



Supplement of

Measurement Report: Investigation on the sources and formation processes of dicarboxylic acids and related species in urban aerosols before and during the COVID-19 lockdown in Jinan, East China

Jingjing Meng et al.

Correspondence to: Pingqing Fu (fupingqing@tju.edu.cn)

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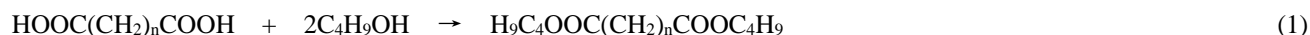
Supporting Information

Text S1. Stable carbon isotopic composition of diacids and related compounds

The isotopic analytical precision was $\leq 0.5\text{‰}$ for a peak height (m/z 44) of 0.2 – 8 V. The isotopic composition of diacids and their derivatives is reported in the δ notation relative to the Pee Dee Belemnite (PDB) standard as follows:

$$\delta^{13}\text{C} (\text{‰}) = \left[\left(\frac{^{13}\text{C}/^{12}\text{C}}{^{13}\text{C}/^{12}\text{C}} \right)_{\text{sample}} / \left(\frac{^{13}\text{C}/^{12}\text{C}}{^{13}\text{C}/^{12}\text{C}} \right)_{\text{PDB}} - 1 \right] \times 10^3$$

After the measurement of isotopic compositions of butyl esters, $\delta^{13}\text{C}$ of individual diacids is calculated by the following mass balance equation:



(unknown $\delta^{13}\text{C}_{\text{Diacid}}$) (known $\delta^{13}\text{C}_{\text{BuOH}}$) (measured $\delta^{13}\text{C}_{\text{DIBE}}$)

$$\delta^{13}\text{C}_{\text{DIBE}} = f_{\text{Diacid}}\delta^{13}\text{C}_{\text{Diacid}} + f_{\text{BuOH}}\delta^{13}\text{C}_{\text{BuOH}} \quad (2)$$

where $\delta^{13}\text{C}_{\text{Diacid}}$, $\delta^{13}\text{C}_{\text{BuOH}}$, and $\delta^{13}\text{C}_{\text{DIBE}}$ represent the carbon isotopic composition of diacid, 1-butanol, and diacid dibutyl ester, respectively; f_{Diacid} and f_{BuOH} are fractions of carbon in the ester derived from diacid and butanol, respectively. f_{Diacid} and f_{BuOH} of C_2 are 0.2 and 0.8, respectively. The f_{Diacid} and f_{BuOH} values of other organic acids have been described in Kawamura and Watanabe (2004). Before actual sample analysis, we confirmed that $\delta^{13}\text{C}$ values of the working standards (a mixture of normal C_{16} – C_{30} alkanes with 0.55 – 2.83 $\text{ng } \mu\text{L}^{-1}$) were equivalent to the theoretical values within an analytical error of $<0.2\text{‰}$.

In addition, $2 \mu\text{L}$ internal standard ($n\text{-C}_{13}$ alkane, -27.24‰) was spiked to the ester fraction, the esters were injected to a gas chromatograph interfaced to isotope ratio mass spectrometer to determine their stable carbon isotopic composition. GC was installed with a HP manual on-column injector and a capillary column (CIP-Sil 8CB, $60 \text{ m} \times 0.32 \text{ mm} \times 0.25 \mu\text{m}$) was used with a column oven temperature programmed from 50 to 120°C at a rate of $30^\circ\text{C min}^{-1}$ and then to 300°C at a rate of 6°C min^{-1} . Flow rate of carrier gas (He) was maintained at 1.7 mL min^{-1} .

