



Dimensionality-reduction techniques for complex mass spectrometric datasets: application to laboratory atmospheric organic oxidation experiments

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Received: 17 May 2019 – Discussion started: 21 June 2019

Revised: 26 November 2019 – Accepted: 1 December 2019 – Published: 27 January 2020

Abstract. Oxidation of organic compounds in the atmosphere produces an immensely complex mixture of product species, posing a challenge for both their measurement in laboratory studies and their inclusion in air quality and climate models. Mass spectrometry techniques can measure thousands of these species, giving insight into these chemical processes, but the datasets themselves are highly complex. Data reduction techniques that group compounds in a chemically and kinetically meaningful way provide a route to simplify the chemistry of these systems but have not been systematically investigated. Here we evaluate three approaches to reducing the dimensionality of oxidation systems measured in an environmental chamber: positive matrix factorization (PMF), hierarchical clustering analysis (HCA), and a parameterization to describe kinetics in terms of multi-generational chemistry (gamma kinetics parameterization, GKP). The evaluation is implemented by means of two datasets: synthetic data consisting of a three-generation oxidation system with known rate constants, generation numbers, and chemical pathways; and the measured products of OH-initiated oxidation of a substituted aromatic compound in a chamber experiment. We find that PMF accounts for changes in the average composition of all products during specific periods of time but does not sort compounds into generations or by another reproducible chemical process. HCA, on the other hand, can identify major groups of ions and patterns of behavior and maintains bulk chemical prop-

erties like carbon oxidation state that can be useful for modeling. The continuum of kinetic behavior observed in a typical chamber experiment can be parameterized by fitting species' time traces to the GKP, which approximates the chemistry as a linear, first-order kinetic system. The fitted parameters for each species are the number of reaction steps with OH needed to produce the species (the generation) and an effective kinetic rate constant that describes the formation and loss rates of the species. The thousands of species detected in a typical laboratory chamber experiment can be organized into a much smaller number (10–30) of groups, each of which has a characteristic chemical composition and kinetic behavior. This quantitative relationship between chemical and kinetic characteristics, and the significant reduction in the complexity of the system, provides an approach to understanding broad patterns of behavior in oxidation systems and could be exploited for mechanism development and atmospheric chemistry modeling.

1 Introduction

Air quality and climate change are major threats to the quality of millions of human lives across the globe (IPCC, 2014; Landrigan et al., 2018). An important scientific component of both topics is the photooxidation chemistry of organic compounds in the atmosphere, which can lead to the formation

of ozone and fine particulate matter, both of which can affect the radiative budget of the atmosphere and can harm human health. A detailed understanding of this chemistry is necessary to predict and mitigate these effects. However, this is challenging because of the diversity and number of species involved. Gas-phase organic compounds emitted directly into the atmosphere have a wide range of functionality and reactivity, and oxidation of these precursors by O_3 , OH, or NO_3 can further functionalize or fragment the molecules. The number and diversity of the product species increases with the number of generations of reaction, and key properties of these product species, such as volatility, reactivity, and concentration, can vary over orders of magnitude (Glasius and Goldstein, 2016; Goldstein and Galbally, 2007).

This complexity presents several challenges. In order to fully characterize oxidation of organic compounds, analytical techniques must be able to detect hundreds to thousands of individual species and accommodate the diversity of functionality and concentration. Advances in instrumentation, especially high-resolution time-of-flight chemical ionization mass spectrometry (CIMS), have enabled the detection of a large number of oxidation products in chamber and field experiments. CIMS involves the introduction of a reagent ion, which then reacts with the analyte, forming product ions that are detected with mass spectrometry. Chemical selectivity can be achieved through choice of the reagent ion, and fast, online measurement of air samples is possible. CIMS instruments with high mass resolution (maximum FWHM $m/\Delta m > 3000$) can unambiguously determine the elemental composition of most detected ions with m/z less than 200, and the elemental composition of ions with $m/z > 200$ can usually be determined with some certainty (Junninen et al., 2010). The analytical capability of atmospheric CIMS instrumentation is rapidly improving, and modern instruments can have sensitivities on the order of $10\,000\text{ cps ppbv}^{-1}$ and a resolution greater than $10\,000\text{ }m/\Delta m$ (Breitenlechner et al., 2017; Krechmer et al., 2018), allowing the measurement of hundreds to thousands of species on a rapid time base (Isaacman-VanWertz et al., 2017; Müller et al., 2012).

While this represents a major advance in our ability to detect and characterize trace atmospheric chemical components, these large datasets can be difficult and time-consuming to interpret, and it is not clear how the full information content from thousands of ions can be best used. Further, secondary ion processes, such as cluster formation or ion fragmentation, can occur within the mass spectrometer, complicating the mass spectra, and different CIMS techniques have differing chemical specificities that can be hard to predict. Data analysis techniques are therefore needed to efficiently reduce the amount of data to more manageable and interpretable sizes. Further, the interpretation of these measurements in terms of chemical mechanisms is often not straightforward. Most laboratory studies use CIMS measurements to support, refute, or suggest new chemical mechanisms; this is typically done by hand, focusing on several

key species of interest. Data analysis techniques that allow for the extraction of useful chemical and mechanistic information from entire mass spectra are valuable and necessary but have not been systematically explored.

Simplification is also needed to incorporate oxidation chemistry into climate and air quality models. Large-scale regional and global models (e.g., chemical transport models and earth system models) cannot currently incorporate a high level of chemical detail. Photochemical mechanisms commonly used to incorporate chemistry into regional and global models typically include 30–200 species and 100–400 reactions (Brown-Steiner et al., 2018; Jimenez et al., 2003), which is much lower than the number of product species from individual precursors included in explicit chemistry mechanisms such as the Master Chemical Mechanism (300–1000+ product species, e.g., Bloss et al., 2005a; Jenkin et al., 2003; Saunders et al., 2003) or GECKO-A ($\sim 10^5$ species, Aumont et al., 2005). In order to reduce the number of species in models, volatile organic compounds (VOCs) are represented by groups, or are lumped, and the choice of lumping criterion can affect the derived ozone, aerosol, and product VOC formation values (Jimenez et al., 2003; Zhang et al., 2012). In gas-phase mechanisms, compounds have been lumped by degree of unsaturation, emission rates, functional groups, or reactivity towards OH (Brown-Steiner et al., 2018; Crassier et al., 2000; Houweling et al., 1998; Jimenez et al., 2003; Gery et al., 1989; Carter, 1990; Stockwell et al., 1997). Similarly, secondary organic aerosol formation has been parameterized by lumping organic species by volatility, O : C ratio, number of carbon and oxygen atoms, or polarity and assigning kinetic properties to each group (Cappa and Wilson, 2012; Donahue et al., 2012; Lane et al., 2008; Pankow and Barsanti, 2009). Lumping schemes could be improved by using laboratory data to define important groups of compounds and assign experimentally derived chemical and kinetic properties to each group to act as a surrogate species.

Several methods have been used to categorize mass spectra and to group compounds. We consider two methods previously used to reduce the dimensionality of complex atmospheric chemistry measurements, positive matrix factorization (PMF) and hierarchical clustering analysis (HCA). Both methods have seen substantial use in the simplification and interpretation of field measurements but have seen far less use in the laboratory, and there has been little exploration of how they can be used to gain useful chemical or mechanistic information from laboratory mass spectrometric datasets. We additionally address a fundamental, underexplored problem in laboratory chamber studies: how to systematically characterize the kinetics of an oxidation system. The systematic characterization is achieved through the gamma kinetics parameterization (GKP) and can be used to group compounds based on similar kinetic properties. The three methods (PMF, HCA, and GKP) have different mathematics but the same goals: to identify groups of compounds and replace each group with a chemically meaningful surrogate.

The three methods are evaluated in terms of the following criteria: whether the resulting surrogates have chemically realistic behavior; whether the surrogates have the same range of chemical properties as the original dataset; which subjective choices the researcher needs to make when implementing the method; and what other new information about the system can be learned. We additionally discuss the extent to which different methods agree in their identification of major groups of compounds. The output of these dimensionality-reduction techniques can be used to quickly analyze and interpret chamber experiments and could be used to reduce the complexity of chemical mechanisms included in models.

2 Methods

2.1 Data collection

We use two datasets: a synthetic dataset describing a simple multigenerational kinetic system and measurements of the OH-initiated oxidation of 1,2,4-trimethylbenzene in an environmental chamber. The synthetic dataset is useful for evaluating the various dimensionality-reduction schemes used here, because the reaction rate constants and generation of each species are known exactly. The chamber data demonstrate the application of the data reduction techniques to a real-world system measured with online mass spectrometry.

2.1.1 Synthetic dataset

A schematic of the simple synthetic kinetic system is shown in Fig. 1. The precursor molecule A reacts with OH to produce first-generation species (B), which in turn react with OH to produce second-generation (C) and third-generation species (D). Only reactions with OH are considered. The system includes three pathways with differing yields, and each pathway includes a product with a fast, a slow, and an intermediate OH rate constant. The different rate constants (randomly generated) and yields simulate a range of product behavior. To enable PMF measurements, artificial noise was added to the synthetic data. The noise is normally distributed with a standard deviation proportional to the square root of the signal. The proportionality constant, based on a typical proton-transfer-reaction mass spectrometer (PTR-MS) sensitivity of $10\,000\text{ counts ppb}^{-1}\text{ s}^{-1}$, was chosen to generate signal-to-noise ratios between 10 and 100, a reasonable range for chamber experiments.

2.1.2 Chamber oxidation of 1,2,4-trimethylbenzene

An oxidation experiment was conducted in the MIT environmental chamber, which consists of a 7.5 m^3 PFA enclosure. The chamber conditions were controlled at 20°C and 2 % relative humidity. The chamber is illuminated by forty-eight 40 W black lights with a 300–400 nm spectrum peaking at 350 nm. During experiments the chamber maintains a con-

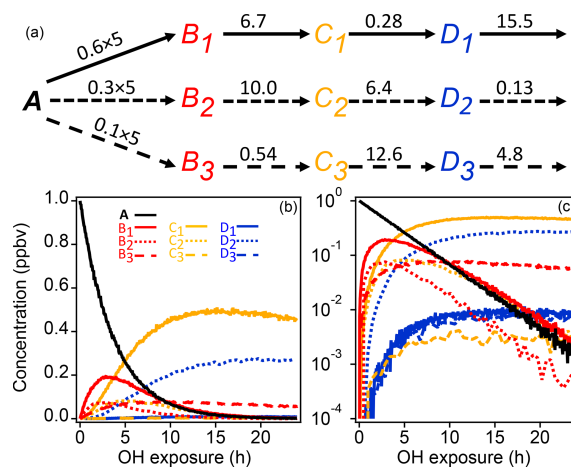


Figure 1. Schematic of reaction pathways with OH (a) of synthetic data and time series shown with linear and log concentration (b, c). Arrows represent a reaction with OH. Reaction rate constants with OH are written above the arrows (units are in $10^{-11}\text{ cm}^3\text{ molecule}^{-3}\text{ s}^{-1}$). Precursor species A reacts at a rate of $5 \times 10^{-11}\text{ cm}^3\text{ molecule}^{-3}\text{ s}^{-1}$ with yields of 0.6, 0.3, and 0.1 for the three pathways, respectively. Products of pathways 1, 2, and 3 are drawn with solid, short-dash, and long-dash lines, respectively, and the first-, second-, and third-generation products are drawn in red, yellow, and blue. The total OH exposure is equal to 24 h at an average OH concentration of $1.5 \times 10^6\text{ molecule cm}^{-3}$.

stant volume, and clean air is continuously added at a rate equal to the instrument sample flow (15 L min^{-1}). Additional details of chamber operation have been previously reported (Hunter et al., 2014).

Dry ammonium sulfate seed (which provide a surface area onto which low-volatility vapors can condense) was first added to the chamber to reach a number concentration of $5.7 \times 10^4\text{ cm}^{-3}$ ($19.7\text{ }\mu\text{g m}^{-3}$). Nitrous acid (HONO, the OH precursor) was added by bubbling clean air through a drop-wise addition of H_2SO_4 to NaNO_2 to reach a concentration of 45 ppbv in the chamber. Several parts per billion by volume of an unreactive tracer, hexafluorobenzene, were added to provide a measure of chamber dilution. A total of $3\text{ }\mu\text{L}$ of neat 1,2,4-trimethylbenzene (Sigma-Aldrich) was added by injection into a 70°C heated inlet with a flow rate of 15 L min^{-1} , resulting in an initial concentration of 69 ppbv in the chamber. The reagents were allowed to mix for 15 min and then the experiment was initiated by turning on lights to photolyze nitrous acid and generate OH. Measurements were conducted for 7 h. During this time three additional aliquots of nitrous acid (27, 10, and 18 ppbv) were added at regularly spaced intervals to roughly maintain the OH concentration. The OH concentration was determined by fitting a double-exponential function to the measured decrease of 1,2,4-trimethylbenzene, including a known dilution term (determined from hexafluorobenzene dilution) and an OH reaction term. A total atmospheric-equivalent exposure of

16.5 h (assuming an average atmospheric OH concentration of 1.5×10^6 molecule cm^{-3}) was achieved.

