



Supplement of

Parameterizations of size distribution and refractive index of biomass burning organic aerosol with black carbon content

Biao Luo et al.

Correspondence to: Ye Kuang (kuangye@jnu.edu.cn) and Shan Huang (shanhuang_eci@jnu.edu.cn)

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28 1. Methods



29 1.1 Determination of PMF factors from SP-AMS measurements

Figure S1. Diagnostics plots for PMF analysis of the SP-AMS OA mass spectra.

The source apportionment of organic aerosols (OA) was performed by positive matrix factorization (PMF) based on high resolution OA data from SP-AMS. The input data were collected in AMS V-mode using only tungsten vaporizer. PMF has been described in many other papers (Paatero, 1997; Paatero and Tapper, 1994) and has been used for AMS-based aerosol source apportionment as a common method (e.g. Huang *et al.*, 2018; Lanz *et al.*, 2007; Ng *et al.*, 2011; Sun *et al.*, 2012; Ulbrich *et al.*, 2009). In this study, high resolution AMS data including organic ion mass concentrations and their uncertainties were input to the PMF. An Igor Pro-based panel, i.e., PMF Evaluation Tool (PET,

v2.06, Ulbrich et al., 2009) were used to perform PMF analysis. Before the calculation, isotopes and 37 ions with m/z > 120 were excluded to minimize the interference from repeatedly calculated 38 uncertainties and noise signals. In total, 454 ions were considered in PMF. After investigating different 39 solutions with factor number from 2 to 10 with fPeak varying between -1 and 1, a six-factor solution 40 was selected based on the best performance shown by PMF quality parameters and most reasonable 41 source identification. In the five-factor solution, two primary OA factors were mixed, while the seven 42 and more factor solutions make factors split over. More parameters of the model performance could 43 44 be found in the diagnostic plots as shown in Fig.S1.

The mass spectra, time series and diurnal variations of factors are illustrated in Fig.S2. Overall, the total OA measured in the Heshan campaign were apportioned to six sources or processes, including two primary sources : 1) a hydrocarbon-like OA (HOA, containing cooking emissions), 2) a biomass burning OA (BBOA); and four secondary sources: 3) more oxygenated OA (MOOA, regional transport), 4) less oxygenated OA (LOOA, related to daytime photochemical formation), 5) nighttimeformed OA (Night-OA), and 6) aged BBOA (aBBOA).

As the key organic aerosols in the present study, BBOA took 16.5% of the total OA mass loading. 51 52 BBOA was identified by good correlation with the biomass burning emission tracer acetonitrile which measured by the parallel gaseous pollutant measurements. The correlation coefficient (R=0.60) is 53 higher than those between other OA factors and acetonitrile. The BBOA factor was also recognized by 54 pronounced intensity of ions $C_2H_4O_2^+$ (m/z 60.022, 0.5%) and $C_3H_5O_2^+$ (m/z 73.029, 0.4%) which 55 were also tracers for biomass burning OA (Lanz et al., 2007; Ng et al., 2011). The diurnal pattern of 56 BBOA showed a remarkable peak in the early evening (~19:00), corresponding to the local dinner 57 58 period during which the people in the surrounding valleys often used biomass materials for cooking or other residential use (the smokes were observed as shown in Fig.S7). The other primary factor, HOA, 59 60 was characterized by abundant hydrocarbon ions $C_xH_y^+$, such as $C_4H_7^+$ (m/z 55.055, 8.3%) and $C_4H_9^+$ (m/z 57.070, 9.5%). Because of the remarkable hydrocarbon ions and the low oxidation state (O:C 61 =0.39), HOA was attributed to primary emissions such as vehicle exhausts; while in this study, cooking 62 emissions cannot be fully excluded since the cooking emission tracer $C_6H_{10}O^+$ (m/z 98.073, 0.3%) was 63 64 mainly distributed in the HOA factor. However, Heshan supersite is far away from the main roads and cooking is not allowed inside of the site, thus the HOA factor accounted for only 8.7% of the total OA 65 and the diurnal variation of HOA showed no morning/evening peaks as the typical pattern of vehicle 66

emissions at urban site. Regarding the OA factors related to the secondary processes, MOOA presented the highest oxidation state (O:C ratio = 1.0) with highest fraction of CO_2^+ ($f_{CO2^+} = 18\%$) among all factors. The negligible diurnal variation of MOOA also suggested that aerosols were probably aged during the long-range transport rather than local formed via photo-oxidation. In contrast, LOOA was recognized as local formed secondary OA via photochemical process, traced by prominent ion $C_2H_3O^+$ (m/z 43.018, 10.3%) and clear afternoon-peak in the diurnal pattern. With a similar



Figure S2. (a) Mass spectral profile in family groups and (b) time series of PMF OA components. Also exhibited are concentration variations of tracer compounds on the right axes.