In our manuscript, we stated that the recoveries of the target compounds were from 80% to 85%, which was not for yield during the derivatization reaction. The yield of derivatization reaction with 14% $\text{BF}_3/n\text{-butanol}$ was 100%, during which the carboxyl functional group was fully derivatized to butyl ester, and the aldehyde and keto groups were fully derivatized to dibutoxy acetal. The $\delta^{13}\text{C}$ values for diacids and related compounds could not be directly determined, so these organic species are derivatized with $\text{BF}_3/n\text{-butanol}$ to dibutyl esters which are analyzed for the stable carbon isotopic composition using a GC-IR-MS. The $\delta^{13}\text{C}$ values for individual diacids are then calculated from $\delta^{13}\text{C}$ of 1-butanol and butyl ester derivative using a mass balance equation. The accuracy of the $\delta^{13}\text{C}$ measurement for these target compounds is within 0.8‰ . The method developed for the determination of stable carbon isotopic composition of diacids and related compounds isolated from aerosols is reliable and scientific, and has been used widely in many studies (Aggarwal and Kawamura, 2008; Meng et al., 2020; Mkombe et al., 2014; Pavuluri et al., 2011; Qi et al., 2022; Shen et al., 2022; Shen et al., 2023; Wang et al., 2012; Wang and Kawamura, 2006; Xu et al., 2022a; Xu et al., 2022b; Zhang et al., 2016; Zhao et al., 2018).

Text S2. Potential source contribution function

35 To analyze air masses from different origins reaching the sampling site, 48 h backward trajectory was performed using a Hybrid Single Particle Lagrangian Integrated Trajectory (HYSPLIT4) model. To determine the potential source regions of total diacids and related compounds, the potential source contribution function (PSCF) was used and detailed information was described in our previous study (Li et al., 2021).

Text S3. Positive Matrix Factorization (PMF) model

In this study, the Environmental Protection Agency (EPA) PMF version 5.0 was used to perform the analysis (Norris et al., 2014). The PMF model has been widely applied to obtain source apportionment of organic aerosols in PM_{2.5}. It decomposes the sample data matrix into factor contribution matrix and factor outline matrix, and then the source types and their contribution are calculated based on the input data. The PMF model is described as follows:

$$X_{ij} = \sum_{k=1}^P g_{ik} f_{ki} + e_{ij} \quad (3)$$

where X_{ij} represents the concentration of the species j detected in the sample i , p represents the number of contribution factors of the samples, g_{ik} represents the relative contribution of factor k to the sample i , f_{kj} represents the mass contribution of the species j in the factor k , e_{ij} represents the error of the species j measured in the sample i . The matrix x is continuously decomposed to get minimum values of Q in order to obtain the optimal matrices g and f . The Q is defined as follows:

$$Q = \sum_{i=1}^n \sum_{j=1}^m \left(\frac{x_{ij} - \sum_{k=1}^P g_{ik} f_{ki}}{u_{ij}} \right)^2 = \sum_{i=1}^n \sum_{j=1}^m \left(\frac{e_{ij}}{u_{ij}} \right)^2 \quad (4)$$

where u_{ij} is the uncertainty of the concentration of the species j in the sample i . If the measured species concentration is below the method detection limit (MDL), it is replaced by 1/2 MDL, and the corresponding uncertainty is 5/6 MDL. When the concentration is higher than the MDL, the uncertainty is calculated as follows:

$$\text{Uncertainty} = \sqrt{(\text{error fraction} * \text{concentration})^2 + (0.5 * \text{MDL})^2} \quad (5)$$

where the error fraction is the estimation of the analytical uncertainty of the measured concentration or flux (Hovorka et al., 2015; Shrivastava et al., 2007; Wu et al., 2020).

Based on Norris et al. (2004) and Nayebar et al. (2018), the comparatively high signal-to-noise (S/N) ratios > 5 could be labeled as “strong” to enhance their effect on the model results for all species. The S/N ratios of species used in PMF model were higher than 8, thus these species were labeled as “strong”.

