2</sub> being well reproduced at the lowest NO mixing ratios (i.e.  $< 0.1$  ppb) due to the relative increase in the importance of HO<sub>2</sub> uptake as a sink of HO<sub>2</sub>. Modelled RO<sub>2</sub> is not significantly affected by the addition of HO<sub>2</sub> uptake at any NO mixing ratio. The modelled concentration of OH is under-predicted for the entire range of NO mixing ratios compared to measured values, though only slightly between  $\sim 1$  and 6 ppb NO. Below  $\sim 4$  ppb NO, the underprediction of OH by MCM\_SA increases compared to MCM\_base due most likely to loss of HO<sub>2</sub> onto aerosols competing with loss via NO to give OH. Budget analysis done by Whalley et al., 2021, showcases that with a reduction in over-prediction of modelled HO<sub>2</sub>, OH is under-predicted revealing a missing OH source.



**Figure S3.** Ratio of measured to modelled OH, HO<sub>2</sub> and RO<sub>2</sub> radical concentrations using the MCM\_base (blue) and MCM\_SA (yellow) model binned over the range of NO mixing ratios (ppb) for the summer AIRPRO campaign. Solid lines show the median average measured to modelled radical concentration. Dashed lines show the 25<sup>th</sup>/75<sup>th</sup> percentile.

## S1.4 HO<sub>2</sub> measured:modelled dependence on aerosol surface area

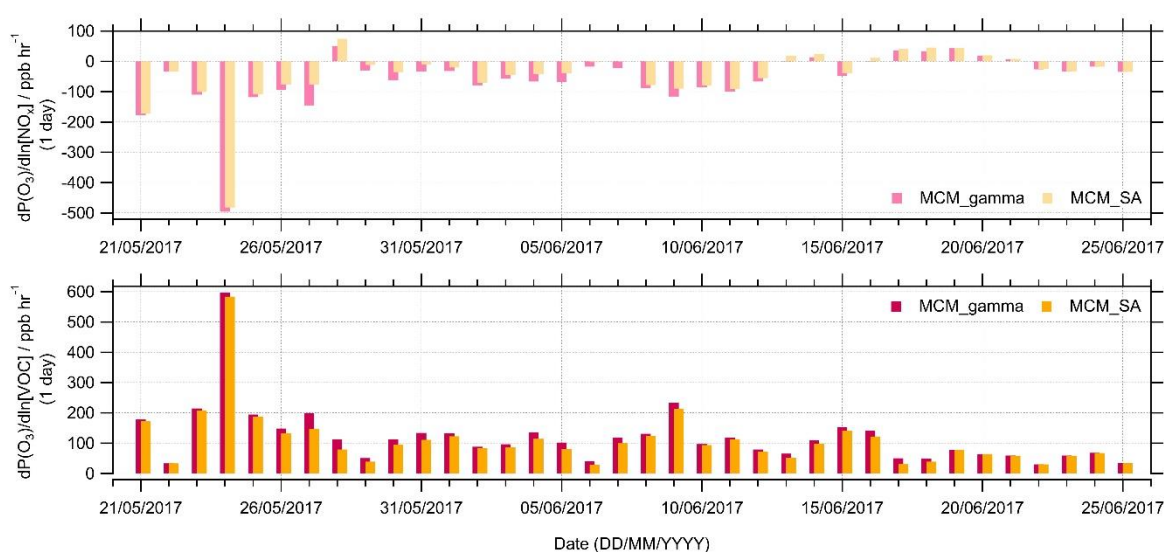
Figure S4 shows the ratio of measured to modelled HO<sub>2</sub> radical concentrations as a function of aerosol surface area concentration ( $\text{cm}^2 \text{cm}^{-3}$ ) for the entire Summer AIRPRO campaign.



**Figure S4.** Ratio of measured to modelled HO<sub>2</sub> radical concentrations using the MCM\_gamma (pink) and MCM\_SA (yellow) models binned over the range of aerosol surface area concentrations ( $\text{cm}^2 \text{cm}^{-3}$ ) for the summer AIRPRO campaign.

