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Supplement of

Real-time measurements of secondary organic aerosol formation and aging from ambient air in an oxidation flow reactor in the Los Angeles area

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1 Quantification of AMS Reactor Data

All aspects of quantification of AMS data are the same as described by Hayes et al. (2013). Here we describe only those aspects where additional analysis or corrections are needed specifically for the reactor output data.

1.1 AMS Collection Efficiency

Quantification of AMS concentration data requires a correction for particle bounce at the vaporizer, referred to as the collection efficiency (CE; Canagaratna et al., 2007). The composition-dependent CE formulation of Middlebrook et al. (2012) was used by Hayes et al. (2013) to estimate CE for the ambient data, leading to good intercomparisons with multiple collocated instruments as documented by that study. The same methodology has also been applied to reactor output measurements.

Although the focus of this paper is OA formation and aging, a brief summary of the observed evolution of the inorganic species: (a) Sulfate formation proceeds as expected from the $\text{OH} + \text{SO}_2$ reaction. A quantitative analysis of sulfate formation is shown in Palm et al. (2016), which reports results from a similar experiment from our group, but in a forest environment. That analysis provides evidence that the corrections for losses of low volatility species developed in that work are appropriate. (b) Nitrate formation is more complex since $\text{OH} + \text{NO}_2$ is a fast reaction, but HNO_3 is semivolatile and the formation of NH_4NO_3 also depends on the availability of $\text{NH}_3(\text{g})$. (c) The aerosols in the output of the flow reactor during CalNex are neutralized, similarly to the ambient aerosols (Hayes et al., 2013).

23 Fig. S2a shows the time series of reactor and ambient aerosol concentrations and estimated CE.
24 Ambient CE periodically rises above 0.5 due to larger fractions of ammonium nitrate aerosol,
25 which leads to reduced particle bounce (Middlebrook et al., 2012). The reactor typically formed
26 additional ammonium sulfate and ammonium nitrate beyond ambient concentrations at the same
27 time as ambient levels peak for those compounds, thus the reactor CE profile has a very similar
28 temporal structure to ambient. However, Fig. S2b shows that the estimated CE increases at the
29 highest reactor OH_{exp} , due to additional ammonium nitrate formation in the reactor with increased
30 photochemical age.

31 Highly acidic particles, as indicated by the ammonium balance, can also lead to increased CE in
32 the AMS (Middlebrook et al., 2012). The ammonium balance method compares the measured
33 ammonium to that required to fully neutralize observed sulfate, nitrate, and chloride (Zhang et al.,
34 2007), as shown in Fig. S3. Ambient and reactor results have near identical slopes that are
35 indistinguishable from the one-to-one line within the uncertainties of the measurements, signifying
36 full neutralization for both. Furthermore, this comparison indicates that the reactor is producing
37 similar inorganic composition to that observed in the atmosphere as nitric acid and sulfuric acid
38 gases are formed in the reactor and fully neutralized by ammonium forming ammonium nitrate
39 and ammonium sulfate. Thus no correction of CE due to the presence of highly acidic particles are
40 needed in this study.

41 Comparison of AMS and SMPS measurements for ambient and reactor data, shows that ambient
42 data falls along a one-to-one line, indicating both instruments are measuring the same amount of
43 mass within the uncertainties (Fig. S4a). Reactor output data has a slightly higher slope of 1.14,
44 i.e. the AMS measures ~14% higher mass than the SMPS from the reactor and also shows a cluster
45 of points where $SMPS > AMS$ due to periods where substantial mass is formed at small particle

46 sizes (see below). Both slopes are within the combined uncertainties of the two measurements.
47 Fig. S4b shows the relative increase in aerosol concentration in the reactor (i.e. ratio of reactor to
48 ambient concentrations) for the SMPS vs. AMS which also compare well, on average (slope =
49 1.05), but with considerable scatter, most of which is likely due to additional measurement noise
50 introduced from ratioing multiple short measurements. Evaporation of freshly formed NH_4NO_3 in
51 the longer residence times in the SMPS (compared to the faster AMS analysis) where the sheath
52 flow may have reduced NH_3 and HNO_3 gas concentrations, has been observed with this
53 experimental setup, and may be a cause of the slightly larger slope for reactor output conditions.
54 A small underestimation of AMS CE for the reactor conditions could also result in this observation.
55 It is also possible that the AMS relative ionization efficiency (RIE) of organic species is lower for
56 more oxidized species (Jimenez et al., 2003; D. Murphy, pers. Comm. 2015), although no clear
57 evidence has been reported for ambient data (e.g. Docherty et al., 2011). If that effect played a
58 dominant role here, we would expect the reactor slopes to be lower, rather than slightly higher than
59 1. Thus we conclude that any RIE changes are small and cannot be separated from other effects
60 such as small changes in CE, nitrate evaporation in the SMPS, or differences in particle
61 transmission (next section).

62 **1.2 Accounting for Particle Mass below the AMS Lens Transmission**

63 As the reactor exposed ambient air to high levels of OH and O_3 , new particle formation and growth
64 was sometimes observed. To fully account for the mass of all particles formed in the reactor, it is
65 necessary to quantify the mass of small particles below the AMS lens transmission size (Zhang et
66 al., 2004). SMPS data was used to estimate the total mass concentration below the AMS size cut.
67 First, particle transmission from plumbing line losses was corrected using the Particle Loss

68 Calculator (von der Weiden et al., 2009) for this experimental plumbing and flowrate configuration
69 for both reactor and ambient SMPS data, with transmission curves as shown in Fig. S5. Second,
70 the measured SMPS mass that is below the AMS transmission curve was estimated using a
71 published AMS lens transmission parameterization (Knote et al., 2011) multiplying the SMPS
72 size-dependent mass by the size-dependent AMS lens fractional loss (1-transmission). Figure S6a
73 shows a time series of estimated reactor and ambient mass missed by the AMS due to transmission
74 losses. Since corrections needed to account for the contribution of these small sizes to total mass
75 is small for ambient data (on average 1.7%), Hayes et al. (2013) did not apply a correction to AMS
76 ambient data. Fig. S6b shows the estimated fraction of the reactor output mass that is below the
77 AMS lens transmission size vs. total photochemical age in days (at $\text{OH} = 1.5 \times 10^6 \text{ molec. cm}^{-3}$).
78 An average of 6.2% of the total reactor output mass is estimated to be below the AMS lens
79 transmission, with no dependence on photochemical age except possibly at the highest values (>20
80 days of age).

81 We note that the AMS measurements from the reactor may be biased ~6% low, on average, and
82 sometimes as much as 20%. This non-measured mass likely has a large OA fraction (see Fig. S8).
83 Thus, reactor-reported mass enhancement above ambient may be underestimated by these
84 amounts. Given the 6.2% AMS underestimation from particle transmission of small sizes in the
85 reactor, and the apparent 14% overestimation in the AMS vs SMPS comparison, but overall good
86 agreement in the relative enhancement of total aerosol between both instruments, we have not
87 corrected for these differences as the net correction would be small and within the uncertainties of
88 the measurement, while the correction process would introduce additional noise.

