



Fungal spores overwhelm biogenic organic aerosols in a midlatitudinal forest

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Abstract. Both primary biological aerosol particles (PBAPs) and oxidation products of biogenic volatile organic compounds (BVOCs) contribute significantly to organic aerosols (OAs) in forested regions. However, little is known about their relative importance in diurnal timescales. Here, we report biomarkers of PBAP and secondary organic aerosols (SOAs) for their diurnal variability in a temperate coniferous forest in Wakayama, Japan. Tracers of fungal spores, trehalose, arabinol and mannitol, showed significantly higher levels in nighttime than daytime ($p < 0.05$), resulting from the nocturnal sporulation under near-saturated relative humidity. On the contrary, BVOC oxidation products showed higher levels in daytime than nighttime, indicating substantial photochemical SOA formation. Using tracer-based methods, we estimated that fungal spores account for 45 % of organic carbon (OC) in nighttime and 22 % in daytime, whereas BVOC oxidation products account for 15 and 19 %, respectively. To our knowledge, we present for the first time highly time-resolved results that fungal spores overwhelmed BVOC oxidation products in contributing to OA especially in nighttime. This study emphasizes the importance of both PBAPs and SOAs in forming forest organic aerosols.

1 Introduction

Biogenic particles and gases emitted from various vegetation types affect earth climate, biogeochemical cycles of carbon and nitrogen, air quality and human health. Particles such as fungal and fern spores, bacteria, viruses, algae, vegetation debris and pollen are directly emitted and thus termed as primary biological aerosol particles (PBAPs) (Després et al., 2012; Elbert et al., 2007). Being mostly in the size range of 0.1–100 μm , PBAPs often contribute to a large mass fraction of organic aerosols (OAs) but a small fraction of total particles (Pöschl et al., 2010). On the other hand, biogenic volatile organic compounds (BVOCs) are highly reactive and thus subjected to photooxidation to form fine particles of secondary organic aerosols (SOAs) and gas–particle partitioning of the reaction products (Claeys et al., 2004). There are substantial progresses in understanding the source strengths, composition and atmospheric transformations of biogenic particles, however, their contributions to OA and hence forces of their effects on climate are still uncertain.

Global PBAP emission is as high as 1000 Tg yr⁻¹ (Jaenicke, 2005). Fungal spore is the dominant component with the emission strengths of 8–186 Tg yr⁻¹ (Després et al., 2012; Elbert et al., 2007; Heald and Spracklen, 2009). Fungal spores (1–30 μm) could account for up to 45 % of coarse par-

ticulate matter (Fröhlich-Nowoisky et al., 2009). In contrast, global bacteria emissions are much smaller with abundances of 0.04–1.8 Tg yr⁻¹ (Burrows et al., 2009). Pollen could contribute up to 65 % of total PBAP mass in the flowering season, although their number contribution is only ~ 1 % of total PBAP (Manninen et al., 2014). Certain PBAPs showed close correlations with atmospheric cloud condensation nucleus (CCN) and ice nucleus (IN) (Hiranuma et al., 2015; Huffman et al., 2013). However, simulations indicated that PBAPs are of regional importance for IN formation as they contribute very little to global average ice nucleation rates (Hoose et al., 2010; Spracklen and Heald, 2014). Unveiling PBAP mass concentrations is necessary to provide benchmarks for simulating its nucleation potentials and the effect on the climate.

The most widely studied BVOCs are isoprene (C₅H₈), monoterpenes (C₁₀H₁₆) and sesquiterpenes (C₁₅H₂₄). Global emission of isoprene (309–706 Tg C yr⁻¹) accounts for ~ 50 % of total BVOC (Acosta Navarro et al., 2014; Guenther et al., 2006). Monoterpenes (~ 10 % of total VOCs) and sesquiterpenes are other important SOA precursors (Ziemann and Atkinson, 2012). Atmospheric oxidation of these BVOCs depends on oxidant levels (NO_x and O₃), humidity, light and temperature. Recent studies indicated that biogenic SOA production could be enhanced under the input of anthropogenic oxidants (Carlton et al., 2010). Under pristine conditions, BVOC oxidation and SOA formation are often accompanied by new particle formation, involving precipitation through increasing CCN and IN numbers (Pöschl et al., 2010). The formation of SOA is often in the fine particle size (Pöschl et al., 2010); however, the relative importance of PBAPs and SOA in contributing to OA is still ambiguous.

