



Assessment of fire emission inventories during the South American Biomass Burning Analysis (SAMBBA) experiment

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Abstract. Fires associated with land use and land cover changes release large amounts of aerosols and trace gases into the atmosphere. Although several inventories of biomass burning emissions cover Brazil, there are still considerable uncertainties and differences among them. While most fire emission inventories utilize the parameters of burned area, vegetation fuel load, emission factors, and other parameters to estimate the biomass burned and its associated emissions, several more recent inventories apply an alternative method based on fire radiative power (FRP) observations to estimate the amount of biomass burned and the corresponding emissions of trace gases and aerosols. The Brazilian Biomass Burning Emission Model (3BEM) and the Fire Inventory from NCAR (FINN) are examples of the first, while the Brazilian Biomass Burning Emission Model with FRP assimilation (3BEM_FRP) and the Global Fire Assimilation System (GFAS) are examples of the latter. These four biomass burning emission inventories were used during the South American Biomass Burning Analysis (SAMBBA) field campaign. This paper analyzes and inter-compared them, focusing on eight regions in Brazil and the time period of 1 September–31 October 2012. Aerosol optical thickness (AOT_{550 nm}) derived from measurements made by the Moderate Resolution Imaging Spectroradiometer (MODIS) operating on board the Terra and Aqua satellites is also applied to

assess the inventories' consistency. The daily area-averaged pyrogenic carbon monoxide (CO) emission estimates exhibit significant linear correlations (r , $p > 0.05$ level, Student t test) between 3BEM and FINN and between 3BEM_FRP and GFAS, with values of 0.86 and 0.85, respectively. These results indicate that emission estimates in this region derived via similar methods tend to agree with one other. However, they differ more from the estimates derived via the alternative approach. The evaluation of MODIS AOT_{550 nm} indicates that model simulation driven by 3BEM and FINN typically underestimate the smoke particle loading in the eastern region of Amazon forest, while 3BEM_FRP estimations to the area tend to overestimate fire emissions. The daily regional CO emission fluxes from 3BEM and FINN have linear correlation coefficients of 0.75–0.92, with typically 20–30 % higher emission fluxes in FINN. The daily regional CO emission fluxes from 3BEM_FRP and GFAS show linear correlation coefficients between 0.82 and 0.90, with a particularly strong correlation near the arc of deforestation in the Amazon rainforest. In this region, GFAS has a tendency to present higher CO emissions than 3BEM_FRP, while 3BEM_FRP yields more emissions in the area of soybean expansion east of the Amazon forest. Atmospheric aerosol optical thickness is simulated by using the emission inventories with two operational atmospheric chemistry transport models: the

IFS from Monitoring Atmospheric Composition and Climate (MACC) and the Coupled Aerosol and Tracer Transport model to the Brazilian developments on the Regional Atmospheric Modelling System (CCATT-BRAMS). Evaluation against MODIS observations shows a good representation of the general patterns of the AOT_{550 nm} time series. However, the aerosol emissions from fires with particularly high biomass consumption still lead to an underestimation of the atmospheric aerosol load in both models.

1 Introduction

Biomass burning is a global phenomenon and an ancient practice of human occupation, as well as a natural process. It consumes large amounts of vegetation across wide areas and modifies Earth surface characteristics. Fires still play a key role in ecosystem services, opening areas for livestock, agriculture, and pest control (Shimabukuro et al., 2013). In the last 5 decades, biomass burning has been extensively and persistently used all over the tropics for these purposes and has been involved in widespread deforestation and forest degradation (Crutzen and Andreae, 1990; Bustamante et al., 2016). Biomass burning emissions inject a considerable amount of oxidants and aerosols into the atmosphere, modifying atmospheric composition and reactivity and therefore disturbing the regional climate, water, and biogeochemical cycles (Andreae et al., 2004; Bowman et al., 2009). Moreover, fire emissions in tropical areas are not only felt regionally but also rapidly and strongly affect the global scale due to the efficiency of the atmospheric transport processes of trace gases and aerosols emitted and the high heat release, which reinforced the intense tropical convective activity (Kaufman et al., 1995; Val Martin et al., 2010). Extensive fire activity also disturbs the environmental system, producing soil depletion, damaging flora and fauna, decreasing biodiversity, and even affecting human life (Fearnside, 2000).

An accurate temporal and spatial estimate of biomass burning emissions is critical to a reliable analysis of the associated effects. Nowadays, efforts to quantify emissions from biomass burning from space-borne instruments have increased considerably in scope, but uncertainties in the input data and within the different methodologies can still lead to errors of up to an order of magnitude for trace gases and aerosol emission totals (Vermote et al., 2009; Baldassarre et al., 2015).

Several products of burned area (BA) and fire emission inventories, such as the Global Fire Emissions Database (GFED; van der Werf et al., 2010), the Brazilian Biomass Burning Emission Model (3BEM; Longo et al., 2010), the Global Fire Assimilation System (GFAS; Kaiser et al., 2009, 2012), and the Fire INventory from NCAR (FINN; Wiedinmyer et al., 2011) have been published. Most of the fire emission inventories utilize active fire locations and burned area

to estimate the trace gases and aerosol emissions released into the atmosphere (Mao et al., 2014). However, the temporal and spatial distribution of biomass burning emissions is affected by several sources of errors related, e.g., to the lack of detection of small fires during prescribed and agricultural burning. Also, global burned area products are unsuitable to estimate the burned area of small fires due to the limitations of their algorithms (Giglio et al., 2006, 2010; Randerson et al., 2012).

Most methodologies to estimate biomass burning emissions utilize the relationship proposed by Seiler and Crutzen (1980), which we will call the “BA-based approach”:

$$M^{[\epsilon]} = AB\beta EF^{[\epsilon]}, \quad (1)$$

where $M^{[\epsilon]}$ is the emission load of species ϵ (g), A is the burned area (in km²), B is the fuel load (kg km⁻²), β is the combustion completeness (unitless), and $EF^{[\epsilon]}$ is the emission factor released species ϵ (g kg⁻¹). In this method, the burned area is one of the parameters necessary to estimate the emission of trace gases and aerosols, usually estimated by Earth Observation (EO) satellites; typically, since fires must have already occurred to produce a burned area, this is not that well suited to providing near-real-time air quality forecasts in models that include vegetation fires. Other factors in Eq. (1) can also be difficult to determine; for example, the combustion completeness typically depends on the moisture present in the biomass (and thus in part on micrometeorology), and the fuel load (average biomass and its fraction above ground) is hardly homogeneous (Chuvieco et al., 2004; Yebra et al., 2009; De Santis et al., 2010).

In the last decade, the development of sensors more targeted at active fire observations has allowed the estimation of the radiative energy flux released by fires or the fire radiative power (FRP, in J s⁻¹). The FRP accuracy has been evaluated by Kaufman et al. (1998) and Wooster et al. (2003), showing an average error of 16 and 6.5 %, respectively. However, the error could be largely due to the spatial resolution of several sensors; basically, the atmospheric transmittance and the cloud obscuration can cause an omission error of 15 and 11 % in FRP estimates, respectively (Schroeder et al., 2008). Also, according to Vermote et al. (2009), the integration of FRP over the fire cycle and its conversion to burned biomass can cause an error of 21 and 10–30 %, respectively, varying according to heterogeneity of regional/zonal characteristics.