CO and formaldehyde were measured by tunable infrared laser differential absorption spectroscopy (TILDAS, Aerodyne Research Inc.). Other gas-phase organic species were measured by chemical ionization, followed by analysis with high-resolution time-of-flight (HR-ToF) mass spectrometry. Three chemical ionization mass spectrometry (CIMS) techniques were used: I^- reagent ion, H_3O^+ reagent ion, and NH_4^+ reagent ion. The I^- CIMS instrument is from Aerodyne Research Inc. and is described by Lee et al. (2014). H_3O^+ and NH_4^+ CIMS involved proton-transfer-reaction mass spectrometers with switchable reagent ion chemistry (PTR3- H_3O^+ and PTR3- NH_4^+ , Ionicon Analytik). The PTR3 H_3O^+ CIMS and NH_4^+ CIMS techniques are described by Breitenlechner et al. (2017) and Zaytsev et al. (2019), respectively. H_3O^+ CIMS was also carried out using a second proton-transfer-reaction mass spectrometer (Vocus-2R-PTR, TOFWERK AG), which is described by Krechmer et al. (2018). Total organic aerosol mass was measured using a high-resolution time-of-flight aerosol mass spectrometer (AMS) from Aerodyne Research Inc. (DeCarlo et al., 2006), calibrated with ammonium nitrate and assuming a collection efficiency of 1. Organic aerosol accounted for approximately 2 % of the secondary carbon, and individual ion measurements from the AMS are not considered separately. The TILDAS was calibrated directly for CO and formaldehyde. The Vocus-2R-PTR was calibrated directly for 1,2,4-trimethylbenzene and acetone. The PTR3 H_3O^+ CIMS was calibrated directly for 15 individual species and an average calibration factor was applied to other species. The PTR3- NH_4^+ and I^- CIMS were calibrated using a combination of direct calibration and collision-induced dissociation (Lopez-Hilfiker et al., 2016; Zaytsev et al., 2019). We note however that the calibration of each instrument does not affect any results presented in this work, since the analysis techniques used examine the time-dependent behavior, and not the absolute concentrations, of the measured species.

Sampling from the chamber to CIMS instruments was designed to reduce inlet losses of compounds as much as possible, within the physical constraints of the chamber. Each instrument used a 0.1875 in. (3/16") ID PFA Teflon line of 1 m or less in length, with a flow of 2 L min^{-1} . Inlets extended 10 cm into the chamber and no metal fittings were used. The PTR instruments additionally have instrument inlets and ion-molecule-reaction chambers that minimize gas contact with walls (Breitenlechner et al., 2017; Krechmer et al., 2018). In this study, CIMS inlet (including chamber and instrument inlet) loss timescales were 15 s or less for test compounds with saturation concentrations between 10^2 and $10^7 \mu\text{g m}^{-3}$, and therefore wall interactions for these species are unlikely to affect the observed kinetics, which occur over tens of minutes (Krechmer et al., 2016).

Chamber background for each measurement was determined from measurements taken prior to precursor injection,

which are subtracted from each chamber measurement reported. All measurements were also corrected for dilution by normalizing to the hexafluorobenzene tracer (for gas-phase data) or to measured $(\text{NH}_4)_2\text{SO}_4$ aerosol seed (for particle-phase data, which also correct for wall loss and AMS collection efficiency).

Between 1000 and 3000 peaks with variability above background were observed in the mass spectra from each CIMS instrument; these include chemically relevant ions related to oxidation products, as well as other ion signals from sources such as instrument ion sources, the hexafluorobenzene dilution tracer, tubing and inlets, and interferences from large neighboring peaks in the mass spectrum (Cubison and Jimenez, 2015). Two data-processing steps were used to identify the chemically relevant ions.

First, the elemental formulas of all ions were determined. With the resolution of the instruments used here (maximum $\sim 10000 m/\Delta m$ for Vocus-2R-PTR and PTR3; ~ 3000 for I^- CIMS), elemental composition can become ambiguous at high m/z values. We first assigned all unambiguous peaks, where only one reasonable formula within 10 ppm of the peak was possible, beginning with the largest peaks in order to identify and exclude isotopes. Then, we used trends observed in Kendrick mass defect plots to suggest formulas for species expected at higher masses. Remaining peaks (< 1 % of instrument signal) were assigned the formula with the nearest mass that included C, H, N, and O; had nine or fewer carbon atoms; and had positive, integer double-bond equivalency (again, beginning with the largest peaks and excluding isotopes). A mass defect plot showing unambiguous ions, and the complete set of ions, is shown in Fig. S1 in the Supplement.

Second, chemically relevant ions were separated from all other ion signals using hierarchical clustering. Chemically relevant ions are those which result from oxidation products. They are enhanced above background during the oxidation experiment and do not have sudden, stepwise changes, which would indicate an instrument interference. A difference mass spectrum, which compares the average signal of each ion before chemistry is initiated to the average signal during oxidation, is a simple way to identify relevant ions but can be misleading for ions with low signal-to-noise ratios or variability unrelated to oxidation chemistry. Hierarchical clustering provides an alternative method, involving the systematic examination of the time-dependent behavior of all measured species. Chemically relevant ions exhibit a time dependence that is consistent with chemical kinetics (formation of the product, often followed by reactive loss) that is different from that of ions not resulting from oxidation. These two classes are clustered separately from each other, enabling the straightforward selection of only chemically relevant ions. The hierarchical clustering algorithm is described in Sect. 2.2.2. An example for the PTR3 H_3O^+ mode instrument is shown in Fig. S2. This approach was used to identify

chemically relevant ions and to exclude all background ions from each CIMS instrument.

Compounds that were measured by more than one instrument, identified as having the same elemental composition (after correction for any reagent ion chemistry) and similar time-series behavior (Pearson's $R > 0.9$), were included only once in the dataset with all product species. When selecting compounds measured by more than one instrument, data from PTR-MS instruments, which have the smallest calibration uncertainties, were used first, followed by I^- CIMS and NH_4^+ CIMS. In the final combined dataset, approximately half the carbon in oxidation products was measured by PTR-MS, with about 15 % measured each by I^- CIMS, NH_4^+ CIMS, and TILDAS, and an additional 2 % by AMS. We recognize that there is a great deal of uncertainty associated with calibrating CIMS instrumentation and identifying detected ions. This is an active area of research that we do not attempt to address fully here. Calibration and identification of species measured by more than one instrument do not affect the major conclusions of this paper.

2.2 Implementation of data simplification tools

2.2.1 Positive matrix factorization (PMF)

In atmospheric chemistry, PMF analysis typically involves representing a time series of mass spectra (or other chemical measurements), recorded as a matrix of m measurements by t time points, as a linear sum of factors (Paatero, 1997; Ulbrich et al., 2009; Zhang et al., 2011). Each factor is fixed in chemical composition but varies in intensity over time.

PMF analysis of ambient air measurements has in many situations been shown to be robust and meaningful and has contributed greatly to our understanding of atmospheric and aerosol chemistry. PMF is frequently used for source apportionment and characterization of organic aerosol in field studies, for example, to sort aerosol as more or less oxidized or from a specific source such as biomass burning (Zhang et al., 2011). PMF is also frequently applied to VOC measurements in field studies. In this application, each factor indicates a particular VOC class (which can be associated with a specific source) and its magnitude, which is a powerful tool to support regulation.

Some aspects of atmospheric chemistry can complicate PMF analysis. Oxidation chemistry during transport from the source to the measurement location can change the chemical composition, causing a single source to appear as several factors, or causing oxidized species from several sources to be grouped together, and adding substantial uncertainty to the derived source profiles (Sauvage et al., 2009; Wang et al., 2013; Yuan et al., 2012). Factors including oxidation products, described as secondary or long-lived species, or that require correction for photochemistry have been reported in a number of studies from diverse locations (e.g., Abeleira et al., 2017; Sarkar et al., 2017; Shao et al., 2016; Stojić et al.,

2015), but the interpretation of such factors within the context of a continually evolving system is unclear.

Finally, PMF has been applied to measurements of oxidizing chemical systems to greatly reduce the complexity of the dataset and identify key shifts in chemistry, including aerosol in laboratory experiments (e.g., Craven et al., 2012; Fortenberry et al., 2018), VOCs in chamber experiments (Rosati et al., 2019), and gas-phase highly oxidized molecules in field studies (Massoli et al., 2018; Yan et al., 2016). Therefore, it is important to understand whether PMF analysis of an oxidizing system returns chemically distinct, reproducible factors that correspond to a physical or chemical aspect of the system.

The algorithm was implemented using the PMF Evaluation Tool v2.08 (Ulbrich et al., 2009) using the PMF2 algorithm (Paatero, 2007). We chose this implementation because it is widely used in atmospheric science and has been optimized for atmospheric chemistry data. Briefly, the algorithm takes as input an $m \times n$ matrix of measured data \mathbf{M} , containing n measured compounds at m time points, and a matrix of estimated error (1 standard deviation, σ) for each point in the measured data matrix. The solution for a given number of factors p is given as an $m \times p$ matrix \mathbf{G} of factor time series, a $p \times n$ matrix \mathbf{F} of factor profiles, and a matrix \mathbf{E} that contains the residual ($\mathbf{M} - \mathbf{GF}$). \mathbf{F} and \mathbf{G} are iteratively adjusted to minimize the quality-of-fit parameter Q :

$$Q = \sum_{i=1}^m \sum_{j=1}^n (e_{ij}/\sigma_{ij})^2,$$

where e_{ij} is the residual between the measurement and the PMF reconstruction of compound j at time point i , and σ_{ij} is the estimated error of that measurement.

The factors and their profiles are constrained to be non-negative. The measured data matrix \mathbf{M} for the synthetic dataset was constructed using all 10 species (precursor plus nine products) with artificial noise. The measured data matrix \mathbf{M} for the chamber dataset was constructed using all measured product species (defined as all chemically relevant ions from CIMS instruments plus total organic aerosol, CO, and formaldehyde), after background subtraction, dilution correction, and calibration in units of parts-per-billion carbon (ppbC). Duplicate measurements of individual species from multiple instruments were excluded. Although calibrated data are used here, because PMF operates on the unitless quality-of-fit parameter Q , the results are not sensitive to calibration, only to the signal-to-noise ratio of the individual measurements.

Because the precursor compound (1,2,4-trimethylbenzene) has an average intensity an order of magnitude larger than any other species, and therefore a very high signal-to-noise ratio, if it is included in \mathbf{M} , the quality-of-fit parameter Q and the resulting solution are dominated by the precursor. As this is not of interest, the precursor was also excluded in PMF analysis. Data were

interpolated to 500 points evenly spaced with respect to OH exposure (0–16.5 atmospheric-equivalent hours).

The matrix of estimated errors for the synthetic dataset was taken as the standard deviation used to generate the artificial noise. The matrix of estimated errors for the chamber dataset was generated by smoothing the data using a running 20 min linear best fit and subtracting these smoothed data from the original measurement. The standard deviation of the residual within a 20 min window was determined for each time point. Signal-to-noise ratios for both synthetic and chamber data are shown in Fig. S3. The overall relationship between the standard deviation determined for chamber data and the measured concentration is reasonable (Fig. S4).

Rotational forcing, which examines linear combinations of possible solutions using the parameter fPeak, was explored through fPeak values between -1 and 1 . The selected fPeak was chosen to avoid factor time series with multiple maxima, which are not physically realistic in the chamber system. Solutions were also explored using different random initialization values, or seeds. No significant differences were found between solutions with random seed values 1–10.

When PMF is used to reduce the complexity of a dataset, the number of factors must be chosen by the researcher, a choice that is inherently subjective. Solutions were explored with 1 to 10 factors for the synthetic dataset and the chamber data.

2.2.2 Hierarchical clustering analysis (HCA)

A second technique is to group or cluster individual measurements based on the similarity of their behavior over time. While a measurement of a single chemical species can contribute to more than one PMF factor, it can belong to only one cluster. Several approaches to clustering exist. The approach we consider here is agglomerative hierarchical clustering, which describes the degree of similarity between any two measurements and can be used to sort species into categories of behavior (Bar-Joseph et al., 2001; Müllner, 2011). Hierarchical clustering analysis (HCA) has been used to group aerosol particles based on the similarity between individual mass spectra determined by aerosol mass spectrometers (Marcolli et al., 2006; Murphy et al., 2003; Rebotier and Prather, 2007), describe time series of thermally desorbed organics measured by CIMS (Sánchez-López et al., 2014, 2016), and recently to determine the appropriate number of PMF factors used to analyze PTR-MS data from chamber studies (Rosati et al., 2019). To our knowledge it has not yet been used to group compounds with similar time-varying behaviors to understand chemical transformation in an oxidation system. In this work we show how this technique can be implemented and assess its ability to reduce the complexity of a dataset while maintaining chemical information.

Agglomerative hierarchical clustering sorts measurements by similar time-series behavior and displays the relative similarity between measurements. First, all measurements were

normalized so that the time-series behavior could be directly compared despite differences in absolute concentrations or detection efficiencies. Data are noisy, and noise can contribute to the absolute highest point in a time series. To account for this, we normalized data to the average of the 10 points surrounding the highest point in each time series. Then, the distance between each pair of measurements A and B was determined. The distance describes the dissimilarity between any two time-series measurements: two identical time series have a distance of zero, and measurements with different time-series behavior have larger distance values. Distance was calculated by summing the differences between the normalized measurement intensities A and B over all time points t :

$$d_{AB} = \sum_t \text{abs}(A_t - B_t).$$

Other distance metrics are possible, including using a correlation coefficient or the sum of squared residuals. This particular approach was chosen because it resulted in the grouping that was most reproducible and understandable as well as least sensitive to outlier points in the time series.

