



Supplement of

Exploring the amplified role of HCHO in the formation of HMS and O₃ during the co-occurring PM_{2.5} and O₃ pollution in a coastal city of south-east China

Youwei Hong et al.

Correspondence to: Youwei Hong (ywhong@iue.ac.cn) and Likun Xue (xuelikun@sdu.edu.cn)

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31 **The Observation-based model (OBM)**

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33 A chemical box model, as one of the important methods for analyzing atmospheric
34 chemical processes, was run based on the platform of the Framework for 0-Dimensional
35 Atmospheric Modeling (F0AM), which has broad application potential in deeply
36 exploring atmospheric observation data and comprehensively understanding the
37 regional atmospheric pollution. About the chemical mechanism, the F0AM
38 incorporating the latest chemical mechanism version of MCM-v3.3.1 (MCM,
39 <http://mcm.leeds.ac.uk/MCM/>, last access: 13 May 2022) was applied to simulate the
40 detailed photochemical processes and quantify the reaction rates of HCHO mechanism,
41 and the MCM mechanism introduced 142 VOCs and about 20,000 chemical reactions.

42 About the uncertainties of the model simulation results, the index of agreement (IOA)
43 was used to judge the reliability of the model simulation results, as follows:

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$$IOA = 1 - \frac{\sum_{i=1}^n (O_i - S_i)^2}{\sum_{i=1}^n (|O_i - \bar{O}| + |S_i - \bar{O}|)^2}$$

45 where S_i is simulated value, O_i represents observed value, \bar{O} is the average observed
46 values, and n is the sample number. The IOA range is 0-1, and the higher the IOA value
47 is, the better agreement between simulated and observed values is. In this study, the
48 simulation results (the IOA is approximately 0.80) are reasonable, and the performance
49 of the OBM-MCM model was acceptable.

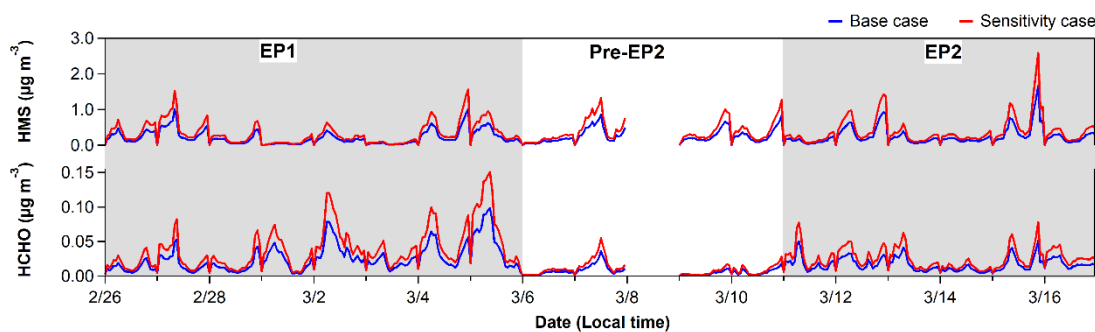
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52 **Uncertainty evaluation of OBM analysis:**

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54 Since aqueous HCHO was not available during the observation, which was the key
55 chemical components influencing the subsequent HMS modeling, we established the
56 aqueous HCHO concentrations by the mass transfer processed between the gas- and
57 particle- phase. Among this processes, the uncertainties were introduced somehow by
58 the Henry's law constant adopted in the model. We conducted a sensitivity test with 2
59 folds of current used Henry's law constant of HCHO, with a value of $6 \times 10^7 \text{ M atm}^{-1}$.
60 As shown in Fig S0, the modeled aqueous HCHO as well as HMS concentrations
61 increased with the increase of Henry's law constant in the sensitivity case, with increase
62 of 0.01 and $0.13 \mu\text{g m}^{-3}$ for aqueous HCHO and HMS, respectively. It is quite
63 reasonable considering the increased solubility of HCHO. On the other hand, the
64 modeled HCHO and HMS were still exhibited higher concentrations during the
65 pollution episodes (EP1 and EP2), of which higher precursors and favorable aerosol
66 properties enhanced the heterogeneous processes. Therefore, the impacts of HCHO
67 Henry's law constant approximations on the conclusions are supposed to be minor.



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 69 **Fig S1. Time series of the modeled aqueous HCHO and HMS concentrations with the model**
 70 **used Henry's law constant of HCHO (referred as base case, the blue line) and 2 folds of the**
 71 **model used Henry's law constant of HCHO (referred as sensitivity case, the red line),**
 72 **respectively.**
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74 In the atmosphere, gaseous HCHO with high Henry's law constant can partition into
 75 aerosol or cloud/fog water, then the aqueous HCHO can react with H₂O₂ to form HMHP,
 76 with an upper bound forward rate constant of 100 (± 35) M⁻¹ s⁻¹ and reverse rate constant of
 77 0.6 (± 0.2) s⁻¹ (Dovrou et al., 2022). In the gas phase, HMHP is mainly formed by the
 78 hydration of CH₂OO Criegee radicals and then can partition into aerosol water, which have
 79 the potential to oxidize SO₂(aq) to form sulfate. From the perspective of contributions of
 80 HMHP paths to sulfate formation, the researcher revealed that the HCHO-catalysis path
 81 (HCHO + H₂O₂) under fast equilibrium is more significant than HMHP-direct path
 82 (hydration of CH₂OO in the gas-phase) to global sulfate, so HCHO-catalysis path is
 83 significant for HMHP production. So, in the future, we will carry out the observation of
 84 H₂O₂ for further evaluate the effects of HCHO + H₂O₂ reaction on the HMHP production.