The concentrations of selected major diacids (including C₂, C₃, C₄, C₉, and Ph), oxoacids (including Pyr and ωC₂), and α-dicarbonyls used in the PMF model accounted for 76.1 ± 6.0% and 82.8 ± 6.6% of total concentration of detected organic components (TDOCs) before and during the LCD, respectively. The other organic compounds (including C₅-C₈, C₁₀, C₁₁, iC₄-iC₆, M, F, mM, iPh, tPh, kC₃, kC₇, ωC₃, ωC₄ and ωC₇-ωC₉) presented relatively low concentrations and occasionally were below the detection limit, which could enhance the uncertainties. Moreover, these unconsidered organic compounds were less indicative for identified sources than the selected major acids. Thus, we kept those organic compounds with low concentrations out of the model.

As discussed in Section 3.3, SO₄²⁻ and LWC could be considered as significant influencing factors during the aqueous formation of C₂ before the LCD, and O₃ was a key influencing factor during the photochemical formation of C₂ during the

LCD. Therefore, those species have been put in the PMF model to distinguish the aqueous pathway from the gaseous photochemical pathway of the diacids formation.

70 The PMF model provides three uncertainty estimation methods to examine the robustness of the solution, namely, bootstrap (BS), displacement (DISP), and bootstrap combined displacement (BS-DISP), which can also provide a distribution range of the factor profiles through resampling for multi times and displacing with the Q value change within the preset value. Further details related to the error estimation methods can refer to Brown et al. (2015), Vossler et al. (2016), and Wang et al. (2018). In this study, PMF solutions of 4-6 factors were analyzed and showed the convergence result. The relevant Q values and $Q_{\text{true}}/Q_{\text{robust}}$ for these solutions were shown in Table S5. Finally, the five-factor solution was selected as the best solution with physical means before ($Q_{\text{true}}/Q_{\text{robust}}=1.05$, $Q_{\text{true}}/Q_{\text{expected}}=1.27$) and during the LCD ($Q_{\text{true}}/Q_{\text{robust}}=1.05$, $Q_{\text{true}}/Q_{\text{expected}}=1.36$),
75 respectively. The error code of DISP before or during the LCD was 0 (Table S6), indicating no error. The values in the first row of $dQ^{\text{max}} = 4$ were zero before and during the LCD (Table S6), indicating that there was no significant rotational ambiguity and that the solution was sufficiently robust to be used. Mapping over 82% of the factors before or during the LCD indicated that the BS uncertainties could be interpreted and the number of factors may be appropriate. The Largest decrease in Q before
80 and during the LCD was -0.048% and -0.061%, respectively (Table S6), suggesting the five-factor solution was more reliably attributed to sources than the four-factor solution and six-factor solution.

Text S4. Quality assurance/quality control (QA/QC) of online air quality data

85 Those measured parameters were obtained from one of the State Controlling Air Sampling Sites in Jinan (<https://www.aqistudy.cn/>). The QA/QC was operated strictly by professionals followed by “Technical Specification for Automatic Monitoring of Ambient Air Quality”, “Technical Standard for Ambient Air Quality Monitoring”, and “Technical Instruction” formulated by the Ministry of Ecology and Environment of the People's Republic of China.

Text S5. Aqueous phase formation of SO_4^{2-} before the LCD

90 The aqueous phase conversion of dissolved SO_2 to SO_4^{2-} is probably the most important chemical formation pathway in the North China Plain during the wintertime (Sun et al., 2013; Wang et al., 2014; Zhang et al., 2015). A few studies have demonstrated that LWC and RH are considered as key factors determining SO_4^{2-} formation via aqueous reactions (Bikkina et al., 2017; Cheng et al., 2016). We observed a robust relationship of SO_4^{2-} with RH ($R^2 = 0.58$) and LWC ($R^2 = 0.71$) before the LCD (Fig. S3), indicating that the aqueous oxidation determined by LWC and RH played a dominant role in the formation
95 of SO_4^{2-} before the LCD.

Table S1. Correlation coefficients (R^2 , $P < 0.01$) of levoglucosan and K^+ with EC, OC, WSOC, C_2 , C_9 , diacids, and total detected organic components (TDOCs).