The algorithm begins with the distances between all original measurements. The pair of measurements s and t with the lowest distance value is found, and these two measurements are assigned to a new cluster u . The two original measurements s and t are removed from the set, and the new cluster u is added. Then, the distances between the new cluster u and all the remaining measurements are determined. The algorithm then iteratively searches for the next smallest distance value and combines the pair into a new cluster. As the algorithm iterates, new clusters can be formed from two original measurements, from an original measurement and a cluster, or from two clusters. The distance between the new cluster u and any other measurement or cluster in the set v is calculated as the average of the distances between each of the n individual members of u and m individual members of v , over all points i in cluster u and points j in cluster v :

$$d_{uv} = \sum_{i,j} \frac{d(u_i, v_j)}{m \times n}.$$

The algorithm continues until only one cluster remains. Clustering was implemented using the open-source `scipy.cluster.hierarchy.linkage` package (SciPy.org, 2018). The relationships between each of the different measurements and clusters are visualized using a dendrogram.

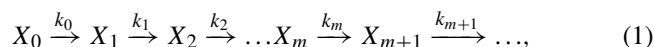
Compounds must be grouped into a specific number of clusters in order to use HCA to define surrogate species. The average chemical and kinetic properties of each cluster can be used to define a surrogate species. As with the number of factors from PMF, the number of clusters is subjectively chosen by the researcher. The clusters could be selected by hand or by choosing a threshold for distance d_{AB} to automatically define clusters. We chose to use a threshold to define

the number of clusters and considered several different values of thresholds that result in different numbers of clusters. The effect of threshold value on the interpretation of the data is discussed in Sect. 3.2.

2.2.3 Gamma kinetics parameterization (GKP)

To date, bulk characterization of oxidation products in photochemical chamber experiments has largely focused on their chemical composition and not their reactivity or mechanistic relationship. A few studies have derived kinetic information from time-series data (Smith et al., 2009; Wilson et al., 2012), but this has been limited to aerosol-aging experiments and not to atmospheric oxidation generally. A chamber oxidation experiment with speciated mass spectrometric measurements also contains a great deal of kinetic information, because the rates of formation and decay of each species are measured. In this work we show how the kinetic behavior of any particular measurement can be parameterized using a simple function, the gamma kinetics parameterization (GKP), which describes a system of first-order linear multi-step reactions. The function returns parameters that describe generation number (how many OH addition steps are needed on average to create the molecule) and reactivity (the relative rates of formation and decay), which are shown to correlate with key chemical characteristics. Grouping by similar kinetic parameters suggests a new, experimentally derived approach to lumping mechanisms.

A multigeneration reaction system can be described as a linear system of first-order reactions:



where k_i is the rate constant and m is the number of reactions needed to produce species X_m (i.e., the generation number). When all k_i 's are equal, the series of differential equations that describe the kinetics of Eq. (1) can be solved analytically, with the time dependence of any compound X_m described by

$$[X_m](t) = a(kt)^m e^{-kt}, \quad (2)$$

where a is a scaling factor that depends on both instrument sensitivity and stoichiometric yield (Smith et al., 2009; Wilson et al., 2012; Zhou and Zhuang, 2007). This function is related to the probability density function of the gamma distribution, a continuous probability distribution that has been previously used in chemistry to characterize protein kinetics (Pogliani et al., 1996; Zhou and Zhuang, 2007).

Oxidation reactions in a chamber experiment can be parameterized as a linear system of reactions, but the reactions between organic compounds and OH are bimolecular. This can be adjusted to a pseudo-first-order system by considering the integrated OH exposure $[\text{OH}]\Delta t = \int_0^t [\text{OH}]dt$ instead of reaction time t . In this case, the observed behavior of an

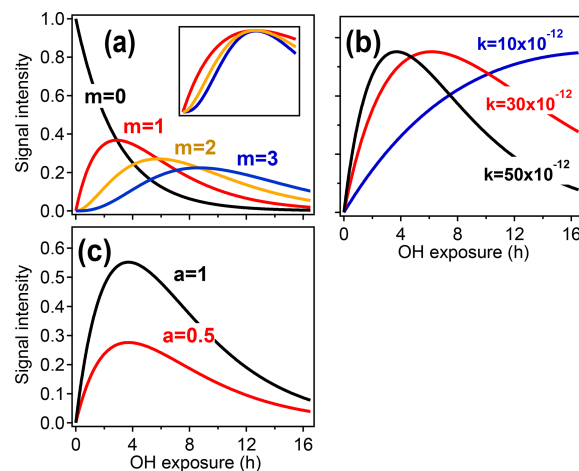


Figure 2. Illustration of the relationships between the different GKP parameters (m , k , and a) and the time dependence of a given species, using synthetic data. (a) Parameterizations with different generation m . In the subpanel, the traces with $m=2$ and $m=3$ have been scaled to allow comparison of the curvature, which differs with generation. (b) Parameterizations with different rate constant k . Increasing k does not change the shape of the curve but causes the maximum to occur at lower OH exposures. (c) Parameterizations with different scaling constant a , which changes neither curvature nor location of the maximum but only the height of the curve.

organic compound X that reacts with OH in the chamber can be parameterized by

$$[X_m](t) = a(k[\text{OH}]\Delta t)^m e^{-k[\text{OH}]\Delta t}, \quad (3)$$

where k is the second-order rate constant (units of $\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$), m is the number of reactions with OH needed to produce the compound (generation number), and $[\text{OH}]\Delta t$ is the integrated OH exposure (units of $\text{molecule s cm}^{-3}$). This parameterization is exact in the situation where all rate constants k in the system are equal and is an approximation otherwise, in which k is an effective rate constant representing the overall rate of reactions in the pathway.

Figure 2 illustrates how the parameters a , k , and m relate to the shape of the function described in Eq. (3). The parameter m (Fig. 2a) returns the generation number and is determined by the curvature of $[X]$ as $[\text{OH}]\Delta t \rightarrow 0$ (Zhou and Zhuang, 2007).

Equation (3) can be fit to time-dependent concentration (or ion intensity) data to return a , k , and m . The fitted value of m can be affected by noise or by fitting to a time step that is too long (Zhou and Zhuang, 2007). The optimum time step depends on the signal-to-noise ratio of the data and the compound's reaction rate but can be determined empirically. The fit can also be improved by integrating the data with respect to OH exposure over the experimental time period and fitting the integrated form of Eq. (3), which reduces random Gaussian noise (Sect. S1 in the Supplement). When all rate

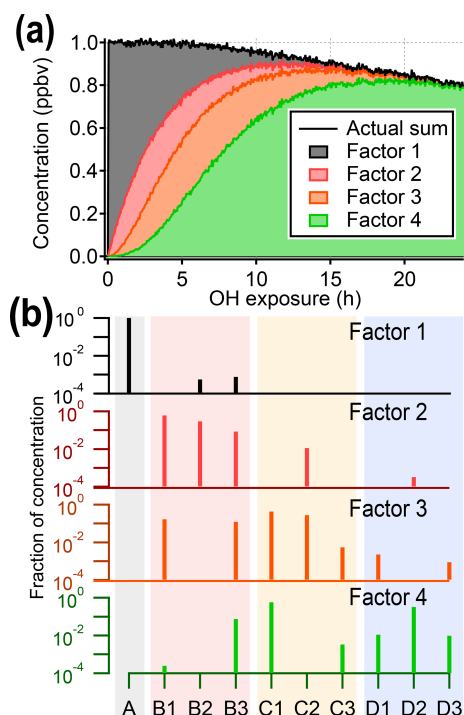


Figure 3. Results from PMF analysis of the synthetic dataset, showing the four-factor solution. **(a)** Total intensity of synthetic data compared to stacked time series of PMF factors. **(b)** Profiles of PMF factors, illustrating that factors do not correspond to individual generations. The shaded background corresponds to generation: precursor (black), first generation (red), second generation (yellow), and third generation (blue). The color of the mass spectra corresponds to panel (a). Solutions with different numbers of factors are given in the Supplement.

constants within a reaction sequence are not identical (which is typically the case), there is no direct analytical relationship between the effective rate constant k (Fig. 2b) and the individual rate constants in the pathway. However, the effective rate constant k provides a rough measure of the reactivity of the compound and its precursors. A higher effective k indicates higher formation and/or reaction rates and is affected by rate-limiting steps. The scaling constant a (Fig. 2c) ensures that the returned values of k and m are insensitive to instrument calibration and stoichiometric yields.

Compounds can be grouped by similar k and m in order to reduce the complexity of the dataset. The k , m , and average chemical properties of the group can be used to define a surrogate species. The choice of the number of groups and the method of grouping are subjective. GKP could be used alone, by binning compounds by similar k and m , or it could be used in combination with another analysis technique, such as HCA. Several approaches to using GKP to define surrogate species are discussed in Sect. 3.3.2.

3 Results and discussion

3.1 PMF

3.1.1 PMF of synthetic data

A set of PMF solutions for the synthetic data, including 2–10 factors, is shown in the Supplement (Fig. S5). The quality of the PMF reconstruction can be evaluated in two ways: the residual between the PMF reconstruction and the original data (lower residual indicates better agreement), and the normalized mutual information (NMI) (Vinh et al., 2010) between PMF factors and photochemical generation. The PMF residual is high for the 2-factor solution (13 % on average) and low for 3- to 10-factor solutions (less than 5 %).

The normalized mutual information metric describes the correlation between PMF factors and generation. A value of 0 means no correlation, and a value of 1 indicates that generations are perfectly assigned to distinct factors. Because species can be assigned to multiple factors, we used the relative intensities of each generation in each factor as input to the NMI calculation. For instance, if PMF factor 2 accounted for 66 % of the total integrated intensity of first-generation product B1, 97 % of the intensity of B2, and 12 % of the intensity of B3, we assigned a value of 1.75 for first-generation products to factor 2. The mutual information describes the probability that products of a particular generation are assigned to the same cluster. Mutual information must be normalized so that it can be compared between solutions with different numbers of factors or clusters. As the normalization factor, we used the arithmetic average of the generation and factor entropy, which is a quantity that describes the size and diversity of values in the two classification schemes (generation and PMF factor).

NMI values are provided in Table 1. For purposes of comparison, Table 1 also includes the NMI values calculated from hierarchical clustering analysis. HCA of the synthetic dataset is described in Sect. 3.2.1. Because there are only 10 species in the synthetic dataset, a solution with 10 groups, each of which contains a single species, has no correlation between generation and groups, and the NMI is zero.

Figure 3 shows the four-factor solution. The four PMF factors are able to reconstruct the total signal with excellent agreement, but they do not correspond to the four original generations of compounds (precursor plus three product generations). There is some relationship between early-, middle-, and late-generation species and the PMF factors (indicated by nonzero NMI values), but regardless of the selected rotational forcing, all PMF factors contain species from more than one generation. For instance, because both C1 and D2 are long-lived species, they are correlated over the time period of the experiment and so are assigned to the same factor. More importantly, many species are included in two or more PMF factors, despite being formed by only one pathway. Eight to 10 factors (approximately the number of species in

Table 1. Synthetic data. Normalized mutual information index quantifying the correlation between PMF factor or HCA cluster and photochemical generation.

Number of groups (PMF factors or HCA clusters)	PMF NMI	HCA NMI
2	0.402	0.397
3	0.381	0.467
4	0.436	0.521
5	0.427	0.683
6	0.442	0.745
7	0.761	0.835
8	0.733	0.799
9	0.679	0.756
10	0	0

the dataset) are needed to separate generations, which is not a useful simplification of the dataset (which is made up of only 10 species).

3.1.2 PMF of chamber data

Figure 4 illustrates positive matrix factorization of chamber data, including 463 individual calibrated product species from CIMS, optical, and AMS instruments; these exclude the precursor and overlapped species and are corrected for background and dilution. A three-factor, a four-factor, and a six-factor solution are shown. Additional solutions are shown in Fig. S6. In each of the solutions, a linear combination of PMF factors can reconstruct the measured intensity with negligible residual (also called reconstruction error) (within 10 ppbC, or about 2 %, for each solution, regardless of aging time). Each solution includes factors that peak in intensity at early, middle, and late times. There are no factors that retain a consistent time series or chemical profile between solutions with different numbers of factors, and in fact the time series do not have shapes consistent with chemical kinetics. Rather, each solution includes factors that peak in intensity at roughly regularly spaced intervals, apportioning the time series into discrete pieces (Fig. 4a). This suggests that the PMF factors are not chemically meaningful, even though the data are fit with low residual.

As in the PMF solution of the synthetic dataset, most species appear in the profiles of more than one factor (Fig. 4b). The time series of acetone (from calibrated m/z 59 $\text{C}_3\text{H}_6\text{OH}^+$ measured by PTR-MS), a species with a large signal and a long lifetime against OH, is shown in Fig. 4c as an example. As oxidative aging progresses, acetone and other long-lived species, including butadiene, acetic acid, and CO, are successively assigned to later-peaking factors, although mechanisms suggest that compounds such as butadiene are formed in the first one to two generations of reaction (Bloss et al., 2005a; Jenkin et al., 2003; Li and Wang, 2014). Relat-

edly, two compounds that are formed in the same generation but exhibit different reactivity are not necessarily assigned to the same factor.

