

# Surveillance of SO<sub>2</sub> and NO<sub>2</sub> from ship emissions by MAX-DOAS measurements and the implications regarding fuel sulfur content compliance

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Abstract. Due to increased concerns regarding air pollutants emitted from shipping, feasible technology for the surveillance of these pollutants is in high demand. Here, we present shore-based multi-axis differential optical absorption spectroscopy (MAX-DOAS) measurements of SO<sub>2</sub> and NO2 emitted from ships under different traffic conditions in China's ship emission control areas (ECAs) in Shanghai and Shenzhen, China. Three typical measurement sites were selected in these two regions to represent the following emission scenarios: ships docked at berth, ships navigating in an inland waterway and inbound/outbound ships in a deep-water port. Using 2-D scanning, the observations show that SO<sub>2</sub> and NO<sub>2</sub> hot spots can be quickly and easily located from multiple berths. Although MAX-DOAS measurements can not distinguish plumes from specific ships in the busy shipping lanes of the inland waterway area, they certify that variations in the SO<sub>2</sub> and NO<sub>2</sub> levels are mainly impacted by the ship traffic density and the atmospheric dispersion conditions. In the open water area, which has a lower vessel density, MAX-DOAS measurements can capture the pulse signal of ship-emitted SO<sub>2</sub> and NO<sub>2</sub> very well; they can also characterize the peak's altitude and the insistent duration of the individual ship plumes. Combined with the ship activity data, information on the rated power of the engine and the fuel sulfur content, it was found that the SO<sub>2</sub>/NO<sub>2</sub> ratio in a single plume is usually low (< 1.5) for inbound vessels due to the usage of the auxiliary engine, which has less power and uses "clean" fuel with a low sulfur content. Thus,

an unexpectedly high  $SO_2/NO_2$  ratio implies the use of fuel with a sulfur content exceeding the regulation limits. Therefore, the observed  $SO_2/NO_2$  ratio in the plume of a single ship can be used as an index to indicate compliance (or noncompliance) with respect to the fuel sulfur content, and the suspicious ship can then be flagged for further enforcement. Combining the ship emissions estimated by actual operation parameters and the logical sulfur content, shore-based MAX-DOAS measurements will provide a fast and more accurate way to surveil ship emissions.

# 1 Introduction

Sulfur dioxide (SO<sub>2</sub>) and nitrogen dioxide (NO<sub>2</sub>) are important air pollutants, and are also recognized as non-negligible components of ship emissions (Corbett et al., 1999; Endresen et al., 2003; Eyring et al., 2010; Matthias et al., 2010). Both of these species can engage in atmospheric chemical reactions to produce aerosols and acid rain, and also have negative effects on air quality, the climate system and human health, as well as contributing to the acidification of terrestrial and marine ecosystems (Berglen et al., 2004; Seinfeld and Pandis, 2006; Singh, 1987). Moreover, NO<sub>2</sub> is also the key substance involved in the formation of photochemical smog (Dimitriades, 1972). With the rapid growth of marine transportation, air pollution has become the most challenging environmental issue in the shipping industry, specifically regarding the emission of CO<sub>2</sub>, SO<sub>2</sub>, NO<sub>x</sub>, particulate matter and greenhouse gases. CO<sub>2</sub> and NO are the main pollutants emitted by ships, and NO is rapidly converted to NO<sub>2</sub> by reaction with O<sub>3</sub> (Eyring et al., 2005; Becagli et al., 2012; Coggon et al., 2012; Diesch et al., 2013; Lauer et al., 2007; Seyler et al., 2017). Eyring et al. (2010) reported that ships contribute 15% of global NO<sub>x</sub> emissions and 4%–9% of SO<sub>2</sub> emissions.

Regarding the spatial distribution of these pollutants, global hot spots with high  $SO_2$  and  $NO_2$  emissions have been identified in the Eastern China Sea and the Southern China Sea, in marine areas in southeastern and southern Asia, in the Red Sea, in the Mediterranean, in the North Atlantic near the European coast, and along the western coast of North America, among others (Johansson et al., 2017).

In China, ship-emitted pollutants play important roles regarding the air quality, human health and climate (Lai et al., 2013; Liu et al., 2016; Yang et al., 2007). They not only affect the air quality in coastal areas, but even influence inland areas hundreds of kilometers from emission sources (Lv et al., 2018). Port cities are most affected by ship pollution, followed by cities along rivers. For example, ship-emitted SO<sub>2</sub> and NO<sub>x</sub> were shown to contribute 12.4 % and 11.6 % of the total emissions of the whole city of Shanghai in 2012, respectively, while up to 64 % of primary PM<sub>2.5</sub> has been attributed to ships at Shanghai Port; this pollution can then also be transported to inland regions (Fan et al., 2016; Zhao et al., 2013).

In order to reduce the negative impacts of ship emissions, the EU and the US have implemented regulations in an effort to decrease ship emissions (Kattner et al., 2015); among these regulations, fuel quality regulation has proven to be the most effective measure for addressing the issue of sulfur oxide  $(SO_x)$  and particulate matter (PM) emission in many countries. In addition, the International Maritime Organization (IMO, 2009) has also set up multiple emission control areas (ECAs) worldwide. By 2020, the maximum fuel sulfur content (FSC) is set to be 0.50 % S m/m worldwide; it is worth noting that the current maximum FSC in ECAs on the US coast and in Europe is 0.10 % S m/m, whereas in China's ECAs it is 0.50 % S m/m. The regulations also set limits regarding pollutant emissions, such as  $NO_x$  and  $CO_2$ , in exhaust gas. Thus, an exhaust gas treatment system could be another option to reduce the emissions of these species. Since 1 January 2017, ships berthed at the core ports of three designated domestic emission control areas (DECAs) in the Pearl River Delta (PRD), the Yangtze River Delta (YRD) and the Bohai Rim (Beijing-Tianjin-Hebei area) in China have been restricted to fuels with a sulfur content of less than or equal to 0.50 % S m/m (MOT, 2015). However, as of 1 January 2019 ships entering the ECAs are restricted to fuels with a sulfur content of no more than 0.50 % S m/m, regardless of whether they are sailing or docking. Currently, scrubbers are also an alternative method of reducing ship emissions in China. As a consequence, it is obvious that a reliable and practical monitoring system is urgently required for the implementation of ECA regulations.

