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### Sulfuric acid-amine nucleation in urban Beijing

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Abstract. New particle formation (NPF) is one of the major sources of atmospheric ultrafine particles. Due to the high aerosol and trace gas concentrations, the mechanism and governing factors for NPF in the polluted atmospheric boundary layer may be quite different from those in clean environments, which is however less understood. Herein, based on long-term atmospheric measurements from January 2018 to March 2019 in Beijing, the nucleation mechanism and the influences of H<sub>2</sub>SO<sub>4</sub> concentration, amine concentrations, and aerosol concentration on NPF are quantified. The collision of H<sub>2</sub>SO<sub>4</sub>-amine clusters is found to be the dominating mechanism to initialize NPF in urban Beijing. The coagulation scavenging due to the high aerosol concentration is a governing factor as it limits the concentration of H<sub>2</sub>SO<sub>4</sub>amine clusters and new particle formation rates. The formation of H<sub>2</sub>SO<sub>4</sub>-amine clusters in Beijing is sometimes limited by low amine concentrations. Summarizing the synergistic effects of H<sub>2</sub>SO<sub>4</sub> concentration, amine concentrations, and aerosol concentration, we elucidate the governing factors for H<sub>2</sub>SO<sub>4</sub>-amine nucleation for various conditions.

### 1 Introduction

New particle formation (NPF) is a major source of ambient particles in terms of particle number concentration. During a typical NPF process, gaseous precursors form stable clusters via nucleation. Some of these clusters survive and grow into cloud condensation nuclei and hence have the potential to influence the global climate (Kuang et al., 2009; Gordon et al., 2017). There has been a considerable number of NPF studies in various atmospheric environments (Kulmala et al., 2004; Kerminen et al., 2018; Lee et al., 2019), but current knowledge on NPF in the polluted atmospheric boundary layer (e.g., the urban environment in megacities) is still limited. In the presence of a high aerosol concentration in the polluted environment, a considerable fraction of the newly formed clusters and particles are scavenged by coagulation within minutes; hence, NPF may be significantly suppressed (McMurry et al., 2005; Kuang et al., 2010). However, frequent NPF events with high formation rates have been reported in polluted environments (Wu et al., 2007; Iida et al., 2008; Wang et al., 2015; Cai and Jiang, 2017; Yao et al., 2018; Deng et al., 2020b). Such unique characteristics

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of NPF with a high aerosol concentration, moderate gaseous precursor (e.g., H<sub>2</sub>SO<sub>4</sub>) concentrations, and a high particle formation rate indicate a fast nucleation mechanism in these environments. Various nucleation mechanisms for the real atmosphere have been reported, such as H<sub>2</sub>SO<sub>4</sub>-amine nucleation (Chen et al., 2012; Yao et al., 2018), H<sub>2</sub>SO<sub>4</sub>-NH<sub>3</sub> nucleation (Chen et al., 2012; Jokinen et al., 2018), oxidized organics nucleation (Bianchi et al., 2016), and HIO<sub>3</sub> nucleation (Sipilä et al., 2016). The second and third mechanisms are not efficient enough to explain the observed high particle formation rates in the polluted environment, and the last mechanism may dominate NPF only in the coastal regions.

The clustering of H<sub>2</sub>SO<sub>4</sub> and amines produces new particles at a high formation rate. Laboratory studies showed that some amines enhance H<sub>2</sub>SO<sub>4</sub> nucleation more efficiently than NH<sub>3</sub> (Kirkby et al., 2011; Erupe et al., 2011; Yu et al., 2012; Jen et al., 2014; Dunne et al., 2016; Yu et al., 2018). Atmospheric measurements in Boulder (Zhao et al., 2011) and urban Atlanta (Chen et al., 2012) indicate the H<sub>2</sub>SO<sub>4</sub>amine nucleation mechanism. The CLOUD (Cosmics Leaving OUtdoor Droplets) chamber experiments and theoretical calculations based on quantum chemistry reported that, in the presence of  $\sim 5$  ppt dimethylamine (DMA, (CH<sub>3</sub>)<sub>2</sub>NH) as the stabilizing base, H<sub>2</sub>SO<sub>4</sub> under a typical atmospheric concentration ( $\sim 10^6$ – $10^7$  molec. cm<sup>-3</sup>) nucleates at a rate approaching the kinetic collision limit where cluster evaporation is negligible (Almeida et al., 2013; Kürten et al., 2014, 2018). Atmospheric measurements in urban Shanghai (Yao et al., 2018) provide support for the view that in polluted megacities in China, H<sub>2</sub>SO<sub>4</sub> initiates NPF, and DMA is perhaps the dominating base to stabilize H<sub>2</sub>SO<sub>4</sub> clusters. Elucidating the governing factors for atmospheric nucleation and their quantitative impacts on particle formation rate is key to understanding the nucleation mechanism in the real atmosphere.