Continuous acquisition of FRP over a fire’s lifetime provides the fire radiative energy (FRE, in MJ) released by the fire process. New fire emission methods utilize the FRE to calculate the amount of biomass burned and/or the emission of trace gases and aerosols (Wooster et al., 2005; Ichoku and Kaufman, 2005; Ellicot et al., 2009; Freeborn et al., 2011; Kumar et al., 2011):

$$\text{FRE}_{\text{grid}(\text{long},\text{lat})} = \frac{1}{2} \sum_{i=1}^n (\text{FRP}_n + \text{FRP}_{n+1}) \cdot (T_{n+1} - T_n), \quad (2)$$

$$M^{[\epsilon]} = \text{FRE}_{\text{grid}(\text{long},\text{lat})} \gamma \text{EF}^{[\epsilon]}, \quad (3)$$

where $\text{FRE}_{\text{grid}(\text{lon},\text{lat})}$ provides radiative energy at a geographical location in terms of longitude and latitude of a specific centered point of a regular grid; T is the time sequence of FRP acquisitions; n represents the n th sample; and γ is the FRP-biomass conversion factor (kg MJ^{-1} ; Wooster et al., 2005; Kaiser et al., 2012). In this method, parameters such as fuel load, burning efficiency, and the presence of moisture in the soil and in vegetation directly influence the observed energy radiated by the fires and do not have to be separately considered.

The South American Biomass Burning Analysis (SAMBBA) was an airborne experiment design to characterize the smoke physical and chemical properties in Amazonian rainforest and central area of Brazil. The SAMBBA campaign took place in September 2012. The operational smoke forecasting system built to support SAMBBA flight planning utilized four fire emission inventories, which deployed the aforementioned approaches: (i) 3BEM (Longo et al., 2010), (ii) Brazilian Biomass Burning Emission Model with FRP assimilation (3BEM_FRP; Pereira et al., 2009), (iii) GFAS (Kaiser et al., 2012), and (iv) FINN (Wiedinmyer et al., 2011). This study provides an intercomparison and evaluation of these inventories, with focus on the SAMBBA field campaign experiment.

2 Data and methodology

2.1 Inventories description

2.1.1 Brazilian biomass burning emission models (3BEM and 3BEM_FRP)

3BEM is a model developed to estimate the daily fire emissions based on the location of actively burning fire “hotspots” (i.e., areas of combustion detected using active fire/thermal anomaly detection algorithms) derived via orbital remote sensing. 3BEM utilizes the Wildfire Automated Biomass Burning Algorithm (WF_ABBA) applied in the Geostationary Operational Environmental Satellite (GOES) constellation data (Prins et al., 1998); the MOD14 and MYD14 products of the Moderate Resolution Imaging Spectroradiometer (MODIS) aboard the Terra and Aqua satellites (Justice et al., 2002); and the fire product developed by the Environmental Satellite Division (DSA) of National Institute for Space Research (INPE), which uses the Spinning Enhanced Visible and Infrared Imager (SEVIRI) on board Meteosat Second Generation, GOES, MODIS, and Advanced Very High Resolution Radiometer (AVHRR) on board the

National Oceanic and Atmospheric Administration (NOAA) constellation (Setzer et al., 1994).

The 3BEM version initially developed by Longo et al. (2010) estimates the emitted mass of trace gases and aerosols related to each fire detected by remote sensing as described in Eq. (1). Therefore, the locations of the detected fire hotspots are cross-tabulated with MODIS Land Cover map to allow the estimation of aboveground biomass density, combustion factor, and emission factor from literature values (Olson et al., 2000; Andreae and Merlet, 2001; Houghton et al., 2001; Sestini et al., 2003; Akagi et al., 2011). Also, the model filters all fires located in a 1 km radius to prevent double counting between fire products.

The updated 3BEM includes FRP assimilation (3BEM_FRP; Pereira et al., 2009) and utilizes the algorithm of the earlier 3BEM version but with burned biomass directly estimated using FRE estimates, as described in Eq. (3). The 3BEM_FRP model groups all FRP values estimated by MODIS, GOES, and METEOSAT products according to their time acquisition, eliminating the low confidence fire pixels (values below 50 % for MOD14, MYD14, and METEOSAT products, and flags 4 and 5 for WFABBA/GOES product) and minimizing the impact of the MODIS bow-tie effect as described in Freeborn et al. (2011). Also, due to the high frequency of observations (mainly in GOES and SEVIRI data) 3BEM needs only one fire detection to estimate the biomass burned and its associated emissions due to the filtering process (Longo et al., 2010). If the active fire has no subsequent observations in the next 4 h, the algorithm assumes that the fire event is over. Thus, missed detections due to cloud cover have an impact on the FRE integration if the cloud persists for more than eight satellite acquisitions.

Some of these FRP estimates are compromised by sensor saturation over larger fires, particularly GOES over South America (Xu et al., 2010). Thus, to not ignore important episodes of biomass burning by removing GOES saturated pixels, for which FRP values are not provided, 3BEM_FRP utilizes Eq. (4) to estimate the energy released by fires, based on the premise that emitted spectral radiance (M_λ) in spectral band centered at $3.9 \mu\text{m}$ is proportional to FRP (Wooster et al., 2003).

$$\text{FRP}_{\text{MIR}} = \frac{A_g}{a} \sigma \int_{3.76}^{4.03} M(\lambda, T) d\lambda - M_b, \quad (4)$$

where A_g represents the area of the GOES pixel (km^2); a is an empirical constant adapted for the GOES MIR spectral band ($\text{W m}^{-2} \text{sr}^{-1} \mu\text{m}^{-1} \text{K}^{-4}$; Wooster et al., 2005); σ is the Stefan–Boltzmann constant ($5.66 \times 10^{-8} \text{W m}^{-2} \text{K}^{-4}$); M is the Planck curve ($\text{W m}^{-2} \mu\text{m}^{-1}$); λ is the wavelength (μm); T represents the temperature (K); and M_b is the radiance emitted by the background (110 MW). FRP values estimated by GOES satellites below 1000 MW are corrected by

+17 % and FRP values higher than 1000 MW are corrected by +41 % (Xu et al., 2010). This procedure is also applied to SEVIRI data but, due to its spatial coverage, we decided not to include these data in the present analysis. Pereira et al. (2009) describe the 3BEM_FRP method in detail.

