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Supplement of

Dramatic changes in atmospheric pollution source contributions for a coastal megacity in northern China from 2011 to 2020

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Text S1 Determination details of elements, water-soluble ions and carbon species.

Element analysis: For samples in 2011-2012, half of polypropylene fiber filters were cut into fragments and placed into a conical flask with the deionized water. Acid solutions (15 mL of HNO₃ and 5 mL of HClO₄) were added into the flask and the flask was heated by electric stove, and then heat on the electric furnace maintaining below 100 °C. The solution was evaporated until about 3 mL residual left. After being cooled and filtered, the solution was decanted into a test tube and diluted to 10 mL with deionized water. Alkali solution was used for measuring the concentrations of silicon. For QA/QC, standard reference materials were pre-treated and analyzed with the same procedure, with the recovered values for all the target elements falling into the range or within 5% of certified values. For samples in 2016, one-fourth of polypropylene filters were cut into fragments and placed into a microwave digestion tube. 10 mL mixture of nitric acid and hydrochloric acid were put into the tube, and then tube was digested by microwave oven. The temperature was raised to 200 °C and kept for 15 min. After digestion and component cooling, add 10 mL ultrapure water into each tube, let stand for 30 min, and then fix the volume to 50 mL. Alkali solution was used for measuring the concentrations of silicon. After centrifugation, take the supernatant for subsequent instrumental analysis. For quality assurance and quality control (QA/QC), standard reference materials were pre-treated and analyzed with the same procedure, with the recovered values for all the target elements falling into the range or within 5% of certified values. For samples in 2019, one-eighth of polypropylene filter was cut it into small pieces with ceramic scissors and put in the microwave digestion tube. Then add 10ml mixed digestion solution of nitric acid, hydrochloric acid and hydrogen peroxide with ratio of 1:3:1 and cover and screw the tube for digestion. The temperature was successively raised to 120 °C, 150 °C, 180 °C, and 200 °C and kept for 8, 8, 8, and 15 min, respectively. After the cooling of the tube, add the ultrapure water to fix the volume to 25 mL, and then use the 0.22 µm microporous filter to take the supernatant for subsequent instrumental analysis. For QA/QC, standard reference materials were pretreated and analyzed with the same procedure, with the recovered values for all the target elements falling into the range

or within 5% of certified values.

Water-soluble inorganic ions: For samples in 2011-2012, one-fourth of quartz membrane sample is soaked in 10 ml deionized water. Then place the well-shaken shake mixture under an ultrasonic bath for 20 minutes, stand for a while, and filter the solution with 0.45 µm of filter head for analysis. The extraction procedure was conducted for at least three times so that the water-soluble ions of samples were extracted adequately into the solution. Before detection of ions, standard solutions were prepared and were detected for over three times; and low relative standard deviations were observed. For samples in 2016, one-fourth of the quartz filter membrane was cut up and put into a 20 mL glass tube with 15 mL ultrapure water. After 30 min of ultrasonic extraction, the solution was filtered with a disposable filter head (0.45 µm) for subsequent instrumental analysis. Take two blank filters of the same batch number, and operate at the same time according to the same steps of sample treatment. The blank solution was prepared and analyzed by chromatography. Before the ion detection, standard solutions were detected for over three times and low relative standard deviations were obtained. Analytical quantification was conducted by using calibration curves made from standard solutions prepared. The concentration of the blank sample was less than the detection limit. For samples in 2019, one-eighth of quartz filter membrane was cut up into the sample bottle. The glass tube was cleaned for three times using the ultrasonic cleaning equipment and then dried by baking oven. Next, the sample was put into the prepared glass tube and soaked by 8 mL ultrapure water. After 20 min of ultrasonic extraction, the glass tube was stored in refrigerator for 24 h. Then the intermediate clarified solution was filtered with two disposable filter heads of 0.2 µm for subsequent instrumental analysis. Before the ion detection, standard solutions were detected for over three times and low relative standard deviations were obtained. Analytical quantification was conducted by using peak retention time and peak area made from standard solutions prepared.

OC and **EC** analysis: For samples in 2011-2012, OC and EC were determined according to the IMPROVE protocol, and the more details were given in Han et al.

(2009). For samples in 2016 and 2019, OC and EC were determined based on the IMPROVE A protocol. The more details were following: a circular quartz filter was heated stepwise to temperatures of 140 °C, 280 °C, 480 °C, and 580 °C in a non-oxidizing helium (He) oven to analyze OC1, OC2, OC3, and OC4, respectively. Then, the oven was added to an oxidizing atmosphere of 2% oxygen (O2) and 98% He, and the quartz filter membrane was gradually heated to 580 °C, 740 °C, and 840 °C to analyze EC1, EC2, and EC3, respectively. The POC is defined as the carbon combusted after the initial introduction of oxygen and before the laser reflectance signal achieves its original value and the POC is specified as the fraction of OC. According to the IMPROVE A protocol, OC is defined as CC1+EC2+EC3-POC. For QA/QC, we carried out the measurements with the field blank filter membranes, standard sucrose solutions, and repeated analyses in the study. In order to ensure the precision of instrument, a replicate sample was analyzed for every ten samples, and the standard deviation < ± 5% was accepted.

Text S2 Input data treatment and uncertainty estimation in PMF analysis.