Previous laboratory and atmospheric studies provide important understanding of H<sub>2</sub>SO<sub>4</sub>–DMA nucleation (Chen et al., 2012; Almeida et al., 2013; Jen et al., 2014; Kürten et al., 2014, 2018). Due to the high aerosol concentration in the polluted atmospheric boundary layer, however, the loss rates of new particles during NPF in urban Beijing and Shanghai are usually  $\sim 10$  times higher than the total loss rate in laboratory chambers. Under such a high aerosol concentration, coagulation scavenging has been found to govern the concentrations of new particles (Cai et al., 2017b; Deng et al., 2020a). In addition, high DMA concentrations were used in the previous laboratory experiments, e.g., up to 200 ppt in Jen et al. (2014) and up to 140 ppt in the CLOUD experiments (Almeida et al., 2013; Kürten et al., 2014, 2018), whereas the DMA concentration in urban Beijing was observed in this study to be usually lower than 5 ppt. As also pointed out in previous studies (Kürten et al., 2014; Yao et al., 2018), the characteristics and limiting factors of nucleation in the polluted atmosphere may be different from those in laboratory experiments, due to differences in the particle loss rate and precursor gas concentrations. As a result, the molecular-scale understanding of  $\rm H_2SO_4$ –DMA nucleation under laboratory conditions (Jen et al., 2014; Kürten et al., 2014, 2018) may not be directly applicable for the real atmosphere with low DMA concentrations (< 5 ppt) and high aerosol concentrations.

For a better understanding of the nucleation mechanism in the polluted atmosphere, long-term atmospheric measurements were conducted in urban Beijing from January 2018 to March 2019. Gaseous  $H_2SO_4$  and cluster concentrations, amine concentrations, and particle size distributions ranging from 1 nm to 10  $\mu$ m were measured. The formation mechanism and the factors governing the initial steps of NPF in the polluted environment were explored. A model based on kinetic nucleation theory is shown to explain the measured concentrations of  $H_2SO_4$  dimer, trimer, and tetramer and the formation rate of 1.4 nm particles in urban Beijing. The roles of coagulation scavenging and amine concentrations in  $H_2SO_4$ —amine nucleation are revealed and quantified.

#### 2 Measurements

The atmospheric measurement campaign was conducted in urban Beijing. The dataset for this study covers the periods from January to April and from October 2018 to March 2019. The observation site is located at the campus of the Beijing University of Chemical Technology (39°56′ N, 116°17′ E). The West 3rd Ring Road is  $\sim 500 \,\mathrm{m}$  away from this observation site. State-of-the-art instruments were deployed to capture the whole NPF process from the very initial step of nucleation to particle growth. The aerosol size distributions ranging from 1 nm to 10 µm were measured using a diethylene glycol scanning mobility particle spectrometer (1-4.5 nm; Jiang et al., 2011; Cai et al., 2017a; Fu et al., 2019) and a particle size distribution system (3 nm-10 µm; Liu et al., 2016). The neutral gaseous H<sub>2</sub>SO<sub>4</sub> molecule and cluster concentrations were measured using chemical ionization time-of-flight mass spectrometers (ToF-CIMS, Aerodyne Research, Inc.; Jokinen et al., 2012). H<sub>2</sub>SO<sub>4</sub> molecules and neutral clusters are charged using a nitrate chemical ionization source. A long ToF-CIMS and a high-resolution ToF-CIMS were used before and after September 2018, respectively. These two instruments were calibrated separately and compared during a short-term parallel measurement. The H<sub>2</sub>SO<sub>4</sub> monomer concentrations reported by these two instruments agree with each other within a systematic relative difference of  $1.4 \pm 0.3$  (Fig. S1 in the Supplement). In addition to H<sub>2</sub>SO<sub>4</sub> molecules, H<sub>2</sub>SO<sub>4</sub> clusters and organics with low volatilities were measured using ToF-CIMS. Their calibration factors were assumed to be equal to that for H<sub>2</sub>SO<sub>4</sub> molecules. The mass dependency of transmission efficiency was calibrated and corrected using the method reported in Heinritzi et al. (2016). The amine molecule in a stable, neutral H<sub>2</sub>SO<sub>4</sub>-amine cluster and may detach during the detection using ToF-CIMS. Hence, the concentrations of H<sub>2</sub>SO<sub>4</sub> clusters containing the same number of H<sub>2</sub>SO<sub>4</sub> molecules were summed up. NH<sub>3</sub> was not found in the detected neutral H<sub>2</sub>SO<sub>4</sub> clusters. Neutral amine and NH<sub>3</sub> concentrations were measured using a modified ToF-CIMS (Aerodyne Research, Inc.; Zheng et al., 2015) since October 2018. The reagent ions to charge amines and NH<sub>3</sub> are H<sub>3</sub>O<sup>+</sup> or its hydrated clusters. The sampling line for this ToF-CIMS was  $\sim 1.5 \, \mathrm{m}$ long and it was heated to  $\sim 60$  °C. The temperature of the sampled air was < 40 °C. A sample flow rate of 6.1 L min<sup>-1</sup> was used to reduce aerosol deposition onto the tube wall. Because the ToF-CIMS cannot separate isomers, the measured C<sub>2</sub>-amine concentration is taken as DMA concentration. Ethylamine is thought to be less efficient as a stabilizing base for  $(H_2SO_4)_1$  (amine)<sub>1</sub> than DMA (Xie et al., 2017); thus, the measured effective DMA concentration for stabilizing H<sub>2</sub>SO<sub>4</sub> clusters might be overestimated. Similarly, the measured C<sub>3</sub>-amine concentration is taken as trimethylamine (TMA) concentration. The naturally charged negative clusters were measured using an atmospheric pressure interface time-of-flight mass spectrometer (APi-ToF-MS, Aerodyne Research, Inc.; Junninen et al., 2010). Ambient temperature, pressure, and relative humidity were measured using a weather station (AWS310, Vaisala Inc.).