A clustering process performs the combination of all detected fires from different sensors. In this step, the size of a matrix that merges FRP data can be defined according to the spatial resolution and grid configuration of the atmospheric model. Consequently, the convolution mask $\eta(\gamma, \kappa)$ of size $M \times N$ (rows \times columns), running over the grid with FRP areal density (FRP_{ad} , in W m^{-2} ; defined by weighting the FRP values by pixel area) values estimated by different satellites $\xi(\text{long}, \text{lat})$, will result in the grid (FRP_{grid}) containing all clustered fires for a given time step.

$$\text{FRP}_{\text{grid}(\text{lon}, \text{lat}, t)} = \sum_{\gamma=-\alpha}^{\alpha} \sum_{\kappa=-\beta}^{\beta} \eta(\gamma, \kappa) \xi(\text{long}+\gamma, \text{lat}+\kappa, t), \quad (5)$$

where the clustered grid is defined to all points where the mask of $M \times N$ size overlaps the image completely ($\text{lon} \in [\alpha, M-\alpha]$, $\text{lat} \in [\beta, N-\beta]$). Moreover, if the interval between two acquisitions is greater than 4 h ($\Delta T > 14\,400$ s) at any time step of the FRP integration, the observations are assumed to originate from two or more independent fires.

2.1.2 GFAS

The GFAS is an approach used to map daily global fire emission through FRP observations. Therefore, GFAS also assumes that the electromagnetic radiation emitted by fires is related to the consumption of burned biomass (Wooster et al., 2005). In the GFAS version (1.1) used in this study, FRP values of MOD14 and MYD14 fire products from Terra and Aqua satellites, respectively, are used to estimate the average of observed FRP areal density. GFAS estimates open vegetation fire trace gas and particulate emissions from each fire detected as described in Eq. (3).

The model performs a clustering process of observed FRP (F_i), pixel area (A_i), and view zenith angle (θ) to sensor pixels with valid observations (i) to estimate the FRP_{ad} . The clustering process also takes observations of $F_i = 0$, i.e., no-fire, into account. Thus, for each individual grid cell the estimated FRP_{ad} could be calculated as

$$\text{FRP}_{\text{ad}} = \frac{\sum_k \sum_{i_k \in j} F_{i_k} \cdot \cos^2(\theta_{i_k})}{\sum_k \sum_{i_k \in j} A_{i_k} \cdot \cos^2(\theta_{i_k})},$$

where i_k represents the pixel i of satellite product k (MOD14, MYD14). This formulation implicitly corrects the MODIS bow-tie effect and partial cloud/ice/snow/water cover of a grid cell. Observation gaps are subsequently filled with a data assimilation approach that currently assumes persistence of the fires.

In GFAS, the coefficient that converts the FRP_{ad} to dry matter combustion rate is based on eight land cover classes

Table 1. Land Use and Land Cover (LULC) classes based on Global Land Cover Characterization (GLCC).

1	Crop/mixed farming	11	Semi-desert
2	Short grassland	12	Ice cap/glacier
3	Evergreen needleleaf tree	13	Bog or marsh
4	Deciduous needleleaf tree	14	Inland water
5	Deciduous broadleaf tree	15	Ocean
6	Evergreen broadleaf tree	16	Evergreen shrub
7	Tall grass	17	Deciduous shrub
8	Desert	18	Mixed woodland
9	Tundra	19	Forest/field mosaic
10	Irrigated crop	20	Water and land mixture

(cf. Heil et al., 2010, Kaiser et al., 2012). In addition, the emission load of 40 species is calculated using emission factors from Andreae and Merlet (2001), with subsequent updates from Christian et al. (2003). Kaiser et al. (2012) describe the inventories method in detail.

2.1.3 FINN

FINN is an approach to estimate daily fire emissions at 1 km resolution through satellite observation of active fires. The FINN model produces global estimates of aerosols and trace gases of open vegetation fires, as described in Eq. (1). The FINN utilizes the MOD14 and MYD14 fire products, processed by MODIS Rapid Response (MRR) or by the MODIS Data Processing System (MODAPS) Collection 5. However, since MODIS observations do not cover the entire globe daily, due to orbital gaps, FINN smears MODIS detections of active fire over 2 days. For each fire located in the equatorial region, it is assumed that the next day the fire will be half of its original size (Wiedinmyer et al., 2011).

Similarly to 3BEM, the FINN model removes multiple detections of same fire pixel prior to estimation of the trace gases and aerosols released. In addition, for each active fire, FINN estimates as 1 km^2 the burned area for most of land use classes, with the exception of grasslands/savannas in which the burned area is assumed to be 0.75 km^2 . To estimate the emissions of trace gas and aerosol species, emission factors derived from Andreae and Merlet (2001) and Akagi et al. (2011) are used, as well as the MODIS Land Cover Type for 2005 and the MODIS Vegetation Continuous Fields (VCF) product. Wiedinmyer et al. (2011) describe the procedures adopted in detail.

2.2 MODIS aerosol optical thickness (AOT)

The aerosol optical depth (AOD) product from MODIS sensors aboard the Aqua and Terra satellites is used to provide a first order assessment of the impacts of the two distinct methods to estimate biomass burning emission: BA-based and FRP-based. In this work, the MODIS Level 2.0 Collection 5.1 (051) data and Level 3 atmospheric prod-

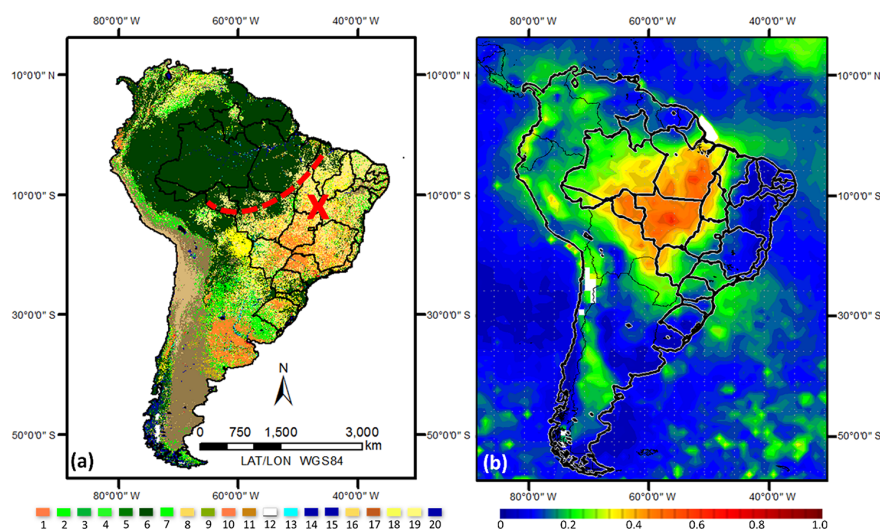


Figure 1. (a) MCD12 Land Use and Land Cover (LULC) for South America with the dotted red line representing the Amazonian Arc of Deforestation and soybean expansion in Brazilian Cerrado (marked as X); (b) time-averaged Aqua MODIS aerosol optical thickness at 550 nm during 1 September–31 October 2012 over South America.

uct denominated MYD08_D3 (mean aerosol optical thickness at 550 nm, i.e., AOT_{550nm}) are utilized to compare the fire emission inventories used during the SAMBBA campaign between 1 September and 31 October 2012. The MODIS daily global product merges all MODIS acquisitions over the globe on an equal-angle grid with 1° resolution (Kaufman and Tanré, 1998). Figure 1a shows the MODIS MCD12 Land Use and Land Cover (LULC) product for South America with LULC classes described in Table 1. During the SAMBBA field campaign, the highest biomass burning aerosol loadings were observed over Brazilian territory, mainly over the southeastern edge of the Amazonian forest (dotted red line, known as Arc of Deforestation) and in the soybean expansion area in the Brazilian savannah (marked as X), as shown in the time-averaged (1 September–31 October 2012) AOD_{550nm} field, derived from MODIS sensor aboard Aqua satellite (Fig. 1b). The high values of AOD related to fires located in eastern Mato Grosso (mainly in the secondary forest) and in transition areas of Amazon rainforest and Brazilian savannah are noteworthy; soybean expansion areas, which present high concentration of fires, have a lower amount of biomass than Amazon rainforest and the strong ventilation favors the transport of smoke to west.