Various observational methods have been applied to explore PBAPs and SOA in natural environment both qualitatively and quantitatively. Applications of ultraviolet aerodynamic particle sizers and wideband integrated bioaerosol sensor based on fluorescence provided much information on the abundances and size distributions of PBAP (Huffman et al., 2010; Pöhlker et al., 2012). In situ observations of SOA are frequently conducted using an aerosol mass spectrometer (AMS) (Zhang et al., 2007). BVOC could also be monitored in real time using proton-transfer reaction mass spectrometry (PTR-MS) technology (Blake et al., 2009). Detections of organic tracer compounds using gas chromatography combined with mass spectrometry (GC/MS) make it possible to compare both PBAP and SOA sources simultaneously (Graham et al., 2003). Progress is achieved in boreal forest (Kourtchev et al., 2008) and the Amazon (Graham et al., 2003). However, few studies focused on the midlatitudinal forest with high time resolution (Miyazaki et al., 2012; Mochizuki et al., 2015).

In this study, we collected daytime and nighttime samples in a midlatitude forest in Wakayama, Japan and analyzed the organic samples for various PBAP and SOA tracers us-

ing GC/MS. We report for the first time that fungal spores overwhelm BSOA in contributing to aerosol organic carbon (OC). We also report diurnal variations of PBAP tracers that reflect the natural rhythm of the forest.

2 Experimental

2.1 Site and sampling

Aerosol sampling was conducted 20–30 August 2010 during a field campaign at Wakayama Forest Research Station, Kyoto University, in the Kii Peninsula, Japan (34.06° N, 135.52° E, ca. 750 m a.s.l.) (Han et al., 2013, 2014). The dominant species of the forest vegetation are coniferous trees of *Cryptomeria japonica* (Japanese cedar), *Chamaecyparis obtuse* (Japanese cypress) and *Pinus densiflora* (Japanese red pine). Previous studies indicated that monoterpenes are abundantly emitted from these tree species (Matsunaga et al., 2011; Mochizuki et al., 2011). Total suspended particulate (TSP) samples were collected on precombusted (450 °C, 6 h) quartz filters (Pallflex 2500QAT-UP, 20 cm × 25 cm) using a high-volume air sampler (Kimoto AS-810B) at a flow rate of ~ 60 m³ h⁻¹ at the ground level. Each of the quartz samples were kept in a sealed glass jar at < -20 °C during transport. The quartz filters were then stored at ~ -20 °C in the laboratory prior to chemical analysis. We believe that the chemical losses were not large under the low temperature and sealed condition (Kawamura et al., 2010). During the campaign, four field blanks were collected following the sampling procedures, except for the running of pump for only 10 s.

Nighttime (from ~ 18:00 to ~ 06:00 LT) and daytime (from ~ 06:00 to ~ 18:00 LT) samples were collected separately each day. Of the whole sampling period, the backward trajectories were calculated using HYSPLIT dispersion model starting at every hour of the day and for different starting height (500, 1000 and 1500 m a.g.l.) (Han et al., 2013, 2014). It can be seen that the air masses were mainly originated from the Asian continent or passed over the Japanese islands during 20–25 August, which were expected to deliver pollutants from anthropogenic sources. On the other hand, air masses were mainly originated from the western North Pacific during 26–30 August, which create an environment close to the pristine forest. On 23 and 28 August, samples were collected every 3 h (namely, 06:00–09:00, 09:00–12:00, 12:00–15:00, 15:00–18:00, 18:00–21:00, 21:00–24:00, 24:00–3:00 and 03:00–06:00 LT). A total of 36 samples were collected during the campaign.

2.2 Chemical analysis

For the determination of organic molecular tracers in aerosols, filter samples were first extracted with dichloromethane/methanol and derivatized by N,O-bis-(trimethylsilyl)trifluoroacetamide (BSTFA) mixed with

