



# Relationship between chemical composition and oxidative potential of secondary organic aerosol from polycyclic aromatic hydrocarbons

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Received: 29 October 2017 – Discussion started: 10 November 2017

Revised: 16 February 2018 – Accepted: 19 February 2018 – Published: 22 March 2018

**Abstract.** Owing to the complex nature and dynamic behaviors of secondary organic aerosol (SOA), its ability to cause oxidative stress (known as oxidative potential, or OP) and adverse health outcomes remains poorly understood. In this work, we probed the linkages between the chemical composition of SOA and its OP, and investigated impacts from various SOA evolution pathways, including atmospheric oligomerization, heterogeneous oxidation, and mixing with metal. SOA formed from photooxidation of the two most common polycyclic aromatic hydrocarbons (naphthalene and phenanthrene) were studied as model systems. OP was evaluated using the dithiothreitol (DTT) assay. The oligomer-rich fraction separated by liquid chromatography dominates DTT activity in both SOA systems ( $52 \pm 10\%$  for naphthalene SOA (NSOA), and  $56 \pm 5\%$  for phenanthrene SOA (PSOA)). Heterogeneous ozonolysis of NSOA was found to enhance its OP, which is consistent with the trend observed in selected individual oxidation products. DTT activities from redox-active organic compounds and metals were found to be not additive. When mixing with highly redox-active metal (Cu), OP of the mixture decreased significantly for 1,2-naphthoquinone ( $42 \pm 7\%$ ), 2,3-dihydroxynaphthalene ( $35 \pm 1\%$ ), NSOA ( $50 \pm 6\%$ ), and PSOA ( $43 \pm 4\%$ ). Evidence from proton nuclear magnetic resonance (<sup>1</sup>H NMR) spectroscopy illustrates that such OP reduction upon mixing can be ascribed to metal–organic binding interactions. Our results highlight the role of aerosol chemical composition under atmospheric aging processes in determining the OP of SOA, which is needed for more accurate and explicit prediction of the toxicological impacts from particulate matter.

## 1 Introduction

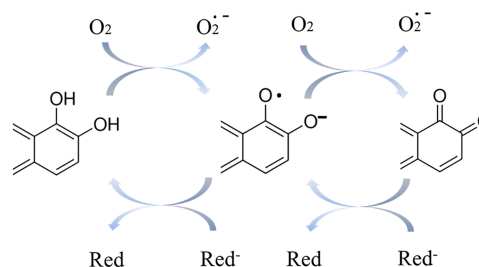
Exposure to particulate matter (PM) has been associated with various adverse health outcomes, such as increased risks of myocardial infarction, ischemic heart disease, lung cancer, exacerbation of asthma, and chronic obstructive pulmonary disease (de Kok et al., 2006; Li et al., 2003a, b; Nel, 2005; Risom et al., 2005; Thurston et al., 2016). As a result, ambient PM<sub>2.5</sub> exposure ranks among the top five global mortality risk factors (Cohen et al., 2017). Meanwhile, decreased ambient PM levels have been associated with longer life expectancies (Pope et al., 2009). To establish causal links between aerosol exposure and health outcomes, cytotoxic and carcinogenic potential has been investigated by epidemiological studies in the past decades (Brunekreef and Holgate, 2002; Beelen et al., 2014; Lelieveld et al., 2015; Pope et al., 2002), but the underlying mechanistic pathways by which PM causes adverse health outcomes still remain poorly understood.

Oxidative stress has been proposed as one of the main mechanisms for PM toxicity in recent years, and is caused by oxidative potential (OP; Li et al., 2003b). OP is exhibited as the capacity of inhaled PM to induce oxidative stress, the redox imbalance generated through consumption of antioxidants and production of reactive oxygen species (ROS) (Antiñolo et al., 2015; Shen et al., 2011; Shiraiwa et al., 2012). ROS include a variety of oxidants such as superoxide ( $O_2^{\bullet-}$ ), hydroxyl radical ( $\bullet OH$ ), and hydrogen peroxide ( $HO_2H$ ), which could either be introduced into human body directly from inhaled PM or generated by targeted cells (Nel et al., 1998; Pöschl and Shiraiwa, 2015; Rhee, 2006; Verma et al., 2015b). The generation of ROS during multiphase in-

teractions between air pollutants and the human respiratory tract is closely related to the chemical composition, since the combination of various pollutants may influence chemical reactivity as well as bioavailability of PM while having synergistic or nonlinear influences on its OP (Antiñolo et al., 2015; Charrier et al., 2015; Fang et al., 2015; Shiraiwa et al., 2012; Xiong et al., 2017).

OP of both organic and inorganic PM components have been evaluated by both cellular and acellular assays. In vitro cellular assays were conducted by detecting biological endpoints of the exposure, including heme oxygenase-1 (HO-1) and other cytokines as well as macrophage-related biomarker expressions (Krapf et al., 2017; Li et al., 2003b). On the other hand, acellular assays use specific chemicals, such as dithiothreitol (DTT), ascorbate (AA), and glutathione (GSH), as surrogates of low molecular weight (MW) antioxidants (Fang et al., 2016; Godri et al., 2011; McWhinney et al., 2013). Among acellular assays, the DTT assay quantifies OP by measuring the DTT depletion rate over a fixed time interval, which mimics the physiological process of electron transfer from biological antioxidants to dissolved  $O_2$  (Cho et al., 2005). DTT assay is one of the most commonly used OP evaluation methods since DTT is a potent surrogate for the total thio-pools (glutathione and protein thiols), and this assay can be conducted under biologically relevant conditions (37 °C, pH = 7.4) with relatively simple procedures (Cleland, 1964; Hansen et al., 2009; McWhinney et al., 2011). OP levels measured by this assay have been found to correlate well with cellular ROS expression as well as several airway inflammation biomarkers, such as HO-1, tumor necrosis factor- $\alpha$  (TNF- $\alpha$ ) and fractional exhaled nitric oxide (FENO) (Delfino et al., 2013; Li et al., 2003b; Tuet et al., 2017a).

Secondary organic aerosol (SOA) from atmospheric oxidation of gaseous precursors comprises a major fraction of submicron particulate matter. To date, the DTT assay has been applied in a few studies to evaluate the OP of both laboratory and ambient SOA (McWhinney et al., 2013; Tuet et al., 2017b; Verma et al., 2015a). However, owing to the complex and dynamic property of SOA, there is limited understanding of the relationship between detailed SOA composition and its OP (Charrier and Anastasio, 2012; Pöschl and Shiraiwa, 2015; Tuet et al., 2017b). Tuet et al. (2017b) studied the DTT activity of chamber-generated SOA from both biogenic and anthropogenic volatile organic compounds under various conditions, showing that naphthalene SOA (NSOA) has the highest OP. Previous work (Antiñolo et al., 2015; Bolton et al., 2000; Charrier and Anastasio, 2012; Cho et al., 2005; Jiang et al., 2017; Tuet et al., 2017b) indicated that polycyclic aromatic hydrocarbons (PAHs) are unreactive towards DTT while their oxidation products, such as quinones, can be highly redox-active. Quinones can be directly emitted from traffic or formed from secondary oxidation (Cho et al., 2004; McWhinney et al., 2013), and are able to consume antioxidants in a catalytic cycle (Fig. 1; Bolton et al., 2000; Valavanidis et al., 2005). McWhin-



**Figure 1.** A simplified mechanism of redox cycling of quinone-like substances and formation of superoxide anion radicals. Red shows a general reductant. In the redox cycle, regenerated quinone serves as a chemical intermediate to transfer electrons from reductants to oxygen to form superoxide ( $O_2^{\bullet-}$ ).

ney et al. (2013) found three quinones (1,2-naphthoquinone, 1,4-naphthoquinone, and 5-hydroxy-1,4-naphthoquinone) in NSOA could only account  $30 \pm 5\%$  for the observed OP of NSOA, and the source of the remaining DTT activity remains unknown. Peroxides, which are the major contributors to the OP of isoprene SOA (Jiang et al., 2017; Lin et al., 2016; Sur-ratt et al., 2010), may also be abundant in NSOA (Kautzman et al., 2010), but their contribution to the OP of NSOA has not been evaluated. In addition, the composition of SOA may evolve upon atmospheric aging. A previous study (Verma et al., 2009) found ambient samples collected in the afternoon had a larger fraction of water soluble organic carbon and higher OP, suggesting the photochemical aging effects. There may also be correlations between the average carbon oxidation state ( $\overline{OS}_C$ ) and OP of SOA at different stages of oxidation (Tuet et al., 2017a). Moreover, mixing between organics and metals was also found to change the OP of specific components in SOA (Xiong et al., 2017), but the mechanisms remain the tip of iceberg.

