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Kinetics, SOA yields, and chemical composition of secondary organic aerosol from β -caryophyllene ozonolysis with and without nitrogen oxides between 213 and 313 K

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Abstract. β -caryophyllene (BCP) is one of the most important sesquiterpenes (SQTs) in the atmosphere, with a large potential contribution to secondary organic aerosol (SOA) formation mainly from reactions with ozone (O_3) and nitrate radicals (NO₃). In this work, we study the temperature dependence of the kinetics of BCP ozonolysis, SOA yields, and SOA chemical composition in the dark and in the absence and presence of nitrogen oxides including nitrate radicals (NO₃). We cover a temperature range of 213–313 K, representative of tropospheric conditions. The oxidized components in both gas and particle phases were characterized on a molecular level by a chemical ionization mass spectrometer equipped with a filter inlet for gases and aerosols using iodide as the reagent ion (FIGAERO-iodide-CIMS). The batch mode experiments were conducted in the 84.5 m³ aluminium simulation chamber AIDA at the Karlsruhe Institute of Technology (KIT). In the absence of nitrogen oxides, the temperature-dependent rate coefficient of the endocyclic double bond in BCP reacting with ozone between 243-313 K is negatively correlated with temperature, corresponding to the following Arrhenius equation: $k = (1.6 \pm 0.4) \times 10^{-15} \times \exp((559 \pm 97)/T)$. The SOA yields increase from $16 \pm 5\%$ to $37 \pm 11\%$, with temperatures decreasing from 313 to 243 K at a total organic particle mass of $10 \,\mu g \,m^{-3}$. The variation in the ozonolysis temperature leads to a substantial impact on the abundance of individual organic molecules. In the absence of nitrogen oxides, monomers $C_{14-15}H_{22-24}O_{3-7}$ (37.4%), dimers $C_{28-30}H_{44-48}O_{5-9}$ (53.7%), and trimers $C_{41-44}H_{62-66}O_{9-11}$ (8.6 %) are abundant in the particle phase at 213 K. At 313 K, we observed more oxidized monomers (mainly $C_{14-15}H_{22-24}O_{6-9}$, 67.5 %) and dimers (mainly $C_{27-29}H_{42-44}O_{9-11}$, 27.6 %), including highly oxidized molecules (HOMs; $C_{14}H_{22}O_{7,9}$, $C_{15}H_{22}O_{7,9}C_{15}H_{24}O_{7,9}$), which can be formed via hydrogen shift mechanisms, but no significant trimers. In the presence of nitrogen oxides, the organonitrate fraction increased from 3 % at 213 K to 12 % and 49 % at 243 and 313 K, respectively. Most of the organonitrates were monomers with C15 skeletons and only one nitrate group. More highly oxygenated organonitrates were observed at higher temperatures, with their signal-weighted O: C atomic ratio increasing from 0.41 to 0.51 from 213 to 313 K. New dimeric and trimeric organic species without nitrogen atoms (C_{20}, C_{35}) were formed in the presence of nitrogen oxides at 298-313 K, indicating potential new reaction pathways. Overall, our results show that increasing temperatures lead to a relatively small decrease in the rate coefficient of the endocyclic double

bond in BCP reacting with ozone but to a strong decrease in SOA yields. In contrast, the formation of HOMs and organonitrates increases significantly with temperature.

1 Introduction

Biogenic volatile organic compounds (BVOCs), which are emitted mainly from plants, are the largest source of precursors of atmospheric secondary organic aerosol (SOA), which has profound impacts on visibility, air quality, human health, clouds, and climate change (Fehsenfeld et al., 1992; Mellouki et al., 2015; Laothawornkitkul et al., 2009; Charnawskas et al., 2017).

Sesquiterpenes (SQTs; $C_{15}H_{24}$) are a class of BVOCs mainly emitted from coniferous trees (Kleist et al., 2012; Matsunaga et al., 2013), Scots pines (Kivimäenpää et al., 2020; Weikl et al., 2016), and deciduous trees (Li et al., 2019). Despite their lower atmospheric emission concentrations compared to isoprene and monoterpenes (Faiola et al., 2018), SQTs are important in the atmosphere because of their high reactivities with ozone and large aerosol formation potentials (Ciccioli et al., 1999; Lee et al., 2006a, b; Ng et al., 2007). Depending on region and season they can even play a dominating role (Geron and Arnts, 2010; Tarvainen et al., 2005; Jardine et al., 2011; Shrivastava et al., 2019).

Among all the sesquiterpene species, β -caryophyllene (BCP) is one of the most abundant and is emitted, for example, from various pine trees, but also from plants used for agricultural purposes (Duhl et al., 2008), like foliage of orange trees (Ciccioli et al., 1999; Hansen and Seufert, 2003), potato plants (Agelopoulos et al., 2000), and cotton (Rodriguez-Saona et al., 2001). BCP is a bicyclic compound with two double bonds with different reactivities towards ozone. The rate coefficient for ozone reacting with the endocyclic double bond at 296 K was determined as $(1.2 \pm 0.4) \times 10^{-14}$ cm³ molec.⁻¹ s⁻¹ (Shu and Atkinson, 1994; Richters et al., 2015) (Cox et al., 2020). It is estimated to be about 100 times larger than the rate coefficient of $(1.1 \pm 0.4) \times 10^{-16}$ cm³ molec.⁻¹ s⁻¹ for the exocyclic double bond (Shu and Atkinson, 1994) of the first-generation products. Therefore, the atmospheric lifetime of the endocyclic double bond towards ozone is estimated to be about 2 min, and the first-generation products can exist about 3.5 h in typical tropospheric ozone levels of 30 ppb (Winterhalter et al., 2009).

The SOA mass yield of BCP ozonolysis is expected to be dependent on organic particle mass concentration (Odum et al., 1996), temperature (Saathoff et al., 2009) and ozone levels (Chen et al., 2012). Previous studies give a relatively large range of SOA yields of 5 %–70 % depending on the experimental conditions (Jaoui et al., 2003; Lee et al., 2006a; Winterhalter et al., 2009; Chen et al., 2012; Tasoglou and Pandis,

2015) for a temperature range of 287–298 K. A comparison of these studies is given in the Supplement (see Table S1).

Yields, products, and mechanisms of BCP ozonolysis were investigated extensively for room temperature (Jaoui et al., 2003; Griffin et al., 1999; Lee et al., 2006a, 2006b; Winterhalter et al., 2009; Nguyen et al., 2009; Jenkin et al., 2012; Richters et al., 2016). The oxidation products mainly include aldehydes, small acids, and acetone in the gas phase (Grosjean et al., 1993; Calogirou et al., 1999; Zhao et al., 2010; Larsen et al., 1998) as well as ketones, aldehydes, alcohols, and carboxylic acids in the particle phase (Dekermenjian, 1999; Lee et al., 2006a; Jaoui et al., 2003; van Eijck et al., 2013; Jaoui et al., 2007; Li et al., 2011; Alfarra et al., 2012). Major oxidation products identified are, for example, β -caryophyllonic acid (C₁₅H₂₄O₃), β -caryophyllinic acid ($C_{14}H_{22}O_4$; BCA), and β -hydroxycaryophyllonic acid $(C_{15}H_{24}O_4)$ (Table S3). BCA has also been used as a tracer to estimate the emissions of BCP in the real atmosphere (Jaoui et al., 2007; Hu et al., 2008; Parshintsev et al., 2008; Haque et al., 2016; Verma et al., 2021; Cheng et al., 2021). In addition, highly oxidized multifunctional organic molecules (HOMs) were observed from BCP ozonolysis. Richters et al. (2016) studied their formation mechanism at 295 K in a free-jet flow system using isotopic labelling. Major HOMs identified are $C_{14-15}H_{22}O_7$, $C_{14-15}H_{22}O_9$, and $C_{15}H_{22}O_{11,13}$ (Richters et al., 2016).

SQT oxidation products are expected to have low or extremely low vapour pressures owing to their long carbon skeletons with multiple functional groups. For example, the saturation vapour pressure of β -caryophyllinic acid was estimated as low as 3.3×10^{-13} Pa (Li et al., 2011), which leads to a high condensation and SOA formation potential. As a consequence, these species are expected to play an important role in new particle formation (NPF) (Kirkby et al., 2016; Kammer et al., 2020; Huang et al., 2021).

Especially during night-time, the reaction between nitrate (NO₃) radicals and unsaturated hydrocarbons is of substantial importance and results in SOA formation. BCP has a high reactivity towards NO₃ radicals, with a rate coefficient of $(1.93 \pm 0.35) \times 10^{-11}$ cm³ molec.⁻¹ s⁻¹ (Shu and Atkinson, 1995; Winterhalter et al., 2009; Cox et al., 2020). SOA formation from BCP ozonolysis in the presence of NO₃ radicals has only been the subject of few studies so far. SOA yields from dark reactions between BCP and NO₃ were estimated to be 1.46 and 0.91 for SOA mass concentrations of 113.4 and 60.3 µg m⁻³, respectively (Jaoui et al., 2013). The formation mechanism of several organic nitrates including C₁₅H_{23,25}O₄N and C₁₅H₂₅O_{6,7}N is described in the Master Chemical Mechanism (MCM v3.2) by Jenkin et al. (2012).

The troposphere covers a wide temperature range which depends on the altitude from the planetary boundary layer (PBL) to the upper troposphere, latitude, and seasonal variations. Typically, the tropospheric temperatures range between 300 and 200 K. In summer, some regions can even reach to 318 K. BVOCs and their oxidation products are proposed to undergo transport from the PBL to the upper troposphere, e.g. via convective systems typical for tropical regions (Andreae et al., 2018). In this conceptual picture of a potential aerosol life cycle the temperature dependence of the different processes involved is quite important. Besides, multiple studies have shown that SOA yields typically increase with decreasing temperature, e.g. for α -pinene and limonene ozonolysis (Saathoff et al., 2009), β -pinene ozonolysis (von Hessberg et al., 2009), and isoprene ozonolysis (Clark et al., 2016). In addition, at lower temperatures also semi-volatile vapours alter their partitioning behaviour due to reduced vapour pressures (Stolzenburg et al., 2018; Ye et al., 2019). Despite the potentially important role of BCP oxidation products in NPF and their high condensation and SOA formation potential, studies on the temperature dependence of SOA formation from ozonolysis of BCP are still scarce.

The objectives of this work were to study the temperature dependence on the kinetics, yields, and chemical composition distribution of the aerosol produced by the reaction of BCP with O_3 in the temperature range between 213–313 K. Furthermore, we investigated the impact of the presence of nitrogen oxides including NO₃ radicals on the SOA chemical composition and volatility.

2 Methodology

2.1 Experimental conditions

A schematic of the AIDA (Aerosol Interaction and Dynamics in the Atmosphere) simulation chamber at the Karlsruhe Institute of Technology (KIT) and the main instrumentation used in this work is shown in Fig. 1. The chamber is an 84.5 m³ aluminium vessel equipped with a LED solar radiation simulator and with precisely controlled temperature, humidity, and gas mixtures. It is operated as a continuously stirred reactor with mixing times of 1-2 min achieved using a fan about 1 m above the bottom of the chamber (Saathoff et al., 2009). Wall and gas temperature inside the chamber is controlled at ± 0.3 K over a wide range of temperatures (313–183 K) (Wagner et al., 2006). The pressure inside the chamber can be varied from 0.01 to 1000 hPa, and the relative humidity (RH) can be modified between close to 0 % to up to 100 % and even to supersaturated conditions (Möhler et al., 2003). Water vapour is measured in situ by a tuneable diode laser (TDL) hygrometer with an accuracy of ± 5 % and a calibrated reference dew point mirror hygrometer (MBW373LX, MBW Calibration Ltd.) with an accuracy of ± 1 % (Fahey et al., 2014).



Figure 1. Schematic of the AIDA simulation chamber and its instrumentation for this study.