The most commonly used options for monitoring ship emissions can be classified into two categories: estimates based on activity data or written documentation, and measurements of fuel samples and exhaust gas made onboard the ship. Basically, the continuous online monitoring of fuel and exhaust gas onboard is the most effective and accurate means of supervision, but is much less feasible in practice. With respect to the regulatory party, fuel sampling and document inspection are currently the most commonly used supervision measures, and the sulfur content in the fuel is usually quickly detected after a ship has docked (in roughly 10 min). Other technical methods have also been developed to determine both  $SO_2$  and  $NO_x$  emissions, including a new type of ship exhaust gas detection technology that mounts a portable sniffer instrument to a ship or helicopter (Beecken et al., 2015; Berg et al., 2012; Murphy et al., 2009; Villa et al., 2016). Alternatively, shore-based remote sensing is another effective way of measuring ship plumes and estimating the sulfur content in these plumes as ships pass in shipping lanes or dock at berth (Kattner et al., 2015; Seyler et al., 2017).

Remote sensing techniques have the advantages of fast detection, easy operation and high automation. In addition to the passive "sniffing" method using in situ instrumentation, optical remote sensing techniques, such as differential optical absorption spectroscopy (DOAS), light detection and ranging (LIDAR) and the ultraviolet camera (UV-CAM) technique, can detect the variation of light properties after interaction with the exhaust plume and, thus, the corresponding SO<sub>2</sub> and NO<sub>2</sub> emission levels in the plume (Balzani Lööv et al., 2014; Seyler et al., 2017). LIDAR systems can be used to retrieve a 2-D concentration distribution by scanning through the ship plume and by combing the wind and concentration profiles to obtain the ship emissions. McLaren et al. (2012) employed an active long-path DOAS technique to measure the  $SO_2/NO_2$  ratios in ship plumes and speculated on their relationship with the sulfur content of fuels. A UV camera has also been successfully applied to measure the SO<sub>2</sub> concentrations and emission rates of moving and stationary ship plumes (Prata, 2014).

The DOAS technique allows for the identification and quantification of the absorption of a variety of species that show characteristic absorption features in the measured wavelength range (Platt and Stutz, 2008). It has been widely used for trace gas measurements over several decades, especially for very mature NO<sub>2</sub> and SO<sub>2</sub> (Edner et al., 1993; Mellqvist and Rosén, 1996; Platt et al., 1979). As an expanded apparatus, multi-axis differential optical absorption spectroscopy (MAX-DOAS) measurements are highly sensitive to aerosols and trace gases in the lower troposphere, as the instrument observes scattered sunlight at different viewing angles close to the horizontal and zenith directions (Hön-

ninger et al., 2004; Ma et al., 2013; Sinreich et al., 2005; Wang et al., 2014). Due to its portability, the MAX-DOAS instrument can be carried onto ships to observe the vertical column densities (VCDs) of NO2 and SO2 during the cruise, as high levels of pollutants have been found close to busy ports and dense shipping lanes (Hong et al., 2018; Schreier et al., 2015; Takashima et al., 2012; Tan et al., 2018). In addition, MAX-DOAS has been successfully employed for the direct monitoring of shipping emissions. Premuda et al. (2011) used ground-based MAX-DOAS measurements to evaluate the NO<sub>2</sub> and SO<sub>2</sub> levels in ship plumes discharged from individual ships in the Giudecca Strait of the Venetian Lagoon. Seyler et al. (2017) utilized MAX-DOAS to perform long-term measurements of NO2 and SO2 from shipping emissions in the German Bight, and evaluated the reduction in SO<sub>2</sub> levels after the implementation of stricter fuel sulfur limits.

In this study, shore-based MAX-DOAS measurements were employed to measure the NO<sub>2</sub> and SO<sub>2</sub> in ship plumes in different ship traffic environments in Shanghai and Shenzhen, China. Combined with photographs taken by the instrument's camera and AIS (Automatic Identification System) information, it was verified that the emissions of ships at berth, those from ships passing through an inland waterway and those from ships in open-sea areas can be successfully detected. The measurements can also provide fuel sulfur content information for individual ships by comparing the  $NO_2$  and  $SO_2$  emitted in the plume (considering the effects) of plume age) with fuel sample analysis and the ship activity data. This suggests that the shore-based MAX-DOAS method is a feasible and reliable method for surveilling shipemitted SO<sub>2</sub> and NO<sub>2</sub> and, further, enables compliance or noncompliance to be established with respect to fuel sulfur content regulations.

#### 2 Measurements and method

#### 2.1 Instrument setup and sites

In this study, the MAX-DOAS instrument was designed and assembled by the authors (Zhang et al., 2018). It mainly consists of a receiving telescope, a spectrometer (Ocean Optics, QE65 Pro) and a computer to control the measurements. Driven by a 2-D stepper motor system, the telescope can collect scattered sunlight from different elevation angles in the vertical direction and different azimuth angles in the horizontal direction. The lens then converges the scattered light onto the fiber bundle connected to the spectrometer. Following this, the sunlight received is dispersed by a grating, detected by a CCD detector and recorded by the spectrometer covering the wavelength range from 300 to 480 nm with a resolution about 0.5 nm full width at half maximum (FWHM). As a new design feature, a camera has been configured on the MAX-DOAS apparatus that moves coaxially with the re-

**Figure 1.** The YRD and PRD domestic emission control areas (DE-CAs) in China, and the locations of the MAX-DOAS measurements in the coastal cities of Shanghai and Shenzhen: Waigaoqiao Port and Wusong Wharf in Shanghai, and Yantian Port in Shenzhen. The viewing direction of instrument azimuth angle is indicated using a red arrow. Panels (**b**) to (**d**) are sourced from ©Google Earth.

ceiving telescope and can record the scene and sky conditions viewed by the telescope. The scanning of the telescope can be set in a sequence of several elevation angles that begins close to horizontal and at  $90^{\circ}$  and then moves to next azimuth angle for another vertical scanning sequence. Due to the different ship traffic conditions, the types of ship passing in inland waterways and at seaside ports are different with respect to size and tonnage. Therefore, the configuration of the observing geometric angles was adjusted depending on the ship conditions (see Table 1).

The MAX-DOAS measurements of ship emissions were performed in two typical port cities – Shanghai and Shenzhen – in China. As shown in Fig. 1a, sea areas surrounding Shanghai and Shenzhen are located in the Yangtze River Delta and Pearl River Delta ECAs, respectively. In Shanghai, two different ship traffic scenarios were considered: ships at berth in the Waigaoqiao container port area ( $31.36^{\circ}$  N,  $121.58^{\circ}$  E; Fig. 1b), and ships passing through the inland waterway in the downstream region of Huangpu River around Wusong ( $31.37^{\circ}$  N,  $121.50^{\circ}$  E; Fig. 1c). In Shenzhen, the measurements were carried out in the deep-water port of Yantian ( $114.29^{\circ}$  E,  $22.56^{\circ}$  N; Fig. 1d). More details about the environments and operational configurations with respect to the measurements are given in Table 1.



