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Supplement of

European NO_x emissions in WRF-Chem derived from OMI: impacts on summertime surface ozone

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1 WRF-Chem namelist

Table S1: Parameterization schemes used in the WRF-Chem setup.

WRF-Chem option	Parameterization scheme (reference)
Physics	
Microphysics	Morrison double-moment (Morrison <i>et al.</i> , 2009)
Longwave radiation	CAM (Collins <i>et al.</i> , 2004)
Shortwave radiation	CAM (Collins <i>et al.</i> , 2004)
Surface layer	MYNN2 (Nakanishi & Niino, 2006)
Land surface physics	Noah land surface model (Tewari <i>et al.</i> , 2004)
Boundary layer physics	MYNN2 (Nakanishi & Niino, 2006; Nakanasi & Niino, 2009)
Cumulus parameterization	Grell 3D Ensemble Scheme (Grell & Devenyi, 2002)
Lightning physics	PR92 neutral buoyancy (Price & Rind, 1993)
Chemistry	
Gas-phase chemistry	CBM-Z (Zaveri & Peters, 1999)
Photolysis parameterization	Madronich F-TUV (Tie <i>et al.</i> , 2003)

2 Meteorology evaluation

2.1 Meteorological reanalysis data

A European-wide meteorology evaluation performed by Mar *et al.* (2016) and numerous other studies demonstrated the skill of WRF-Chem to simulate several meteorological variables relevant to O₃ formation (radiation, temperature, wind speed and wind direction, boundary layer height). We further evaluated WRF-Chem’s performance to simulate meteorology by comparing to the ERA-Interim reanalysis product (Dee *et al.*, 2011), for five variables that are important for surface ozone: surface pressure, 2m temperature, relative humidity, wind speed and wind direction. This complements the comparison with meteorological station observations (e.g. Mar *et al.*, 2016), and has the additional advantage that it is continuous in space.

2.2 Results

To evaluate the meteorology in WRF-Chem we perform a comparison with the state-of-the-art ECMWF operational reanalysis product (hereafter referred to as ECMWF reanalysis). Model performance metrics for the meteorological evaluation for the two simulated months are shown in Table S2, for which show the monthly average of single-day comparisons. We only calculate performance metrics for land-based pixels, as the oceanic pixels generally contribute less to the overall bias. Overall, WRF-Chem shows good performance compared to ECMWF-reanalysis data, and WRF-Chem-ECMWF differences between March and July are consistent in sign.

WRF-Chem performs best at simulating surface temperature and pressure, but relative humidity and wind speed and -direction are simulated with less accuracy. Surface temperature is slightly underestimated, which agrees well with the cold bias generally found in WRF(-Chem) (e.g. Holtslag *et al.*, 2013; Kleczek *et al.*, 2014). Surface pressure is in general slightly underestimated, although we must note that this comparison is limited by terrain height differences in ECMWF reanalysis compared to WRF-Chem. Relative humidity is overestimated substantially in WRF-Chem, by

approximately 10%. This potentially impacts simulated surface ozone in WRF-Chem, as there is an important role for surface atmospheric humidity, which governs the VPD in combination with temperature, in describing ozone removal at the surface (Kavassalis & Murphy, 2017).

We found an approximately linear increase in the model bias (defined in this section as WRF-Chem - ECMWF reanalysis) for RH in July, with a slope of $0.2\% \text{ d}^{-1}$. This coincides with a linear decrease in the bias from 0.12 K to -0.98 K, which would suggest that the domain-averaged latent energy flux is overestimated, leading to an enhanced moisture flux to the atmosphere and underestimated temperatures. For all other variables we did not observe a clear change in domain-average model biases with time, indicating that model performance is robust over the simulation period. Overall, this evaluation, in combination with recent WRF-Chem meteorology evaluation studies (e.g. Mar et al., 2016) provides confidence in WRF-Chem’s skill to reproduce domain-averaged surface meteorological conditions.