The results presented in this work are from dark BCP ozonolysis experiments with or without addition of NO₂ from a campaign in November and December 2019 covering five different temperatures between 213-313 K. The experimental conditions are listed in Table 1. Two additional experiments were undertaken in March 2020 to study the rate coefficients of BCP reacting with ozone at 243 and 258 K. BCP (98%, Carl Roth GmbH) was added to the AIDA chamber with a flow of 0.01 m³ min⁻¹ of synthetic air potentially saturated with its vapour at 298 K. Please note that the BCP concentrations for the experiments at 213 K could not be measured due to the low vapour pressure and strong wall losses at lower temperatures (more details are given in Sect. 2.2). During the experiment at 213 K the initially added BCP was lost to the walls, so almost no SOA was formed after the addition of ozone. However, when adding BCP in the presence of ozone, SOA was formed in quantities comparable to the other experiments. In all experiments, ozone was typically in excess and generated by a silent discharge generator (Semozon 030.2, Sorbios) in pure oxygen (99.9999%). The relative humidity ranged from 96% to 13% for experiments at 213 and 313 K, respectively. This corresponds to water vapour concentrations of 1 Pa $(3.4 \times 10^{14} \text{ cm}^{-3})$ at 213 K and 952 Pa $(2.2 \times 10^{17} \text{ cm}^{-3})$ at 313 K, respectively, and reflects the variability in the water vapour concentrations throughout the troposphere. At the initial phase of each experiment, BCP was depleted completely by ozonolysis, and SOA was formed. After depletion of the BCP at a lower ozone level to facilitate the kinetic study we increased the excess of ozone to accelerate the oxidation of remaining double bonds. Yields and chemical composition are determined and compared for the time period after increasing the ozone level. Then a second addition of BCP generated more SOA mass. Due to the large excess of ozone, BCP could not be measured during the subsequent additions. The corresponding conditions are marked as 1a-5a in Table 1. Subsequently, NO2 (1000 ppm of 99.5 % purity in nitrogen 99.999 %; Basi Schöberl GmbH) was added to the reaction mixture still containing an excess of ozone. Another step of SOA formation was then initialized by adding more BCP in the presence of NO₃ radicals. This series of experiments is marked as 1b–5b in Table 1. Hence, the BCP and ozone concentrations listed in Table 1 for experiments 1b-5b also include the amounts added in experiments 1a-5a. Ozone was at nearly the same concentrations as for the initial experiments without NO₂, except for the experiments at 273 K. To get a slower decay of BCP to better determine its rate coefficient at 273 K, we added 73 ppb ozone into the chamber first and then added more ozone to the same concentration level ($\sim 300 \text{ ppb}$) as used for the other temperatures for comparison. All experiments were performed in the dark, and no hydroxyl radical scavenger was used. Hence, the OH radicals generated in the ozonolysis reaction with a yield of (8 ± 3) % (Cox et al., 2020), (10.4 ± 2.3) % (Winterhalter et al., 2009), or 6 % (Shu and Atkinson, 1994) also contributed to SOA formation, as is detailed in Sect. 3.1. Time zero in the plots refers to the first addition of ozone to the reaction mixture. Please note that we tried to have similar organic particle mass concentrations among the different experiments in order to limit a potential influence of absorptive partitioning on the comparison among different temperatures. The organic aerosol mass levels were in the range of $4.5-31.5 \,\mu g \, m^{-3}$ for the initial ozonolysis and between 14 and 41 μ g m⁻³ for the reaction step with NO₃ radicals. Since we had relatively similar and sufficient organic aerosol mass levels available for partitioning, the observed differences in chemical composition should be governed by the changing chemistry, mainly in the gas phase.

2.2 Instrumentation

The main instruments used for this study are shown in Fig. 1. They were installed outside the AIDA chamber and operated at 296 K. To avoid potential artefacts, e.g. due to the phase partitioning of semi-volatile species in the sampling lines from the chamber to the instruments, the sampling lines were partially insulated, and the residence time was below 1-2 s. The concentrations of O₃ and NO₂ were measured using the gas monitors O₃41M and AS32M (both Environment S.A.), respectively.

The concentrations of BCP and lowly oxygenated gaseous oxidation products were measured by a proton-transferreaction time-of-flight mass spectrometer (PTR-ToF-MS 4000, Ionicon Analytic GmbH). Data were analysed using PTR viewer 3.3.12. The PTR-ToF-MS is also interfaced with a particle inlet (Chemical Analysis of Aerosol Online, CHARON), which allows the semi-volatile particle components to be measured. A detailed description of the CHARON-PTR-MS has been provided elsewhere (Müller et al., 2017; Piel et al., 2021). During the period of BCP in-

Experiment no.	[K]	RH	ВСР [µg m ⁻³]	Total O ₃ [ppb]	Total NO ₂ [ppb]	NO ₃ radicals	SOA mass ⁺ (SMPS) $[\mu g m^{-3}]$	Mean geometric diameter (nm)	O:C	SOA density [g cm ⁻³]	SOA yield [%]	BCA yield [%]
19	213	96	#	0.05	0	I	144+36	673+12	0.26	0 75 + 0 06	I	1
2a	243	88	$15.8 \pm 3.1 \ (15.6 \pm 3.1^{t})$	317	0	I	4.5 ± 1.1	73.3 ± 1.1	0.32	0.86 ± 0.14	28.5 ± 8	11.1 ± 7.3
3a	273	67	109.5 ± 21.9	73	0	I	31.5 ± 7.9	73.5 ± 1.9	0.38	1.0 ± 0.08	29.7 ± 8	4.1 ± 2.7
4a	298	27	$65.0 \pm 13~(23.2 \pm 4.6^{\mathrm{t}})$	325	0	I	13.6 ± 3.4	68.5 ± 1.5	0.41	1.09 ± 0.12	20.9 ± 6	$0.8 \pm 0.5 \ (4^*)$
5a	313	13	78.6 ± 15.7	290	0	I	13.7 ± 3.4	52.9 ± 3.6	0.42	$1.09\pm0.16^{\mathrm{ass}}$	17.4 ± 5	0.6 ± 0.4
6a	243	59	65.0 ± 13.0	6-505	0	I	1	I	I	I	I	I
7a	258	50	94 ± 19	13-419	0	I	Ι	I	I	I	I	I
1b	213	96	#	320	27	I	$13.9 \pm 3.5 \ (28.3 \pm 7.1^{t})$	92.7 ± 0.5	0.27	I	I	I
2b	243	88	$14.2\pm2.8^{ m t}$	317	32	0.8	$18.0 \pm 4.5~(52.5 \pm 13.1^{\rm t})$	96.0 ± 1.1	0.31	I	I	I
3b	273	67	$109.5\pm21.9^{\mathrm{t}}$	361	39	3.5	$30.8 \pm 7.7 \ (62.4 \pm 15.6^{t})$	114.9 ± 1.6	0.40	I	I	I
4b	298	27	$27.9\pm5.6^{\mathrm{t}}$	325	42	5.3	$32.1 \pm 8.0 \ (49.1 \pm 12.3^{t})$	106.8 ± 1.2	0.44	I	I	I
5b	313	13	$24.6\pm4.9^{ m t}$	290	46	5.6	$21.8 \pm 5.5 \; (34.4 \pm 8.6^{t})$	87.6 ± 1.7	0.45	I	I	I

Table 1. Compilation of experimental conditions, O : C ratios, SOA densities, and yields

jection and ozonolysis, the PTR-ToF-MS was operated for measuring gaseous volatile organic compounds (VOCs) only. After BCP was fully depleted, we switched the PTR-ToF-MS to an alternating measurement mode for detecting the lowly oxygenated organic molecules in both gas and particle phases. This alternating measurement mode included 3 min HEPA filter measurement for the particle background, 5 min CHARON particle measurement, 1 min transition for instrument equilibration, 5 min of VOC measurement, and another 1 min transition. The CHARON-PTR-MS measured the particle phase at a sampling flow of 500 SCCM via a 1/4'' silcosteel tube, while the gas phase was measured at a flow rate of 100 SCCM via a 1/16" PEEK (polyether ether ketone) tube taken from the particle measurement flow. Furthermore, a flow rate of $3.9 \,\mathrm{L\,min^{-1}}$ was added to the total flow to minimize the residence time in the sampling tube. For measuring gases, the drift tube of the PTR-MS was kept at 393 K and 2.8 mbar, leading to an electric field (E/N)of 127 Td. During alternating measurements, the drift tube was automatically optimized to 100 Td for particle measurement. BCP was calibrated using a liquid calibration unit (LCU-a, Ionicon Analytic GmbH). The PTR-ToF-MS at an E/N of 127 Td showed significant fragmentation of BCP, in agreement with previous studies (Kim et al., 2009; Kari et al., 2018). In this study, we observed that the parent ion $(m/z \ 205.20, \ C_{15}H_{25}^+)$ contributed $(29 \pm 1)\%$ to total signals of BCP-related ions including m/z 81.07, m/z 95.09, m/z 109.10, m/z 121.10, m/z 137.13, m/z 149.13, and m/z 205.20 (Fig. S1). Therefore, we scaled the concentration of m/z 205.20 (C₁₅H⁺₂₅) by a factor of 3.45 for the quantification of BCP. The total uncertainty in BCP quantification was estimated as ± 20 % by including all errors mainly related to the uncertainties in the LCU and the fragmentation pattern of BCP. In this study, we calibrated the mixture of toluene and BCP with a solvent of *n*-hexane using LCU. We obtained a similar sensitivity for toluene using the LCU and a gas cylinder with a toluene standard. Then we calculated the sensitivity of BCP relative to the sensitivity of toluene. This resulted in 36.2 ncps ppb^{-1} (ncps: normalized counts per second) for the parent ion (C₁₅H⁺₂₅, m/z 205) of BCP. The total uncertainty in the quantification of BCP was estimated to $\sim 20 \,\%$ by including the uncertainties in toluene in the gas standard $(\sim 10\%)$, the LCU calibration procedure $(\sim 15\%)$, and the fragmentation pattern of BCP ($\sim 5\%$).

Particle size distributions and number concentrations were measured by a scanning mobility particle sizer (SMPS) utilizing a differential mobility analyser (DMA; 3071, TSI Inc.) connected to a CPC (3772, TSI Inc.). Particle number concentrations were measured by two condensation particle counters (3022a and 3776, TSI Inc.). The particle number size distributions of the SMPS were corrected for the total number concentration measured by a calibrated CPC and used to calculate the SOA mass concentration by applying an effective particle density. This particle density was determined by comparing the mobility and aerodynamic size distributions measured by SMPS and AMS, calculated increasing from 0.9 ± 0.1 at 243 K to 1.1 ± 0.1 at 313 K, respectively (Saathoff et al., 2009).

A high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS; Aerodyne Inc.) was used to continuously measure the total organic particle mass with a time resolution of 30 s at a total flow rate of $0.0011 \text{ m}^3 \text{ min}^{-1}$ (with only $0.0008 \text{ m}^3 \text{ min}^{-1}$ going into the instrument). The data were analysed using the PIKA v1.60C software. For calculation of the organic particle concentration from the AMS mass spectra, a collection efficiency of 0.6 (determined by comparison with the SMPS results) and an ionization efficiency of $(1.52 \pm 0.1) \times 10^{-7}$ (calibrated with 300 nm ammonium nitrate particles) were used. In the absence of inorganic species, we calculated the mass concentration of the organonitrates (OrgNO3, i.e. organonitrates) from AMS data by assuming an average OrgNO3 molecular weight of 331 and 363 g mol⁻¹ based on the most abundant molecules (C15H25O7N and C15H25O9N) detected by FIGAERO-iodide-CIMS.

Furthermore, we used a chemical ionization mass spectrometer (CIMS) equipped with a filter inlet for gases and aerosols (FIGAERO) (Aerodyne Research Inc.) using iodide as the reagent ion to measure both gas-phase and particlephase components. The FIGAERO-iodide-CIMS has been described in detail by Lopez-Hilfiker et al. (2014). It allows very sensitive detection of oxygenated compounds, avoiding significant fragmentation. In our experiments, the gas-phase compounds were measured online by FIGAEROiodide-CIMS with a total flow rate of $0.005 \text{ m}^3 \text{ min}^{-1}$, with only $0.002 \text{ m}^3 \text{ min}^{-1}$ going into the instrument. The bypass flow of $0.003 \,\mathrm{m^3 \,min^{-1}}$ was used to reduce the residence time in the sampling line. In parallel, particles were deposited offline on prebaked Teflon filters (polytetrafluorethylene (PTFE), 1 µm, SKC Inc.) using a stainless-steel filter holder with a flow rate of $0.006 \,\mathrm{m^3 \, min^{-1}}$ for 5 min, the same offline sampling procedure as described by Huang et al. (2019). After particle deposition, the filter samples were stored at 243 K for analysis after the experiment. These filters were then heated by FIGAERO-iodide-CIMS using a flow of ultra-high-purity nitrogen as carrier gas following a thermal desorption procedure from 296 K to a maximum temperature of 473 K with a total desorption time of 35 min (Huang et al., 2018). We collected four, three, three, two, and two particle filter samples for the periods when the particle concentrations reached stable levels in experiments at 213, 243, 273, 298, and 313 K, respectively. The data analysis was done with the Tofware software (version 3.1.2). Note that the reagent ion I⁻ (m/z = 126.9) was subtracted from the mass-to-charge ratio of all the molecules shown in this work. For the comparison of measured total particle mass concentration between AMS, SMPS, and FIGAERO-iodide-CIMS, we use the maximum sensitivity of 22 cps ppt^{-1} (ppt: parts per trillion) to convert the signals to mass concentration (Lee et al., 2014; Lopez-Hilfiker et al., 2016). In addition, we did a mass calibration

for β -caryophyllinic acid (95%, Toronto Research Chemicals), resulting in a sensitivity of $(2.4^{+0.96}_{-0.63})$ cps ppt⁻¹ (details are given in Fig. S2). We included I⁻ (m/z = 127 Th), I(H₂O)⁻ (m/z = 145 Th), I(CH₂O₂)⁻ (m/z = 173 Th), I⁻₃ (m/z = 381 Th), and a dominating product ion C₃₀H₄₈O₅I⁻ (m/z = 615 Th) for our mass calibration. With this procedure we found the mass defect of most compounds in a good linear correlation and without significant deviations at high masses (see Fig. S3). We constrained the peak assignment errors to 20 ppm.