The data of all sites were chronologically ordered end to end for each PMF analysis. In this study, altogether 20 components were resolved, including 2 carbon compositions (i.e. OC and EC), 4 water-soluble inorganic ions (i.e. NO₃-, SO₄²⁻, NH₄+, and Cl⁻), and 14 elements (i.e. Na, Mg, Al, Si, K, Ca, Ti, V, Mn, Fe, Ni, Cu, Zn, and Pb). The optimal result of PMF is to adjust estimated data uncertainties (μ_{ij}) to minimize the value of objective function *Q* which is expressed as Eq. (1). In this work, the missing values were replaced by median concentration of a given component with an uncertainty of four times of the median. For data below method detection limit (MDL), its uncertainty was set by 5/6 MDL. While for data above MDL, its corresponding uncertainty was calculated via Eq. (2). In addition, PM_{2.5} was also included as the fitting species, which was classified as the total variable with uncertainties set as 4 times of the concentration to reduce the impact of PM_{2.5} values on the modelling. Moreover, signal-to-noise ratio (S/N) was applied to address weak

and bad species when running PMF (Paatero and Hopke, 2003). According to US EPA PMF v5.0 fundamentals and user guide, species are categorized as "Bad" if the S/N ratio is less than 0.5 and "Weak" if the S/N ratio is greater than 0.5 but less than 1.

$$Q = \sum_{i=1}^{n} \sum_{j=1}^{m} \left[\frac{x_{ij} - \sum_{k=1}^{p} g_{ik} f_{kj}}{\mu_{ij}} \right]^{2}$$
 (1)

$$Uncertainty = \sqrt{(Error\ Fraction \times concentration)^2 + (0.5 \times MDL)^2}$$
 (2)

Text S3 Principle of potential source contribution function (PSCF) analysis

Air parcel trajectories for 72 h were calculated backward from Qingdao (36.10° N, 120.32°E) at starting times of 0:00, 06:00, 12:00, 18:00 LT for each day. Based on the trajectory analysis, PSCF model was applied to identify the potential source areas. The study region was divided into $i \times j$ small equal grid cells. Thus, the PSCF value of each cell was defined as:

$$PSCF = \frac{m_{ij}}{n_{ij}} \tag{3}$$

where i and j were the latitude and longitude indices, n_{ij} was the number of endpoints falling in the ij cell, and m_{ij} was the number of endpoints beyond the threshold criterion (i.e., the mean contribution concentration of each source-category in 2011-2012, 2016, and 2019, respectively, Table S15) in the same cell.

The PSCF value was usually weighted to obtain more reasonable results. When n_{ij} is smaller than three times the grid average number of trajectory endpoint (n_{ave}), a weighting function $W(n_{ij})$ was used to reduce uncertainty in cells (Dimitriou et al., 2015). The weighting function was defined by:

$$WPSCF_{ij} = \frac{m_{ij}}{n_{ij}}W(n_{ij}) \tag{4}$$

$$W(n_{ij}) = \begin{cases} 1.0, & 3n_{ave} < n_{ij} \\ 0.7, & 1.5n_{ave} < n_{ij} \le 3n_{ave} \\ 0.4, & n_{ave} < n_{ij} \le 1.5n_{ave} \\ 0.2, & n_{ij} \le n_{ave} \end{cases}$$
(5)

The studying field ranged from 25° N to 65° N, and 90° E to 140° E, thus the covered region was divided into 2000 grid cells of 1.0°×1.0°. The endpoints of

backward trajectories separately were 11972, 16352, and 17520 in each sampling year. Accordingly, there were 6, 8, and 9 trajectory endpoints in per cell, i.e., $n_{ave} = 6$, $n_{ave} = 8$, $n_{ave} = 9$, in every sampling periods.

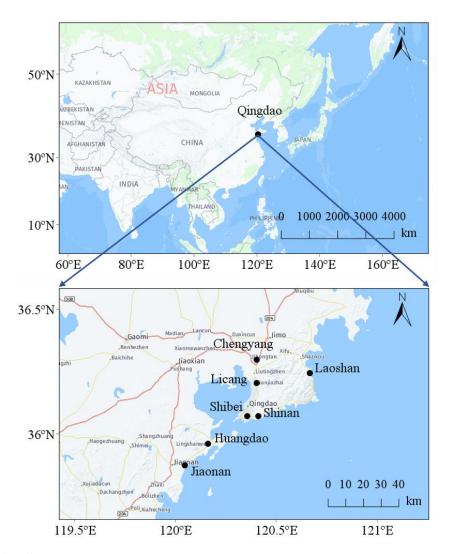


Fig. S1. Map of the study area and sampling sites (from Yahoo! Maps).

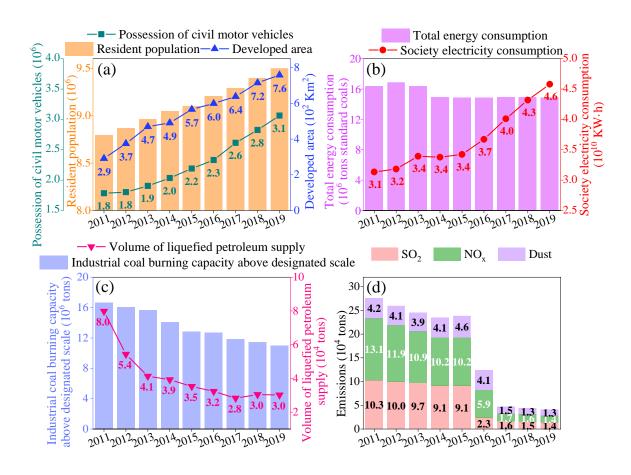


Fig. S2. Variation of local economic and social development from 2011 to 2019 in Qingdao. Data were derived from Qingdao Statistical Yearbook (http://qdtj.qingdao.gov.cn/n28356045/n32561056/n32561073/index.html, last access: 27 October, 2021).