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91 **Supplementary Captions**

92 **Figure S1:** Results of computerized fluid dynamics (CFD) simulations comparing two OFR
93 configurations. **(a)** Tube inlet, similar to Lambe et al. (2011); **(b)** Large open face inlet (11.9 cm
94 diameter) as used in this field study. Colors are contours of positive horizontal velocity. White
95 regions involve horizontal velocities, i.e. recirculation regions. The extensive recirculation regions
96 of case (a) are almost completely removed in case (b), resulting in a narrower residence time
97 distribution. Simulations were conducted using the FLUENT software, using cylindrical
98 symmetry, with air at 1 atm and 293 K.

99 **Figure S2:** **(a)** Estimated AMS collection efficiency (CE) and corresponding AMS mass
100 concentration time series for ambient and reactor data (after applying CE correction). **(b)**
101 Estimated CE vs. OH exposure (OH_{exp}) in the reactor for all reactor measurements and averages
102 for 7% quantiles.

103 **Figure S3:** Measured vs. predicted ammonium assuming full neutralization (“Ammonium
104 balance”) for ambient and reactor data. Linear orthogonal distance regression fit lines, slope and
105 R^2 for each are also shown.

106 **Figure S4:** **(a)** Scatter plot of AMS mass vs. mass estimated from SMPS measurements for
107 ambient and reactor data, with linear orthogonal distance regression fit slope and R^2 for each. A
108 one-to-one line and +/-15% region is shown for reference. **(b)** Relative enhancement ratio from
109 AMS and SMPS data with raw data, 20-minute averaged smooth data, linear orthogonal distance
110 regression, line, fit slope and R^2 for each.

111 **Figure S5:** Estimated particle transmission of inlet plumbing vs. particle diameter for reactor and
112 ambient sampling lines for both AMS and SMPS measurements, calculated using the particle loss
113 calculator of von der Weiden et al. (2009).

114 **Figure S6:** **(a)** Time series of SMPS mass measured below the AMS lens transmission size for
115 ambient and reactor measurements. **(b)** Percent of estimated mass not measured by AMS, due to
116 on particle losses in sampling lines and the AMS lens transmission at small sizes, for the reactor
117 vs. total photochemical age in days (at $\text{OH} = 1.5 \times 10^6 \text{ molec. cm}^{-3}$), where all data is colored by
118 ΔOA mass with average 5% quantiles and standard error bars.

119 **Figure S7:** Modeled fate of low volatility organic gases (LVOCs) formed in the reactor vs. OH_{exp}
120 including wall loss, reaction with OH, condensation on aerosol, and exiting the reactor, with a fit
121 for the fraction condensing on aerosols in the reactor.

122 **Figure S8:** AMS mass size distribution (vs. vacuum aerodynamic diameter, d_{va}) for reactor and
123 ambient OA, averaged from 20:00 on 2 June 2010 – 00:20 on 9 June 2010 for average nighttime
124 ambient and reactor with no internal OH_{exp} (dark reactor), and for ~ 3.7 days and ~ 23.5 days aging.

125 **Figure S9:** Times series of benzene, 1,3,5-trimethylbenzene, and toluene on top panel. Time
126 series of ambient OOA, reactor OA mass enhancement, maximum reactor mass enhancement,
127 and O_x on bottom panel.

128 **Figure S10:** Ratio of organic aerosol to excess carbon monoxide (above background) vs. total
129 photochemical age in days (at $\text{OH} = 1.5 \times 10^6 \text{ molec. cm}^{-3}$) for (a) the same data as Fig. 9,
130 showing all data used to produce averages for quantiles of ambient and reactor vapor-loss
131 corrected data. Also shown are the expected decays of benzene, toluene, and 1,3,5-
132 trimethylbenzene in the reactor vs. total photochemical age in days (at $\text{OH} = 1.5 \times 10^6 \text{ molec. cm}^{-3}$),
133 using reaction rates from Atkinson et al. (2006). (b) The same data as Fig. 9, showing reactor
134 vapor loss-corrected data, but where excess CO is decreased by reaction with OH in the reactor,
135 including means for 12% quantiles. Results from field studies in the northeastern US and Mexico
136 City are shown for comparison to previous observations (DeCarlo et al., 2010). A fit to the data
137 when CO is assumed to react with OH is shown.

138 **Figure S11:** The ratio of the gain of oxygen of OA observed in the reactor ($\Delta \text{Oxygen in OA} =$
139 $\text{O}_{\text{atoms, reactor}} - \text{O}_{\text{atoms, ambient}}$) to the total number of OH collisions with OA in the reactor vs. total
140 photochemical age. The estimated number of OH collisions is calculated based on the
141 methodology outlined in appendix A of DeCarlo et al. (2008).

142 **Figure S12:** Top panel: Mass fraction remaining (MFR) for OA vs. thermal denuder
143 temperature for this CalNex-LA dataset, using the methods described in Huffman et al. (2008;
144 2009). Bottom panel: estimated volatility distribution of particle- and gas-phase species,
145 calculated from the thermal denuder profile using the method of Faulhaber et al. (2009), on
146 bottom panel.

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Figure S1.