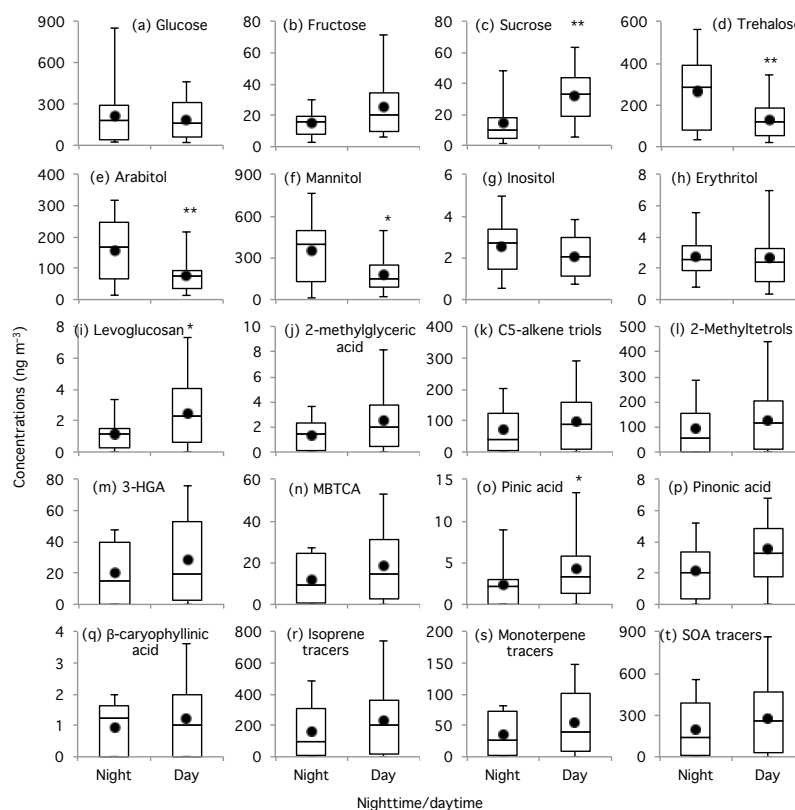


Figure 1. Diurnal variations of organic compounds in forest aerosols in Wakayama, Japan during 20–30 August 2010. The 10th, 25th, 50th, 75th and 90th percentiles of the nighttime and daytime samples were shown in the box-and-whisker plot. The means were shown in solid circles. Statistically significant values were also shown (t test, * $p < 0.05$; ** $p < 0.01$).

pyridine to convert OH groups to trimethylsilyl (TMS) ethers and COOH groups to TMS esters (Simoneit et al., 2004). The extracts were kept at $\sim -20^{\circ}\text{C}$ and then organic compounds were quantified within 24 h by GC/MS using an Agilent 7890A GC equipped with HP-5ms capillary column ($30\text{ m} \times 0.25\text{ mm} \times 0.25\text{ }\mu\text{m}$) coupled to Agilent 5975C mass selective detector (MSD). Individual compounds were identified by comparing the mass spectra with those of authentic standards or literature data and quantified by comparing the peak areas with the internal standard (C_{13} n-alkane) (Zhu et al., 2015b). Recoveries for the authentic standards or surrogates were generally better than 80 %, except for pinonic acid of $\sim 60\%$, as obtained by spiking standards to precombusted quartz filters following extraction and derivatization. No peak was found for the target compounds in the field and laboratory blanks. The analytical errors by duplicate analyses were less than 15 %. The detection limits of the reported compounds were $105\text{--}557\text{ }\mu\text{g }\mu\text{L}^{-1}$, which correspond to ambient concentrations of $12\text{--}64\text{ }\mu\text{g m}^{-3}$ under a typical sampling volume of 700 m^3 and usage of 10 cm^2 filter section.

Aerosol OC was determined using a thermal/optical carbon analyzer (Sunset Laboratory Inc., USA) following the Interagency Monitoring Protected Visual Environments (IM-

PROVE) thermal evolution protocol (Birch and Cary, 1996). The analytical error in replicate analyses was within 8 %. Samples were corrected for field blanks, in which OC level was $< 5\%$ of the samples. Inorganic ions were measured using an ion chromatograph (761 Compact IC, Metrohm, Switzerland). SO_4^{2-} and NO_3^- were discussed in the study. The detection limits were $\sim 0.1\text{ ng m}^{-3}$. The laboratory analytical error was within 5 %. Samples were corrected for field blanks.

Statistical significance of daytime and nighttime organic tracers was evaluated based on Student's t test at $p < 0.05$ and $p < 0.01$ level, respectively. To explore the effects of anthropogenic pollutant inputs on BSOA formation, linear correlations (Pearson) between SOA tracers and NO_3^- and SO_4^{2-} were investigated. The contributions from various sources to OC were estimated using tracer-based methods using conversion factors according to the preceding studies.

3 Results and discussion

3.1 PBAP tracers reflecting diurnal rhythm of fungal spores and pollen

3.1.1 Fungal spore tracers

Four primary saccharides (trehalose, glucose, fructose and sucrose) and four sugar alcohols (arabitol, mannitol, inositol and erythritol) were detected in the forest aerosols as PBAP tracers. Trehalose, arabitol and mannitol were the major species of fungus and have been used as tracers for fungal spores (Jia and Fraser, 2011). These compounds showed significantly higher levels in nighttime (262 ± 168 , 156 ± 102 and $349 \pm 243 \text{ ng m}^{-3}$ for trehalose, arabitol and mannitol, respectively) than in daytime (128 ± 90.2 , 76.0 ± 54.9 and $183 \pm 139 \text{ ng m}^{-3}$, respectively) (t test, $p < 0.05$, Fig. 1d–f). High abundances of these fungal spore tracers were resulting from the nocturnal sporulation under near saturated relative humidity (Fig. 2). In nighttime, the mean relative humidity was 96 % in the sampling period. Nevertheless, the mean daytime relative humidity was still as high as 88 %. Such a moist environment caused enrichment of fungal spores and their discharges, resulting high abundances of their tracer in aerosols. The sum of trehalose, arabitol and mannitol contributed to 70 % of total PBAP tracer mass, demonstrating an important role of fungi in regulating the land–atmosphere exchange in the forest region.