Table S2: Meteorological evaluation of two one-month WRF-Chem simulations with ECMWF operational reanalysis fields for five key surface meteorological variables. Only land-based pixels are used in the evaluation.

	March			July						
	μ_{ERA}	μ_{WRF}	r^2	MB	RMSE	μ_{ERA}	μ_{WRF}	r^2	MB	RMSE
T_{2m} [K]	281.51	280.52	0.77	-0.95	2.86	296.33	295.38	0.87	-0.95	2.69
P_{sfc} [hPa]	978.90	976.61	0.83	-2.29	18.56	972.94	972.28	0.96	-0.67	7.98
RH [%]	61.74	71.55	0.41	9.81	18.19	53.21	63.81	0.42	10.60	19.11
WS_{10m} [m s^{-1}]	4.78	5.45	0.56	0.68	2.08	3.34	4.51	0.50	1.17	1.95
WD_{10m} [$^{\circ}$]	181.70	180.39	0.44	1.30	82.53	215.18	209.83	0.34	-5.35	85.71

3 Emission speciation

Table S3: Distribution of TNO-MACC non-methane VOC emission categories over VOC species in CBM-Z.

CBM-Z	TNO-MACC-III
e_ch3oh	alcohols
e_c2h5oh	alcohols
e_hc3	propane, butanes, ethyne
e_hc5	pentanes
e_hc8	hexanes & higher alkanes
e_ol2	ethene
e_olt	propene
e_oli	other alk(adi)enes & alkynes
e_tol	benzene, toluene, other aROUatics
e_xyl	xylene, trimethylbenzenes
e_hcho	methanal
e_ald	other alkanals, ethers
e_ket	ketones
e_ora2	acids

4 Spatial plots of emission scaling parameters

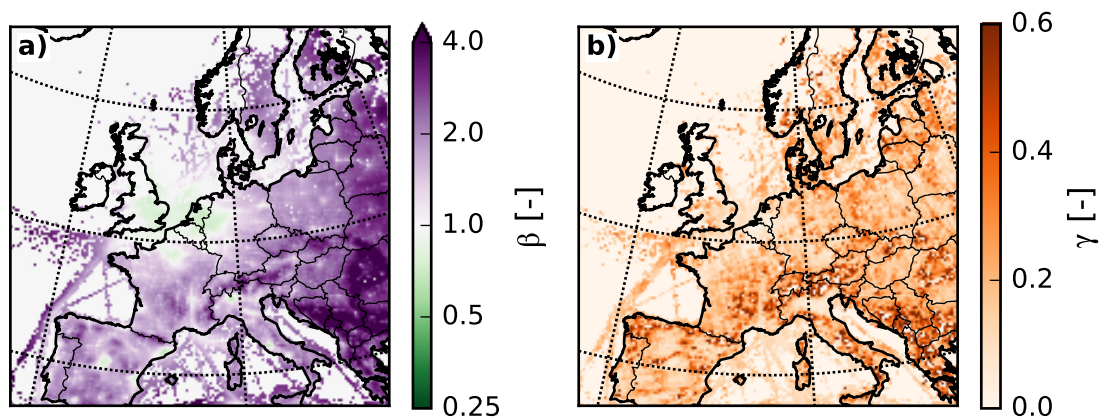


Figure S1: Spatial plots of monthly-averaged values of a) β and b) γ , calculated following Eqns. 3 and 4 (in main text), respectively.