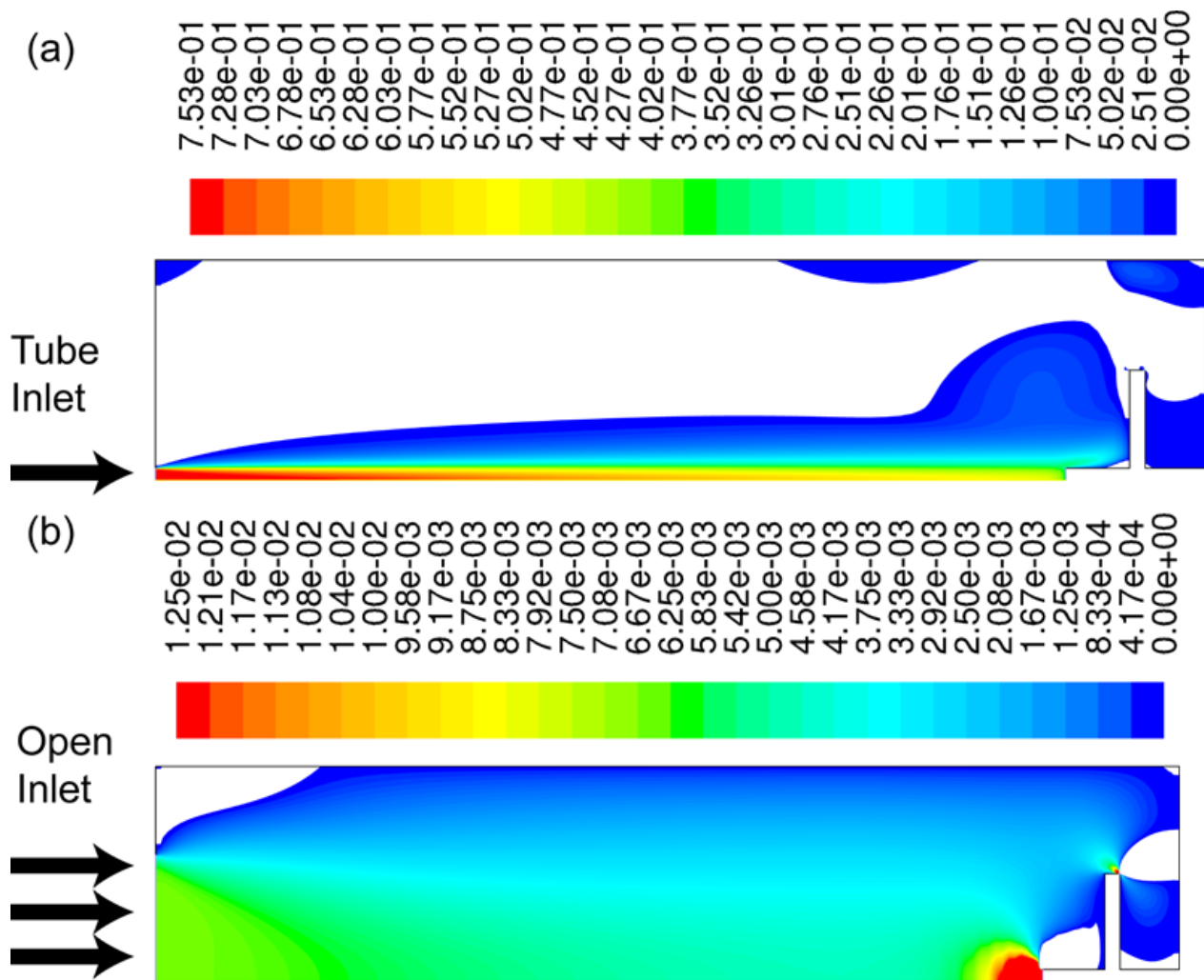


Figure S2.

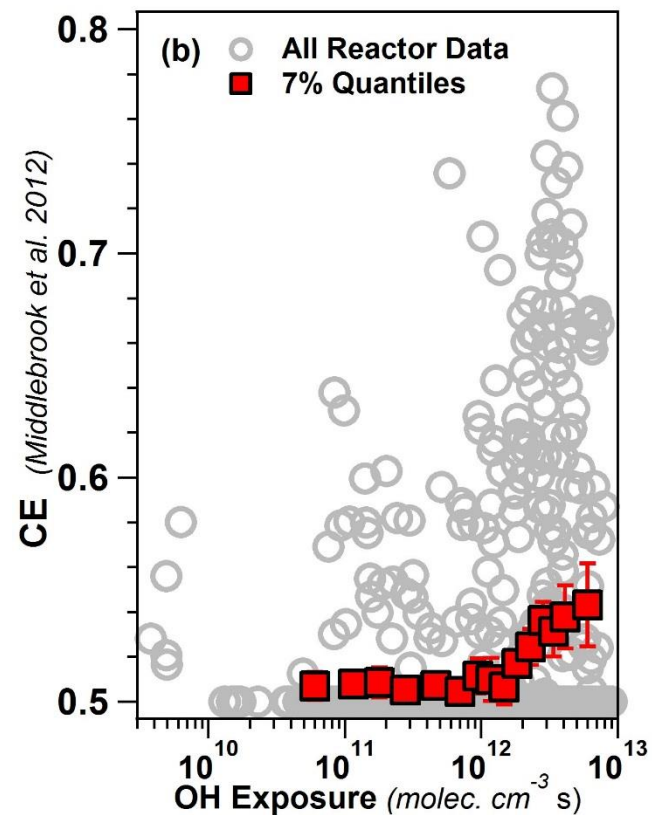
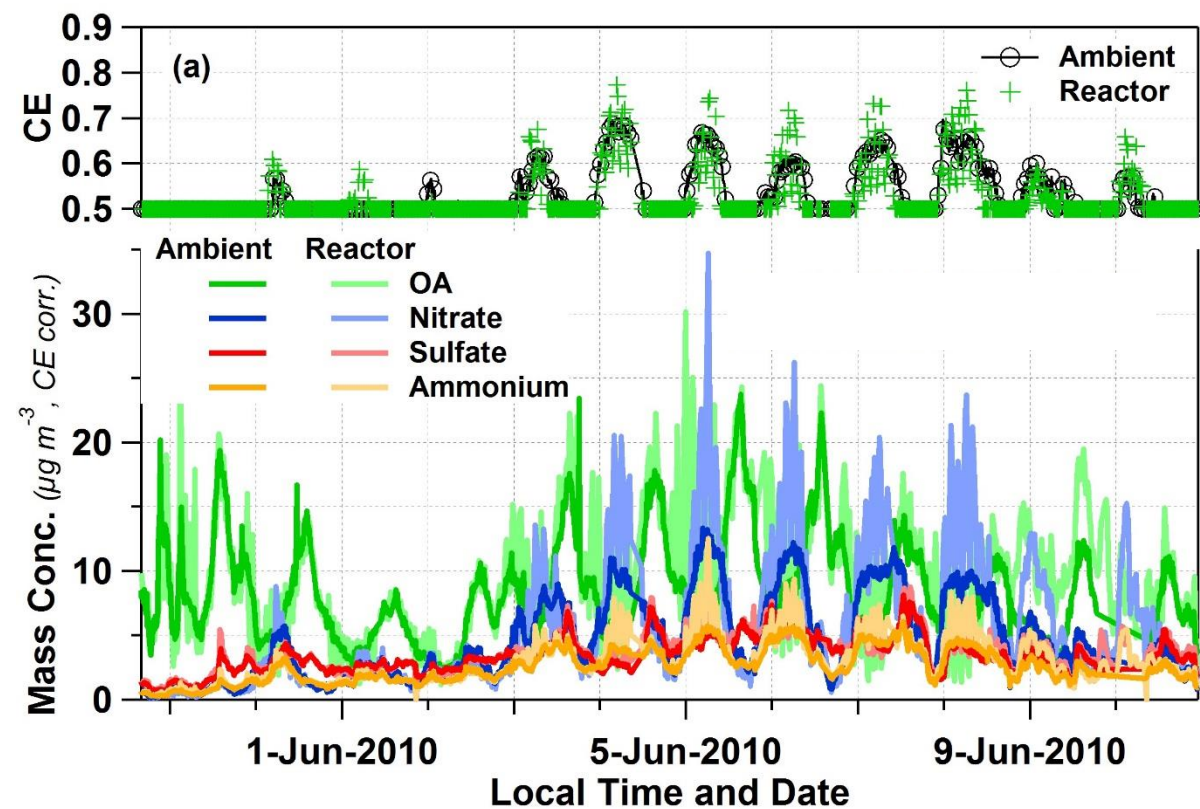


Figure S3.