Typically, background measurements for both the gas and particle phase were done before and after the first addition of BCP to identify any contamination inside the chamber. However, gas background levels were almost negligible for most experiments, and most of the particle background signals were from filter matrix contaminations mainly due to fluorinated constituents. We subtracted the mass spectra of the background filter samples from those of the particle-loaded filter samples for the same experiments.

2.3 Rate coefficient calculation

Based on the PTR-MS measurements of the BCP decay as well as the ozone measurements, the rate coefficients of the reaction of BCP with ozone can be determined. As BCP has two double bonds with reactivities with ozone differing by a factor of 100 (Shu and Atkinson, 1994), here we discuss only the rate coefficient for the reaction of the most reactive endocyclic double bond. Since we did not use an OH radical scavenger, the reaction between BCP and OH radicals, which are generated from BCP ozonolysis, was included in our analysis. Employing the following reaction scheme, the observed decays of ozone and BCP were fitted by adjusting only the rate for Reaction (R1).

$$BCP + O_3 \rightarrow products + \gamma \cdot OH \tag{R1}$$

$$BCP + \bullet OH \rightarrow products$$
 (R2)

For the reaction of OH radicals with BCP, the rate coefficient of $(1.97 \pm 0.25) \times 10^{-10}$ cm³ molec.⁻¹ determined for 296 K (Shu and Atkinson, 1995; Winterhalter et al., 2009; Mellouki et al., 2021) was used for all temperatures due to the lack of its temperature dependence.

To integrate this simple model and to fit the rate coefficient for Reaction (R1) we used the software KinSim (Kinetics Simulator for Chemical-Kinetics and Environmental-Chemistry Teaching, version 4.14) (Peng and Jimenez, 2019). The OH radical yields (γ in R1) varied between 5%–15% at different temperatures (see Sect. 3.1).

2.4 SOA yield calculation

The SOA yields (Y_{SOA}) were calculated as $Y_{SOA} =$ $\Delta M_{\rm org} / \Delta \text{VOC}$, where $\Delta M_{\rm org}$ is the SOA mass formed from the reacted mass of BCP (Δ VOC). Similarly, the yields (Y_{BCA}) of BCA, the typical product of BCP ozonolysis, were calculated as $Y_{BCA} = \Delta M_{BCA} / \Delta VOC$, where ΔM_{BCA} is the mass concentration of β -caryophyllinic acid formed from the reacted mass of BCP (Δ VOC). We used a more than 5fold ozone excess in this study for all temperatures to facilitate oxidation of all double bonds in BCP and its oxidation products. This should lead to more comparable yields for all conditions studied (Li et al., 2011; Chen et al., 2012). Wall losses of particles and semi-volatile trace gases were calculated with the aerosol dynamic model COSIMA (Naumann, 2003; Saathoff et al., 2009) and used to correct the yields. The yields were calculated for the initial period of the experiments, which lasted about 90 min. During this relatively short time period and due to the large size of the simulation chamber, particle losses contributed typically 6% or less to the total SOA mass. Due to the relatively small particle sizes at the beginning of the experiments, diffusional losses dominated.

3 Results and discussion

In this chapter we discuss the typical course of the experiments, the kinetics of the reaction of BCP with ozone, the SOA mass yields as well as the SOA chemical composition in the absence and in the presence of nitrogen oxides.

3.1 Overview of the experiments and kinetics of BCP ozonolysis

Figure 2 shows the evolution of trace gases as well as particle mass and size distribution for BCP ozonolysis at 298 K first without and then in the presence of nitrogen oxides in experiment 4a and 4b. The time point when ozone was firstly added is regarded as the start of each experiment and is set to zero. As shown in Fig. 2, following the nucleation, the particle size increased to 36 nm and grew further to 57 and 69 nm after subsequent additions of ozone and BCP and finally to 122 nm after NO₂ addition. The particle mass increased in the presence of the low initial ozone concentration (25 ppb) and stabilized within 20 min after increasing the ozone level to above 300 ppb, reaching $13.3 \,\mu g \, m^{-3}$. A second addition of more BCP led to another increase in the particle mass stabilizing at a level of $17.0 \,\mu g \, m^{-3}$. To this SOA, 42 ppb of NO₂ were added, which reacted with the excess of ozone that forms NO₃ radicals and consequently led to a small increase in particle size and mass $(1.9 \text{ nm and } 0.9 \text{ µg m}^{-3}, \text{ re-}$ spectively) due to their reaction with BCP oxidation products. After the third addition of BCP to the reaction mixture, the SOA mass concentration increased to $50.6 \,\mu g \, m^{-3}$. For the other four temperatures, these values are given in Table 1, and their time evolution is shown in Fig. S4. During the final addition of BCP, the existing particles grew quickly, and new particle formation was observed. The discrepancy in mass concentrations obtained from SMPS and HR-ToF-AMS in the initial phase of the experiment is due to the fact that smaller particles are poorly detected by the HR-ToF-AMS due to the lower transmission of sub-100 nm particles in the aerodynamic lens (Liu et al., 2007; Williams et al., 2013). After 160 min, the particle size reached 122 nm, and both mass concentrations agreed well assuming a collection efficiency of 0.6 for the AMS. Please note that the FIGAEROiodide-CIMS is sensitive to more polar oxidized compounds; thus, the sum of all compounds detected by CIMS can only be a fraction of the total organic aerosol compounds measured by HR-ToF-AMS. The total organic mass concentration detected by FIGAERO-iodide-CIMS, if assuming an average maximum sensitivity for all compounds, corresponds to 36 %-61 % of the total organic aerosol mass measured by AMS.

After the first addition of an excess of ozone, BCP was depleted within less than 5 min. An example of the kinetic model results for the experiment at 313 K is compared to the measured data in Fig. S5, and all kinetic parameters fitted are listed in Table S2. The OH radical yields from the ozonolysis reaction increase from (5 ± 2) % at 243 K to (15 ± 2) % at 313 K, and 91%–92% of the BCP is calculated to react with ozone under 243–313 K.

Figure 3 shows the temperature dependence of the rate coefficients for BCP ozonolysis. The rate coefficient we determined for 298 K of $(1.09 \pm 0.21) \times 10^{-14} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$ agrees very well with values from Shu and Atkinson (1994) of $(1.16 \pm 0.43) \times 10^{-14} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$ and from Richters et al. (2015) of $(1.1 \pm 0.3) \times 10^{-14} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$ for 296 K. At lower temperatures, the rate coefficient increases. This is in agreement with the density-functional theory (DFT) quantum chemical calculations by Nguyen et al. (2009) but with a slightly higher slope. The Arrhenius equation fitted to our measured values is k = $(1.6 \pm 0.4) \times 10^{-15} \times \exp((559 \pm 97)/T)$. This corresponds to a reaction enthalpy of (5.6 ± 1.0) kJ mol⁻¹ in our analysis. Please note that the experiment at 273 K was not used for this analysis because of an unusual background signal in the PTR-MS measurement (Fig. S7b).

3.2 Particle mass yields

Figure 4 presents the yields from BCP ozonolysis in the absence of NO₂ as a function of the temperature for a constant total organic particle mass concentration (M_{org}) of 10 µg m⁻³. The SOA yields decrease with increasing temperatures from 243 to 313 K. For this organic particle mass concentration, we determined a SOA yield of (19.4±6)% at 298 K, which was lower by around 55% and 40% than those reported by Chen et al. (2012) and Tasoglou and Pandis (2015), respectively, for a similar particle mass. However, the results are still within the combined uncertainty limits. Please note that these two previous studies used OH radical scavengers, which should lead to lower yields, since OH radicals may potentially oxidize more VOCs, causing higher SOA yields for the ozonolysis without the OH scavenger system. The temperature dependence we observed is significantly lower than, for example, for α -pinene (Saathoff et al., 2009) and hence reflects the generally lower vapour pressure of the condensable oxidation products compared to the monoterpene. In our study, the SOA formation time of about 90 min was longer than the typical lifetime of the first-generation products from BCP ozonolysis (20 min) at an ozone level of 300 ppb; thus the difference cannot be explained by potential incomplete reactions. Different initial ozone levels may also contribute to the higher yields reported by Chen et al. (2012) and Tasoglou and Pandis (2015). For our study the wall losses and corresponding corrections were relatively small and cannot explain deviations of 40 %-55 %. Typically, the losses of acidic gases are larger in an aluminium chamber compared to a Teflon chamber, and it may be the opposite for particle losses. This depends on the age of the chamber walls, potential electrostatic losses, and the volume-to-surface ratio of the chamber. It is therefore not easy to determine the impact of the different wall losses and the wall loss corrections. We can only speculate if too-high wall loss corrections for the studies in the Teflon chamber contribute to the different yields.

An overview of the SOA yields determined by Lee et al. (2006), Winterhalter et al. (2009), Chen et al. (2012), and Tasoglou and Pandis (2015) is given in Fig. S6 in comparison with our results in the temperature range between 243 and 313 K. The SOA masses formed reflect the different oxidation products of differing volatilities (Chan et al., 2007) formed at low and high temperatures, which is discussed in the following section.

3.3 Chemical characterization of SOA from BCP ozonolysis without nitrogen oxides

In this section the gas- and particle-phase molecular composition is discussed mainly based on the FIGAERO-iodide-CIMS mass spectra, complemented by the CHARON-PTR-MS.

3.3.1 Composition of the gas-phase products

Figure 5 shows the averaged gas-phase mass spectra for all five temperatures in the range 213–313 K before adding NO₂ and when the SOA concentration was in a relatively stable state. Only compounds with mass-tocharge ratios less than 400 Th contributed significantly to the FIGAERO-iodide-CIMS spectra. The signals were normalized to the total gas-phase $C_x H_y O_z$ compounds at each temperature. In the temperature range of 213–



Figure 2. Evolution of trace gases as well as particle mass and size for β -caryophyllene ozonolysis at 298 K first without and then in the presence of NO₂. The time axis is relative to the first addition of ozone. The top panel shows the concentration of ozone (blue) and NO₂ (yellow) as well as the experiment steps; the middle panel shows the change in β -caryophyllene (pink) and particle concentration from the SMPS (blue dots), COSIMA model (black line), AMS (green dots), and CIMS (red triangle); the bottom panel shows the particle size distribution over the course of the whole experiment.

273 K, $C_{15}H_{24}O_3$ (likely β -caryophyllonic acid), $C_{15}H_{24}O_4$ (likely β -hydroxycaryophyllonic acid), C₁₅H₂₆O₄ (not identified yet), and $C_{14}H_{22}O_4$ (β -nocaryophyllonic acid or β caryophyllinic acid) were the most dominant monomeric compounds in the gas phase. These compounds have been identified as products from BCP ozonolysis at room temperature in previous studies (Jaoui et al., 2003; Winterhalter et al., 2009; Chan et al., 2011; Li et al., 2011). The absolute signals of all detected gas-phase species are shown in the Fig. S8. In addition, we also observed several abundant ions, such as $C_{14}H_{20}O_2$, $C_{15}H_{22}O_2$, and $C_{14}H_{20}O_3$, in the PTR-MS mass spectra (Fig. S7). As mentioned above, the PTR-MS is more sensitive to less-oxygenated organic compounds and is prone to fragmentation, e.g. via loss of H₂O. In combination with FIGAERO-iodide-CIMS data, it is reasonable to interpret that C₁₅H₂₂O₂ in PTR-MS mass spectra is most likely the fragmentation ion from C₁₅H₂₄O₃ via loss of H₂O. Besides, C₁₅H₂₄O₂ and its potential fragmentation ion C₁₅H₂₂O could be tentatively identified as BCP aldehyde based on previous studies (Li et al., 2011). This indicates that the lowly oxidized monomeric compounds (O < 5), which are abundant at lower temperatures, can also be formed at higher temperatures (273–313 K) but react (see Fig. S10) to form more highly oxygenated and less volatile compounds. It cannot be excluded that low-molecular-weight compounds, like $C_4H_6O_4$ and $C_2H_4O_3$, might be from a similar process.