2.3 Inventories configuration and analysis description

To evaluate the fire emission inventories utilized during the SAMBBA experiment, we used the 3BEM preprocessor to generate the gridded data in geographical coordinates with a spatial resolution of 0.1° . The 3BEM preprocessor generates daily emission fluxes of several species, but in this work we selected only carbon monoxide (CO) emission fields from 1 September to 31 October 2012 to analyze. The four in-

ventories are compared over eight sub-domains windows with sizes of $10^\circ \times 10^\circ$ that typically cover different Brazilian states (Fig. 2a) and different fire regimes and biomes (Fig. 2b).

Two models were selected for real-time applications during SAMBBA experiment: the Monitoring Atmospheric Composition and Climate (MACC) from European Centre for Medium-Range Weather Forecasts (ECMWF) with GFAS fire emission inventory (described in Sect. 2.1.2) and the Coupled Chemistry Aerosol and Tracer Transport model to the Brazilian developments on the Regional Atmospheric Modelling System (CCATT-BRAMS) with 3BEM fire emission inventory (described in Sect. 2.1.1). These models were selected due to their distinct methodologies for estimating the biomass burning emissions (the first a FRP-based approach, and the second the BA-based approach). The MACC/ECMWF model simulation had a global domain, about 0.8° (T255) horizontal resolution, 60 levels vertical levels, and a forecast frequency of 6 h. It assimilates MODIS AOD observations and the full set of meteorological observations that are used for the operational weather forecasts of ECMWF. The CCATT-BRAMS simulations were performed for the South America domain, 0.22° of horizontal resolution, 24 levels of vertical resolution, forecast operation mode, forecast frequency of 3 h, and boundary conditions from Center for Weather Forecasting and Climate Research (CPTEC) model.

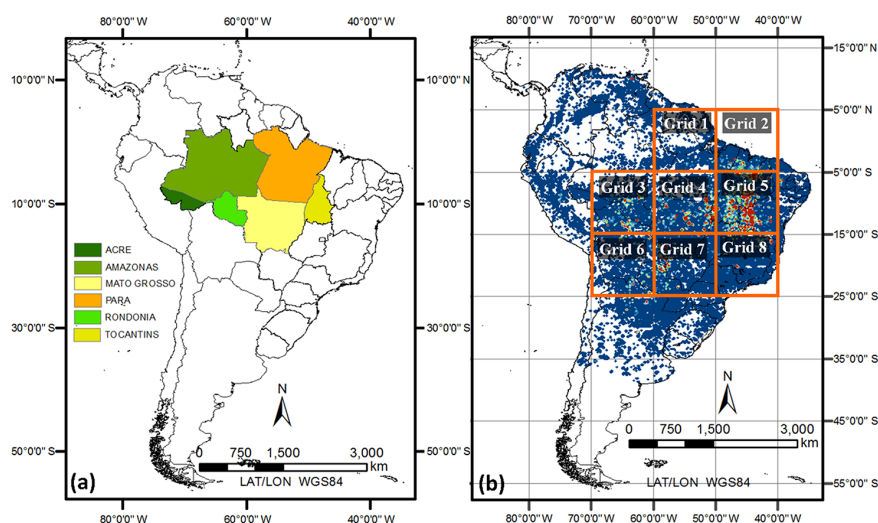


Figure 2. South America maps highlighting (a) the Brazilian states of Acre, Amazonas, Mato Grosso, Para, Rondonia, and Tocantins, where SAMBBA field campaign took place; and (b) the eight sub-domains windows for the biomass burning inventories comparison.

3 Results and discussion

3.1 Comparing emission inventories spatially

During the SAMBBA experiment, four inventories were used in near real time within atmospheric chemistry transport models to support the SAMBBA flight planning (FINN, GFAS, 3BEM, 3BEM_FRP). Each inventory's spatial distribution of total CO emission (10^4 kg m^{-2}) over South America, from 1 September to 31 October 2012, is depicted in Fig. 3, according to the respective methodologies used to estimate the emission loading.

The intercomparison shown in Fig. 3 demonstrates that inventories that utilize the same (i.e., FRP-based or BA-based) methodology show similar spatial patterns in CO emissions (kg m^{-2}), not only in the Amazon basin but across all of South America, although they differ in their absolute values. The CO emissions estimated by the 3BEM and FINN emission inventories (Fig. 3a and b, respectively) present higher values than 3BEM_FRP (Fig. 3c) and GFAS (Fig. 3d) in the regions where the main processes of forest logging and subsequent agricultural expansion (grids 3, 4, and 5) occur. The highest emissions in 3BEM and FINN are located mainly in Rondonia State (grid 3) and in Mato Grosso State (grid 4), where most of the SAMBBA flights took place.

Cardozo et al. (2014) analyzed the fires pattern in Rondonia between 2000 and 2011 and identified that most fires result in relatively small “burn scars” on the landscape, with areas of 20 to 80 ha (64 % of cases). In addition, only 6.5 % of all burned areas in Rondonia are associated with recently deforested areas. This could indicate that 3BEM and FINN are overestimating the CO emission load due to an erroneous assumption that observed fires occur in forest instead of livestock and permanent crops, which have lower aboveground

biomass. Furthermore, 3BEM_FRP and GFAS inventories do not display a similar pattern to 3BEM and FINN because their emissions are directly related to FRP, with a weaker dependency on land cover type. Thus, areas with low above-ground biomass will provide low values of FRE due the observed characteristics of the fire activity and thus low values of CO.

The spatial distribution of the emission inventories suggests that fires in the region are strongly related to deforestation activity and therefore to the general economy, with a strong trend in recent years of fires in secondary forests (Cardozo et al., 2014). In Fig. 3c, higher emission loads are located in the east of Brazil (grid 5), mainly in the Cerrado biome, a vegetal formation composed of savanna and other typically low-density vegetation formations, which include trees of 15 m height (as shown in Fig. 1a). This region is now economically used, constituting a new agricultural frontier of Brazil (with more than 100 million hectares suitable for modern mechanized crop agriculture, mainly soybean). In this area, the four fire inventories differ considerably. The inventories that use the FRP approach show much higher emission loads compared to 3BEM and FINN.