The chemical composition of each PMF profile can be summarized by calculating the average carbon oxidation state and average number of carbon atoms per molecule in the factor (Fig. 5). The contribution of each species to the average is weighted by its intensity in the factor profile. As the precursor species becomes more oxygenated and fragments to smaller product species, the average composition moves towards CO and CO_2 , which are in Fig. 5b (Kroll et al., 2011). This trajectory is observed from early- to late-peaking PMF factors, as expected. Regardless of the number of factors chosen for the solution, the average chemical composition of each factor falls within the same range of oxidation state and molecular size. The various PMF factors appear to show the average composition of the mixture during early, moderate, and high OH exposures. This is consistent with the time series of PMF factors, which appear at discrete intervals (Fig. 4), and with the calculated average compositions of the mixture at specific time periods, which fall within the range of the PMF factors (Fig. 5). In other words, solutions with a larger number of factors do not add new groups of species not represented by solutions with a smaller number of factors, even though the PMF residuals are low.

We conclude that, in chamber experiments such as the one considered here, the PMF factors generally cannot be attributed to distinct chemical groups, oxidation generations, or chemical processes. Surrogate species derived from PMF factors do not have chemically realistic behavior or the same range of chemical properties as the original dataset. The information about the system that can be determined from PMF factors is the average composition during specific time periods of the experiment. The researcher must subjectively choose the number of factors. These factors are not chemically robust, and this should be considered when comparing PMF factors between oxidation experiments or chemical systems. PMF is certainly well suited for cases in which groups of compounds have distinct and constant composition (Ulbrich et al., 2009), such as field measurements near fresh emission sources and/or when using instruments that classify mixtures into a small number of types (e.g., the AMS). However, in a chamber oxidation experiment there are instead continuous, dynamic changes in composition as a function of time. Species created in the same oxidation generation often do not have similar time-series behavior, given differences in reactivity of different cogenerated species. This could be a useful first-level simplification of the data but suggests that PMF factors derived from chamber experiments cannot be used as surrogates for groups of reaction products within 3-D models because surrogate species should have a chemical behavior that emulates real species.

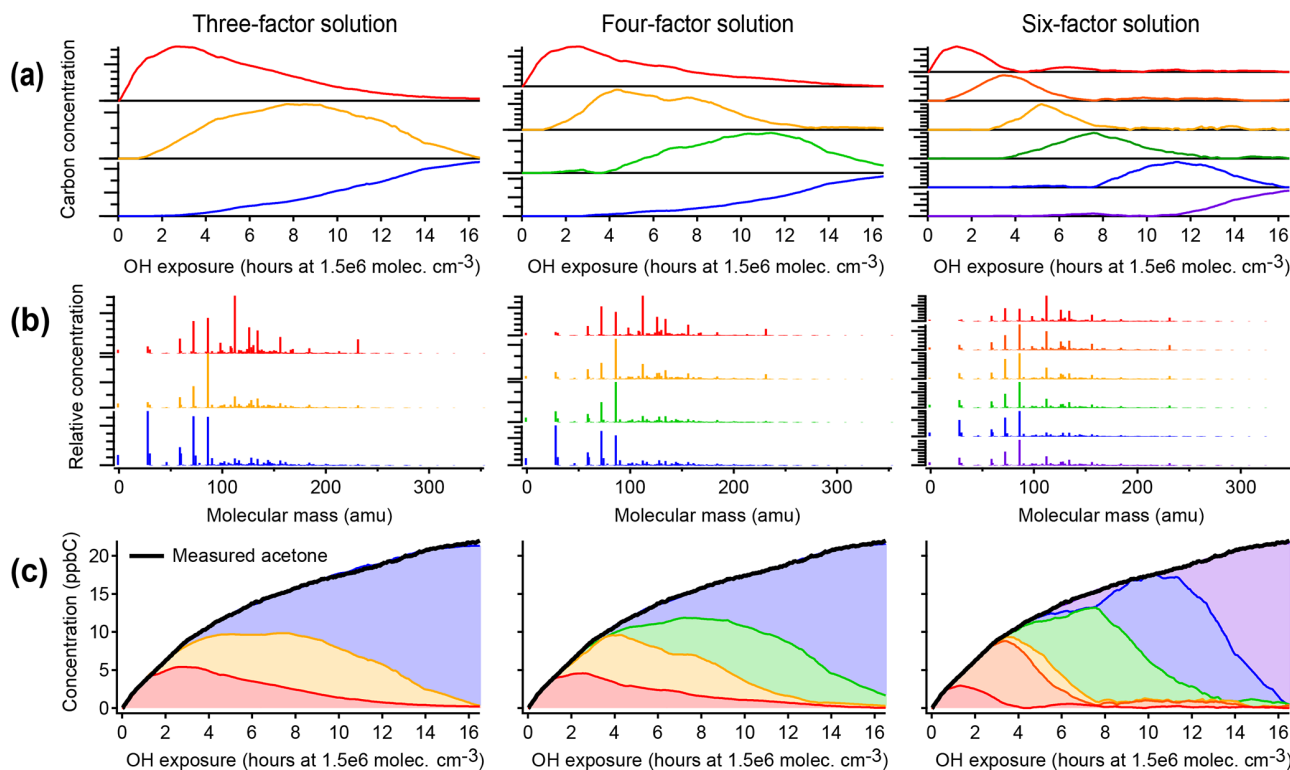


Figure 4. Positive matrix factorization of chamber data, showing solutions with three, four, and six factors. (a) Time series of PMF factors. (b) Compositional profiles of factors, shown as combined mass spectra from all instruments with CO, CH₂O, and CIMS measurements at their exact molecular masses and OA shown with a molecular mass of -1 . (c) Apportionment of the concentration of acetone (a long-lived oxidation product signal) across all factors. Within each column, the assigned color of each factor is consistent. As in the PMF analysis of the synthetic dataset (Fig. 3), factors do not correspond to generations, and long-lived species (such as acetone) are assigned to successively later peaking factors over the course of the time series.

3.2 HCA

Hierarchical clustering can be used to identify major chemical groups in processed data. This could be used to reduce the complexity of a dataset by analyzing the chemical properties of the clusters rather than individual species.

3.2.1 HCA of synthetic data

An example of the use of HCA to cluster chemical species within complex oxidation mixtures is shown in Fig. 6 using the synthetic dataset. Species D1 and D3, with very similar time-series behavior, are the two most closely related compounds and are assigned to cluster D*. The next two most similar groups are species D2 and cluster D*, which are assigned to a new, higher-level cluster. Species are clustered together until all have been grouped into a single cluster.

In this example with simulated data, HCA generally clusters together compounds of a similar generation, though not perfectly. HCA clusters together compounds that have similar time-series behavior, and time-series behavior is determined not only by generation, but also by formation and reaction rate constants. For example, species B1, B2, and C2

all have fast formation and reaction rates, resulting in similar time series. HCA groups these three species together. The algorithm further suggests that the first-generation products B1 and B2 are much more similar to one another than they are to the second-generation product C2.

The ability of HCA to separate compounds of different generations was quantified by the normalized mutual information (NMI). NMI values are provided in Table 1. For all solutions with more than two clusters (or factors), NMI values for HCA are higher than those of PMF, indicating that HCA more successfully sorts compounds by generation.

The results of HCA applied to synthetic data indicate several strengths and weaknesses of the HCA algorithm. Most importantly, the algorithm provides a clear way to visualize the behavior and relationships between all measurements in a dataset. The precursor compound can be included in the analysis, because data are normalized and the high intensity of the precursor does not skew the results. Compounds with similar kinetic properties are mostly grouped together, but some generational miscategorization still occurs. It may be difficult to use HCA to separate compounds which have dif-

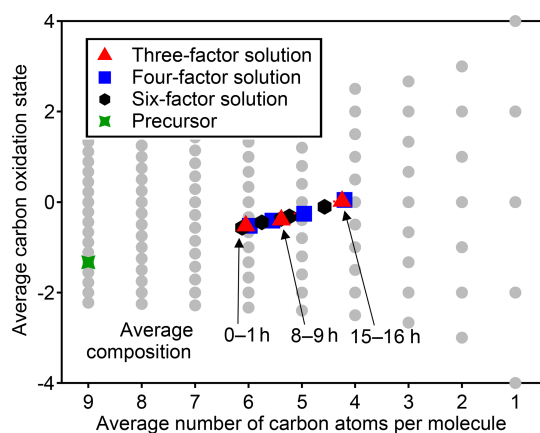


Figure 5. Average carbon oxidation state and number of carbon atoms per molecule in each PMF factor from analysis of chamber data for solutions with three, four, and six factors. Also noted is the average composition of the mixture during low (1 h atmospheric-equivalent aging), medium (8–9 h), and high (15–16 h) OH exposures. Factors cover a relatively small region in this chemical space, which is unaffected by the number of factors chosen for the solution.

ferent generation numbers but similar formation and reaction rates.

HCA can be used to simplify the dataset by replacing clusters of compounds with surrogates. If the surrogate time-series behavior is determined by averaging the time series of the individual members of the cluster, then the surrogate will have chemically realistic behavior. As noted previously, the researcher must subjectively choose the number of clusters.

3.2.2 HCA of chamber data

There are some significant differences between the synthetic dataset and real-world datasets collected from chamber experiments. Most importantly, the actual chamber experiment includes many more species (10 species in the synthetic system compared to thousands of detected ion masses and hundreds of measured species in the chamber experiment). The real chamber dataset includes many nonmeaningful measurements whose time series have no structure. Additionally, many species in the real-world dataset have much more similar time-series behavior to one another than any two of the species in the synthetic system. Conversely, there are also distinct outliers in the real-world dataset, whose time-series behavior does not resemble any other compound. HCA effectively separates meaningful from nonmeaningful measurements, groups together very similar compounds, and highlights outliers.

A diagram showing the hierarchical distance between all species measured in the chamber study is shown in Fig. 7. This dataset includes measured, calibrated, and background-subtracted species from all instruments and excludes overlaps. We use calibrated data here, but an advantage of this

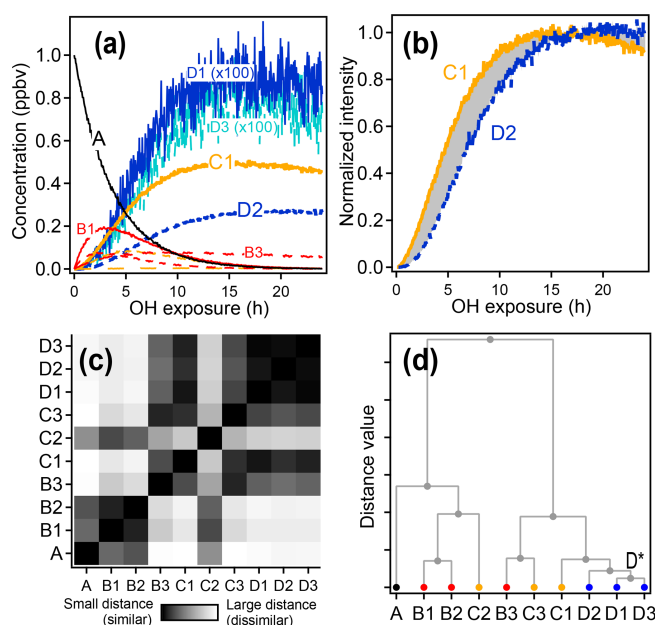


Figure 6. Hierarchical clustering procedure applied to synthetic data. First-, second-, and third-generation species are shown in red, yellow, and blue, respectively, and the precursor is shown in black. (a) Time series data. (b) Time series of species C1 and D2 normalized between 0 and 1. The gray shaded area is integrated to give the distance between the two time series. (c) Matrix showing the relative distance between each pair of species. (d) Hierarchical cluster relationship; D1 and D3 are the most similar species and so are the first to be clustered together (forming a new cluster D*).

method is that it is insensitive to calibration: data are normalized, and only relative behavior is important. In Fig. 7a, individual species are arrayed across the bottom, and their accumulation into clusters is denoted by gray lines linking species and clusters. As with PMF, the user must choose the number of groups (factors or clusters) in the solution. Here we have selected a maximum threshold relative distance that places the precursor, 1,2,4-trimethylbenzene, in a cluster separate from all product species. The individual clusters that fall below this threshold are distinguished by color in Fig. 7a. The resulting groups include 10 individual species that do not fall into a cluster (including the precursor, 1,2,4-trimethylbenzene) and nine clusters that incorporate at least two species. Figure 7b shows the time series of a selection of these clusters (all time series are included in Fig. S7). The cluster average was determined by summing the individual species contributors to the cluster, weighted by parts-per-billion carbon.

The chemical properties of each cluster, described as average oxidation state and average number of carbon atoms per molecule, are shown in Fig. 8. Clusters lie on a diagonal trajectory between the precursor and highly oxidized, small molecules (CO and CO₂), and clusters that peak earlier in time appear closer to the precursor. This indicates that

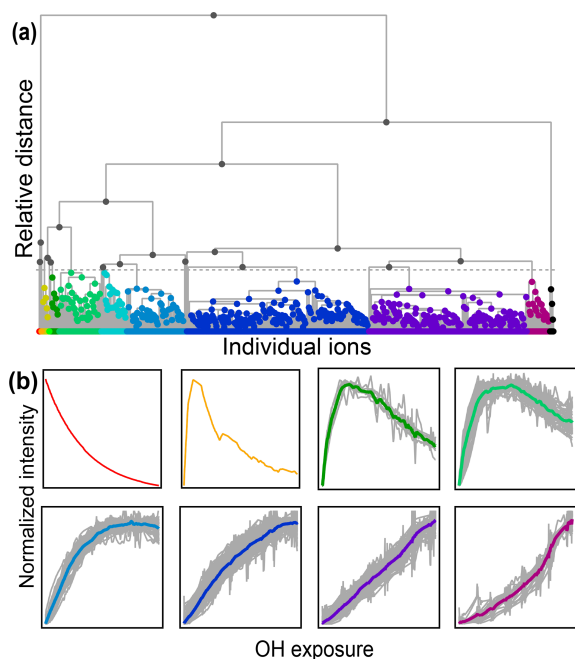


Figure 7. (a) Hierarchical cluster relationship of all measured species from the chamber experiment. Clusters are colored at a relative distance cutoff (gray dashed line) that separates 1,2,4-trimethylbenzene from all other products, with gray lines showing linkages between species and clusters. The individual clusters are distinguished by different colors. (b) Time series of eight example clusters. The x axis in each plot is OH exposure, and the y axis is the normalized intensity. The cluster average is shown by a thick colored line, and individual species contributors are shown as thinner gray lines. Colors correspond to those in panel (a).