Here, we focused on understanding how the composition of PAH-derived SOA is related to its strong OP (Tuet et al., 2017b). We aim to address the following specific questions in this work: what are the compounds within SOA that are important for DTT activity? How does the OP change upon atmospheric aging processes, including oligomerization, heterogeneous oxidation, and mixing with transition metals (Gao et al., 2004; Rudich et al., 2007)? In this work, SOA from photooxidation of naphthalene and phenanthrene was studied as model systems, and compared to SOA from ozonolysis of  $\alpha$ -pinene and limonene. The relative OP contributions of peroxides and high MW oligomers were evaluated. Effects of aerosol aging on OP was evaluated by examining the OP of individual oxidation products known to be present in NSOA, and the OP of NSOA samples that were further oxidized in the condensed phase. Lastly, the impacts of SOA mixing with metal was explored by mixing SOA or redox-active SOA components with Cu (II), a transition metal which has been identified with the highest OP in am-

bient particles (Charrier and Anastasio, 2012). Additivity of OP revealed by DTT assay was investigated, and the mechanisms of Cu–organic interactions were examined in detail using proton nuclear magnetic resonance spectroscopy ( $^1\text{H}$  NMR).

## 2 Methods

### 2.1 Flow tube experiments

SOA was produced in a custom-built 10 L quartz flow tube. Details about the flow tube conditions have been described in previous work (Ye et al., 2016, 2017). Prior to each experiment, the flow tube was flushed with purified compressed air at a flow rate of  $20\text{ L min}^{-1}$  for over 24 h.

To produce SOA from naphthalene or phenanthrene, solid PAH was placed in a heated container ( $80^\circ\text{C}$ ), and the sublimed vapor was carried into the flow tube at a  $0.2\text{ L min}^{-1}$  flow of purified compressed air.  $\text{O}_3$  and water vapor were also added into the flow tube.  $\text{O}_3$  was produced by passing  $0.5\text{ L min}^{-1}$  oxygen (99.6 %, Linde, Mississauga, Canada) through an UV  $\text{O}_3$  generator (no. 97006601, UVP, Cambridge, UK). Water vapor was produced by bubbling purified air through a custom-made humidifier with a flow rate of  $1.3\text{ L min}^{-1}$ . Residence time inside the flow tube was maintained around 5 min. The flow tube was housed inside an aluminum enclosure, which was equipped with 254 nm UV lamps (UVP, Cambridge, UK). The photolysis of  $\text{O}_3$  produces  $\text{O} (^1\text{D})$ , which reacts with water vapor to produce  $\cdot\text{OH}$  and initiate photooxidation of naphthalene/phenanthrene as well as SOA formation. During naphthalene/phenanthrene photooxidation,  $\text{O}_3$  concentration inside the flow tube was controlled around 1 ppm. In addition, blank experiments were conducted under the same conditions, without injecting any hydrocarbons.

Two types of SOA from monoterpene ozonolysis were also synthesized under similar conditions. Separately,  $\alpha$ -pinene (Sigma-Aldrich, 98 %) and limonene (Sigma-Aldrich, 97 %) were pre-dissolved in cyclohexane (Sigma-Aldrich, 99.5 %) with a volumetric ratio of 1 : 500 and 1 : 1500, respectively. At these ratios, the reaction rates between  $\cdot\text{OH}$  and cyclohexane are at least a hundred times higher than that of SOA precursors (Atkinson and Arey, 2003; Keywood et al., 2004). The experimental solution was injected continuously into purified air flow by a syringe (1000 mL, Hamilton) installed on a syringe pump (KDS Legato100) to achieve an initial concentration of  $588 \pm 16$  or  $298 \pm 24$  ppb of  $\alpha$ -pinene or limonene, respectively.  $\text{O}_3$  was produced by passing oxygen through the  $\text{O}_3$  generator at a flow rate of 0.2 or  $0.3\text{ L min}^{-1}$  for  $\alpha$ -pinene or limonene, respectively, such that the  $\text{O}_3$  concentrations were at least 5 times higher than  $\alpha$ -pinene or limonene. Both experiments were conducted in the same flow tube without irradiation of UV lights.

Temperature and relative humidity were monitored by an Omega HX94C RH/T transmitter. The concentrations of SOA precursors at the inlet and outlet of the flow tube were measured by a gas chromatography–flame ionization detector (GC-FID, model 8610C, SRI Instruments Inc., LV, USA) equipped with a Tenax<sup>®</sup> TA trap. SOA size distribution and volume concentration at the flow tube reactor outlet were monitored using a custom-built scanning mobility particle sizer (SMPS), which was composed of a differential mobility analyzer column (DMA, model 3081, TSI, Shoreview, MN, USA) with flow controls and a condensation particle counter (CPC, model 3772, TSI, Shoreview, MN, USA). The SMPS data were inverted to particle size distributions using custom code written in Igor Pro (Wavemetrics, Portland, OR, USA). By assuming a particle density of  $1.25\text{ g cm}^{-3}$  for monoterpene SOA (Kostenidou et al., 2007; Shilling et al., 2009) and  $1.55\text{ g cm}^{-3}$  for PAH SOA (Chan et al., 2009; McWhinney et al., 2011), particle volume concentrations were converted into mass concentrations and integrated as a function of sample collection time and flow rates to obtain the total mass of collected SOA.

### 2.2 SOA sampling and extraction

All SOA samples were collected on 47 mm prebaked ( $500^\circ\text{C}$ , 24 h) quartz fiber filters (Pall, Ann Arbor, MI, USA) in a stainless-steel filter holder after reaching a steady state yield, and then wrapped in prebaked aluminum foil before being stored in sterile petri dishes sealed with Parafilm M<sup>®</sup> at  $-20^\circ\text{C}$ . Within 3 days of collection, the filters were extracted in methanol (HPLC grade, 99.9 %, Sigma-Aldrich, St. Louis, MO, USA), by ultra-sonication at room temperature for more than 3 min. After sonication, insoluble materials were filtered by a PTFE (polytetrafluoroethylene) syringe filters (Fisherbrand<sup>™</sup>) with pore size of  $0.22\text{ }\mu\text{m}$ . Chemical analysis and DTT activities of the filter extracts were conducted within hours of extraction. As negative control, filter samples were also collected during blank experiments and extracted in the same manner.

### 2.3 DTT assay

OP of SOA and selected quinone/peroxide standards were quantified by the depletion rate of DTT, an antioxidant surrogate that can be consumed by oxidative components in PM (Kumagai et al., 2002). The protocols used in this work are adapted from those of McWhinney et al. (2011, 2013). The SOA extracts were first evaporated to complete dryness in a  $5.0\text{ L min}^{-1}$  of  $\text{N}_2$  using a blow-off system (N-EVAP, Organomation, USA). Phosphate buffer (0.1 M, pH 7.4) was then added to dissolve the SOA to achieve a concentration of 0.2 mM. For quinones, copper (II) sulfate, and peroxides, each pure compound was weighed and dissolved in 0.1 M phosphate buffer. The concentration of quinones and copper (II) sulfate solutions was  $1\text{ }\mu\text{M}$ , and the concentration

of each peroxide solution was 120  $\mu\text{M}$ . The specific solution was then added into multiple wells in a 96-well UV plate (Greiner Bio-One, Kremsmünster, AT) with 160  $\mu\text{L}$  per well, and immediately covered with adhesive plate sealer (Edge-Bio, Gaithersburg, USA). The plate was then placed in a UV-Vis spectrophotometer (Spectramax 190, Molecular Devices Corporation, Sunnyvale, CA) for incubation. The incubation temperature was maintained at 37  $^{\circ}\text{C}$  inside the spectrophotometer to mimic human physiological conditions. Then, 0.02 mL DTT (0.2 mM) was added into each well to initiate the redox reactions. At each time point, 0.02 mL of 5,5'-dithiobis(2-nitrobenzoic acid) (DTNB; 2 mM) was added, which immediately consumed all the remaining DTT to form a yellow product, 2-nitro-5-thiobenzoic acid (TNB; Fig. S1 in the Supplement). DTNB was added in excess (10 times) of DTT for the rapid and complete consumption of the remaining DTT. The initial DTT concentration was determined in order to maintain an eventual consumption of around 50 % (Fig. S2c, d) for PAH-derived SOA. TNB was quantified by the light absorption at a wavelength of 412 nm, which was further converted to the DTT amount by calibration curve (Fig. S2b) in order to obtain the DTT decay rate. The reaction was quenched by adding DTNB at different wells at different times (all containing the same initial mixture), allowing for quantification of the DTT decay rate over a 30 min time interval (every 5 min for the first 10 min and then every 10 min). Blank control and the calibration curve for DTT quantification are shown in Fig. S2. SOA methanol extractions were blown off by various extents in order to assess whether the loss of volatile compounds along with the  $\text{N}_2$  blow-off procedure affects the DTT activity of SOA (Fig. S3). Here, we used the total DTT decay rate,  $\text{DTT}_t$  ( $\mu\text{M DTT min}^{-1}$ ), to report the total oxidative capacity, as well as the mass normalized DTT decay rate,  $\text{DTT}_m$  ( $\text{pmol DTT min}^{-1} \mu\text{g}^{-1}$  organics), to report the OP (Charrier et al., 2016; Jiang et al., 2017; Xiong et al., 2017). Detailed information about the chemicals used in this assay is shown in Sect. S1 in the Supplement.