The difference in South American daily area-averaged emission of CO (kg m^{-2}) between 1 September and 31 October 2012 was quantified via linear correlation coefficient analysis (all significant at $p < 0.05$) (Fig. 4). The highest linear correlation coefficients were found between 3BEM and FINN with 0.86, between 3BEM_FRP and GFAS with 0.85, and between GFAS and FINN with 0.84. These high linear correlations indicate that inventories produced using similar emission methods tend to agree with each other. The third correlation reflects that both inventories use the same active fire observations (MODIS), albeit with different data processing. We highlight that all were significant at $p > 0.05$

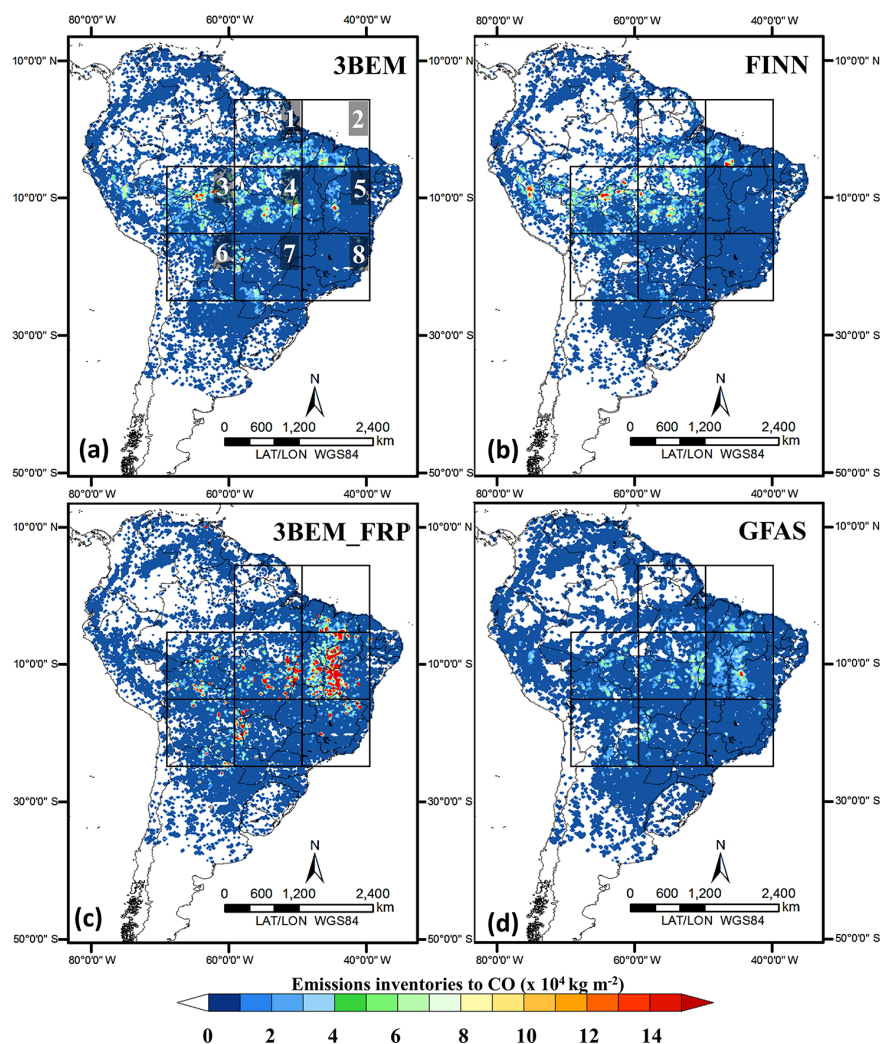


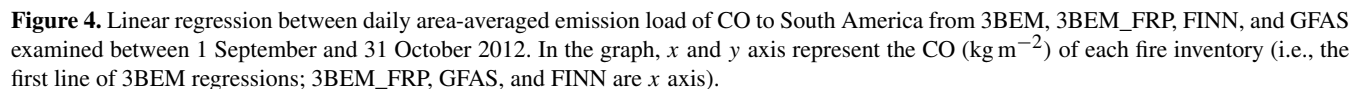
Figure 3. Spatial distribution of total CO emissions (10^4 kg m^{-2}) over South America from 1 September to 31 October 2012, estimated by (a) 3BEM, (b) FINN, (c) 3BEM_FRP, and (d) GFAS.

level, using the Student t test. To analyze the measures of accuracy of regression between fire inventories we used the bootstrap technique (Efron, 1982). In this technique, a population of 1.0×10^4 reconstructs the regression and provides the parameters to create the confidence interval and error analysis of model estimation.

The bootstrap regression among the daily area-averaged CO values to 3BEM and FINN emission inventories present a linear correlation between 0.75 and 0.92 with a tendency of FINN to overestimate, relative to 3BEM, the emission load of CO in 20–30 % (Fig. 4). This apparent FINN overestimation is seen in the majority of grids, with the exception of grids 2, 5, and 7 (Fig. 5a). In areas where FINN emissions are lower than 3BEM emissions, vegetation is mainly composed of Brazilian Cerrado (grids 2 and 5), while the Brazilian Pantanal wetland biome (grid 7) is predominantly composed of savannah and grassland, with the additional presence of de-

ciduous and semi-deciduous forest. Since 3BEM and FINN implement similar methodologies to estimate the emission load of species released by wildfires, the parameters used in Eq. (1), such as aboveground biomass, are likely to be associated with the relatively high estimation in CO values.

The intercomparison analysis of the daily area-averaged CO values to 3BEM and GFAS inventories shows a linear correlation coefficient between 0.75 and 0.85 with higher values over the Arc of Deforestation in the Amazon rainforest (Fig. 1a). However, GFAS has a tendency to present higher CO emissions (by 10–20 %) in grids 5 and 8 and to underestimate by 20–30 % in grids 3 and 4 (Fig. 5b) when compared to 3BEM. In these areas, the vegetation is dominated by the Amazon biome, along with a small area of Cerrado (located in the south of the grids), with presence of evergreen broadleaf forest, tropical degraded forest, and cropland/livestock areas. Moreover, the daily area-averaged CO



Similar to GFAS, the relationship among the daily area-averaged CO values to 3BEM and 3BEM_FRP shows a linear correlation coefficient between 0.65 and 0.75. However, 3BEM_FRP has a tendency to overestimate the CO values by 60–85 % in grid 5 and to underestimate by 10–15 % in grids 3 and 4 (Fig. 5c). The 3BEM_FRP model presented an elevated emission of CO in grid 5, possibly due the GOES viewing zenith angle (high viewing angles results in erroneous values of infrared brightness and present a large pixel area.; Roberts et al., 2005; Vermote et al., 2009; Peterson et al., 2012). In addition, in grid 5, cloud absence may influence in the FRP cycle, leading to an overestimation due the large number of acquisitions in high view angles. The relationship between the daily area-averaged CO values to 3BEM_FRP and GFAS shows a linear correlation between 0.82 and 0.90. Accordingly, the 3BEM_FRP model has a tendency to present higher emission fluxes of CO than GFAS,

The fieldwork measurements acquired during the SAMBBA campaign indicate that the near-real-time air smoke forecasts based on BA-based inventories, such as 3BEM used by CCATT-BRAMS, typically underestimated smoke loading in the central to eastern region of the Amazon forest (near Mato Grosso and Tocantins States), as described in Rosário et al. (2013). In addition, the emission inventories tended to overestimate the smoke loading in the northwestern part of Rondonia (indicated in Fig. 2a). However, during the SAMBBA campaign lower values of CO emissions were identified, showing an overestimation of emission flux estimates in this region. In general, all fire emission inventories present a good agreement, with most regressions significant at $p > 0.05$ level, according to the Student t test (Table 2, non-significant regression marked in