3.3.2 Composition of the particle-phase products

Figure 6 shows the particle-phase mass spectra from BCP ozonolysis for 213 and 313 K. For both the lowest and highest temperatures, several first- and second-generation oxidation products, some of which have been identified also in previous studies, were observed, e.g. C₁₄H₂₂O₄ $(\beta$ -caryophyllinic acid or β -hydroxynocaryophyllon aldehyde), $C_{15}H_{24}O_3$ (β -hydroxycaryophyllon aldehyde or β -caryophyllonic acid), and C₁₄H₂₂O₇ (2,3-dihydroxy-4-[2-(4-hydroxy-3-oxobutyl)-3,3-dimethylcyclobutyl]-4oxobutanoic acid) (Jaoui et al., 2007; Li et al., 2011; Winterhalter et al., 2009; Chan et al., 2011; Jaoui et al., 2003; Griffin et al., 1999; Lee et al., 2006a, b; Nguyen et al., 2009; Jenkin et al., 2012; Richters et al., 2016). Table S3 lists all major particulate $C_x H_y O_z$ compounds detected by FIGAERO-iodide-CIMS in this work. Furthermore, we also observed, for example, $C_{14}H_{20}O_2$ H⁺, $C_{14}H_{20}O_3$ H⁺,



Figure 3. Arrhenius plot of the rate coefficients determined for the reaction of ozone with the endocyclic double bond of β -caryophyllene (blue diamonds) compared to the values measured by Shu and Atkinson et al. (1994) (green circle) and Richter et al. (2016) at 296 K (grey triangle) as well as model calculations by Nguyen et al. (2009) (dashed yellow line). The dashed blue line is a fit to the rate coefficients determined in this work.



Figure 4. SOA mass yield from β -caryophyllene ozonolysis at a constant total organic aerosol mass (M_{org}) of 10 µg m⁻³ for temperatures between 243–313 K from this study in comparison with literature data. The dashed line represents a single exponential fit to the data. Chen et al. (2012) used OH/CI scavenger and ammonium sulfate as seeds; Tasoglou and Pandis (2015) used OH/CI scavenger. Another comparison with literature data is given in Fig. S6.

 $C_{14}H_{22}O_3$ H⁺, $C_{15}H_{22}O_2$ H⁺, and $C_{15}H_{22}O_3$ H⁺ with the CHARON-PTR-MS (Fig. S9). From previous studies we know that CHARON-PTR-MS measurements are affected by fragmentation, e.g. by losing H₂O, CO, and CO₂, due



Figure 5. Averaged CIMS gas-phase mass spectra (background subtracted) for all temperatures, indicating the gas information when the SOA got stable in the absence of nitrogen oxides. The particle information sampled at the same time is given in Figs. 6 and 7.

to the relatively high collisional energy in the drift tube (100–170 Td) (Gkatzelis et al., 2018). In this study, we operated the CHARON-PTR-MS at 100 Td for the particle measurement. Taking fragmentation by losing one H₂O molecule into account, it is reasonable to speculate that $C_{14}H_{20}O_3$ and $C_{15}H_{22}O_2$ are from the fragmentation of $C_{14}H_{22}O_4$ and $C_{15}H_{24}O_3$, respectively. Similarly, $C_{14}H_{20}O_2$ and $C_{15}H_{22}O_3$ in CHARON-PTR-MS mass spectra could be identified as the fragmentation ions from $C_{14}H_{22}O_3$ and $C_{15}H_{24}O_4$, respectively, which were the most abundant ions in the FIGAERO-iodide-CIMS mass spectra as shown in Fig. 6.

At 213 K, monomers, dimers, and trimers are clearly visible. The most abundant compounds measurable with our FIGAERO-iodide-CIMS in each group are $C_{15}H_{24}O_3$, $C_{30}H_{48}O_5$, and $C_{44}H_{68}O_9$, respectively. At 313 K, the monomers dominate, and only a few dimers are observed, with the most abundant signals by compounds $C_{14}H_{22}O_7$ and $C_{29}H_{44}O_{10}$, respectively.

Compared to 213 K, it is evident that the monomeric compounds formed at 313 K are more oxygenated and have higher elemental O: C ratios (Fig. 6). It is inferred that at higher temperatures, once the first-generation products are formed in the gas phase, some of them can remain in the gas due to higher saturation vapour pressures and undergo further oxidation reactions of their unsaturated exocyclic double bonds with the excess of ozone. Since the saturation vapour pressures of the compounds at 213 K are substantially lower than at 313 K, more relatively less oxidized molecules are found in the particle phase due to rapid con-



Figure 6. CIMS mass spectra of particle-phase compounds obtained from ozonolysis of β -caryophyllene without NO₂ at 213 K (**a**) and 313 K (**b**). Colours refer to monomeric (brown), dimeric (blue), and trimeric (purple) compounds; labels represent the assignment of individual molecules.

densation. For the compounds remaining in the gas phase, HOMs such as, for example, C14H22O7 could be formed via simple or extended autoxidation (Richters et al., 2016). In addition to C14H22O7, we also detected several other compounds (e.g. C₁₄H₂₂O₉, C₁₅H₂₂O_{7.9}, and C₁₅H₂₄O_{7.9}) that are likely products of autoxidation reactions (Jokinen et al., 2016). At 313 K, the monomeric highly oxidized compounds (MHOCs; $C \le 15$, $O \ge 6$) dominate the monomers with a signal fraction of 42.5 % to total organic signals, which exceeds the contribution of other monomeric lowly oxygenated organic compounds (MLOCs; $C \le 15$, O < 6) with a signal fraction of 24.9%. For comparison, at 213 K, MHOCs and MLOCs contribute 6.9 % and 30.5 %, respectively. Furthermore, the six confirmed HOMs have signal fractions of 9.2 % at 313 K and nearly 0% at 213 K. Note that more highly oxygenated autoxidation products like $C_{15}H_{22}O_{11}$ (detected by atmospheric-pressure-interface time-of-flight mass spectrometer (CI-APi-ToF) using nitrate as the reagent ion; Jokinen et al., 2016) and C₁₅H₂₂O₁₃ could not be detected in this work due to instrument limitations (Riva et al., 2019). To summarize, the relative abundance of HOMs was higher at increasing temperatures (Fig. 8), indicating that the autoxidation showed a strong temperature dependence, slowing down at reduced temperatures.

The dimer groups show two completely different patterns at 213 and 313 K. At 213 K, the dimers mainly consist of lowly oxygenated organic compounds (DLOCs), with 28–30 carbon atoms and 5–8 oxygen atoms, and the total signal fraction of the dimeric compounds is 53.7 %. At 313 K, dimers are more oxygenated, with 9–11 oxygen atoms with the same carbon number of 28–30 (DHOCs), and they con-



Figure 7. CIMS mass spectra for particle-phase compounds for all five temperatures (213–313 K). Colours refer to monomeric (brown), dimeric (blue), and trimeric (purple) compounds. The table lists the signal-weighted bulk O : C ratios for each temperature. The mass-to-charge ratios on the *x* axis are subtracted by I⁻ (m/z = 126.9).

tribute less to the total normalized organic signal (27.6%). Here, we put forward esterification as a potential pathway for forming the most abundant dimers at both 213 and 313 K, similar to the dimer formation of other biogenic VOCs suggested by Yasmeen et al. (2010) and Müller et al. (2008).



Figure 8. Product groups contributing to SOA from ozonolysis of β -caryophyllene without NO₂ at temperatures between 213–313 K. Autoxidation-product calculation includes C₁₄H₂₂O_{7,9}, C₁₅H₂₂O_{7,9}, and C₁₅H₂₄O_{7,9} identified in previous studies (Jokinen et al., 2016; Richters et al., 2016).

The dominating dimeric molecule at 213 K, $C_{30}H_{48}O_5$, could potentially be formed via esterification of two of the most abundant monomers ($C_{15}H_{26}O_3$ and $C_{15}H_{24}O_3$). In a similar way, the dimer $C_{29}H_{44}O_9$ could be formed at 313 K from $C_{14}H_{22}O_7$ and $C_{15}H_{24}O_3$. These reactions are described in Fig. S11 in the Supplement.

However, we cannot exclude the reaction of a BCP-derived stabilized Criegee intermediate and an abundant acid as a potential pathway for $C_{30}H_{48}O_5$ formation, similar to mechanisms suggested for α -pinene (Kristensen et al., 2016; Wang et al., 2016; Witkowski and Gierczak, 2014; Lee and Kamens, 2005), for example. The dimers we observed had no more than 11 oxygen atoms, which is due to low sensitivities of FIGAERO-iodide-CIMS to these compounds (Riva et al., 2019).

Significant signals of trimers at 213 K are assigned to $C_{41-43}H_{62-68}O_{9-11}$, but they are not detected at 313 K. The potential formation pathway of $C_{44}H_{68}O_9$ at 213 K is also included in Fig. S11. Please note that the assignment for ions at mass-to-charge ratios of more than 700 Th has significantly higher uncertainties as for smaller mass peaks.

Figure 7 presents the averaged mass spectra of SOA particles formed from BCP ozonolysis before NO₂ addition at all five temperatures. With increasing temperature, the most abundant monomeric compounds shift to higher masses with higher elemental O : C ratios, e.g. from C₁₅H₂₄O₃ (m/z =252.2, O : C = 0.20, 213–243 K) and C₁₅H₂₄O₄ (m/z =268.2, O : C = 0.27, 273 K) to C₁₄H₂₂O₆ (m/z = 286.1, O : C = 0.43, 298 K) and C₁₄H₂₂O₇ (m/z = 302.1, O : C = 0.5, 313 K). This indicates again that at higher temperatures, after ozonolysis of the endocyclic bond (formation of firstgeneration oxidation products), the unsaturated compounds can react further with the excess of ozone and form more highly oxidized products (e.g. HOMs), while those formed at lower temperatures would partition into the condensed phase before further oxidation can occur. As shown in Fig. 8, the monomer groups (MHOCs+MLOCs) contribute 39.2%, 64.7%, 85.3%, 68.9%, and 67.5% to the total signal from 213 to 313 K, respectively. Among all the monomers, the signal fraction of the six identified HOMs to the total signals of all organic species has a monotonic positive temperature dependence (see Fig. 8), increasing from 0.1% to 9.2% for temperatures increasing from 213 to 313 K. This is a similar correlation between HOM formation and temperatures as observed by (Bianchi et al., 2019).

Two different dimeric patterns appear in the temperature range of 213–313 K (Figs. 6 and 7). One pattern is represented by molecular formulae of $C_{28-30}H_{42-48}O_{5-8}$ at 213–243 K (marked as the low-temperature group, LT group), and the other pattern is represented by $C_{28-30}H_{36-44}O_{9-11}$ at 273–313 K (marked as the high-temperature group, HT group). The LT group contributes 53.7 % at 213 K and 32.8 % at 243 K, with a negative temperature dependence, while the dimeric signal fraction of the HT group is lower than the LT group, contributing 13.8 % at 273 K, 24.6 % at 298 K, and 27.6 % at 313 K, respectively, as shown in Fig. 8. After 273 K, dimer formation is enhanced with increasing temperatures. The contributions in the gas and particle phase of the major compounds are shown in Fig. S12.

3.4 Chemical characterization of SOA from BCP ozonolysis in the presence of nitrogen oxides

In this section we discuss the chemical composition of SOA from BCP ozonolysis in the presence of nitrogen oxides including NO₃ radicals, which refers to the SOA formed after the last addition of BCP, e.g. 150-160 min in Fig. 2. Please note that the results given here refer to the total SOA, which was formed in two steps, first by pure ozonolysis and in a subsequent step including nitrogen oxides. The SOA mass formed in the presence of nitrogen oxides compared to the total SOA mass detected corresponds to 49 % at 213 K, 34 % at 243 K, 49 % at 273 K, 65 % at 298 K, and 63 % at 313 K. Due to the excess of ozone besides NO₂, also NO₃ radicals and N₂O₅ were present. Thus, the BCP was now also oxidized by reaction with NO₃ radicals, with a major pathway to produce organonitrates (org-N's) (Kiendler-Scharr et al., 2016; Wu et al., 2021). The concentrations of NO₃ radicals before the final addition of BCP were estimated for each experiment using a kinetic box model (details in Fig. S13), and the results are given in Table 1.