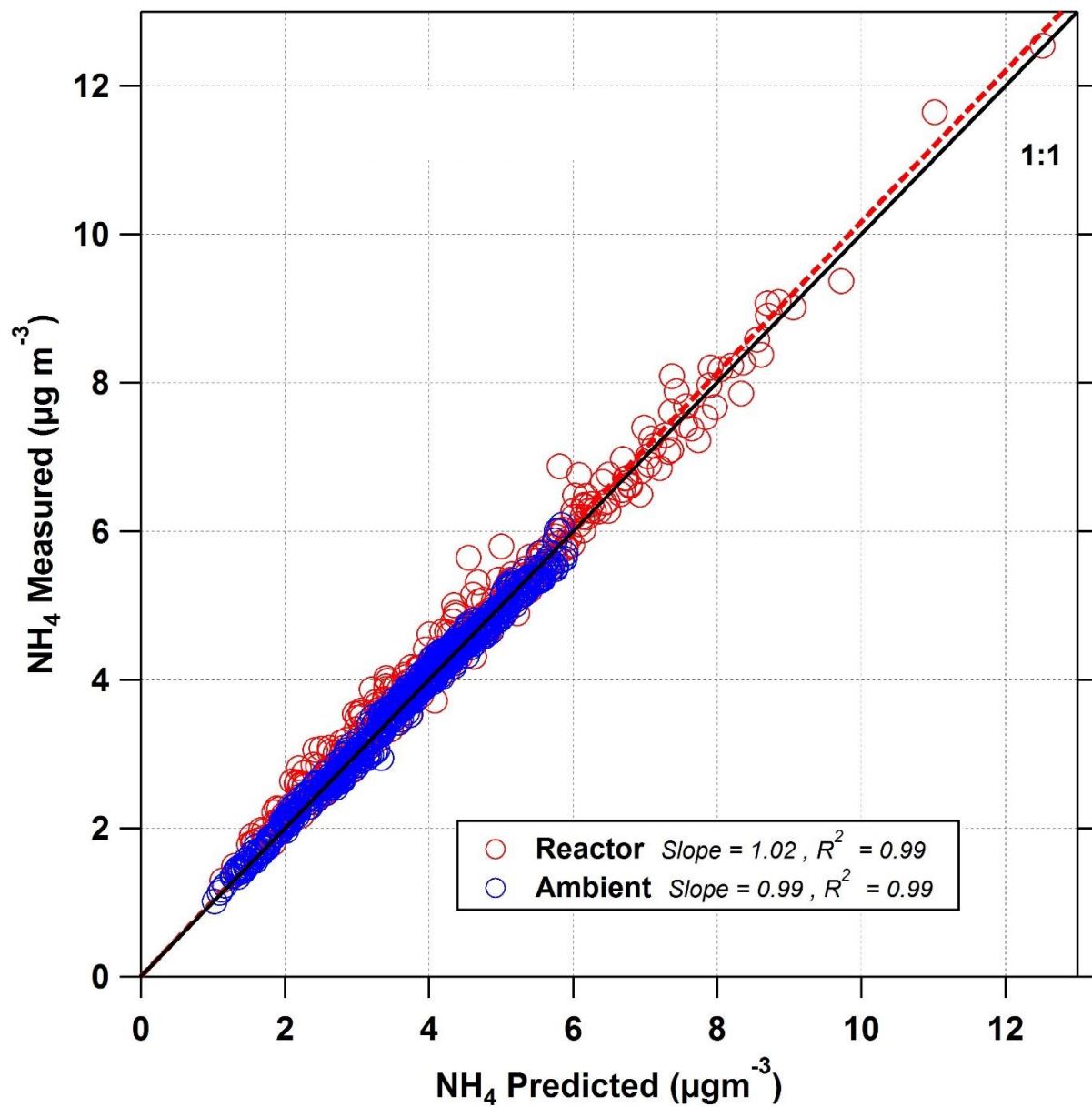
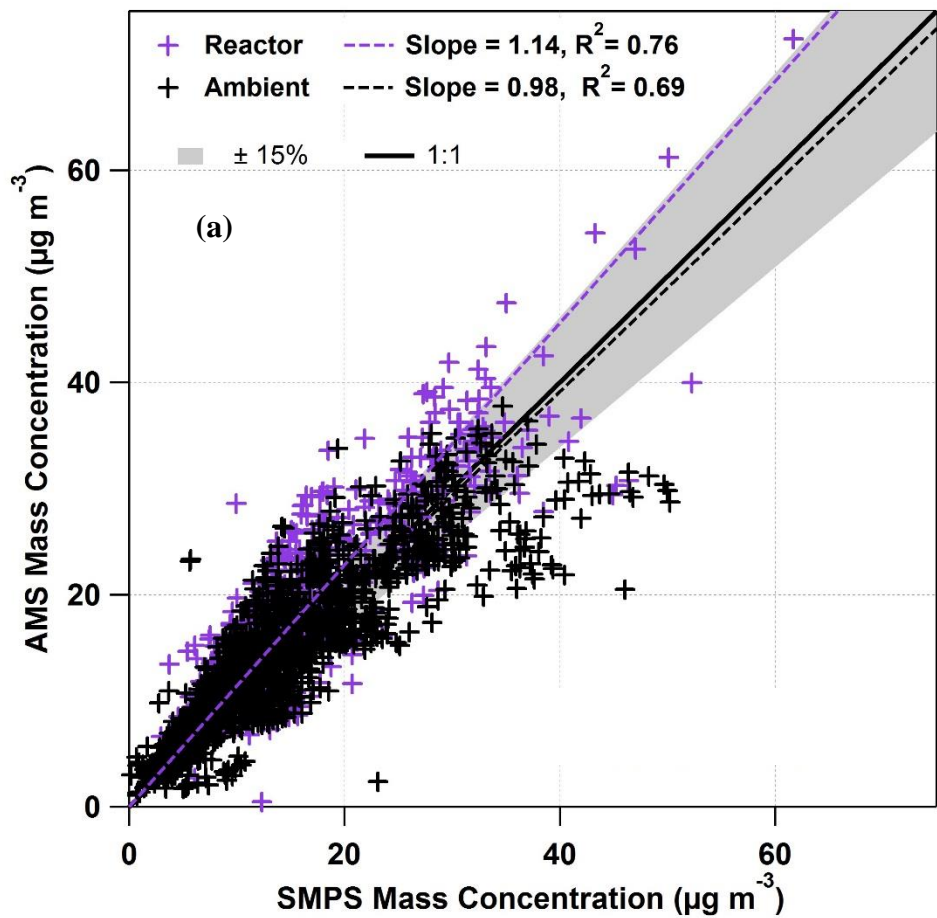


Figure S4.



(b)

Figure S5.