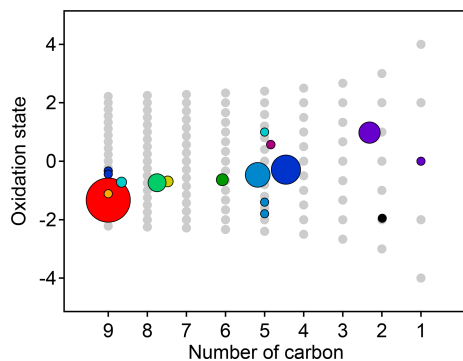


Figure 8. Average oxidation state and number of carbon atoms per molecule for each cluster determined from HCA of chamber data. The individual clusters are distinguished by color, and the color scheme is the same as in Fig. 7. The contribution of each species to the cluster average is weighted by parts-per-billion carbon (averaged over the entire experiment). The marker area is proportional to the averaged concentration (parts-per-billion carbon) of all species in the cluster, with the marker size of the precursor (red) decreased by a factor of 2 for legibility. Clusters cover a substantially wider area of chemical space than PMF factors (Fig. 5).

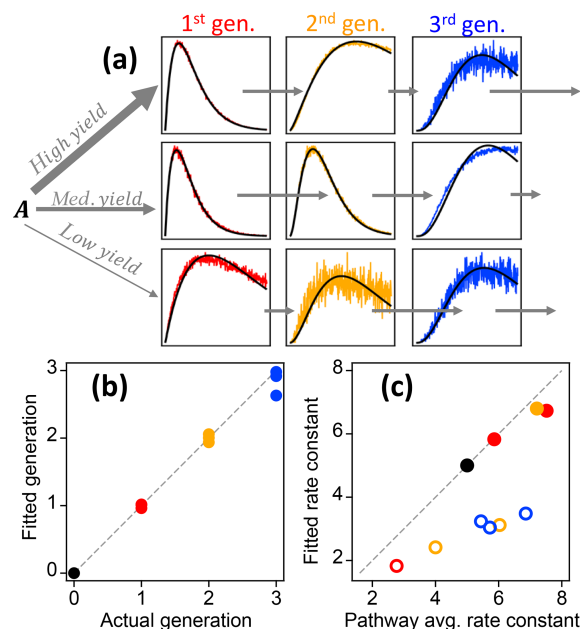


Figure 9. Best fit of the gamma kinetic parameterization to synthetic data (GKP, Eq. 3). (a) Time series of synthetic data (colored lines) and best fit (black lines). First-generation species are shown in red, second-generation species in yellow, and third-generation species in blue. The relative rate constants are indicated by short arrows (slow rate constant) or long arrows (fast rate constant). (b) Fitted generation compared to actual generation. The colors correspond to the generations shown in panel (a). (c) Effective rate constant compared to the average of the rate constants in the pathway that produces each particular species. Pathways that include slow steps are shown with open circles.

species with similar time-series behavior have similar chemical properties. Compared to the chemical properties of the PMF factors (Fig. 5), the clusters lie along the same diagonal trajectory but are substantially more varied in terms of average carbon number and oxidation state and cover a wider range of chemical space. As the threshold for separating clusters is lowered, resulting in more clusters with fewer species per cluster, a wider range of chemical properties is observed (Fig. S8). This is in contrast to PMF analysis, in which increasing numbers of factors does not increase the range of chemical properties (Fig. 5). As shown in Sect. 2.2.1, increasing the number of PMF factors provided the average composition of the mixture at more time points. HCA does not always separate generations perfectly (as can be seen in Table 1 and Fig. 6d), but the generational mixing is not as severe as with PMF and can be reduced by choosing a lower threshold for separating clusters.

The surrogate species derived from HCA clusters have chemically realistic behavior and have a similar range of chemical properties to the original dataset. As with PMF, the choice of the number of clusters is subjective. In addition to defining surrogate species, HCA can be used to visualize the

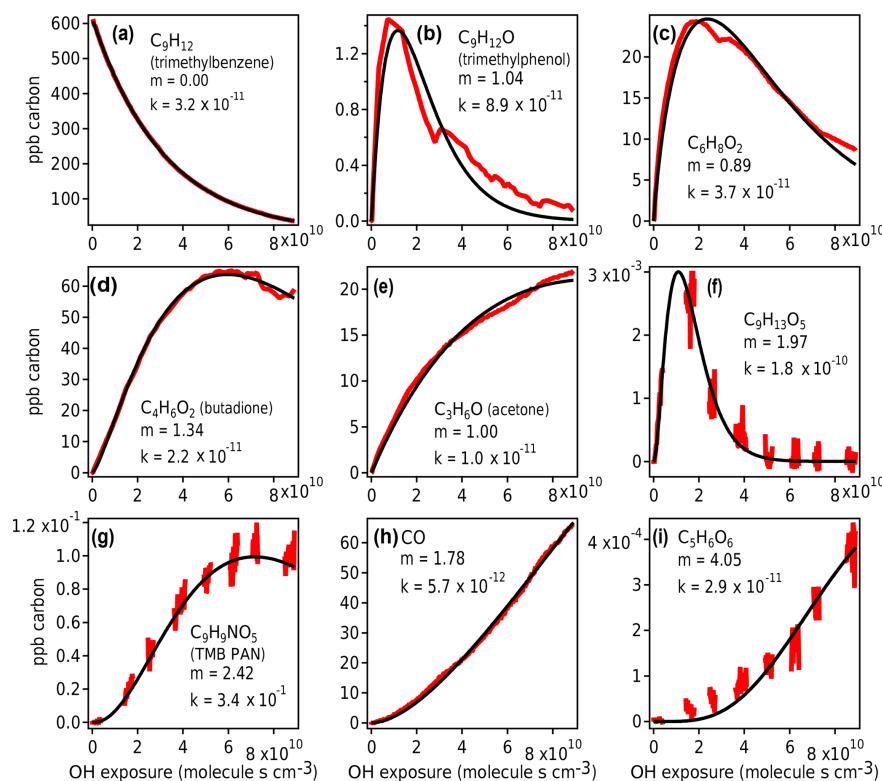


Figure 10. Measured species from chamber experiment (red) and GKP best fit (black). Data in panels (a), (c), and (e) are from Vocus-2R-PTR; in panels (b) and (d) from PTR3- H_3O^+ ; in panels (f), (g), and (i) from I^- CIMS; and in panel (h) from TILDAS. The data gaps in panels (f), (g), and (i) arise from the I^- CIMS instrument measuring particle-phase composition, measurements that are not considered in this work.

range of behavior and degree of similarity between all compounds in a dataset. The clustering algorithm is thus a viable approach for describing a continuum of kinetic behavior and chemical properties.

3.3 GKP

3.3.1 GKP fit to synthetic data

The gamma kinetics parameterization (GKP, Eq. 3) provides a method for determining kinetic and mechanistic information from chamber experiments. The parameterization returns an effective rate constant k and generation number m . To investigate the extent to which fitting kinetic data to Eq. (3) yields reasonable values for rate constants (k) and generation number (m), we first apply the parameterization to the synthetic dataset described in Sect. 2.1.2, which has known rates and generation numbers. Figure 9a shows the time series of synthetic data and the parameterized best fit, using the integrated signal as described in Sect. S1. The parameterization can reproduce a range of kinetic behavior, even in situations where the formation and loss rate constants k_m are very different (for which the assumption of uniform reactivities is poor). Figure 9b shows the fitted generation compared to the actual generation. The actual generation

numbers are correctly returned in all cases (with errors within 12 %). Figure 9c shows the parameterized k compared to actual pathway-average k_m rate constants in the pathway. The effective rate constant k cannot be calculated directly from the actual k_m in the system but is rather a best-fit value in the approximation of equal k_m . The returned values of k are in the same range as the actual k_m and are larger for pathways that generally involve faster rate constants. The average rate constant in a particular pathway and the fitted effective rate constant k are similar, except when the pathway includes a very slow step. In this case the fitted value of k is closer to that of the rate-limiting step (Fig. 9c). We conclude that the fit parameters m and k are reasonable, physically meaningful values that provide information on the kinetics of the system.

3.3.2 GKP fit to chamber data

The GKP was applied to the chamber data, with the time dependence of all measured compounds fit to Eq. (3). More than 95 % of measured compounds are fit with a correlation coefficient R^2 of 0.9 or higher, meaning the function generally describes the kinetic behavior of species measured in oxidation systems well. Examples of fitted chamber measure-

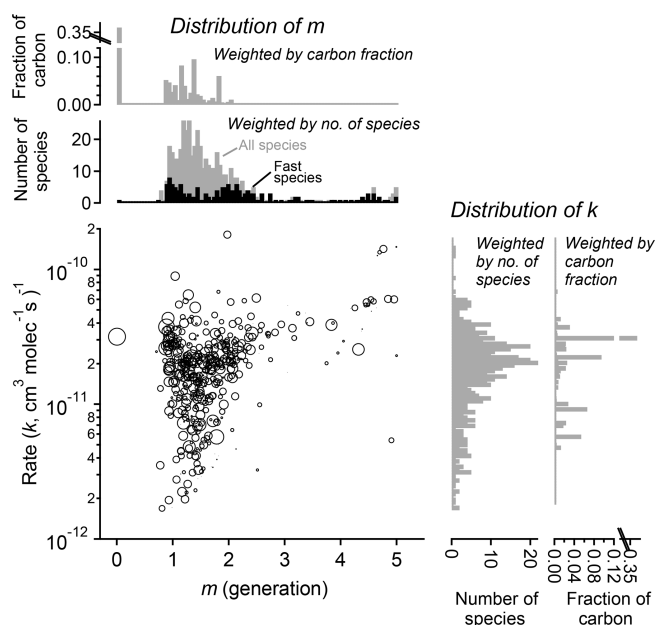


Figure 11. Parameterized rate constant and generation number for 463 species detected during the chamber experiment OH-initiated oxidation of trimethylbenzene. The marker area corresponds to $\log(\text{ppb carbon})$ of detected species, averaged over the duration of the experiment. Fast-reacting species, defined as having an effective rate constant at least 75 % that of the precursor, are highlighted as black bars in the histogram of m . These tend to center on integer values of generation number.

ments are shown in Fig. 10. In some cases, noninteger values of m are returned, which may occur for several reasons.

First, noise can contribute to uncertainty in m . At low generations ($m = 1$ – 2), the standard deviation of the fit is about 0.1, and at high generations ($m \geq 3$) it is somewhat higher, with standard deviation up to 0.8 (Fig. S9). Especially for measurements with low signal-to-noise ratios and limited data near the beginning of the experiment, m may not be fit with high precision. For example, the fits using $m = 3$ and $m = 5$ to $\text{C}_5\text{H}_6\text{O}_6$ (Fig. 10i) are not significantly worse than $m = 4$.

Second, the generation number can be distorted if the compound is produced by or reacts significantly via channels other than OH reaction (e.g., by ozone reaction, NO_3 radical reaction, or photolysis), in which case the assumption of linear, first-order kinetics with respect to OH exposure is not necessarily applicable. For example, $\text{C}_6\text{H}_8\text{O}_2$ (Fig. 10c) may correspond to 3,4-dimethyl-2(5H)-furanone (Bloss et al., 2005b), which reacts with O_3 under experimental conditions at a comparable rate to OH, or an unsaturated diketone (Li and Wang, 2014), which has a high photolysis rate. In Fig. 10b and c, the curves are also distorted due to repeat injections of HONO, which abruptly changes the NO concentration in the experiment and clearly affects the reaction of these compounds. Any of these processes can distort the

shape of the curve, making it more difficult to fit m correctly. Because m is related to the slow (rate-limiting) steps in a mechanism, specifically OH additions, it is not affected by faster radical chemistry such as autoxidation and intramolecular arrangements.

Finally, if the compound is produced by more than one pathway with a differing number of reaction steps, such as butadiene (Fig. 10d), the resulting generation parameter is noninteger. This is also demonstrated using a synthetic system in Fig. S10.

In addition, if physical (nonchemical) processes have a major influence on species concentrations, and occur on the same timescale as the chemical reaction, they may impact the fitted kinetic parameters. In particular, delays caused by strong interactions of gas-phase compounds with surfaces (chamber walls or instrument inlets) can shift the fitted m to higher values and the fitted k towards the time constant of the surface interaction. As noted above, the timescales of surface equilibration processes in the present experiments are < 15 s, which is much shorter than the timescales of the chemical changes observed. Thus, such processes are unlikely to affect the analysis of the present chamber results but could introduce substantial errors if they occur over longer timescales or are competing against much more rapid chemical transformations. GKP analysis is therefore only valid when the equilibration times of such processes are short compared to the timescales of the chemical processes being studied.