Based on the ozone and NO₃ concentration levels as well as the corresponding reaction rates, we estimated the fraction of NO₃ radicals contributing to the initial BCP oxidation to be 84 %, 90 %, and 72 % for 273, 298, and 313 K. Please note that these values should be considered to be upper limits due to other potential sinks for NO₃ radicals.

3.4.1 Mass spectra from the gas-phase products

Figure 9 shows the averaged gas-phase mass spectra including molecules without a nitrogen atom (org) and organonitrates (org-N's) for all temperatures as detected by CIMS after the particle concentration became stable after the last BCP addition (e.g. at 298 K, referring to the time period after 160 min in Fig. 2). The gas-phase organonitrates showed an abundance increasing with temperature from 20.7 % at 213 K, 26.0 % at 243 K, 38.3 % at 273 K, and 46.5 % at 298 K to 48.9 % at 313 K. These compounds consisted of three groups of different carbon numbers (C₅H₇O₆N, $C_{10}H_{15}O_{5-7}N$ and $C_{15}H_{23,25}O_{6-8}N$). To illustrate more clearly the organonitrate formation, we show the time evolution of the three most abundant org-N's in these three groups in the right panel of Fig. 9. The signals of C₅H₇O₆N and C₁₀H₁₅O₆N increased after the NO₂ addition immediately, indicating that their formation was related to reactions between BCP oxidation products (BCP+O₃) and NO₃ radicals. After the start of the last BCP addition, C15H25O7N started to increase significantly, indicating that its formation was linked to the reaction of BCP and NO₃ directly, but not the reaction between the nitrate radicals and oxidation products from BCP ozonolysis.

3.4.2 Mass spectra from the particle phase

Figure 10 shows the particle chemical composition of SOA from BCP ozonolysis in the presence of NO₂ at 213 K (upper panel) and 313 K (lower panel). At 213 K (upper panel), org dominated the particle composition, with a signal fraction of 97.1 % to total signals (org-N's + org), similar to the SOA before NO₂ addition at 213 K. Org groups of monomers (41.2 % in signal fraction), dimers (48.2 %), and trimers (7.4 %) were observed. The most abundant signals in each group were $C_{15}H_{24}O_3$, $C_{30}H_{48}O_5$, and $C_{44}H_{68}O_9$, respectively. The largest org-N signal was from $C_{15}H_{25}O_7N$, contributing 0.6 % to the total signals. It is obvious that only few org-N's were formed at 213 K in our study, with a contribution ([org-N's]/[total org.]) of 2.8 % measured by FIGAERO-iodide-CIMS (normalized signal fraction, Fig. 12) and 0.08 detected by HR-AMS (Fig. S14).

Compared to the SOA chemical composition at 213 K, substantially more org-N's (48.9 % to total signals) were formed at 313 K. Monomeric org-N's ($C_{15}H_{23,25}O_{6-10}N$) made up 35.5 % of total signal (and the AMS measured 61 % of org-N's to total mass concentration of organic components; Fig. S14). The most abundant monomeric org-N signals at 313 K were $C_{15}H_{23-25}O_{7-9}N$, which had also been detected by Wu et al. (2021). In our study, 51.1 % of normalized signals were contributed by org species, with 28.3 % of monomeric org ($C_{13-15}H_{20-24}O_{4-8}$), 16.6 % of dimeric org ($C_{20}H_{22-24}O_{7-9}$), and 6.3 % of trimeric org ($C_{35}H_{48}O_{10-13}$).

The particle mass spectra for all temperatures are shown in Fig. 11, with relative signal abundance of individual molecules to total detected species, including pure organic components (org; red) and organonitrates (org-N's; blue). The most abundant mass spectral peak of all organic compounds without nitrogen at 298–313 K was not C₁₄H₂₂O₇ as for pure ozonolysis, but $C_{15}H_{24}O_4$ in the presence of NO₂ and NO₃ radicals. This is attributed to additional formation of pure MLOCs. This could be confirmed by comparison of the absolute signals of $C_{14}H_{22}O_7$ and $C_{15}H_{24}O_4$ in the particles before and after the last BCP addition (Fig. S15). Moreover, the dimeric compounds with 20 carbon atom skeletons (i.e. $C_{20}H_{24}O_{7-8}$; insertion in Fig. 10) and the trimeric compounds with 35 carbon atom skeletons (i.e. $C_{35}H_{48}O_{12}$; Figs. 10 and 11) were newly formed at higher temperatures (> 273 K) in the presence of nitrogen oxides. However, it cannot be excluded that they could be formed by the reaction of the oxidation products from pure BCP ozonolysis with those formed in the presence of NO₂ and NO₃ radicals. In contrast, the mass spectra of non-N-containing organic species (org) at 213-243 K showed no substantial changes compared to the species from ozonolysis without nitrogen oxides present (see Fig. 7). One obvious reason for this may be the lower NO₃ radical concentrations at lower temperatures, but also changes in the active reaction pathways may play a role.

In addition, more organonitrates were detected at increased temperatures, ranging from 2.8 % (213 K) to 51.5 % (298 K) and 48.9 % (313 K), dominated by monomeric org-N's ($C_{15}H_{23,25}O_{6-10}N$), wherein the most abundant signals were from C₁₅H₂₅O₇N at 243-273 K and C₁₅H₂₃O₉N at 298-313 K. Monomeric organonitrates contributed 1.7 % to 40.1 % from 213 to 298 K, and 35.5 % at 313 K to total organic signals, as shown in Fig. 12. Besides, the signalweighted averaged O: C of organonitrates monotonically increased from 0.41 to 0.51 from 213 to 298 K and 0.50 at 313 K. We assume that the positive impact of temperature on the oxygenation and formation of organonitrates could also be relevant for the highest temperature (313 K). However, the more highly oxidized organonitrates may be out of the detection range of the FIGAERO-iodide-CIMS, resulting in lower signal fractions and lower signal-weighted averaged O: C ratios of organonitrates at 313 K. This is supported by HR-ToF-AMS measurements (Fig. S14), which show increasing organonitrate fractions from 8% to 61% also for 313 K. On the other hand, thermal instability of some N-containing compounds formed at 313 K, e.g. peroxy nitrates (Francisco and Krylowski, 2005; Lee et al., 2016), can also be an explanation for the weaker increase in their fraction observed for 313 K. For example, the potential contributions from thermal decomposition $(C_{1-13}H_vO_zN_1I^- \text{ ions})$ showed a positive temperature dependence. Also, abundant N-containing ions, such as low-molecular-weight molecule C₅H₇O₆N₁, desorbed substantially between 120–200 °C and had an unexpectedly high T_{max} of about 160 °C.



Figure 9. CIMS gas-phase mass spectra for all temperatures in the presence of NO₂ (**a**). These spectra correspond to the same time as the particle-phase spectra shown in Figs. 10 and 11. Colours refer to compounds without nitrogen atoms (black) and compounds with one nitrogen atom (blue). The I⁻ (m/z = 126.9) was subtracted, and here "mass/charge" refers to the weight of the molecules. Panel (**b**) shows the FIGAERO-iodide-CIMS measurement for the signal change of three typical organonitrate molecules for C5 (C₅H₇O₆N; red line), C10 (C₁₀H₁₅O₆N; blue line), and C15 (C₁₅H₂₅O₇N; black line) corresponding to the evolution starting after NO₂ addition at 298 K.



Figure 10. CIMS mass spectra of particle-phase compounds obtained from ozonolysis of β -caryophyllene in the presence of nitrogen oxides at 213 K (**a**) and 313 K (**b**). Colours refer to compounds without nitrogen atoms (red) and compounds with one nitrogen atom (blue).

It should be noted that the nitrate radical levels were estimated to be higher at higher temperatures (Table 1), which partially explains the increasing organonitrate formation at higher temperatures. The organonitrates detected in this study were dominantly with the same carbon atom skeleton (C_{15}) and only one organonitrate functional group (-ONO₂). The O : C ratios for organonitrates increase with

temperature. Thus, it can be concluded that higher temperatures favour the formation of more highly oxygenated organonitrates from the BCP oxidation in the presence of O_3 , NO₂, and NO₃ radicals. The contributions in the gas and particle phase of the major compounds are shown in Fig. S16.



Figure 11. CIMS mass spectra for particle-phase compounds for all five temperatures (213–313 K). Colours refer to $C_x H_y O_z N_0$ (brown) and $C_x H_y O_z N_1$ (blue) compounds. The table lists the average O : C ratios for each temperature.



Figure 12. Product groups (monomeric org, dimeric org, trimeric org, monomeric org-N's, dimeric org-N's, and trimeric org-N's) contributing to SOA from ozonolysis of β -caryophyllene in the presence of nitrogen oxides at temperatures between 213–313 K.

4 Summary and conclusions

In this work, a series of experiments conducted in the dark AIDA chamber with temperatures covering the whole tropospheric range (213–313 K) were analysed to investigate the yields, kinetics, and chemistry of BCP ozonolysis in the absence and presence of nitrogen oxides. The rate coefficient of the endocyclic double bond in BCP reacting with ozone was determined in the temperature range between 243 and 313 K, showing decreasing values with increasing tempera-

ture. The rate coefficients agree well with literature data at 296 K and with quantum chemical calculations of the temperature dependence. We determined a reaction enthalpy of (5.6 ± 1.0) kJ mol⁻¹ and OH radical yields increasing from (5 ± 2) % at 243 K to (15 ± 2) % at 313 K.

SOA yields determined for the ozonolysis of BCP show about 50% smaller values than literature data at 298 K, which is still within the combined uncertainty limits. For the first time the temperature dependence of the SOA yield was determined, showing values decreasing from (37 ± 11) % to (16 ± 5) % for temperatures increasing from 243 to 313 K and a constant organic particle mass of $10 \,\mu g \,m^{-3}$. This allows the calculation of the potential of BCP to contribute to SOA formation, e.g. if reaching higher altitudes in the atmosphere.

The chemical characteristics of BCP pure organic species and organonitrates were determined by the FIGAERO-CIMS using I⁻ as the reagent ion. Major products and different chemical composition of the gas and particle phase with and without NO₃ present at all temperatures were resolved. The variation in the ozonolysis temperature revealed substantial impact on the abundance of individual pure organic molecules without nitrogen atoms (org) and organonitrate molecules (org-N's). In the first generated SOA without nitrogen oxides present, monomers (mainly C₁₄₋₁₅H₂₂₋₂₄O₃₋₇) and dimers (mainly C₂₈₋₃₀H₄₄₋₄₈O₅₋₉) were abundant, wherein minor signals of trimers (mainly C₄₁₋₄₄H₆₂₋₆₆O₉₋₁₁) were mainly detected at 213 K. Potential dimer and trimer formation pathways are proposed. With temperature increasing to 313 K, monomers and dimers (C14-15H22-24O6-9 and C₂₇₋₂₉H₄₂₋₄₄O₉₋₁₁, respectively) became more oxidized, and no significant trimeric signals were detected.

The positive temperature dependence of the oxygenation of the products was also observed in the BCP organonitrates. In the presence of nitrogen oxides, most of the organonitrates were found as monomers with a C_{15} skeleton with one nitrate group, with their signal-weighted O : C ratio increasing from 0.41 to 0.51 for temperatures in the range between 213 and 313 K. Dimeric and trimeric pure organic species without nitrogen atoms (C_{20} , C_{35}) were newly formed in the presence of nitrogen oxides at 298–313 K, which substantially changed the chemical composition of pure organic components and indicates that more termination ways might exist.

Please note that we cannot exclude completely that the changing humidity in our experiments has no influence on our results. Nonetheless, we think that the SOA composition we have observed is dominated by gas-phase reactions with sufficient water concentrations compared to ambient conditions and that the potential impact of relative humidity on condensed phase reactions was minimized by increasing relative humidity with decreasing temperature (Li and Shiraiwa, 2019; Maclean et al., 2021), maybe except for the lowermost temperature (213 K).

This work helps to get a better understanding of the yields, kinetics, and chemical composition of SOA from BCP ozonolysis over a relatively wide range of temperatures, 213 to 313 K, which is representative of the real atmosphere from the boundary layer to the upper troposphere. In addition, SOA volatility is expected to have a strong impact on the SOA formation and will be discussed in our upcoming paper focusing on the desorption of BCP SOA from filters but also the warming-up of the aerosol in the simulation chamber overnight.

Data availability. The data on SOA yields and chemical composition from FIGAERO-iodide-CIMS are accessible from the KITopen data repository (https://doi.org/10.5445/IR/1000144632; Gao et al., 2022). Data are available upon request to the corresponding author.

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Competing interests. At least one of the (co-)authors is a member of the editorial board of *Atmospheric Chemistry and Physics*. The peer-review process was guided by an independent editor, and the authors also have no other competing interests to declare.