110°E

115°E

120°E

Sites	Locations and periods	Operations*	Environment types
Waigaoqiao, Shanghai	31.36° N, 121.58° E; 28/08/2017	AZ: 26 to 34° ELE: 3, 4, 5, 6 and 7° Spectrum temporal resolution: 15–30 s Completed scanning cycle: 15 min	Viewing the berths. Ships: container ships
Wusong, Shanghai	31.37° N, 121.50° E; 30/12/2017–18/05/2018	AZ: 85° ELE: 0, 1, 2, 3, 4, 5, 6, 8 and 65° Spectrum temporal resolution: 40 s Completed scanning cycle: 7 min	Viewing the inland waterway with a high traffic volume. Ships: a wide variety of ships and small in size
Yantian, Shenzhen	114.29° E, 22.56° N; 23/05/2018–30/06/2018	AZ: 75° ELE: 2, 3, 5, 7, 10, 15, 30 and 90° Spectrum temporal resolution: 60 s Completed scanning cycle: 9 min	Viewing open-sea areas with a lower traffic volume. Ships: primarily container ships

 Table 1. Measurement details and the operational configurations of the MAX-DOAS.

\* AZ denotes the azimuth angle in the horizontal direction; ELE denotes the elevation angle in the vertical direction.

#### 2.2 DOAS spectral analysis

The DOAS algorithm is basically based on the Beer-Lambert law, which describes the extinction of radiation through the atmosphere (Platt and Stutz, 2008). The MAX-DOAS instrument observes scattered sunlight from various viewing directions and records the spectrum (e.g., Hönninger et al., 2004; Sinreich et al., 2005). The spectral analysis generates the measured SCD (slant column density), defined as the integral of the trace gas concentration along the entire optical path including the SCDs in the troposphere and the stratosphere (e.g., Platt and Stutz, 2008; Wagner et al., 2010). The stratospheric absorption is assumed to be at the same level in all spectra taken within one scan cycle; therefore, we generally choose the spectrum with the lowest less trace gas absorption as the reference (such as the spectrum measured in zenith direction). The slant column concentration of the trace gas measured at each lower elevation angle ( $\alpha$ ) is represented by the DSCD (differential SCD), which is the gas information of the measured slant column densities minus background densities in the reference spectrum:

$$DSCD(\alpha) = SCD(\alpha) - SCD(ref)$$
  
= SCD(\alpha)\_{trop} + SCD\_{strat} - SCD(ref)\_{trop}  
- SCD\_{strat}  
= SCD(\alpha)\_{trop} - SCD(ref)\_{trop}. (1)

Based on the DOAS principle, the measured scattered sunlight spectra are analyzed using the QDOAS spectral fitting software, which was developed by the Royal Belgian Institute for Space Aeronomy (BIRA-IASB) (http://uv-vis. aeronomie.be/software/QDOAS/, last access: 1 November 2019). The strong absorption band of SO<sub>2</sub> is below 325 nm, where NO<sub>2</sub> absorption are relatively weak. Therefore, the fitting wavelength intervals for SO<sub>2</sub> and NO<sub>2</sub> are 307.5–320 nm and 338–370 nm, respectively. Although the DSCDs generated are related to the wavelength of the fitting interval, the usage of the  $SO_2/NO_2$  ratios of DSCDs retrieved at different spectral ranges will not be impacted by the effect of wavelength dependency. Trace gases with absorptions in the relevant fitting windows and ring spectrum were included. The details of the spectral fitting configuration are listed in Table 2. Wavelength calibration was performed using a high-resolution solar reference spectrum (Chance and Kurucz, 2010). The offset and the signal of the dark current were measured every night and were automatically extracted from the measured spectra before spectral analysis. Consequently, the DSCDs of  $SO_2$  and  $NO_2$  were acquired by taking the measured spectrum at 90° as the Fraunhofer reference spectrum.

Figure 2 shows the typical spectral fitting of measured spectra with and without ship plume contamination. The obvious absorbing structures of SO<sub>2</sub> and NO<sub>2</sub> as well as fairly low residuals can be observed in both clean (collected at elevation angle of 5° at 10:39 LT on 22 June 2018) and polluted spectrum (collected at elevation angle of 5° at 09:53 LT on 22 June 2018). By contrast, the retrieved SO<sub>2</sub> and NO<sub>2</sub> DSCDs of  $8.11 \times 10^{16}$  and  $3.08 \times 10^{16}$  molec cm<sup>-2</sup> in the polluted case are significantly higher than those of  $2.24 \times 10^{16}$ and  $1.61 \times 10^{16}$  molec cm<sup>-2</sup> under clean conditions. This demonstrates the high sensitivity of the measurements to ship plumes and the good performance of the spectral fitting. In this study, a threshold of the residual of  $< 1 \times 10^{-3}$  was used to screen out any cases of unsatisfactory spectral fitting; following this screening process 94.57 % of the NO<sub>2</sub> DSCDs and 76.26 % of SO<sub>2</sub> fitting results remained for further discussion. The uncertainties in the spectral analysis of SO<sub>2</sub> were higher due to the weak scatter intensity of sunlight and lower signal-to-noise ratio at shorter wavelengths.

Parameters	SO <sub>2</sub>	NO <sub>2</sub>
Fitting window	307.5–320 nm	338–370 nm
NO <sub>2</sub>	298 K (Vandaele et al., 1998)	298 K (Vandaele et al., 1998)
SO <sub>2</sub>	293 K (Bogumil et al., 2003)	/
O <sub>4</sub>	/	293 K (Thalman and Volkamer, 2013)
O <sub>3</sub>	223 and 243 K (Serdyuchenko et al., 2014)	223 K (Serdyuchenko et al., 2014)
BrO	/	293 K (Fleischmann et al., 2004)
CH <sub>2</sub> O	298 K (Meller and Moortgat, 2000)	298 K (Meller and Moortgat, 2000)
Ring	Calculated by QDOAS	Calculated by QDOAS
Polynomial degree	3	5
Intensity offset	Constant	Constant

**Table 2.** The configuration of spectral fitting for SO<sub>2</sub> and NO<sub>2</sub>.



**Figure 2.** Typical DOAS spectral fitting for SO<sub>2</sub> and NO<sub>2</sub>. Panels (**a**) and (**b**) show spectrum collected at an elevation angle of  $5^{\circ}$  at 10:39 LT on 22 June 2018 under clean conditions, whereas panels (**c**) and (**d**) are spectrum measured at an elevation angle of  $5^{\circ}$  at 09:53 LT on 22 June 2018 under the influence of ship plume pollution. Black lines show the measured atmospheric spectrum, and the red line shows the reference absorption cross section.