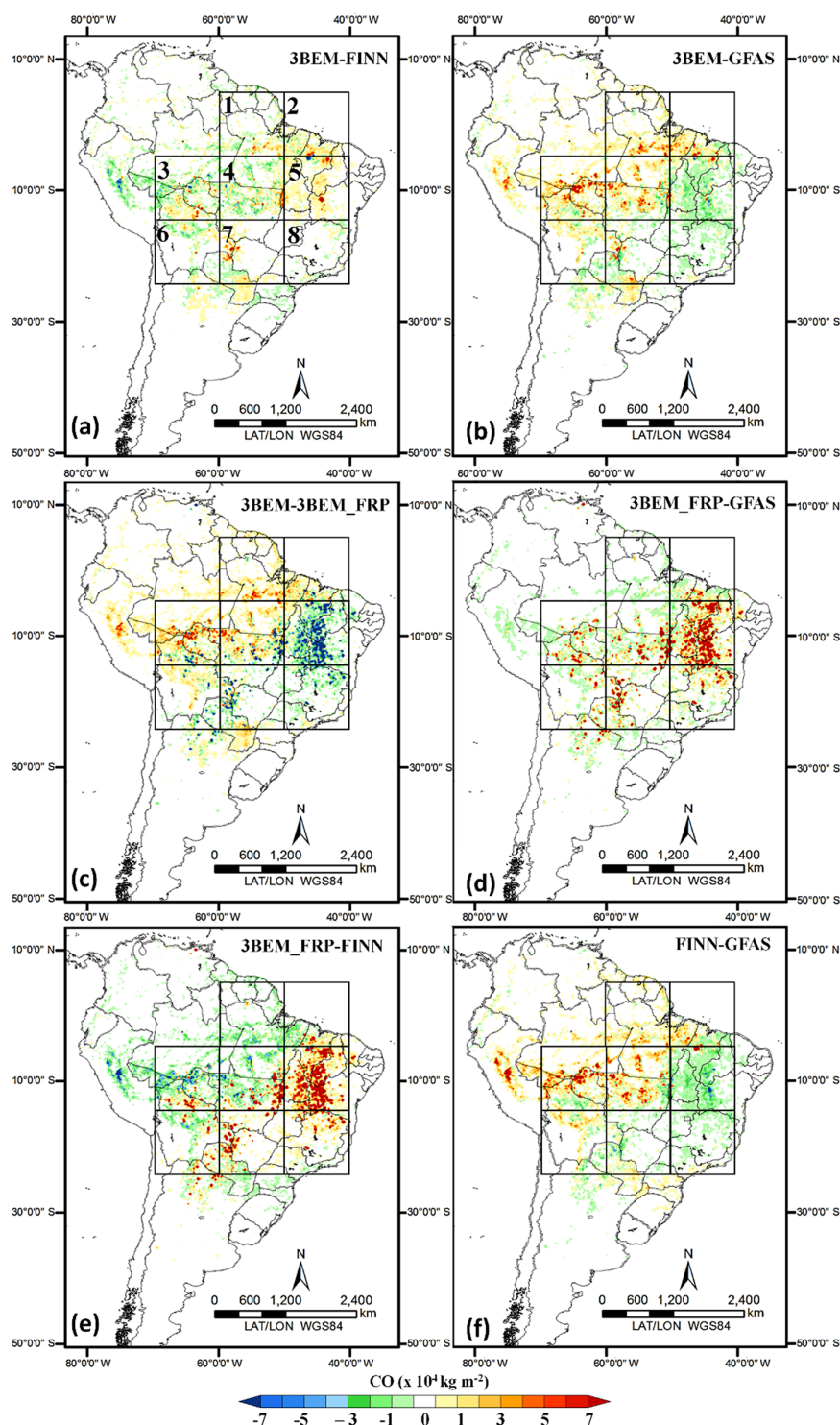


Figure 5. Difference between fire inventories. (a) 3BEM-FINN; (b) 3BEM-GFAS; (c) 3BEM-3BEM_FRP; (d) 3BEM_FRP-GFAS; (e) 3BEM_FRP-FINN; (f) FINN-GFAS.

red). The only regression that presented low values of linear correlation coefficients are located in grid 5 with values lower than 0.30. Moreover, regression between 3BEM and FINN presents the highest correlation in most of the grids.

Also, the absolute bias analysis indicate a high variability in daily area-averaged emission load of CO between the four fire inventories.