	EC	OC	WSOC	K^+	C_2	C_9	Diacids	TDOCs
Before the LCD								
Levoglucosan	0.62	0.66	0.45	0.77	0.61	0.74	0.62	0.73
K^+	0.57	0.56	0.55	1.00	0.68	0.64	0.59	0.65
During the LCD								
Levoglucosan	0.15	0.09	0.05	0.86	0.02	0.06	0.02	0.03
K^+	0.10	0.08	0.02	1.00	0.03	0.02	0.11	0.13

Table S2. Statistic test for the differences in concentrations and mass ratios of major species in PM_{2.5} before and during the LCD in Jinan.

Species/ Ratios	<i>p</i>
PM _{2.5}	0.000^a
PM ₁₀	0.000^a
CO	0.000^a
SO ₂	0.000^a
NO _x	0.000^a
O ₃	0.000^a
RH (Relative humidity)	0.000^a
Wind speed	0.000^a
OC (Organic carbon)	0.000^a
EC (Elemental carbon)	0.000^a
WSOC (Water-soluble organic compounds)	0.098 ^b
SIA (Total concentration of SO ₄ ²⁻ , NO ₃ ⁻ , and NH ₄ ⁺)	0.001^a
LWC (Liquid water content)	0.000^a
pH _{is} (Particle in-situ pH)	0.256 ^b
C ₂ (Oxalic acid)	0.004^b
C ₉ (Azelaic acid)	0.000^a
Ph (Phthalic acid)	0.026^b
Diacids (Dicarboxylic acids)	0.018^b
Oxoacids	0.003^a
α-Dicarbonyls	0.881 ^b
TDOCs (Total detected organic components)	0.014^a
Levoglucosan	0.003^a
C ₂ /Diacids	0.000^a
C ₂ /TDOCs	0.005^a
C ₂ /C ₄ (Succinic acid)	0.000^a
C ₃ (Malonic acid)/C ₄	0.000^a
C ₉ (Azelaic acid)/Diacids	0.000^a
Ph/Diacids	0.027^a

^aWhen the variables follow a normal distribution, ANOVA was employed to analyze the statistically significant differences between these two datasets before and during the LCD. ^bOtherwise, the Kruskal-Wallis test was used. Bold font represented the species (or ratios) that were statistically significantly different ($p < 0.05$).

Table S3. Statistic test for the day-night differences in concentrations and mass ratios of major species in PM_{2.5} before or during the LCD, respectively.

Species/ Ratios	Before the LCD	During the LCD
	<i>p</i>	<i>p</i>
C ₂	0.011^a	0.042^a
Diacids	0.032^a	0.046^b
C ₂ /Diacids	0.018^a	0.041^b
C ₂ /TDOCs	0.017^b	0.029^a
Na ⁺	0.764 ^b	0.362 ^b
Ca ²⁺	0.443 ^b	0.637 ^b
Mg ²⁺	1.000 ^b	0.784 ^b
C ₂ /Levogluconan	0.037^a	0.074 ^b
C ₂ /K ⁺	0.000^a	0.081 ^b

^aWhen the variables follow a normal distribution, ANOVA was employed to analyze the statistically significant differences during the daytime and nighttime. ^bOtherwise, the Kruskal-Wallis test was used. Bold font represented the species (or ratios) that were statistically significantly different ($p < 0.05$).

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Table S4. Comparison of C₂, diacids, and TDOCs concentrations (ng m⁻³), and C₃/C₄ ratio in Jinan with those in other Asian megacities during the wintertime.