# 2.3 MAX-DOAS measurements for ship emissions

For ship emission measurements, the DSCDs of pollutants at low elevation angles should express the change in the integrated concentrations along the light path after contamination by the exhaust plume, which collects scattered sunlight passing through ship plumes. Figure 3a depicts a schematic diagram of ground-based MAX-DOAS measurements of ship emissions. The telescope is pointed towards the shipping lanes or in the direction of a ship that is passing through. Consequently, the spectra measured at a low elevation angle will be impacted by the plumes of ship emissions. In order to better demonstrate the background concentration of NO<sub>2</sub> and SO<sub>2</sub>, several typical observation cycles from 29 June were selected as examples. Figure 3b and c show the vertical distributions of NO<sub>2</sub> and SO<sub>2</sub> DSCDs with elevation angle when there is a ship passing through and when no ship is present, respectively. It can be observed that the DSCDs of NO<sub>2</sub> and SO<sub>2</sub> decreased slowly with increasing angle under clean conditions, during which the maximum values of the NO<sub>2</sub> and SO<sub>2</sub> DSCDs are  $5.03 \times 10^{16}$  molec cm<sup>-2</sup> at an elevation of 3° and  $1.78 \times 10^{16}$  molec cm<sup>-2</sup> at an elevation of 2°, respectively. In contrast, the NO<sub>2</sub> and SO<sub>2</sub> DSCDs increased significantly when ships passed through, showing maximum NO<sub>2</sub> and SO<sub>2</sub> DSCD values of  $7.36 \times 10^{16}$  molec cm<sup>-2</sup> and  $4.15 \times 10^{16}$  molec cm<sup>-2</sup> at an elevation of 5°, respectively; the highest value of SO<sub>2</sub> generally appeared between an elevation angle of 5 and 10°.

#### 3 Results and discussion

# 3.1 Identifying ship emissions at berth

The measurement site at the Waigaoqiao container terminal is located on the southern bank of the Yangtze River, close to the confluence of the Yangtze and Huangpu rivers. The terminal has a total quay length of more than 1 km, and its three container berths are able to accommodate fifth- and sixthgeneration container ships. The terminal's location means that it is an important traffic route for ships entering or leaving the Yangtze River, the Shanghai Port and the Waigaoqiao Wharf. Ship emission are also the only obvious source of SO<sub>2</sub> or NO<sub>2</sub> in this area.

In order to detect the emissions of ship at berths, the MAX-DOAS instrument is placed on a fifth floor of the Pudong MSB (Maritime Safety Bureau) building. The distance between the instrument and the berths is about 1.4 km. As the view from the building is not obscured, multiple berths are in the MAX-DOAS instrument's direct field of view. Considering the size and chimney height of berthed ships, the MAX-DOAS telescope was set to scan vertically in sequence at 3, 4, 5, 6 and 7° (indicated by angle  $\alpha$  in Fig. 4). In the horizontal direction, the telescope ranged from 26 to 34° (the viewing angle from north in the clockwise direction) and yielded a range of angle  $\beta$ . After completing one full scanning sequence in both the vertical and horizontal, a 2-D distribution of DSCDs in the instrument's field of view can be generated. To avoid interference from pollutants' absorptions in the reference spectrum, the spectrum measured at azimuth angle of 10° was considered as the reference background spectrum (no direct pollution from ship emissions).

Figure 5a and b show the respective spatial distributions of the SO<sub>2</sub> and NO<sub>2</sub> DSCDs in the horizontal and vertical directions during a complete scanning sequence. Large spatial gradients can be observed for both SO<sub>2</sub> and NO<sub>2</sub> levels, and the pollutant concentrations are generally higher at lower elevation angles and decline with an increase in height. The highest SO<sub>2</sub> levels, i.e., DSCDs up to  $2.5 \times 10^{16}$  molec cm<sup>-2</sup>, appeared at a horizontal azimuth of 31° and an elevation of 3° and attenuated toward the upper left (see Fig. 5a). Similarly, hot spots of NO<sub>2</sub> with DSCDs of 7.0–8.0  $\times 10^{16}$  molec cm<sup>-2</sup> were centered between  $31^{\circ}$  and  $33^{\circ}$  in the horizontal at an elevation of  $3^{\circ}$ , and decreased in the periphery. It should be noted that the hot spots of the SO<sub>2</sub> and NO<sub>2</sub> distributions are shifted to the left accordingly, whereas the height is raised. This implies that plumes containing SO<sub>2</sub> and NO<sub>2</sub> emitted at the bottom of the observational field of view, disperse and dilute in an upper-left direction, during which the wind at this test site mainly came from the south. Combined with the real scene, shown in Fig. 5c, the dashed rectangle indicates the range of the MAX-DOAS telescope scanning sequence. It can be seen that there are smoke clusters discharged by a ship in the right-hand section of the picture, which corresponds to the azimuth angle between 31 and 33°; under the action of the wind, the plume then spreads to the left of the observational view.

As a full 2-D scanning sequence in the horizontal and vertical directions took about 15 min, more than 10 cycles could be performed in total during an afternoon. In view of the previously identified emission source position, and considering that the NO<sub>2</sub> and SO<sub>2</sub> hot spots in each 2-D distribution are related to the azimuth of the berth where the ship is docked and the corresponding ship's operation status, the DSCDs of NO<sub>2</sub> and SO<sub>2</sub> observed at an elevation of 4° and an azimuth angle of between 31 and 33° were selected to display the temporal pattern of emissions at berth without averaging. In general, the level of the NO<sub>2</sub> DSCDs was much higher than that of  $SO_2$ ; this is due to the fact that there were considerable  $NO_x$  emissions from port trucks between the berth and the instrument, whereas there were no other obvious emission sources of SO<sub>2</sub> in the area. In order to show the variations in the DSCDs with less interference due to changes in the light path, we used a mathematical method to remove the steady change from the trend lines of the NO<sub>2</sub> and SO<sub>2</sub> DSCDs, and kept the residual after background subtraction (see Fig. 6). The baseline is modeled as a low-pass signal, whereas the series of peaks is modeled sparsely with sparse derivatives. Moreover, to account for the positivity of peaks, both asymmetric and symmetric penalty functions were utilized (Ning et al., 2014). Following this process, four significant increases in the SO<sub>2</sub> levels can be observed during the afternoon, accompanied by NO2 enhancements at approximately the same moments, which can be verified by the real scene photographs that evidently show the plumes emitted from the expected exhaust position.

Therefore, it can be concluded that the concentration of  $NO_2$  and  $SO_2$  gases contained in the plumes emitted by the container ships and the corresponding discharge position at berth were identified and monitored remotely by the 2-D MAX-DOAS observations. This application of 2-D MAX-DOAS is similar to the imaging differential optical absorption spectroscopy (IDOAS) technique, which is also used to map the 2-D spatial distribution of polluted gases, such as the distribution of SO<sub>2</sub> in plumes from industrial point sources (General et al., 2014; Pikelnaya et al., 2013). This suggests that the 2-D DOAS technique has the potential to measure the polluted gases mapping from the ships.