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References

- Agelopoulos, N. G., Chamberlain, K., and Pickett, J. A.: Factors Affecting Volatile Emissions of Intact Potato Plants, Solanum tuberosum: Variability of Quantities and Stability of Ratios, J. Chem. Ecol., 26, 497–511, https://doi.org/10.1023/A:1005473825335, 2000.
- Alfarra, M. R., Hamilton, J. F., Wyche, K. P., Good, N., Ward, M. W., Carr, T., Barley, M. H., Monks, P. S., Jenkin, M. E., Lewis, A. C., and McFiggans, G. B.: The effect of photochemical ageing and initial precursor concentration on the composition and hygroscopic properties of β-caryophyllene secondary organic aerosol, Atmos. Chem. Phys., 12, 6417–6436, https://doi.org/10.5194/acp-12-6417-2012, 2012.
- Andreae, M. O., Afchine, A., Albrecht, R., Holanda, B. A., Artaxo, P., Barbosa, H. M. J., Borrmann, S., Cecchini, M. A., Costa, A., Dollner, M., Fütterer, D., Järvinen, E., Jurkat, T., Klimach, T., Konemann, T., Knote, C., Krämer, M., Krisna, T., Machado, L. A. T., Mertes, S., Minikin, A., Pöhlker, C., Pöhlker, M. L., Pöschl, U., Rosenfeld, D., Sauer, D., Schlager, H., Schnaiter, M., Schneider, J., Schulz, C., Spanu, A., Sperling, V. B., Voigt, C., Walser, A., Wang, J., Weinzierl, B., Wendisch, M., and Ziereis, H.: Aerosol characteristics and particle production in the upper troposphere over the Amazon Basin, Atmos. Chem. Phys., 18, 921–961, https://doi.org/10.5194/acp-18-921-2018, 2018.
- Bianchi, F., Kurtén, T., Riva, M., Mohr, C., Rissanen, M. P., Roldin, P., Berndt, T., Crounse, J. D., Wennberg, P. O., Mentel, T. F., Wildt, J., Junninen, H., Jokinen, T., Kulmala, M., Worsnop, D. R., Thornton, J. A., Donahue, N., Kjaergaard, H. G., and Ehn, M.: Highly Oxygenated Organic Molecules (HOM) from Gas-Phase Autoxidation Involving Peroxy Radicals: A Key Contributor to Atmospheric Aerosol, Chem. Rev., 119, 3472–3509, https://doi.org/10.1021/acs.chemrev.8b00395, 2019.
- Calogirou, A., Larsen, B., and Kotzias, D.: Gas-phase terpene oxidation products: a review, Atmos. Environ., 33, 1423–1439, 1999.
- Chan, A. W. H., Kroll, J. H., Ng, N. L., and Seinfeld, J. H.: Kinetic modeling of secondary organic aerosol formation: effects of particle- and gas-phase reactions of semivolatile products, Atmos. Chem. Phys., 7, 4135–4147, https://doi.org/10.5194/acp-7-4135-2007, 2007.
- Chan, M. N., Surratt, J. D., Chan, A. W. H., Schilling, K., Offenberg, J. H., Lewandowski, M., Edney, E. O., Kleindienst, T. E., Jaoui, M., Edgerton, E. S., Tanner, R. L., Shaw, S. L., Zheng, M., Knipping, E. M., and Seinfeld, J. H.: Influence of aerosol acidity on the chemical composition of secondary organic aerosol from β-caryophyllene, Atmos. Chem. Phys., 11, 1735–1751, https://doi.org/10.5194/acp-11-1735-2011, 2011.
- Charnawskas, J. C., Alpert, P. A., Lambe, A. T., Berkemeier, T., O'Brien, R. E., Massoli, P., Onasch, T. B., Shiraiwa, M., Moffet, R. C., and Gilles, M. K.: Condensed-phase biogenic– anthropogenic interactions with implications for cold cloud formation, Faraday Discuss., 200, 165–194, 2017.
- Chen, Q., Li, Y. L., McKinney, K. A., Kuwata, M., and Martin, S. T.: Particle mass yield from β -caryophyllene ozonolysis, At-

mos. Chem. Phys., 12, 3165–3179, https://doi.org/10.5194/acp-12-3165-2012, 2012.

- Cheng, Y., Ma, Y., and Hu, D.: Tracer-based source apportioning of atmospheric organic carbon and the influence of anthropogenic emissions on secondary organic aerosol formation in Hong Kong, Atmos. Chem. Phys., 21, 10589–10608, https://doi.org/10.5194/acp-21-10589-2021, 2021.
- Ciccioli, P., Brancaleoni, E., Frattoni, M., Di Palo, V., Valentini, R., Tirone, G., Seufert, G., Bertin, N., Hansen, U., and Csiky, O.: Emission of reactive terpene compounds from orange orchards and their removal by within-canopy processes, J. Geophys. Res.-Atmos., 104, 8077–8094, 1999.
- Clark, C. H., Kacarab, M., Nakao, S., Asa-Awuku, A., Sato, K., and Cocker III, D. R.: Temperature effects on secondary organic aerosol (SOA) from the dark ozonolysis and photo-oxidation of isoprene, Environ. Sci. Technol., 50, 5564–5571, 2016.
- Cox, R. A., Ammann, M., Crowley, J. N., Herrmann, H., Jenkin, M. E., McNeill, V. F., Mellouki, A., Troe, J., and Wallington, T. J.: Evaluated kinetic and photochemical data for atmospheric chemistry: Volume VII – Criegee intermediates, Atmos. Chem. Phys., 20, 13497–13519, https://doi.org/10.5194/acp-20-13497-2020, 2020.
- Dekermenjian, M.: FTIR analysis of aerosol formed in the ozone oxidation of sesquiterpenes, Aerosol Sci. Tech., 30, 349–363, 1999.
- Duhl, T. R., Helmig, D., and Guenther, A.: Sesquiterpene emissions from vegetation: a review, Biogeosciences, 5, 761–777, https://doi.org/10.5194/bg-5-761-2008, 2008.
- Fahey, D. W., Gao, R.-S., Möhler, O., Saathoff, H., Schiller, C., Ebert, V., Krämer, M., Peter, T., Amarouche, N., Avallone, L. M., Bauer, R., Bozóki, Z., Christensen, L. E., Davis, S. M., Durry, G., Dyroff, C., Herman, R. L., Hunsmann, S., Khaykin, S. M., Mackrodt, P., Meyer, J., Smith, J. B., Spelten, N., Troy, R. F., Vömel, H., Wagner, S., and Wienhold, F. G.: The AquaVIT-1 intercomparison of atmospheric water vapor measurement techniques, Atmos. Meas. Tech., 7, 3177–3213, https://doi.org/10.5194/amt-7-3177-2014, 2014.
- Faiola, C. L., Buchholz, A., Kari, E., Yli-Pirilä, P., Holopainen, J. K., Kivimäenpää, M., Miettinen, P., Worsnop, D. R., Lehtinen, K. E. J., and Guenther, A. B.: Terpene composition complexity controls secondary organic aerosol yields from scots pine volatile emissions, Sci. Rep., 8, 1–13, 2018.
- Fehsenfeld, F., Calvert, J., Fall, R., Goldan, P., Guenther, A. B., Hewitt, C. N., Lamb, B., Liu, S., Trainer, M., and Westberg, H.: Emissions of volatile organic compounds from vegetation and the implications for atmospheric chemistry, Global Biogeochem. Cy., 6, 389–430, 1992.
- Francisco, M. A. and Krylowski, J.: Chemistry of Organic Nitrates: Thermal Chemistry of Linear and Branched Organic Nitrates, Ind. Eng. Chem. Res., 44, 5439–5446, https://doi.org/10.1021/ie049380d, 2005.
- Gao, L., Song, J., Mohr, C., Huang, W., Vallon, M., Jiang, F., Leisner, T., and Saathoff, H.: Kinetics, SOA yields and chemical composition of secondary organic aerosol from βcaryophyllene ozonolysis with and without nitrogen oxides between 213 and 313 K, Repository KITopen [data set], https://doi.org/10.5445/IR/1000144632, 2022.

- Geron, C. D. and Arnts, R. R.: Seasonal monoterpene and sesquiterpene emissions from Pinus taeda and Pinus virginiana, Atmos. Environ., 44, 4240–4251, 2010.
- Gkatzelis, G. I., Tillmann, R., Hohaus, T., Müller, M., Eichler, P., Xu, K.-M., Schlag, P., Schmitt, S. H., Wegener, R., Kaminski, M., Holzinger, R., Wisthaler, A., and Kiendler-Scharr, A.: Comparison of three aerosol chemical characterization techniques utilizing PTR-ToF-MS: a study on freshly formed and aged biogenic SOA, Atmos. Meas. Tech., 11, 1481–1500, https://doi.org/10.5194/amt-11-1481-2018, 2018.
- Griffin, R. J., Cocker III, D. R., Flagan, R. C., and Seinfeld, J. H.: Organic aerosol formation from the oxidation of biogenic hydrocarbons, J. Geophys. Res.-Atmos., 104, 3555–3567, 1999.
- Grosjean, D., Williams, E. L., Grosjean, E., Andino, J. M., and Seinfeld, J. H.: Atmospheric oxidation of biogenic hydrocarbons: reaction of ozone with. beta.-pinene, D-limonene and transcaryophyllene, Environ. Sci. Technol., 27, 2754–2758, 1993.
- Hansen, U. and Seufert, G.: Temperature and light dependence of β -caryophyllene emission rates, J. Geophys. Res.-Atmos., 108, 4801, https://doi.org/10.1029/2003JD003853, 2003.
- Haque, M. M., Kawamura, K., and Kim, Y.: Seasonal variations of biogenic secondary organic aerosol tracers in ambient aerosols from Alaska, Atmos. Environ., 130, 95–104, 2016.
- Hu, D., Bian, Q., Li, T. W., Lau, A. K., and Yu, J. Z.: Contributions of isoprene, monoterpenes, β -caryophyllene, and toluene to secondary organic aerosols in Hong Kong during the summer of 2006, J. Geophys. Res.-Atmos., 113, D22206, https://doi.org/10.1029/2008JD010437, 2008.
- Huang, W., Saathoff, H., Pajunoja, A., Shen, X., Naumann, K.-H., Wagner, R., Virtanen, A., Leisner, T., and Mohr, C.: α-Pinene secondary organic aerosol at low temperature: chemical composition and implications for particle viscosity, Atmos. Chem. Phys., 18, 2883–2898, https://doi.org/10.5194/acp-18-2883-2018, 2018.
- Huang, W., Saathoff, H., Shen, X., Ramisetty, R., Leisner, T., and Mohr, C.: Seasonal characteristics of organic aerosol chemical composition and volatility in Stuttgart, Germany, Atmos. Chem. Phys., 19, 11687–11700, https://doi.org/10.5194/acp-19-11687-2019, 2019.
- Huang, W., Li, H., Sarnela, N., Heikkinen, L., Tham, Y. J., Mikkilä, J., Thomas, S. J., Donahue, N. M., Kulmala, M., and Bianchi, F.: Measurement report: Molecular composition and volatility of gaseous organic compounds in a boreal forest – from volatile organic compounds to highly oxygenated organic molecules, Atmos. Chem. Phys., 21, 8961–8977, https://doi.org/10.5194/acp-21-8961-2021, 2021.
- Jaoui, M., Leungsakul, S., and Kamens, R.: Gas and particle products distribution from the reaction of β -caryophyllene with ozone, J. Atmos. Chem., 45, 261–287, 2003.
- Jaoui, M., Lewandowski, M., Kleindienst, T. E., Offenberg, J. H., and Edney, E. O.: β-caryophyllinic acid: An atmospheric tracer for β-caryophyllene secondary organic aerosol, Geophys. Res. Lett., 34, L05816, https://doi.org/10.1029/2006GL028827, 2007.
- Jaoui, M., Kleindienst, T. E., Docherty, K. S., Lewandowski, M., and Offenberg, J. H.: Secondary organic aerosol formation from the oxidation of a series of sesquiterpenes: α -cedrene, β -caryophyllene, α -humulene and α -farnesene with O₃, OH and NO₃ radicals, Environ. Chem., 10, 178–193, 2013.