5 Sensitivity study on inorganic reaction rates

A recent study investigated the representation of inorganic rate constants for tropospheric O_3 formation in WRF-Chem, and found a strong impact on the monthly average of $8 \mu\text{g m}^{-3}$ when using

MOZART inorganic rate constants in RADM2 (Mar *et al.*, 2016). To evaluate the potential impact of this on our simulations, we apply the mixed-layer and chemistry model MXLCH (Janssen *et al.*, 2012), which uses a simplified version of the MOZART mechanism. We set up this case as follows in order to reproduce polluted conditions occurring in the Mediterranean, in order to determine the impact of inorganic reaction rates on the production of ozone in a well-mixed boundary layer: The location is set at 45.5°N/3.4°E (Southern France), initial O₃ concentrations in the mixed layer and the free troposphere are set to 62 ppbv and 78 ppbv, respectively, initial NO and NO₂ concentrations in the mixed layer are set to 1.6 ppb and 4.0 ppb, respectively, we apply NO and CO emission fluxes representative for relatively polluted conditions (0.15 ppb s⁻¹ and 2.0 ppb s⁻¹, respectively), and we add two reactions to this mechanism representing HO_x cycling via reaction with O₃.

From a comparison of rate constants among the mechanisms CBM-Z, RADM2 and MOZART, we found the largest differences in rate constants for the reaction forming HNO₃ (NO₂ + OH + M → HNO₃ + M), while other inorganic rate constants are much more comparable. This is in line with the rate constant comparison by Knote *et al.* (2015). We modify the temperature-dependent rate constants of the reaction forming HNO₃ ($k_{\text{NO}_2 + \text{OH}}$) according to Fig. S2 (panel a), and subsequently we study the sensitivity of afternoon ozone concentrations to $k_{\text{NO}_2 + \text{OH}}$.

The NO₂ concentration and lifetime increase with decreasing rate constants, but the impact of $k_{\text{NO}_2 + \text{OH}}$ on NO₂ concentrations is rather small (Fig. S2c). The relative impact on OH is stronger (Fig. S2d): the NO₂ availability in combination with $k_{\text{NO}_2 + \text{OH}}$ drives OH loss, causing increasing OH concentrations for a decrease in $k_{\text{NO}_2 + \text{OH}}$.

HNO₃ formation in CBM-Z has a somewhat lower rate constant compared to other mechanisms, and therefore leads to a longer NO₂ lifetime. This accelerates O₃ formation, and thus leads to higher afternoon O₃ concentrations. The upper right panel of Fig. S2 shows that the inter-mechanism spread is ±2 ppbv. From this sensitivity analysis with a simplified representation of atmospheric chemistry within the atmospheric boundary layer, we conclude that there is some sensitivity of afternoon O₃ concentrations to the representation of inorganic reactions, particularly HNO₃ formation, involved in O₃ chemistry.