Diurnal profiles of fungal spore tracers showed similar pattern on 23 and 28 August, two typical days representing regional and local phenomenon, respectively. Minima of trehalose, arabitol and mannitol appeared around noon and increased dramatically in the evening on both days (Fig. 3). On 23 August, fungal spore tracers peaked at 21:00–24:00 LT and decreased slowly thereafter. In comparison, these compounds peaked slightly earlier at 18:00–21:00 LT on 28 August. The maxima of these compounds on 23 August were higher by a factor of ~ 1.5 than those on 28 August. This phenomenon was likely related to the precipitation that occurred at 13:00–17:00 LT, 23 August (Fig. 2), which favors the growth and release of fungal spores.

3.1.2 Other PBAP tracers and levoglucosan

Sucrose, a key component of pollen grain (Yttri et al., 2007), showed significantly higher levels ($p < 0.01$) in daytime ($31.9 \pm 17.7 \text{ ng m}^{-3}$) than nighttime ($14.6 \pm 14.2 \text{ ng m}^{-3}$) (Fig. 1c). The diurnal profile of sucrose showed peaks in the afternoon (Fig. 3a), indicating the flowering time of the vegetation. As the blossom season of the dominant coniferous trees is March–April in the region, the source of sucrose is likely from the pollen of the widely distributed floor vegetation, *Ambrosia artemisiifolia*, *Artemisia indica* and plants of *Poaceae* family, which are flowering during the sampling period.

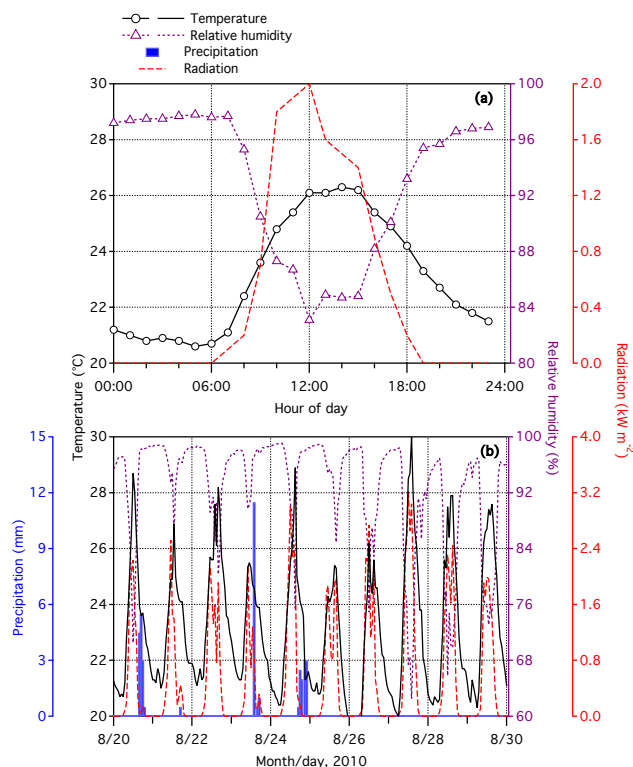


Figure 2. Meteorological parameters during 20–30 August 2010 in the Wakayama forest for (a) averaged diurnal variations of temperature, relative humidity and solar radiation and (b) hourly means over the periods. Hourly precipitation was also shown in (b).

Glucose and fructose are dominantly originated from plant debris (Medeiros et al., 2006; Zhu et al., 2015a) and sometimes from microorganisms and soil dust (Rogge et al., 2007; Simoneit et al., 2004). There was no serious continental outflow of Asian dust over the sampling period (based on observations by Japan Meteorological Agency). Insignificant diurnal variations of these two compounds imply that they were emitted both in the daytime and the nighttime in the forest possibly from plants and microorganisms (Figs. 1a–b, 3a). The diurnal variation of planetary boundary layer height could also contribute to the observed variations of glucose and fructose and other PBAP tracers. In the daytime, enhanced vertical mixing might cause dilution effect to the observed PBAP tracers. While in the nighttime, the low wind speed favors the buildup of PBAP tracers in the canopy (Fig. S1 in the Supplement), resulting in the elevated levels (Fig. 3a, b). Interestingly, the mass concentrations of glucose were 1 order of magnitude higher than those of fructose. This might be related to the higher abundance of glucose in trees. For example, tissues of a wide range of subtropical plants were found to contain more glucose than fructose (Baker et al., 1998).