- Jardine, K., Yañez Serrano, A., Arneth, A., Abrell, L., Jardine, A., Van Haren, J., Artaxo, P., Rizzo, L. V., Ishida, F. Y., and Karl, T.: Within-canopy sesquiterpene ozonolysis in Amazonia, J. Geophys. Res.-Atmos., 116, D19301, https://doi.org/10.1029/2011JD016243, 2011.
- Jenkin, M. E., Wyche, K. P., Evans, C. J., Carr, T., Monks, P. S., Alfarra, M. R., Barley, M. H., McFiggans, G. B., Young, J. C., and Rickard, A. R.: Development and chamber evaluation of the MCM v3.2 degradation scheme for β-caryophyllene, Atmos. Chem. Phys., 12, 5275–5308, https://doi.org/10.5194/acp-12-5275-2012, 2012.
- Jokinen, T., Kausiala, O., Garmash, O., Peräkylä, O., Junninen, H., Schobesberger, S., Yan, C., Sipilä, M., and Rissanen, M. P.: Production of highly oxidized organic compounds from ozonolysis of beta-caryophyllene: laboratory and field measurements, Boreal Environ. Res., 21, 262–273, 2016.
- Kammer, J., Flaud, P. M., Chazeaubeny, A., Ciuraru, R., Le Menach, K., Geneste, E., Budzinski, H., Bonnefond, J. M., Lamaud, E., Perraudin, E., and Villenave, E.: Biogenic volatile organic compounds (BVOCs) reactivity related to new particle formation (NPF) over the Landes forest, Atmos. Res., 237, 104869, https://doi.org/10.1016/j.atmosres.2020.104869, 2020.
- Kari, E., Miettinen, P., Yli-Pirilä, P., Virtanen, A., and Faiola, C. L.: PTR-ToF-MS product ion distributions and humidity-dependence of biogenic volatile organic compounds, Int. J. Mass Spectrom., 430, 87–97, https://doi.org/10.1016/j.ijms.2018.05.003, 2018.
- Kiendler-Scharr, A., Mensah, A. A., Friese, E., Topping, D., Nemitz, E., Prevot, A. S. H., Äijälä, M., Allan, J., Canonaco, F., Canagaratna, M., Carbone, S., Crippa, M., Dall Osto, M., Day, D. A., De Carlo, P., Di Marco, C. F., Elbern, H., Eriksson, A., Freney, E., Hao, L., Herrmann, H., Hildebrandt, L., Hillamo, R., Jimenez, J. L., Laaksonen, A., McFiggans, G., Mohr, C., O'Dowd, C., Otjes, R., Ovadnevaite, J., Pandis, S. N., Poulain, L., Schlag, P., Sellegri, K., Swietlicki, E., Tiitta, P., Vermeulen, A., Wahner, A., Worsnop, D., and Wu, H.-C.: Ubiquity of organic nitrates from nighttime chemistry in the European submicron aerosol, Geophys. Res. Lett., 43, 7735–7744, https://doi.org/10.1002/2016GL069239, 2016.
- Kim, S., Karl, T., Helmig, D., Daly, R., Rasmussen, R., and Guenther, A.: Measurement of atmospheric sesquiterpenes by proton transfer reaction-mass spectrometry (PTR-MS), Atmos. Meas. Tech., 2, 99–112, https://doi.org/10.5194/amt-2-99-2009, 2009.
- Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon, M., Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim, A., Bernhammer, A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A. P., Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Carslaw, K.

S., and Curtius, J.: Ion-induced nucleation of pure biogenic particles, Nature, 533, 521–526, https://doi.org/10.1038/nature17953, 2016.

- Kivimäenpää, M., Babalola, A. B., Joutsensaari, J., and Holopainen, J. K.: Methyl Salicylate and Sesquiterpene Emissions Are Indicative for Aphid Infestation on Scots Pine, Forests, 11, 573, https://doi.org/10.3390/f11050573, 2020.
- Kleist, E., Mentel, T. F., Andres, S., Bohne, A., Folkers, A., Kiendler-Scharr, A., Rudich, Y., Springer, M., Tillmann, R., and Wildt, J.: Irreversible impacts of heat on the emissions of monoterpenes, sesquiterpenes, phenolic BVOC and green leaf volatiles from several tree species, Biogeosciences, 9, 5111– 5123, https://doi.org/10.5194/bg-9-5111-2012, 2012.
- Kristensen, K., Watne, Å. K., Hammes, J., Lutz, A., Petäjä, T., Hallquist, M., Bilde, M., and Glasius, M.: High-molecular weight dimer esters are major products in aerosols from α -pinene ozonolysis and the boreal forest, Environ. Sci. Technol. Lett., 3, 280–285, 2016.
- Laothawornkitkul, J., Taylor, J. E., Paul, N. D., and Hewitt, C. N.: Biogenic volatile organic compounds in the Earth system, New Phytol., 183, 27–51, 2009.
- Larsen, B., Lahaniati, M., Calogirou, A., and Kotzias, D.: Atmospheric oxidation products of terpenes: A new nomenclature, Chemosphere, 37, 1207–1220, 1998.
- Lee, A., Goldstein, A. H., Keywood, M. D., Gao, S., Varutbangkul, V., Bahreini, R., Ng, N. L., Flagan, R. C., and Seinfeld, J. H.: Gas-phase products and secondary aerosol yields from the ozonolysis of ten different terpenes, J. Geophys. Res.-Atmos., 111, D07302, https://doi.org/10.1029/2005JD006437, 2006a.
- Lee, A., Goldstein, A. H., Kroll, J. H., Ng, N. L., Varutbangkul, V., Flagan, R. C., and Seinfeld, J. H.: Gas-phase products and secondary aerosol yields from the photooxidation of 16 different terpenes, J. Geophys. Res.-Atmos., 111, D17305, https://doi.org/10.1029/2006JD007050, 2006b.
- Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., Kurtén, T., Worsnop, D. R., and Thornton, J. A.: An Iodide-Adduct High-Resolution Time-of-Flight Chemical-Ionization Mass Spectrometer: Application to Atmospheric Inorganic and Organic Compounds, Environ. Sci. Technol., 48, 6309–6317, https://doi.org/10.1021/es500362a, 2014.
- Lee, B. H., Mohr, C., Lopez-Hilfiker, F. D., Lutz, A., Hallquist, M., Lee, L., Romer, P., Cohen, R. C., Iyer, S., Kurtén, T., Hu, W., Day, D. A., Campuzano-Jost, P., Jimenez, J. L., Xu, L., Ng, N. L., Guo, H., Weber, R. J., Wild, R. J., Brown, S. S., Koss, A., Gouw, J. d., Olson, K., Goldstein, A. H., Seco, R., Kim, S., McAvey, K., Shepson, P. B., Starn, T., Baumann, K., Edgerton, E. S., Liu, J., Shilling, J. E., Miller, D. O., Brune, W., Schobesberger, S., D'Ambro, E. L., and Thornton, J. A.: Highly functionalized organic nitrates in the southeast United States: Contribution to secondary organic aerosol and reactive nitrogen budgets, P. Natl. Acad. Sci. USA, 113, 1516–1521, https://doi.org/10.1073/pnas.1508108113, 2016.
- Lee, S. and Kamens, R. M.: Particle nucleation from the reaction of α -pinene and O₃, Atmos. Environ., 39, 6822–6832, 2005.
- Li, S.-J., Yuan, X.-Y., Li, Q., and Feng, Z.-Z.: Inventory and Characteristics of Biogenic Volatile Organic Compounds (BVOCs) for 12 Deciduous Fruit Trees, Huan jing ke xue = Huanjing kexue, 40, 2078–2085, 2019.

- Li, Y. and Shiraiwa, M.: Timescales of secondary organic aerosols to reach equilibrium at various temperatures and relative humidities, Atmos. Chem. Phys., 19, 5959–5971, https://doi.org/10.5194/acp-19-5959-2019, 2019.
- Li, Y. J., Chen, Q., Guzman, M. I., Chan, C. K., and Martin, S. T.: Second-generation products contribute substantially to the particle-phase organic material produced by β caryophyllene ozonolysis, Atmos. Chem. Phys., 11, 121–132, https://doi.org/10.5194/acp-11-121-2011, 2011.
- Liu, P. S. K., Deng, R., Smith, K. A., Williams, L. R., Jayne, J. T., Canagaratna, M. R., Moore, K., Onasch, T. B., Worsnop, D. R., and Deshler, T.: Transmission Efficiency of an Aerodynamic Focusing Lens System: Comparison of Model Calculations and Laboratory Measurements for the Aerodyne Aerosol Mass Spectrometer, Aerosol Sci. Technol., 41, 721–733, https://doi.org/10.1080/02786820701422278, 2007.
- Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, Th. F., Lutz, A., Hallquist, M., Worsnop, D., and Thornton, J. A.: A novel method for online analysis of gas and particle composition: description and evaluation of a Filter Inlet for Gases and AEROsols (FIGAERO), Atmos. Meas. Tech., 7, 983–1001, https://doi.org/10.5194/amt-7-983-2014, 2014.
- Lopez-Hilfiker, F. D., Iyer, S., Mohr, C., Lee, B. H., D'Ambro, E. L., Kurtén, T., and Thornton, J. A.: Constraining the sensitivity of iodide adduct chemical ionization mass spectrometry to multifunctional organic molecules using the collision limit and thermodynamic stability of iodide ion adducts, Atmos. Meas. Tech., 9, 1505–1512, https://doi.org/10.5194/amt-9-1505-2016, 2016.
- Maclean, A. M., Smith, N. R., Li, Y., Huang, Y., Hettiyadura, A. P. S., Crescenzo, G. V., Shiraiwa, M., Laskin, A., Nizkorodov, S. A., and Bertram, A. K.: Humidity-Dependent Viscosity of Secondary Organic Aerosol from Ozonolysis of β -Caryophyllene: Measurements, Predictions, and Implications, ACS Earth Space Chem., 5, 305–318, https://doi.org/10.1021/acsearthspacechem.0c00296, 2021.
- Matsunaga, S. N., Niwa, S., Mochizuki, T., Tani, A., Kusumoto, D., Utsumi, Y., Enoki, T., and Hiura, T.: Seasonal variation in basal emission rates and composition of mono-and sesquiterpenes emitted from dominant conifers in Japan, Atmos. Environ., 69, 124–130, 2013.
- Mellouki, A., Wallington, T., and Chen, J.: Atmospheric chemistry of oxygenated volatile organic compounds: impacts on air quality and climate, Chem. Rev., 115, 3984–4014, 2015.
- Mellouki, A., Ammann, M., Cox, R. A., Crowley, J. N., Herrmann, H., Jenkin, M. E., McNeill, V. F., Troe, J., and Wallington, T. J.: Evaluated kinetic and photochemical data for atmospheric chemistry: volume VIII – gas-phase reactions of organic species with four, or more, carbon atoms ($\geq C_4$), Atmos. Chem. Phys., 21, 4797–4808, https://doi.org/10.5194/acp-21-4797-2021, 2021.
- Möhler, O., Stetzer, O., Schaefers, S., Linke, C., Schnaiter, M., Tiede, R., Saathoff, H., Krämer, M., Mangold, A., Budz, P., Zink, P., Schreiner, J., Mauersberger, K., Haag, W., Kärcher, B., and Schurath, U.: Experimental investigation of homogeneous freezing of sulphuric acid particles in the aerosol chamber AIDA, Atmos. Chem. Phys., 3, 211–223, https://doi.org/10.5194/acp-3-211-2003, 2003.
- Müller, L., Reinnig, M.-C., Warnke, J., and Hoffmann, Th.: Unambiguous identification of esters as oligomers in secondary organic aerosol formed from cyclohexene and cyclohexene/ α -

pinene ozonolysis, Atmos. Chem. Phys., 8, 1423–1433, https://doi.org/10.5194/acp-8-1423-2008, 2008.