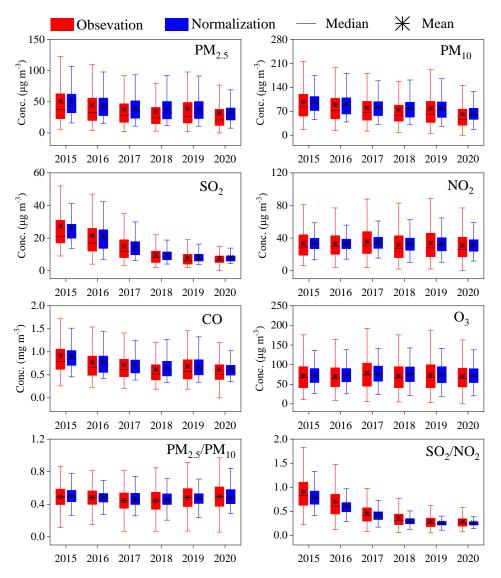


Fig. S3. Annual observed and normalized pollutant concentrations and $PM_{2.5}/PM_{10}$ and SO_2/NO_2 during different years.

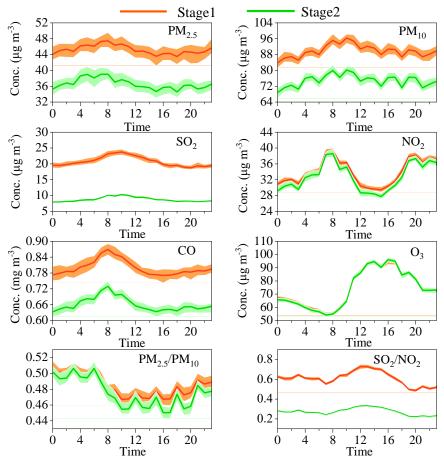


Fig. S4. Diurnal variation of air pollutant concentrations and $PM_{2.5}/PM_{10}$ and SO_2/NO_2 during the stages based on the weather normalized data. The line chart represents the mean value, while the shaded part illustrates its 95% confidence interval.

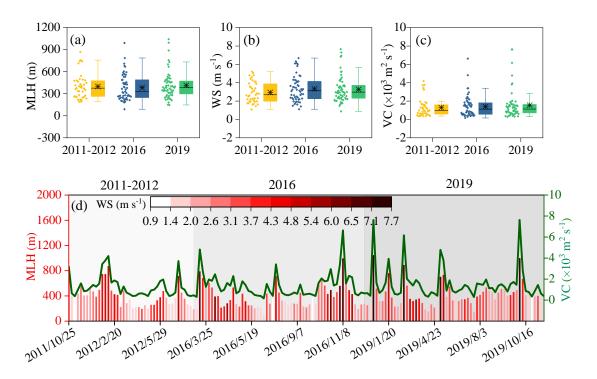


Fig. S5. The daily average (a) mixing layer height (MLH), (b) wind speed (WS), and (c) ventilation coefficient (VC), and (d) their time series during sampling periods.

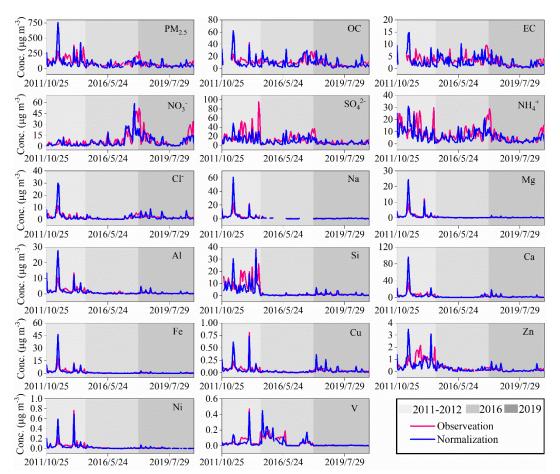


Fig. S6. Time series of concentrations of the observed and dispersion normalized chemical components during sampling periods.

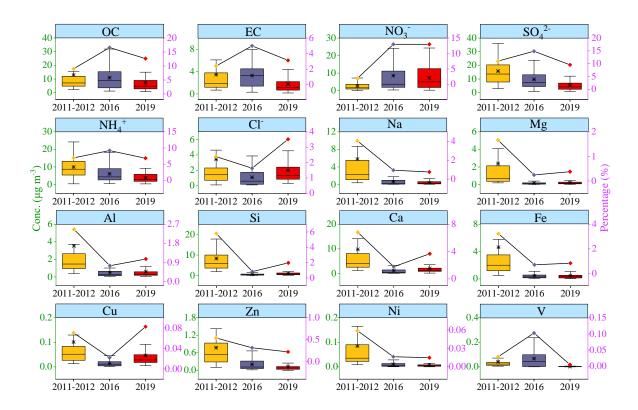


Fig. S7. The average concentrations and percentages of dispersion normalized chemical compositions in 2011-2012, 2016, and 2019. Box charts represent concentrations, and line charts represent percentages.

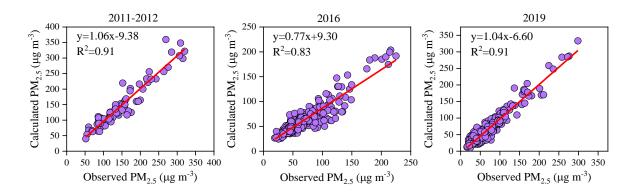


Fig. S8. The correlations between observed and calculated $PM_{2.5}$ concentrations from PMF in 2011-2012, 2016, and 2019.