The fitted values of k and m for all species are shown in Fig. 11. The returned k 's fall within 1 order of magnitude of the OH rate constant of the precursor species ($k_{\text{TMB}} = 3.2 \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$). Most m 's are between 1 and 2, meaning most measured compounds are produced after one or two reaction steps (assuming OH is the dominant oxidant). When the data are restricted to fast-reacting compounds, major modes at integer values of m are observed (black bars in Fig. 11). However, when all compounds are considered, major modes at integer values are not observed, which suggests that many compounds are formed by more than one pathway and/or have significant reactions with O_3 or another oxidant. The generation numbers of compounds with $m \geq 4$ are less certain due to data gaps, limited experimental duration, and low signal-to-noise ratio in the fits. Higher-generation ($m > 2$) compounds are uniformly the fast-reacting (high k) species. Conversely, no species are observed with high m (> 2) and low k . This area of the diagram corresponds to slow-forming, slow-reacting species that are created after multiple OH additions; such species are unlikely to be formed at observable concentrations within the time-frame of the experiment. Were the experiment to be run at higher OH exposures, it is possible that these species would be observed as well.

The kinetic parameters derived from fitting the gamma distribution are correlated with individual species' chemical composition. Figure 12 shows that species that involve the fastest reactions (high values of the effective rate constant,

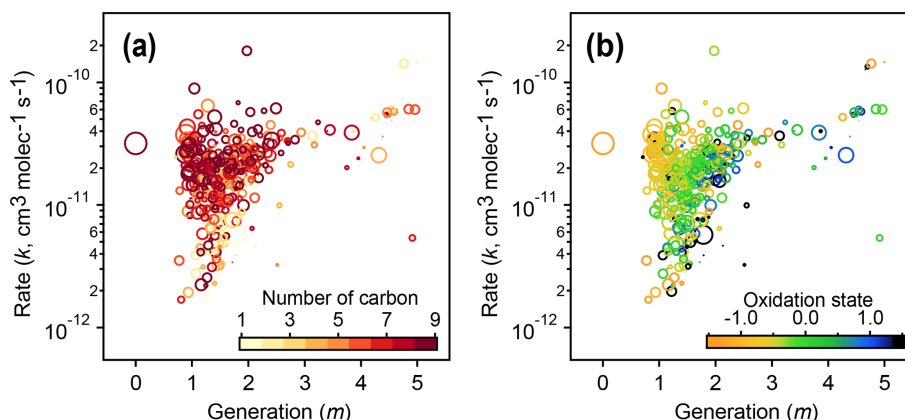


Figure 12. Relationships of kinetic parameters (from the GKP of chamber data) with key chemical properties of reactive species. **(a)** Generation (m) and rate (k) values of 1,2,4-trimethylbenzene precursor and products, colored by number of carbon atoms. **(b)** Same as **(a)** but with k and m colored by carbon oxidation state. The marker area corresponds to $\log(\text{ppb carbon})$. The early-generation and fast-reacting products tend to have higher numbers of carbon atoms and are less oxidized, while later-generation and slow-reacting products tend to be smaller and more oxidized.

k) and earliest formation (lowest values of m) tend to be large and relatively unoxidized, with oxidation states similar to that of the 1,2,4-trimethylbenzene precursor. Species that form or react slowly (low values of k) or that form in later generations (higher values of m) tend to be smaller and more oxidized.

3.3.3 Clustering of GKP results

The GKP can be used not only to describe individual species, but also to group compounds and reduce the complexity of the system. If compounds are grouped by similar k and m , compounds in the group will have similar chemical composition and similar kinetic behavior, and the chemical and kinetic properties of the groups will include a range of variability similar to the individual species. Here we test three methods of using GKP to group compounds: (1) fitting the GKP to time series of HCA-derived clusters, (2) using HCA to cluster compounds based on their GKP-derived time series (based on fitted values k and m), and (3) using fixed bins to group compounds based on k and m . Groups derived from PMF analysis cannot be fit with the GKP because the factor time series are not consistent with chemical kinetics.

Results from each approach, showing both kinetic characteristics (k and m) and chemical properties (oxidation state and carbon number) of each group, are given in Fig. 13, which includes an overview and comparison of grouped species derived from PMF (Fig. 13a), HCA (Fig. 13b), and GKP (Fig. 13c and d). Figure 13b shows results from applying the GKP to HCA data. For each of the nine HCA clusters (described in Sect. 3.2.2), the GKP was fit to the cluster's average time series, determined from a carbon-weighted average of the time series of all individual species in the cluster. This provided values of k and m for each cluster. (For the 10 species that did not fit into any cluster, the k and

m of these were determined as well.) Figure 13c shows the reversed approach, the application of HCA to GKP results. Here, the time series of each individual species was fit with GKP, and the distances between the time series of the best fits were determined and used as input into the HCA algorithm. The k and m of the resulting cluster were calculated by averaging the k and m of the individual compounds in the cluster, weighted by parts-per-billion carbon. A potential advantage of this approach is that the GKP fitting reduces the noise of the signals used in HCA analysis, possibly allowing for more precise determinations of clusters. Finally, shown in Fig. 13d are results from an alternate approach for grouping compounds by GKP parameters (k and m), binning all the species by their values of k and m . This is analogous to the 2-D volatility basis set developed by Donahue et al. (2011, 2012), which bins species based on saturation mass concentration and O : C ratio.

Surrogate species defined by GKP have by definition kinetically realistic behavior. The resulting groups of compounds have a range of chemical properties similar to that of the original dataset, regardless of whether they are grouped using HCA or grouped by similar k and m . The method of grouping is subjective, as is the choice of the number of clusters (if HCA is used) or the number of bins (if compounds are grouped by similar k and m). A particular strength of GKP is the resulting kinetic characterization of each compound. The effective rate constant and generation number provide new information that can be used to assess proposed mechanisms or to guide the reactive behavior of surrogate species in a model.

3.4 Comparison of PMF, HCA, and GKP

A comparison of compound groups derived from PMF, HCA, and GKP is also shown in Fig. 13. Included in this figure are

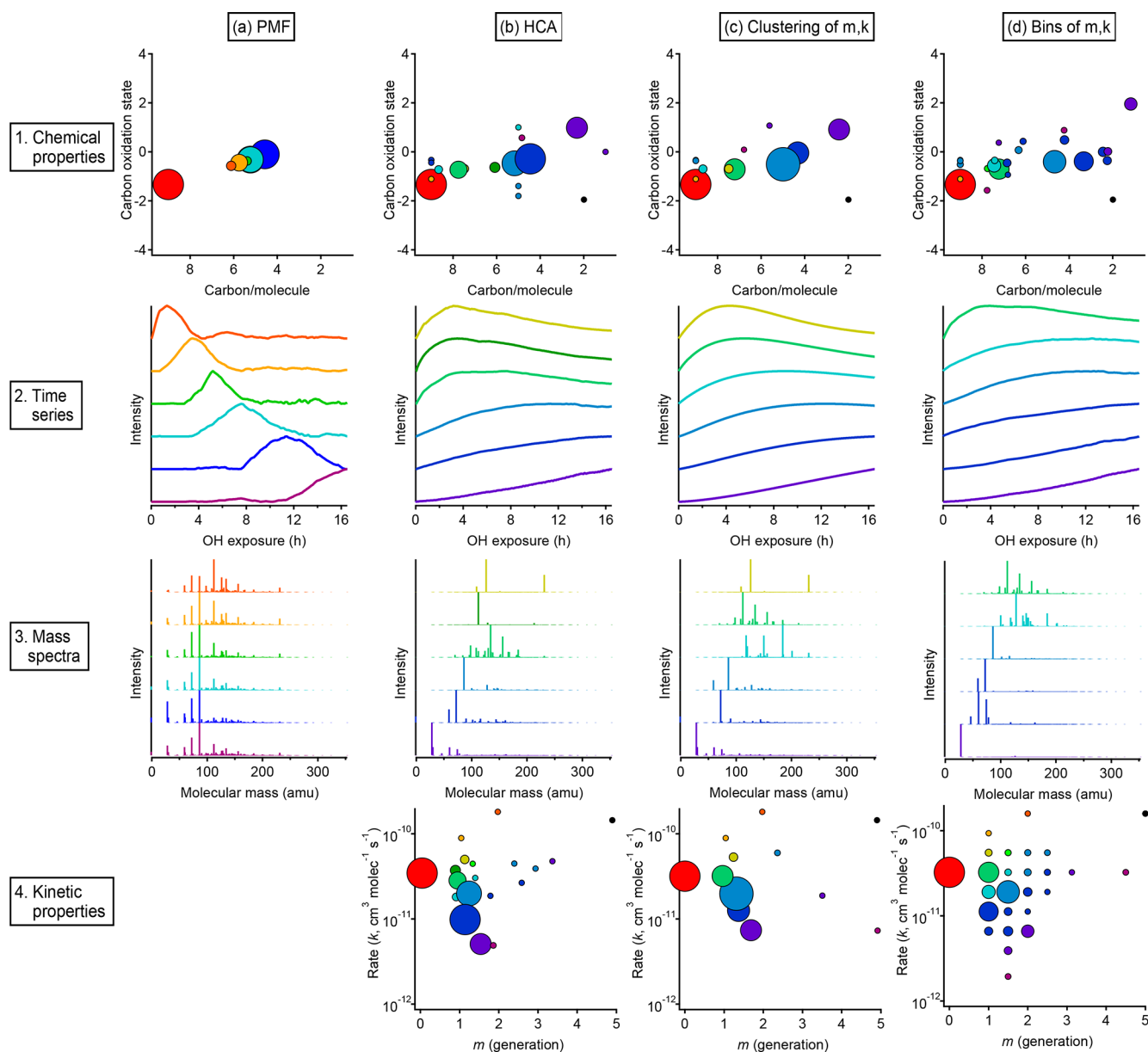


Figure 13. Overall comparison of groups derived from PMF, HCA, and GKP of chamber data. The columns show, from left to right, the results of (a) PMF, (b) HCA, (c) GKP best fits grouped using HCA, and (d) measurements grouped by GKP fit parameters. The rows show, from top to bottom, (1) the average carbon oxidation state and number of carbon atoms per molecule for each group, (2) the time series of the six groups containing the most carbon, (3) the mass spectra of those six groups, and (4) the rate constant and generation number of each group. Within each column, each chemical group is assigned a specific color. This color scheme is the same for each plot within a column. The marker area is proportional to the averaged concentration (ppb carbon) of all species in the group, with the marker size of the precursor (red) decreased by a factor of 2 for legibility. The marker area scheme is consistent across all plots. PMF factors do not have kinetically realistic time series; therefore, there is no plot a4.

the chemical properties (oxidation state vs. number of carbon atoms), time series, mass spectra, and kinetic properties (k vs m) of the compound groups. For each technique, solutions with different numbers of groups are possible. Figure 13 shows the solution discussed most extensively in the text: the six-factor solution for PMF; the HCA solution with

nine major clusters; and the two GKP solutions discussed in Sect. 3.3.3, which have seven major clusters and 25 bins, respectively. For clarity, the time series and mass spectra for only six groups derived from HCA and GKP are shown. These six groups contain cumulatively about 80 % of the total product carbon in the system.

In all cases, the majority of the carbon can be represented by a manageable number of groups. The relationship between oxidation state and number of carbon per molecule is similar, regardless of the grouping technique. The PMF factors have a smaller range of chemical properties than chemical groupings derived from HCA or GKP. The range of chemical properties is similar for HCA and GKP. The time series of PMF factors are clearly different from those of HCA- and GKP-derived groups and have non-kinetically realistic shapes with sharp maxima.

The PMF factors each contain many more compounds than the groups derived from HCA or GKP. Many of the same compounds are consistently grouped together by HCA and GKP, regardless of whether HCA, HCA of GKP, or binning of GKP is used. Additionally, the range of kinetic properties and the locations of major compound groups in kinetic space are similar between the HCA and GKP approaches. This reproducibility suggests that these are chemically meaningful compound groupings. Some groups derived from HCA or GKP contain only a single species. These could be chemically important compounds whose unique behavior should be considered when modeling the system; conversely, they could be measurement outliers which should be discarded. The interpretation of these species is subjective.

Regardless, the combination of fitting using the GKP and grouping based on kinetic behavior may provide a viable approach for greatly simplifying the time-dependent behavior of complex mixtures of reaction products in a laboratory oxidation system.

4 Conclusions

Hundreds to thousands of individual chemical species can be produced in a typical organic photooxidation chamber experiment. This chemical complexity presents a number of analytical challenges, including organizing and processing large mass spectrometric datasets, identifying major groups of compounds, providing kinetic and mechanistic information, and simplifying the chemistry in a way that can be implemented in large-scale regional and global models.

In this paper, we evaluated three methods to simplify a description of atmospheric chemistry in chamber studies. The methods explored include positive matrix factorization (PMF), which represents data as a linear sum of factors; hierarchical clustering analysis (HCA), which describes the similarity of species in terms of their time-series behavior; and the gamma kinetics parameterization (GKP), which characterizes species in terms of the effective rate constant and generation. All three approaches require a subjective choice of the number of compound groups.