Location	Period	Size	C ₂	Diacids	TDOCs	C ₂ /C ₄	C ₃ /C ₄	Data source
14 Chinese cities	13–14 Jan 2003	PM _{2.5}	558±351	904±480	974±499	7.0	0.5	(Ho et al., 2007)
Beijing, China	Dec 2013	PM _{2.5}	149±123	366±261	450±368	4.7	0.6±0.1	(Zhao et al., 2018)
Chengdu, China	7–23 Jan 2013	PM _{2.1}	1380	3390	4285	3.2	0.5±0.1	(Li et al., 2015)
Xi'an, China	Jan–Feb 2009	PM ₁₀	1162±570	1843±810	2207±1044	11.9	1.0	(Cheng et al., 2013)
Tianjin, China	23 Nov–11 Dec 2016	PM _{2.1}	526±335	1223±623	1438±739	3.2	0.3	(Devineni et al., 2023)
Liaocheng, China	17 Jan–16 Feb 2016	PM _{2.5}	817±544	1415±899	1677±1026	3.6±0.9	0.4±0.2	(Meng et al., 2020)
Guangzhou, China	Dec 2006–Jan 2007	PM _{2.5}	182±106	384±171	414±184	9.9	0.7	(Ho et al., 2011)
Tianjin, China	1–28 Feb 2019	PM _{2.5}	668±470	1137±707	1561	7.6	1.0	(Zhao et al., 2023)
Padori, South Korea	1–28 Feb 2019	PM _{2.5}	389±204	582±266	750	8.3	0.7	(Zhao et al., 2023)
Daejeon, South Korea	1–28 Feb 2019	PM _{2.5}	516±327	784±446	1108	10.0	0.7	(Zhao et al., 2023)
Ulaanbaatar, Mongolia	Nov 2007–Jan 2008	PM _{2.5}	107±28	536±156	697±207	1.8±0.8	0.2±0.1	(Jung et al., 2010)
Chennai, India	23–28 Jan 2007	PM ₁₀	472±137	695 ± 176	765	11.8	1.5	(Pavuluri et al., 2010)
Chennai, India	29 Jan–6 Feb 2007	PM ₁₀	380±89	641 ± 157	719	7.9	1.2	(Pavuluri et al., 2010)
Tokyo, Japan	20–21 November 1989	TSP	186	438	595	3.9	0.9	(Kawamura and Yasui, 2005)
Jinan, China	6–23 Jan 2020	PM _{2.5}	181±48	351±92	437±117	3.9±1.5	0.3±0.1	This study
Jinan, China	31 Jan–17 Feb 2020	PM _{2.5}	239±108	386±127	486±144	8.4±3.4	1.6±0.4	This study

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Table S5. Q values for PMF Analysis with different number of factors.

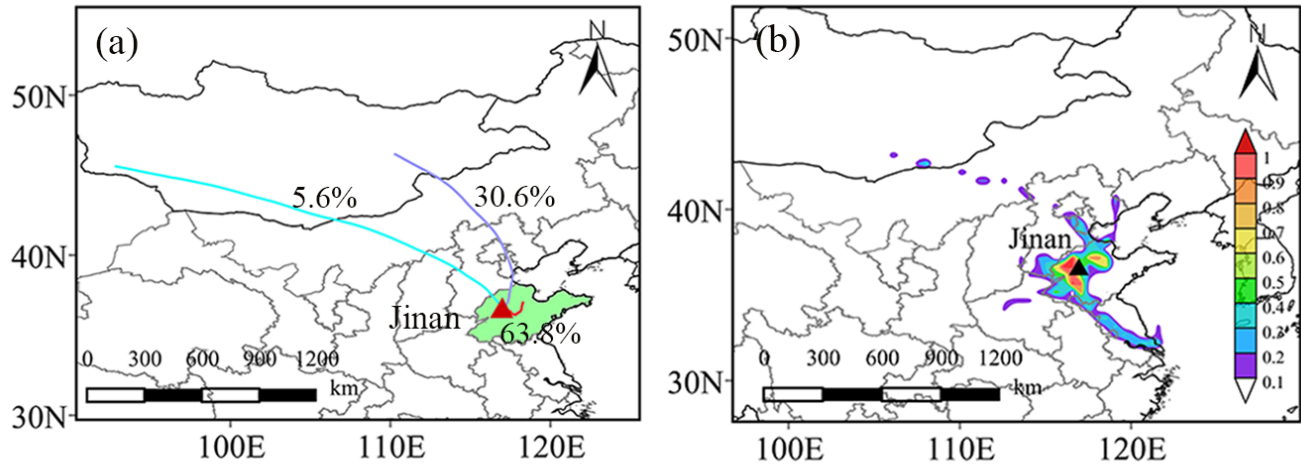
Number of factors	*R ² for all input species	Q _{true}	Q _{robust}	Q _{true} /Q _{expected}	Q _{true} /Q _{robust}
(a) Before the LCD					
4	0.63-0.98	1556.1	1480.2	2.241	1.051
5	0.68-0.99	1176.9	1123.5	1.273	1.047
6	0.60-0.98	905.2	851.6	2.454	1.062
(a) During the LCD					
4	0.32-0.96	2722.0	2396.5	2.624	1.136
5	0.52-0.99	1859.3	1768.2	1.360	1.052
6	0.37-0.84	1334.1	1294.6	2.912	1.071