## 2.4 Quantification of peroxides in SOA

Quantification of total peroxides in the four types of SOA was conducted using the iodometric–spectrophotometric method, which quantifies total aerosol peroxides in all three forms ( $\text{H}_2\text{O}_2$ , ROOH, ROOR; Banerjee and Budke, 1964; Docherty et al., 2005):



where  $\text{I}^-$  is oxidized to  $\text{I}_2$  by peroxides under acidic conditions, and then complexes with remaining  $\text{I}^-$  to form  $\text{I}_3^-$ , a compound with brown color detected spectrophotometrically at a wavelength of 470 nm. The concentration of each SOA solution was first adjusted to 5 mM; 0.02 mL of potassium iodide solution (1 g  $\text{mL}^{-1}$ , KI dissolved in DI water), which provided the  $\text{I}^-$  in Reaction (R1), and 0.02 mL of formic acid

(Sigma-Aldrich,  $\geq 95\%$ ), which maintained the acidity, were added to 0.16 mL of the SOA solution in each well of a 96-well UV plate. The plate was immediately sealed and incubated for 1 h following the same procedures as in the DTT assay, and the UV-Vis absorbance was measured at 470 nm. After testing for the sensitivities of various peroxides in KI assay (Fig. S4), and following the previous work by Kautzman et al. (2010), benzoyl peroxide (Sigma-Aldrich,  $\geq 98\%$ ) was chosen to represent peroxides in NSOA and used as standards for mass calibration in this study. All values are reported as mass fraction of peroxides in the total SOA.

## 2.5 Heterogeneous oxidation

Heterogeneous oxidation of NSOA was conducted by first cutting a filter with freshly collected NSOA into halves (within 3 days of flow tube synthesis and stored at  $-20^{\circ}\text{C}$ ). One half of the filter was placed in a sealed container, and an  $\text{O}_3$  stream ( $\sim 3$  ppm, from the previously mentioned  $\text{O}_3$  generator) was passed through the filter at a flow rate of  $0.2 \text{ L min}^{-1}$ . The other half of the filter was treated in parallel with a  $0.2 \text{ L min}^{-1}$  flow of  $\text{N}_2$  over the same time intervals, to account for evaporation and/or decomposition of SOA components at room temperature. Three sets of experiments were conducted with exposure times of 1, 12, and 24 h where the  $\text{O}_3$  exposure can be determined by

$$\text{O}_3 \text{ exposure} = \int_0^t [\text{O}_3] dt = \langle \text{O}_3 \rangle_t \times t, \quad (1)$$

where  $\langle \text{O}_3 \rangle_t$  is the time averaged  $\text{O}_3$  concentration at a total flow rate of  $0.2 \text{ L min}^{-1}$ . The DTT activity of each  $\text{O}_3$ -exposed aerosol sample was normalized to that of the corresponding  $\text{N}_2$  exposure group. Changes in organic carbon mass on NSOA filters exposed to  $\text{O}_3/\text{N}_2$  for 1, 12, and 24 h were also monitored with a thermal optical organic carbon / elemental carbon (OC / EC) aerosol analyzer instrument (Sunset Laboratory Inc., Tigard, OR, USA). OC / EC content was measured following the IMPROVE OC / EC protocol (Chow et al., 1993). Blanks were measured before each run and subtracted from the sample measurements.

## 2.6 Chromatographic separation of NSOA

To identify the relative contributions of monomers and oligomers in NSOA and phenanthrene SOA (PSOA) to DTT activity, the SOA extract was separated using ultra-high-performance liquid chromatography (UHPLC), and analyzed using electrospray ionization/ion mobility/time-of-flight mass spectrometry (ESI/IMS-TOF MS, TOFW-ERK, Switzerland; hereafter referred to as IMS-TOF). SOA methanol extract ( $30 \text{ g L}^{-1}$ ) was separated on a reverse-phase column (ZORBAX Eclipse Plus C18, Agilent, USA) with an initial mobile phase of 90 % DI water and 10 % HPLC methanol at a flow rate of  $0.15 \text{ mL min}^{-1}$  (1290 Infinity II,

Agilent, USA). The ratio of water to methanol was gradually adjusted from 9 : 1 to 1 : 9 between 25 and 30 min. Separation temperature was set to 30 °C with a pressure setting of 150 bar. The outlet flow was regulated using LC-MS post column flow splitters (Supelco, Sigma-Aldrich, USA) at a ratio of 30 : 1. The major flow was collected in two different fractions: the first fraction was collected between 6 and 14 min, and the second fraction was collected between 14 and 33 min for NSOA (3–17 and 17–28 min for PSOA). DTT assay was conducted on each fraction to assess their OPs. The minor flow injected into IMS-TOF was controlled at  $5 \mu\text{L min}^{-1}$  for mass spectrometric analysis in the negative mode.

A deactivated fused silica capillary (360  $\mu\text{m}$  OD, 50  $\mu\text{m}$  ID, 50 cm length, New Objective, Woburn, MA, USA) was used as the sample transfer line between the UHPLC and the IMS-TOF. The ESI source was equipped with an uncoated SilicaTip Emitter (360  $\mu\text{m}$  OD, 50  $\mu\text{m}$  ID, 50  $\mu\text{m}$  tip ID, New Objective, Woburn, MA, US). Charged SOA droplets generated from the tip of the emitter were transferred through a desolvation region by a  $1 \text{ L min}^{-1}$   $\text{N}_2$  flow at room temperature, and ions produced from the evaporated droplets were introduced into the drift tube for ion mobility separation. The IMS drift voltage was set to  $-1.2 \text{ kV}$  for the negative mode. The separation temperature was set to  $80 \pm 1 \text{ }^\circ\text{C}$  with an operation pressure setting of 1.2 bar for the whole mass spectrometer. After separation in the ion mobility region, the ion  $m/z$  is measured by high-resolution time-of-flight mass spectrometry within an  $m/z$  range of 40 to 800. Resolution ( $m/\text{dm}_{50}$ ) of the time-of-flight mass spectrometer is typically 3500–4000 FWHM at  $m/z$  250 (Groessl et al., 2015; Krechmer et al., 2016). Spectra recording and data processing of the current study were performed using routines written in Igor Pro (6.37, Wavemetrics, OR, USA): “Acquility” (version 2.1.0, <http://www.tofwerk.com/acquility>) for raw data acquisition and “Tofware” (version 2.5.3, [www.tofwerk.com/tofware](http://www.tofwerk.com/tofware)) for post-processing.

## 2.7 $^1\text{H}$ NMR spectroscopy

$^1\text{H}$  NMR spectroscopy was used to further investigate the mechanism behind the OP reduction upon mixing specific organics and transition metals (Simpson et al., 2011; Simpson and Simpson, 2014; Smith and van Eck, 1999). NMR measurements were performed on a Bruker Avance III NMR spectrometer (11.7 T), equipped with a four-channel liquid state ( $^1\text{H}$ ,  $^{13}\text{C}$ ,  $^{15}\text{N}$ ,  $^2\text{H}$ ) inverse detection probe (QXI) fitted with an actively shielded Z gradient. Typical parameters used for 1-D  $^1\text{H}$  experiments were a  $9.5 \mu\text{s}$   $^1\text{H}$  pulse, 64k acquisition points, and 14 ppm spectral width; eight transients were collected, with a total of 5.7 s between scans. Before NMR analysis, each sample was dissolved into deuterium oxide ( $\text{D}_2\text{O}$ , Cambridge Isotope Laboratories, 99.96 %) and dimethyl sulfoxide (DMSO, Fisher Scientific, 99.9 %) at a ratio of 9 : 1. Here we used  $\text{D}_2\text{O}$  as a lock reagent by suppressing it through pre-saturation.

In order to monitor the duration for the nuclear spin magnetization returning to an equilibrium state, NMR relaxation times T1 (longitudinal direction) and T2 (transverse direction) were analyzed. T1 of the sample was measured through the standard inversion recovery experiment. The delay periods used for T1 measurements ramped from 0.001 to 15 s in 16 increments with a delay of 60 s between scans, which represented  $> 5 \times \text{T1}$  time to permit full signal recovery. For each delay period, 16 transients were collected. The T2 of the sample was measured through the standard Carr–Purcell–Meiboom–Gill (CPMG) sequence. The delay periods used for T2 measurements ramped from 1.2 to 614 ms in 16 increments with a delay of 60 s between scans, which represented  $> 5 \times \text{T1}$  time to permit full signal recovery. For each delay period, 16 transients were collected. A total of 16 free induction decays were collected for each of the relaxation experiments, and relaxation time calculations were done using Bruker Dynamics Centre (v2.4.5) with mono-exponential fitting functions (Eqs. 3 and 4 below).

$$f(t) = I_0 \times [1 - 2e^{-t/\text{T1}}] \quad (2)$$

$$f(t) = I_0 \times e^{-t/\text{T2}} \quad (3)$$

Equations (2) and (3) are the fitting functions for T1 and T2, respectively, where  $I_0$  is the thermal equilibrium state of the overall proton nuclear spin magnetization and  $t$  is the variable delay time.