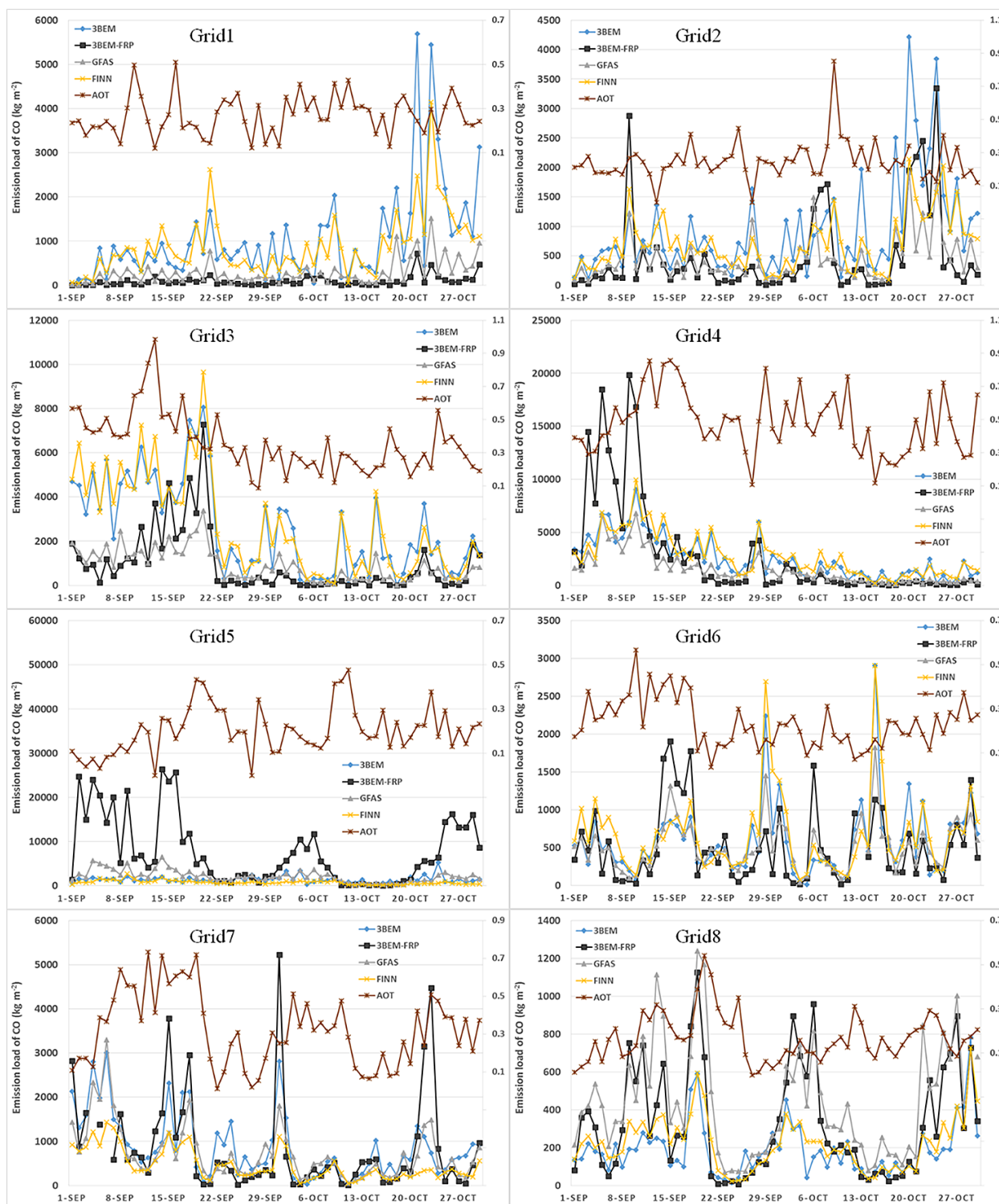


Figure 6. Lat–long average time series of Aqua MODIS AOT_{550nm} and 3BEM, 3BEM_FRP, GFAS, and FINN emission load of CO time series.

Table 2. Linear correlation coefficients (r) between 3BEM, 3BEM_FRP, FINN, and GFAS daily area-averaged emission inventories to the eight grids and average absolute bias of CO (in kg m^{-2}). In the table, the first value indicates the r and the second represents the absolute bias (r /bias). Bold value are not significant ($p > 0.05$ level, Student t test).

	3BEM_FRP	FINN	GFAS
3BEM_G1	0.84/933	0.85/145	0.82/736
3BEM_FRP_G1		0.67/−788	0.76/−197
FINN_G1			0.82/591
3BEM_G2	0.70/436	0.75/458	0.79/210
3BEM_FRP_G2		0.61/22	0.78/−226
FINN_G2			0.82/−249
3BEM_G3	0.78/1642	0.94/1674	0.90/−136
3BEM_FRP_G3		0.77/32	0.87/−1778
FINN_G3			0.91/−1811
3BEM_G4	0.81/−245	0.93/996	0.90/−263
3BEM_FRP_G4		0.71/1241	0.90/−18
FINN_G4			0.88/−1260
3BEM_G5	0.16/−6345	0.30/−793	0.36/605
3BEM_FRP_G5		0.47/5552	0.84/6950
FINN_G5			0.80/1399
3BEM_G6	0.46/12	0.89/68	0.88/−86
3BEM_FRP_G6		0.32/55	0.65/−99
FINN_G6			0.81/−154
3BEM_G7	0.68/−14	0.89/151	0.85/413
3BEM_FRP_G7		0.60/165	0.79/427
FINN_G7			0.85/261
3BEM_G8	0.63/−138	0.84/−251	0.79/−31
3BEM_FRP_G8		0.65/−113	0.65/106
FINN_G8			0.91/219

3.2 Assessment of fire inventories with AOT

Figure 6 shows the daily emission estimates from each of the emission inventories used in SAMBBA campaign integrated over the eight grids and the area-averaged values of AOT_{550nm} for 1 September to 31 October 2012. During the SAMBBA campaign, the grids located in the Arc of Deforestation and in Mato Grosso state exhibited the highest values of averaged AOT_{550nm}. Generally, the inventories that use FRP to estimate the emission of CO produce similar patterns. Likewise, the emission inventories that utilize the relationship between the burned area, fuel load, and the combustion completeness lead to comparable patterns. This is evident in Fig. 6a, in which 3BEM_FRP and GFAS have lower values during all periods, with daily emission loads of CO between 20 and 500 kg m^{-2} and an average value of 93 kg m^{-2} . In addition, the 3BEM and FINN emission inventories presented a high variability during the analyzed period, with a daily area-averaged CO emission flux that reaches values greater than 3000 kg m^{-2} . The linear correlation coefficient of 3BEM_FRP and GFAS and of 3BEM and FINN are 0.85 and 0.76, respectively, with a tendency of GFAS

Table 3. Total mean emissions (kg m^{-2}) of CO in the eight grids shown in Fig. 2.

	Total emission of CO (kg m^{-2})			
	3BEM	FINN	GFAS	3BEM_FRP
Grid 1	61.660	52.971	17.485	5.636
Grid 2	55.234	42.631	27.706	29.059
Grid 3	155.346	163.552	54.891	56.815
Grid 4	151.720	167.518	91.905	166.423
Grid 5	82.271	45.945	129.866	274.896
Grid 6	35.199	40.412	31.118	29.811
Grid 7	53.072	28.266	43.981	48.541
Grid 8	11.342	13.252	25.674	19.633
Total	605.844	554.547	422.626	630.814

and 3BEM to present higher CO emission fluxes. In grid 1, the total emission fluxes of CO in 3BEM and FINN from 1 September to 31 October 2012 are, respectively, 61.660 and 52.971 kg m^{-2} (Table 3), an estimation that is 4–10 times more than GFAS and 3BEM_FRP.

Figure 6b shows the area-averaged time series of CO and AOT_{550nm} in grid 2, with the four emission inventories showing similar patterns with high emission values on 9 September and between 20 and 30 October 2012. The linear correlation coefficients of 3BEM_FRP and GFAS and of 3BEM and FINN are 0.78 and 0.75 with a tendency of 3BEM (55.234 kg m^{-2}) and 3BEM_FRP (29.059 kg m^{-2}) to present higher CO emission fluxes in this region. Also, in the general pattern of temporal evolution, we could observe a good agreement between CO emission load estimated by the four emission inventories and MODIS AOT_{550nm}. However, the four emission inventories exhibit distinct CO flux distributions in grid 3 (Fig. 6c). In September, FINN and 3BEM presented higher values than 3BEM_FRP and GFAS during September and October. However, 3BEM_FRP shows higher values than GFAS in September and smaller values in October. The linear correlation coefficients of 3BEM_FRP and GFAS and of 3BEM and FINN are 0.87 and 0.94, respectively, with a tendency of FINN and 3BEM_FRP to yield higher estimates of the CO emission flux. The relationship between GFAS and FINN presents a linear correlation coefficient of 0.91. However, the total emission load of CO estimated by FINN (163.552 kg m^{-2}) is 3 times more than GFAS (54.891 kg m^{-2}).