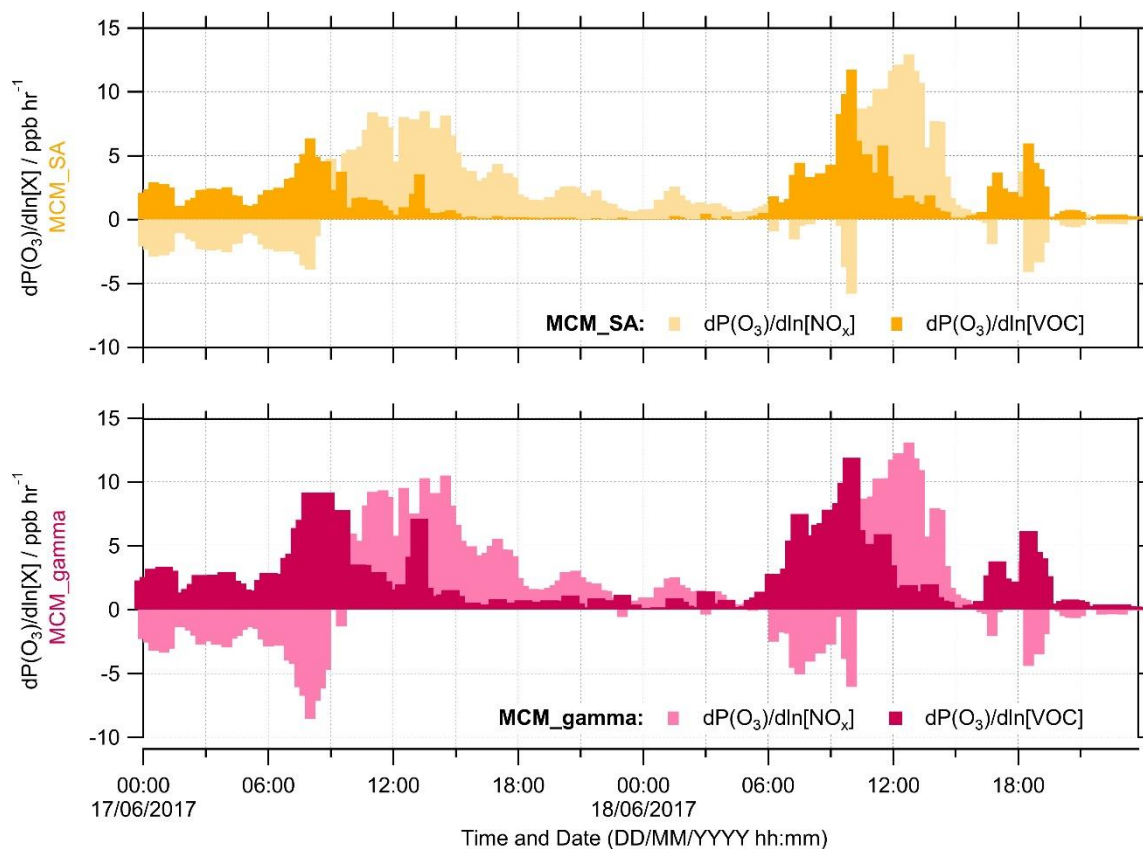
## S1.5 Time series of absolute O<sub>3</sub> sensitivity across AIRPRO Summer campaign

The time series of absolute P(O<sub>3</sub>), averaged up to a daily time resolution, across the entire measurement period is shown in Figure S5. Calculations of absolute O<sub>3</sub> production sensitivity showcase that for both MCM\_gamma and MCM\_SA, on average, the O<sub>3</sub> regime was VOC sensitive throughout the day with NO<sub>x</sub> sensitivity increasing in the afternoons. As shown in Figure S6, when the NO mixing ratio was low and this coincided with high aerosol surface area, the O<sub>3</sub> sensitivity shifted from VOC to NO<sub>x</sub> limited.



**Figure S5.** Time series of absolute O<sub>3</sub> sensitivity to NO<sub>x</sub> (top panel) and VOC (bottom panel) in ppb h<sup>-1</sup> for MCM\_gamma (pink) and MCM\_SA (yellow) averaged daily across the entire summer AIRPRO campaign. MCM\_gamma includes  $\gamma_{HO_2}$  calculated using the Song parameterisation (av.  $0.070 \pm 0.035$ ) while MCM\_SA includes  $\gamma_{HO_2}$  at a fixed value of 0.2.