85 We used 2 folds of the HCHO Henry's law constant as the sensitivity case and conducted
 86 the simulation of the aqueous HCHO as well as HMS concentrations. We found that
 87 aqueous HCHO and HMS exhibited slight higher concentrations due to the increased
 88 solubility of HCHO, with increases of 0.01 and 0.13 µg m⁻³ for HCHO and HMS,
 89 respectively. However, the modeled HCHO and HMS were still exhibited higher
 90 concentrations during the pollution episodes (EP1 and EP2), which still able to support our
 91 major conclusions. On the other hand, we compared our modeling aqueous HCHO
 92 concentrations with previous observations. As shown in the following tables, our modeling
 93 aqueous HCHO concentrations are comparable and in the same order of magnitude with
 94 previous observations. Therefore, uncertainties due to the lack of aqueous HCHO
 95 measurement and the HCHO Henry's law constant approximations would take minor
 96 impacts on our major conclusions.

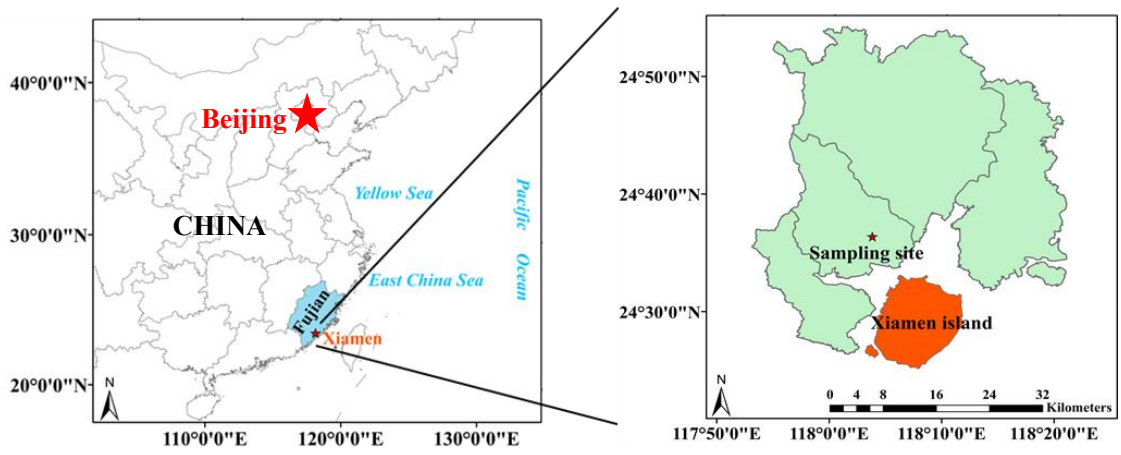
102 **Backward trajectory analysis**

103 Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) was used to
104 analyze the air masses before and during the PM_{2.5} and O₃ pollution period.
105 Meteorological data used here were obtained from the Global Data Assimilation System
106 (GDAS) with a 1°×1° spatial resolution and 3-h temporal resolution. The 72-h backward
107 trajectories at a height of 100 m obtained from the National Oceanic and Atmospheric
108 Administration were run every hour. Cluster analysis was performed, and four clusters
109 were determined based on the total spatial variance (TSV) (Chen et al., 2022; Ji et al.,
110 2022).

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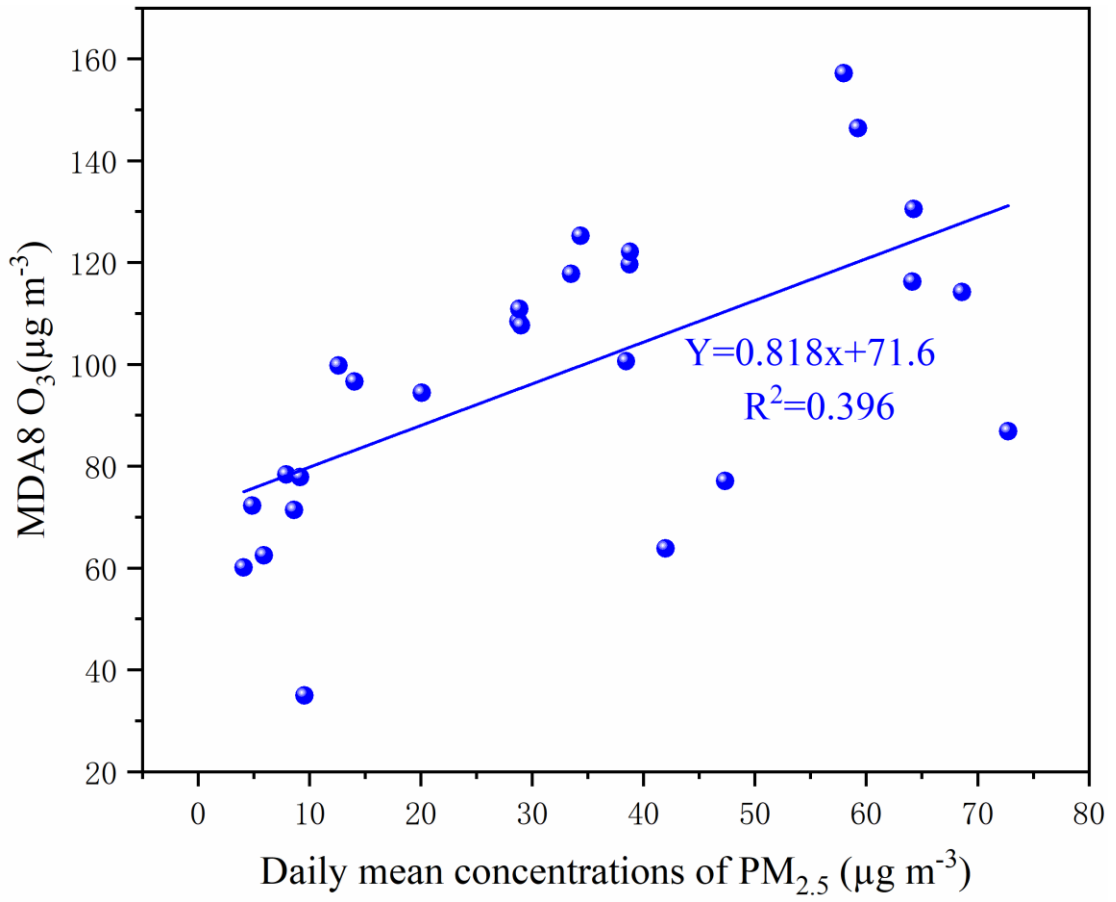
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Fig S2. Location of the observation site in a coastal city of Southeast China