In grid 4, all inventories showed a good agreement with linear correlations coefficients greater than 0.90. However, during the period from 3 to 10 September, 3BEM_FRP presents six episodes with higher emission loads, which remain undetected in the other inventories (Fig. 6d). The outlier values in 3BEM_FRP are likely to be related to GOES FRP acquisitions, which suggests inconsistencies in GOES data acquisition due to viewing angle and errors in data process-

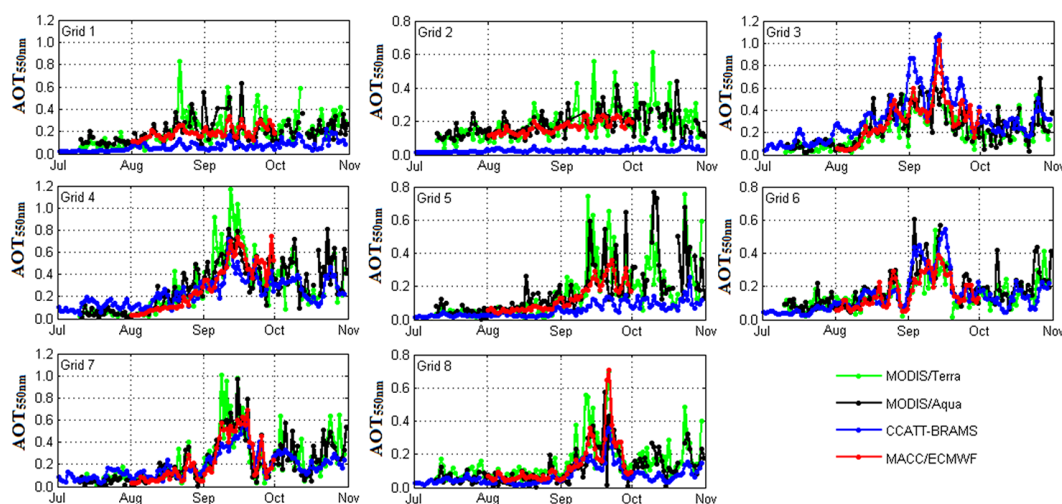


Figure 7. Lat-long average time series of Terra/MODIS AOT_{550nm} (in green), Aqua/MODIS AOT_{550nm} (in black), CCATT-BRAMS simulated AOT_{550nm} (in blue), and MACC/ECMWF (in red) AOT_{550nm} analysis. Grids are those shown in Fig. 2.

ing. Grid 5 shows the lowest linear correlation coefficient between the 3BEM and FINN emission inventories during the SAMBBA campaign with 0.30. In this grid, FINN presents a low estimation of the CO emission flux when compared with the other emission inventories (as demonstrate in Table 3). Here, FRP-based methods showed larger emissions than the BA-based ones. The 3BEM_FRP model overestimated the emission values by more than 100 % compared to GFAS (Fig. 6e) However, the linear correlation between these emission inventories is 0.84. In the other grids, the emissions of CO provided by the four emission inventories show a considerable consistency, with few differences in absolute values, indicating a good agreement across these other regions of South America (Fig. 6f–h).

3.3 Evaluating the emissions against observations

During the SAMBBA campaign, operational atmospheric chemistry transport models were applied in real time: (i) the MACC/ECMWF with GFAS and (ii) CCATT-BRAMS with 3BEM. Figure 7 shows the simulated AOT_{550nm} from MACC/ECMWF (red line), CCATT-BRAMS (blue line), and estimated by Terra (green line) and Aqua (black line) to the eight grid cells. Although 3BEM presents the highest emission flux in grid 1 and grid 2, the CCATT-BRAMS simulation underestimate the MODIS AOT_{550nm}. In these grids, the MACC/ECMWF model presents a better consistency with the satellite measurements. One potential reason for the underestimation of CCATT-BRAMS is related to background aerosol unrelated to the fires directly being observed and that is not included in this model, causing low values of simulated AOT_{550nm} for these grids. Also, MACC/ECMWF assimilates MODIS AOD observations and, even with low emission values in grids 1 and 2, the atmospheric aerosol load is strongly constrained by the MODIS observations.

The assessment of the MACC/ECMWF and CCATT-BRAMS models against MODIS AOT_{550nm} reveals a good agreement in terms of the general pattern of the temporal evolution. As shown in the AOT_{550nm} time series, both methods appear to be able to estimate the influence of aerosols released by fires rather well. However, when the intensity of the biomass burning is too high, the values simulated by MACC/ECMWF and CCATT-BRAMS for the grids 4–8 appear underestimated, possibly due to the influence of smoke on the FRP measurements, lack of fire observations, clouds, aboveground biomass data, or fire size. Moreover, it is possible to identify an overestimation in grid 3, mainly due to an out-of-date land use and land cover map that inserts fires over forest areas even if the area had been burned/deforested during earlier years, and a very noticeable underestimation in grid 5, as demonstrated in the inventories.

4 Conclusions

Full characterization of the emissions of trace gases and aerosols is often essential for assessing the atmospheric impacts of fire and for constructing fire inventories. These inventories generally rely on data from environmental satellites and provide useful information for weather, climate, and air quality models. In this study, we analyzed data from the four biomass burning emission inventories used during the SAMBBA airborne atmospheric sampling and remote sensing campaign that took place in Rondonia (Brazil) between 1 September and 31 October 2012. Each inventory utilizes distinct methodology, with 3BEM and FINN deriving the emission of trace gases and aerosols through a combination of BA and fuel load metrics based on vegetation maps and field-location-specific coefficients, whilst 3BEM_FRP

and GFAS estimate biomass burned more directly from FRP measurements made by the EO satellite instruments.

The evaluation of the emission inventories focused on eight pre-defined grid cells (Fig. 2) and demonstrates that inventories that utilize the same methodology, such as 3BEM and FINN on the one hand (BA-based) and GFAS and 3BEM_FRP on the other (FRP-based), show similar patterns in emission spatial distribution, not only in the Arc of Deforestation but also throughout South America. However, they can differ in their absolute values. As such, each inventory has particular characteristics, with 3BEM and FINN showing more emissions of CO in the Amazon forest logging area, where most of the SAMBBA campaign flights occurred. Furthermore, these emission inventories typically underestimated the smoke loading in the southeastern region of Amazon forest and in the northwestern region of Rondonia, where 3BEM_FRP and GFAS present larger emission fluxes of CO. The best overall linear correlations coefficients were found between 3BEM and FINN, with approximately 0.86, and 3BEM_FRP and GFAS, with approximately 0.85, which indicate that similar emission methods tend to agree with each other. Furthermore, the comparison of the 3BEM and FINN fire inventories reveal linear correlation coefficients between 0.75 and 0.92, with larger total emissions of CO in FINN than in 3BEM for grids 3, 4, 6, and 8. 3BEM_FRP and GFAS show linear correlation coefficients between 0.65 and 0.95, with more emissions of CO in GFAS than in 3BEM_FRP for grids 1, 6, and 8.

During the SAMBBA campaign, the assessment of simulated AOT_{550nm} from MACC/ECMWF and CCATT-BRAMS operational atmospheric chemistry transport models with MODIS AOT_{550nm} Terra and Aqua measurements show a good agreement related with general pattern of the time series. Also, MACC/ECMWF and CCATT-BRAMS models are capable of simulating the aerosols released by fires. However, when the intensity of the biomass burning is too high, the aerosol values simulated by the two models underestimate the MODIS measurements.

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