#### 3.2 Ship emissions in an inland waterway

In addition to ocean-going ships, vessels in inland waterways also significantly contribute to the amount of ship emissions present in a region (Kurtenbach et al., 2016; Pillot et al., 2016; Zhang et al., 2017). In order to consider emissions



Figure 3. (a) A schematic diagram of the MAX-DOAS measurement geometry for monitoring ship emissions, and the distributions of (b)  $NO_2$  and (c)  $SO_2$  DSCDs with elevation angle on 29 June 2018.



**Figure 4.** The observational geometry of the MAX-DOAS for identifying the emissions of ships at berth in the Waigaoqiao port, Shanghai: (a) top view and (b) side view.

from an inland waterway, the MAX-DOAS instrument was installed on the external windowsill on the third floor of the Wusong MSB building (121°29' E, 31°22' N) from 30 December 2017 to 18 May 2018. The measurement site is located in the downstream region of the Huangpu River, close to its confluence with the Yangtze River. It is the only channel with access to the upstream region of the Huangpu River. There are some non-container terminals near the measurement site, which mainly handle goods related to domestic trading. As a consequence, a large number of ships enter and leave the wharf area every day, and the shipping lanes in the downstream region of Huangpu River suffer from dense ship traffic. By checking the synchronized photographs taken by the camera attached to the instrument, it was found that the types of vessels passing through were diverse, including medium and small container ships, passenger vessels,



Figure 5. Distributions of the measured DSCDs of (a)  $SO_2$  and (b)  $NO_2$  from emissions of a ship at berth from 12:04 to 12:20 on 28 August 2017, and (c) a live photo captured by the MAX-DOAS instrument's camera. The dashed rectangle indicates the observational view of the MAX-DOAS.

bulkers and cargo ships. In addition, the traffic volume in the area is quite high, even as high as hundreds of ships per hour. As shown in Fig. 1c, the viewing direction of the MAX-DOAS instrument in the Wusong area was perpendicular to the river's shipping lane. Thus, the pollutant signals observed mainly came from the emissions of ships navigating the shipping channel. Nevertheless, a small dock and a station for transport containers are located opposite the river, and may slightly influence the measurements. The elevation angles were set in a scanning sequence of 0, 1, 2, 3, 4, 5, 6 and 8°. The spectrum measured at 65° was utilized as the reference, as the zenith direction was blocked to some extent by the MSB building.



**Figure 6.** Time series of DSCDs of (a) NO<sub>2</sub> and (b) SO<sub>2</sub> measured at a  $4^{\circ}$  elevation angle at three azimuths on 28 August 2017.

In order to illustrate the impact of the volume of ship traffic and the meteorological conditions, the 30 min averaged wind speed data, the observed NO<sub>2</sub> and SO<sub>2</sub> DSCDs on 2 representative days (1 January, and 9 March 2018), and the corresponding number of passing ships are shown in Fig. 7. It can be seen from Fig. 7a that 1 January and 9 March were selected to represent days with stabile and unstable atmospheric conditions, respectively. On 1 January, the average SO<sub>2</sub> DSCDs changed from  $1.8 \times 10^{16}$  to  $3.0 \times 10^{16}$  molec cm<sup>-2</sup>, whereas the average NO<sub>2</sub> DSCDs varied between  $4.0 \times 10^{16}$  and  $1.1 \times 10^{17}$  molec cm<sup>-2</sup>. On 9 March, the SO<sub>2</sub> and NO<sub>2</sub> average DSCDs ranged from  $1.0 \times 10^{16}$  to  $2.7 \times 10^{16}$  molec cm<sup>-2</sup> and from  $2.5 \times 10^{16}$  to  $1.0 \times 10^{17}$  molec cm<sup>-2</sup>, respectively. According to the ship traffic density shown in Fig. 7b, the diurnal variations of the SO<sub>2</sub> and NO<sub>2</sub> DSCDs were obviously closely related to the flow of ships (quantitative information). Although the averaged SO<sub>2</sub> and NO<sub>2</sub> DSCD levels are comparable during these 2 days, the ship traffic flow on 9 March was 50 % higher overall than on 1 January, which may imply the important role of the meteorological conditions. Considering the dense shipping lanes, the ship-emitted pollutants are easily accumulated under unfavorable conditions with lower wind speeds (less than  $2 \text{ m s}^{-1}$ ) on 1 January. On the contrary, the ship emissions along the lanes can be spread and more effectively diffused when the averaged wind speed is around  $5 \text{ m s}^{-1}$  on 9 March.

In order to investigate the impacts of wind on the observed DSCDs, the wind rose diagrams of NO<sub>2</sub> and SO<sub>2</sub> DSCDs at an elevation of  $5^{\circ}$  from January to March 2018 are shown in Fig. 8a and b, respectively. It can be seen that the wind mainly comes from north-northwest during the obser-



**Figure 7.** The 30 min averaged wind speed, ship traffic volume, observed NO<sub>2</sub> and SO<sub>2</sub> DSCDs from 10:00 to 13:00 LT on 1 January and 10:00 to 14:00 LT on 9 March at the Wusong Wharf measurement site. The hollow squares in the middle of the boxes represent the mean values, and the solid lines in the middle of the boxes represent the medians. The whiskers extend from each end of the box to the internal and external limits. "–" represents the maximum and minimum, and the "×" symbols are the 1 % and 99 % quantiles. The upper and lower edges of the boxes are the 25 % and 75 % quantiles, respectively.

vation period. The average of the NO<sub>2</sub> DSCDs is low under northerly wind conditions. When the wind direction is parallel to the observation direction (i.e., easterly and westerly, as the viewing direction of the telescope is pointing to the east), the average of the NO2 DSCDs is significantly higher. Similarly, the averaged SO<sub>2</sub> DSCDs are higher under easterly and westerly wind conditions than under southerly and northerly conditions. This suggests that the optical length inside the polluted air and, therefore, the response signal is probably increased when the wind transports the polluted air parallel to the DOAS viewing direction. In Fig. 8c and d, the perpendicular direction for north and south is considered to be wind from  $0^{\circ} \pm 15^{\circ}$  to  $180^{\circ} \pm 15^{\circ}$ , and the parallel direction for east and west is considered to be wind from  $90^{\circ} \pm 15^{\circ}$ to  $270^{\circ} \pm 15^{\circ}$ . It can be seen that the NO<sub>2</sub> and SO<sub>2</sub> DSCDs are quite different under these two types of wind conditions. When the wind is parallel to the observation direction (east and west), 34 % of NO2 DSCDs and 31 % of SO2 DSCDs are greater than  $3.00 \times 10^{16}$  and  $1.50 \times 10^{16}$  molec cm<sup>-2</sup>, respectively. However, under perpendicular direction conditions (north and south), the occurrence of high DSCDs of



**Figure 8.** The dependence of NO<sub>2</sub> (**a**, **c**) and SO<sub>2</sub> (**b**, **d**) DSCDs at an elevation of  $5^{\circ}$  on wind directions from January to March 2018.