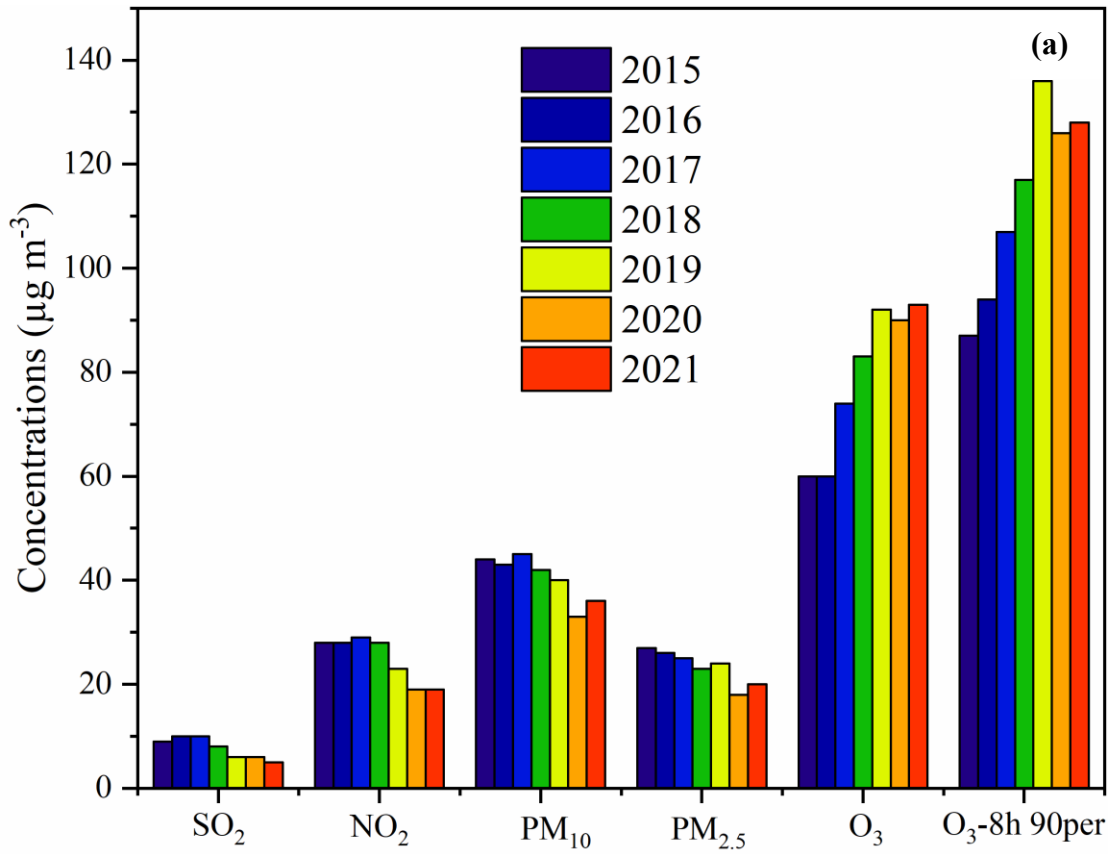


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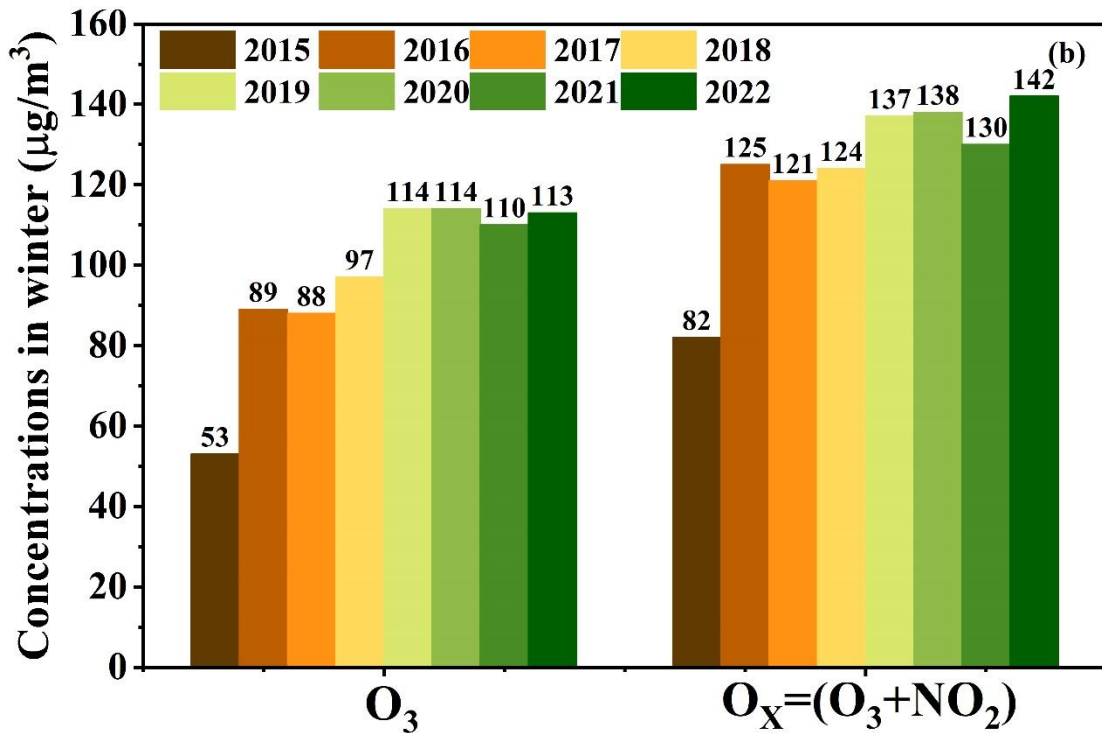
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Fig S3. The correlation of PM_{2.5} and MDA8 O₃ concentrations during the whole periods