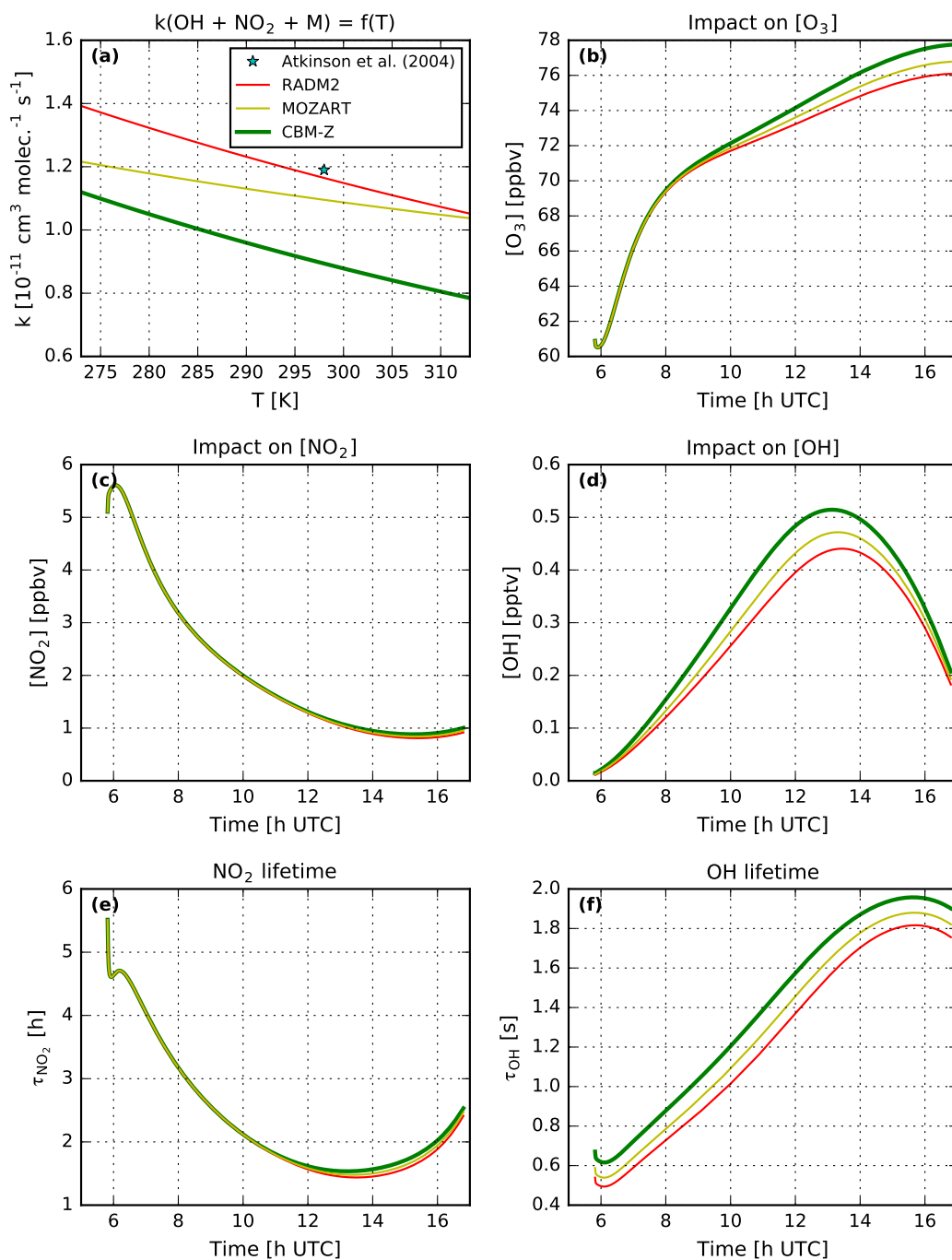


Figure S2: Temperature-dependence of rate constants for the reaction $\text{NO}_2 + \text{OH} + \text{M} \rightarrow \text{HNO}_3 + \text{M}$ from three different mechanisms (panel a), and the resulting impacts on O_3 (panel b), NO_2 (c) and OH (d). Panel a additionally gives the IUPAC-recommended value under standard conditions ($P = 1 \text{ bar}$, $T = 298 \text{ K}$) given by Atkinson et al. (2004). The lifetimes of NO_2 and OH are given in panels e and f, respectively.