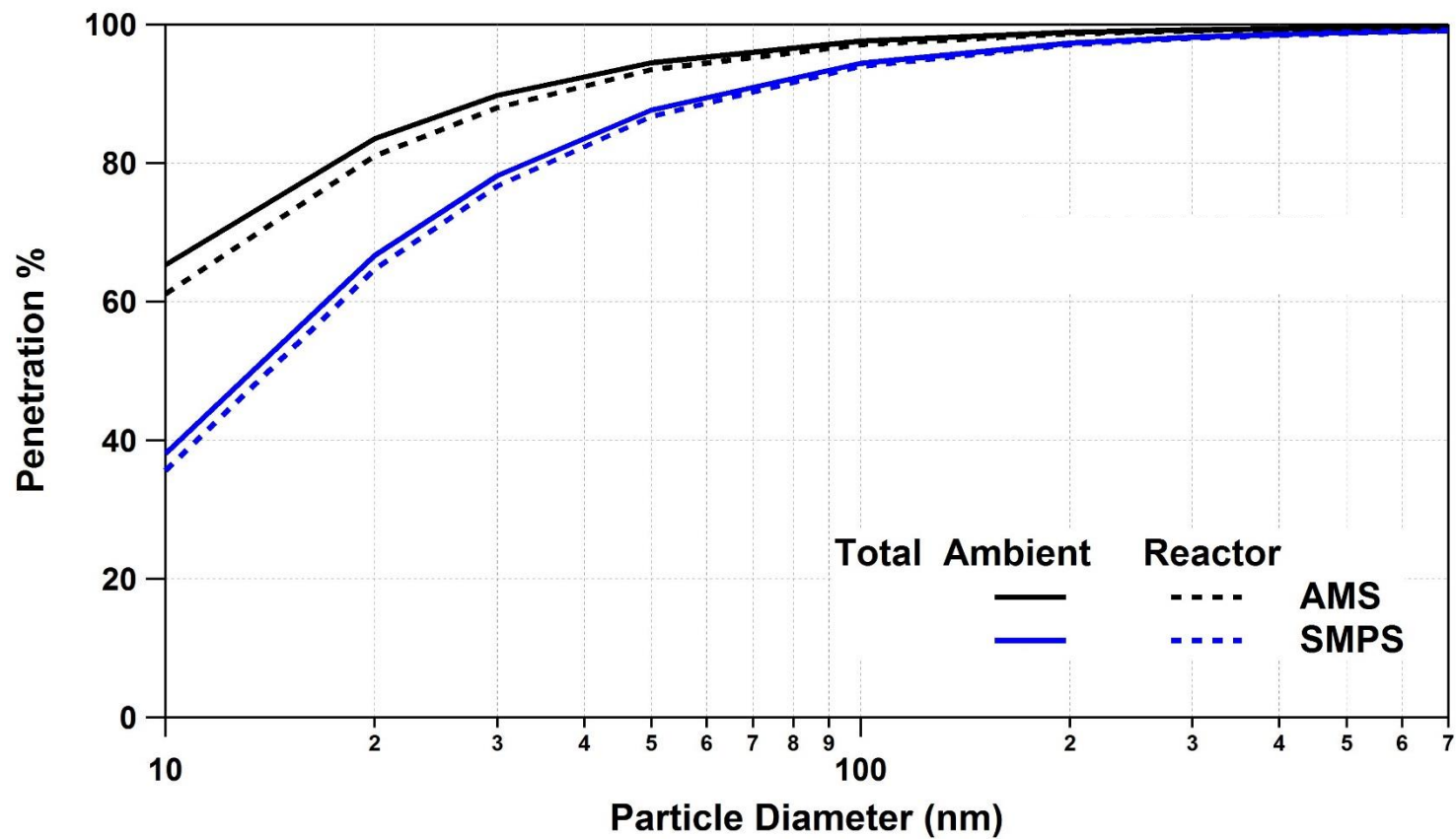


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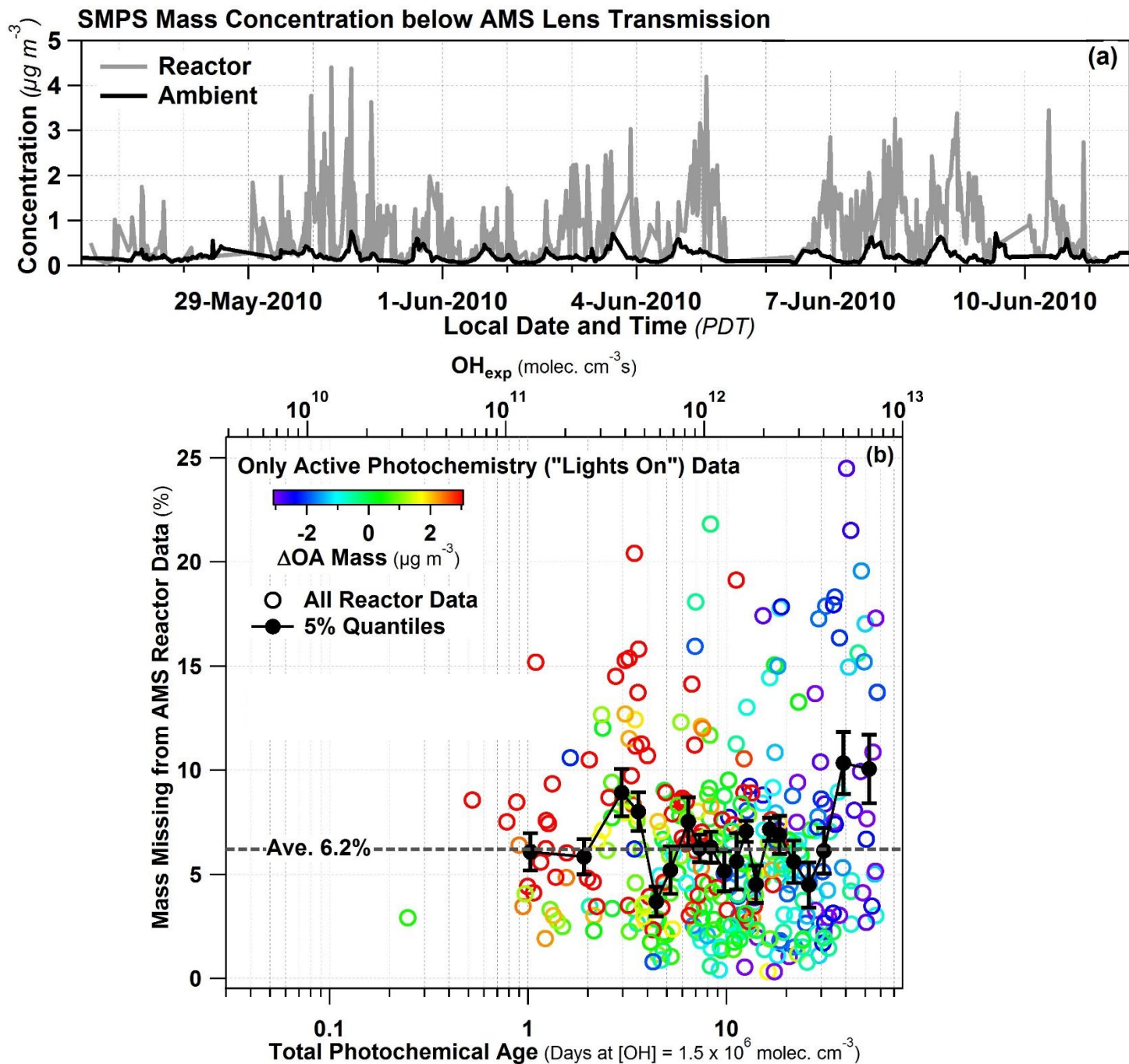