*R² between the measured and predicted species

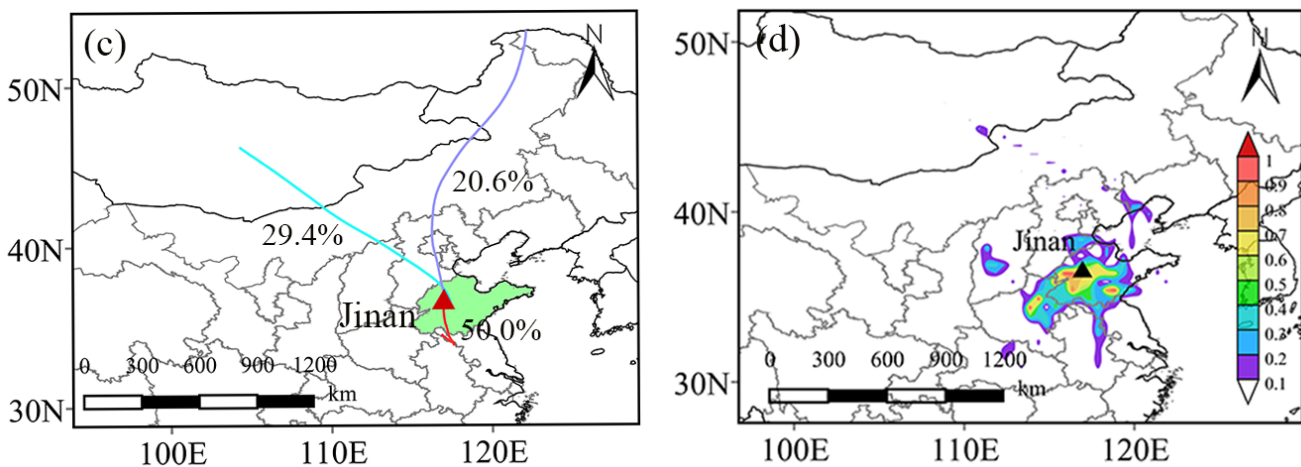
Table S6. Summary of error estimation diagnostics from BS and DISP for PMF (a) Before the LCD and (b) During the LCD.

(a) Before the LCD						
DISP Diagnostics		Error code: 0		Largest decrease in Q:-0.048		
Factor	$dQ^{\max}=4$	0	0	0	0	0
	$dQ^{\max}=8$	0	0	0	1	1
Swaps	$dQ^{\max}=15$	0	1	3	6	0
	$dQ^{\max}=25$	0	1	5	8	10
(b) During the LCD						
DISP Diagnostics		Error code: 0		Largest decrease in Q:-0.061		
Factor	$dQ^{\max}=4$	0	0	0	0	0
	$dQ^{\max}=8$	0	0	0	0	0
Swaps	$dQ^{\max}=15$	0	0	1	1	0
	$dQ^{\max}=25$	0	1	8	10	6

Before the LCD



During the LCD



125 **Figure S1.** 48-h backward trajectories of air masses arriving at Jinan (a) before the LCD and (b) during the LCD. Potential source contribution function (PSCF) analysis of total detected organic compounds (TDOCs) in PM_{2.5} (c) before the LCD and (d) during the LCD.