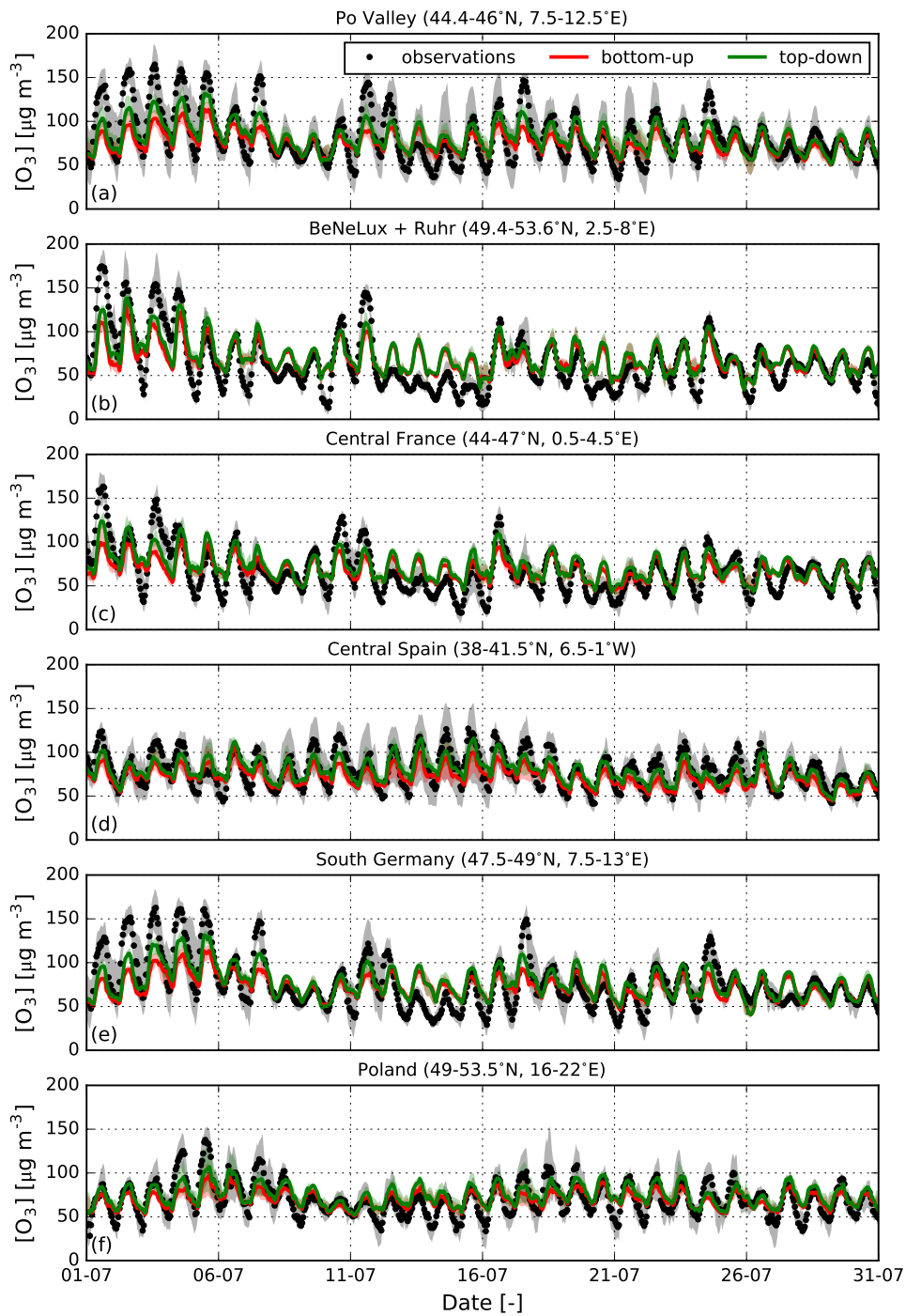


Figure S3: July 2015 time series of the median (shaded areas show inter-quartile range) O_3 concentrations as observed at AirBase stations (black dots), and as simulated by WRF-chem with bottom-up (red) and top-down emissions (green). Medians are calculated by including all stations (resp. co-sampled simulations) in the latitude/longitude range specified in the subplot titles.

Table S4: Model performance statistics for surface ozone concentration time series of the WRF-Chem simulation with bottom-up and top-down emissions for six European regions.

	Po Valley	BeNeLux + Ruhr	Central France	Central Spain	South Germany	Poland
n (stations)	59	32	29	24	39	18
Bottom-up						
MBE	-20.14	16.82	3.35	-22.40	-11.15	1.74
RMSE	68.07	71.48	59.92	45.39	68.68	43.64
r	0.80	0.78	0.76	0.81	0.74	0.77
Top-down						
MBE	-1.58	25.94	17.29	-1.33	5.02	16.10
RMSE	55.08	68.57	56.48	36.44	58.13	41.81
r	0.85	0.81	0.79	0.83	0.81	0.80

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