**Figure S6.** Time series of absolute  $O_3$  sensitivity to  $NO_x$  (lighter shades) and VOC (darker shades) for MCM\_SA (top panel) and MCM\_gamma (bottom panel) for 17/06/2017-18/06/2017 showing an example where  $dP(O_3)/d\ln[NO_x] > dP(O_3)/d\ln[VOC]$  in the afternoon of both days. The average  $NO$  mixing ratio for this period was  $0.41 \pm 0.50$  ppbV while the average SA concentration was  $(8.4 \pm 6.2) \times 10^{-6} \text{ cm}^2 \text{ cm}^{-3}$ . MCM\_gamma includes  $\gamma_{HO_2}$  calculated using the Song parameterisation (av.  $0.070 \pm 0.035$ ) while MCM\_SA includes  $\gamma_{HO_2}$  at a fixed value of 0.2.

## S1.6 Comparison of $\gamma_{HO_2}$ for Beijing, 2017 and Wangdu, 2014 campaigns

In the Song et al., 2020, the  $\gamma_{HO_2}$  was calculated using measured data from a comprehensive 2014 field campaign at a rural site in Wangdu on the North China Plain. In Song et al., 2020,  $\gamma_{HO_2}$  was calculated as  $0.116 \pm 0.086$  (for  $a_{HO_2} = 0.5$ ) for the conditions of the Wangdu campaign. Another study by Tan et al., 2020 calculated the  $\gamma_{HO_2}$  value by comparing  $HO_2$  field data with  $HO_2$  modelled concentrations and estimating the  $HO_2$  aerosol uptake coefficient required to close the radical budget for each day. This method lead to an average value of  $\gamma_{HO_2} = 0.08 \pm 0.13$ , within the uncertainty of the  $\gamma_{HO_2}$  value calculated by Song et al., 2020.

The average  $\gamma_{HO_2}$  value of  $0.116 \pm 0.086$  for the Wangdu campaign is comparable, though slightly higher than the  $\gamma_{HO_2}$  value of  $0.070 \pm 0.035$  calculated in this study for the summer AIRPRO Beijing campaign. A comparison of the measured parameters used within the Song parameterisation for both the Wangdu and Beijing campaigns are shown in Table S1.

Parameters	Average Values for Wangdu	Average Values for Beijing
ALWC ( $\mu\text{g m}^{-3}$ )	68.2	6.94 (fixed as 14 in eqn 1)
Relative humidity (%)	61	43
[PM] ( $\mu\text{g m}^{-3}$ )	67.2	38
[Cu] <sub>eff</sub> <sup>2+</sup> ( $\text{mol L}^{-1}/\text{ng m}^{-3}$ )	0.0005 / 35.8	0.0008 / 4
Count median radius, $R_d$ (cm)	$3 \times 10^{-6}$	$1.5 \times 10^{-6}$

**Table S1.** Comparison of important measured parameters within the Song parameterisation for the Summer AIRPRO campaign in Beijing 2017 and the Wangdu campaign on the North China Plain, 2014. Wangdu campaign values from Song et al., 2020 and Song et al., 2022.

In the Song parameterisation, the uptake coefficient depends on [PM], [ALWC] and [Cu<sup>2+</sup>]<sub>eff</sub>, with the strongest dependence on [Cu<sup>2+</sup>]<sub>eff</sub>. As shown in Table S1 **Table S1**, both PM and ALWC concentrations were lower on average for the Beijing AIRPRO campaign compared to the Wangdu campaign. While the total copper mass concentration in  $\text{ng m}^{-3}$  was significantly lower ( $4 \text{ ng m}^{-3}$  on average) for the Beijing campaign than the Wangdu campaign ( $35.8 \text{ ng m}^{-3}$ ), due to differences in the count median radius and aerosol volume, the [Cu<sup>2+</sup>]<sub>eff</sub> calculated in  $\text{mol L}^{-1}$  was slightly higher for the Beijing campaign. This is due to a larger  $R_d$  value for the Wangdu campaign and therefore higher aerosol volume, due mostly likely in part to higher average RH across the campaign, which makes up for the difference in copper mass concentration between

the two campaigns. Comparable values of  $[\text{Cu}^{2+}]_{\text{eff}}$ , but lower [ALWC] and [PM] for the Beijing campaign explain why the estimated uptake coefficient values for Beijing are slightly lower than those for the Wangdu campaign.

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