All the  $^1\text{H}$  NMR spectra were collected using TopSpin (Bruker, v3.2). Post-NMR data processing was conducted in MestReNova (Mestrelab Research, v11.0.4) and Origin (OriginLab, v9).

## 2.8 Statistical analysis

Data in this study were interpreted as mean  $\pm$  standard error of the mean (SEM,  $n = 3$ ), and significance analyses among DTT activities were performed by Student's  $t$  test with a 95 % confidence interval. A statistical value of  $p < 0.05$  was considered significant.

## 3 Results and discussion

### 3.1 DTT activity of laboratory-generated SOA

Here, we chose two types of SOA derived from PAHs, naphthalene and phenanthrene, as the model SOA systems. The OP of NSOA has been shown to be the highest among various types of SOA previously studied (Tuet et al., 2017a, b). At the same time, both NSOA and PSOA contain quinones which are known to be highly redox active and exhibit high OP (Cho et al., 2004; McWhinney et al., 2013). As a comparison,  $\alpha$ -pinene and limonene SOA from ozonolysis were chosen to represent biogenic SOA derived from monoterpenes. Experimental conditions and SOA yield information are summarized in Table 1.

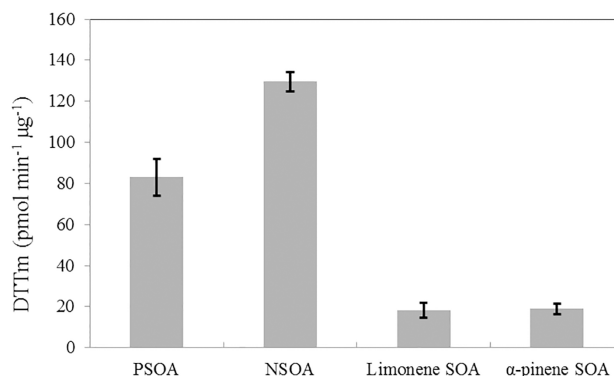
**Table 1.** Flow tube experimental conditions.

Compound	Reaction <sup>a</sup>	$\Delta\text{HC}$ ppb	$Y^{\text{b,c}}$ %	RH %
Limonene	ozonolysis	$251 \pm 23$	$25 \pm 3.9$	$14 \pm 1$
$\alpha$ -Pinene	ozonolysis	$304 \pm 17$	$19 \pm 4.2$	$14 \pm 1$
Naphthalene	photooxidation	$6436 \pm 402$	$28 \pm 6.7$	$57 \pm 5$
Phenanthrene	photooxidation	$4050 \pm 578$	$12 \pm 2.6$	$57 \pm 5$

<sup>a</sup> Temperature in all experiments is around room temperature (22–25 °C). <sup>b</sup> SOA mass yields were calculated without particle wall loss correction. <sup>c</sup> SOA density in this study was assumed to be  $1.25 \text{ g cm}^{-3}$  for monoterpene SOA (Shilling et al., 2009), and  $1.55 \text{ g cm}^{-3}$  for PAHs SOA (Chan et al., 2009).

The mass-normalized DTT decay rate,  $\text{DTT}_m$ , was applied here for OP evaluation (Fig. 2). Similar  $\text{DTT}_m$  has been reported for NSOA, with values of  $118 \text{ pmol min}^{-1} \mu\text{g}^{-1}$  by McWhinney et al. (2013) and  $110 \text{ pmol min}^{-1} \mu\text{g}^{-1}$  by Tuet et al. (2017b) for NSOA generated from chamber photooxidation under dry conditions. The  $\text{DTT}_m$  of  $\alpha$ -pinene SOA of  $19.1 \pm 2.5 \text{ pmol min}^{-1} \mu\text{g}^{-1}$  in this study is also consistent with those reported by Tuet et al. (2017b) and Jiang et al. (2017). The similar values among the different studies highlight the reproducibility of results from the DTT assay. To the best of our knowledge, the OP of PSOA and limonene SOA from this study are the first ones reported in the literature. Both SOA values derived from PAHs yield a higher OP than the two types of monoterpene SOA. The similarities in OP between limonene and  $\alpha$ -pinene SOA (cyclic monoterpenes) and between naphthalene and phenanthrene SOA (PAHs) observed in this study further confirm the hypothesis proposed by Tuet et al. (2017b) that the intrinsic OP of SOA is closely related to the molecular skeleton of the precursor.

One of the reasons for the high OP exhibited by PAH-derived SOA is the abundance of redox-active quinone moieties in SOA compounds (Lee and Lane, 2009). The cytotoxic and carcinogenic effects from quinone-like compounds are well recognized in the field of biochemistry (Bolton et al., 2000; Valavanidis et al., 2005), and the toxicity of PM has been attributed to the presence of quinones. Charrier and Anastasio (2012) have found the OP of several quinones are comparable to transition metals in ambient particles. The importance of quinone-like components to OP was also evaluated by examining the changes in DTT activity in response to the presence of 2,4-dimethylimidazole, which has been shown to be the co-catalyst of the quinone redox cycle (Jiang et al., 2017). However, McWhinney et al. (2013) quantified three quinones in NSOA using GC/MS and found that these quinones can only account for 30 % of the total NSOA DTT response. The remaining DTT activity may arise from other quinone-like compounds that have not been identified, or from other oxidation products in NSOA. Given this knowledge gap, we examine the potential roles of peroxides, oligomers and other more oxygenated products that may explain the high- $\text{DTT}$  activities of NSOA in the next sections.



**Figure 2.**  $\text{DTT}_m$  ( $\text{pmol min}^{-1} \mu\text{g}^{-1}$ ) for SOA formed from various types of hydrocarbons (phenanthrene, naphthalene, limonene, and  $\alpha$ -pinene). Each measurement was conducted in triplicates, and the error bar represents the standard error of the mean (SEM).

### 3.2 OP contribution from peroxides

One of the main hypothesis in this study is that organic peroxides contribute to OP. Organic peroxides have been identified to be major components in both laboratory and ambient OA (Jokinen et al., 2014; Lin et al., 2016; Surratt et al., 2010; Zhang et al., 2015, 2017). They can play important roles in forming high MW oligomers (Docherty et al., 2005) and highly oxygenated molecules (Mentel et al., 2015). Recent studies have shown that peroxides may also be important for OP. Kramer et al. (2016) suggested that isoprene-derived hydroxyhydroperoxide (ISOPOOH) is an essential contributor to the OP of isoprene SOA, consistent with the results of bulk peroxide measurements using 4-nitrophenylboronic acid assay (NPBA assay) by Jiang et al. (2017). Since peroxides have been proposed to be a major component in NSOA (Kautzman et al., 2010), it is essential to determine whether or not these peroxides can account for the remaining OP contribution (McWhinney et al., 2013).

Here we compare the NSOA and  $\alpha$ -pinene SOA systems to determine the role of peroxides in OP. The KI assay is known to be sensitive to all types of SOA peroxides (ROOR, ROOH, and HOOH; Banerjee and Budke, 1964), and we confirmed its sensitivity by conducting calibrations with four different peroxides (Fig. S4). Similar KI response factors were observed with hydrogen peroxide, cumene hydroperoxide, *tert*-butyl peroxide, and benzoyl peroxide. Since it is likely that the peroxides in NSOA with one aromatic ring are mostly in the form of ROOR (Kautzman et al., 2010), we used benzoyl peroxide as the mass calibration standard. The mass fraction of peroxides and  $\text{DTT}_m$  of each SOA system are shown in Table 2. A high percentage of peroxide (40–100 %) was observed in  $\alpha$ -pinene SOA, which was consistent with the results (47 %) in the study by Docherty et al. (2005). Meanwhile, a very low percentage of peroxides (< 3 %) was found in NSOA system, a result that is inconsis-



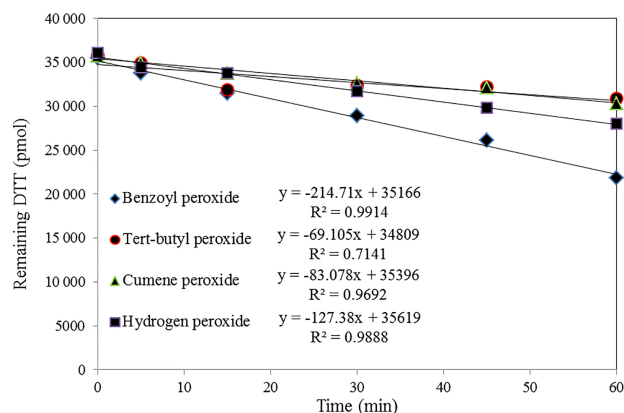
**Table 2.** SOA peroxide content and OP.