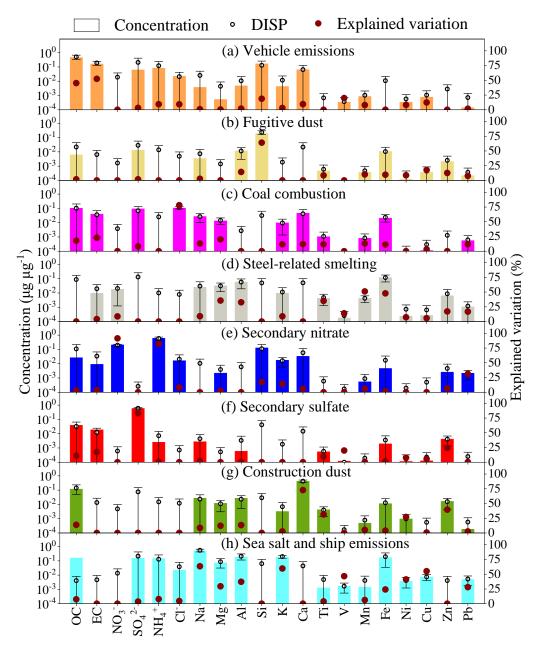


Fig. S9. Factor profiles resolved from PMF analysis from 2011 to 2012. The top and bottom of the error bars and the hollow points represent the maximum, minimum, and average DISP values, respectively.

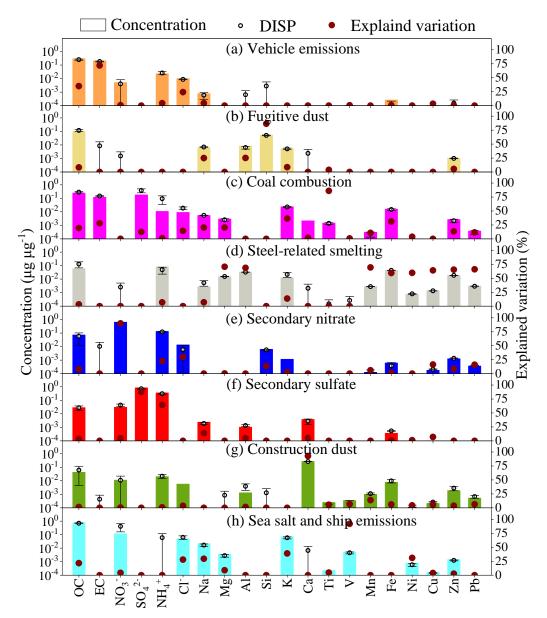


Fig. S10. Factor profiles resolved from PMF analysis in 2016. The top and bottom of the error bars and the hollow points represent the maximum, minimum, and average DISP values, respectively.

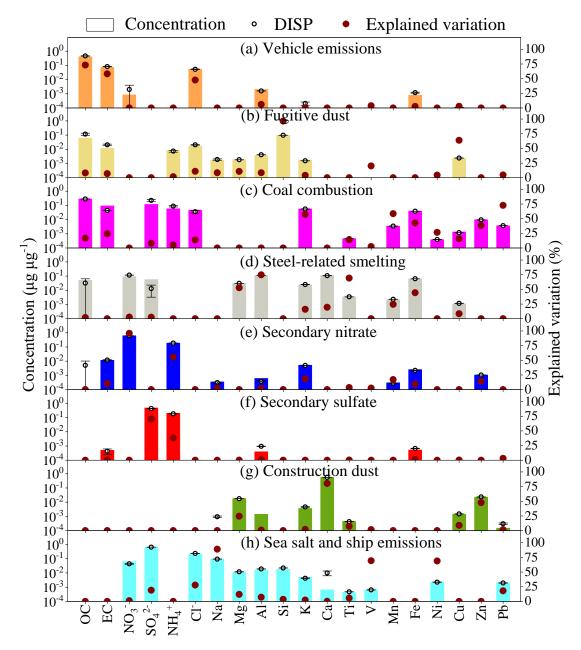


Fig. S11. Factor profiles resolved from PMF analysis in 2019. The top and bottom of the error bars and the hollow points represent the maximum, minimum, and average DISP values, respectively.

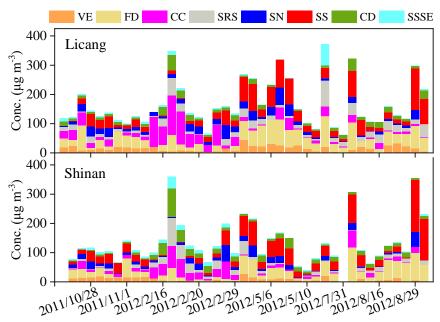
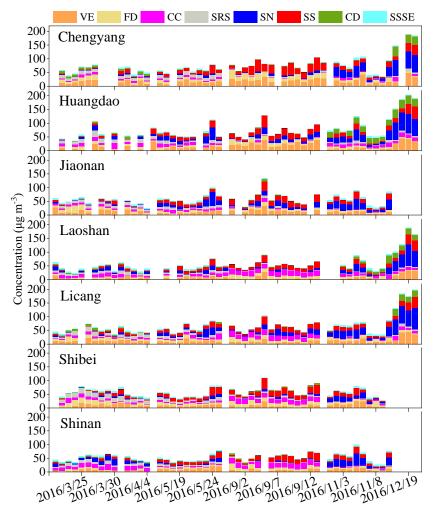
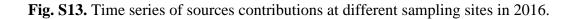


Fig. S12. Time series of sources contributions at different sampling sites in 2011-2012. VE represents vehicle emissions, FD represents fugitive dust, CC represents coal combustion, SRS represents steel-related smelting, SN represents secondary nitrate, SS represent secondary sulfate, CD represents construction dust, and SSSE represents sea salt and ship emissions.