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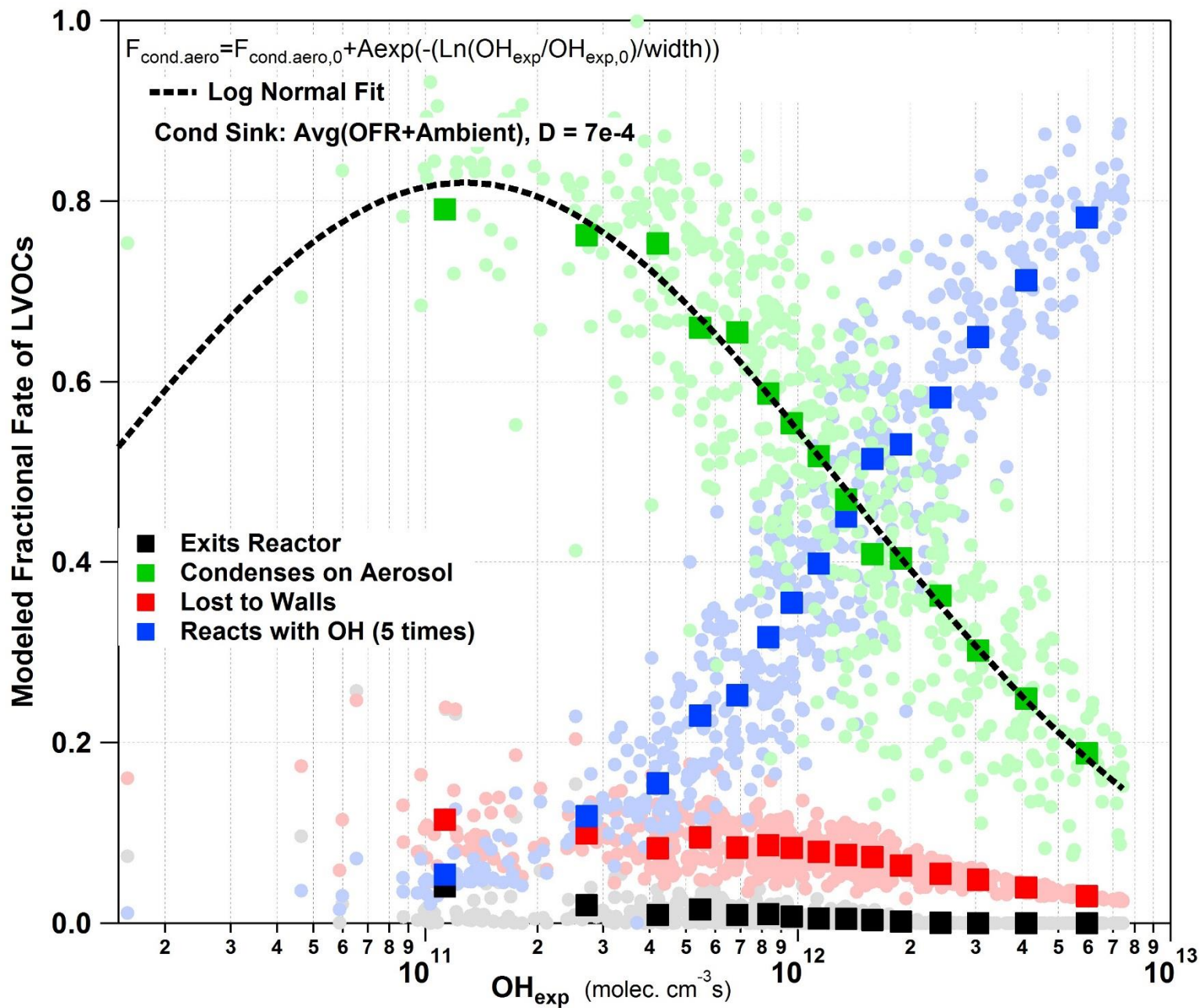


Figure S8.

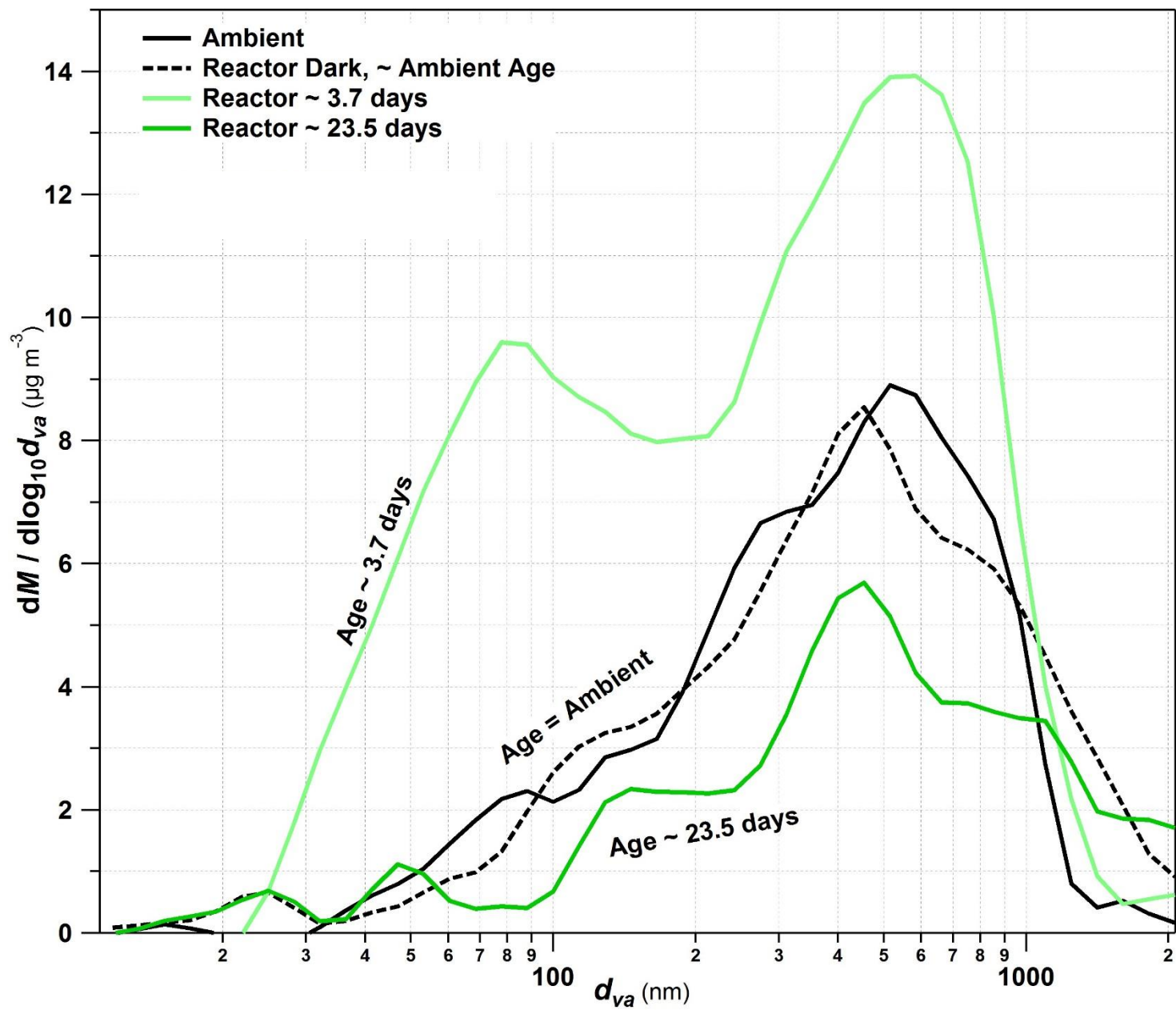


Figure S9.