When using the measured  $H_2SO_4$  monomer or n-mer (n=2,3,4) concentrations, the concentrations of clusters containing the same  $H_2SO_4$  molecules are summed up and written as  $[(H_2SO_4)_{n,\text{tot}}]$ , because amines may detach from  $H_2SO_4$ -amine clusters during the ionization process imposed by the instrument. For instance, an  $H_2SO_4$  dimer refers to a cluster containing two  $H_2SO_4$  molecules regardless of its base number.

The loss rates of gaseous precursors and clusters onto particles, i.e., condensation sinks (CSs) and coagulation sinks, respectively, were calculated using the measured aerosol size distributions (Kulmala et al., 2001). The reported CS was calculated for H<sub>2</sub>SO<sub>4</sub>. The coagulation sinks of clusters and particles are usually characterized by CS or the Fuchs surface area (McMurry et al., 2005). From the perspective of molecular kinetics, we do not distinguish between coagulation and condensation in this study. Hence, CS is used to characterize the condensation and coagulation scavenging effects of aerosols on gaseous precursors, clusters, and new particles. The formation rate of 1.4 nm (in electrical mobility diameter) particles,  $J_{1,4}$ , is calculated using a population balance formula (Cai and Jiang, 2017). This formula improves the estimation of particle coagulation scavenging compared to previous formulae.

The uncertainty of the measured aerosol size distributions was estimated to be  $\pm 10\%$  (Wiedensohler et al., 2012) and +100%/-50% (Kangasluoma et al., 2020) for particles larger and smaller than 10 nm, respectively. The CS in urban Beijing is mainly contributed by accumulation mode particles (Cai et al., 2017b); hence, the uncertainty of CS was estimated to be  $\pm 10\%$ . The formation rate is mainly deter-

mined by the product of new particle concentration and CS in urban Beijing (Cai and Jiang, 2017); hence, the uncertainty of the measured  $J_{1.4}$  was estimated to be +100%/-50%. The uncertainty of the measured  $H_2SO_4$  concentration is +100%/-50% according to Fig. S1 and previous studies (Kürten et al., 2012; Jokinen et al., 2012). The uncertainty of the measured amine concentrations is estimated to be similar to that of  $H_2SO_4$  concentration.

The occurrence of NPF was determined according to the evolution of the measured aerosol size distributions. A day was classified as an NPF day if a clear new particle formation and growth pattern was observed. If no NPF event occurs on a day, it was classified as a non-event day. The rest of the days, mainly with weak NPF events that are difficult to distinguish, were classified as undefined days. From January 2018 to March 2019, the frequencies of NPF days and undefined days are 35 % and 5 %, respectively.

#### 3 Model

A kinetic model was used to illustrate the nucleation process of H<sub>2</sub>SO<sub>4</sub>-amine clusters. Similar models have been reported in previous studies (Chen et al., 2012; McGrath et al., 2012; Jen et al., 2014). The cluster evaporation rate used in the model was based on quantum chemistry calculations but modified to fit the experimental data. The standard molar Gibbs free energy of formation of  $(H_2SO_4)_1(DMA)_1$ ,  $\Delta_{\rm f} G_{\rm m,A1B1}^{\theta}$  (298.15 K), was assumed to be -14.0 kcal mol<sup>-1</sup> in this study, which is in the range of values reported in previous studies (Ortega et al., 2012; Myllys et al., 2019). The quantum chemistry results for  $\Delta_{\rm f} G_{\rm m,A1B1}^{\theta}$  (298.15 K) using the  $\omega$ B97X-D/6-31++G\*\*, CBS-QB3, and RICC2B3 levels of theory were reported to be -13.5 (Myllys et al., 2019), -14.4, and -15.4 kcal mol<sup>-1</sup> (Ortega et al., 2012), respectively. The evaporation rate of (H<sub>2</sub>SO<sub>4</sub>)<sub>1</sub>(TMA)<sub>1</sub> is assumed to be 5 times that of  $(H_2SO_4)_1(DMA)_1$  according to previous experimental results (Jen et al., 2014). Since the measured TMA concentration was usually lower than or comparable to the measured DMA concentration in this study, the uncertainty of the evaporation rate of (H<sub>2</sub>SO<sub>4</sub>)<sub>1</sub>(TMA)<sub>1</sub> does not significantly affect the simulated particle formation rate or cluster concentrations. The value of free energy at different temperatures is calculated using Eq. (S15). The evaporation rate of a cluster was derived from its corresponding standard molar Gibbs free energy of formation (Ortega et al., 2012).

In the proposed kinetic model, the evaporation rates of other  $(H_2SO_4)_n(amine)_n$  clusters (n=2, 3, 4) are negligible compared to their coagulation sinks and hence not included. The very unstable clusters, e.g.,  $(H_2SO_4)_1(amine)_3$ , are omitted. Some studies indicate that other  $H_2SO_4$ -amine clusters, e.g.,  $(H_2SO_4)_3(amine)_2$ ,  $(H_2SO_4)_3(amine)_4$ , and  $(H_2SO_4)_4(amine)_3$ , and their corresponding reaction pathways may contribute to NPF (McGrath et al., 2012; Olenius et al., 2017), but these studies are not consistent with each

other due to the uncertainties in quantum chemistry calculations. Within the ranges of these predicted Gibbs free energies of formation, it is similarly arbitrary to assume negligible or high evaporation rates of these clusters. The kinetic model used in this study does not include the reaction pathways via these clusters (see the Supplement); i.e., it assumes infinitely high evaporation rates of these clusters. Ioninduced nucleation is neglected in this model, because it has a minor contribution to  $\rm H_2SO_4$ -amine nucleation for typical ambient conditions in the polluted environment (Yao et al., 2018).