Because PMF is so widely used in atmospheric chemistry to characterize organic aerosol and for source apportionment in field studies, it is important to understand how oxidation systems are described by PMF. We found that PMF analy-

sis of the chamber experiment described here did not sort species into clear generations, since different species formed in a single generation can exhibit highly variable reactivities. Oxidized factors appearing in PMF analysis of chamber studies, and in ambient air, may be able to reproduce observations as a linear sum of a fresh factor and a highly aged factor with low residual, but these factors do not necessarily represent distinct chemical groups. This is because PMF assumes constant factor composition, which is useful when distinguishing fresh emission sources but does not apply to evolving oxidation systems.

Hierarchical clustering, which also does not depend on calibration, can be used to quickly identify major groups of ions and patterns of behavior. The derived clusters maintain more chemical information (including average oxidation state and molecular size) than do PMF factors. HCA is therefore useful to identify chemically meaningful ions in mass spectrometry data and to group compounds into a smaller number of groups with consistent chemical characteristics.

A continuum of kinetic behavior is observed and can be described using the gamma kinetics parameterization of individual species (or clusters of species). The parameterization is derived from first-order kinetics and thus provides a physically meaningful fit to the kinetics of the species. The two returned parameters, effective rate constant and generation number, correlate with oxidation state and molecular size. The parameterization provides a way to derive mechanistic information from an oxidation system, in addition to describing chemical composition.

Future directions of this work include evaluation of mechanisms, mechanism development, and applications to lumping schemes in models. The current analysis is based on two systems, a synthetic system and a chamber experiment, and more work is needed to see how these analysis approaches perform with other systems. The gamma kinetics parameterization can be used to support complex chemical mechanisms by determining whether the experimentally determined generation and rate constants are consistent with a proposed pathway or mechanism. Further, with well-calibrated, high-quality laboratory data, it may be possible to derive yields, formation rate constants, and reaction rate constants separately, which would be invaluable in model and mechanism development. Finally, HCA-derived clusters, or groups of compounds with a similar effective rate constant and generation, could be used as surrogates or lumps in aerosol or air quality models as an experimentally supported way of simplifying a complex system.

Data availability. Data are available through the Kroll group publications website, <http://krollgroup.mit.edu/publications.html> (Kroll, 2020). Both datasets (chamber data and the synthetic dataset) are included in this repository as comma-separated-value (csv) files available for public download.

Supplement. The supplement related to this article is available online at: <https://doi.org/10.5194/acp-20-1021-2020-supplement>.

Author contributions. ARK, MRC, AZ, JEK, MB, KN, CL, JCR, and JRR collected and analyzed data. ARK implemented PMF and HCA algorithms, developed the GKP analysis, and wrote the manuscript. FNK and JHK provided project guidance. All authors were involved in helpful discussion and contributed to the manuscript.

Competing interests. The authors declare that they have no conflict of interest.

Financial support. This work was supported by NSF grant AGS-1638672. We additionally acknowledge the Harvard Global Institute for funding. ARK acknowledges support from the Dreyfus postdoctoral program. MB acknowledges support from the Austrian science fund 10 (FWF), Erwin-Schrödinger-Stipendium, grant J-3900.

Review statement. This paper was edited by Joel Thornton and reviewed by two anonymous referees.

References

- Abeleira, A., Pollack, I. B., Sive, B., Zhou, Y., Fischer, E. V., and Farmer, D. K.: Source characterization of volatile organic compounds in the Colorado Northern Front Range Metropolitan Area during spring and summer 2015, *J. Geophys. Res.-Atmos.*, 122, 3595–3613, <https://doi.org/10.1002/2016JD026227>, 2017.
- Aumont, B., Szopa, S., and Madronich, S.: Modelling the evolution of organic carbon during its gas-phase tropospheric oxidation: development of an explicit model based on a self generating approach, *Atmos. Chem. Phys.*, 5, 2497–2517, <https://doi.org/10.5194/acp-5-2497-2005>, 2005.
- Bar-Joseph, Z., Gifford, D. K., and Jaakkola, T. S.: Fast optimal leaf ordering for hierarchical clustering, *Bioinformatics*, 17, S22–S29, https://doi.org/10.1093/bioinformatics/17.suppl_1.S22, 2001.
- Bloss, C., Wagner, V., Jenkin, M. E., Volkamer, R., Bloss, W. J., Lee, J. D., Heard, D. E., Wirtz, K., Martin-Reviejo, M., Rea, G., Wenger, J. C., and Pilling, M. J.: Development of a detailed chemical mechanism (MCMv3.1) for the atmospheric oxidation of aromatic hydrocarbons, *Atmos. Chem. Phys.*, 5, 641–664, <https://doi.org/10.5194/acp-5-641-2005>, 2005a.
- Bloss, C., Wagner, V., Bonzanini, A., Jenkin, M. E., Wirtz, K., Martin-Reviejo, M., and Pilling, M. J.: Evaluation of detailed aromatic mechanisms (MCMv3 and MCMv3.1) against environmental chamber data, *Atmos. Chem. Phys.*, 5, 623–639, <https://doi.org/10.5194/acp-5-623-2005>, 2005b.
- Breitenlechner, M., Fischer, L., Hainer, M., Heinritzi, M., Curtius, J., and Hansel, A.: PTR3: An Instrument for Studying the Lifecycle of Reactive Organic Carbon in the Atmosphere, *Anal. Chem.*, 89, 5824–5831, <https://doi.org/10.1021/acs.analchem.6b05110>, 2017.
- Brown-Steiner, B., Selin, N. E., Prinn, R., Tilmes, S., Emmons, L., Lamarque, J.-F., and Cameron-Smith, P.: Evaluating simplified chemical mechanisms within present-day simulations of the Community Earth System Model version 1.2 with CAM4 (CESM1.2 CAM-chem): MOZART-4 vs. Reduced Hydrocarbon vs. Super-Fast chemistry, *Geosci. Model Dev.*, 11, 4155–4174, <https://doi.org/10.5194/gmd-11-4155-2018>, 2018.
- Cappa, C. D. and Wilson, K. R.: Multi-generation gas-phase oxidation, equilibrium partitioning, and the formation and evolution of secondary organic aerosol, *Atmos. Chem. Phys.*, 12, 9505–9528, <https://doi.org/10.5194/acp-12-9505-2012>, 2012.
- Carter, W. P. L.: A detailed mechanism for the gas-phase atmospheric reactions of organic compounds, *Atmos. Environ.*, 24, 481–518, 1990.
- Crassier, V., Suhre, K., Tulet, P., and Rosset, R.: Development of a reduced chemical scheme for use in mesoscale meteorological models, *Atmos. Environ.*, 34, 2633–2644, [https://doi.org/10.1016/S1352-2310\(99\)00480-X](https://doi.org/10.1016/S1352-2310(99)00480-X), 2000.
- Craven, J. S., Yee, L. D., Ng, N. L., Canagaratna, M. R., Loza, C. L., Schilling, K. A., Yatavelli, R. L. N., Thornton, J. A., Ziemann, P. J., Flagan, R. C., and Seinfeld, J. H.: Analysis of secondary organic aerosol formation and aging using positive matrix factorization of high-resolution aerosol mass spectra: application to the dodecane low-NO_x system, *Atmos. Chem. Phys.*, 12, 11795–11817, <https://doi.org/10.5194/acp-12-11795-2012>, 2012.
- Cubison, M. J. and Jimenez, J. L.: Statistical precision of the intensities retrieved from constrained fitting of overlapping peaks in high-resolution mass spectra, *Atmos. Meas. Tech.*, 8, 2333–2345, <https://doi.org/10.5194/amt-8-2333-2015>, 2015.
- DeCarlo, P. F., Kimmel, J. R., Trimborn, A., Northway, M. J., Jayne, J. T., Aiken, A. C., Gonin, M., Fuhrer, K., Horvath, T., Docherty, K. S., Worsnop, D. R., and Jimenez, J. L.: Field-Deployable, High-Resolution, Time-of-Flight Aerosol Mass Spectrometer, *Anal. Chem.*, 78, 8281–8289, <https://doi.org/10.1021/AC061249N>, 2006.
- Donahue, N. M., Epstein, S. A., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set: 1. organic-aerosol mixing thermodynamics, *Atmos. Chem. Phys.*, 11, 3303–3318, <https://doi.org/10.5194/acp-11-3303-2011>, 2011.
- Donahue, N. M., Kroll, J. H., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set – Part 2: Diagnostics of organic-aerosol evolution, *Atmos. Chem. Phys.*, 12, 615–634, <https://doi.org/10.5194/acp-12-615-2012>, 2012.
- Fortenberry, C. F., Walker, M. J., Zhang, Y., Mitroo, D., Brune, W. H., and Williams, B. J.: Bulk and molecular-level characterization of laboratory-aged biomass burning organic aerosol from oak leaf and heartwood fuels, *Atmos. Chem. Phys.*, 18, 2199–2224, <https://doi.org/10.5194/acp-18-2199-2018>, 2018.
- Gery, M., W., Whitten, G. Z., Killus, J. P., and Dodge, M. C.: A photochemical kinetics mechanism for urban and regional scale computer modeling, *J. Geophys. Res.-Atmos.*, 94, 12925–12956, 1989.
- Glasiu, M. and Goldstein, A. H.: Recent Discoveries and Future Challenges in Atmospheric Organic Chemistry, *Environ. Sci. Technol.*, 50, 2754–2764, <https://doi.org/10.1021/acs.est.5b05105>, 2016.

- Goldstein, A. H. and Galbally, I. E.: Known and Unexplored Organic Constituents in the Earth's Atmosphere, *Environ. Sci. Technol.*, 41, 1514–1521, <https://doi.org/10.1021/es072476p>, 2007.
- Houweling, S., Dentener, F., and Lelieveld, J.: The impact of nonmethane hydrocarbon compounds on tropospheric photochemistry, *J. Geophys. Res.-Atmos.*, 103, 10673–10696, <https://doi.org/10.1029/97JD03582>, 1998.
- Hunter, J. F., Carrasquillo, A. J., Daumit, K. E., and Kroll, J. H.: Secondary Organic Aerosol Formation from Acyclic, Monocyclic, and Polycyclic Alkanes, *Environ. Sci. Technol.*, 48, 10227–10234, <https://doi.org/10.1021/es502674s>, 2014.
- IPCC: Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, edited by: Core Writing Team, Pachauri, R. K., and Meyer, L. A., IPCC, Geneva, Switzerland, 151 pp., 2014.
- Isaacman-VanWertz, G., Massoli, P., O'Brien, R. E., Nowak, J. B., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., Su, L., Knopf, D. A., Misztal, P. K., Arata, C., Goldstein, A. H., and Kroll, J. H.: Using advanced mass spectrometry techniques to fully characterize atmospheric organic carbon: current capabilities and remaining gaps, *Faraday Discuss.*, 200, 579–598, <https://doi.org/10.1039/C7FD00021A>, 2017.
- Jenkin, M. E., Saunders, S. M., Wagner, V., and Pilling, M. J.: Protocol for the development of the Master Chemical Mechanism, MCM v3 (Part B): tropospheric degradation of aromatic volatile organic compounds, *Atmos. Chem. Phys.*, 3, 181–193, <https://doi.org/10.5194/acp-3-181-2003>, 2003.
- Jimenez, P., Baldasano, J. M., and Dabdub, D.: Comparison of photochemical mechanisms for air quality modeling, *Atmos. Environ.*, 37, 4179–4194, [https://doi.org/10.1016/S1352-2310\(03\)00567-3](https://doi.org/10.1016/S1352-2310(03)00567-3), 2003.
- Junninen, H., Ehn, M., Petäjä, T., Luosujärvi, L., Kotiaho, T., Koskinen, R., Rohner, U., Gonin, M., Fuhrer, K., Kulmala, M., and Worsnop, D. R.: A high-resolution mass spectrometer to measure atmospheric ion composition, *Atmos. Meas. Tech.*, 3, 1039–1053, <https://doi.org/10.5194/amt-3-1039-2010>, 2010.
- Krechmer, J., Lopez-Hilfiker, F., Koss, A., Hutterli, M., Stoermer, C., Deming, B., Kimmel, J., Warneke, C., Holzinger, R., Jayne, J., Worsnop, D., Fuhrer, K., Gonin, M., and de Gouw, J.: Evaluation of a New Reagent-Ion Source and Focusing Ion-Molecule Reactor for Use in Proton-Transfer-Reaction Mass Spectrometry, *Anal. Chem.*, 90, 12011–12018, <https://doi.org/10.1021/acs.analchem.8b02641>, 2018.
- Krechmer, J. E., Pagonis, D., Ziemann, P. J., and Jimenez, J. L.: Quantification of Gas-Wall Partitioning in Teflon Environmental Chambers Using Rapid Bursts of Low-Volatility Oxidized Species Generated in Situ, *Environ. Sci. Technol.*, 50, 5757–5765, <https://doi.org/10.1021/acs.est.6b00606>, 2016.
- Kroll, J.: Kroll Group: Publications, available at: <http://krollgroup.mit.edu/publications.html>, last access: 23 January 2020.
- Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K. R., Altieri, K. E., Mazzoleni, L. R., Wozniak, A. S., Bluhm, H., Mysak, E. R., Smith, J. D., Kolb, C. E., and Worsnop, D. R.: Carbon oxidation state as a metric for describing the chemistry of atmospheric organic aerosol, *Nat. Chem.*, 3, 133–139, <https://doi.org/10.1038/nchem.948>, 2011.
- Landrigan, P. J., Fuller, R., Acosta, N. J. R., Adeyi, O., Arnold, R., Basu, N. (Nil), Baldé, A. B., Bertollini, R., Bose-O'Reilly, S., Boufford, J. I., Breyse, P. N., Chiles, T., Mahidol, C., Coll-Seck, A. M., Cropper, M. L., Fobil, J., Fuster, V., Greenstone, M., Haines, A., Hanrahan, D., Hunter, D., Khare, M., Krupnick, A., Lanphear, B., Lohani, B., Martin, K., Mathiasen, K. V., McTeer, M. A., Murray, C. J. L., Ndahimananjara, J. D., Perera, F., Potočník, J., Preker, A. S., Ramesh, J., Rockström, J., Salinas, C., Samson, L. D., Sandilya, K., Sly, P. D., Smith, K. R., Steiner, A., Stewart, R. B., Suk, W. A., van Schayck, O. C. P., Yadama, G. N., Yumkella, K., and Zhong, M.: The Lancet Commission on pollution and health, *Lancet*, 391, 462–512, [https://doi.org/10.1016/S0140-6736\(17\)32345-0](https://doi.org/10.1016/S0140-6736(17)32345-0), 2018.
- Lane, T. E., Donahue, N. M., and Pandis, S. N.: Simulating secondary organic aerosol formation using the volatility basis-set approach in a chemical transport model, *Atmos. Environ.*, 42, 7439–7451, <https://doi.org/10.1016/J.ATMOSENV.2008.06.026>, 2008.
- Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., Kurtén, T., Worsnop, D. R., and Thornton, J. A.: An Iodide-Adduct High-Resolution Time-of-Flight Chemical-Ionization Mass Spectrometer: Application to Atmospheric Inorganic and Organic Compounds, *Environ. Sci. Technol.*, 48, 6309–6317, <https://doi.org/10.1021/es500362a>, 2014.
- Li, Y. and Wang, L.: The atmospheric oxidation mechanism of 1,2,4-trimethylbenzene initiated by OH radicals, *Phys. Chem. Chem. Phys.*, 16, 17908, <https://doi.org/10.1039/C4CP02027H>, 2014.
- Lopez-Hilfiker, F. D., Iyer, S., Mohr, C., Lee, B. H., D'Ambro, E. L., Kurtén, T., and Thornton, J. A.: Constraining the sensitivity of iodide adduct chemical ionization mass spectrometry to multifunctional organic molecules using the collision limit and thermodynamic stability of iodide ion adducts, *Atmos. Meas. Tech.*, 9, 1505–1512, <https://doi.org/10.5194/amt-9-1505-2016>, 2016.
- Marcolli, C., Canagaratna, M. R., Worsnop, D. R., Bahreini, R., de Gouw, J. A., Warneke, C., Goldan, P. D., Kuster, W. C., Williams, E. J., Lerner, B. M., Roberts, J. M., Meagher, J. F., Fehsenfeld, F. C., Marchewka, M., Bertman, S. B., and Middlebrook, A. M.: Cluster Analysis of the Organic Peaks in Bulk Mass Spectra Obtained During the 2002 New England Air Quality Study with an Aerodyne Aerosol Mass Spectrometer, *Atmos. Chem. Phys.*, 6, 5649–5666, <https://doi.org/10.5194/acp-6-5649-2006>, 2006.
- Massoli, P., Stark, H., Canagaratna, M. R., Krechmer, J. E., Xu, L., Ng, N. L., Mauldin, R. L., Yan, C., Kimmel, J., Misztal, P. K., Jimenez, J. L., Jayne, J. T., and Worsnop, D. R.: Ambient Measurements of Highly Oxidized Gas-Phase Molecules during the Southern Oxidant and Aerosol Study (SOAS) 2013, *ACS Earth Sp. Chem.*, 2, 653–672, <https://doi.org/10.1021/acsearthspacechem.8b00028>, 2018.
- Müller, M., Graus, M., Wisthaler, A., Hansel, A., Metzger, A., Dommen, J., and Baltensperger, U.: Analysis of high mass resolution PTR-TOF mass spectra from 1,3,5-trimethylbenzene (TMB) environmental chamber experiments, *Atmos. Chem. Phys.*, 12, 829–843, <https://doi.org/10.5194/acp-12-829-2012>, 2012.
- Müllner, D.: Modern hierarchical, agglomerative clustering algorithms, available at: <http://arxiv.org/abs/1109.2378> (last access: 8 November 2018), 2011.