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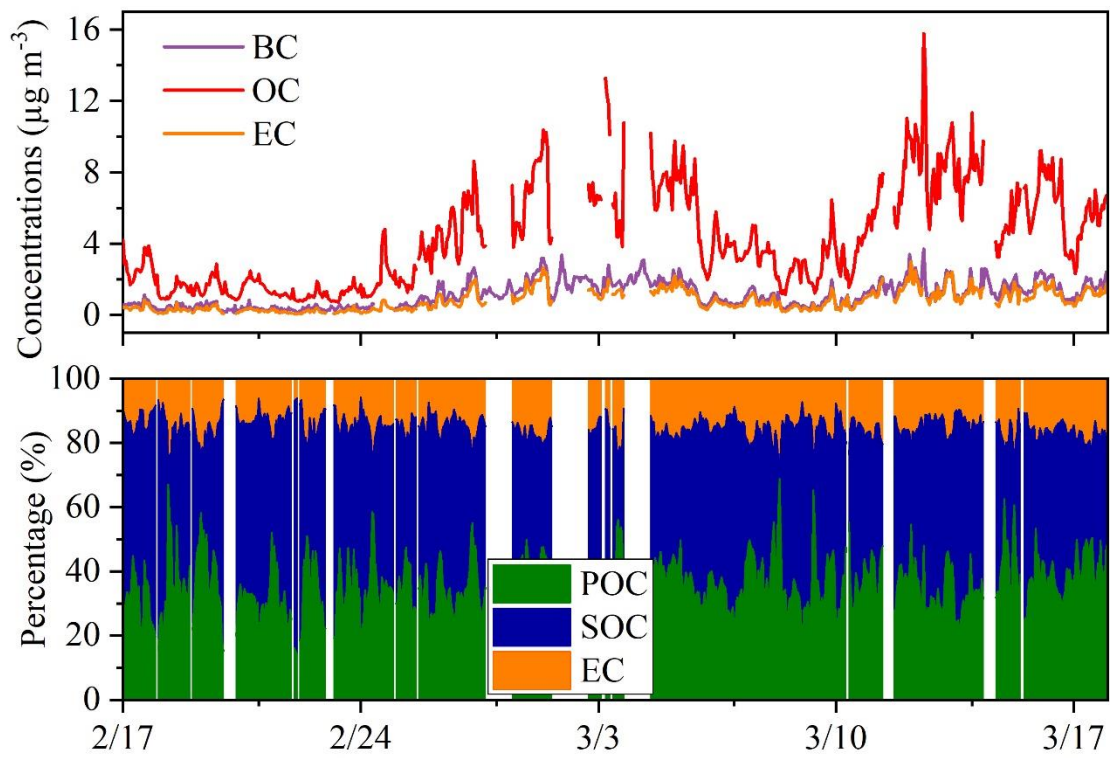
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Fig S4. Inter-annual mean concentrations of criteria air pollutants (a) and averaged concentrations of O₃ and O_x in winter in Xiamen, a coastal city of Southeast China



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135 **Fig S5. The concentrations and percentages of OC, EC, and BC during the whole periods**