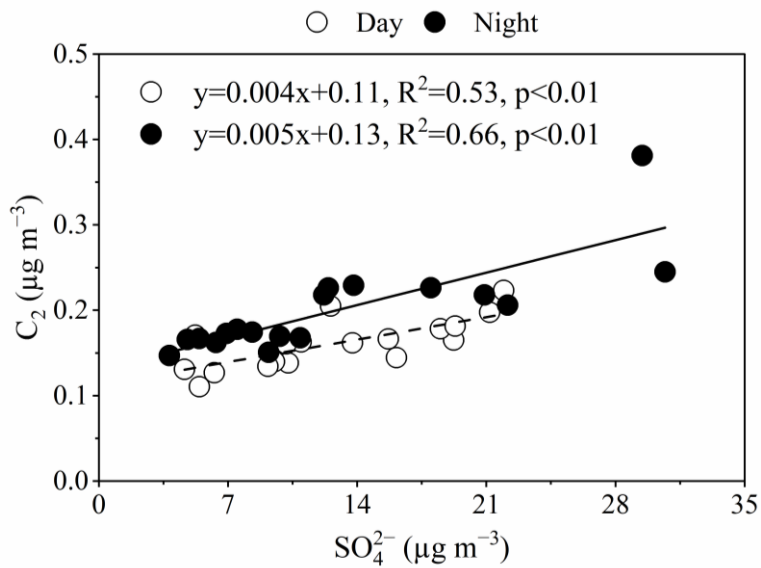


Figure S2. Correlation analysis for C_2 and SO_4^{2-} during the daytime and nighttime before the LCD.