NO<sub>2</sub> and SO<sub>2</sub> significantly decreased compared with the parallel direction.

Based on the real-time photographs taken by the instrument, we manually counted the ship density in order to discuss its relationship to the observed DSCDs of NO2 and SO2 at an elevation of  $5^{\circ}$ , as shown in Fig. 9. It is obvious that the hourly means of the NO2 and SO2 DSCDs show an upward trend as the ship density increases. As the fuels used by ships are inconsistent, and the wind speed and direction also affect the DSCDs, it is difficult to find a clear linear relationship between the hourly data of ship density and the DSCDs in this busy inland waterway environment. From a statistical prospective, the averaged NO<sub>2</sub> and SO<sub>2</sub> DSCDs of the binned ship density group (the hollow squares in the middle of the boxes in Fig. 9) show a strong positive correlation with the ship density, with a correlation coefficient (R) of 0.86 and 0.97, respectively. The relatively higher R of SO<sub>2</sub> suggests that the main impacts of SO<sub>2</sub> stem from ship emission sources; however, the R value of NO<sub>2</sub> suggests that there may be a more complex  $NO_2$  source nearby.

In general, it can be concluded from the several months of continuous measurements at the Wusong MSB site that the ship density and meteorological conditions are the two major factors influencing the observed  $NO_2$  and  $SO_2$  levels in this typical inland waterway environment. For similar diffusion situations, the MAX-DOAS instrument can accurately detect elevated pollutants concentrations with an increased number of ships. However, due to the busy shipping lanes in front of the instrument, the MAX-DOAS instrument usually observes pollutant signals in plumes from multiple ships con-



**Figure 9.** Relationship between the DSCDs of (**a**) NO<sub>2</sub> and (**b**) SO<sub>2</sub> at an elevation of 5° and ship density. The hollow squares in the middle of the boxes represent the mean values, and the solid lines in the middle of the boxes represent the medians. Whiskers extend from each end of the boxes to the internal and external limits. "–" represents the maximum and minimum, and the "×" symbols are the 1% and 99% quantiles. The upper and lower edges of the boxes are the 25% and 75% quantiles, respectively.

currently, and the navigation speed of the ships is relatively faster than the period of a completed scan sequence. Therefore, it is very hard to distinguish a single plume from the mixture. Another shortcoming of this measurement method is that the MAX-DOAS measured NO2 levels are considerably impacted by emission sources in the surrounding area, such as nearby main roads and highways, which make nonnegligible contributions to the amount of ambient NO<sub>2</sub>. As it is not practical for regulatory authorities to undertake fuel detection for each ship in this busy inland waterway environment, MAX-DOAS measurements present a more pragmatic opportunity to surveil ship emissions. Based on the legally permitted sulfur content and ship activity data, the theoretical NO<sub>2</sub> an SO<sub>2</sub> concentrations of a plume exhausted from a chimney can be calculated. By combining this calculation with the diffusion model of the plume, the theoretical concentration of SO<sub>2</sub> on the light path of the MAX-DOAS observations can be obtained. Following this, suspicious ships thought to be using fuel with an excess sulfur content can be identified.

## 3.3 Ocean-going ship emissions

Another shipping traffic scenario was also considered with respect to the MAX-DOAS measurements at Yantian Port ( $114^{\circ}29'$  E,  $22^{\circ}56'$  N), which is located on the east side of Shenzhen in the Pearl River Delta emission control zone (see Fig. 1). Yantian Port is one of the largest container ports in China and even in the world, and has 20 large deep-water berths with a quay length of 8212 m and a water depth alongside the quay of 17.4 m, which is beneficial for docking ocean-going vessels with a length of more than 300 m. Unlike the measurement sites discussed above, a distinct feature of shipping traffic in the Yantian Port is the huge size of inbound and outbound vessels and the much lower traf-

fic density. The MAX-DOAS instrument was installed on the shore in the central operation zone of the Yantian Port from 23 May 2018 to 30 June 2018. As can be seen in Fig. 1d, the viewing direction of the MAX-DOAS was pointed toward the shipping lanes in the eastward sea area. Due to the lack of other emission sources in the instrument's field of view, the MAX-DOAS observations could easily capture the pollutants in a single plume from individual inbound and outbound ships (as shown in Fig. 10).

Figure 10a and b present the altitude dependence of the observed  $SO_2$  and  $NO_2$  DSCDs around noontime on 26 May 2018. During the observational period, there were three apparent peaks in the SO<sub>2</sub> and NO<sub>2</sub> DSCDs, i.e., 13:00, 13:30 and 14:10 LT. The increases of both pollutant levels occurred simultaneously. For the first pulse around 13:00 LT, the higher SO<sub>2</sub> DSCD levels are distributed above a 10° elevation, whereas the strong signals of NO2 are concentrated below an elevation angle of  $5^{\circ}$ . This can be explained by the fact that ocean-going container vessels and tugboats behave differently with respect to emissions and operations. Figure 10c shows that a large container ship is outbound at 12:55 LT with the assistance of two tugboats. It is obvious that the exhaust outlet height is very high for large container ships, but quite low for tugboats. As tugboats are usually operated within the port area, their fuel usage is always in accordance with ECA regulations and is of high quality. Thus, the stronger SO<sub>2</sub> signals that appeared at a higher altitude were attributed to container ship emissions, whereas NO<sub>2</sub> hot spots close to the sea surface were contributed by tugboat emissions. During the period around 13:30 LT, the DSCDs of both SO<sub>2</sub> and NO<sub>2</sub> were slightly increased and were located below an elevation of 7°. According to the live photo in Fig. 10d, there was only a small container ship passing through at a distance of about 1 km (in the front view) from the instrument during this time. Considering the distance between the ship and instrument, the height of the exhaust outlet should be related to a lower elevation angle, where the corresponding strong signals of the emitted pollutants are expected to be observed. Additionally, obvious SO<sub>2</sub> and NO<sub>2</sub> signals were found at around 14:10 LT, during which high SO<sub>2</sub> levels were distributed from an elevation angle of 10 to 15°, but strong signals of SO<sub>2</sub> and NO<sub>2</sub> were both found near the sea surface; however, no ship was captured in the live photographs. It can be inferred from the AIS information that the signal observed was a dispersed plume from a ship in another shipping lane, and was not from the front view of the instrument. In addition, the AIS information also confirmed that there were no other ship emission disturbances for the two earlier measurements.