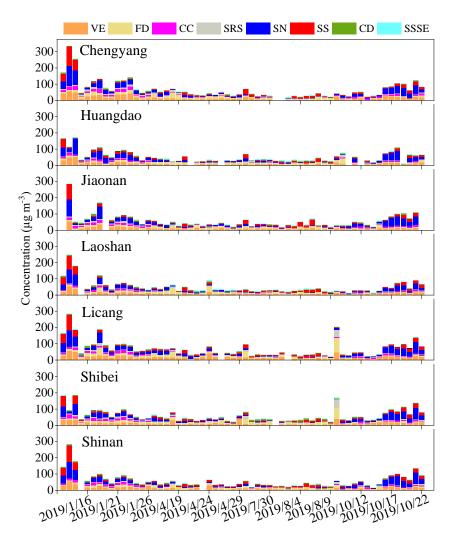


Fig. S14. Time series of sources contributions at different sampling sites in 2019.

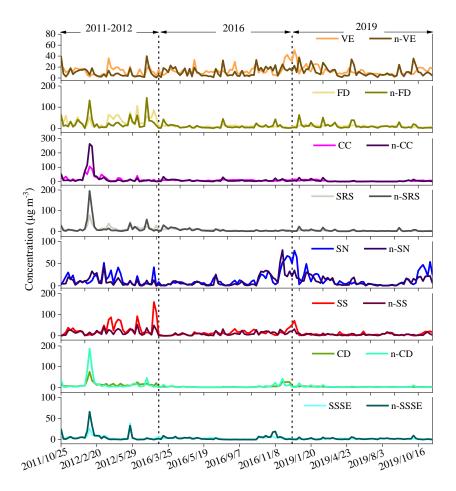


Fig. S15. Time series of source contributions before and after dispersion normalization. n- represents the results after dispersion normalization.

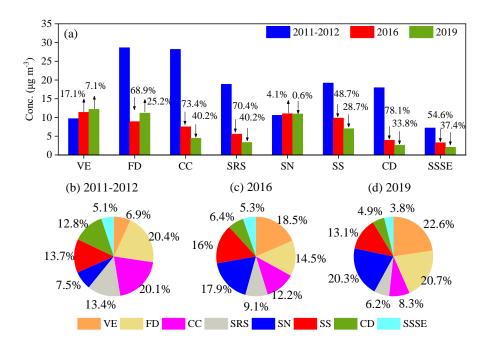


Fig. S16. The changes in the dispersion normalized source contributions in 2011-2012, 2016, and 2019.

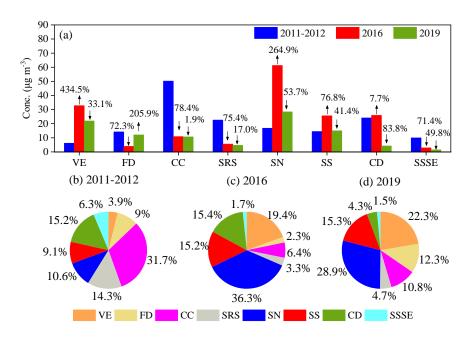


Fig. S17. The change in the source contributions during heating seasons of 2011-2012, 2016, and 2019.

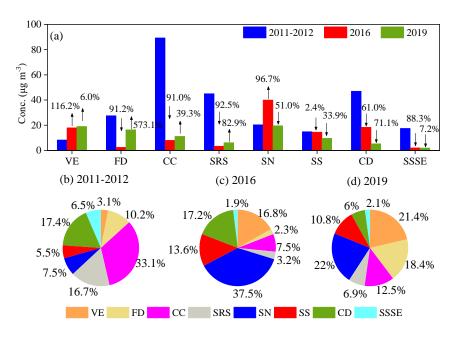


Fig. S18. The changes in the dispersion normalized source contributions during heating seasons of 2011-2012, 2016, and 2019.

 $\label{eq:Table S1.} \textbf{Table S1.} \ \textbf{The descriptions of the sampling sites.}$

| Site | Longitude | Latitude | Functional Zone |
|-----------|------------|-----------|-------------------------------------|
| Shibei | 120°20′45″ | 36°04′05″ | Traffic and landscape area |
| Shinan | 120°24′46″ | 36°03′57″ | Traffic and residential area |
| Licang | 120°24′10″ | 36°12′15″ | Residential area |
| Laoshan | 120°39′55″ | 36°14′25″ | Administrative and educational area |
| Chengyang | 120°23′59″ | 36°17′47″ | Administrative and traffic area |
| Huangdao | 120°09′37″ | 35°57′40″ | Educational area |
| Jiaonan | 120°02'47" | 35°52'20" | Administrative area |

Table S2. Summary on the sampling periods for different years in Qingdao.

| | 7 1 61 7 | \ \ \ | | |
|------|--|-------|---------|-------|
| Year | Date | Days | Samples | Total |
| 2011 | 25 October to 3 November | 10 | 19 | |
| | 15 to 22, 28 to 29 February | 10 | 20 | 80 |
| 2012 | 3 to 11, 28 to 29 May | 11 | 22 | 80 |
| | 31 July; 6 to 8, 16, 22 to 24, 29 to 30 August | 10 | 19 | |
| | 21 March to 5 April | 16 | 89 | _ |
| 2016 | 16 to 26 May | 11 | 67 | 222 |
| 2016 | 31 August to 13 September; 20 September | 15 | 95 | 332 |
| | 1 to 9, 14 November; 17 to 20 December | 14 | 81 | |
| | 12 to 26 January | 15 | 107 | |
| 2010 | 15 to 29 April | 15 | 101 | 414 |
| 2019 | 26 July to 9 August | 15 | 104 | 414 |
| | 8 to 22 October | 15 | 102 | |