- Müller, M., Eichler, P., D'Anna, B., Tan, W., and Wisthaler, A.: Direct Sampling and Analysis of Atmospheric Particulate Organic Matter by Proton-Transfer-Reaction Mass Spectrometry, Anal. Chem., 89, 10889–10897, https://doi.org/10.1021/acs.analchem.7b02582, 2017.
- Naumann, K.-H.: COSIMA a computer program simulating the dynamics of fractal aerosols, J. Aerosol Sci., 34, 1371–1397, https://doi.org/10.1016/S0021-8502(03)00367-7, 2003.
- Ng, N. L., Chhabra, P. S., Chan, A. W. H., Surratt, J. D., Kroll, J. H., Kwan, A. J., McCabe, D. C., Wennberg, P. O., Sorooshian, A., Murphy, S. M., Dalleska, N. F., Flagan, R. C., and Seinfeld, J. H.: Effect of NO_x level on secondary organic aerosol (SOA) formation from the photooxidation of terpenes, Atmos. Chem. Phys., 7, 5159–5174, https://doi.org/10.5194/acp-7-5159-2007, 2007.
- Nguyen, T. L., Winterhalter, R., Moortgat, G., Kanawati, B., Peeters, J., and Vereecken, L.: The gas-phase ozonolysis of β caryophyllene (C 15 H 24). Part II: A theoretical study, Phys. Chem. Chem. Phys., 11, 4173–4183, 2009.
- Odum, J. R., Hoffmann, T., Bowman, F., Collins, D., Flagan, R. C., and Seinfeld, J. H.: Gas/Particle Partitioning and Secondary Organic Aerosol Yields, Environ. Sci. Technol., 30, 2580–2585, https://doi.org/10.1021/es950943+, 1996.
- Parshintsev, J., Nurmi, J., Kilpeläinen, I., Hartonen, K., Kulmala, M., and Riekkola, M.-L.: Preparation of β-caryophyllene oxidation products and their determination in ambient aerosol samples, Anal. Bioanal. Chem., 390, 913–919, 2008.
- Peng, Z. and Jimenez, J. L.: KinSim: A Research-Grade, User-Friendly, Visual Kinetics Simulator for Chemical-Kinetics and Environmental-Chemistry Teaching, J. Chem. Educ., 96, 806– 811, https://doi.org/10.1021/acs.jchemed.9b00033, 2019.
- Piel, F., Müller, M., Winkler, K., Skytte af Sätra, J., and Wisthaler, A.: Introducing the extended volatility range protontransfer-reaction mass spectrometer (EVR PTR-MS), Atmos. Meas. Tech., 14, 1355–1363, https://doi.org/10.5194/amt-14-1355-2021, 2021.
- Richters, S., Herrmann, H., and Berndt, T.: Gas-phase rate coefficients of the reaction of ozone with four sesquiterpenes at 295±2 K, Phys. Chem. Chem. Phys., 17, 11658–11669, https://doi.org/10.1039/C4CP05542J, 2015.
- Richters, S., Herrmann, H., and Berndt, T.: Different pathways of the formation of highly oxidized multifunctional organic compounds (HOMs) from the gas-phase ozonolysis of β -caryophyllene, Atmos. Chem. Phys., 16, 9831–9845, https://doi.org/10.5194/acp-16-9831-2016, 2016.
- Riva, M., Rantala, P., Krechmer, J. E., Peräkylä, O., Zhang, Y., Heikkinen, L., Garmash, O., Yan, C., Kulmala, M., Worsnop, D., and Ehn, M.: Evaluating the performance of five different chemical ionization techniques for detecting gaseous oxygenated organic species, Atmos. Meas. Tech., 12, 2403–2421, https://doi.org/10.5194/amt-12-2403-2019, 2019.
- Rodriguez-Saona, C., Crafts-Brandner, S. J., ParÉ, P. W., and Henneberry, T. J.: Exogenous methyl jasmonate induces volatile emissions in cotton plants, J. Chem. Ecol., 27, 679–695, https://doi.org/10.1023/A:1010393700918, 2001.
- Saathoff, H., Naumann, K.-H., Möhler, O., Jonsson, Å. M., Hallquist, M., Kiendler-Scharr, A., Mentel, Th. F., Tillmann, R., and Schurath, U.: Temperature dependence of yields of secondary or-

ganic aerosols from the ozonolysis of α -pinene and limonene, Atmos. Chem. Phys., 9, 1551–1577, https://doi.org/10.5194/acp-9-1551-2009, 2009.

- Shrivastava, M., Andreae, M. O., Artaxo, P., Barbosa, H. M. J., Berg, L. K., Brito, J., Ching, J., Easter, R. C., Fan, J., Fast, J. D., Feng, Z., Fuentes, J. D., Glasius, M., Goldstein, A. H., Alves, E. G., Gomes, H., Gu, D., Guenther, A., Jathar, S. H., Kim, S., Liu, Y., Lou, S., Martin, S. T., McNeill, V. F., Medeiros, A., de Sá, S. S., Shilling, J. E., Springston, S. R., Souza, R. A. F., Thornton, J. A., Isaacman-VanWertz, G., Yee, L. D., Ynoue, R., Zaveri, R. A., Zelenyuk, A., and Zhao, C.: Urban pollution greatly enhances formation of natural aerosols over the Amazon rainforest, Nature Commun., 10, 1046, https://doi.org/10.1038/s41467-019-08909-4, 2019.
- Shu, Y. and Atkinson, R.: Rate constants for the gas-phase reactions of O_3 with a series of terpenes and OH radical formation from the O_3 reactions with sesquiterpenes at 296 ± 2 K, Int. J. Chem. Kinetics, 26, 1193–1205, 1994.
- Shu, Y. and Atkinson, R.: Atmospheric lifetimes and fates of a series of sesquiterpenes, J. Geophys. Res.-Atmos., 100, 7275– 7281, https://doi.org/10.1029/95JD00368, 1995.
- Stolzenburg, D., Fischer, L., Vogel, A. L., Heinritzi, M., Schervish, M., Simon, M., Wagner, A. C., Dada, L., Ahonen, L. R., Amorim, A., Baccarini, A., Bauer, P. S., Baumgartner, B., Bergen, A., Bianchi, F., Breitenlechner, M., Brilke, S., Buenrostro Mazon, S., Chen, D., Dias, A., Draper, D. C., Duplissy, J., El Haddad, I., Finkenzeller, H., Frege, C., Fuchs, C., Garmash, O., Gordon, H., He, X., Helm, J., Hofbauer, V., Hoyle, C. R., Kim, C., Kirkby, J., Kontkanen, J., Kürten, A., Lampilahti, J., Lawler, M., Lehtipalo, K., Leiminger, M., Mai, H., Mathot, S., Mentler, B., Molteni, U., Nie, W., Nieminen, T., Nowak, J. B., Ojdanic, A., Onnela, A., Passananti, M., Petäjä, T., Quéléver, L. L. J., Rissanen, M. P., Sarnela, N., Schallhart, S., Tauber, C., Tomé, A., Wagner, R., Wang, M., Weitz, L., Wimmer, D., Xiao, M., Yan, C., Ye, P., Zha, Q., Baltensperger, U., Curtius, J., Dommen, J., Flagan, R. C., Kulmala, M., Smith, J. N., Worsnop, D. R., Hansel, A., Donahue, N. M., and Winkler, P. M.: Rapid growth of organic aerosol nanoparticles over a wide tropospheric temperature range, P. Natl. Acad. Sci. USA, 15, 9122-9127, https://doi.org/10.1073/pnas.1807604115, 2018.
- Tarvainen, V., Hakola, H., Hellén, H., Bäck, J., Hari, P., and Kulmala, M.: Temperature and light dependence of the VOC emissions of Scots pine, Atmos. Chem. Phys., 5, 989–998, https://doi.org/10.5194/acp-5-989-2005, 2005.
- Tasoglou, A. and Pandis, S. N.: Formation and chemical aging of secondary organic aerosol during the β caryophyllene oxidation, Atmos. Chem. Phys., 15, 6035–6046, https://doi.org/10.5194/acp-15-6035-2015, 2015.
- van Eijck, A., Opatz, T., Taraborrelli, D., Sander, R., and Hoffmann, T.: New tracer compounds for secondary organic aerosol formation from β-caryophyllene oxidation, Atmos. Environ., 80, 122– 130, 2013.
- Verma, S. K., Kawamura, K., Deshmukh, D. K., Haque, M. M., and Pavuluri, C. M.: Seasonal Characteristics of Biogenic Secondary Organic Aerosols Over Chichijima Island in the Western North Pacific: Impact of Biomass Burning Activity in East Asia, J. Geophys. Res.-Atmos., 126, e2020JD032987, https://doi.org/10.1029/2020JD032987, 2021.

- von Hessberg, C., von Hessberg, P., Pöschl, U., Bilde, M., Nielsen, O. J., and Moortgat, G. K.: Temperature and humidity dependence of secondary organic aerosol yield from the ozonolysis of β -pinene, Atmos. Chem. Phys., 9, 3583–3599, https://doi.org/10.5194/acp-9-3583-2009, 2009.
- Wagner, R., Bunz, H., Linke, C., Möhler, O., Naumann, K.-H., Saathoff, H., Schnaiter, M., and Schurath, U.: Chamber, in: Environmental Simulation Chambers: Application to Atmospheric Chemical Processes, edited by: Barnes, I. and Rudzinski, K. J., Nato Science Series: IV: Earth and Environmental Science, vol 62, Springer, Dordrecht, https://doi.org/10.1007/1-4020-4232-9_5, 2006.
- Wang, M., Yao, L., Zheng, J., Wang, X., Chen, J., Yang, X., Worsnop, D. R., Donahue, N. M., and Wang, L.: Reactions of atmospheric particulate stabilized Criegee intermediates lead to high-molecular-weight aerosol components, Environ. Sci. Technol., 50, 5702–5710, 2016.
- Weikl, F., Ghirardo, A., Schnitzler, J.-P., and Pritsch, K.: Sesquiterpene emissions from Alternaria alternata and Fusarium oxysporum: effects of age, nutrient availability, and co-cultivation, Sci. Rep., 6, 22152, https://doi.org/10.1038/srep22152, 2016.
- Williams, L. R., Gonzalez, L. A., Peck, J., Trimborn, D., McInnis, J., Farrar, M. R., Moore, K. D., Jayne, J. T., Robinson, W. A., Lewis, D. K., Onasch, T. B., Canagaratna, M. R., Trimborn, A., Timko, M. T., Magoon, G., Deng, R., Tang, D., de la Rosa Blanco, E., Prévôt, A. S. H., Smith, K. A., and Worsnop, D. R.: Characterization of an aerodynamic lens for transmitting particles greater than 1 micrometer in diameter into the Aerodyne aerosol mass spectrometer, Atmos. Meas. Tech., 6, 3271–3280, https://doi.org/10.5194/amt-6-3271-2013, 2013.
- Winterhalter, R., Herrmann, F., Kanawati, B., Nguyen, T. L., Peeters, J., Vereecken, L., and Moortgat, G. K.: The gas-phase ozonolysis of β -caryophyllene (C 15 H 24). Part I: an experimental study, Phys. Chem. Chem. Phys., 11, 4152–4172, 2009.
- Witkowski, B. and Gierczak, T.: Early stage composition of SOA produced by α-pinene/ozone reaction: α-Acyloxyhydroperoxy aldehydes and acidic dimers, Atmos. Environ., 95, 59–70, 2014.
- Wu, C., Bell, D. M., Graham, E. L., Haslett, S., Riipinen, I., Baltensperger, U., Bertrand, A., Giannoukos, S., Schoonbaert, J., El Haddad, I., Prevot, A. S. H., Huang, W., and Mohr, C.: Photolytically induced changes in composition and volatility of biogenic secondary organic aerosol from nitrate radical oxidation during night-to-day transition, Atmos. Chem. Phys., 21, 14907–14925, https://doi.org/10.5194/acp-21-14907-2021, 2021.
- Yasmeen, F., Vermeylen, R., Szmigielski, R., Iinuma, Y., Böge, O., Herrmann, H., Maenhaut, W., and Claeys, M.: Terpenylic acid and related compounds: precursors for dimers in secondary organic aerosol from the ozonolysis of α - and β -pinene, Atmos. Chem. Phys., 10, 9383–9392, https://doi.org/10.5194/acp-10-9383-2010, 2010.
- Ye, Q., Wang, M., Hofbauer, V., Stolzenburg, D., Chen, D., Schervish, M., Vogel, A., Mauldin, R. L., Baalbaki, R., Brilke, S., Dada, L., Dias, A., Duplissy, J., El Haddad, I., Finkenzeller, H., Fischer, L., He, X., Kim, C., Kürten, A., Lamkaddam, H., Lee, C. P., Lehtipalo, K., Leiminger, M., Manninen, H. E., Marten, R., Mentler, B., Partoll, E., Petäjä, T., Rissanen, M., Schobesberger, S., Schuchmann, S., Simon, M., Tham, Y. J., Vazquez-Pufleau, M., Wagner, A. C., Wang, Y., Wu, Y., Xiao, M., Baltensperger, U., Curtius, J., Flagan, R., Kirkby, J., Kulmala, M., Volkamer,

R., Winkler, P. M., Worsnop, D., and Donahue, N. M.: Molecular Composition and Volatility of Nucleated Particles from α -Pinene Oxidation between 50 °C and +25 °C, Environ. Sci. Technol., 53, 12357–12365, https://doi.org/10.1021/acs.est.9b03265, 2019.

Zhao, Y., Zhang, R., Sun, X., He, M., Wang, H., Zhang, Q., and Ru, M.: Theoretical study on mechanism for O3-initiated atmospheric oxidation reaction of β -caryophyllene, J. Molecular Structure: THEOCHEM, 947, 68–75, 2010.