The cluster concentrations and particle formation rates were simulated using the kinetic model. Similar to previous studies (Jen et al., 2014), we assume the A<sub>4</sub>B<sub>4</sub> formation rate as the simulated particle formation rate,  $J_{1,4}$ . Currently, knowledge of the exact size of H<sub>2</sub>SO<sub>4</sub>-amine clusters is still limited. Previous studies reported that the electrical mobility of [HSO<sub>4</sub>(H<sub>2</sub>SO<sub>4</sub>)<sub>3</sub>(DMA)<sub>3</sub>]<sup>-</sup> (Jen et al., 2015) and  $[HSO_4(H_2SO_4)_6(DMA)_4]^-$  (Thomas et al., 2016) is  $1.0 \times 10^{-4}$  and  $9.4 \times 10^{-5}$  m<sup>2</sup> V<sup>-1</sup> s<sup>-1</sup>, respectively. According to these values, the electrical mobility diameter of a  $(H_2SO_4)_4(DMA)_4$  cluster was estimated to be  $\sim 1.4$  nm, which is located within the measurement range of the aerosol size spectrometer. Its geometric diameter is estimated to be  $\sim 1.1$  nm according to the relationship between geometric diameter and electrical mobility diameter (Ku and de la Mora, 2009; Larriba et al., 2011).

The uncertainty of the model mainly comes from the uncertainty in the evaporation rate of  $(H_2SO_4)_1(amine)_1$ . We estimated this uncertainty range using one high and one low evaporation rate from the values reported in the literature (Ortega et al., 2012; Myllys et al., 2019). According to this estimation, the uncertainty of the model is of the same order of magnitude as the measurement uncertainties (Fig. S2).

#### 4 Results and discussion

## 4.1 Kinetic nucleation in the presence of a high aerosol concentration

During this measurement, H<sub>2</sub>SO<sub>4</sub>-amine nucleation was found to be the dominating nucleation mechanism in the polluted atmosphere in urban Beijing. This finding is supported by comparing the measured and simulated H<sub>2</sub>SO<sub>4</sub> cluster concentrations and new particle formation rates. The consistency between the measurement results and the H<sub>2</sub>SO<sub>4</sub>-amine nucleation mechanism is shown and discussed below. Meanwhile, other nucleation mechanisms, e.g., H<sub>2</sub>SO<sub>4</sub>-NH<sub>3</sub> nucleation and organics nucleation, were found to be not sufficient to explain the observed high new particle formation rate under the high coagulation sink (see Sect. 4.3 and the Supplement).

As shown in Fig. 1, the  $H_2SO_4$  dimer concentration was simulated at the median values of CS (0.017 s<sup>-1</sup>),  $C_2$ -amine

concentration (1.8 ppt), and temperature (281 K). Good consistency ( $R^2 = 0.75$ ) was observed between the measured and simulated H<sub>2</sub>SO<sub>4</sub> dimer concentrations. The measured H<sub>2</sub>SO<sub>4</sub> trimer and tetramer concentrations provide further evidence for the kinetic nucleation mechanism of H<sub>2</sub>SO<sub>4</sub>amine clusters in urban Beijing. Figure 2 shows that the measured H<sub>2</sub>SO<sub>4</sub> dimer, trimer, and tetramer concentrations are in accordance with their corresponding simulated concentrations when considering the uncertainties in determining the detection efficiencies of H<sub>2</sub>SO<sub>4</sub> trimer and tetramer concentrations. The systematic difference that the measured H<sub>2</sub>SO<sub>4</sub> trimer and tetramer concentrations are lower than the simulated concentrations is presumably caused by measurement uncertainties, e.g., cluster fragmentation (Zapadinsky et al., 2019). In a CLOUD chamber study on kinetic H<sub>2</sub>SO<sub>4</sub>–DMA nucleation (Kürten et al., 2014) with a high DMA concentration (5-32 ppt), such differences between simulated and measured H<sub>2</sub>SO<sub>4</sub> n-mer concentrations were also observed. If the H<sub>2</sub>SO<sub>4</sub> *n*-mer concentrations were overestimated in the kinetic model, the measured particle formation rate should also be lower than the simulated rate, which is inconsistent with the results shown in Fig. 3.

Besides the concentrations of H<sub>2</sub>SO<sub>4</sub> clusters, there is also a consistency between the measured and simulated formation rate of 1.4 nm (electrical mobility diameter) particles (Figs. 3 and S3). This consistency indicates that the clustering of H<sub>2</sub>SO<sub>4</sub> and amine is the governing mechanism for nucleation and the initial growth of new particles up to 1.4 nm in urban Beijing. To compare the formation rates measured at different CSs, the measured  $J_{1.4}$  values in Fig. 3a were scaled to the median CS (0.017 s<sup>-1</sup>) during the observed NPF events in this study. The scaling formula is  $J_{1.4,scaled} =$  $J_{1.4} \times (\text{CS}/0.017\,\text{s}^{-1})^2$ , which is based on the fact that under the high CS,  $J_{1.4}$  is theoretically inversely proportional to  $CS^2$ . Similarly, the measured  $J_{1.4}$  in Fig. 3b were scaled by  $J_{1.4,\text{scaled}} = J_{1.4} \times ((3.5 \times 10^6 \text{ cm}^{-3})/[(\text{H}_2\text{SO}_4)_{1,\text{tot}}])^4$  to compare the formation rates measured at different H<sub>2</sub>SO<sub>4</sub> monomer concentrations. The derivations for these scaling formulae are detailed in the Supplement. The dependency of measured  $J_{1,4}$  on  $H_2SO_4$  monomer concentrations in Fig. 3a provides support for the scaling in Fig. 3b and vice versa.