73 afternoon-peak, aBBOA factor was attributed to the quick aging process of biomass burning emissions. Note that gaseous pollutants could also be released during biomass burning and then be oxidized to 74 form organic aerosols. As shown in Fig.S2, the aBBOA factor is positively correlated with C₆H₂NO₄⁺ 75 $(m/z \ 151.998, R = 0.31)$, which is a typical fragment of the aged BBOA component nitrocatechol 76 (Bertrand et al., 2018). Therefore, we speculate this factor was related to the aged biomass burning 77 emissions. In Heshan site, a night-formed OA factor (Night-OA) was resolved with a pronounced 78 evening elevation in the diurnal circle. Night-OA is correlated with nitrate (R = 0.67) which was likely 79 associated with NO₃ nighttime chemistry as discussed previously (Kuang et al., 2021). 80

81 **1.2** The modelling method of $AAE_{BC,\lambda-880}$

82 The absorption coefficient $\sigma_{abs,Mie,\lambda}$ was obtained by the following:

83
$$\sigma_{abs,Mie,\lambda} = \int Q_{abs,Mie,\lambda}(Dp) * \left(\frac{\pi}{4}D_p^2\right) N(logD_p) * d_{logD_p}(1)$$

where $N(logD_p)$ is the PNSD function, $Q_{abs,Mie,\lambda}$ is absorption efficiency which is simulated using the Mie-theory, D_p is the particle. For D_p bin, particles are classified into three types: non-lightabsorbing particles (Non-BC), BC-containing particles and pure BC particles. Parameters r_{ext} defined as mass fraction of pure externally mixed BC (M_{ext}) to total BC mass (M_{BC}) in different diameter bins and R_NBC defined as the number fraction of particles that does not contain BC are used to represent BC mixing state. The PNSD of Non-BC particles ($N(logD_p)_{Non-BC}$) and BC-containing particles ($N(logD_p)_{BC}$) can be given by the following equations:

91
$$N(log D_p)_{Non-BC} = N(log D_p)_{measure} * R_NBC$$
 (2)

- 92 $N(logD_p)_{BC} = N(logD_p)_{measure} * (1 R_NBC) (3)$
- Where R_NBC is the number fraction of BC-free particles, $N(logD_p)_{measure}$ is the PNSD measured by the SMPS and APS. The PNSD of pure BC particles $(N(logD_p)_{ext})$ and core-shell mixed BC particles $(N(logD_p)_{core-shell})$ can be given by the following equations:

96
$$r_{ext} = \frac{M_{ext}}{M_{BC}}$$
 (4)

97
$$N(logD_p)_{ext} = N(logD_p)_{BC} * f_{ext}(Dp)$$
(5)

98
$$N(logD_p)_{core-shell} = N(logD_p)_{BC} * (1 - f_{ext}(Dp)) (6)$$

99 $f_{ext}(Dp)$ is volume fraction of pure BC at each diameter bin which can be calculated by :

100
$$f_{ext}(Dp) = \frac{M_{BC}(Dp) * r_{ext}}{\rho_{BC} * N(log D_p)_{BC} * (\frac{\pi}{6} D_p^3)} (7)$$

where ρ_{BC} is the density of BC (1.5 g cm⁻³)(Ma *et al.*, 2012); M_{BC} is derived from AE33. Size distribution of BC mass $M_{BC}(Dp)$ is calculated based on the normalized Cx fraction $f_{Cx}(Dp)$ at each diameter bin:

104
$$M_{BC}(Dp) = M_{BC} \times f_{Cx}(Dp)$$
(8)

In addition to the PNSD of the three types of particles mentioned above, another key parameter of

106 the core-shell model is the diameter of the BC core, at each diameter bin the D_{core} is calculated as:

107
$$D_{core} = \left(\frac{6 \times M_{BC}(Dp) * (1 - r_{ext})}{\rho_{BC} * \pi}\right)^{1/3}$$
 (9)

108 After obtaining the absorption coefficient at each wavelength, the AAE at two wavelengths is 109 calculated as the following equation:

110
$$AAE_{\lambda 1-\lambda 2} = \frac{\ln(\sigma_{abs,\lambda 1}) - \ln(\sigma_{abs,\lambda 2})}{\ln(\lambda_1) - \ln(\lambda_2)}$$
(10)

111 When retrieving BrC absorptions, calculations of $R_{AAE}(\lambda)$ were achieved through assuming r_{ext} and R_NBC values 112 of 0.5 and 0.5 due to small influences of r_{ext} and R_NBC variations on $R_{AAE}(\lambda)$.

113 **1.3 Mass Scattering Efficiency calculations for different aerosol components.**

114 The size distributions of ammonium (AN) nitrate and ammonium sulfate (AS) paired from measured size 115 distributions of nitrate, ammonium and sulfate by the SP-AMS under different non-refractory mass concentrations of 116 PM_1 (NR_PM_1, PM_1 corresponds to aerosols with aerodynamic diameter less than 1 μ m) is shown in Fig.S3a and 117 Fig.S3b. Note that the D_m converted by assuming an aerosol density of 1.6 g/cm³ from the SP-AMS vacuum



Figure S3. Average size distributions of AN (a) and AS (b) under different NR_PM1 conditions, solid lines are measurement averages of SP-AMS measurements and dashed lines are corresponding lognormal fitting curves; (c) Simulated PM1 MSE of AS and AN under different NR_PM1 conditions and corrected ones

aerodynamic diameter Dva. The MSE_{AN,PM1} defined as $MSE_{PM1} = \frac{\sigma_{SP,PM1}}{AN_{PM1}}$ under different NR_PM1 conditions and results of MSE_{AS,PM1} are shown in Fig.S3c. The density of AS and AN are 1.769 and 1.72 g/cm³, and used refractive index of AS and AN is 1.53-10⁻⁷i. However, in this study, the aerosol scattering coefficient of PM₁₀ (aerosols with aerodynamic diameter less than 10 μ m) and the aerosol mass concentration of PM₁ is measured by the SP-AMS. Therefore, the corrected MSE defined as $MSE_{Corrected} = \frac{\sigma_{SP,PM10}}{AN_{PM1}} = MSE_{PM1} \times C$ is simulated and also shown in Fig.S3c, and C is the correction factor. The average $MSE_{Corrected}$ values for AS and AN are 4.6 and 4.8 m²/g, and used in this study.



Figure S4. (a) Average OA size distributions, and parts associated with HOA and BBOA, Dva is the vacuum aerodynamic diameter. **(b)** The average OA size distributions with contributions of HOA and BBOA subtracted, and the remaining can be well fitted using two lognormal modes.

The MSE_{HOA} is simulated on the basis of identified average HOA D_{gv} and σ_g of 180 nm and 1.46 using the Mie theory in combination with calculated HOA density of 1.15 g/cm3. The calculated MSE_{HOA} is 3.2 m²/g. The MSE_{BC} is simulated using the normalized Cx distribution shape shown in Fig.s4, and assuming an BC density of 1.5 g/cm3 and BC refractive index of 1.8-0.54i. The calculated MSE_{BC} is 2.8 m²/g.

129 The average size distributions of OA measured by the SP-AMS during the entire campaign is shown in Fig.S4a. 130 The average size distributions of HOA and BBOA calculated from average HOA and BBOA mass concentrations 131 together with identified average Dgv and σ_q of HOA and BBOA identified from SP-AMS measurements (175, 395 132 and 1.46, 1.55 for HOA and BBOA) are also shown in Fig.S4a. If the contributions of HOA and BBOA are subtracted 133 (Fig.S2b), the OA size distribution can be well fitted using two lognormal modes. The mass concentration of fitted 134 Mode 1 is consistent that of aBBOA and obviously different with mass concentrations of other remaining OA factors, 135 and mass concentration of fitted Mode 2 is consistent with the sum of LOOA, MOOA and Night-OA. Thus, the 136 Mode 1 is identified as the size distribution of aBBOA. The Mode 2 of LOOA, MOOA and Night-OA suggests that most secondary organic aerosols during this campaign are likely internally mixed. Thus the MSE_{aBBOA} is simulated 137 138 as 4.5 m2/g with identified size distribution of Mode 1 and calculated aBBOA density g/cm3 using the scheme 139 proposed by (Kuwata et al., 2012). Corrected MSE of MOOA, Night-OA, LOOA (MSE_{SOA}) is calculated as 6.3 140 m2/g with identified size distribution of Mode 2 and estimated average density of 1.31 g/cm3 using volume weighting 141 rule.