In general, the characteristics of the observed  $SO_2$  and  $NO_2$  DSCD distributions with respect to height are to some extent related to the ship size, its distance from the instrument and operational status, as well as the atmospheric stability. Based on the example discussed above, the MAX-DOAS measurements at Yantian Port can detect pollutants



Figure 10. Measured DSCDs of (a)  $SO_2$  and (b)  $NO_2$  from 12:55 to 14:20 LT, and live photographs taken by the camera at (c) 12:56 and (d) 13:22 LT on 26 May 2018.



Figure 11. A typical distribution of  $SO_2$  DSCDs in the smoke plume of a ship on 22 June 2018.

in a single plume from an individual ship and can provide information regarding the vertical distribution of the pollutants under low ship traffic volume conditions. Considering the large discrepancies of SO<sub>2</sub> signals with altitude, we try to analyze the plumes detected in more detail and obtain a representative observation elevation. According to the live photographs, a large container ship entered the field of view at 09:51 LT on 22 June 2018 and moved very slowly, emitting a distinct black smoke plume. Figure 11 shows the distribution of the SO<sub>2</sub> DSCDs in plumes at different elevation angles. The SO<sub>2</sub> DSCDs peaked at  $8.17 \times 10^{16}$  molec cm<sup>-2</sup> between an elevation angle of 5 and 7°, and decreased with height. It was found that the DSCDs observed at an elevation angle of  $7^{\circ}$  are suitable to represent the peak concentrations in the plumes considering the chimney height of ship and its horizontal distance from the MAX-DOAS instrument.



**Figure 12.** Diurnal variations of (a) NO<sub>2</sub> and (b) SO<sub>2</sub> DSCDs measured at a  $7^{\circ}$  elevation angle on 26 June 2018.

Rather than scanning the full range of elevations, it is possible to individually resolve the plume signal of passing ships using high temporal resolution measurements (60 s) at a single  $(7^{\circ})$  elevation. Thus, the diurnal profiles of the SO<sub>2</sub> and NO<sub>2</sub> DSCDs at a 7° elevation for 26 June 2018 were further investigated, and are presented in Fig. 12. The multiple peaks of SO<sub>2</sub> and NO<sub>2</sub>, with the highest DSCDs of SO<sub>2</sub> and NO<sub>2</sub> exceeding  $6.00 \times 10^{16}$  molec cm<sup>-2</sup>, occurred due to the emissions of the occasional passing ship. By applying a mathematical method, a baseline (the blue dotted line in Fig. 12) can be extracted from the DSCD trend lines (the black solid line in Fig. 12). The baseline represents the diurnal variations in the DSCDs, mostly due to the change in the light path caused by the solar zenith angle and the background emissions. Finally, it can be seen that seven synchronous peaks of  $SO_2$  and  $NO_2$  levels higher than  $2.00 \times 10^{16}$  molec cm<sup>-2</sup> are present in the trend line (the red solid line in Fig. 12). Validated by the live photographs from the instrument and the AIS information, these sharp pollutant concentration peaks originated from the ship plumes of passing ships. This emphasizes the high sensitivity of MAX-DOAS measurements to the change in  $SO_2$  and  $NO_2$  levels in the atmosphere. The increases in the pollutant levels lasted from 10 min to 30 min, which is related to the duration of the ship movement in the instrument's field of view.

Moreover, it should be noted that the amplitude of each peak varied differently, which implies that the DSCDs'  $SO_2/NO_2$  ratio for each peak may reveal emission information regarding the fuel sulfur content of individual vessels (Seyler et al., 2017; Mellqvist et al., 2017; Van Roy

and Scheldeman, 2016). However, it is important to note that NO<sub>2</sub> is formed by the reaction of NO and O<sub>3</sub> in the plume; thus, the SO<sub>2</sub>/NO<sub>2</sub> ratio depends to a certain extent on the age of the plume. The ambient O<sub>3</sub> between 08:00 and 17:00 LT averaged at 63.7 ppb in Yantian during the campaign. Considering the abundance of ozone, the NO emitted by the ship would have rapidly (within a few minutes or even faster) reacted with O<sub>3</sub> to form NO<sub>2</sub> (Seyler et al., 2017). In addition, the conversion between NO and NO<sub>2</sub> is very fast and maintains a dynamic balance with sunlight during the daytime considering the photolysis of NO<sub>2</sub> (Singh et al., 1987). Therefore, the SO<sub>2</sub>/NO<sub>2</sub> ratio in an observed plume would be less impacted by the freshly emitted NO. Thus, a linear regression analysis between the SO<sub>2</sub> and NO<sub>2</sub> DSCDs was performed to infer the fuel sulfur content.

Figure 13 presents the analysis results for 24 different vessels. The strong correlation between the SO<sub>2</sub> and NO<sub>2</sub> DSCDs is obvious evidence of the significant homologies of emission sources between SO<sub>2</sub> and NO<sub>2</sub>. Nevertheless, the slope of different vessels varied greatly from 0.28 to 2.90, indicating the diversity of the SO<sub>2</sub> emission intensity in the ship plumes. In general, SO<sub>2</sub> emissions are directly related to the fuel sulfur content and engine operation status of the ships, according to the emission model estimation, e.g., the power, activity time and the speed of ship (Fan et al., 2016; Zhang et al., 2019). Outbound vessels usually leave the shore slowly with the help of tugboats, and then speed up when sailing into the sea. During the latter process, the main engine is utilized to power navigation, and the fuel used by the main engine has a higher sulfur content than that utilized by the auxiliary engine. In contrast, the main engine of a vessel is usually shutdown during the inbound process. Therefore, the ratio of  $SO_2/NO_2$  in the DSCDs in plumes emitted by outbound vessels could be higher than that those in the plumes of inbound ships.