Organics	Peroxide percentage %	DTT <sub>m</sub> pmol min <sup>-1</sup> µg <sup>-1</sup>
Naphthalene SOA	< 3	100–129
α-Pinene SOA	40–100	10–20
Benzoyl peroxide	100	37

tent with previous work (> 20 %) by Kautzman et al. (2010). The difference in measured peroxide content is most likely due to the difference in the UV light source. Kautzman et al. (2010) and McWhinney et al. (2013) used H<sub>2</sub>O<sub>2</sub> photolysis under black lights (~ 350 nm), whereas in our study 254 nm UV lamps were used to photolyze O<sub>3</sub> and generate O (<sup>1</sup>D). Organic peroxides in SOA are known to be photolabile (Banerjee and Budke, 1964; Krapf et al., 2016; Wang et al., 2011) and had likely decomposed rapidly under the shorter UV wavelengths used in our studies. Despite the differences in light conditions and peroxide content, the DTT<sub>m</sub> measured for NSOA in this study is consistent with those measured in two separate studies (McWhinney et al., 2013; Tuet et al., 2017b). Also, the DTT<sub>m</sub> of NSOA was found to be significantly higher than that of α-pinene SOA, which contains a large fraction of peroxides. Therefore, from our work, there is no evidence showing that peroxides contribute significantly to the high DTT<sub>m</sub> observed in NSOA. Even if organic peroxides were present at a mass fraction of around 20 %, as reported by Kautzman et al. (2010), we expect these peroxides would react with DTT at a similar rate as benzoyl peroxide, which has a similar structure to that of the proposed peroxides. The DTT<sub>m</sub> of benzoyl peroxide (ROOR-type) is 37 pmol min<sup>-1</sup> µg<sup>-1</sup> (Fig. 3), which is around 3 times lower than the DTT<sub>m</sub> of NSOA. It should also be noted that organic hydroperoxides are the major OP contributors for biogenic SOA, such as isoprene and monoterpene SOA (Jiang et al., 2017). One of the potential mechanisms is the formation of hydroxyl radicals from the decomposition of organic hydroperoxides in water (Tong et al., 2016). For the NSOA system, our results suggest that other non-peroxide species are likely to serve as major contributors to OP.

### 3.3 OP of oligomers in NSOA

Atmospheric OA also undergoes extensive oligomerization, forming high MW compounds that have profound impacts on SOA physicochemical properties (Hallquist et al., 2009; Rudich et al., 2007; Trump and Donahue, 2014; Wang et al., 2011). Laboratory photooxidation of aromatic compounds produces a substantial fraction of oligomers in the SOA (Kalberer et al., 2004) and these oligomers may be highly functionalized (Gao et al., 2004; Tolocka et al., 2004). IMS-TOF analysis reveals that a substantial fraction of the signals in NSOA and PSOA are located in the high *m/z* range, which are associated with high MW oligomeric products. Since pre-

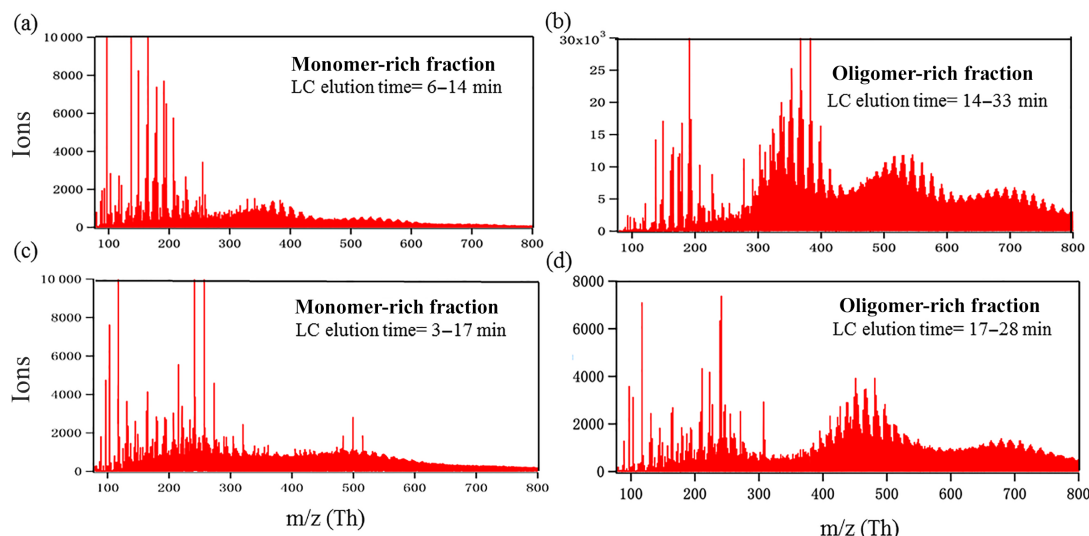


**Figure 3.** DTT activity of various types of peroxides (hydrogen peroxide, cumene peroxide, *tert*-butyl peroxide, benzoyl peroxide). With the same initial concentration of peroxide (120 µM), benzoyl peroxide has the highest DTT activity (converted to DTT<sub>m</sub> of 37 pmol min<sup>-1</sup> µg<sup>-1</sup>).

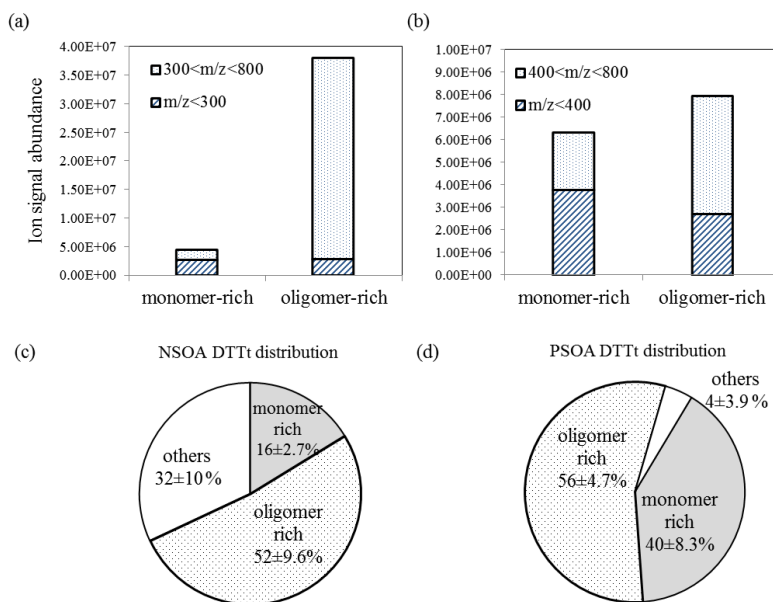
vious studies have largely focused on monomeric quinones (such as 1,2-naphthoquinone or 9,10-phenanthrenequinone), the contribution of high MW products to OP have not been studied and may explain the “missing” OP contributors.

To evaluate OP of high MW products in NSOA and PSOA, solutions of SOA extract were separated in a C18 reverse-phase column into two major fractions. As shown in Fig. 4, when analyzed by IMS-TOF, the first fraction was found to contain relatively higher signals at *m/z* associated with monomers, and the second fraction contain products with higher signals located in a higher *m/z* range. It should be noted that complete separation could not be achieved in this work. Other techniques, such as size-exclusion chromatography (Di Lorenzo and Young, 2016; Di Lorenzo et al., 2017), may yield better separation based on molecular weights, but may not be able to resolve compounds in the relatively low molecular weight range in the current study. Nonetheless, the first fraction can be qualitatively described as a “monomer-rich” fraction, and the second fraction can be regarded as an “oligomer-rich” fraction (Figs. 4, 5a, b). The DTT assay was then conducted on both of the original SOA solution as well as the two separated fractions. Since the amount of organic material in each fraction is not known, we use the total DTT activity (DTT<sub>t</sub>, in µM min<sup>-1</sup>) to qualitatively compare the oxidative capacities of the two fractions.

As shown in Fig. 5, both the monomer-rich and oligomer-rich fractions are reactive towards DTT. Compared to the DTT<sub>t</sub> of NSOA before LC separation, relative DTT<sub>t</sub> of monomer-rich fraction and the oligomer-rich fraction were 16 ± 3 and 56 ± 10 %, respectively (Fig. 5c). For PSOA, relative DTT<sub>t</sub> from monomer-rich fraction and oligomer-rich fraction were 40 ± 8 and 50 ± 5 %, respectively (Fig. 5d). In both systems, the oligomer-rich fraction caused a more rapid decay in DTT than the monomer-rich fraction even with a



**Figure 4.** IMS-TOF mass spectra for (a) monomer-rich fraction and (b) oligomer-rich fraction in NSOA, and for (c) monomer-rich fraction and (d) oligomer-rich fraction in PSOA. During a total of 46 min elution, the majority of NSOA monomers eluted at 6–14 min, and most of the oligomers eluted at 14–33 min. For PSOA system, the majority of monomers eluted at 3–17 min, and most of the oligomer-rich fraction eluted at 17–28 min.