Being exclusively produced by the thermal depolymerization and dehydration of cellulose and hemicellu-

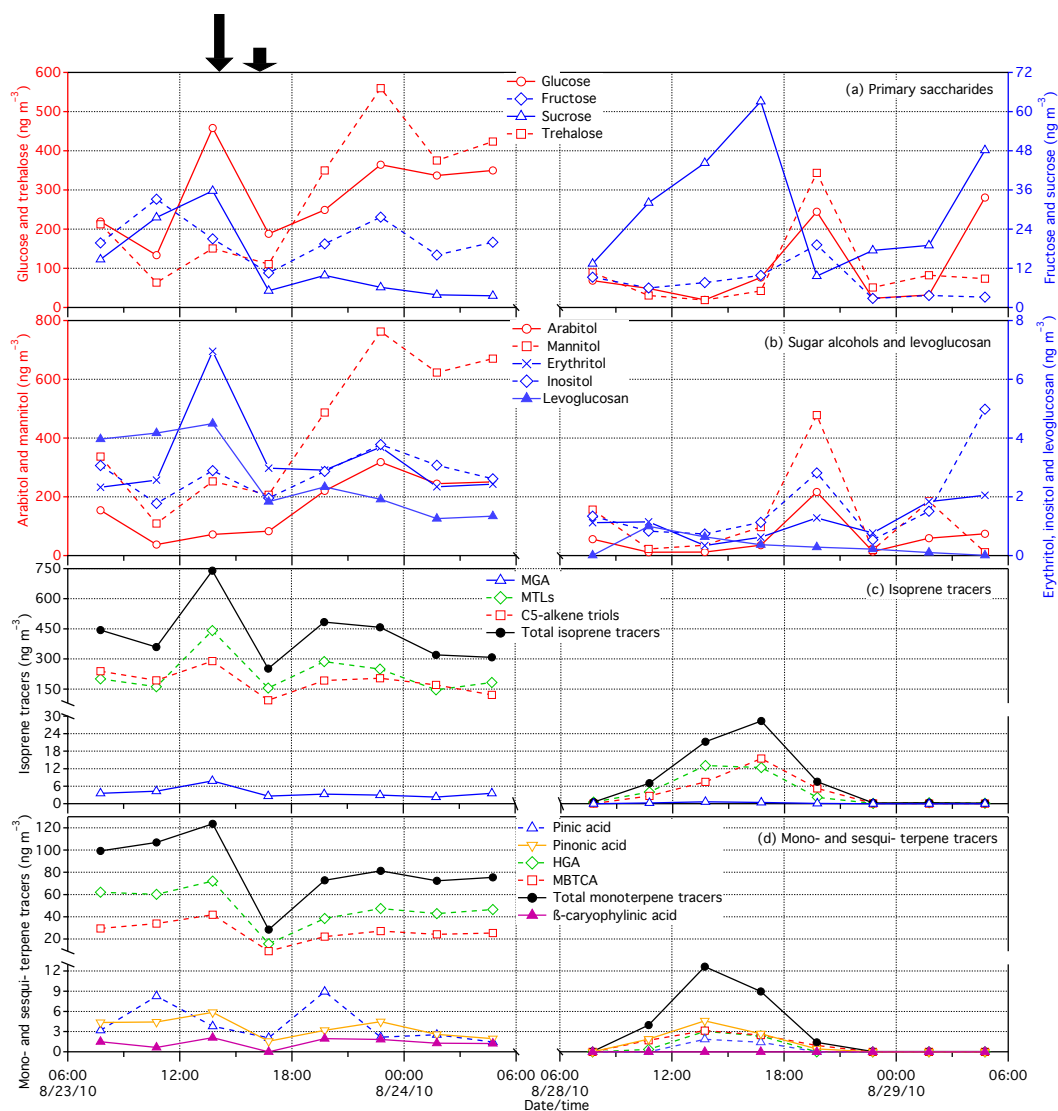


Figure 3. Diurnal profiles of organic compounds in forest aerosols in Wakayama on 23–24 August (left) and 28–29 August 2010 (right).

lose, levoglucosan has been used as a specific tracer for biomass burning (Simoneit et al., 1999; Zhu et al., 2015b). Levoglucosan showed significantly higher levels ($p < 0.05$) in daytime ($2.52 \pm 2.08 \text{ ng m}^{-3}$) than nighttime ($1.14 \pm 0.94 \text{ ng m}^{-3}$). Moreover, we found higher levels of levoglucosan 20–25 August ($2.97 \pm 1.64 \text{ ng m}^{-3}$) than 26–30 August ($0.53 \pm 0.44 \text{ ng m}^{-3}$); those concentrations in the latter period are close to the regional background level from Okinawa Island (Zhu et al., 2015b). In the nighttime, levoglucosan may have been deposited to the canopy as a result of dampened vertical mixing and relatively low wind speed (Fig. S1). In addition, local burning activities could also contribute to the daytime elevation of levoglucosan although they might be insignificant.