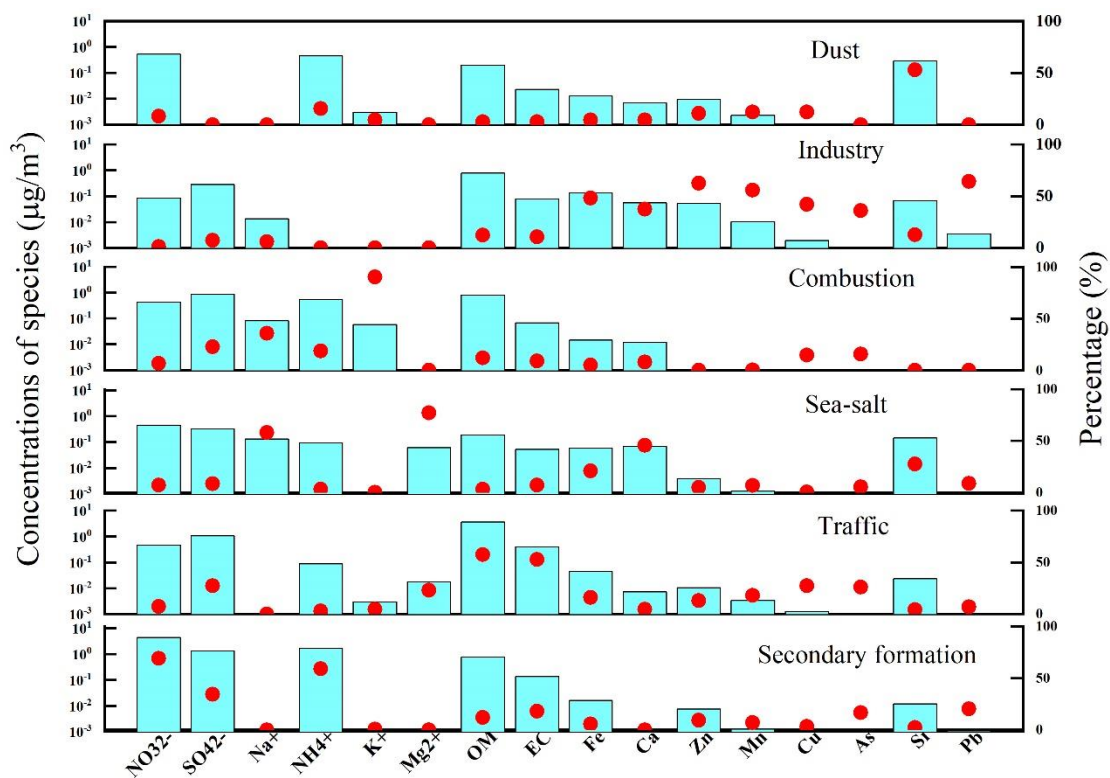
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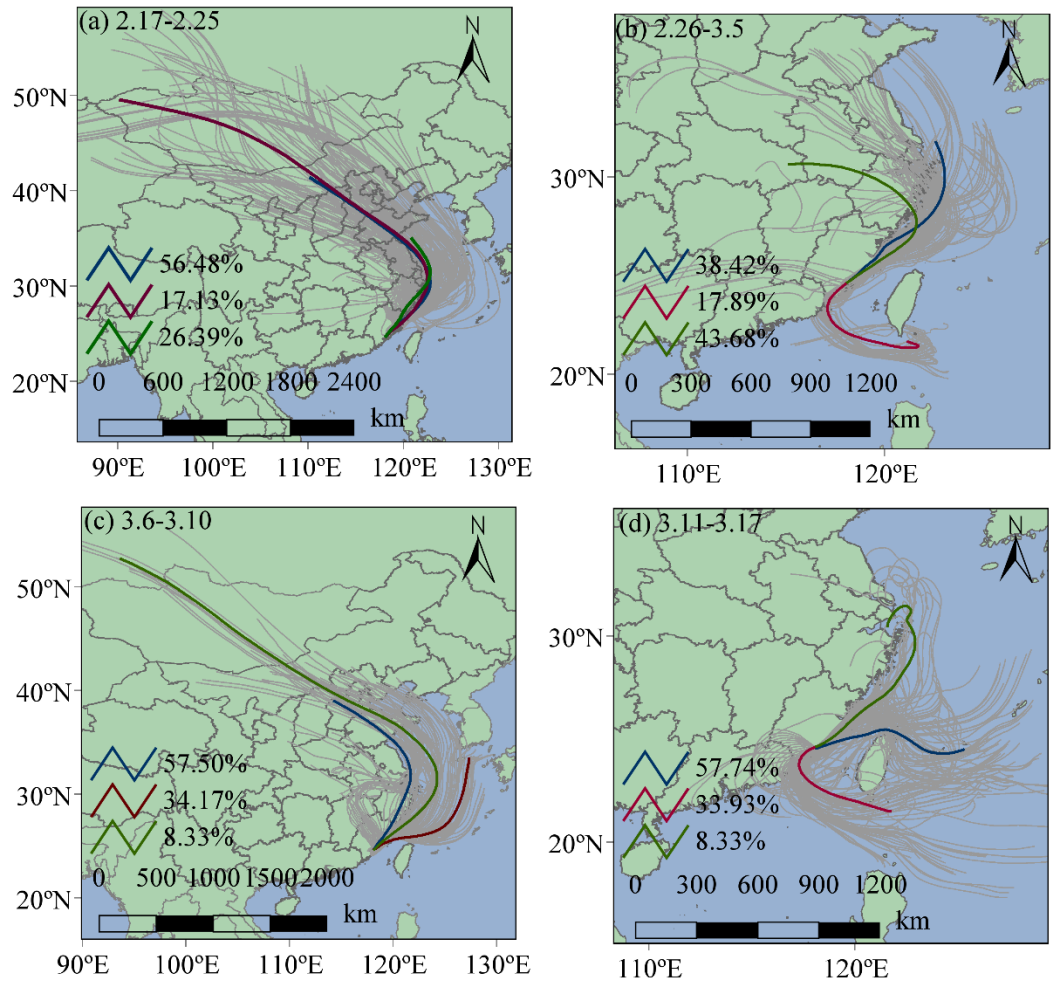
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Fig S6. The factor profiles and the contribution of various sources to $\text{PM}_{2.5}$ by the positive matrix factorization (PMF) model analysis. The bars represent the concentrations of various species and the dots represent the contributions of species to the factors.