**Figure 5.** Sum of ion signals for monomers and oligomers in monomer-rich fraction and oligomer-rich fraction for NSOA (a) and PSOA (b) systems. Relative DTT<sub>t</sub> from monomer-rich fraction and oligomer-rich fraction in NSOA (c) and PSOA (d) systems. The remaining DTT activity (others, white) is attributed to residual SOA fractions that did not clearly elute out.

lower summed ion signal of low MW constituents (Fig. 5a, b for NSOA and PSOA, respectively). These qualitative results suggest that while the current focus of health studies has been on monomeric quinones, other higher-MW products may be important for the OP of NSOA and PSOA. Specific molecular characteristics of these high MW OP contributors are currently unknown, and understanding them will be the

subject of future research. It is very likely that the oligomers also contain redox-active quinone functional groups, such as those formed on the surface of oxidized soot (Antinolo et al., 2015), and are therefore important for OP.

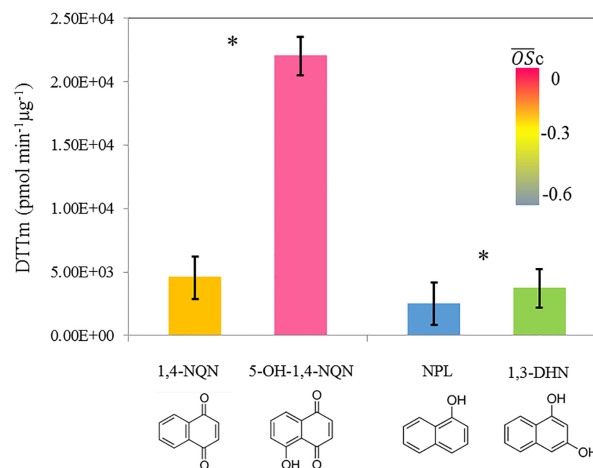


### 3.4 OP from heterogeneous oxidation

In addition to its complexity, the composition of SOA is also highly dynamic and evolves upon atmospheric oxidation (Jimenez et al., 2009). Heterogeneous oxidation in the particle phase is one of the major pathways in aerosol aging (George and Abbatt, 2010; Rudich et al., 2007) and can increase oxygen content during the functionalization processes. McWhinney et al. (2013) attributed 21 % of the NSOA's DTT activity to two quinone isomers (1,2-NQN and 1,4-NQN) in NSOA while finding a higher DTT contribution (30 %) when they took 5-hydroxy-1,4-naphthoquinone (5-OH-1,4-NQN) into consideration. Thus, we expect that oxygenated derivatives produced upon heterogeneous oxidation may also contribute to the OP of SOA, and the OP of SOA could be enhanced by heterogeneous oxidation.

We first examined the changes in OP with additional functional groups in known organic compounds. As shown in Fig. 6, two pairs of organic compounds in NSOA were chosen: 1,4-NQN and 5-OH-1,4-NQN were used to study quinone-like compounds, while naphthol (NPL) and 1,3-dihydroxy naphthalene (1,3-DHN) were used to compare phenol-like compounds. The  $\overline{\text{OSc}}$  of those four components were calculated (Kroll et al., 2011), as shown in Fig. 6. Our results demonstrated a higher DTT<sub>m</sub> for standards with higher oxidation states. For each addition of an OH group to the selected molecule, the OP increases. OP of an aromatic compound is therefore shown here to be associated with its degree of oxygenation and is demonstrated here fundamentally using individual organic compounds.

More broadly, oxidation also increases the degree of oxygenation in the bulk aerosol phase, and increases OP. Here we conducted heterogeneous oxidation by exposing filter-collected SOA to O<sub>3</sub> / N<sub>2</sub> with the same flow rate. The N<sub>2</sub> exposure group is used as the control group in order to isolate the effects of evaporation and/or decomposition at room temperature from those of heterogeneous O<sub>3</sub> oxidation. For each of the exposure (1, 12, 24 h), the DTT<sub>t</sub> of O<sub>3</sub> exposure group was normalized by the DTT<sub>t</sub> of the corresponding N<sub>2</sub> exposure group (Fig. 7a). OC loss was determined by thermal optical OC / EC analysis, and was observed to be 17 and 13 % for the O<sub>3</sub> and N<sub>2</sub> exposure groups after 24 h, respectively (Fig. 7b). Generally, DTT<sub>t</sub> of NSOA filter under O<sub>3</sub> exposure was higher than that of N<sub>2</sub> exposure. The study of Antiñolo et al. (2015) also showed an increased redox activity of soot accompanied by an increased amount of oxygenated derivatives (quinone) under heterogeneous oxidation. However, the enhanced oxidative capacity from heterogeneous ozonolysis appeared to decrease with longer exposure to O<sub>3</sub> (Fig. 7a), which we hypothesize may result from functionalization as well as fragmentation of organic molecules during heterogeneous oxidation (Kroll et al., 2009). This can be further confirmed with the observed changes in the different OC fractions, as shown in Fig. S5. Within the 24 h exposure, the volatile fractions (OC1 and OC2) of the O<sub>3</sub> exposed group

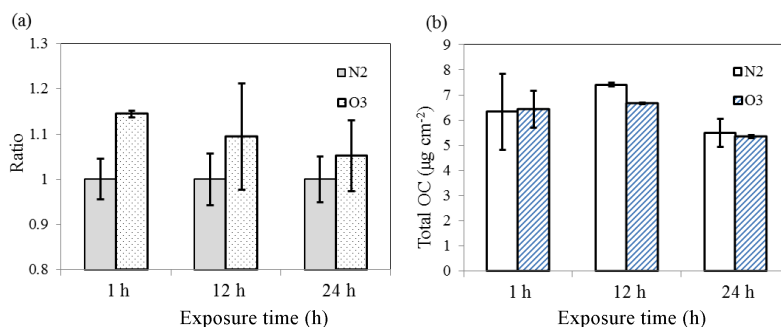


**Figure 6.** DTT<sub>m</sub> for two selected pairs of oxygenated derivatives in NSOA system (1,4-NQN vs. 5-OH 1,4-NQN and NPL vs. 1,3-DHN). Averaged carbon oxidation state ( $\overline{\text{OSc}}$ ) of each component is shown in color (color scale shown on top right). Each measurement was conducted in triplicates, and the error bar here represents the SEM. The asterisk indicates significant difference between each pair of measurements at the 95 % confidence level.

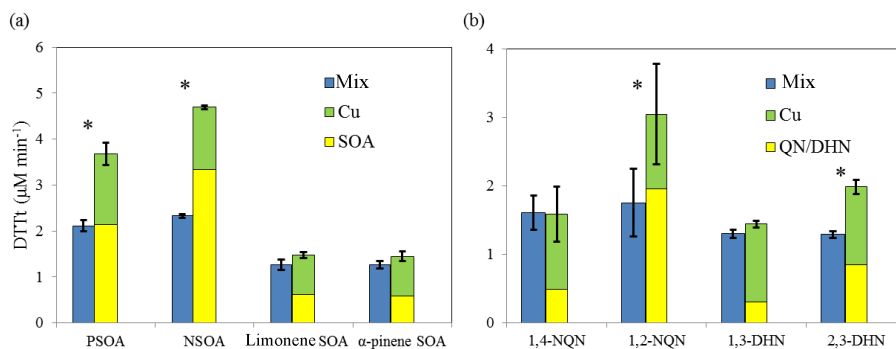
increased while the less volatile fractions (OC3 and OC4) decreased compared to the N<sub>2</sub> group, which suggested the decomposition of high MW (low volatility) species into low MW (high volatility) compounds. Previous work has also shown fragmentation can play a dominant role in a late stage of heterogeneous oxidation (Kroll et al., 2011). The overall increased volatility may lead to evaporation of smaller redox-active molecules and decrease the DTT<sub>t</sub> compared to the N<sub>2</sub> exposure group. It should also be noted that the O<sub>3</sub> concentrations to which NSOA are exposed here are about 100 times higher than typical ambient levels (Finlayson-Pitts and Pitts Jr., 2000). Assuming heterogeneous oxidation mechanisms are linear and an ambient O<sub>3</sub> concentration of 30 ppb, O<sub>3</sub> exposure for 1, 12, and 24 h under our experimental conditions represent 4, 50, and 100 days of aging in the atmosphere, respectively. Therefore, we anticipate an overall enhancement in OP under ambient conditions. Though similar observations have been made in soot (Antiñolo et al., 2015) and diesel exhausted particles (McWhinney et al., 2013), this enhancement in OP by heterogeneous oxidation is shown here for the first time in SOA particles.