The measured particle formation rate is then compared to previous studies. The CLOUD study reported that particle formation rate for  $H_2SO_4$ –DMA nucleation (red curve in Fig. 3) was obtained at a high DMA concentration (5–32 ppt), with a low cluster loss rate (Kürten et al., 2014, 2018). The wall loss and dilution rates in that study sum up to be  $\sim 0.002 \, {\rm s}^{-1}$ . The particle formation rate under the same  $H_2SO_4$  monomer concentration measured in these CLOUD experiments deviates from the measured formation rate in urban Beijing, and the reason for this deviation will be discussed in Sect. 4.2 below. The curves from other previous studies are simulated using their reported equations (Chen et al., 2012; Jen et al., 2014; Hanson et al., 2017) and the parameters measured in this study. Some of these studies re-

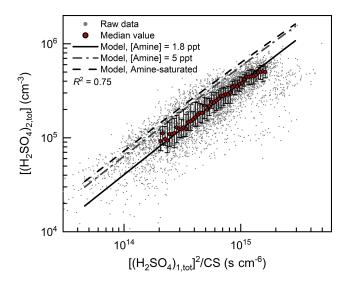


Figure 1. H<sub>2</sub>SO<sub>4</sub> dimer concentration for urban Beijing.  $[(H_2SO_4)_{1,tot}]$  and  $[(H_2SO_4)_{2,tot}]$  are the measured  $H_2SO_4$ monomer and dimer concentrations, respectively. The horizontal axis is proportional to the theoretical steady-state H<sub>2</sub>SO<sub>4</sub> dimer concentration at given amine concentrations. The measured raw data on NPF days with a temporal resolution of 5 min are shown with small dots. The median values of these dots grouped by the horizontal coordinate are shown with big red markers, and the error bars indicate the lower and upper quartiles. The condensation sink (CS) and temperature for the simulated formation rate are their median values during the observed NPF events, i.e., 0.017 s<sup>-1</sup> and 281 K, respectively. The amine-saturated limit (dashed black line) is simulated at an ultrahigh DMA concentration (10<sup>6</sup> ppt), so the evaporation rates of H<sub>2</sub>SO<sub>4</sub>-DMA clusters are negligible compared to their formation rates. The  $R^2$  value was calculated using logarithmic values.

ported higher evaporation rates of H<sub>2</sub>SO<sub>4</sub>-amine clusters according to their experimental data (Chen et al., 2012; Jen et al., 2014). However, due to these high evaporation rates, the simulated particle formation rates using these models are orders of magnitude lower than the measured particle formation rates in urban Beijing.

In addition to  $H_2SO_4$  nucleation with amines, the nucleation of oxidized organics with low vitalities was also reported in the atmosphere (e.g., Bianchi et al., 2016). Various organic vapors were observed in urban Beijing (Fig. S4). However, there was a considerable discrepancy between the absolute value and diurnal trend of particle formation rate contributed by organics and those obtained by measurements (Figs. S5 and S6), indicating that oxidized organics nucleation is not a governing mechanism to initialize NPF in urban Beijing during this campaign. Note that organics with low volatilities may contribute to the growth of larger (e.g., > 2 nm) particles (Deng et al., 2020b).

The consistency between the measured particle formation rate and the kinetic model also provides hints on the sticking probability between H<sub>2</sub>SO<sub>4</sub>-amine clusters and parti-

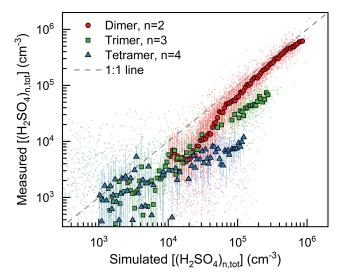
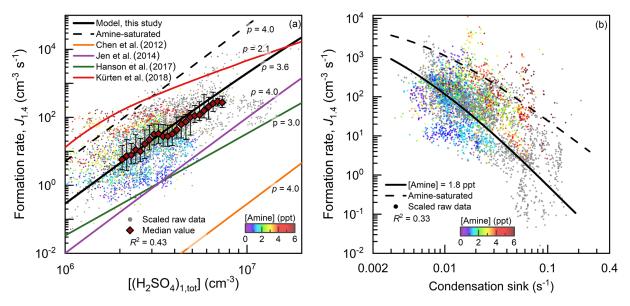


Figure 2. The measured and simulated  $H_2SO_4$  dimer, trimer, and tetramer concentrations. The concentrations of clusters containing the same  $H_2SO_4$  molecule number are summed together. The small dots are the raw data on both NPF and non-NPF days with a temporal resolution of 10 min, and their median values are shown with big markers. The simulation condition was case specific for each raw data dot. The error bars indicate the lower and upper quartiles. The uncertainty of instrument detection efficiency contributes to the difference between the measured and simulated values, as explained in the main text.

cles. In a previous study (Kulmala et al., 2017), it was discussed that other condensable vapors in addition to H<sub>2</sub>SO<sub>4</sub> and amine may contribute to the initial growth of new particles and that the coagulation scavenging effect in the polluted environment may be overestimated because of the overestimated sticking probability between particles or clusters. However, in this study, we found that for up to  $\sim 1.4$  nm particles (in electrical mobility diameter), the particle formation rate estimated using H<sub>2</sub>SO<sub>4</sub>-amine clustering and a sticking probability of 1.0 is consistent with the measured formation rate. The measured H<sub>2</sub>SO<sub>4</sub> trimer and tetramer concentrations are even lower than the simulated concentrations (Fig. 2) due to potential measurement uncertainties, whereas a significant contribution of other condensable vapors or an overestimated coagulation sink will theoretically result in higher measured concentrations compared to the simulated concentrations. However, the further growth of particles beyond  $\sim 1.4$  nm in polluted environments still needs further exploration.