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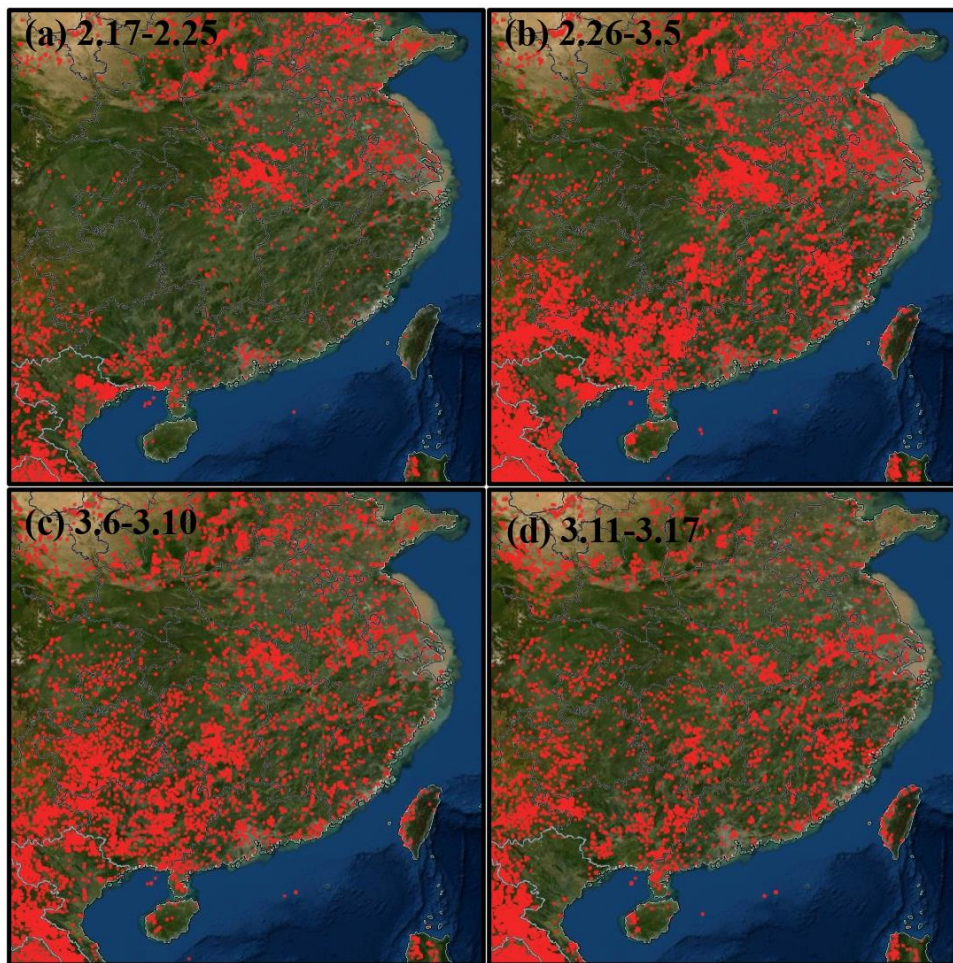
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Fig.S7 Backward trajectory clusters at the monitoring site under different periods

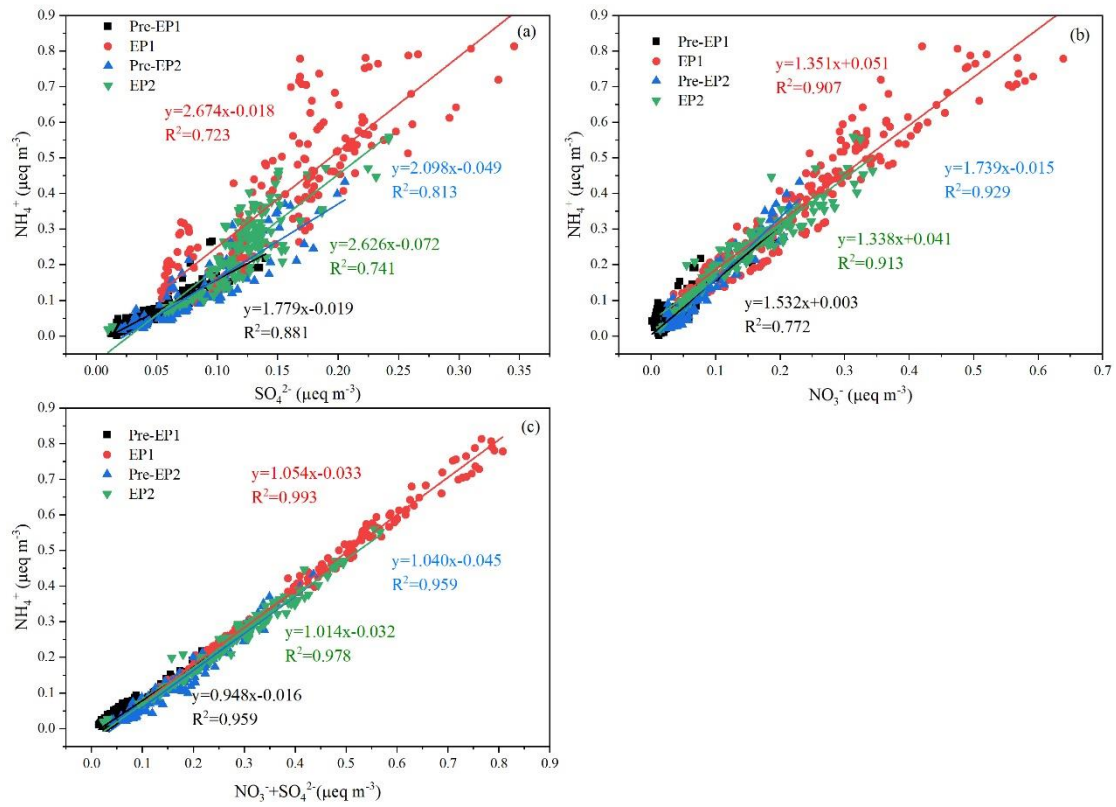


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156 Fig.S8 The distribution of fire spots around the monitoring site under different periods (From
157 the Fire Information for Resource Management System,
158 <https://firms.modaps.eosdis.nasa.gov/firemap/>)

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Fig S9.The correlations between (a) NH_4^+ and SO_4^{2-} , (b) NH_4^+ and NO_3^- , and (c) NH_4^+ and $\text{NO}_3^- + \text{SO}_4^{2-}$