### 3.5 OP changes upon mixing with Cu

Ambient PM forming from mixed sources is frequently composed of both organics and metals. To date, both organics (quinone) and transition metals (Cu, Fe, Mn, etc.) have been shown to be redox-active (Charrier and Anastasio, 2012; Xiong et al., 2017). Metals in ambient particles can range from insoluble substances to soluble cations, leading to var-



**Figure 7.** (a) Relative DTT<sub>t</sub> for NSOA after O<sub>3</sub> and N<sub>2</sub> exposure. For each of the exposure duration (1, 12, 24 h), the DTT<sub>t</sub> of O<sub>3</sub> exposure group was normalized by the DTT<sub>t</sub> of the corresponding N<sub>2</sub> exposure group. Generally, DTT<sub>t</sub> of NSOA that underwent heterogeneous ozonolysis was higher than that of the N<sub>2</sub> control. (b) OC / EC measurement results show total OC mass loss (17 and 13 % for O<sub>3</sub> and N<sub>2</sub> exposure, respectively) after 24 h exposures. Each measurement was conducted in triplicates, and the error bar represents the SEM.



**Figure 8.** Significant OP depletions were observed when PSOA ( $43 \pm 4\%$ ), NSOA ( $50 \pm 6\%$ ), 1,2-NQN ( $42 \pm 7\%$ ), and 2,3-DHN ( $35 \pm 1\%$ ) mixed with Cu (II). The asterisk indicates significant difference between a pair of bars at a 95 % confidence level.

ious health outcomes after deposition onto the human respiratory tract (Gojova et al., 2007; Oberdörster et al., 2005).

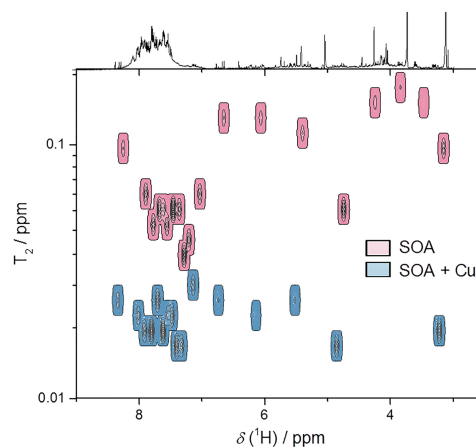
Based on the chemical composition and the assumption that DTT activities of quinones and transition metals are additive, previous studies have attempted to reconstruct the overall OP in ambient particles based on the chemical composition (Charrier and Anastasio, 2012; Charrier et al., 2015). However, addition of a transition-metal chelator did not result in significant changes in the expression of inflammatory biomarkers (Donaldson et al., 2001), suggesting that oxidative activities from different transition metals may not be additive. In our study, significant reductions in OP were observed for PSOA ( $43 \pm 4\%$ ) and NSOA ( $50 \pm 6\%$ ), when they were mixed with Cu (II) (Fig. 8a). Conversely, no significant OP reduction was observed when α-pinene or limonene SOA was mixed with Cu (II). To further investigate the cause of this reduction, we examined the OP of 1,2-NQN, 1,4-NQN, 1,3-DHN, and 2,3-DHN and the effects from mixing with Cu (II). Significant OP reductions for 1,2-NQN ( $42 \pm 7\%$ ) and 2,3-DHN ( $35 \pm 1\%$ ; Fig. 8b) were observed, but no such changes were observed with 1,4-NQN or 1,3-DHN upon metal mixing. It should be noted that a mixture of phenanthrenequinone and Cu (II) did not show a signif-

icant reduction in DTT activity in a study by Charrier and Anastasio (2012), while it is likely that the DTT measurements may be affected by the inefficiency of the quench reagent, trichloroacetic acid (Curbo et al., 2013). Furthermore, an increased level of OP reduction was observed when an increasing amount of Cu (II) was mixed with the same amount of 1,2-NQN (Fig. S7). Based on the observation that OP reduction occurs only when there are neighboring oxygenated functional groups, we hypothesize that the OP reduction is related to formation of covalent bonds between the electron-deficient Cu (II) and the electron-donating polar functional groups. The formation of quinone–copper complexes have been demonstrated previously (Dooley et al., 1990; Klinman, 1996), and may be responsible for reducing the overall OP.

To understand the underlying mechanism, <sup>1</sup>H NMR spectroscopy was used to monitor the formation of the organic–Cu complex. <sup>1</sup>H NMR has previously been applied to study the binding between metals and organics (Peana et al., 2015; Syme and Viles, 2006). The relaxometry (T2) of NSOA illustrated in Fig. 9 shows a decreasing trend on average for T2 relaxation time when Cu (II) was added to the system. Such a decrease in relaxation time indicates interactions between copper and SOA components. In general, protons adjacent

to and within aliphatic hydroxyl groups resonate between 3 and 5 ppm, while aromatic groups and double bonds arise between 5 and 9 ppm. Before the addition of copper, NSOA T<sub>2</sub> relaxation time shows a range of values consistent with a complex material with a diversity of functional groups and dynamics. After the addition of copper, both of the aliphatic hydroxyl and aromatic region showed a decrease in T<sub>2</sub> and inherit roughly similar relaxation values. This suggested copper is able to influence a wide range of SOA components, with both structural categories (aliphatic hydroxyl and aromatic) involved in copper binding to some extent. Due to the complexity in the SOA NMR spectra, it is still currently challenging to specifically identify binding between individual NSOA components and Cu. To further investigate the binding reactions, <sup>1</sup>H NMR measurements were made for the compounds present in NSOA, as previously mentioned, in the presence and absence of Cu, as shown in Fig. 10. Significant interactions between Cu and 1,2-NQN are evidenced by the broadened peak shapes (Fig. 10a), caused by the coordination of adjacent oxygen groups with the copper ion, which has also been well documented in previous studies with similar compounds (Inoue and Gokel, 1990; Schmidt et al., 1990; Tolman, 1977). Protons proximal to the binding sites are more significantly broadened, while protons further away are less affected. Conversely, mixing of 1,4-NQN with Cu led to little changes in peak shape, indicating the lack of any interactions with copper, as shown in Fig. 10b. Similar phenomena were also observed with another pair of isomers (2,3-DHN and 1,3-DHN). As shown in Fig. 10c, 2,3-DHN shows a clear change in peak shape indicating that these hydroxyl moieties on adjacent carbons are very important for copper coordination. On the other hand, the 1,3-DHN structure shows very little peak broadening when mixed with copper.

Such binding evidence was further supported by NMR relaxation data (Fig. 11). As copper is paramagnetic, it is an effective relaxation agent, and protons brought into its proximity undergo faster T<sub>1</sub> and T<sub>2</sub> relaxation that manifest themselves as spectral broadening in 1-D NMR (Fig. 10). For the epitope maps, the largest circles indicate the least interactions with copper. For both T<sub>1</sub> and T<sub>2</sub> data (Fig. 11, Table S1), protons adjacent to the binding sites a and f (proton peak assignment based on Figs. 10a, S6) underwent significant changes in relaxation indicating the metal coordinating to the neighboring oxygen groups. Proton e changed less as it is located further away from the copper binding site. While copper has a relatively mild effect on protons b and d, a significant reduction in T<sub>1</sub> and T<sub>2</sub> for proton c was observed. Such interesting phenomenon is due to the increased nuclear Overhauser effect (NOE) from the b and d sites. In the absence of binding, protons b and d would in part relax via NOE with protons a and e, which relax rapidly due to the copper binding. As such protons b and d can no longer lose magnetization via an Overhauser effect owing to a and e. Instead, they pass magnetization to position c, which underwent an

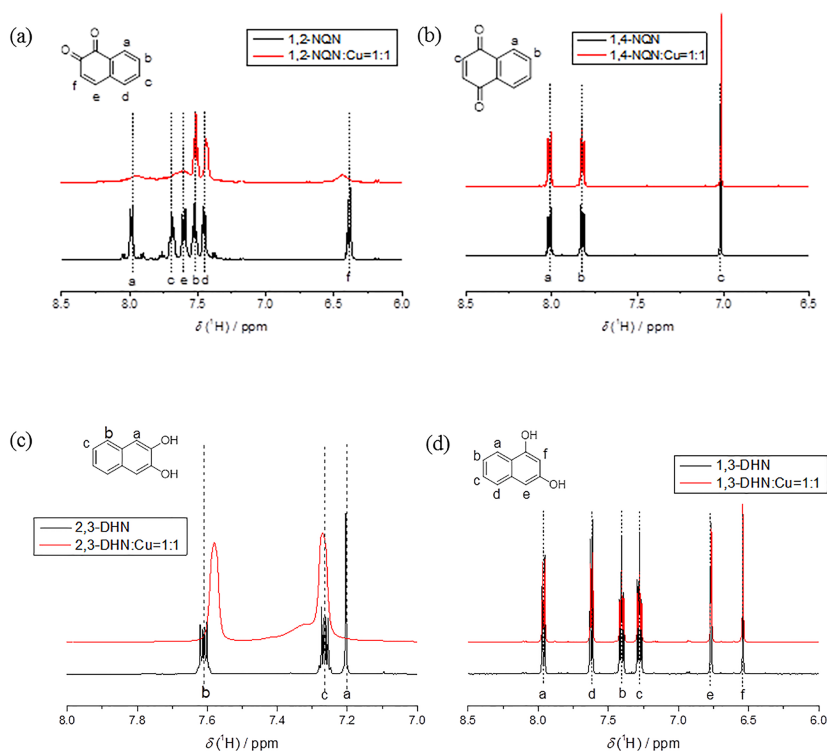


**Figure 9.** 2-D <sup>1</sup>H-NMR T<sub>2</sub> relaxation contour map for NSOA with (blue) and without copper (red) with 1-D NMR projections from the top. A general decreasing trend in T<sub>2</sub> is observed here, which indicates interactions (binding) of Cu (II) with many NSOA components.