#### 4.2 The influence of coagulation scavenging

The scavenging of H<sub>2</sub>SO<sub>4</sub>-amine clusters due to coagulation with larger particles is a major limiting factor for NPF in urban Beijing. After normalizing the influence of H<sub>2</sub>SO<sub>4</sub> monomer concentration, the particle formation rate decreases with an increasing CS (Fig. 3b). Negative dependencies of



**Figure 3.** The measured and simulated particle formation rates  $(J_{1,4})$  as a function of (a)  $H_2SO_4$  monomer concentration ( $[(H_2SO_4)_{1,tot}]$ ) and (b) condensation sink (CS). The measured raw data on NPF days with a temporal resolution of 5 min are shown with small dots and colored by the measured effective amine concentration. When amine concentrations are not available, the corresponding dots are shown in grey. The median values of these dots grouped by the horizontal coordinate in (a) are shown with big red markers, and the error bars indicate the lower and upper quartiles. The measured formation rates shown in (a) and (b) are scaled to the median CS and  $[(H_2SO_4)_{1,tot}]$  for the measured NPF events, respectively. The scaling formulae are given in the main text and illustrated in the Supplement. Note that the simulated formation rates are not scaled. p is the simulated power of  $J_{1,4}$  with respect to  $[(H_2SO_4)_{1,tot}]$ . The  $R^2$  values were calculated using logarithmic values.

 $\rm H_2SO_4$  cluster concentration and sub-3 nm particle concentration on CS were also reported in our previous studies in urban Beijing (Cai et al., 2017b; Deng et al., 2020a). However, there was usually a good positive correlation between CS and amine concentrations in urban Beijing, presumably due to the correlation between their sources. As a result, the apparent dependency between the measured  $J_{1.4}$  and CS in Fig. 3b was also influenced by amine concentrations.

Although coagulation scavenging does not affect the detailed equilibrium of reactions, it can have significant impacts on the steady-state cluster concentrations. For instance, the particle formation rate of the CLOUD chamber experiments deviates from the measured and simulated formation rates in urban Beijing. This indicates that although the H<sub>2</sub>SO<sub>4</sub> concentration in these chamber studies was in the typical ambient range, these results from chamber experiments are not directly applicable to represent the real atmospheric conditions in urban Beijing due to the difference in coagulation scavenging rates characterized by CS. The median CS in urban Beijing during the NPF events in this field measurement was  $\sim 0.017 \,\mathrm{s}^{-1}$ , which is nearly an order of magnitude higher than the total loss rates in the chamber studies (Kürten et al., 2014, 2018; Hanson et al., 2017). Hence, the curve for the CLOUD chamber experiments in Fig. 3 deviates from the measured data in urban Beijing.

The power of particle formation rate to  $H_2SO_4$  monomer concentration,  $[(H_2SO_4)_{1,tot}]^p$  is consistent with the argu-

ment that coagulation scavenging is a limiting factor for NPF in the polluted environment. It can be proven that under the CS-controlled regime, the power of the formation rate of  $(H_2SO_4)_4$  (amine)<sub>4</sub> to  $H_2SO_4$  concentration, p, is  $\sim 4.0$ rather than 2.0 (Fig. S7), which is consistent with the measured p in urban Beijing (Fig. 3a). In the perspective of conventional kinetic nucleation theory, the critical step of nucleation is the formation of H<sub>2</sub>SO<sub>4</sub> dimer clusters. Accordingly, the p value is expected to be 2.0 (Kuang et al., 2008). This power dependency was also used to prove that there was no significant evaporation of  $(H_2SO_4)_n(amine)_n$  for n > 2 in a previous CLOUD chamber study (Kürten et al., 2014). However, this theorem is valid only when the external cluster losses are negligible (Ehrhart and Curtius, 2013; Kupiainen-Määttä et al., 2014; Elm et al., 2020), whereas in the presence of a high aerosol concentration the loss rates of H<sub>2</sub>SO<sub>4</sub>amine clusters are usually an order of magnitude higher than their growth rates into large clusters. As a result, the cluster concentrations maintain a pseudo-steady-state, and their growth fluxes into the next larger clusters are proportional to, rather than independent of, H<sub>2</sub>SO<sub>4</sub> concentration (as detailed in the Supplement).