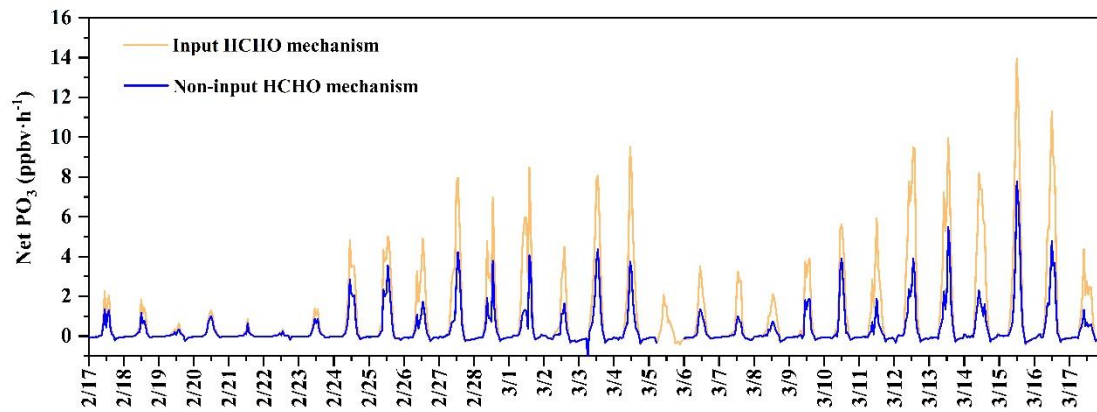
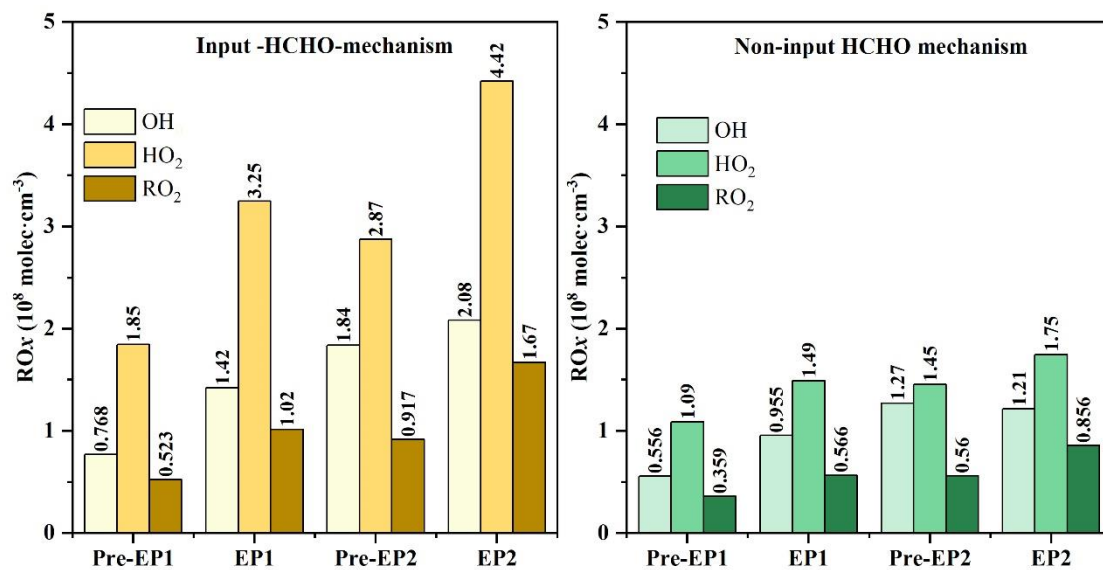


Fig S10. The net O₃ production rates modeled by the OBM with or without HCHO mechanism

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Fig S11. The ROx concentrations modeled by the OBM with or without HCHO mechanism

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207 Table S1 Mean concentrations of air pollutants, chemical compositions and meteorological
208 parameters under different pollution stages

	Pre-EP1	EP1	Pre-EP2	EP2
PM _{2.5} ($\mu\text{g m}^{-3}$)	9.03±6.28	51.94±19.02	33.53±10.41	35.34±23.51
PM ₁₀ ($\mu\text{g m}^{-3}$)	13.68±10.51	74.43±24.40	57.46±21.11	58.21±28.98
O ₃ ($\mu\text{g m}^{-3}$)	60.82±18.57	67.12±44.73	76.58±31.83	58.98±47.97
CO($\mu\text{g m}^{-3}$)	531.02±63.18	742.86±121.34	379.17±95.65	577.98±97.24
SO ₂ ($\mu\text{g m}^{-3}$)	9.16±3.47	13.74±7.61	16.26±2.68	18.20±6.58
NO ₂ ($\mu\text{g m}^{-3}$)	12.40±7.00	31.38±14.20	22.01±10.82	32.56±11.08
NO($\mu\text{g m}^{-3}$)	1.5±1.31	5.56±9.51	2.04±14.29	6.13±8.02
Ox($\mu\text{g m}^{-3}$)	73.22±12.78	98.50±29.45	88.59±21.32	91.54±29.57
BC($\mu\text{g m}^{-3}$)	0.45±0.19	1.69±0.56	0.78±0.32	1.61±0.54
OC($\mu\text{g m}^{-3}$)	1.69±0.81	6.36±2.16	3.20±1.52	7.48±5.03
EC($\mu\text{g m}^{-3}$)	0.27±0.17	1.23±0.50	0.59±1.51	1.29±0.51
SO ₄ ²⁻ ($\mu\text{g m}^{-3}$)	1.96±1.26	7.07±2.91	3.87±2.24	5.87±1.70
NO ₃ ⁻ ($\mu\text{g m}^{-3}$)	2.02±1.81	14.95±8.34	4.83±2.83	9.69±4.89
NH ₄ ⁺ ($\mu\text{g m}^{-3}$)	0.96±0.90	6.77±3.43	2.16±1.63	4.46±1.88
NOR	0.16±0.07	0.32±0.08	0.18±0.06	0.24±0.05
SOR	0.19±0.07	0.38±0.18	0.19±0.06	0.26±0.08
T(°C)	12.38±3.89	20.56±3.90	18.65±4.17	23.93±4.30
RH(%)	69.41±12.45	63.17±10.32	54.30±15.52	62.17±11.05
P(kPa)	101.14±0.36	100.71±0.30	100.68±0.15	100.17±0.21
WS(m/s)	1.64±0.70	1.02±0.51	1.37±0.59	0.93±0.54
UV(W/m ²)	5.88±9.58	9.31±13.94	11.08±16.09	11.10±15.05
JNO ₂ ($\times 10^{-3} \text{ s}^{-1}$)	1.50±2.45	2.49±3.57	2.91±3.94	2.95±3.79
HONO($\mu\text{g m}^{-3}$)	0.75±0.36	3.10±2.25	1.19±0.58	4.19±2.68
HCHO(ppb)	0.68±0.71	2.94±1.01	1.59±0.71	3.59±1.36
TVOCs(ppb)	14.47±7.14	29.39±11.65	14.28±6.18	29.91±12.05
Aromatics(ppb)	1.29±1.36	3.21±1.92	1.54±1.29	3.06±1.87