3.1.3 Daytime abundances of SOA tracers

A total of 11 compounds were detected as SOA tracers. Six isoprene tracers, including 2-methylglyceric acid (MGA), two 2-methyl tetrols (2-methylthreitol and 2-methylerythritol, MTLs) and three C5-alkene triols (3-methyl-2,3,4-trihydroxy-1-butene, cis-2-methyl-1,3,4-trihydroxy-1-butene and trans-2-methyl-1,3,4-trihydroxy-1-butene, MTHBs), showed higher levels in daytime (Fig. 1j–l). Similar daytime abundances were also found for four monoterpene tracers: pinonic acid (PNA), pinic acid (PA) ($p < 0.05$), 3-hydroxyglutaric acid (HGA) and 3-methyl-1,2,3-butanetricarboxylic acid (MBTCA) (Fig. 1m–p). Total isoprene–SOA tracers (281 ± 274 and $199 \pm 207 \text{ ng m}^{-3}$ in daytime and nighttime, respectively) were more abundant by a factor of ~ 5 than total monoterpene–SOA tracers

Table 1. Linear correlation coefficients (r) between secondary organic aerosol tracers and NO_3^- and SO_4^{2-} in forest aerosols during 20–30 August 2010, Wakayama, central Japan.

| | | MGA ^a | C5T | MTLs | Isoprene tracers | 3-HGA | MBTCA | PA | PNA | Monoterpene tracers |
|--------------------|--------------|-------------------|--------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|---------------------|
| NO_3^- | All data set | 0.08 | -0.13 ^b | -0.06 | -0.09 | -0.11 | -0.01 | -0.02 | 0.18 | -0.06 |
| | Daytime | -0.07 | -0.26 | -0.15 | -0.2 | -0.27 | -0.22 | -0.25 | 0.01 | -0.25 |
| | Nighttime | -0.32 | -0.36 | -0.34 | -0.35 | -0.30 | -0.18 | -0.33 | -0.19 | -0.27 |
| SO_4^{2-} | All data set | 0.84 ^e | 0.87 ^e | 0.84 ^e | 0.86 ^e | 0.91 ^e | 0.9 ^e | 0.70 ^e | 0.65 ^e | 0.91 ^e |
| | Daytime | 0.86 ^e | 0.89 ^e | 0.84 ^e | 0.88 ^e | 0.91 ^e | 0.92 ^e | 0.76 ^e | 0.67 ^d | 0.92 ^e |
| | Nighttime | 0.82 ^e | 0.82 ^e | 0.82 ^e | 0.83 ^e | 0.90 ^e | 0.86 ^e | 0.53 ^c | 0.58 ^c | 0.89 ^e |

^a Secondary organic aerosol tracer compounds are as follows: MGA, 2-methylglyceric acid; C5T, C5-alkene triols; MTLs, 2-methyl tetrols; PA, pinic acid; PNA, pinonic acid; 3-HGA, 3-hydroxyglutaric acid, MBTCA, 3-methyl-1, 2, 3-butanetricarboxylic acid. ^b Negative values indicate negative correlations. ^c $p < 0.05$;

^d $p < 0.01$; ^e $p < 0.001$.

(54.6 ± 50.2 and $36.3 \pm 33.6 \text{ ng m}^{-3}$, respectively). Isoprene is more abundantly emitted than monoterpenes from this coniferous forest. This is consistent with the observations at a mountain site in central Japan (Fu et al., 2014). Sesquiterpene tracer, caryophyllinic acid, showed consistently low levels in both daytime ($1.22 \pm 1.23 \text{ ng m}^{-3}$) and nighttime ($0.94 \pm 0.83 \text{ ng m}^{-3}$) (Fig. 1q). The daytime abundances of isoprene and monoterpene tracers reflected the production of SOA by the photooxidation of BVOC in the forest.

3.2 Influence of anthropogenic oxidant on SOA formation

The diurnal profile of SOA tracers showed a clear single-peak pattern in the afternoon on 28–29 August. Total isoprene tracers peaked at 15:00–18:00 LT (28.4 ng m^{-3}), while total monoterpene tracers peaked slightly earlier at 12:00–15:00 LT (12.7 ng m^{-3}) (Fig. 3c–d). Their afternoon peaks are consistent with the maxima of ambient temperature and solar radiation (Fig. 2). These results indicate that SOA is formed by photooxidation of isoprene and monoterpenes emitted from local vegetation under natural conditions. Our results are consistent with the afternoon peaks of semivolatile oxygenated organic aerosol (SVOOA) component determined by the positive matrix factorization of the AMS spectra collected during the same campaign (Han et al., 2014), which were produced by photooxidation of freshly emitted BVOCs from the local forest.