#### 4.3 The influence of amine concentrations

In addition to coagulation scavenging, the low effective amine concentration is another limiting factor for NPF in urban Beijing. During the measurement period in urban Beijing, the median C<sub>2</sub>-amine concentrations for the daytime NPF period and all the observation periods were 1.8 and 2.7 ppt, respectively. Meanwhile, measured H<sub>2</sub>SO<sub>4</sub> dimer concentration and particle formation rate in urban Beijing were lower than the amine-saturation limit (Figs. 1 and 3). Amine saturation means that further increasing the amine concentrations does not significantly enhance the nucleation rate. In contrast, under unsaturated amine concentrations, with respect to the formation of H<sub>2</sub>SO<sub>4</sub>-amine clusters, they are not stable against evaporation. The measured H<sub>2</sub>SO<sub>4</sub> dimer concentration and particle formation rate indicate a moderate evaporation rate of H<sub>2</sub>SO<sub>4</sub>-amine clusters, because they were close to but lower than their corresponding aminesaturation limits. This unsaturated particle formation rate with a low effective amine concentration is consistent with the saturation concentration of amines ( $\sim 5-20 \, \mathrm{ppt}$ ) reported in previous chamber experiments (Almeida et al., 2013; Jen et al., 2014).

In addition, the dependency of the measured particle formation rate on amine concentrations in Fig. 3a provides support for the view that unsaturated  $H_2SO_4$ -amine nucleation occurs in urban Beijing. Note that the apparent correlation between the effective amine concentration and CS was minimized in Fig. 3a by scaling the measured  $J_{1.4}$  with respect to CS. In contrast, the apparent negative correlation between the measured  $J_{1.4}$  and the effective amine concentration in Fig. 3b is governed by the positive correlation between the effective amine concentration and CS in urban Beijing. For the same reason, a negative correlation between NPF and amine concentrations was also reported in central Germany (Kürten et al., 2016).

In the above analysis, DMA is thought to be a major base that stabilizes H<sub>2</sub>SO<sub>4</sub>-amine clusters, and TMA may also contribute. Other bases, e.g., monomethylamine and NH<sub>3</sub>, were measured but are not included in our analysis due to their relatively weak bond to H<sub>2</sub>SO<sub>4</sub> molecules. Although NH<sub>3</sub> concentrations are high (with a median value of 789 ppt during the observed NPF in this study), theoretically NH<sub>3</sub> cannot be the major base to stabilize H<sub>2</sub>SO<sub>4</sub> in urban Beijing due to the high evaporation rate of the NH<sub>4</sub>HSO<sub>4</sub> molecule (Ortega et al., 2012; Jen et al., 2014; Olenius et al., 2017). However, these relatively weak bases may contribute to the particle growth, and their synergistic effects and base substitutions have been reported in previous studies (Kupiainen et al., 2012; Glasoe et al., 2015; Myllys et al., 2019). Trimers of neutral H<sub>2</sub>SO<sub>4</sub> stabilized by C<sub>1</sub> and C<sub>4</sub> amines are detected during NPF events in this study, as shown in Fig. S4a. Despite these potential contributions, the formation of large clusters until  $(H_2SO_4)_4$  (amine)<sub>4</sub> and  $\sim 1.4$  nm particles (in electrical mobility diameter) in urban Beijing can be quantitatively explained by the kinetic model (Figs. 1–3 and S5). Some other compounds in addition to amines, such as diamines (Jen et al., 2016) and guanidine (Myllys et al., 2019), are also reported to be possible bases to stabilize H<sub>2</sub>SO<sub>4</sub> clusters; however, they were not observed during this measurement campaign.

# 4.4 The synergistic influences of H<sub>2</sub>SO<sub>4</sub>, coagulation scavenging, and amine concentrations

Summarizing the synergistic effect of H<sub>2</sub>SO<sub>4</sub> monomer concentration, CS, and the effective amine concentration, the governing factors for H<sub>2</sub>SO<sub>4</sub>-amine nucleation at different regimes are illustrated in Fig. 4. The horizontal coordinate is equal to  $0.5\beta_{11} \cdot [(H_2SO_4)_{1,tot}]/CS$ , where  $\beta_{11}$  is the collision coefficient of two (H<sub>2</sub>SO<sub>4</sub>)<sub>1</sub>(amine)<sub>1</sub> clusters. It characterizes the ratio of the condensational growth rate ( $\beta_{11}$ .  $[(H_2SO_4)_{1,tot}])$  of a molecule or cluster to its loss rate (CS). The vertical coordinate is equal to  $\beta_{AB}$  [amine]/(CS +  $\gamma$ ), where  $\beta_{AB}$  is the collision coefficient of an  $H_2SO_4$  molecule and an amine molecule, and  $\gamma$  is the evaporation rate of (H<sub>2</sub>SO<sub>4</sub>)<sub>1</sub>(amine)<sub>1</sub>. As indicated by this formula, this vertical coordinate is linearly proportional to the effective amine concentration.  $J_{1.4}$  was estimated using the proposed model, and it is normalized by dividing it by its collision limit  $(J_c)$ at the same H<sub>2</sub>SO<sub>4</sub> monomer concentration. The collision limit refers to the collision rate of two H<sub>2</sub>SO<sub>4</sub> monomers, which is theoretically the maximum steady-state particle formation rate. The effective amine concentration is assumed to be independent of [(H<sub>2</sub>SO<sub>4</sub>)<sub>1,tot</sub>]; i.e., it is assumed that the formation of (H<sub>2</sub>SO<sub>4</sub>)<sub>1</sub>(amine)<sub>1</sub> does not cause a change in amine concentrations. It should be clarified that due to evaporation and the minor reaction pathways, the normalized particle formation rate is governed but not only determined by the normalized sulfuric and amine concentrations. Hence, the background color map shown in this figure is illustrative rather than quantitative.