Alkynes(ppb)	0.78±0.41	1.04±1.00	0.12±0.06	0.77±0.72
Alkanes(ppb)	6.48±2.06	10.27±5.07	4.46±1.92	9.73±4.46
Alkenes(ppb)	0.75±0.52	1.28±0.84	0.49±0.42	1.37±1.09
OVOCs(ppb)	0.87±0.35	1.21±0.23	1.01±0.41	1.22±0.25
Halohydrocarbon(ppb)	4.30±3.69	12.40±4.77	6.65±3.03	13.76±5.76

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Table S2 Comparisons of atmospheric HCHO in China

Location	Type	Seasons	Mean(\pm SD) (ppbv)	Range (ppbv)	Reference
Xiamen, China	Urban	Spring	2.9 \pm 0.3	0.25-8.34	[12]
	Urban	Autumn	3.2 \pm 1.4	0.38-7.56	[12]
	Urban	Winter	1.95 \pm 1.06	0.23-6.22	This study
Shenzhen, China	Urban	Spring	3.4 \pm 1.6	N.A. ^a	[2]
Hong Kong, China	Urban	Spring (2012)	3.02 \pm 0.91	N.A.	[6]
Shanghai, China	Suburban	Spring (2018)	6.7 \pm 3.6	N.A.-20.9	[4]
Shanghai, China	Suburban	Spring (2018)	5.01 \pm 3.80	N.A.-18.69	[10]
Beijing, China	Urban	Summer (2013)	11.39 \pm 5.58	N.A.	[5]
Shanghai, China	Urban	Summer (2018)	3.31 \pm 1.43	N.A.	[8]
Shenzhen, China	Urban	Summer	5.0 \pm 4.4	N.A.	[2]
Hong Kong, China	Urban	Summer (2011)	8.07 \pm 1.94	N.A.	[6]
Guangzhou, China	Urban	Summer (2010)	6.69 \pm 1.98	N.A.	[3]
Yantai, China	Urban	Summer	3.90 \pm 1.12	N.A.	[3]
Beijing, China	Suburban	Summer	11.17 \pm 5.32	3.14-35.08	[9]
Shanghai, China	Suburban	Summer	2.2 \pm 1.8	N.A.-9.4	[7]
Wuhan, China	Suburban	Summer	2.1 \pm 0.2	0.6-4.1	[1]
Hong Kong, China	Urban	Autumn (2011)	2.96 \pm 0.70	N.A.	[6]
Guangdong, China	Urban	Autumn	4.12 \pm 1.02	2.56-7.31	[11]
Beijing, China	Urban	Winter (2013)	7.39 \pm 5.26	N.A.	[5]
Shenzhen, China	Urban	Winter	4.2 \pm 2.2	N.A.	[2]
Hong Kong, China	Urban	Winter (2012)	2.70 \pm 1.20	N.A.	[6]
Guangzhou, China	Urban	Winter	3.35 \pm 1.38	N.A.	[3]

216 Note: (a) N.A. means no relevant data available.

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220 Table S3. The detection limits, time resolutions and measured uncertainties of air
221 pollutants

Species	Measurement Techniques	Uncertainties	Detection limits	Time resolution
HCHO	FMS-100, Focused Photonics Inc., Hangzhou, China	≤5%	50 pptv	1 s
PAN	PANs-1000, Focused Photonics Inc., Hangzhou, China	±10%	50 pptv	5 min
O ₃	Model 49i, Thermo Fischer Scientific, USA	±5%	1 ppbv	1 h
NO _x	Model 42i, Thermo Fischer Scientific, USA	±10%	0.5 ppbv	1 h
CO	Model 48i, Thermo Fischer Scientific, USA	±5%	40 ppbv	1 h
SO ₂	Model 43i, Thermo Fischer Scientific, USA	±10%	0.5 ppbv	1 h
VOCs	GC-FID/MS, TH-300B, Wuhan, China	±10%	20-300 pptv	1 h
HONO	MARGA, ADI 2080, Applikon Analytical B.V., the Netherlands	±20%	50 pptv	1 h

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