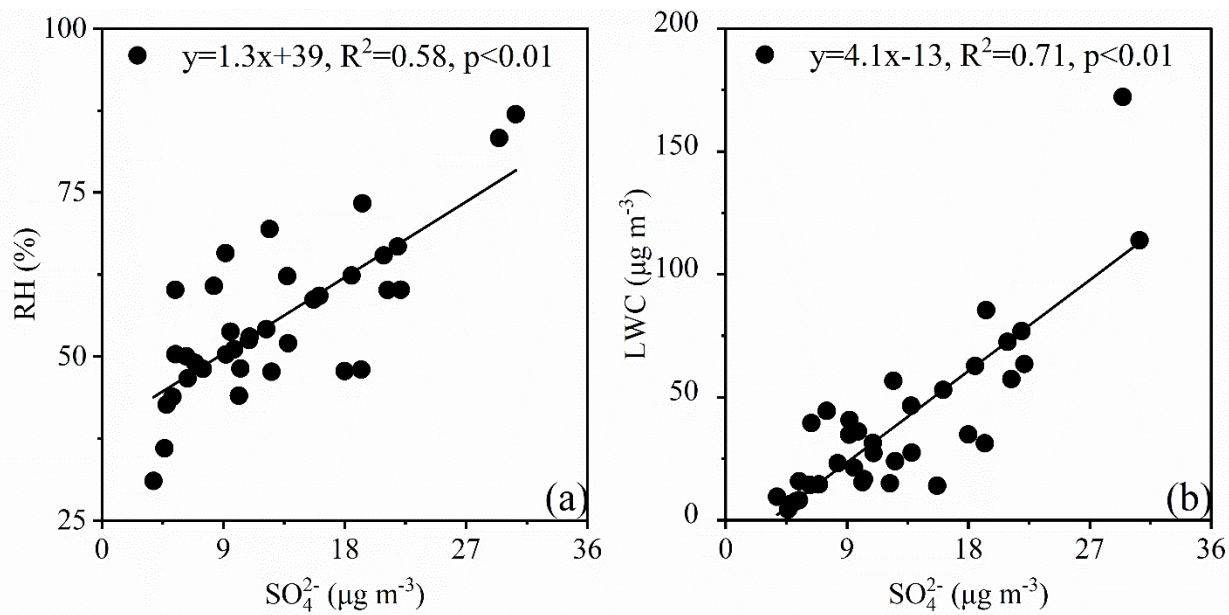
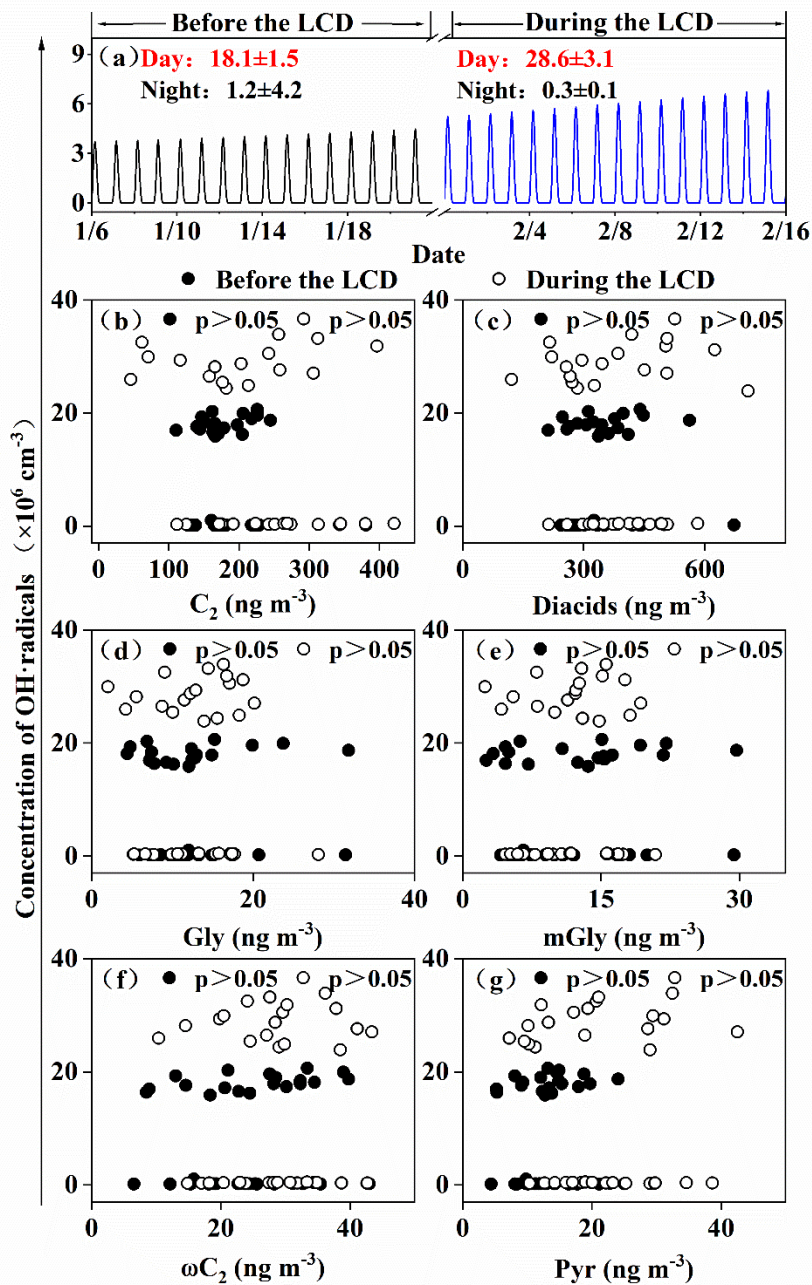


Figure S3. Relationships of SO_4^{2-} with (a) RH and (b) LWC before the LCD.



135 Figure S4. (a) Time series of OH· radical concentration per hour, and correlations of OH· radicals with (b) C₂, (c) Diacids, (d) Gly, (e) mGly, (f) ωC₂, and (g) Pyr before and during the LCD.

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