143

144 **1.4 Refractive index retrieval**

145 Determined BBOA MSEs and MAEs were converted to Volume scattering or absorption 146 efficiency (VSE or VAE) through VSE=MSE*density. With given geometric mean (Dgn) and standard 147 deviation (σ g) values of the PNSD. The assuming total number concentration of 1000 (Ntot=1000 148 /cm3), PNSD can be given as :

149
$$N(logDp) = \frac{dN}{dlogDp}(Dp) = \frac{Ntot}{\sqrt{2\pi}\log(\sigma_g)} \exp\left[-\frac{(\log(D_p) - \log(D_{gn}))^2}{2\log^2\sigma_g}\right]$$
(11)

The with given refractive index of m= $m_{R,BBOA} + m_{i,BBOA} \times i$, the aerosol scattering efficiency Q_{sca} and absorption efficiency Q_{abs} can be calculated using the Mie theory. And then scattering and absorption coefficients of bulk aerosols can be derived as:

153
$$\sigma_{sca}(\lambda) = \int_0^{D_p^{max}} Q_{sca}(m,\lambda,D_p) \times \frac{\pi}{4} D_p^2 \times N(logD_p) \times dlogD_p$$
(12)

154
$$\sigma_{abs}(\lambda) = \int_0^{D_p^{max}} Q_{abs}(m,\lambda,D_p) \times \frac{\pi}{4} D_p^2 \times N(logD_p) \times dlogD_p$$
(13)

155 Where λ is the optical wavelength, and D_p^{max} of 2500 nm is set. The total volume concentration can be 156 calculated as:

157
$$Vtot = \int_0^{D_p^{max}} \frac{\pi}{6} D_p^3 \times N(logD_p) \times dlogD_p$$
(14)

158 Then the VSE and MAE can be calculated as:

159
$$VSE(\lambda) = \frac{\sigma_{sca}(\lambda)}{Vtot}$$
 (15)

160
$$VAE(\lambda) = \frac{\sigma_{abs}(\lambda)}{Vtot}$$
 (16)

161 The $m_{R,BBOA was}$ retrieved through varying $m_{R,BBOA}$ in iterations to find a $m_{R,BBOA}$ with which the derived 162 VSE can be reproduced, in these iterations $m_{i,BBOA}$ was parameterized with the corresponding 163 $\Delta CO/\Delta BBOA$ using the relationships determined in Sect.3.4. The $m_{i,BBOA was}$ retrieved through varying 164 $m_{i,BBOA}$ in iterations to find a $m_{R,BBOA}$ with which the derived VAE can be reproduced, and $m_{R,BBOA}$ are 165 fixed as 1.6 due to that sensitivity tests show that very small influences of $m_{R,BBOA}$ variations on $m_{i,BBOA}$ 166 retrieval.

- 167
- 168

2. Other figures



Figure S5. Time series of SSA during the campaign.







Figure S6	Normalized	average	Cv distribution	measured h	v the SP_AMS
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Figure S7. Photoed at at the Heshan supersite during the observation period.





243 Figure S8. (a) Diurnal variations of BBOA and HOA; (b) The distribution of BBOA/HOA ratio.



Figure S9. (a) to (c) Time series of resolved OA factors by SP-AMS measurements, and (d) is the derived BrC absorption coefficients at 370 nm. Shaded areas represent identified spikes.









Figure S11. Average OA size distribution differences for spikes in Fig.3b, and the difference can be fitted using two lognormal modes, corresponding to HOA and BBOA respectively.



Figure S12. Multilinear fitting of BrC absorptions at 370 nm with OA factors



Figure S13. Variations of BBOA MSE under different Dgv, Dgn and σ_g conditions.

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