Table S3. Details of sampling instruments and filters during different sampling years.

| Year | Instrument | Model | Corporation | Country | Flow rate (L min ⁻¹) | Filter diameter (mm) | Filter category | Corporation | Country |
|---------------|--|----------|--|---------|----------------------------------|----------------------|---------------------------|--|---------------|
| 2011- 2012 | Four channel air particulate matter sampler | TH-16A | Wuhan Tianhong Instrument Co., Ltd | China | 16.7 | 47 | Polypropylene / Quartz | Beijing Synthetic Fiber Research Institute/Pall Life Sciences | China/ USA |
| 2016 | Multichannel ambient air particulate sampler | ZR-3930D | Qingdao Junray Intelligent Instrument Co., Ltd | China | 16.7 | 47 | Polypropylene / Quartz | Munktell | Sweden |
| 2019 | Multichannel ambient air particulate sampler | ZR-3930D | Qingdao Junray Intelligent Instrument Co., Ltd | China | 16.7 | 47 | Polypropylene / Quartz | Pall Life Sciences | USA |

Table S4. Details of instruments of gravimetric analysis during different sampling periods.

| Year | Instrument | Model | Corporation | Country | Resolution | Error range |
|-----------|--------------|---------|----------------|-------------|------------|-------------|
| 2011-2012 | Microbalance | MX5 | Mettler Toledo | Switzerland | 1µg | 5μg |
| 2016 | Microbalance | XS105DU | Mettler Toledo | Switzerland | 10µg | 50μg |
| 2019 | Microbalance | XPE105 | Mettler Toledo | Switzerland | 10μg | 50μg |
| | | | | | | |

Table S5. Details of instruments of chemical analysis during different sampling periods.

| Year | Composition | Instrument | Model | Corporation | Country |
|---------------|--|---|------------------|--------------------------|---------|
| 2011- 2012 | Na, Mg, Al, K, Ca, Ti, V, Mn, Fe, Ni, Cu, Zn, Si, and Pb | Inductively coupled plasma-mass spectrometer (ICP-MS) | IRIS Intrepid II | Thermo Fisher Scientific | USA |
| | NO ₃ -, SO ₄ ²⁻ , NH ₄ +, and Cl | Ion chromatograph | DX-120 | DIONEX | USA |

| | OC and EC | Thermal/optical carbon analyzer | DRI 2001A | Desert Research Institute | USA |
|------|---|---|------------------------------|----------------------------------|-------------|
| | Na, Mg, Al, K, Ca, Ti, V, Mn, Fe, Ni, Cu, Zn, Si, and Pb | Inductively coupled plasma-mass spectrometer (ICP-MS) | 7500A | Agilent Technologies Co., Ltd | USA |
| 2016 | NO ₃ -, SO ₄ ² -, NH ₄ +, and Cl ⁻ Ion chromatograph | | 940 Professional IC Vario | Metrohm | Switzerland |
| | OC and EC | Thermal/optical carbon analyzer | DRI 2001A | Desert Research Institute | USA |
| | Na, Mg, Al, K, Ca, Ti, V, Mn, Fe, Ni, Cu, Zn, and Pb and Si | | | Thermo Fisher Scientific | USA |
| 2019 | NO ₃ -, SO ₄ ²⁻ , NH ₄ +, and Cl ⁻ | Ion chromatograph | Thermo ICS-900 | Thermo Fisher Scientific | USA |
| | OC and EC | Thermal/optical carbon analyzer | DRI 2001A | Desert Research Institute | USA |

Table S6. The method detection limits (MDLs) of chemical compositions for PMF calculation.

| Components | MDLs ($\mu g m^{-3}$) | Components | MDLs (µg m ⁻³) |
|-----------------------------|-------------------------|------------|----------------------------|
| OC | 0.2985 | Ca | 0.3775 |
| EC | 0.0728 | Ti | 0.0003 |
| NO_3^- | 0.0371 | V | 0.0012 |
| $\mathrm{SO_4}^{2\text{-}}$ | 0.0458 | Cr | 0.0010 |
| $\mathrm{NH_4}^+$ | 0.0197 | Mn | 0.0006 |
| Cl- | 0.0116 | Fe | 0.0603 |
| Na | 0.0304 | Ni | 0.0048 |
| Mg | 0.0183 | Cu | 0.0106 |
| Al | 0.0690 | Zn | 0.0087 |
| Si | 0.0441 | Pb | 0.0064 |
| K | 0.0277 | | |

Table S7. Statistical summary (μg m⁻³·yr⁻¹) of air pollutant concentrations from Theil–Sen trend analysis after weather normalization by random forest modelling.

| | J | · · · · · · · · · · · · · · · · · · · | |
|-----------------------|--------|---------------------------------------|----------------|
| Pollutants | Median | Percentile 5% | Percentile 95% |
| PM _{2.5} | -2.8 | -3.6 | -2.2 |
| PM_{10} | -5.4 | -6.4 | -4.3 |
| SO_2 | -3.4 | -4.3 | -2.8 |
| NO_2 | -0.3 | -0.6 | -0.1 |
| CO | -42.8 | -55.1 | -29.4 |
| O_3 | -0.1 | -0.5 | +0.3 |
| | | | |

Table S8. The annual mean concentrations for air pollutants and $PM_{2.5}/PM_{10}$ and SO_2/NO_2 during two pollution-control stages based on the observed and weather normalized data.