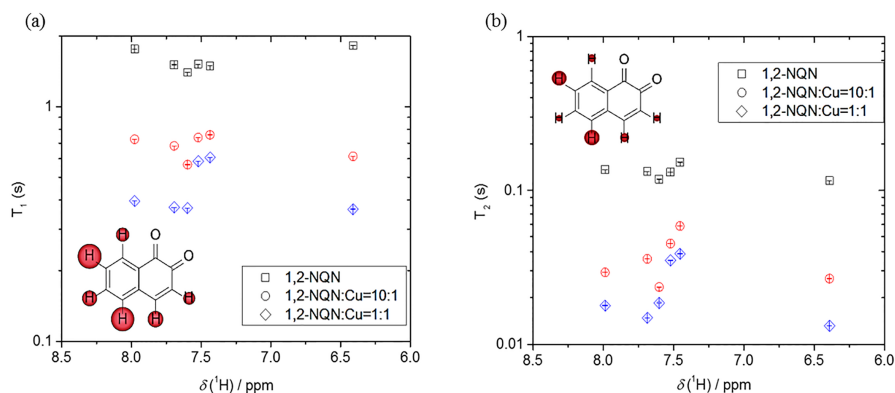
enhanced level of relaxation as a result. The NOE effect for ring systems with similar structure has also been demonstrated in several publications (Kowalewski and Maler, 2006; Rehmann and Barton, 1990). It should be noted that the increased level of OP depletion is accompanied by an increased ratio of Cu to 1,2-NQN (Fig. S4a), the degree of broadening in the 1-D NMR proton peaks becomes more significant (Fig. S4b), and the relaxation times T<sub>1</sub> and T<sub>2</sub> decrease (Fig. 11). All these observations together illustrate that the reduction of OP is proportional to the binding between Cu and organics, and further supports the mechanism behind OP depletion in the DTT assay. Based on the conclusions from the individual organic standards, the overall decrease in relaxation times for NSOA mixing with Cu (shown in Fig. 9) likely indicates that Cu (II) are binding with NSOA components, limiting the redox activities and OP of both the Cu (II) ions and the redox-active NSOA components.

#### 4 Implications

Oxidative stress caused by ROS production and antioxidant consumption is one of the most commonly studied mechanisms for PM toxicity (Nel, 2005; Rhee, 2006; Manke et al., 2013). Here we performed OP evaluation of two SOA formed from PAHs (naphthalene, phenanthrene) by the DTT assay, and investigated the linkages between SOA OP and chemical composition upon various atmospheric aging processes. SOA derived from ozonolysis of monoterpenes ( $\alpha$ -pinene, limonene) have a lower DTT<sub>m</sub> than that of the two SOA-derived PAHs, which could be attributed to the high redox-active quinone-like components. This is also consistent with the previous hypothesis that OP of SOA is highly dependent on the identity of its precursor (Tuet et al., 2017a). To fur-



**Figure 10.** 1-D  $^1\text{H}$ -NMR spectra of (a) 1,2-NQN, (b) 1,4-NQN, (c) 2,3-DHN, and (d) 1,3-DHN, and their mixture with 1 : 1 ratio of Cu (II). Both 1,2-NQN and 2,3-DHN show the broadening of  $^1\text{H}$ -NMR peaks (protons at a, c, e, f and a, b, c for 1,2-NQN and 2,3-DHN, respectively) after mixing with Cu (II).



**Figure 11.**  $^1\text{H}$  NMR relaxometry analyses for (a)  $T_1$  and (b)  $T_2$  of 1,2-NQN mixed with Cu (II) at different ratios: 1,2-NQN with no Cu (black), Cu (II): 1,2-NQN = 1 : 10 (red), and Cu (II): 1,2-NQN = 1 : 1 (blue). Both  $T_1$  and  $T_2$  decreased when Cu was introduced into the system, indicating a smaller scale of nuclear spin dynamic resulted from organic–metal binding. The molecular epitopes illustrate the influence of Cu (II) binding on an individual proton. A smaller sphere shadow on a proton denotes a larger relaxation influence from Cu (II)–organic binding.

ther link SOA OP to its chemical composition, this study also explored the possible impacts from atmospheric aging processes so as to provide mechanistic understanding for ambient observations.

Over the span of atmospheric lifetime, the mass and chemical composition of SOA can be affected by aging processes

(Kroll et al., 2009; Lim et al., 2017). The aerosol aging processes that we studied here include oligomerization, heterogeneous oxidation, and metal mixing. Apart from quinones that are well known to exhibit high OP in aerosol samples, OP contributions from peroxides in our NSOA system are likely to be insignificant. Rather, oxygenated derivatives

were shown here to contribute greater OP than their precursors in our study of selected organic individuals, and heterogeneous oxidation of NSOA was shown to lead to greater OP as well. Moreover, DTT activities of the monomer-rich and oligomer-rich fractions in NSOA separated by liquid chromatography showed oligomers are OP contributors in SOA. While organic peroxides have been proved to be very labile components with half-lives of minutes at room temperature (Krapf et al., 2016), SOA oligomers are relatively stable and highly oxygenated with their ratio of total organic molecular weight per organic carbon weight (OM/OC) similar to that of atmospheric humic-like substances (HULIS; Altieri et al., 2008). Consistent with OP contributions from the oxygenated-component enriched HULIS fraction in ambient PM (Verma et al., 2015b), this study also shows evidence for OP contributors from atmospheric aging products of PAH-derived SOA, indicating that major organic PM OP contributors could be less volatile than previously thought, and may more readily remain in the particle phase under atmospheric aging. Nevertheless, future work should focus on improving separation methods, allowing for more precise measurements of OP from SOA oligomeric constituents.

A reduction in OP was observed when mixing NSOA/PSOA with Cu (II), resulting in a non-additive effect. However, no such reduction was observed in  $\alpha$ -pinene SOA or limonene SOA. Using  $^1\text{H}$  NMR spectroscopy, we demonstrate that the reduction in OP is likely caused by binding between Cu (II) and redox active organic compounds. Both the peak broadening in 1-D NMR spectra and shorter relaxation times are observed for compounds that exhibited OP reduction (1,2-NQN, 2,3-DHN) upon Cu (II) mixing. Additionally, with a greater amount of Cu (II) mixed in, an enhanced OP reduction and a decrease in relaxation times appeared. While it is still challenging to determine which NSOA components are binding with Cu (II), the overall relaxation time also decreased when NSOA was mixed with Cu (II), indicating binding between Cu (II) and various NSOA components. Based on our work, previously recognized redox-active organic and inorganic components in ambient particles (Charrier and Anastasio, 2012; McWhinney et al., 2011; Monks et al., 1992; Turski and Thiele, 2009) may bind with each other once mixed during atmospheric aging processes (externally) or within the physiological environment of the human body (internally). Similarly, Hems et al. (2017) found that suppression of OH generation (photo-Fenton reaction) occurred, and attributed the decrease to the formation of metal–SOA complex. Although current studies demonstrate such metal–SOA binding leads to a lower OP, other studies (Gonzalez et al., 2017; Yu et al., 2018) have found an increased level of OH radicals through the formation of Fe (II)–HULIS complex. These synergistic/antagonistic metal–organic interactions are relevant to multiple health outcomes. Figuring out the dominant interactions under atmospherically/physically relevant conditions may provide hints for alleviating the health

impact of ambient particles. In the future, a more detailed understanding of SOA binding with metal components and the effects on the oxidative health outcomes will be essential. It should also be noted that the DTT assay alone may not be entirely representative of physiological ROS variations (Tuet et al., 2017a; Xiong et al., 2017). More in vitro and in vivo work should be performed to establish the relationship between chemical composition and the OP of aerosol.

**Data availability.** Data used in this study can be provided upon request to the corresponding author.

**The Supplement related to this article is available online at <https://doi.org/10.5194/acp-18-3987-2018-supplement>.**

**Competing interests.** The authors declare that they have no conflict of interest.

**Acknowledgements.** This work was supported by Natural Sciences and Engineering Research Council Discovery Grant, Canadian Foundation for Innovation John R Evans Leaders Fund, and the Ontario Early Researcher Award. The authors would like to thank Jon Abbatt, Greg Evans, and Manpreet Takhar for helpful discussion.

Edited by: Manabu Shiraiwa

Reviewed by: two anonymous referees

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