Interestingly, on 23–24 August, days with regional anthropogenic oxidant input, isoprene and monoterpene tracers showed not only afternoon peaks at 12:00–15:00 LT but also nighttime peaks (Fig. 3c–d) following the intensive precipitation during 13:00–17:00 LT (Fig. 2). These results suggest that SOA is continuously produced even in nighttime with sufficient oxidants. AMS observation in the same campaign indicated that highly oxidized low volatility oxygenated organic aerosols (LVOOA) are even more abundant than that of SVOOA on 23–24 August, especially in nighttime (Han et

al., 2014). Nighttime production of SOA was also observed in the United States (Zhao et al., 2013).

Moreover, on days with regional input of oxidants, the SOA tracer levels were 1–2 orders of magnitude higher than those under natural conditions both in daytime and nighttime. Meanwhile, we found positive relations between SO_4^{2-} and isoprene–SOA tracers as well as monoterpene–SOA tracers (Table 1). These phenomena could be explained by the following reasons. Firstly, regional BSOA could be transported along with anthropogenic oxidants. Observations at Okinawa Island, southern Japan indicated that the variations of monoterpene–SOA tracers and caryophyllinic acid are related to the continental outflow of BVOC oxidation products (Zhu et al., 2016). Secondly, the earlier-generation products of BVOC oxidation tend to shift to the particle phase during gas–particle partitioning given their semivolatile nature (Cahill et al., 2006), such a shift might be more prominent on days with aged regional aerosols accompanied by a daytime increase of organics determined by AMS (Han et al., 2014). The third possibility is that the anthropogenic oxidants might promote BSOA formation. We focused on the third possibility and further testified this hypothesis.

We examined the degrees of BSOA tracer buildups in daytime as compared to nighttime (ΔBSOA represents the mathematical difference between the daytime BSOA tracer level and the nighttime BSOA tracer level) (Table 2). ΔBSOA values during 20–25 August were higher by more than 1 order of magnitude than during 26–30 August. These results implied that, although transported BSOA might contribute to their elevation, anthropogenic oxidants substantially promoted the SOA formation in the Wakayama forest. A recent field study in the southeastern US suggested that the formation of methyl tetrols is mediated by SO_4^{2-} , being consistent with our results (Xu et al., 2015). They also reported that NO_3^\bullet radical was involved in the nighttime monoterpene oxidation. However, inorganic NO_3^- did not correlate with monoterpene–SOA tracers during nighttime, indicating that the contribution of NO_3^\bullet radical to nighttime monoterpene

Table 2. Daytime buildups (differences between daytime and nighttime levels) of SOA tracers on days with (20–25 August) and without (26–30 August) input of anthropogenic oxidants in forest aerosols, Wakayama, central Japan.

| Period | MGA* | C5T | MTLs | Isoprene tracers | 3-HGA | MBTCA | PA | PNA | Monoterpene tracers | CPA | BSOA tracers |
|--------------|------|------|------|------------------|-------|-------|------|------|---------------------|------|--------------|
| 20–25 August | 2.06 | 47.1 | 60.9 | 110 | 14.1 | 12.0 | 3.31 | 1.52 | 30.9 | 0.10 | 141 |
| 26–30 August | 0.02 | 1.73 | 2.33 | 4.08 | 0.16 | 0.22 | 0.37 | 1.18 | 1.93 | 0.51 | 6.53 |
| 20–30 August | 1.17 | 27.2 | 35.3 | 63.7 | 7.98 | 6.87 | 2.03 | 1.37 | 18.2 | 0.28 | 82.2 |

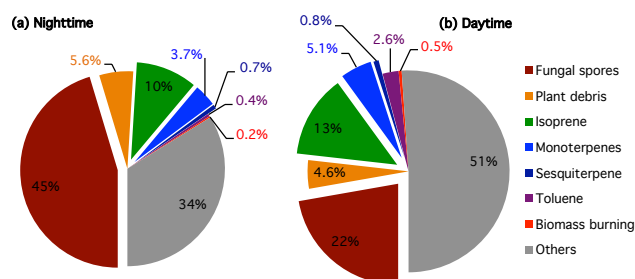
* Secondary organic aerosol tracer compounds are as follows: MGA, 2-methylglyceric acid; C5T, C5-alkene triols; MTLs, 2-methyl tetrols; PA, pinic acid; PNA, pinonic acid; 3-HGA, 3-hydroxyglutaric acid; MBTCA, 3-methyl-1, 2, 3-butanetricarboxylic acid; CPA, β -caryophyllinic acid; BSOA tracers are the sum of isoprene, monoterpene and sesquiterpene (CPA) tracers.

oxidation is not clear in Wakayama aerosols, which deserve further investigation.