| Data | Periods | $PM_{2.5}$ | PM_{10} | SO_2 | NO_2 | CO | O_3 | $PM_{2.5}/PM_{10}$ | SO ₂ /NO ₂ |
|-----------------|---------|------------|-----------|--------|--------|------|-------|--------------------|----------------------------------|
| Normalized data | Stage1 | 45 | 90 | 21 | 34 | 0.80 | 74 | 0.49 | 0.61 |
| | Stage2 | 37 | 75 | 9 | 33 | 0.66 | 74 | 0.48 | 0.27 |
| Observed data | Stage1 | 45 | 89 | 21 | 34 | 0.80 | 73 | 0.47 | 0.69 |
| | Stage2 | 35 | 71 | 8 | 32 | 0.64 | 71 | 0.47 | 0.32 |

Unit for CO is mg m^{-3} , other pollutant is μg m^{-3} . Stage 1: the Air Pollution Prevention and Control Action Plan Period (2015-2017); Stage 2: the Blue-Sky Defense War Period (2018-2020).

Table S9. The observed pollutant concentrations and PM_{2.5}/PM₁₀ and SO₂/NO₂ during different lockdown stages in 2020 and the corresponding periods (*) in 2019 and 2018.

| Year | Period | $PM_{2.5}$ | PM_{10} | SO_2 | NO_2 | CO | O_3 | $PM_{2.5}/PM_{10}$ | SO_2/NO_2 |
|------|-------------------|------------|-----------|--------|--------|-----|-------|--------------------|-------------|
| | Pre-lockdown* | 72.6 | 128.9 | 17.9 | 54.8 | 1.1 | 38.5 | 0.5 | 0.4 |
| 2018 | Full lockdown* | 38.8 | 74.8 | 15.1 | 35.8 | 0.7 | 62.6 | 0.5 | 0.5 |
| | Partial lockdown* | 51.3 | 92.9 | 11.8 | 37.7 | 0.7 | 85.0 | 0.5 | 0.3 |
| | Pre-lockdown* | 90.4 | 147.4 | 17.2 | 62.6 | 1.3 | 25.7 | 0.6 | 0.3 |
| 2019 | Full lockdown* | 69.4 | 115.7 | 12.6 | 40.6 | 1.0 | 54.0 | 0.6 | 0.4 |
| | Partial lockdown* | 31.9 | 71.8 | 8.5 | 32.5 | 0.6 | 73.3 | 0.4 | 0.3 |
| 2020 | Pre-lockdown | 83.8 | 118.8 | 13.6 | 51.7 | 1.2 | 37.4 | 0.7 | 0.3 |
| | Full lockdown | 35.6 | 51.1 | 6.8 | 21.6 | 0.6 | 64.8 | 0.7 | 0.4 |
| | Partial lockdown | 26.6 | 70.8 | 7.0 | 29.7 | 0.5 | 73.0 | 0.4 | 0.3 |

Units: $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , O_3 of μg m⁻³, CO of mg m⁻³, and $PM_{2.5}/PM_{10}$ and SO_2/NO_2 of unitless. Prelockdown: 1 to 24 January, 2020; Full lockdown: 25 January to 7 March, 2020; and Partial lockdown: 8 to 31 March, 2020.

Table S10. The weather normalized pollutant concentrations and $PM_{2.5}/PM_{10}$ and SO_2/NO_2 during different lockdown stages in 2020 and the corresponding periods (*) in 2019 and 2018.

| Year | Period | $PM_{2.5}$ | PM_{10} | SO_2 | NO_2 | CO | O_3 | $PM_{2.5}/PM_{10}$ | SO_2/NO_2 |
|------|-------------------|------------|-----------|--------|--------|-----|-------|--------------------|-------------|
| 2018 | Pre-lockdown* | 76.8 | 131.3 | 18.6 | 53.9 | 1.2 | 44.7 | 0.6 | 0.4 |
| | Full lockdown* | 52.1 | 90.9 | 14.7 | 35.1 | 0.8 | 67.3 | 0.6 | 0.4 |
| | Partial lockdown* | 48.0 | 91.2 | 11.7 | 36.3 | 0.7 | 83.2 | 0.5 | 0.3 |
| | Pre-lockdown* | 80.3 | 137.3 | 16.7 | 56.2 | 1.2 | 39.7 | 0.6 | 0.3 |
| 2019 | Full lockdown* | 61.1 | 109.2 | 12.2 | 37.2 | 0.9 | 63.6 | 0.6 | 0.3 |
| | Partial lockdown* | 36.1 | 78.1 | 8.6 | 33.2 | 0.6 | 79.5 | 0.5 | 0.3 |
| 2020 | Pre-lockdown | 76.6 | 119.4 | 14.6 | 49.1 | 1.1 | 43.8 | 0.6 | 0.3 |
| | Full lockdown | 40.0 | 60.3 | 8.9 | 25.1 | 0.7 | 66.1 | 0.7 | 0.4 |
| | Partial lockdown | 30.7 | 72.5 | 7.5 | 30.3 | 0.6 | 78.9 | 0.4 | 0.2 |

Units: $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , O_3 of μg m⁻³, CO of mg m⁻³, and $PM_{2.5}/PM_{10}$ and SO_2/NO_2 of unitless.