- Murphy, D. M., Middlebrook, A. M., and Warshawsky, M.: Cluster Analysis of Data from the Particle Analysis by Laser Mass Spectrometry (PALMS) Instrument, *Aerosol Sci. Technol.*, 37, 382–391, <https://doi.org/10.1080/02786820300971>, 2003.
- Paatero, P.: Least squares formulation of robust non-negative factor analysis, *Chemom. Intell. Lab. Syst.*, 37, 23–35, [https://doi.org/10.1016/S0169-7439\(96\)00044-5](https://doi.org/10.1016/S0169-7439(96)00044-5), 1997.
- Paatero, P.: User's guide for positive matrix factorization programs PMF2.EXE and PMF3.EXE, 2007.
- Pankow, J. F. and Barsanti, K. C.: The carbon number-polarity grid: A means to manage the complexity of the mix of organic compounds when modeling atmospheric organic particulate matter, *Atmos. Environ.*, 43, 2829–2835, <https://doi.org/10.1016/J.ATMOSENV.2008.12.050>, 2009.
- Pogliani, L., Berberan-Santos, M. N., and Martinho, J. M. G.: Matrix and convolution methods in chemical kinetics, *J. Math. Chem.*, 20, 193–210, <https://doi.org/10.1007/BF01165164>, 1996.
- Rebotier, T. P. and Prather, K. A.: Aerosol time-of-flight mass spectrometry data analysis: A benchmark of clustering algorithms, *Anal. Chim. Acta*, 585, 38–54, <https://doi.org/10.1016/J.ACA.2006.12.009>, 2007.
- Rosati, B., Teiwes, R., Kristensen, K., Bossi, R., Skov, H., Glasius, M., Pedersen, H. B., and Bilde, M.: Factor analysis of chemical ionization experiments: Numerical simulations and an experimental case study of the ozonolysis of α -pinene using a PTR-ToF-MS, *Atmos. Environ.*, 199, 15–31, <https://doi.org/10.1016/J.ATMOSENV.2018.11.012>, 2019.
- Sánchez-López, J. A., Zimmermann, R., and Yerezian, C.: Insight into the Time-Resolved Extraction of Aroma Compounds during Espresso Coffee Preparation: Online Monitoring by PTR-ToF-MS, *Anal. Chem.*, 86, 11696–11704, <https://doi.org/10.1021/ac502992k>, 2014.
- Sánchez-López, J. A., Wellinger, M., Gloess, A. N., Zimmermann, R., and Yerezian, C.: Extraction kinetics of coffee aroma compounds using a semi-automatic machine: On-line analysis by PTR-ToF-MS, *Int. J. Mass Spectrom.*, 401, 22–30, <https://doi.org/10.1016/J.IJMS.2016.02.015>, 2016.
- Sarkar, C., Sinha, V., Sinha, B., Panday, A. K., Rupakheti, M., and Lawrence, M. G.: Source apportionment of NMVOCs in the Kathmandu Valley during the SusKat-ABC international field campaign using positive matrix factorization, *Atmos. Chem. Phys.*, 17, 8129–8156, <https://doi.org/10.5194/acp-17-8129-2017>, 2017.
- Saunders, S. M., Jenkin, M. E., Derwent, R. G., and Pilling, M. J.: Protocol for the development of the Master Chemical Mechanism, MCM v3 (Part A): tropospheric degradation of non-aromatic volatile organic compounds, *Atmos. Chem. Phys.*, 3, 161–180, <https://doi.org/10.5194/acp-3-161-2003>, 2003.
- Sauvage, S., Plaisance, H., Locoge, N., Wroblewski, A., Coddeville, P., and Galloo, J. C.: Long term measurement and source apportionment of non-methane hydrocarbons in three French rural areas, *Atmos. Environ.*, 43, 2430–2441, <https://doi.org/10.1016/J.ATMOSENV.2009.02.001>, 2009.
- SciPy.org: `scipy.cluster.hierarchy.linkage`, available at: <https://docs.scipy.org/doc/scipy/reference/generated/scipy.cluster.hierarchy.linkage.html> (last access: 23 December 2019), 2018.
- Shao, P., An, J., Xin, J., Wu, F., Wang, J., Ji, D., and Wang, Y.: Source apportionment of VOCs and the contribution to photochemical ozone formation during summer in the typical industrial area in the Yangtze River Delta, China, *Atmos. Res.*, 176–177, 64–74, <https://doi.org/10.1016/J.ATMOSENV.2016.02.015>, 2016.
- Smith, J. D., Kroll, J. H., Cappa, C. D., Che, D. L., Liu, C. L., Ahmed, M., Leone, S. R., Worsnop, D. R., and Wilson, K. R.: The heterogeneous reaction of hydroxyl radicals with sub-micron squalane particles: a model system for understanding the oxidative aging of ambient aerosols, *Atmos. Chem. Phys.*, 9, 3209–3222, <https://doi.org/10.5194/acp-9-3209-2009>, 2009.
- Stockwell, W. R., Kirchner, F., Kuhn, M., and Seefeld, S.: A new mechanism for regional atmospheric chemistry modeling, *J. Geophys. Res.-Atmos.*, 102, 25847–25879, 1997.
- Stojić, A., Stanišić Stojić, S., Mijić, Z., Šoštarić, A., and Rajšić, S.: Spatio-temporal distribution of VOC emissions in urban area based on receptor modeling, *Atmos. Environ.*, 106, 71–79, <https://doi.org/10.1016/J.ATMOSENV.2015.01.071>, 2015.
- Ulbrich, I. M., Canagaratna, M. R., Zhang, Q., Worsnop, D. R., and Jimenez, J. L.: Interpretation of organic components from Positive Matrix Factorization of aerosol mass spectrometric data, *Atmos. Chem. Phys.*, 9, 2891–2918, <https://doi.org/10.5194/acp-9-2891-2009>, 2009.
- Vinh, N. X., Epps, J., and Bailey, J.: Information Theoretic Measures for Clusterings Comparison: Variants, Properties, Normalization, and Correction for Chance, *J. Mach. Learn. Res.*, 11, 2837–2854, 2010.
- Wang, H. L., Chen, C. H., Wang, Q., Huang, C., Su, L. Y., Huang, H. Y., Lou, S. R., Zhou, M., Li, L., Qiao, L. P., and Wang, Y. H.: Chemical loss of volatile organic compounds and its impact on the source analysis through a two-year continuous measurement, *Atmos. Environ.*, 80, 488–498, <https://doi.org/10.1016/J.ATMOSENV.2013.08.040>, 2013.
- Wilson, K. R., Smith, J. D., Kessler, S. H., and Kroll, J. H.: The statistical evolution of multiple generations of oxidation products in the photochemical aging of chemically reduced organic aerosol, *Phys. Chem. Chem. Phys.*, 14, 1468–1479, <https://doi.org/10.1039/C1CP22716E>, 2012.
- Yan, C., Nie, W., Äijälä, M., Rissanen, M. P., Canagaratna, M. R., Massoli, P., Junninen, H., Jokinen, T., Sarnela, N., Häme, S. A. K., Schobesberger, S., Canonaco, F., Yao, L., Prévôt, A. S. H., Petäjä, T., Kulmala, M., Sipilä, M., Worsnop, D. R., and Ehn, M.: Source characterization of highly oxidized multifunctional compounds in a boreal forest environment using positive matrix factorization, *Atmos. Chem. Phys.*, 16, 12715–12731, <https://doi.org/10.5194/acp-16-12715-2016>, 2016.
- Yuan, B., Shao, M., de Gouw, J., Parrish, D. D., Lu, S., Wang, M., Zeng, L., Zhang, Q., Song, Y., Zhang, J., and Hu, M.: Volatile organic compounds (VOCs) in urban air: How chemistry affects the interpretation of positive matrix factorization (PMF) analysis, *J. Geophys. Res.-Atmos.*, 117, D24302, <https://doi.org/10.1029/2012JD018236>, 2012.
- Zaytsev, A., Breitenlechner, M., Koss, A. R., Lim, C. Y., Rowe, J. C., Kroll, J. H., and Keutsch, F. N.: Using collision-induced dissociation to constrain sensitivity of ammonia chemical ionization mass spectrometry (NH_4^+ CIMS) to oxygenated volatile organic compounds, *Atmos. Meas. Tech.*, 12, 1861–1870, <https://doi.org/10.5194/amt-12-1861-2019>, 2019.

- Zhang, Q., Jimenez, J. L., Canagaratna, M. R., Ulbrich, I. M., Ng, N. L., Worsnop, D. R., and Sun, Y.: Understanding atmospheric organic aerosols via factor analysis of aerosol mass spectrometry: a review, *Anal. Bioanal. Chem.*, 401, 3045–3067, <https://doi.org/10.1007/s00216-011-5355-y>, 2011.
- Zhang, Y., Chen, Y., Sarwar, G., and Schere, K.: Impact of gas-phase mechanisms on Weather Research Forecasting Model with Chemistry (WRF/Chem) predictions: Mechanism implementation and comparative evaluation, *J. Geophys. Res.-Atmos.*, 117, D01301, <https://doi.org/10.1029/2011JD015775>, 2012.
- Zhou, Y. and Zhuang, X.: Kinetic Analysis of Sequential Multistep Reactions, *J. Phys. Chem. B*, 111, 13600–13610, <https://doi.org/10.1021/JP073708+>, 2007.