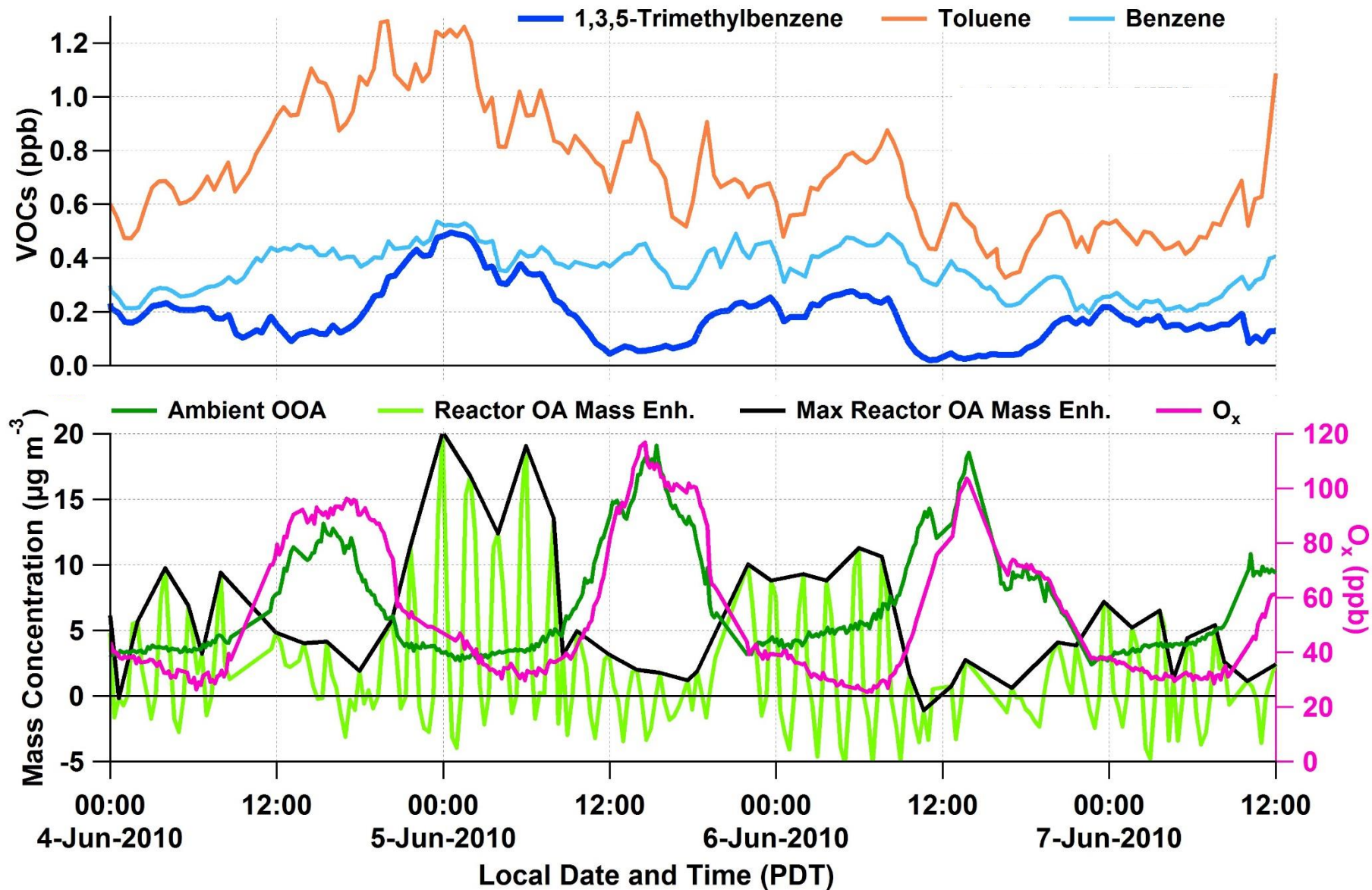
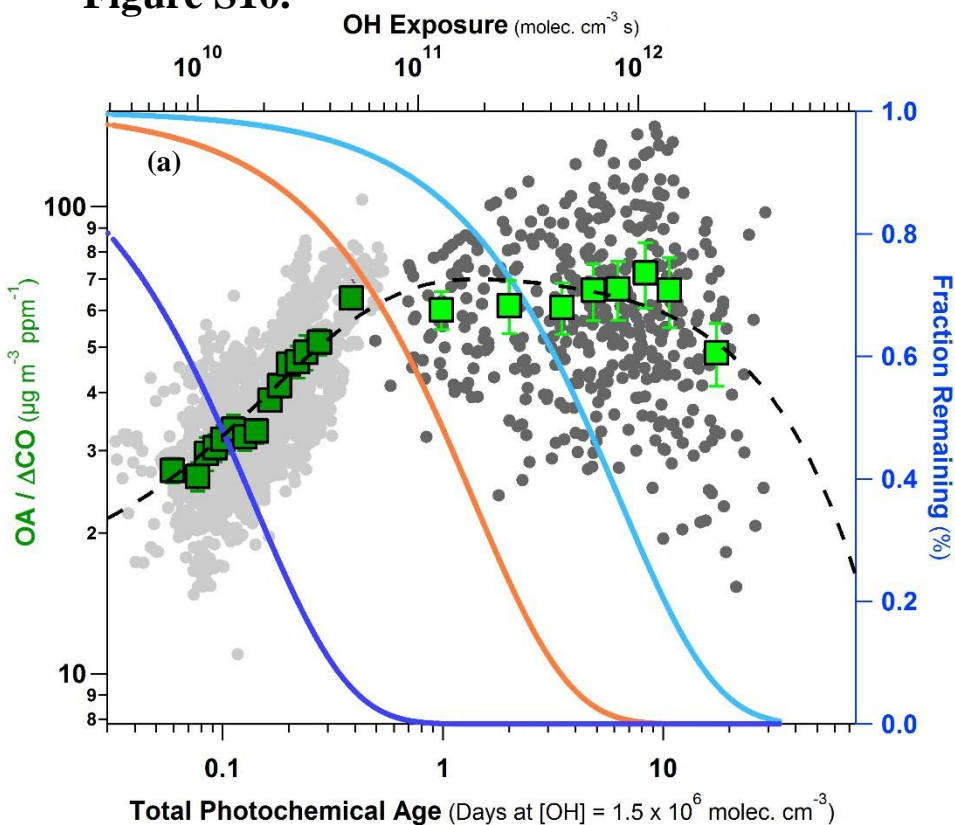
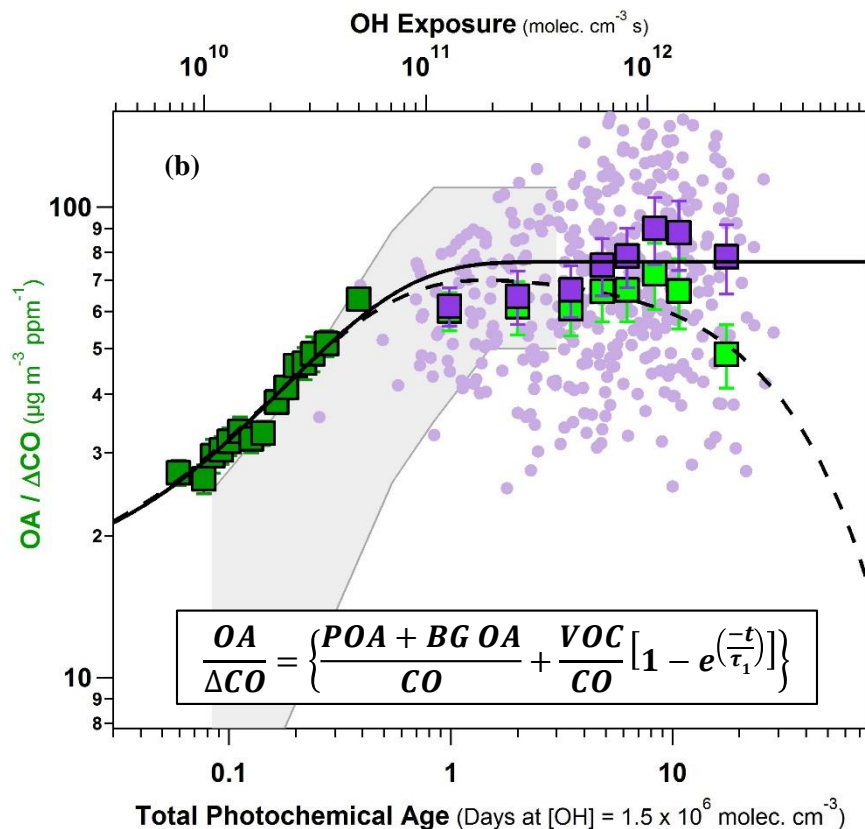