According to the region of measured data in Fig. 4, particle formation rate is sensitive to both CS and amine concentrations in the real atmosphere of urban Beijing and Nanjing, whereas it is insensitive to amine concentrations in most of the experimental conditions in the CLOUD studies (Almeida et al., 2013; Kürten et al., 2014). A high C2-amine concentration was reported for Shanghai ( $40 \pm 12 \,\mathrm{ppt}$ ; Yao et al., 2016); therefore, the effective amine concentration in Fig. 4 is located in a similar range to that in the CLOUD experiments. In the upper right corner of Fig. 4, both the sulfuric and amine concentrations are sufficiently high, so the steadystate cluster and particle formation rates are governed by the collision rate of two H<sub>2</sub>SO<sub>4</sub> monomers. At the upper left corner, nucleation is controlled by CS, because the concentrations of clusters and particles are governed by both their formation and loss rates (see Eqs. (S22)–(S28) in the Supplement). In this CS-controlled regime, increasing the effective amine concentration in a narrow range does not significantly increase the formation rate. This is because the effective amine concentration is sufficient with respect to the evaporation of H<sub>2</sub>SO<sub>4</sub>-amine clusters, and the formation rate is close to its amine-saturation limit. At the lower right cor-

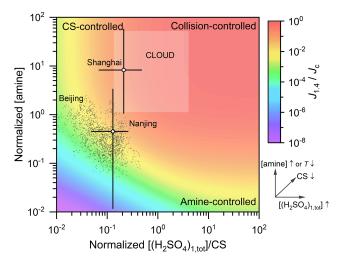


Figure 4. An illustrative figure for the governing factors of H<sub>2</sub>SO<sub>4</sub>– amine nucleation in various conditions. The horizontal and vertical coordinates are the normalized H2SO4 and effective amine concentrations, respectively, and the normalizing formulae are given in the main text. The dependencies of horizontal and vertical coordinates or  $[(H_2SO_4)_{1,tot}]$ , CS, [amine], and temperature (T) are indicated by the arrows. The color indicates the normalized steady-state formation rate of simulated 1.4 nm particles,  $J_{1.4}$ . Then  $J_c$  is the theoretical formation rate at the collision limit. The dark grey markers are the measured data on NPF days in Beijing with a temporal resolution of 5 min. The semitransparent square above the colored contour is the estimated range for experimental conditions of the CLOUD studies (Almeida et al., 2013; Kürten et al., 2014):  $[(H_2SO_4)_{1,tot}]$  between  $5 \times 10^5$  and  $1.5 \times 10^7$  cm<sup>-3</sup>, [amine] between 3 and 140 ppt, T = 278 K, and the sum of wall loss and dilution rates (instead of CS) was estimated to be  $0.002 \,\mathrm{s}^{-1}$ . The open markers and error bars indicate the median values and approximate ranges, respectively, of the normalized H<sub>2</sub>SO<sub>4</sub> and effective amine concentrations in Shanghai and Nanjing. The Shanghai data were reported by Yao et al. (2016, 2018). The Nanjing data were reported by Zheng et al. (2015) and Deng et al. (2020b). Note that for Shanghai and Nanjing, the amine concentrations were measured in different campaigns from those for H<sub>2</sub>SO<sub>4</sub> and CS, and the amine concentrations are the average of all days rather than the new particle formation period only. As a result, there are potential uncertainties in the results for Shanghai and Nanjing.

ner, the formation and growth of  $H_2SO_4$ -amine clusters are not limited by their coagulation losses. However, due to the low effective amine concentration, considerable evaporation of  $H_2SO_4$ -amine clusters limits the formation rate of new particles.

#### 5 Conclusions

The predominating mechanism initiating NPF in urban Beijing is illustrated. Compared to previous studies that investigated H<sub>2</sub>SO<sub>4</sub>-amine nucleation under high amine concentrations in laboratories (Almeida et al., 2013; Jen et al., 2014;

Kürten et al., 2014) and reported a governing mechanism in Shanghai (Yao et al., 2018), we further show the governing roles of H<sub>2</sub>SO<sub>4</sub>, amines, and the coagulation scavenging effect at the molecular level based on long-term ambient atmospheric measurements in urban Beijing. Comparing the measured particle formation rate and cluster concentrations with those simulated using a kinetic model, we demonstrated and quantified the influences of H<sub>2</sub>SO<sub>4</sub> concentration, amine concentration, and coagulation scavenging on new particle formation rate in urban Beijing. The formation and growth of H<sub>2</sub>SO<sub>4</sub>-amine clusters under the strong coagulation scavenging effect is shown to be a major pathway for cluster growth up to  $\sim 1.4$  nm particles (in electrical mobility diameter). Both theoretical analysis and measured data support that unlike previous chamber studies (Kürten et al., 2014; Almeida et al., 2013) and atmospheric measurements (Yao et al., 2018), the typical amine concentrations measured in this study are sometimes insufficient to bound with nearly all the H<sub>2</sub>SO<sub>4</sub> monomers into H<sub>2</sub>SO<sub>4</sub>-amine clusters. The sensitivity of NPF to amine concentrations also indicates that the contributions of NPF to the aerosol number and surface concentrations will decrease if atmospheric amine concentrations are reduced. Due to the correlated variables and measurement uncertainties in atmospheric measurements, the quantitative influences of various amines, water vapor, and coagulation sink on NPF need further verification from experiments in chambers or other controlled systems. For future chamber studies, we recommend that the gaseous precursors and condensation sink should be at their typical ambient levels, as their values not only affect particle formation rate but also the detailed nucleation kinetics.

Code availability. The Julia code for the kinetic model is available upon request.

Data availability. The data that support the findings of this study are available from the corresponding author upon request.

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