3.3 OC fractions from fungal spores and BVOC oxidation

The contributions of each source component to OC were estimated using tracer-based methods. OC fractions from fungal spores were estimated based on mannitol levels assuming that there were 1.7 pg mannitol and 13 pg OC per spore (Bauer et al., 2008). The secondary organic carbon (SOC) fraction resulting from VOC oxidation was estimated following Kleindienst et al. (2007) based on the chamber experiment. Contributions of plant debris to OC were estimated based on glucose following Puxbaum and Tenze-Kunit (2003). Contributions of biomass burning aerosols to OC were estimated based on levoglucosan following Zhu et al. (2015b). It is noteworthy that, although each tracer method gave a reasonable estimate in the original studies, uncertainties would be introduced in the current study considering the site differences. Still, the estimations are inferable for the contributions from each source to OC in the forest aerosols.

Fungal spores contribute a major fraction (45 %) to OC in nighttime. In daytime, although the mass fraction was about half of that in nighttime (22 %), fungal spores are still the largest known source of OC (Fig. 4). In comparison, BVOC oxidation products (BSOC) contribute to 15 and 19 % of OC in nighttime and daytime, respectively. Interestingly, the large BSOC fraction of OC mainly occurred on days with anthropogenic influences (20–25 August) (17 % in daytime and 23 % in nighttime). On days close to natural conditions (26–29 August), BSOC has a small contribution to OC (4.0 % in nighttime and 4.7 % in daytime). The larger contributions of BSOC to OC might be associated with the BSOC that are transported from the Asian continent, a stimulating effect of anthropogenic oxidants on daytime BSOA formation, as well as the enhanced gas to particle conversion of BVOC. These results suggest that increased fossil fuel combustion and the subsequent oxidant emissions would cause a larger contribution of BSOA to OA, which may in part compensate the human-caused global warming by acting as CCN. Such ef-

**Figure 4.** Contributions from different sources to organic carbon in the Wakayama forest aerosols during the nighttime and the daytime.

fects are supposed to be incorporated into model simulations of the future climate in the Anthropocene.

It is noteworthy that the contribution of plant debris to OC (5.6 % in nighttime and 4.6 % in daytime) could not be neglected. These fractions account for only the primary OC sources based on glucose level, and could be underestimated. Nevertheless, there are still quite large amounts of OC not being determined (34 and 51 % in nighttime and daytime, respectively). Other than the uncertainties introduced by the tracer methods, a notable fraction of OC could be contributed from humic-like substances that are chemically and physically similar with terrestrial and aquatic humic and fulvic acids (Graber and Rudich, 2006). Mono- and di-carboxylic acids, originating from a broad range of primary organic compounds, could contribute 38–44 % of OC (Legrand et al., 2013). Pollen could also contribute to OC because of its large size. Amines are another source of OC, but not well studied (Sintermann and Neftel, 2015). These sources from various sectors need to be further characterized for an unambiguous understanding of the sources of forest aerosols.

Among few studies on the observation-based quantification of OC/OA sources of forest aerosols, our results are in accordance with those in the Amazon rainforest, where primary biological aerosol and SOA accounted for 68.5 and 18.8 % of total particle masses, without consideration of the diurnal variations (Pöschl et al., 2010). At an urban site with various OA sources (Bakersfield, California), Zhao et al. (2013) reported that SOA accounted for 78 and 66 % of OA during the daytime and the nighttime, respectively, while

primary sources accounted for 15 % of OA. Under a changing climate and uncertain anthropogenic emission scenario, although OA could be contributed by complicated sources in a pristine forest environment, our results demonstrate that biogenic particles and gases are rather the major sources of organic aerosols in the forest atmosphere.

4 Conclusions

We collected daytime and nighttime samples in a midlatitude coniferous forest in Wakayama, central Japan. Focusing on fungal spores and BSOA, we quantified organic tracers containing information of various aerosol sources. Night-time higher levels of trehalose and sugar alcohols were clarified under near saturated relative humidity, indicating the flourishing of fungal sporulation. BSOA tracers were more abundant in the daytime, indicating their photooxidation production among other sources. Enhanced BSOA tracer levels were observed on days with notable input of anthropogenic oxidant. Using tracer-based methods, we estimated that fungal spores contribute dominantly to OC in forest aerosols not only in the nighttime but also in the daytime. The study emphasizes the importance of fungal spores in affecting biogenic aerosols in forest ecosystem.

5 Data availability

The data of this study is available upon request to C. Zhu.

The Supplement related to this article is available online at doi:10.5194/acp-16-7497-2016-supplement.

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