Figure S10.



- Ambient Data (7%) (Hayes et al. 2013)
- Reactor (12%) Vapor Loss Corrected
- All Data
- All Data
- Trimethylbenzene
- Toluene
- Benzene
- - Ambient + Reactor Fit Vapor Loss Corrected



- Ambient Data (7%) (Hayes et al. 2013)
- Reactor (12%) Vapor Loss Corrected
- All CO-reacted Data
- Reactor CO-reacted (12%)
- Ambient + Reactor CO-reacted Fit
- - Ambient + Reactor Fit Vapor Loss Corrected
- Northeast US & Mexico City (DeCarlo et al. 2010)

$$\frac{POA+BG\ SOA}{CO} = 16 \frac{\mu g}{m^3 ppm}$$

$$\frac{VOC^*}{CO} = 60 \pm 5 \frac{\mu g}{m^3 ppm}$$

$$\tau_1 = 0.3 \pm 0.1 \text{ days}$$

Figure S11.

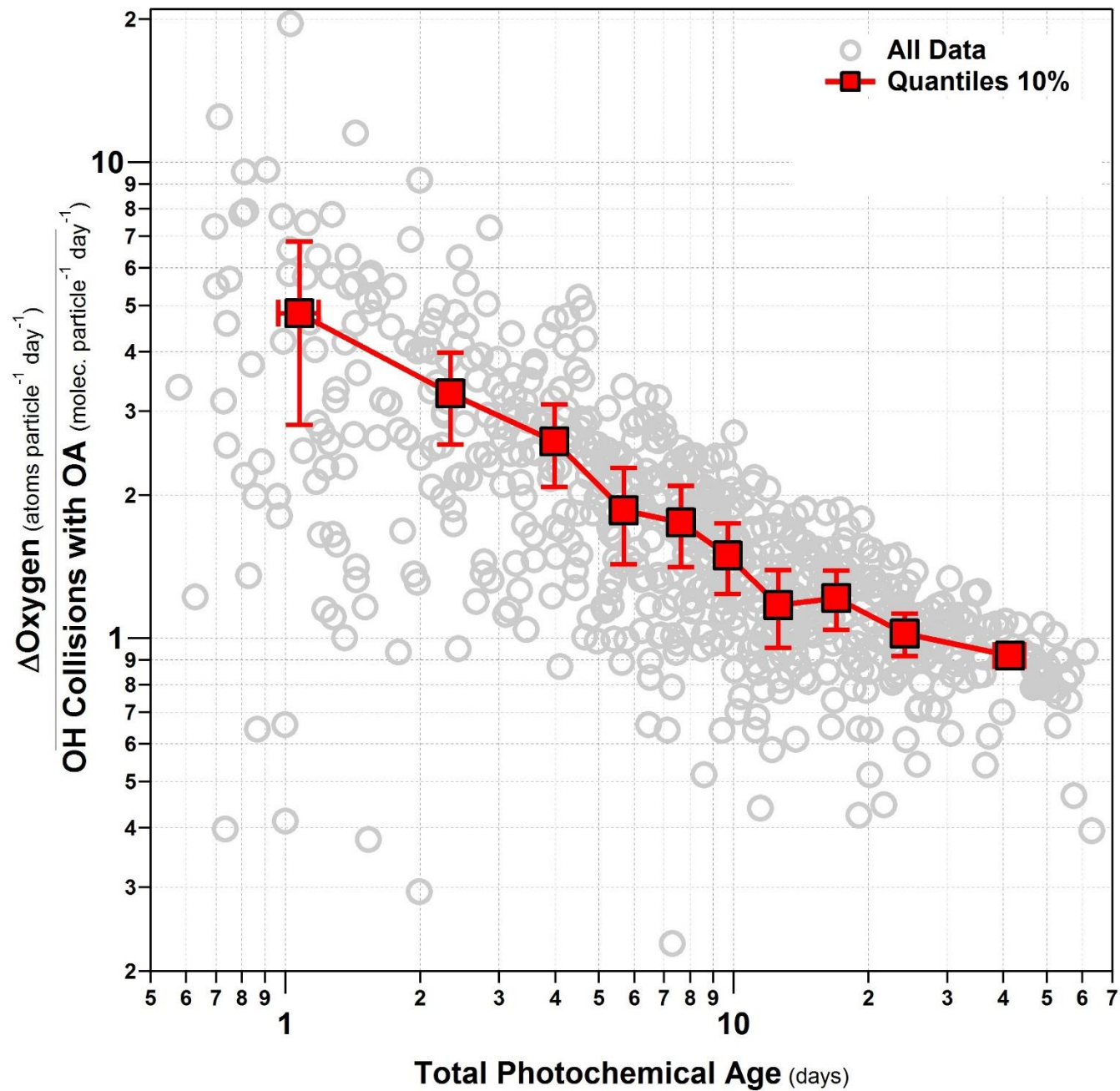


Figure S12.