During the MAX-DOAS observations, we also carried out some fuel sample analyses in addition to the investigation of the activity data and engine parameters of vessels, and five of the vessels in question are shown in Fig. 12. We also indicate the different status of nine individual vessels in Fig. 14, along with information regarding their fuel sulfur content and the rated power of their engines ("inbound" and "outbound" in this figure refer to the rated power of the main and auxiliary engines, respectively). Vessel II is a tugboat that was operating in the port area; thus, it used fuel with the lowest fuel sulfur content of 0.001 % and showed a low  $SO_2/NO_2$  ratio in the DSCDs. Furthermore, the inbound vessels I, VI and VII (indicated by diamonds in Fig. 14) had switched off their main engines when they entered the MAX-DOAS instrument's field of view, and were moved by tugboats before finally docking by inertia. As previously stated, the sulfur content of fuels used to power auxiliary engines are much lower than those utilized to power main engines; thus, SO<sub>2</sub>/NO<sub>2</sub> ratios of inbound vessels are much lower than those of outbound vessels. The other vessels, indicated



Figure 13. The relationship between  $SO_2$  and  $NO_2$  emitted by several typical vessels."O" indicates outbound vessels, "I" indicates inbound vessels and "T" indicates tugboats.

by circles in Fig. 14, are all in outbound ships. As these outbound vessels launched their main engine, and, thus, utilized high sulfur fuels, the vessels exhibited relatively higher  $SO_2/NO_2$  ratios (over 2.0), except for vessel IV. Due to the use of much cleaner fuel (with a sulfur content of 1.28%), vessel IV presented the lowest SO<sub>2</sub>/NO<sub>2</sub> ratio among all of the outbound vessels. Compared with vessel III, which had a similar engine power rating but used fuel with a higher sulfur content, it can be observed that the  $SO_2/NO_2$  ratio in the plume increased with the increase in the fuel sulfur content used by the vessels (Fig. 14). This phenomenon is also notable with respect to vessels V and VIII. It is worth noting that outbound cargo vessel IX deviates from the results seen for the other vessels, as it has very low rated power but a very high  $SO_2/NO_2$  ratio (>2.0). Therefore, vessel IX can be flagged as a suspicious ship that is probably using fuel with a sulfur content exceeding the regulation limit.

Basically, the  $SO_2/NO_2$  ratios in the plumes discharged from inbound vessels and tugboats are usually lower than 1.5 under normal conditions, which are much smaller values than those of outbound vessels using high rated power engines and high sulfur content fuels. For outbound vessels, the  $SO_2/NO_2$  ratios are more related to the fuel sulfur content.



**Figure 14.** The relationship between the  $SO_2/NO_2$  ratio and the fuel sulfur content and engine rated power of all nine vessels, obtained by linear regression. Tugboats, and inbound and outbound vessels are represented by a star, diamonds and circles, respectively.

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An irregular observed  $SO_2/NO_2$  ratio can flag a vessel as possibly not being in conformance with the enforced sulfur content limitation. Therefore, the MAX-DOAS measurement provides promising technology for the compliance monitoring of fuel sulfur content by investigating the  $SO_2/NO_2$  ratio in ship plumes, in addition to more accurate estimation of ship emissions during operation due to the inclusion of the load and emission factors. During the course of this study, the statistical analysis of SO<sub>2</sub>/NO<sub>2</sub> ratios in the plumes of 55 ships were performed. The frequency distribution of the slope of the  $SO_2/NO_2$  ratio is shown in Fig. 15. It can be seen that the  $SO_2/NO_2$  ratios were mostly less than 1.5 (72.7 % of observations). Ships with SO<sub>2</sub>/NO<sub>2</sub> ratios between 0.6 and 0.9 comprised the highest proportion of observations. This indicates that most of the fuels used by ships in Yantian Port could be qualified. However, there are still some ships that may be using noncompliant fuels, as 27 % of the ships observed were found to have a  $SO_2/NO_2$  ratio greater than 1.5.

# 4 Conclusions

In this study, we performed MAX-DOAS measurements to observe ship emissions of SO<sub>2</sub> and NO<sub>2</sub> in Shanghai and Shenzhen, China, under three different typical ship traffic conditions. At Waigaoqiao container terminal in Shanghai, the SO<sub>2</sub> and NO<sub>2</sub> exhausted by ships at berth could be easily identified using 2-D MAX-DOAS observations with respect to the locations and intensities of the emissions. In the inland waterway area at Wusong Wharf it was difficult to determine individual ship emissions due to the dense traffic volume and complex background environment. The long-term MAX-DOAS measurements showed that the changes in SO<sub>2</sub> and NO<sub>2</sub> were correlated with the ship traffic density under stable and unstable atmospheric conditions. However, better dispersion under unstable atmospheric condition was favorable for the decrease of pollutant levels. For open-sea waters in the Yantian deep-water port, the DSCDs of SO<sub>2</sub> and NO<sub>2</sub> measured by MAX-DOAS were highly sensitive to the emitted plumes of vessels passing through the shore-based instrument's field of view, which showed a significant increase in the pollutant concentrations as well as a 10-30 min duration of the emission signals. Considering the distance and the size of the vessels, the DSCDs observed at an elevation angle of  $7^{\circ}$  are hot spots of the concentration in altitude; thus, these values were further selected to investigate the fuel sulfur content. According to the linear regression of the SO<sub>2</sub> and NO<sub>2</sub> DSCDs, the SO<sub>2</sub>/NO<sub>2</sub> ratios are found to be very helpful to infer the levels of sulfur emission. Combined with fuel sample analysis and vessel data, the  $SO_2/NO_2$  ratio in plumes is usually lower than 1.5 for inbound vessels and tugboats, and is much smaller than that of other vessels. The abnormally high  $SO_2/NO_2$  ratio in a plume usually implies that a vessel is not in compliance with the enforced sulfur content limitation.



Figure 15. Frequency distribution of the slope of  $SO_2/NO_2$  from samples of 55 vessels.

In summary, optical remote sensing using the MAX-DOAS technique is highly beneficial for measurements of ship-emitted NO<sub>2</sub> and SO<sub>2</sub>. These applications under different ship traffic conditions demonstrated the feasibility of shore-based MAX-DOAS to observe SO2 and NO2 emitted from vessels docked at berth, vessels navigation in riverine shipping lanes and inbound and outbound operations. Nevertheless, the main ship-emitted pollutants - NO and CO<sub>2</sub> - cannot be monitored due to the limitation of the observed wavelength range. As MAX-DOAS uses solar scattered light as a source, it cannot be utilized at night when there is no sunlight, and there is also large error for observations carried out during twilight and under rainy conditions. In the future, the combination of MAX-DOAS remote sensing of ship plumes and the estimation of emissions using theoretical fuel sulfur contents and actual operation data will provide a promising approach for ship emission surveillance in the future.

*Data availability.* Data are available for scientific purposes upon request from the corresponding authors.

Author contributions. YC, SW and BZ designed and implemented the research, and prepared the paper. JZ, YG and RZ contributed to the MAX-DOAS measurements at different sites. YC and SW carried out the MAX-DOAS retrieval and the combined analysis involving other auxiliary data. YL, YZ, YQ and WM provided constructive comments and support for the ship emissions research portion of this study.

*Competing interests.* The authors declare that they have no conflict of interest.

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