Table S11. The change rates (%) of weather normalized pollutant concentrations and PM_{2.5}/PM₁₀ and SO₂/NO₂ during different stages in 2020 and corresponding periods (*) in 2019 and 2018.

| Year | Comparison | PM _{2.5} | PM ₁₀ | SO_2 | NO ₂ | CO | O ₃ | PM _{2.5} /PM ₁₀ | SO ₂ /NO ₂ |
|------|---------------------------|-------------------|------------------|--------|-----------------|-------|----------------|-------------------------------------|----------------------------------|
| 2018 | Full VS Pre-lockdown* | -32.1 | -30.7 | -20.5 | -34.8 | -27.4 | +50.6 | -1.8 | +19.8 |
| | Partial VS Full-lockdown* | -8.0 | +0.3 | -20.9 | +3.4 | -18.4 | +23.7 | -8.8 | -23.0 |
| 2019 | Full VS Pre-lockdown* | -23.9 | -20.5 | -27.4 | -33.9 | -24.9 | +60.0 | -4.0 | +10.6 |
| | Partial VS Full-lockdown* | -41.0 | -28.5 | -29.4 | -10.6 | -30.9 | +25.1 | -17.2 | -21.9 |
| 2020 | Full VS Pre-lockdown | -47.8 | -49.5 | -39.0 | -49.0 | -37.0 | +50.8 | +3.1 | +19.9 |
| | Partial VS Full-lockdown | -23.4 | +20.3 | -15.4 | +21.1 | -21.3 | +19.5 | -36.0 | -30.7 |

Table S12. The change rates (%) of weather normalized pollutant concentrations and PM_{2.5}/PM₁₀ and SO₂/NO₂ in different lockdown stages in 2020 corresponding to the same periods of other different years.

| Period | Comparison | PM _{2.5} | PM ₁₀ | SO ₂ | NO ₂ | CO | O ₃ | PM _{2.5} /PM ₁₀ | SO ₂ /NO ₂ |
|------------------|--------------|-------------------|------------------|-----------------|-----------------|-------|----------------|-------------------------------------|----------------------------------|
| Due le alcderum | 2020 VS 2018 | -0.2 | -9.1 | -21.6 | -8.8 | -2.7 | -1.9 | +10.1 | -14.3 |
| Pre-lockdown | 2020 VS 2019 | -4.6 | -13.1 | -13.0 | -12.6 | -7.3 | 10.3 | +10.8 | -0.4 |
| Evil lookdores | 2020 VS 2018 | -23.2 | -33.7 | -39.8 | -28.7 | -15.5 | -1.8 | +15.6 | -14.2 |
| Full lockdown | 2020 VS 2019 | -34.5 | -44.8 | -27.0 | -32.6 | -22.3 | +3.9 | +19.0 | +8.0 |
| Partial lockdown | 2020 VS 2018 | -36.1 | -20.5 | -35.6 | -16.5 | -18.5 | -5.2 | -18.9 | -22.8 |
| Partiai iockdown | 2020 VS 2019 | -15.0 | -7.1 | -12.5 | -8.7 | -11.5 | -0.7 | -8.0 | -4.2 |

Table S13. The change rates (%) of major chemical compositions in comparisons of different sampling periods after dispersion normalization.

| Compositions | 2016 VS 2011-2012 | 2019 VS 2016 | 2019 VS 2011-2012 |
|----------------------------|-------------------|--------------|-------------------|
| OC | -15.3 | -34.4 | -44.5 |
| EC | -4.9 | -46.1 | -48.7 |
| NO_3^- | +213.1 | -14.1 | +169.0 |
| $\mathrm{SO_4}^{2	ext{-}}$ | -38.1 | -44.8 | -65.8 |
| $\mathrm{NH_4}^+$ | -39.2 | -36.1 | -61.1 |
| Cl- | -68.7 | +85.6 | -41.9 |
| Na | -89.3 | -31.6 | -92.6 |
| Mg | -93.2 | +32.5 | -91.0 |
| Al | -86.2 | +23.0 | -83.0 |
| Si | -93.3 | +101.3 | -86.5 |
| Ca | -88.0 | +77.1 | -78.8 |
| Fe | -90.0 | +1.8 | -89.8 |
| Cu | -85.6 | +221.3 | -53.8 |
| Zn | -73.2 | -39.3 | -83.7 |
| Ni | -87.1 | -22.0 | -90.0 |
| V | +61.5 | -96.0 | -93.5 |

Table S14. Summary of error estimation diagnostics with eight-factor solution from the PMF during different periods.

| Diagnose | 2011-2012 | 2016 | 2019 | | | |
|------------------------|----------------------------------|--------|--------|--|--|--|
| Base run number | 10 | 10 | 10 | | | |
| Qexp | 792 | 3816 | 4800 | | | |
| Qtrue | 812 | 4094 | 5124 | | | |
| Q_{true}/Q_{exp} | 1.03 | 1.07 | 1.07 | | | |
| DISP %dQ | < 0.1% | < 0.1% | < 0.1% | | | |
| DISP swaps | 0 | 0 | 0 | | | |
| BS run number | 100 | 100 | 100 | | | |
| Easter with DC manning | CC factor 92% | | | | | |
| Factor with BS mapping | CD factor 90% SSSE factor 87% | | | | | |
| <100% | | | | | | |

Table S15. The threshold concentrations involved in PSCF analysis

| | | | 5 |
|-----------------------------|-----------|------|------|
| Sources | 2011-2012 | 2016 | 2019 |
| Vehicle emissions | 12.1 | 13.5 | 13.5 |
| Fugitive dust | 35.0 | 8.4 | 10.1 |
| Coal combustion | 21.1 | 8.1 | 4.4 |
| Steel-related smelting | 15.9 | 5.9 | 2.9 |
| Secondary nitrate | 14.2 | 12.6 | 15.1 |
| Secondary sulfate | 33.5 | 12.8 | 9.8 |
| Construction dust | 14.2 | 3.8 | 2.4 |
| Sea salt and ship emissions | 5.7 | 3.0 | 2.0 |

Unit: µg m⁻³

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